



# 1 Trends and spatial variation of oceanic dimethyl sulfide 2 under a warming climate revealed by an artificial neural 3 network model

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8 **Abstract.** Marine dimethyl sulfide (DMS), a climatically active gas generated through  
9 microbial degradation of dimethyl sulfoniopropionate (DMSP), plays a key role in the  
10 Earth's climate system by modifying its radiation budget. However, the sea-to-air flux  
11 and future variations under climate change are still uncertain. Simulations from Earth  
12 System Models (ESMs) provide divergent trends. Here, we developed an artificial  
13 neural network (ANN) model trained using DMS observations and eight observational  
14 environmental parameters, along with model parameters extracted from the  
15 simulations of CESM2-WACCM to predict variations of DMS concentrations and  
16 sea-to-air flux for both historical (1850–2014) and SSP5-8.5 scenario (2015–2100).  
17 Our simulation indicates that DMS concentrations will generally decline by the end of  
18 this century. Specifically, from 2015 to 2050, the DMS concentrations are projected to  
19 decrease at a rate of  $0.40 \pm 0.13\%$  per decade. From 2050 to 2100, the rate of decrease  
20 is expected to accelerate to  $0.89 \pm 0.08\%$  per decade. The sea-to-air flux of DMS  
21 exhibits a non-monotonic trend. It is projected to increase at a rate of  $0.51 \pm 0.16\%$  per  
22 decade from 2015 to 2050. However, from 2050 to 2100, the flux is expected to  
23 decrease at a rate of  $0.37 \pm 0.11\%$  per decade. We further explore the attribution of  
24 DMS changes by running a series of sensitivity tests. We find that elevated sea surface  
25 temperature (SST) and photosynthetically active radiation (PAR), along with nutrient  
26 depletion, are projected to lead to the decline in DMS concentrations by the end of  
27 this century. Furthermore, our geospatial analysis indicates that mixed layer depth  
28 (MLD) emerges as the predominant driver in the Southern Ocean, and nutrient-  
29 dependent effects strongly correlate with DMS in the open seas (trades and  
30 westerlies). Our findings suggest that site-specific modeling schemes are needed to  
31 accurately model DMS dynamics.

## 32 1 Introduction

33 Oceanic dimethyl sulfide (DMS) is primarily synthesized in seawater through the  
34 enzymatic cleavage of the biogenic compound dimethyl sulfoniopropionate (DMSP)  
35 and released through microalgae exudation and mortality (Galí et al., 2015; Simó &  
36 Dachs, 2002; Stefels, 2000). DMS in the surface ocean is supersaturated compared to  
37 its atmospheric counterpart, and the sea-to-air flux is responsible for more than half of  
38 the total flux of gaseous sulfur to the atmosphere (Lana et al., 2011; Quinn & Bates,  
39 2011). Once in the atmosphere, DMS is oxidized to sulfuric and methane-sulfonic  
40 acids, which contribute to the formation of cloud condensation nuclei (CCN) and  
41 facilitate cloud formation, and thereby has the ability to reduce solar radiation and  
42 affects the Earth's energy budget (Charlson et al., 1987). The CLAW hypothesis  
43 postulates a climate-negative feedback loop among phytoplankton, DMS emissions,  
44 CCN, and Earth's energy budget. In the proposed feedback loop, CCN and cloud  
45 albedo are regulated up or down by oceanic phytoplankton through the medium of  
46 DMS emissions (Charlson et al., 1987).

47 However, the conventional view of DMS induced negative climate feedback is  
48 increasingly challenged by emerging evidence. Both model simulations and  
49 mesocosm studies suggest that DMS may instead exert a positive climate feedback



effect under the global warming scenario (Six et al., 2013; Wang et al., 2018; Webb et al., 2016; Zhao et al., 2024). CO<sub>2</sub> forcing simulations have been conducted to predict DMS distribution and its global-scale emissions, with varying forcing conditions revealing substantial spatial heterogeneity (Bopp et al., 2003; Gabric et al., 2013). As a result of the combined effects of ocean acidification and climate change, the projection of global DMS emissions decreases by about 18(±3)% in 2100 compared to the pre-industrial time in an Earth system model (ESM) climate simulations (Six et al., 2013). Similarly, global mean DMS concentrations are predicted to decrease by 15.1% by the end of this century compared to the historical results (1960-2014), which is primarily driven by rising CO<sub>2</sub> levels (Zhao et al., 2024). In addition, a declining trend in DMS concentrations and sea-to-air flux under the RCP8.5 scenario are also detected using a fully coupled Earth system model (CESM) (Wang et al., 2018).

