

1 **Contrasting Aerosol Mixing States at Inland and Coastal Sites: An**
2 **Entropy-Based Metric for CCN Activity**

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25 **Abstract**

26 Simplified assumptions of aerosol mixing states in modeling studies often
27 introduce substantial uncertainties in estimating cloud condensation nuclei (CCN)
28 concentrations (N_{CCN}) and their climatic impacts. This study systematically investigates
29 the contrasting relationships between mixing states and CCN activity by combining
30 field measurements of the hygroscopicity distribution with the algorithm of entropy at
31 two inland and coastal sites. We show distinct seasonal variations of aerosol mixing
32 state. In winter, externally-mixed particles dominated both sites, with comparable
33 mixing state indices (χ) of 0.38 ± 0.12 and 0.39 ± 0.09 respectively for coastal air and
34 inland air. However, measurements during summer periods showed pronounced
35 differences: aerosols in the coastal atmosphere exhibited a higher degree of internal
36 mixing ($\chi=0.69\pm 0.19$), whereas inland χ values only increased moderately to 0.47 ± 0.12 .
37 Aerosol mixing state is largely influenced by primary emissions and secondary
38 formation process. Externally-mixed particles originate chiefly from anthropogenic
39 emissions in inland or sea salt in coastal. During the aging process, particles become
40 more internally-mixed as the enhanced fraction of more-hygroscopic mode. A universal
41 logarithmic correlation was identified between the critical diameter (D_{cri}) characterizing
42 CCN activity and χ ($D_{\text{cri}} = -32.15\ln(\chi)+84.71$, Pearson $r = -0.74$), but with distinct
43 decrement rates for coastal vs. inland aerosols. Further analysis reveals the covariation
44 relationships between the mixing state and CCN activity: N_{CCN} exhibits heightened
45 sensitivity to fluctuations in χ at low values. These results underscore that mixing states

46 exert different control over N_{CCN} in diverse environments. Our work provides critical
47 constraints for parameterizing fine aerosols CCN activity in climate models.

48 **1. Introduction**

49 Atmospheric cloud condensation nuclei (CCNs) are complex mixtures of organic
50 and inorganic components. Their chemical and physical properties make quantifying
51 aerosol-cloud interactions challenging (Liu et al., 2018; Rosenfeld et al., 2019; Xu et
52 al., 2022, 2024; Virtanen et al., 2025), introducing uncertainties into climate effect
53 assessments (Charlson et al., 1992; Shrivastava et al., 2017; IPCC, 2021; Chen et al.,
54 2022a; Manavi et al., 2025). Accurate climate model predictions of aerosol impacts
55 require understanding aerosol mixing states under different atmospheric conditions and
56 their effects on CCN activity (Ching et al., 2016; Zheng et al., 2021a). Current models
57 often oversimplify mixing states by assuming pure internal or external mixing (Winkler,
58 1973; Stevens et al., 2019; Riemer et al., 2019; Zheng et al., 2021b). This is problematic
59 because mixing states directly determine particle hygroscopicity distribution and CCN
60 estimates (Wang et al., 2010; Tao et al., 2024). For example, internal-mixed aerosol
61 particles have unimodal hygroscopicity distribution, while the external-mixed particles
62 are characterized by the bimodal/trimodal or partly overlapping structures (Spitieri et
63 al., 2023; Liu et al., 2025). Such simplifications can lead to significant errors, e.g.,
64 Sotiropoulou et al. (2007) found that mixing state assumptions caused two-fold N_{CCN}
65 estimation errors in global models.

66 Systematic observations across diverse environments are critical because aerosol
67 mixing states exhibit pronounced spatial-temporal variations (Ye et al., 2018; Hughes

68 et al., 2018; Liu et al., 2025). For example, continental and coastal regions present
69 contrasting scenarios (Ramachandran et al., 2016). The continental areas are dominated
70 by anthropogenic emissions, where aerosol aging is driven by industrial and traffic-
71 related pollutants (Huang et al., 2014; Ren et al., 2023). Particles here undergo
72 progressive internal mixing via photochemical process and heterogenous reactions,
73 altering their hygroscopic properties (Ervens et al., 2010; Tao et al., 2021). While the
74 coastal regions feature dynamic interactions between marine aerosols (e.g., sea salt) and
75 continental pollutants (Schill et al., 2015; Collins et al., 2013; Cheung et al., 2020).
76 Seasonal shifts in air mass sources (e.g., marine vs. continental dominance) might
77 create unique mixing state patterns (Xu et al., 2020, 2021a). For instance, summer
78 photochemical aging and heterogenous processes in coastal areas can enhance the
79 degree of internal mixing, while winter often retains more external mixing due to the
80 presence of the sea-salt particles with less-hygroscopic organic matter.

81 The continental aerosols influence regional cloud formation, while coastal
82 aerosols may provide insights into the characteristics of marine aerosols in region. The
83 properties of marine aerosols are significantly different from those of continental
84 aerosols, and therefore have distinct climate feedback mechanisms (Bellouin et al.,
85 2020; Xu et al., 2024; Liu et al., 2024). However, the current models lack regional-
86 specific mixing state parameters and usually assume uniform mixing in both
87 environments. This could lead to large uncertainties in predicting CCN concentrations,
88 highlighting the need for site-specific observations. For example, Ren et al. (2018)
89 found that the impact of aerosol mixing state on CCN activation characteristics ranged

90 from -34% to +16 % in urban atmosphere. Comparison between a fully internal mixture
91 assumption and using the mixing state index from the particle-resolved model, Ching
92 et al. (2017) found the obvious overestimation in CCN concentration estimation.
93 Especially in the regions eg., Amazon Basin, Central Africa and Indonesia, the particles
94 appeared to be more external, errors in CCN concentration would increase up to 100%
95 (Hughes et al., 2018). A detailed exploration of mixing state on CCN concentration in
96 global scale was conducted by Zheng et al. (2021a), and the results showed that the
97 mixing state varied spatially with more externally mixed over the North Atlantic Ocean,
98 off the coasts of Southern Africa, and Australia. Thus, assuming particles with
99 internally-mixed would introduce errors in CCN concentration of 50-100%.

100 Therefore, for quantifying the aerosol mixing state in the ambient atmosphere,
101 we apply the algorithm of entropy proposed by Riemer and West (2013) to investigate
102 the aerosol heterogeneity. This index has been applied to quantify the mixing state more
103 reasonably both in field campaigns (Zhao et al., 2021; Yuan et al., 2023) and model
104 simulations (Ching et al., 2016; Zheng et al., 2021a). However, most studies focused
105 on quantifying the particle heterogeneity in composition (Ching et al., 2019; Fierce et
106 al., 2020; Zhao et al., 2021). Here we concentrated on evaluating the heterogeneity in
107 aerosol hygroscopicity for sub-micron particles, which directly related to CCN budget.
108 By refereeing to Yuan et al. 2023, the heterogeneity in hygroscopicity was investigated
109 by combining in-situ measurements of probability distribution function of the
110 hygroscopicity with the algorithm of entropy. Briefly, the mixing state index χ , is
111 devised based on the concept of information entropy concerning the distribution of

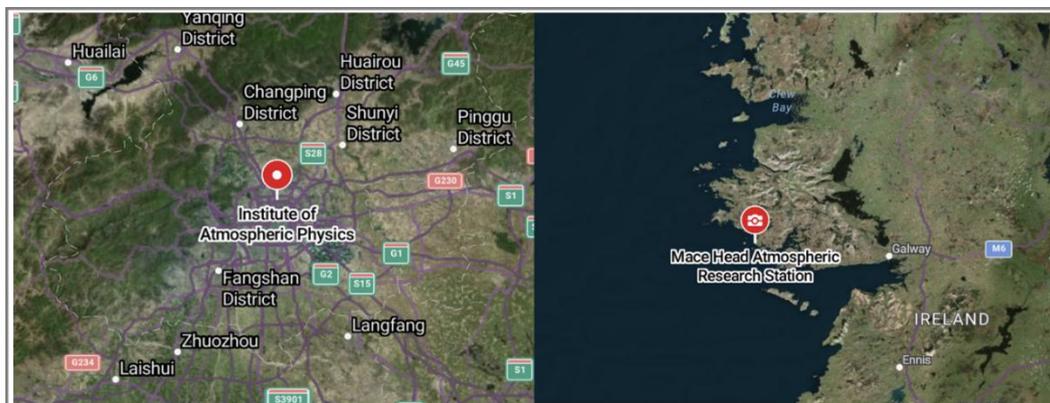
112 hygroscopicity across the aerosol population. It varies between 0 (external mixing
113 completely) and 1 (internal mixing completely). By integrating inland and coastal
114 measurements, this study will focus on addressing two key gaps, (1) How continental
115 vs. marine-dominated environments shape aerosol mixing states and CCN activity; (2)
116 Whether χ -based CCN parameterizations show regional dependencies, providing
117 critical constraints for climate models.

