

Response to reviewers

We gratefully appreciate the reviewers for the valuable time and insightful comments. The manuscript has been greatly improved based on these comments and suggestions. All the comments have been addressed point by point in the revised manuscript, and the revisions are marked in red colour. The responses to the specific comments are listed as follows.

Response to the Comments from Reviewer #2:

Reviewer #2: *This study explores terpinolene's oxidation mechanisms initiated by OH (daytime) and NO₃ (nocturnal) radicals, finding H shift limits continuous autoxidation in OH-induced reactions while radical center transfer via bond breaking drives new autoxidation for high-yield nitrogen-containing OOMs in NO₃-induced ones. it also finds nitrogen-containing OOMs have lower C* (more prone to enter particle phase), significant C* differences among OOM isomers emphasize determining molecular structures (especially ring numbers), and comparing the two mechanisms aids OOM identification in atmospheric monitoring and model refinement. It can be published after addressing the following concerns:*

Response: Response: Thank you for the positive and valuable comments. We have revised the manuscript carefully according to the comments and suggestions, and marked them in red in the manuscript.

Comment #1: *Abstract: Line 3. Higher SOA yield than what? Higher than other limonene-based systems? What is the specific magnitude of this higher yield? SOA yield varies significantly depending on multiple factors and is not a fixed value.*

Response to Comment #1: Thank you for pointing out. We totally agree with the reviewer's comment; the expression "Terpinolene is an isomeride of limonene, with an even higher SOA yield" is not accurate, since the SOA yields of terpinolene and limonene are largely dependent on the experimental conditions. According to the experimental results from Friedman and Farmer (2018), the SOA yields from the sequential photooxidation of seven monoterpene isomers (α -pinene, β -pinene, limonene, sabinene, terpinolene, α -terpinene, and γ -terpinene) using an Oxidative Flow Reactor under dry conditions. SOA yields were highest for terpinolene (33% at 5.7 days of aging), followed by sabinene, β -pinene, α -pinene, limonene, γ -terpinene, and α -terpinene. Whereas, based on the results (Figure 3) from Griffin et al. (1999), the SOA yield of limonene is evidently higher than that of terpinolene. Therefore, we have revised this sentence in the Abstract, as detailed below: **Terpinolene is an isomeride of limonene, with a high SOA yield.**

Comment #2: *Introduction: Authors should review existing literature to clarify whether nitrogen-containing species have lower or higher saturation concentrations compared to non-nitrogen-*

containing species. Additionally, is organic nitrate or peroxy organic nitrate considered in this context?

Response to Comment #2: Thank you for the valuable comment. We have supplemented the related literature review in the section of Introduction. For the estimation of the vapor pressure, Isaacman-VanWertz and Aumont (2021) assumed that the impact of a nitrate group on the vapor pressure is equivalent to a hydroxyl group, which was considered reasonable for the reaction system dominated by R-ONO₂ and peroxy nitrates in organic nitrogen gas-phase oxidation products. Besides, according to the chamber experiments and structure-activity relationships, adding -ONO₂ functional groups to C10 and larger molecules could reduce the vapor pressure (Rollins et al. 2013). In our work, both organic nitrate and peroxy organic nitrate were considered.

Comment #3: *Method: The approach for evaluating uncertainties associated with the calculations is missing.*

Response to Comment #3: Thank you for pointing out. To evaluate the reliability of the computational method used in our work, the accuracy of the computational level in determining the energy barriers of the rate-determining steps was compared with the calculated results obtained by the highly time-consuming “gold standard” method CBS-QB3 (Dunn et al. 2004), which shows good consistency, as shown in the table below.

Table S1. Energy Barriers calculated by CBS-QB3 and the method used in this study

Reactions	Energy Barrier (kcal/mol)	Energy Barrier (kcal/mol)
	(CBS-QB3)	(M06-2X/6-311++G(3df,3pd))
1-IM4 → 1-IM41b	21.5	21.3
1-IM4 → 1-IM41a	18.3	19.6
1-IM4 → 1-IM41d	22.1	23.0

In addition, we also compared our results of key reactions with the previous work of structurally similar monoterpenes (Møller et al. 2020), and the deviations are within a reasonable range, which is summarized in the following table.

Table S2. Energy Barriers calculated by CCSD(T)-F12a/VDZ-F12 (abbreviated F12) and ωB97X-D/aug-cc-pVTZ, and the method used in this study (unit: kcal/mol).