An ensemble of four ESMs (CNRM-ESM2-1, MIROC-ES2L, NorESM2-LM, and UKESM1-0-LL) from the CMIP6 historical and SSP5-8.5 experiments provided divergent trends in DMS concentrations and sea-to-air flux starting from the year of 2015, which means that there is significant uncertainty regarding the changes of projected DMS trends based on Earth System Models in the future (Bock et al., 2021). A recent study by Joge et al. (2025) found that the global mean surface DMS concentration exhibits a decreasing trend with warming. However, in contrast to the decreasing trend of DMS concentration, the combined effects of increasing wind speed, increasing sea surface temperature, and decreasing ice-coverage lead to an increasing trend of sea-to-air DMS flux. Consensus has not been reached regarding future trends in DMS concentrations and sea-to-air flux. Key factors controlling DMS variations have yet to be determined. Consequently, the response of DMS concentrations/flux under a warming climate remains uncertain.

Here, we train an artificial neural network (ANN) model using DMS measurements (Fig. S1) and eight observational environmental variables, including Chl *a*, MLD, dissolved inorganic nitrate (DIN), PAR, dissolved inorganic phosphorus (DIP), silicate (SiO<sub>4</sub>), sea surface salinity (SSS), and sea surface temperature (SST). Subsequently, we feed the model with the monthly outputs from CESM2-WACCM historical and SSP5-8.5 experiments to predict global DMS concentrations and temporal variations (Fig. S2). Furthermore, we investigate the mechanisms driving the variations of DMS concentrations under global warming scenarios. We also conduct a series of sensitivity tests to explore the attribution of DMS changes in a warming climate state, and identify the key factors (SST, PAR and nutrients) in different regions. In the end, we make prospects and suggestions for future study.

## 2 Data and Methods

### 2.1 Observations

Observational DMS concentration data were obtained from two primary sources: 1) the Global Surface Seawater DMS Database (Pacific Marine Environmental Laboratory, PMEL; last access: 1 May 2020) and 2) the North Atlantic Aerosols and Marine Ecosystems Study (NAAMES; Behrenfeld et al., 2019) (Table S1). After quality control, which excluded measurements with DMS concentrations below 0.1 nM or exceeding 100 nM following Galí et al. (2015), a total of 93,571 valid data points were retained (PMEL: 86,785; NAAMES: 6,786). The Global Surface Seawater DMS Database also provides additional in situ measurements, including Chl *a* (PMEL: 11,491; NAAMES: 6,750), SST (PMEL: 81,069; NAAMES: 6,786), and SSS (PMEL: 77,209; NAAMES: 6,786). We used in situ measurements when they are available. Otherwise, we supplemented the missing values with monthly climatology data from auxiliary datasets (Table S1). For example, SeaWiFS monthly averaged Level 3-binned Chl *a* data (9.2 km resolution; last access: 1 May 2020) from December 1997 to March 2010 were spatially and temporally matched to DMS measurements. Similarly, SeaWiFS monthly averaged Level 3-binned PAR data (9.2 km resolution; last access: 1 May 2020) from September 1997 to August 2010 were



also matched with DMS observations. Climatological MLD data were downloaded from the Monthly Isopycnal/Mixed-Layer Ocean Climatology (MIMOC; Schmidt et al., 2013). Nutrients (nitrate, phosphate, silicate) data were obtained from the World Ocean Atlas 2013 (WOA2013). All ancillary data were aligned with DMS measurements based on sampling location and time of year. Rigorous quality control was applied following Galí et al. (2015). For instance, coastal data (salinity < 30), as well as measurements with anomalously low nutrient concentrations (phosphate < 0.01 μM; nitrate < 0.01 μM; silicate < 0.1 μM) and low Chl *a* (Chl *a* < 0.01 mg m<sup>-3</sup>) were excluded to focus on open-ocean conditions.