118 **2. Data and Methods**

119 **2.1 Field Campaigns**

120 The inland atmospheric measurements were conducted for two campaigns from
121 16 November to 6 December 2016 and 29 May to 13 June 2017 as a part of the Air
122 Pollution and Human Health (APHH) project (Shi et al., 2019), at the Institute of
123 Atmospheric Physics, Chinese Academy of Sciences (IAP, 39.97° N, 116.37° E) in
124 urban Beijing. The campaigns were complemented by the hygroscopicity and CCN
125 observations and were conducive to provide information on the aerosol hygroscopicity
126 affecting urban pollutions. This urban site exhibited highly variable aerosol populations
127 dominated by local anthropogenic sources including vehicular, cooking emissions, and
128 residential heating. Coastal measurements were performed at the Mace Head
129 atmospheric research station (MHD, 53.33° N, 9.90° W) from 1 November 2009 to 30
130 January 2010, and summer periods from 11 to 31 August 2009 and July 2010, which
131 located on the west coast of Ireland. Aerosol particles here experience alternating
132 influences from polluted continental and clean marine atmospheres. The map of the

133 sites was shown in Figure 1. More details about the campaigns were given in Fan et al.
134 (2020) and Xu et al. (2021a).



135
136 **Fig 1.** Map of the sites in the Inland of the Institute of Atmospheric Physics (IAP) and
137 Coastal of Mace Head (MHD). (© Google Maps, <https://maps.google.com/>, last access:
138 2 April 2025).

139 2.2 Instrumentation

140 Hygroscopicity measurements

141 The particle hygroscopicity at both sites was characterized using the humidified
142 tandem differential mobility analyzer (HTDMA). The hygroscopic growth factor (Gf),
143 defined as the ratio of the particle diameter at the fixed RH (90%) and dry diameter set
144 in this study for 40, 80, 110, 150, 200 nm at IAP and 35, 50, 75, 110 and 165 nm at
145 MHD, respectively. The RH calibration with ammonium sulfate for HTDMA system
146 was given in Fig. S1. The Gf probability density function (Gf-PDF) was derived using
147 the TDMAinv algorithm (Gysel et al., 2009). The number fraction (NF) of near-
148 hydrophobic mode (NH: $Gf \leq 1.21$), and more hygroscopic mode (MH: $Gf > 1.21$) in
149 IAP site was referred from Chen et al. (2022b). It was integrated into three modes for

150 the MHD site with the near-hydrophobic mode (NH: $1 < Gf < 1.3$), more hygroscopic
 151 mode (MH: $1.3 \leq Gf < 1.85$) and sea salt mode (SS: $Gf \geq 1.85$) for further examination
 152 (Xu et al., 2021a).

153 Here for each particle size, the hygroscopicity parameter κ can be subsequently
 154 calculated using κ -Köhler theory (Petters and Kreidenweis, 2007):

$$155 \quad \kappa = (Gf^3 - 1) \cdot \left[\frac{1}{RH} \exp\left(\frac{4\sigma_s M_w}{RT\rho_w D_d Gf}\right) - 1 \right] \quad (1)$$

156 where RH is the HTDMA relative humidity (90% set in the instrument), $\sigma_{s/a}$ is the
 157 surface tension of pure water (0.072 mN m^{-1}), M_w and ρ_w are the molecular weight and
 158 the density of pure water, R is the gas constant, and T is the absolute temperature, D_d is
 159 the droplet diameter.

160 Then, the κ -PDF is obtained and normalized as $\int_0^\infty c(\kappa)d\kappa = 1$, where $c(\kappa)$ is
 161 normalized as κ -PDF. Further it was used to calculate the particle population
 162 heterogeneity (Calculation seen in Section 2.3).

163 According to the κ -Köhler theory, the critical diameter (D_{cri}) corresponding to the
 164 supersaturation ratio can be expressed as:

$$165 \quad D_{cri} = \sqrt[3]{\frac{4A^3}{27\kappa \ln^2 S}}, \quad A = \frac{4\sigma_s M_w}{RT\rho_w} \quad (2)$$

166 where S is the given supersaturation ratio (here 0.2% used in this study), κ is the mean
 167 value of the hygroscopicity parameter calculated in Equation (1).

168 **Chemical components**

169 For the inland atmospheric measurements, the non-refractory submicron aerosol
 170 (smaller than $1\mu\text{m}$, NR-PM₁) chemical composition was quantitatively characterized
 171 using the Aerodyne High-Resolution Time-of-Flight Aerosol Mass Spectrometer (HR-

172 ToF-AMS) (DeCarlo et al., 2006), including sulfate (SO_4^{2-}), nitrate (NO_3^-), ammonium
173 (NH_4^+), chloride (Chl) and organics (Org). The black carbon (BC) mass concentration
174 was determined from the light absorption with a seven-wavelength aethalometer (AE33,
175 Magee Scientific Corp.).

176 Measurements of PM_{10} in the coastal atmosphere were also performed by the HR-
177 ToF-AMS, including major inorganic salts (non-sea-salt sulfate, nss-SO_4^{2-} ;
178 methanesulfonic acid, MSA; NO_3^- ; NH_4^+) and organic matter. The sea salt was
179 quantified using NaCl ion signal, which has been demonstrated in previous studies
180 (Ovadnevaite et al., 2014). The instrument operation and calibration have been
181 described in previous studies (Ovadnevaite et al., 2014; Xu et al., 2019).

182 **Aerosol number size distribution and CCN number concentration**

183 Particle number size distributions (PNSD) were measured using an integrated
184 system consisting of a Differential Mobility Analyzer (DMA; model 3081, TSI Inc.)
185 coupled with a Condensation Particle Counter (CPC; model 3772, TSI Inc.). During the
186 measurements at IAP, the PNSD covered the size range of 10-550 nm with a 5-minute
187 time resolution. It scanned size range of 20-500 nm at MHD with a 10-minute temporal
188 resolution. The CCN number concentrations were quantified at both sites using a
189 Droplet Measurement Technologies CCN counter (DMT-CCNc) (Lance et al., 2006).
190 The instrument's supersaturation (SS) settings were carefully calibrated before and after
191 each campaign using ammonium sulfate aerosol following Rose et al. (2008) (Fig. S1).
192 Four effective supersaturations (SS) were 0.14%, 0.23%, 0.40% and 0.76% at IAP site.
193 Four SS levels were 0.25%, 0.5%, 0.75% and 1% at MHD site with an uncertainty of

194 $\pm 0.03\%$. Using measurements at set supersaturation of 0.2% as an example explores the
195 CCN activity in the following discussions.

196 **2.3 Calculation the heterogeneity for aerosol particles**

197 To characterize the heterogeneous distribution of the hygroscopic and non-
198 hygroscopic components in populations (Chen et al., 2022b), we calculated the mixing
199 state index (χ) using the κ -PDF, following the methodology of Yuan et al. (2023). Two
200 surrogate groups in a population of N aerosol particles were assumed (Zheng et al.,
201 2021a). One surrogate group consists the non- and/or slightly hygroscopic species with
202 κ_N of < 0.05 and another group contains the more hygroscopic species with κ_H of 0.5-
203 0.6 (Yuan et al., 2023, referred inorganics). Ambient particles typically contain one or
204 two of the components and the κ lies between 0 and 0.6 at IAP or 0.8 at MHD as shown
205 in Figure S2. Taking into account the enhanced hydrophilicity of marine aerosols at
206 MHD site, calculation assuming κ_H values of 0.7 and 0.8 were shown in Fig.
207 S3. While these variations in κ_H introduced a mean uncertainty of 8% in χ values, it
208 did not significantly affect the seasonal or site comparisons. The volume fraction of two
209 surrogate groups can be calculated based on the total κ according to the Zdanovskii–
210 Stokes–Robinson (ZSR) mixing rule (Zdanovskii, 1948; Stokes et al., 1966).