Reactions	CCSD(T)- F12a/VDZ-F12	ωB97X-D/aug-cc- pVTZ	This work
1-IM4 → 1-IM41d	26.3	25.5	23.0
1-IM4 → 1-IM41b	21.7	22.2	21.3
1-IM4 → 1-IM41g	18.5	19.4	20.9

Comment #4: *Line 153: Previous studies suggest that NO₂ release was derived from experimental data, whereas HONO release in this study is based on energy barrier calculations. If NO₂ was indeed measured experimentally, should this be more reliable than theoretical calculations? If so, how can the calculation results in this study be validated? Conversely, if NO₂ was not measured*

directly (e.g., its signal could be interfered with by HONO), then HONO should be the correct pathway. Are you implying that the NO₂ measurements reported in previous studies originated from HONO signals?

Response to Comment #4: Thank you for pointing out this problem. As shown in Figure 2 in the manuscript, the subsequent reactions of 1-IM42, formed by the reaction of the peroxy radical 1-IM41 and NO, have two possible pathways: a. the elimination of HONO to produce 1-P1 overcoming a high energy barrier of 43.6 kcal/mol; b. the formation of 1-P1 via sequential H abstraction by OH, the cleavage of C-C bond and the barrierless elimination of NO₂, overcoming low barriers of 10.6 and 6.5 kcal/mol. Therefore, the generation of NO₂ is evidently more feasible, which is consistent with the experimental results. Although the formation of HONO from the above reaction pathway is difficult, the possibility that HONO is produced by the atmospheric oxidation of other organic compounds cannot be ruled out. For the NO₂ measurements, the influence of HONO signals could not be completely excluded.

Comment #5: *HONO can release HONO upon photolysis—do you have any data to demonstrate the yield of HONO from this process?*

Response to Comment #5: Thank you for your valuable comment. As the formation of HONO is not the dominant pathway, it was not considered in the zero-dimensional chemical model, and therefore, the yield of HONO cannot be determined in this work.

Comment #6: *Line 236: NO₂ is not specified; I assume its concentration is set to 0 in your model? In forested areas and remote regions with low NO levels, most NO is converted to NO₂. NO₂ can undergo photolysis to form NO, so total NO_x (NO + NO₂) should be a more representative metric.*

Response to Comment #6: Thank you for your valuable comment. We agree with your view, in the atmosphere, the conversion between NO and NO₂ can readily occur, with the concentrations reaching a dynamic equilibrium. In our work, the initial concentrations of NO₂ were set in the model, and the re-simulations were conducted under different atmospheric conditions, i.e., (a) in a urban area (50 ppb NO, 39 ppb NO₂, 1 ppt HO₂), (b) in a suburban area (3 ppb NO, 9 ppb NO₂, 3 ppt HO₂), and (c) in a forested area (0.5 ppb NO, 4 ppb NO₂, 40 ppt HO₂) (Guo et al. 2024, Kieloaho et al. 2013, Klemm et al. 2006, Lew et al. 2020, Ma et al. 2019, Mao et al. 2010, Mavroidis and Ilia 2012, Mazzeo et al. 2005, Zhang et al. 2022).

Comment #7: *I am also curious why NO concentrations in forests are set to be lower than in "remote areas". Soil NO_x emissions are a substantial source of NO_x in forest environments.*

Response to Comment #7: Thank you for your valuable comment. We agree with your point of view; soil emission is an important source of NO_x in the forest areas. According to the literature, NO concentration over the forests is not fixed, varying from tens of ppt to hundreds of ppt or even several ppb in different regions, with higher concentration in the low atmosphere due to the soil emission (Kieloaho et al. 2013, Klemm et al. 2006). In our work, we referred to the literature and determined the atmospheric conditions in a suburban area (3 ppb NO, 9 ppb NO₂, 3 ppt HO₂) and a forested area (0.5 ppb NO, 4 ppb NO₂, 40 ppt HO₂) in our simulation (Guo et al. 2024, Kieloaho et al. 2013, Klemm et al. 2006, Lew et al. 2020, Mavroidis and Ilia 2012).