## 2.2 Earth System Models

For the input data to our ANN model, we first used the surface and monthly environmental outputs from CESM2-WACCM. This model ensemble was selected because it demonstrates the best overall results among the CMIP6 ensembles when compared to observational data (see Fig S6). We then compare our predicted DMS concentrations and fluxes with the outputs from four ESMs in CMIP6 (CNRM-ESM2-1, MIROC-ES2L, NorESM2-LM, UKESM1-0-LL). More detailed descriptions of these four ESMs are provided below.

The oceanic components and their respective resolutions for the four ESMs are detailed in Table S2, which includes ensemble numbers for both the historical (1850-2014) and SSP5-8.5 (2015-2100) experiments. All datasets were downloaded from the CMIP6 Earth System Grid Federation (ESGF) nodes. These ESMs simulate the main large-scale features of the ocean circulation. Recent studies have also shown that these models have improved simulations of MLD, a key driver for marine biogeochemistry and marine DMS emissions (Seferian et al., 2020).

For the CNRM-ESM2-1 model, DMS concentrations are computed using the biogeochemical model PISCES, coupled with the global ocean general circulation model (OGCM) NEMO. The version of PISCES used, PISCESv2-gas, includes a module for simulating the cycle of gases relevant to climate. DMS flux is calculated using the parameterization of gas exchange coefficients of Wanninkhof (2014) (Michou et al., 2020).

In the MIROC-ES2L model, DMS concentrations are calculated based on the Aranami and Tsunogai (2004) parameterization, which links sea surface DMS concentrations to MLD and Chl-*a* concentration as follows:

$$\text{DMS} = \begin{cases} \frac{60.0}{\text{MLD}} & \text{if } \frac{\text{Chl}}{\text{MLD}} < 0.02 \\ 55.8 \cdot \left( \frac{\text{Chl}}{\text{MLD}} \right) + 0.6 & \text{if } \frac{\text{Chl}}{\text{MLD}} > 0.02 \end{cases},$$

in which MLD and Chl *a* are simulated by OECO-v2, coupled in MIROC-ES2L (Hajima et al., 2020). DMS flux is calculated using Aranami and Tsunogai (2004) parameterization.

For the NorESM2-LM model, the biogeochemical model iHAMOCC is coupled in the global OGCM BLOM to compute DMS concentrations, which is a function of temperature and export production (Tjiputra et al., 2020).

For the UKESM1-0-LL model, DMS concentrations are computed within the ocean biogeochemistry model MEDUSA (Yool et al., 2013) and interactively coupled with the global OGCM NEMO. DMS concentrations are linearly correlates with a composite variable that includes the logarithm of Chl *a*, light, and nutrients. DMS flux is calculated according to the air-to-sea gas transfer scheme of Liss and Slater (1974). DMS concentrations in the atmosphere are subsequently modified through a number of gas-phase aerosol precursor reactions within the stratospheric and tropospheric chemistry schemes of the UKESM1-0-LL model (Mulcahy et al., 2020).



### 2.3 Artificial neural network model

The ANN model is a branch of artificial intelligence (AI), which builds with a fully connected network of nodes and neurons. Each neuron has an activation function and is connected to other neurons by iteratively determined weights (Gardner & Dorling, 1998). This algorithm has a great advantage because they make no prior assumptions on the data distribution and can fit data in gap area using non-linear equation (Breiman, 2001; Gardner & Dorling, 1998).

The ANN model is trained using the *Keras* deep-learning toolbox in Python 3.8, with eight environmental variables (Chl *a*, MLD, DIN, DIP, PAR, SiO<sub>4</sub>, SST, SSS) as predictants and DMS as predictor. All data are log transformed and normalized to the range of [-1,1].