211 The mixing state index χ is defined as the affine ratio of the average particle species
212 diversity ($D\alpha$) and population species diversity ($D\gamma$) as:

$$213 \chi = \frac{D\alpha - 1}{D\gamma - 1} \quad (3)$$

214 The average per-particle species diversity $D\alpha$ can be calculated as follows. First,

215 the mixing entropies at bin i (H_i) are determined according to equation (4),

$$216 \quad H_i = -P_{i,N} \times \ln P_{i,N} - P_{i,H} \times \ln P_{i,H} \quad (4)$$

217 where $P_{i,N}$ and $P_{i,H}$ are the volume fraction of each group for the κ -PDF with X bins at
 218 bin i ($i=1,2,\dots,X$), and can be determined from the $P_{i,N} + P_{i,H} = 1$ and $P_{i,N} \times \kappa_N +$
 219 $P_{i,H} \times \kappa_H = \kappa_i$. Here $\kappa_N = 0.01, \kappa_H = 0.6$; κ_i represents the hygroscopicity parameter
 220 at bin i .

221 Based on the assumption that particles in the same diameter have the same mixing
 222 entropy $H_\alpha = \sum_{j=1}^N P_j \times H_j$, $P_j = \frac{V_j}{V_{total}} = \frac{1}{N}$; the per-particle mixing entropies H_α is
 223 determined according to equation (5),

$$224 \quad H_\alpha = \sum_{i=1}^X H_i \times c(\kappa)_i \times \Delta\kappa \quad (5)$$

225 where $c(\kappa)_i$ is the probability density of the normalized κ -PDF at bin i , and $\Delta\kappa$
 226 represents the bin width. Then, the average per-particle species diversity D_α can be
 227 determined as $D_\alpha = e^{H_\alpha}$;

228 The bulk population species diversity D_γ can be calculated as follows. First, the
 229 aerosol population of the mixing entropy can be calculated as equation (6):

$$230 \quad H_\gamma = -P_N \times \ln P_N - P_H \times \ln P_H \quad (6)$$

231 where P_N and P_H are the volume fraction of the non-hygroscopic and hygroscopic
 232 components in the population, and can be calculated by equation (7) and (8):

$$233 \quad P_N = \sum_{i=1}^X P_{i,N} \times c(\kappa)_i \times \Delta\kappa \quad (7)$$

$$234 \quad P_H = \sum_{i=1}^X P_{i,H} \times c(\kappa)_i \times \Delta\kappa \quad (8)$$

235 Then, the bulk population species diversity D_γ can be determined as $D_\gamma = e^{H_\gamma}$.

236 Here, the definition of surrogate species as supersets encompassing hygroscopicity

237 heterogeneity implies that the heterogeneity parameter χ ranges from 0 to 1. When the
238 mixing index χ approaches 0, it indicates a completely segregated state where
239 hygroscopic and non-hygroscopic species reside in distinct particles. While for the case
240 the mixing index χ to be 1 represents that the non-hygroscopic and hygroscopic species
241 distributing homogeneously throughout the aerosol population.

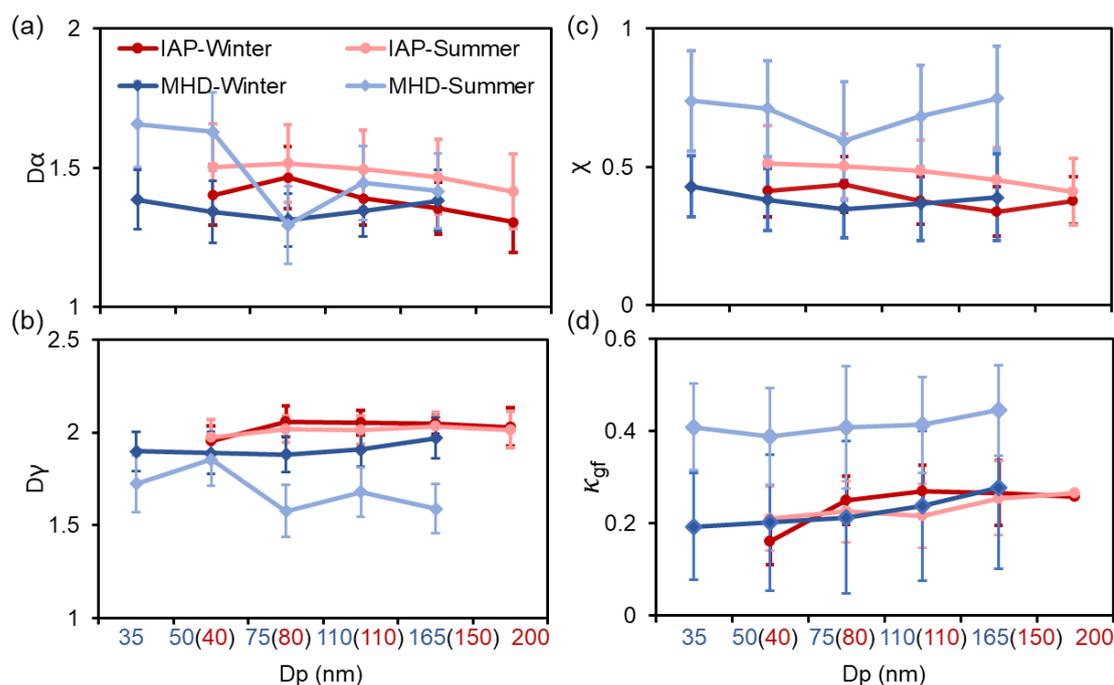
242 **3. Result and Discussion**

243 **3.1 Comparison of the heterogeneity in the inland and coastal atmosphere**

244 To characterize the hygroscopic heterogeneity of atmospheric aerosols, Figure 2
245 depicts variations in mixing state metrics ($D\alpha$, $D\gamma$, χ) and the hygroscopic parameter
246 (κ_{gf}) across particle size distributions. The $D\alpha$ and χ decrease with increasing particle
247 diameter, accompanied by higher κ_{gf} values at IAP site. This trend indicates that inland
248 particle populations tend to homogenize into hygroscopic compositions through
249 primary particle aging or secondary formation processes (Liu et al., 2025; Chen et al.,
250 2022b; Zhong et al., 2022). In contrast, particles exhibit a non-monotonic pattern at
251 MHD site: $D\alpha$ and χ decrease for Aitken-mode particles (<100 nm) but increase for
252 accumulation-mode particles. The κ_{gf} shows consistent size-dependent increases in
253 both winter and summer campaigns.

254 Notably, the mixing state metrics exhibit a pronounced minimum at 75 nm
255 particles, influenced by distinct mechanisms: winter minima reflect the high sea salt
256 fraction, while summer minima are driven by anthropogenic organic matter (Cheung et
257 al., 2020; Xu et al., 2021a). Lower winter χ values—coupled with broader κ -PDF

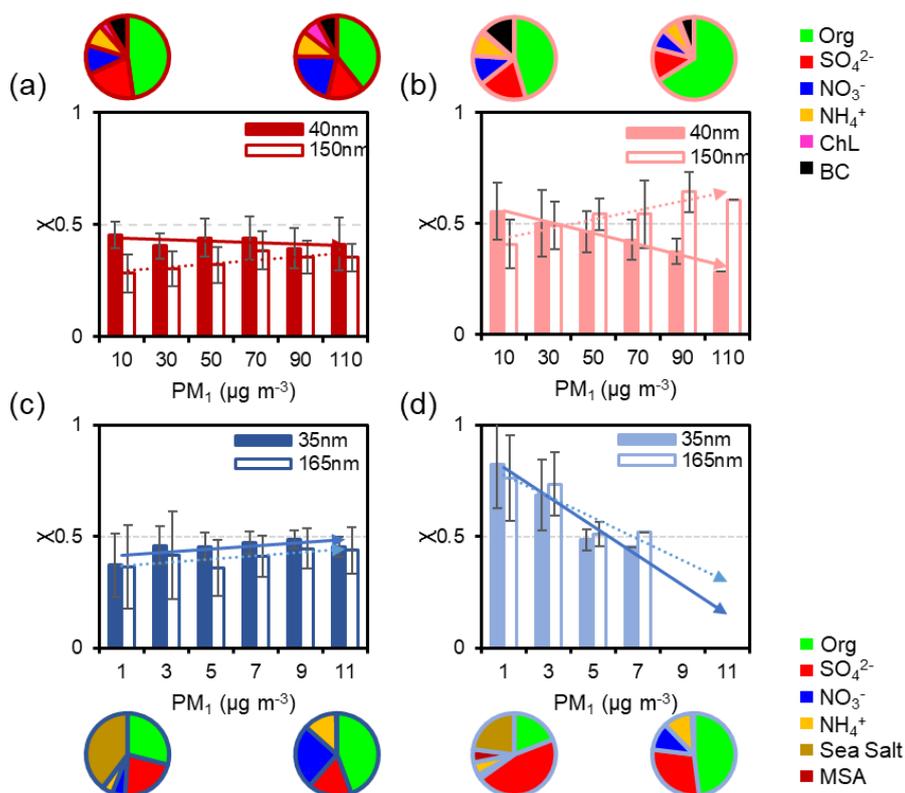
258 distributions—indicate stronger external mixing and compositional diversity compared
 259 to summer (Fig. S2). Seasonal χ and κ_{gf} disparities are more pronounced at MHD site,
 260 primarily driven by the seasonal alternation of marine and anthropogenic emission
 261 sources.