Comment #8: *Line 237: Why is HO₂ concentration fixed at 50 ppt across all scenarios? HO₂ is an intermediate radical, and its formation and levels depend on factors such as VOCs and NO_x. It is not typically constrained to a fixed value in box models.*

Response to Comment #8: Thank you for pointing out. We have modified the setting of HO₂ concentrations in the model based on the previous research. In our simulation, the initial concentrations of HO₂ were set to be 1 ppt in an urban area, 3 ppt in a suburban area, and 40 ppt in a forested area. In addition, the HO₂ concentration is not constrained to a fixed value in our box models.

Comment #9: *Line 349: Can you provide a more detailed description of the results presented in Figure 8? For example, are you stating that C₁₀H₁₆O₄ (classified as an OOM) is identified as an SVOC with a C* of 1 in your analysis, but would be categorized as an SVOC with a C* of 2.5 if calculated using the functional group method?*

Response to Comment #9: Thank you for your valuable comment. For instance, for the product C₁₀H₁₆O₄ (two isomers: 2-P12 & 1-P5), the values of lg C* for two isomers are both ~1 µg/m³ estimated by the molecular formula parameterization method, belonging to SVOC. Whereas, using the functional group contribution method (SIMPOL.1), the lg C* values of the isomers 1-P5 and 2-P12 are calculated to be 2.4 and 2.7 µg/m³, respectively, of which 2-P12 can be categorized as an IVOC.

Comment #10: *Line 350: Which method for volatility estimation—molecular formula-based or functional group-based—is more reliable? Can you discuss the reasons for the discrepancies between these two methods? Additionally, does organic nitrate influence volatility calculations?*

Response to Comment #10: Thank you for your valuable comment. The functional group-based method could be more reliable, since within each compound category the variations in vapor pressure can be attributed to the number and size of functional groups present and the relative positions of those functional groups to each other both positionally and geometrically, these two factors impact upon both the molecules' dipole moments and upon their ability to interact both intramolecularly and intermolecularly via hydrogen bonding (Dang et al. 2019), thus explaining the differences in the estimated vapor pressure.

Yes, organic nitrate can influence the volatility calculations; the reactions of volatile organic precursors with NO_x lead to the formation of nitro and nitrate groups. There exists some data in the literature for polynitrates, but few data are available for compounds with combinations of nitro and/or nitrate with -COOH and -OH groups. Hence, the previous work assumed that nitrogen is predominantly present as nitrate groups, and each nitrate group is treated as being equivalent to a hydroxyl group; this assumption is reasonable for a system dominated by products of gas-phase oxidation, in which R-ONO₂ compounds and peroxy nitrates are the dominant source of organic nitrogen (Bilde et al. 2015, Isaacman-VanWertz and Aumont 2021).

Comment #11: *Results section: Organic nitrate is a key component of nitrogen-containing OOMs. Given this, I recommend focusing discussions on organic nitrate (RONO₂) OOMs specifically, rather than attributing all relevant observations to "nitrogen-containing OOMs" in general.*

Response to Comment #11: Thank you for pointing out. We have modified the description in the revised manuscript for specificity and clarity.

Reference:

- Bilde, M., Barsanti, K., Booth, M., Cappa, C.D., Donahue, N.M., Emanuelsson, E.U., McFiggans, G., Krieger, U.K., Marcolli, C., Topping, D., Ziemann, P., Barley, M., Clegg, S., Dennis-Smith, B., Hallquist, M., Hallquist, Å.M., Khlystov, A., Kulmala, M., Mogensen, D., Percival, C.J., Pope, F., Reid, J.P., Ribeiro da Silva, M.A.V., Rosenoern, T., Salo, K., Soonsin, V.P., Yli-Juuti, T., Prisle, N.L., Pagels, J., Rarey, J., Zardini, A.A. and Riipinen, I. (2015) Saturation Vapor Pressures and Transition Enthalpies of Low-Volatility Organic Molecules of Atmospheric Relevance: From Dicarboxylic Acids to Complex Mixtures. *Chemical Reviews* 115(10), 4115-4156.
- Dang, C., Bannan, T., Shelley, P., Priestley, M., Worrall, S.D., Waters, J., Coe, H., Percival, C.J. and Topping, D. (2019) The effect of structure and isomerism on the vapor pressures of organic molecules and its potential atmospheric relevance. *Aerosol Science and Technology* 53(9), 1040-1055.
- Dunn, M.E., Pokon, E.K. and Shields, G.C. (2004) Thermodynamics of Forming Water Clusters at Various Temperatures and Pressures by Gaussian-2, Gaussian-3, Complete Basis Set-QB3, and Complete Basis Set-APNO Model Chemistries; Implications for Atmospheric Chemistry. *Journal of the American Chemical Society* 126(8), 2647-2653.
- Friedman, B. and Farmer, D.K. (2018) SOA and gas phase organic acid yields from the sequential photooxidation of seven monoterpenes. *Atmospheric Environment* 187, 335-345.
- Griffin, R.J., Cocker III, D.R., Flagan, R.C. and Seinfeld, J.H. (1999) Organic aerosol formation from the oxidation of biogenic hydrocarbons. *Journal of Geophysical Research: Atmospheres* 104(D3), 3555-3567.
- Guo, P., Su, Y., Sun, X., Liu, C., Cui, B., Xu, X., Ouyang, Z. and Wang, X. (2024) Urban–Rural Comparisons of Biogenic Volatile Organic Compounds and Ground-Level Ozone in Beijing. *Atmospheric Environment* 15(3), 508.
- Isaacman-VanWertz, G. and Aumont, B. (2021) Impact of organic molecular structure on the estimation of atmospherically relevant physicochemical parameters. *Atmos. Chem. Phys.* 21(8), 6541-6563.
- Kieloaho, A.-J., Hellén, H., Hakola, H., Manninen, H.E., Nieminen, T., Kulmala, M. and Pihlatie, M. (2013) Gas-phase alkylamines in a boreal Scots pine forest air. *Atmospheric Environment* 80, 369-377.
- Klemm, O., Held, A., Forkel, R., Gasche, R., Kanter, H.J., Rappenglück, B., Steinbrecher, R., Müller, K., Plewka, A., Cojocariu, C., Kreuzwieser, J., Valverde-Canossa, J., Schuster, G., Moortgat, G.K., Graus, M. and Hansel, A. (2006) Experiments on forest/atmosphere exchange: Climatology and fluxes during two summer campaigns in NE Bavaria. *Atmospheric Environment* 40, 3-20.
- Lew, M.M., Rickly, P.S., Bottorff, B.P., Reidy, E., Sklaveniti, S., Léonardis, T., Locoge, N., Dusanter, S., Kundu, S., Wood, E. and Stevens, P.S. (2020) OH and HO₂ radical chemistry in a midlatitude forest: measurements and model comparisons. *Atmos. Chem. Phys.* 20(15), 9209-9230.
- Ma, X., Tan, Z., Lu, K., Yang, X., Liu, Y., Li, S., Li, X., Chen, S., Novelli, A., Cho, C., Zeng, L., Wahner, A. and Zhang, Y. (2019) Winter photochemistry in Beijing: Observation and model simulation of OH and HO₂ radicals at an urban site. *Science of The Total Environment* 685, 85-95.
- Mao, J., Ren, X., Chen, S., Brune, W.H., Chen, Z., Martinez, M., Harder, H., Lefer, B., Rappenglück, B., Flynn, J. and Leuchner, M. (2010) Atmospheric oxidation capacity in the summer of Houston 2006: Comparison with summer measurements in other metropolitan studies. *Atmospheric Environment* 44(33), 4107-4115.
- Mavroidis, I. and Ilia, M. (2012) Trends of NO_x, NO₂ and O₃ concentrations at three different types of air quality monitoring stations in Athens, Greece. *Atmospheric Environment* 63, 135-147.
- Mazzeo, N.A., Venegas, L.E. and Choren, H. (2005) Analysis of NO, NO₂, O₃ and NO_x concentrations measured at a green area of Buenos Aires City during wintertime. *Atmospheric Environment* 39(17),

3055-3068.

Møller, K.H., Otkjær, R.V., Chen, J. and Kjaergaard, H.G. (2020) Double Bonds Are Key to Fast Unimolecular Reactivity in First-Generation Monoterpene Hydroxy Peroxy Radicals. *The Journal of Physical Chemistry A* 124(14), 2885-2896.

Zhang, G., Hu, R., Xie, P., Lou, S., Wang, F., Wang, Y., Qin, M., Li, X., Liu, X., Wang, Y. and Liu, W. (2022) Observation and simulation of HOx radicals in an urban area in Shanghai, China. *Science of The Total Environment* 810, 152275.