The dataset is then divided into three sections: training, internal testing, and external validating datasets. Specifically, data falling into the following 14 latitude bands (64–65°N, 54–55°N, 44–45°N, 34–35°N, 24–25°N, 14–15°N, 4–5°N, 4–5°S, 14–15°S, 24–25°S, 34–35°S, 44–45°S, 54–55°S, 64–65°S) are left out for internal testing (9084 points). Similarly, data falling to the fifteen latitude bands (69–70°N, 59–60°N, 49–50°N, 39–40°N, 29–30°N, 19–20°N, 9–10°N, 1–0°S, 9–10°S, 19–20°S, 29–30°S, 39–40°S, 49–50°S, 59–60°S, 69–70°S) are left out for external validation (10870 points). The remaining data are used as training dataset (63042 data points). Separating the data by latitude bands rather than using random separation helps prevent information leakage, as in situ measurements are internally correlated. The traditional random separation methods tend to overfitting (Wang et al., 2020).

In the training process, we adjust the hyper-parameters, such as dropout ratio, number of hidden layers, and number of nodes on each layer to prevent overfitting while achieving the best goodness of fit to observations. Eventually, the final ANN model adopted consists of one input layer, two dense hidden layers, and one output layer. The input layer comprises nodes corresponding to the predictors. Each hidden layer contains 128 nodes, and the output layer has a single node for DMS concentration simulations. To mitigate overfitting, two dropout layers with a dropout ratio of 0.25 are incorporated into each hidden layer. Additionally, an L2 kernel regularizer with a value of 0.001 is applied to each hidden layer. During network training, the mean squared error of the internal validation data is monitored. After obtaining a satisfactory combination of those hyper-parameters, we fix them and finetune the network using all available data.

### 2.4 Sea-to-air flux of DMS

DMS flux is calculated using an empirical formula, which takes into account sea surface wind (SSW), sea ice coverage, and the viscosity coefficient related to gas transfer velocities in atmosphere and surface ocean.

Air–sea gas transfer is estimated using the following bulk formula:

$$F = K_w(C_w - C_a/H) ,$$

where  $F$  is sea-to-air gas exchange flux,  $C_w$  and  $C_a$  are bulk water and gas concentrations, and  $K_w$  ( $cm\ h^{-1}$ ) is the overall gas transfer velocity, expressed in waterside units (Liss & Merlivat, 1986).  $K_w$  reflects the combined resistance to gas transfer on both sides of the interface, as follows:

$$\frac{1}{K_w} = \frac{1}{k_w} + 1/Hk_a ,$$

where the dimensionless  $H$  is the Henry law constant (gas or liquid), and  $k_a$  and  $k_w$  are gas transfer velocities in air and seawater, respectively. DMS in the surface ocean is strongly supersaturated with respect to that in the overlying atmosphere ( $C_w \gg C_a$ ), so the DMS flux bulk formula is simplified as:

$$F = K_w C_w$$



Our study uses the parameterization for  $K_w$  that refer to Goddijn-Murphy et al. (2012) (hereafter GM12), which is based on regressions between satellite-based wind speed observations and shipboard in situ measurements of DMS gas transfer velocities using eddy covariance method. SSW and sea ice area coverage data are from the CESM-WACCM monthly simulation datasets.

### 3. Results and discussion

#### 3.1 Long-term trends of DMS under global warming scenario

The ANN model captures the major variance in the observed data with the goodness-of-fit  $R^2$  value of 0.68 for the training datasets and 0.66 for the testing datasets (Fig. S3). We then conduct temporal simulations by feeding the ANN model with parameters extracted from CESM2-WACCM, which best reproduces the corresponding observational parameters among CMIP6 ensembles (Fig. S4). Compared to the historical pattern, the model reveals distinct trends in DMS concentration across the global ocean, with notable patterns emerging in several key areas (Fig. 1a). For instance, DMS concentrations exhibit an increasing trend in the Southern Ocean, the eastern equatorial Pacific, the subpolar North Atlantic, and the Arctic Ocean, with the highest concentration increase occurring in the Southern Ocean between 40°S and 60°S. This is particularly important because the Southern Ocean is far from anthropogenic aerosol sources, and the sea-to-air flux of DMS is the major source of atmospheric sulfur. Therefore, it strongly influences the radiative budget in the Southern Hemisphere (Hamilton et al., 2014).