262
 263 **Fig 2.** Mean values of the $D\alpha$ (a), $D\gamma$ (b), χ (c) and κ_{gf} (d) for aerosols of five diameters
 264 during winter and summer periods at IAP and Mace Head sites.

265 Ultrafine particles (40 nm in IAP vs. 35 nm in MHD, Aitken mode) and larger
 266 particles (150 nm in IAP vs. 165 nm in MHD, accumulation mode) are selected to
 267 investigate distinct evolutionary processes of aerosol heterogeneity (Fig. 3 and Fig. S4).
 268 With the increasing of PM_{10} concentration during winter, the variation in χ values exhibit
 269 only minor both at the IAP and MHD sites, generally fluctuating between
 270 approximately -0.04 and 0.08 (Fig. 3a and c). Inland accumulation-mode particles
 271 show a modest increase in χ , corresponding with a higher proportion of inorganic salts.

272 Conversely, at MHD site, the composition fraction shifts from a sea-salt dominance



273

274 **Fig 3.** Variation of the average χ for 40 nm and 150 nm particles at IAP and 35 nm and
 275 165 nm at MHD site with the particle mass concentration in IAP-winter (a), IAP-
 276 summer (b), MHD-winter (c) and MHD-summer (d). The pie charts represent the
 277 average mass fraction during four field measurements.

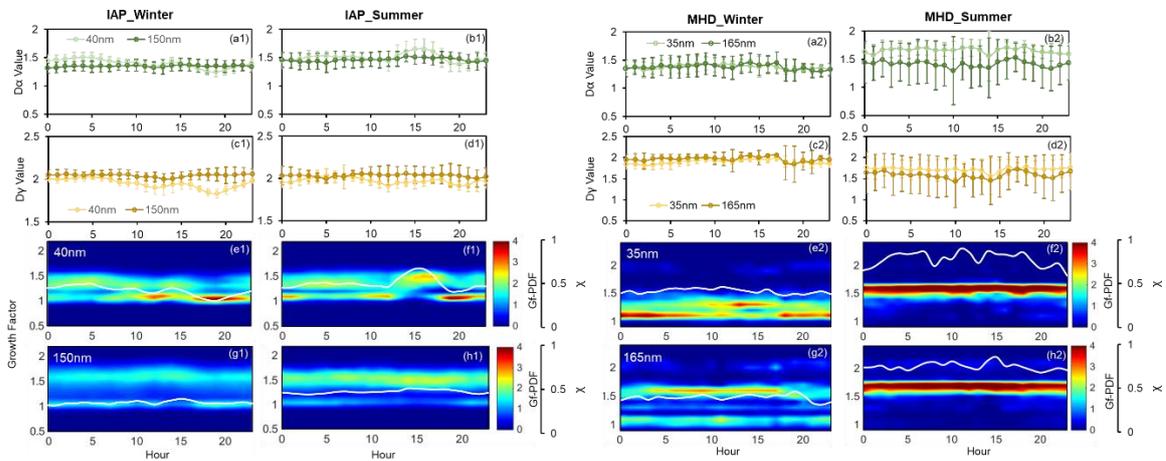
278 toward organic matter, accompanied by a ~20 % increase in nitrate content (Fig. 3c). In
 279 summer, the variation of χ with PM concentration becomes markedly pronounced at
 280 both IAP and MHD stations. For example, χ for 40 nm particles decreases as PM
 281 increases at IAP site (Fig. 3b). The elevated particle heterogeneity mainly arises from
 282 the locally primary emissions, corresponding to the enhanced primary organic
 283 emissions as shown in Fig. S5. It appeared more pronounced during evening rush hours.
 284 In contrast, χ for 150 nm particles increases from ~0.40 to ~0.60 with rising PM,

285 reflecting enhanced secondary formation and internal mixing during pollution process
286 that render the particle population more homogeneous. At coastal sites, χ declines with
287 rising PM by approximately 0.37 for 35 nm particles and 0.24 for 165 nm particles,
288 mirroring the shift in chemical composition makeup from inorganic dominance to
289 greater organic content (Fig. 3d).

290 Diurnal variations of mixing state metrics ($D\alpha$, $D\gamma$, Gf-PDF and χ) at IAP and
291 MHD sites are shown in Figure 4. In IAP-winter, particles exhibited steeper declines in
292 $D\alpha$ and χ during evening rush hours than summer, indicating a higher fraction of non-
293 hygroscopic particles (40 nm) from fresh traffic emissions (Fig. 4a1 and S5).
294 Concurrently, reduced $D\gamma$ values suggest that the bulk population consists of uniformly
295 distributed less-hygroscopic (LH) components (Fig. 4c1). Aitken mode particles
296 showed bimodal and broader Gf-PDF distributions, corresponding to cooking activities
297 (11:00–13:00 LT) and traffic peaks (17:00–20:00 LT) (Cai et al., 2020). Midday
298 photochemical aging promoted more internally mixed aerosols (Yang et al., 2012; Liu
299 et al., 2025), as evidenced by increasing $D\alpha$ at the urban site (Fig. 4b1). Conversely, the
300 χ for accumulation-mode particles showed minimal diurnal variations both in IAP-
301 winter and IAP-summer. This is mainly due to the dominant hygroscopic mode for 150
302 nm particles (Fig. 4g1 and 4h1), especially during summer, which is mainly from
303 secondary formation or aging of the primary particles (such as the transformation from
304 primary organic aerosol (POA) to secondary organic aerosol (SOA) in Fig. S5) (Wang
305 et al., 2019; Fan et al., 2020).

306 For the coastal atmosphere, the mixing state metrics ($D\alpha$, $D\gamma$, and χ) of Aitken and

307 accumulation mode particles in winter exhibited analogous diurnal patterns,
 308 characterized by a descending trend at nightfall (Fig. 4a2-h2). This corresponds to an
 309 enhanced modal distribution of near-hydrophobic (NH) particles at 35 nm and more-
 310 hygroscopic (MH) particles at 165 nm. In summer, $D\alpha$ and $D\gamma$ both trended downward
 311 during daytime, with the decline of $D\gamma$ being more pronounced. A conspicuous seasonal
 312 discrepancy between Aitken and accumulation mode particles was observed in this
 313 region (Fig. 4a2–h2), where the mixing state index χ increased incrementally from
 314 winter to summer. Specifically, the mean χ for 35 nm particles escalated from 0.42 to
 315 0.80, and for 165 nm particles, it rose from 0.39 to 0.76. This trend demonstrates a
 316 strong alignment with the spread factor (used as a measure of particle mixing state)
 317 documented by Xu et al. (2021a).



318

319 **Fig 4.** Diurnal variation of $D\alpha$, $D\gamma$, Gf-PDF, and χ during winter and summer periods
 320 for 40 nm and 150 nm aerosols at IAP (a1-h1) and for 35 nm and 165 nm aerosols MHD
 321 site (a2-h2).

322 Similar to the mixing state χ , a very clear seasonal pattern of the aerosol

323 hygroscopic distribution was found (Fig. 4e-h). In winter, the Gf-PDF diurnal profiles
324 of both Aitken and accumulation mode particles showed bimodal distribution (Fig. 4e2-
325 g2) as evident by the number fraction of nearly-hydrophobic and more hygroscopic
326 modes (Fig. S6). The NH mode is likely composed of anthropogenic organic matter and
327 biogenic species derived from marine air masses (Xu et al., 2020), with this
328 composition being particularly prominent in the Aitken mode. Additionally, Xu et al.
329 (2021a) observed a higher abundance of the NH-mode in marine polar and Arctic air
330 masses, this further supports the notion that NH particles are likely of biogenic origin,
331 aside from contributions from anthropogenic activities. The more hygroscopic and sea
332 salt mode was mostly contributed from the nss-sulfate and sea salt in winter (Xu et al.,
333 2021a). Analogously, accumulation mode particles with a higher proportion of MH and
334 SS mode (Fig. S6) primarily attributed to the prevalence of non-sea-salt sulfate (nss-
335 sulfate) and sea salt in the coastal atmosphere (Xu et al., 2020). The bimodal and broad
336 of hygroscopic distribution suggested that particles were more diverse and external
337 mixed, consistent with the lower χ value in winter.

338 In contrast, summer observations revealed that Gf-PDFs of both Aitken and
339 accumulation mode particles transitioned to unimodal distributions, signifying particles
340 in summer had more homogeneous composition with a large extent of internal mixing
341 particles (with higher χ). Such diurnal trend in Gf-PDFs was consistent along with the
342 high number fraction of MH-mode and low NH-mode (Fig. S6). The higher
343 hygroscopicity and MH mode in summer were largely driven by the enhancement of
344 sulfate and decrease of organic matter (Fig. S6). And a clear shift from NH to MH mode

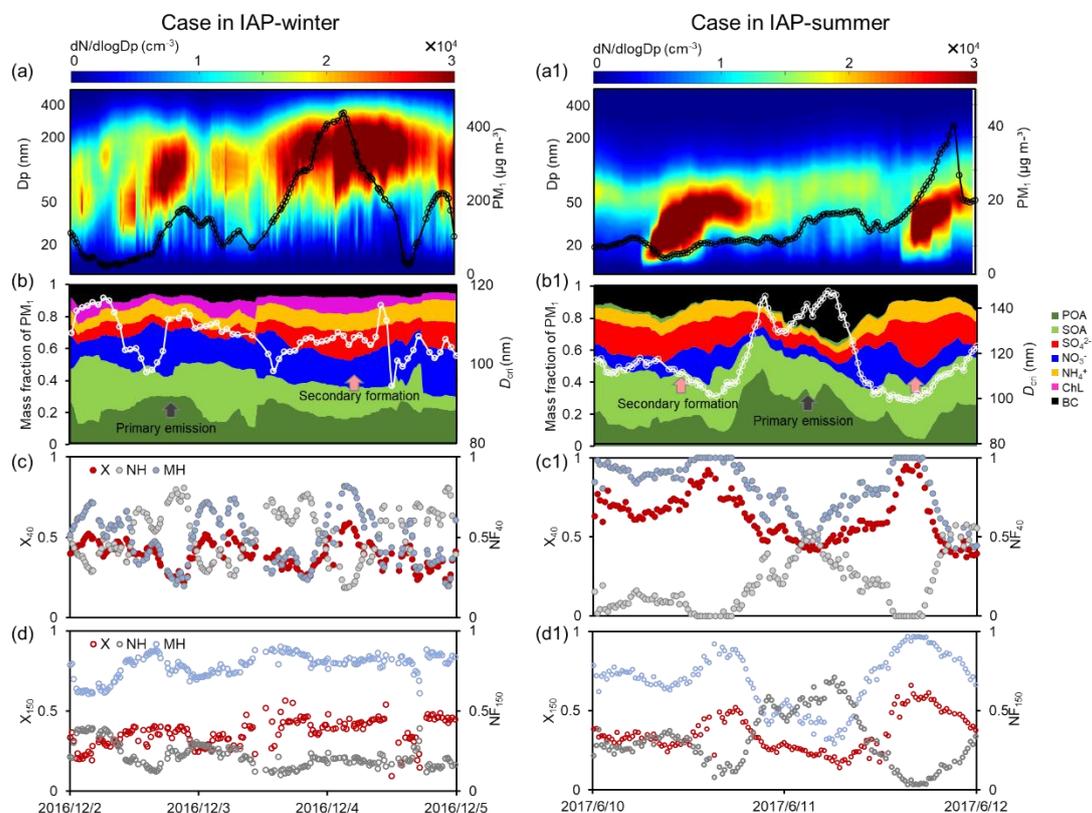
345 at midday might further demonstrate the promotion of photochemical aging in summer
346 (Xu et al., 2021a).

347 **3.2 Impacts of Primary Aerosol Emissions and Secondary Aerosol Formation on** 348 **Aerosol Mixing State**

349 As already noted above, changes in χ were clearly associated with the chemical
350 composition varying with site and season. The relationships between the mixing state
351 index and the number fraction of hydrophobic and hygroscopic mode during four
352 campaigns are presented in Figure S7. The χ exhibited negative correlations with the
353 fraction of hydrophobic mode but a positive relationship with the fraction of
354 hygroscopic particles, highlighting the markedly different effects of the primary
355 emissions and secondary formation on aerosol mixing state (Tao et al., 2024). To gain
356 more insight on this effect between inland and coastal atmosphere, four case are
357 analyzed (Fig. 5 and 6): case for IAP-winter, case for IAP-summer, case for MHD-
358 winter and case for MHD-summer.

359 Case for IAP-winter is a heavy polluted event with the mean PM mass
360 concentration increased from 22 to 437 $\mu\text{g m}^{-3}$ (Fig. 5a-d). The 40- and 150-nm χ
361 patterns shifted quickly during the pollution periods. With the mass fraction of
362 hydrophobic compounds (ie., POA) in PM_{10} increased, the χ of 40-nm particles
363 decreased from 0.5 to 0.2, that is, an enhanced NH mode and a weaken MH mode (Fig.
364 5b-c). At this stage, large particles for 150 nm are mainly from aqueous formation with
365 more proportion of nitrate. The corresponding χ of 150 nm was higher. While with that
366 the mass fractions of secondary organic and inorganic compositions increased, particles

367 were more internal mixed with χ increased to be 0.6 for 40-nm and 0.53 for 150-nm
 368 particles.

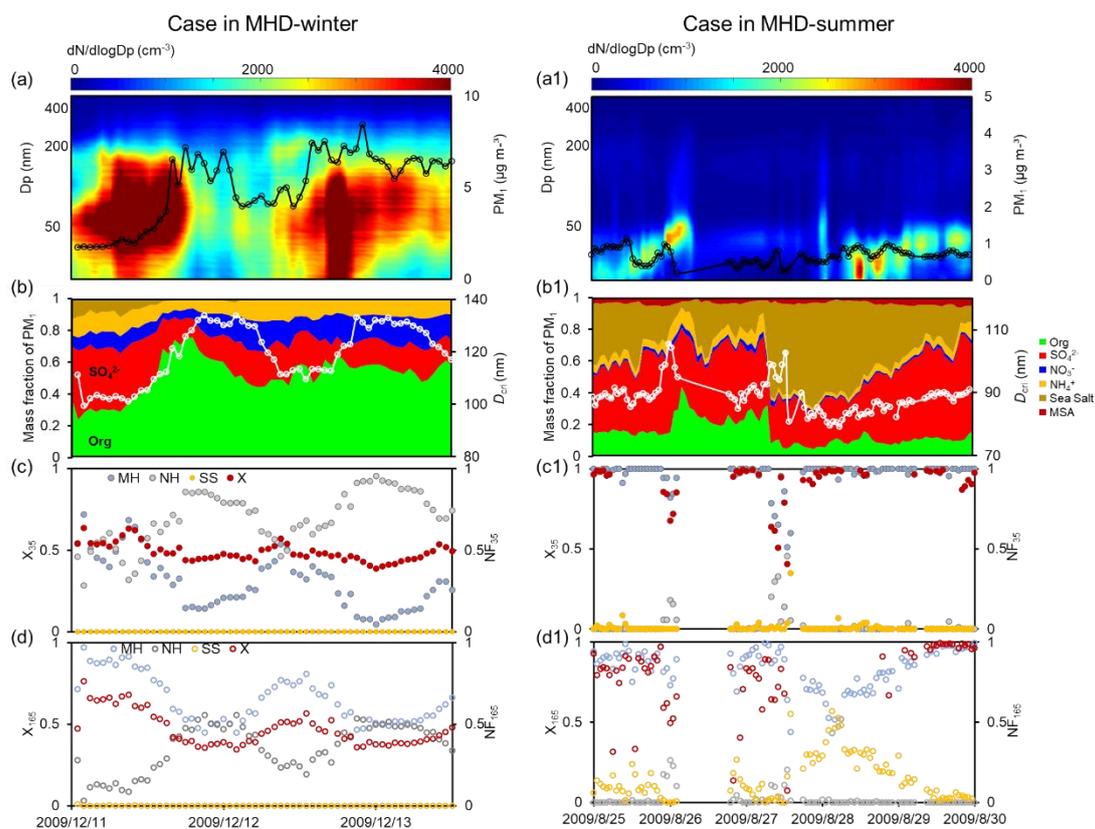


369

370 **Fig 5.** Case in IAP-winter (a-d) and IAP-summer (a1-d1). Particle number size
 371 distribution and PM_{10} (a and a1), mass fraction of the PM_{10} and the critical diameter (D_{cri})
 372 (b and b1), mixing state index (χ), number fraction of the nearly hydrophobic mode
 373 (NH) and more hygroscopic mode (MH) for 40 nm particles (c and c1), χ , NH and MH
 374 for 150 nm particles (d and d1).