Nevertheless, a decreasing trend in DMS concentrations is evident in the low-to-middle latitude regions of the Pacific, Indian, and Atlantic Oceans. Using a similar network model, Joge et al. (2025) found that DMS concentrations increase in the subtropical gyres, whereas we observe a decreasing trend in the same regions. The discrepancy is primarily attributed to two factors: 1) Data sources: We trained our model using predominantly observational parameters, while Joge et al. (2025) used model outputs. Given the evident biases between model simulations and observational data, we believe that the observational parameters are more effective in capturing the true relationship with DMS concentrations than model-derived data. 2) Model Ensembles: We employed output solely from CESM2-WACCM as input to our network model, whereas Joge et al. (2025) used an ensemble of eight models. As illustrated in Fig. S4, CESM2-WACCM demonstrated the best reproduction of observational data, while other models exhibit significant biases. These differences in data sources and model ensembles likely account for the divergent results observed in DMS concentration trends. The sea-to-air flux of DMS generally follows a similar trend to DMS concentrations. The correlation sign is consistent in most of the open oceans, except for regional discrepancies in coastal biomes. These discrepancies are probably caused by the inverse change of wind speed with DMS concentrations.

To elaborate more on the trend of DMS with global warming, we calculate the global area-weighted annual mean DMS concentrations and DMS flux. The temporal trend from our ANN model is plotted alongside another neural network model and four ESM ensembles, all of which explicitly model DMS under historical and SSP5-8.5 scenarios, spanning from 1850 to 2100 (Fig. 2 and Table 1, 2). DMS concentrations and fluxes in all models show a similar flat trend with different magnitudes over the historical period. For future projections, our global mean surface DMS concentration shows a decreasing trend, consistent with Joge et al. (2025). However, the global mean sea-to-air flux of DMS exhibits a non-monotonic trend. From 2015 to 2050, the sea-to-air flux of DMS shows a similar increasing trend to that reported by Joge et al. (2025), but with a higher increasing rate (Table 2). This increasing trend is likely due to the combined effects of decreasing ice coverage, increasing wind speed, and increasing sea surface temperature, which compensate for the decreasing DMS concentration (Fig. 1). However, from 2050 to 2100, the increasing trend reverses to a decreasing trend, with a rate of  $0.37 \pm 0.11\%$  per decade. Among the CMIP6 models, two EMSs (NorESM2-LM, and UKESM1-0-LL) predict decreasing trends in the





future, while the other two models (CNRM-ESM2-1 and MIROC-ES2L) predict the opposite. This divergence has previously been suggested to be explained by the bias in modelled SST (Bock et al., 2021).

### 3.2 Attribution of DMS changes under global warming scenario

To identify the parameter(s) driving the temporal variations of DMS, we conduct eight sensitivity experiments. In each, we hold seven of the eight environmental parameters needed by the ANN at their initial historical values, and allow the remaining one to vary according to historical and SSP5-8.5 simulation. These eight experiments are denoted as VChl, VMLD, VDIN, VPAR, VDIP, VSiO<sub>4</sub>, VSSS, and VSST, representing the varying parameter of Chl *a*, MLD, DIN, PAR, DIP, SiO<sub>4</sub>, SSS, and SST, respectively. For the eight parameters, PAR and SST display an increase trend, and the other six variables show a decrease trend under the SSP5-8.5.

DMS concentration shows a significant increase in the VSST test and a modest increase in the VPAR test compared to the control run, which is consistent with the trends of SST and PAR. The elevated SST, especially in high latitude oceans, promotes phytoplankton production, which is the primary producer of DMS (del Valle et al., 2007; Derevianko et al., 2009; Galí et al., 2013; Watanabe et al., 2007). VPAR test reveals a modest positive correlation between the changing trends of DMS concentrations and PAR (Fig. 3a, b). The higher irradiance inhibits bacterial consumption of DMS, influencing the proportion of high DMSP producers within assemblages (Galí et al., 2011; McNabb & Tortell, 2022; Vance et al., 2013). Conversely, the distribution of PAR shows an overall negative spatial correlation with DMS (Fig. 1a and Fig. S5), which may indicate a role for photolytic degradation in DMS loss (del Valle et al., 2007). These findings suggest that light-induced oxidative stress and inhibited microbial DMS consumption may influence regional DMS distributions. This particularly true in areas where photolysis significantly drives DMS oxidation. The relative contributions of biotic and abiotic processes require further in situ validation.