375 Case for IAP-summer is the typical new particle formation events (NPF) with the
 376 mean PM_{10} of $13 \mu\text{g m}^{-3}$ (Fig. 5a1-d1). With the evolution of NPF events, the χ of 40-
 377 and 150-nm particles increased to be 0.95 and 0.61 with the enhanced proportion of
 378 more-hygroscopic components (ie., SOA, NO_3^- , SO_4^{2-}). The χ pattern is opposite of that

379 the number fraction of NH mode and consistent with the variation of MH mode (Fig.
 380 S7). Note that a sudden decrease in χ on June 11th was disturbed by the strong primary
 381 emission. The chemical mass fractions showed more POA and black carbon with an
 382 enhanced NH mode and a weaker MH mode (Fig. 5b1-d1). The χ of 40-nm particles
 383 decreased to be 0.4 and that of the 150-nm particles decreased to be 0.2. The χ patterns
 384 appear to similar transitions for Aitken and accumulation-mode particles during haze
 385 and NPF events. The increase in χ is synchronous with the increase in MH mode from
 386 secondary formation but opposite with that of LH mode from primary emissions. This
 387 implies that the primary emissions would lead particles more external mixing while
 388 secondary formation would promote aerosol more internal mixed in Inland atmosphere.



389

390 **Fig 6.** Case in MHD-winter (a-d) and MHD-summer (a1-d1). Particle number size
 391 distribution and PM_1 (a and a1), mass fraction of the PM_1 and the critical diameter (D_{cri})

392 (b and b1), mixing state index (χ), number fraction of the nearly hydrophobic mode
393 (NH) and more hygroscopic mode (MH) for 35 nm particles (c and c1), χ , NH and MH
394 for 165 nm particles (d and d1).

395 Case for MHD-winter is a high organic matter pollution event with the mean PM_{10}
396 of $5.2 \mu\text{g m}^{-3}$ and 52% mass fraction of organics (Fig. 6a-d). Larger presence of
397 anthropogenic organic matter resulted the NH mode for 35-nm particles to be 95% and
398 165-nm particles to be 53% (Fig. 6). The χ of 35- and 165-nm particles decreased with
399 the NH mode increased (Fig. S7), similar with the case for IAP site. There was a steady
400 increase in χ when the MH-mode particles started increasing with the increase in mass
401 fraction inorganics, eg., 35 nm particles showed the mean χ increasing from 0.43 to
402 0.57 and 165 nm particles from 0.35 to 0.6. This indicated that the trend of aerosol
403 mixing state closely followed the evolution emission and secondary formation.

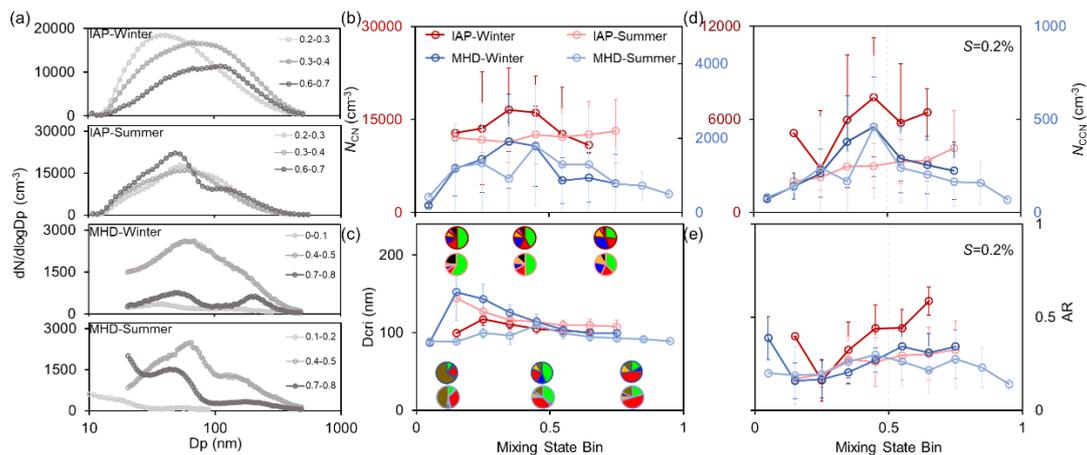
404 Case for MHD-summer is an extremely clean event with the mean PM_{10} of $0.7 \mu\text{g}$
405 m^{-3} (Fig. 6a1-d1). The dominated MH mode was found throughout the case, which
406 could be attributed from the high mass fraction of nss-sulfate (41% average). Compared
407 with the case in MHD-winter, the mean proportion of organic has decreased to be 15%.
408 Therefore, the χ remains at a high value (mean χ of 0.9 for 35-nm and 0.8 for 165-nm
409 particles). Until August 28th, a stronger increase in the mass fraction of sea salt and
410 accordingly SS mode in larger-size particles was observed. The χ decreased rapidly
411 with the decrease in MH mode and enhanced SS mode, especially for the accumulation
412 mode particles, suggesting the sea spray production makes particles more externally
413 mixed.

414 In summary, these results suggest that the primary emission and secondary
415 formation drive the hygroscopic distribution and can result in significant variation of
416 aerosol mixing state χ both in Inland and coastal atmosphere. The pattern of χ varied
417 among site and season, highlighting the importance of considering the impact of mixing
418 state on CCN activity.

419 **3.3 Covariation relationships between the Mixing State and CCN Activity**

420 The mixing state of particle populations undergoes dynamic transformations
421 during atmospheric aging, profoundly influencing their CCN activity. Unlike prior
422 CCN closure studies that assumed mixing states based on chemical component
423 fractions (Yang et al., 2012; Padró et al., 2012; Ren et al., 2018), this work employs the
424 hygroscopicity distribution- and entropy-derived mixing state index χ to clarify the
425 covariation relationships between the mixing state and CCN activity. Given that CCN
426 activity reflects the characteristics of the entire particle population while χ is calculated
427 from HTDMA measurements at specific particle diameters, this study chooses the N_{CCN} ,
428 activation ratio (AR) at supersaturation of 0.2% and the mean χ for accumulation-mode
429 to illustrate the covariation characteristic approximately. This approach is employed to
430 ensure that both variables/parameters reflect the characteristics of the accumulation-
431 mode particle population. The covariations of particle size and chemical composition
432 with the mixing state (χ ranging from 0 to 1 at intervals of 0.1) are illustrated in Figure
433 7. This figure provides key insights into two fundamental determinants of CCN activity
434 (Dusek et al., 2006).

435 As χ increases, the peak diameter (D_{peak}) of the particle number size distribution
 436 (PNSD) shifts toward larger sizes (Fig. 7a and Fig. S8), while peak concentrations occur
 437 within the intermediate χ range (0.3–0.6). This trend indicates that CN number
 438 concentration (N_{CN}) first increases, driven by primary emissions and new particle
 439 formation, then decreases due to mixing and aging processes (Fig. 7b). Notably, new
 440 particle formation events frequently occurred in IAP-summer (Fig. S9), corresponding
 441 the gradually increase of χ . And the χ for Aitken-mode is significantly larger than the
 442 accumulation-mode particles during this period. Thus, N_{CN} exhibits a sustained slight
 443 increase as the degree of the internal mixing increases in IAP-summer.



444
 445 **Fig 7.** Comparison of the average particle number size distribution (PNSD) in different
 446 mixing state index (χ) (a), CN number concentration (N_{CN}) as a function of χ (b),
 447 Critical diameter (D_{cri}) at $S=0.2\%$ and mass fraction of chemical composition as a
 448 function of χ (c), CCN number concentration (N_{CCN}) (d) and activation ratio (AR) at
 449 $S=0.2\%$ a function of χ (e).

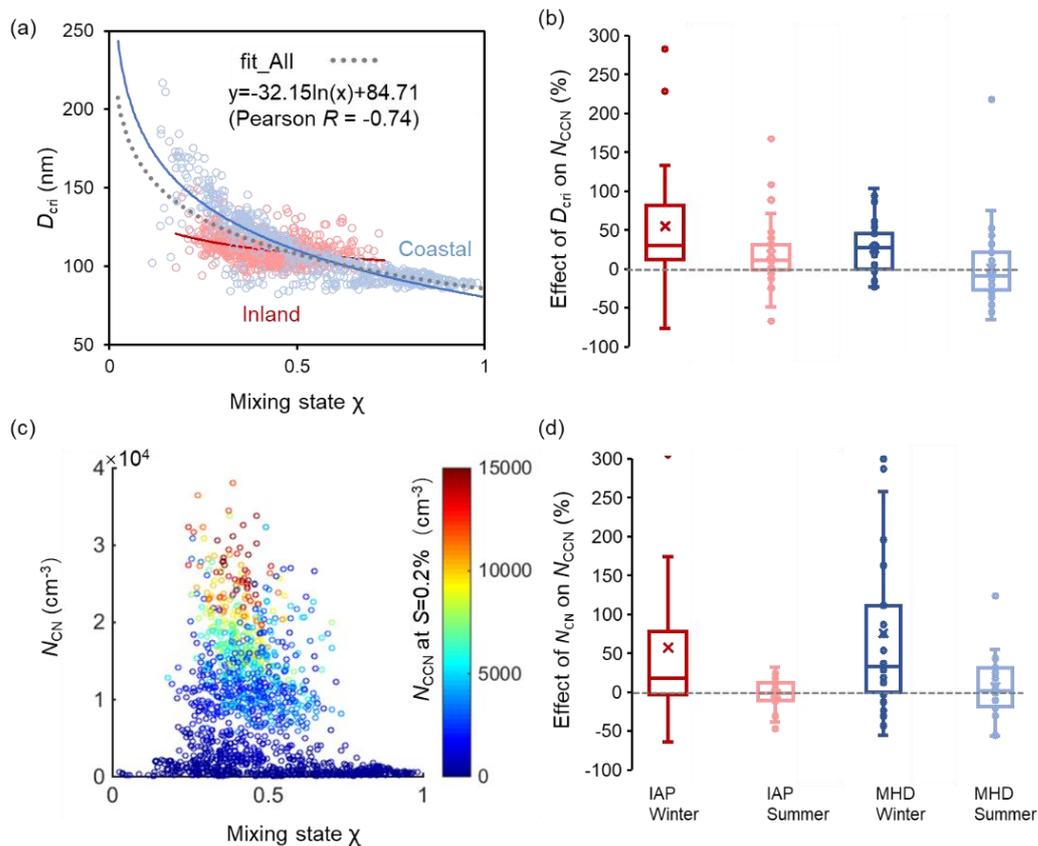
450 The critical diameter (D_{cri})—defined as the minimum size required for activation
 451 at a given supersaturation—depends on aerosol hygroscopicity. In turn, this

452 hygroscopicity is jointly determined by the hygroscopicity of individual soluble
453 components and their mass fractions in the aerosol (Petters and Kreidenweis, 2007).
454 Using measurements at supersaturation of 0.2% as an example, Fig. 7c shows that D_{cri}
455 decreases with increasing highly hygroscopic inorganic components (e.g., sulfate,
456 nitrate) in the inland atmosphere. In contrast, coastal D_{cri} exhibits nonlinear variations
457 with χ : high external mixing (low χ) elevates D_{cri} due to dominant organic components,
458 reducing sea salt particle fractions. As χ increases, the mass fraction of non-sea-salt
459 sulfate (nss-sulfate) rises, enhancing activation potential by decreasing D_{cri} .

460 The covariation characteristic of CCN activity at 0.2% supersaturation with
461 mixing state index χ reveals distinct inter-atmospheric differences, as shown in Fig. 7d-
462 e. In IAP, N_{CCN} at $S=0.2\%$ demonstrates a monotonic increasing trend with χ , attributed
463 to the synergistic effects of rising N_{CN} and decreasing D_{cri} (Fig. S10). By contrast,
464 coastal N_{CCN} follows a pattern analogous to N_{CN} , with peak concentrations shifting
465 toward higher χ values. This highlights the dominant role of particle size effects in
466 enhancing CCN concentrations under marine-influenced conditions (Dusek et al.,
467 2006).

468 Two distinct $D_{\text{cri}}-\chi$ trends underpin these disparities: one remains stable, driven by
469 the inherent hygroscopicity of sea salt, while the other exhibits steep D_{cri} declines
470 associated with anthropogenic pollution as internal mixing intensifies. These
471 discrepancies are further manifested in the nonlinear $D_{\text{cri}}-\chi$ relationship. The activation
472 ratio (AR)—quantifying aerosol cloud droplet formation potential at fixed
473 supersaturation—also varies by site (Fig. 7e). Notably, AR shows a marked increase

474 with χ in IAP-winter, likely due to enhanced N_{CCN} from the elevated inorganic fraction
 475 under higher mixing states (Fig. 3). Conversely, the inorganic fraction decreases during
 476 other sampling periods, dampening AR growth.



477
 478 **Fig 8.** Covariation characteristics of the critical diameter (D_{cri}) with the χ (a), relative
 479 change of CCN number concentration (N_{CCN}) at supersaturation $S = 0.2\%$ with the
 480 reduction in D_{cri} (b); Covariation characteristics of the CN number concentration (N_{CN})
 481 with the χ , different colors represent the N_{CCN} (c), relative change of N_{CCN} with the
 482 change in N_{CN} (d).

483 Fig. 8 further shows the covariation characteristics of CN concentration, chemical
 484 compositions (i.e., D_{cri}) with mixing state across particle populations. D_{cri} exhibits
 485 heightened sensitivity to minor χ fluctuations in the mixing state (χ) when $\chi < 0.5$ (Fig.

486 8a). In contrast, for particles already characterized by internal mixing, further increases
487 in χ (i.e., higher degrees of internal mixing) exert a negligible influence on D_{cri} . This
488 behavior pattern implies that the $D_{\text{cri}}-\chi$ relationship could enable a novel
489 parameterization for D_{cri} estimation, a framework that has not yet been reported in the
490 existing literature.

491 Aerosol observation at MHD site (blue dots) span a broad D_{cri} range (80–220 nm)
492 with χ varying from 0.1 to 1, reflecting alternating influences of highly hygroscopic
493 inorganic salts (sea salt, sulfate) and less-hygroscopic organic matter. In contrast,
494 aerosols—dominated by anthropogenic pollutants—exhibit a narrower D_{cri} range (90–
495 150 nm) at IAP site. Both environments show negative $D_{\text{cri}}-\chi$ correlations, but with
496 distinct functional forms: MHD aerosols feature an exceptional logarithmic fit ($D_{\text{cri}} = -$
497 $42.98\ln(\chi) + 80.36$, $R^2 = 0.75$; Fig. 8a blue line), while IAP aerosols (red line) yield a
498 shallower slope (-12.04). Pooling all data, we derive a generalized parameterization:
499 $D_{\text{cri}} = -32.15\ln(\chi) + 84.71$ (Pearson $r = -0.74$, $R^2 = 0.54$). As discussed above, strong
500 impact of primary emission and secondary formation on aerosol mixing state was
501 observed in both sites (Fig. 5 and 6). One could obtain more details on the $D_{\text{cri}}-\chi$
502 correlations. For example, the D_{cri} exhibited rapidly increase with the increase of
503 primary emissions (ie., mass fraction of POA enhanced) during polluted periods. The
504 D_{cri} pattern appeared opposite with that of the mixing state index, especially for the
505 accumulation-mode particles. More pronounced $D_{\text{cri}}-\chi$ correlations were observed
506 during the new particle formation (Fig. 5a1-d1). The decreasing presence of D_{cri}
507 matched the increasing proportion of SO_4^{2-} and SOA with the χ increased during NPF

508 events. Similar correlations between the critical diameter and mixing state index were
509 also found in the coastal atmosphere, especially for the case of the enhanced
510 anthropogenic organic matter and sea salt production (Fig. 6). This implies that the
511 relationship between the D_{cri} and χ might be disturbed by the variation of emission
512 pollution and secondary formation processes, resulting in spatiotemporal differences.
513 The D_{cri} reduces by 2.2–6.8% with the mixing state increase at a step of 0.1, with the
514 steepest winter decline (Fig. S11).

515 Changes in N_{CN} with differ starkly between environments: positive effects in
516 polluted inland air (+9%) versus negative effects in coastal regions (-2%). Aerosols in
517 IAP, frequently perturbed by primary emissions and new particle formation, exhibit
518 elevated N_{CN} (peaking at $\chi = 0.2\text{--}0.7$), while in MHD N_{CN} remains $\sim 5000 \text{ cm}^{-3}$ across
519 all χ . We categorized data into two groups: C1 (particles within specific N_{CN} ranges)
520 evaluates N_{CCN} covariations mainly driven by $D_{\text{cri}}\text{--}\chi$ relationships, while C2 (particles
521 within fixed D_{cri} intervals) assesses $N_{\text{CN}}\text{--}\chi$ effects (Fig. 8b). Relative changes (RC) in
522 D_{cri} , N_{CN} , and N_{CCN} with χ were calculated by comparing successive χ increments (χ_{i+1}
523 vs. χ_i , $i=0,0.1\dots 1$) within defined N_{CN} or D_{cri} windows.

524 Notably, the covariation relationship between the N_{CCN} and mixing state exerts
525 more pronounced for the case of externally mixed aerosols dominated. For example,
526 MHD-winter aerosols (high external mixing; $\chi_{\text{mean}}=0.38\pm 0.12$) showed RCs in N_{CCN} of
527 23% (C1) and 72% (C2), whereas MHD-summer aerosols (high internal mixing;
528 $\chi_{\text{mean}}=0.69\pm 0.19$) exhibited negligible effects (-2.5% in C1, 0.9% in C2). Inland
529 atmospheres, despite smaller seasonal χ variations, showed analogous trends: RCs of

530 N_{CCN} in winter (55% in C1, 57% in C2 for external mixing) exceeded summer values
531 for more internally mixed populations (Fig. 8d). These results confirm that hygroscopic
532 heterogeneity strongly influences N_{CCN} under external mixing, aligning with prior work
533 (Ching et al., 2017).

534 The covariation characteristics are most pronounced during winter in both
535 environments, attributed to heightened winter D_{crit} sensitivity to χ : a 0.1 χ increase
536 reduces D_{crit} by 5.2% (winter), boosting N_{CCN} by 39%, versus 2.4% D_{crit} reduction
537 (summer) yielding only 6% N_{CCN} enhancement. Concomitantly, winter N_{CN} - χ effects on
538 N_{CCN} reach 65%, far exceeding summer responses.

539 In contrast to previous evaluation methods that oversimplify mixing states (Ren et
540 al., 2018; Xu et al., 2021b), the entropy-based framework employed in this study
541 explicitly characterizes the covariation between the CCN activity and transitions in the
542 mixing state. Aerosols in IAP-winter are presumably shaped by intense urban pollution
543 sources—including traffic emissions, residential heating, and cooking activities—
544 thereby enriching the externally mixed particle fraction (Fan et al., 2020; Xie et al.,
545 2020). Analogously, aerosols in MHD-winter exhibit dominant external mixing,
546 consisting of near-hydrophobic and hydrophilic particle mixtures (Xu et al., 2021a). As
547 illustrated in Fig. S2, the winter aerosol population exhibits bimodal or multimodal κ -
548 PDF distributions—this pattern indicates a high degree of external mixing, with the
549 aerosols characterized by chemically diverse compositions. Collectively, these results
550 underscore the pivotal role of mixing state heterogeneity in modulating CCN activity
551 across different environments.

552 4. Conclusions

553 The mixing state of aerosol populations undergoes complex transformations
554 during atmospheric aging, altering the distribution of hygroscopic and non-hygroscopic
555 components and thus influencing CCN activity (Xu et al., 2021a; Ching et al., 2017).
556 This study derived a mixing state index (χ) from field-measured hygroscopicity
557 distributions, systematically investigating the covariation relationship between the
558 mixing state and CCN activity at two inland and coastal environments. Results provide
559 field evidence that aerosol mixing states generally reside between purely internal and
560 external extremes (Chen et al., 2022b). Aerosol mixing state is largely influenced by
561 the primary emissions and secondary formation process. Externally-mixed particles
562 with more hydrophobic-mode originate chiefly from primary emissions in IAP, while
563 that of more sea-salt mode from sea spray in MHD. While it becomes more internally-
564 mixed as the enhanced fraction of more-hygroscopic mode and decreased of
565 hydrophobic mode during the aging process. This highlights a dual regulatory
566 mechanism of mixing state and its potential impact on hygroscopic distribution and
567 CCN activity.

568 As χ increases, CN number concentrations (N_{CN}) first rise—driven by primary
569 emissions and new particle formation—then decline due to condensation and
570 coagulation during aging. Additionally, a logarithmic decreasing relationship between
571 critical diameter (D_{cri}) and χ was identified for both inland and coastal particles,
572 parameterized as $D_{\text{cri}} = -32.15\ln(\chi) + 84.71$ (Pearson $R = -0.74$, $R^2 = 0.54$). This offers
573 a practical approach to estimate D_{cri} from χ , serving as a general framework for

574 integrating mixing state effects on CCN activity in atmospheric models.

575 Entropy-based analyses further support the covariation relationships between the
576 mixing state and N_{CCN} , especially for externally mixed aerosols. Current models often
577 oversimplify aerosol mixing states as purely internal or external (Stevens et al., 2019;
578 Bauer et al., 2013), the latter being particularly sensitive to organic matter (Ren et al.,
579 2018; Bhattu et al., 2015). Such simplifications introduce significant biases in N_{CCN}
580 estimation (Riemer et al., 2019; Ching et al., 2019). The χ - D_{cri} parameterization
581 proposed here offers a novel approach to reduce model complexity in representing
582 aerosol hygroscopicity and CCN activation, enabling more accurate simulations of
583 aerosol CCN capacity. It is expected mitigate the underestimation in CCN compared
584 with the complete external mixing assumption, while effectively alleviates the
585 overestimation that arises from applying the complete internal mixing assumption in
586 regions characterized by high external mixing (Zheng et al., 2021a). This advancement
587 improves our understanding of aerosol-cloud interactions (IPCC, 2021; Rosenfeld et al.,
588 2019), critical for refining climate effect assessments.

589 **Data availability**

590 All data used in the study are available at <https://doi.org/10.3974/geodb.2019.06.11.V1>
591 (Fan et al., 2019) and <http://doi.org/10.17632/3dx6pnx869.1> (Xu et al., 2021a).

592 **Author contributions**

593 RH and JR conceived the conceptual development of the paper. JR, FZ and WX directed
594 and performed the experiments with YW and LC. FZ and YS provided the dataset in

595 the inland site. JO, DC and CO provided the dataset in the coastal site. JR conducted
596 the data analysis and wrote the draft. All authors edited and commented on the various
597 sections of the paper.

598 **Competing interests**

599 The contact author has declared that none of the authors has any competing interests.

600 **Supporting Information**

601 Additional analysis results that were applied in this study. Example of calibration
602 results of HTDMA and CCN used in this study (Figure S1), mean values of the κ -PDF
603 for aerosols of five diameters (Figure S2), sensitivity of the hygroscopic parameter for
604 the group of the hygroscopic species on the mixing state index χ (Figure S3), time series
605 of the average per-particle species diversity $D\alpha$, the bulk population species diversity
606 $D\gamma$, and their affine ratio χ (Figure S4), diurnal trend of particle size, chemical mass
607 fraction and number fraction (NF) of hydrophobic and hygroscopic mode in IAP
608 (Figure S5) and in MHD (Figure S6), mixing state as a function of number fraction of
609 hydrophobic and hygroscopic mode (Figure S7), variation of the peak diameter (D_{peak})
610 with the mixing state index (Figure S8), particle number size distribution and mixing
611 state during new particle formation events (Figure S9), diurnal variation of χ and CN
612 concentration during winter and summer periods for 40 nm and 150 nm aerosols in
613 inland and for 35 nm and 165 nm aerosols in coastal site (Figure S10), relative change
614 of the critical diameter and CN concentration with the mixing state index χ (Figure S11)
615 (PDF).

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