DMS exhibits a decreasing trend in the VMLD test, likely because the fact that the shoaling of MLD due to global warming inhibits the upwelling of bottom nutrients (Fig. 3a), hinders the vertical mixing of higher nutrients from deeper layers and oxygen-rich waters in the upper ocean, suppressing phytoplankton primary production in low-to-middle latitude oceans and ultimately resulting in a decline of DMS concentrations in the surface ocean. In the VChl test, DMS concentration remains nearly unchanged, which differs from the decreasing trend of global mean Chl *a* concentration. This is in contrast to previous studies that extensively link DMS to Chl *a*, and indicates that the biogeochemical cycle of DMS is far more complex than Chl *a* can represent (Galí & Simó, 2015; Nemcek et al., 2008; Simó & Dachs, 2002). Indeed, applying an algorithm based on Chl *a* yielded little insight into DMS dynamics (Hirata et al., 2011). This is probably because the taxonomic composition of phytoplankton assemblages that differ in their ability of DMS production likely influence the variability of DMS cycling. As such, the bulk Chl *a*, representing a composite signal from all phytoplankton taxa, may have limited utility in predicting the spatial patterns of DMS on a global scale, but may be useful regionally.

For the nutrient tests, DMS concentration show an increasing trend in both VDIN and VDIP, while both DIN and DIP decrease under SSP5-8.5. Overall, DMS–nutrients relationship may be partially attributed to the sulfur overflow hypothesis (Stefels, 2000), which suggests that nutrient-limited phytoplankton increase DMSP production, and its subsequent cleavage to DMS as a mechanism to regulate intracellular sulfur quotas when protein synthesis is limited (Hatton & Wilson, 2007; Kinsey et al., 2016; Simó & Vila-Costa, 2006; Spiese & Tatarkov, 2014; Stefels, 2000). This hypothesis also explains elevated DMS concentrations in the Southern Ocean, subpolar North Atlantic and the Bering Sea, where nutrients concentrations are showing decreasing trends (Fig. S5). Moreover, nutrient-dependent effects significantly explain seasonal



variability, particularly as phytoplankton growth becomes nutrient-limited during summer time.

The ocean represents an intricate system where environmental changes directly modulate DMS concentrations. Under the warming scenario (SSP5-8.5), elevated SST and PAR will strengthen ocean stratification, shoaling the MLD and reducing nutrient supplies from the deep ocean. Our findings suggest that these effects jointly determines the temporal variation of DMS concentrations.

### 3.3 Key factors regulating DMS variation in typical regions

To further investigate these regional variabilities, we divide the ocean into six regions according to Longhurst (1998): polar North, polar South, westerlies North, westerlies South, trades, and coastal (see Fig. 4a). We then examine the key factors influencing DMS concentrations across these regions (Fig. 4b). No key driving factors of DMS variation are identified in the polar North and coastal regions, likely because these areas encompass diverse biomes with site-specific drivers. When analyzed as a whole, no single dominant factor emerges. In contrast, DMS concentrations in the westerlies North region show strong negative correlations with DIN ( $r = -0.46$ ), DIP ( $r = -0.58$ ), and  $\text{SiO}_4$  ( $r = -0.31$ ), and strong positive correlations with PAR ( $r = 0.60$ ), SSS ( $r = 0.52$ ), and SST ( $r = 0.53$ ). These results suggest that low nutrient levels, strong light, and warm surface waters may favor small phytoplankton or *Phaeocystis*, which are more prolific DMS producers compared to diatoms.

Conversely, in the westerlies South region, Chl *a* and MLD emerge as the dominant factors ( $r = 0.44$  and  $-0.48$ , respectively) influencing DMS variation. This indicates that Chl *a* is a strong predictor of DMS concentrations in this region, likely because prolific DMS-producing phytoplankton contribute significantly to Chl-*a* levels. A similarly strong negative correlation is observed between MLD and predicted DMS concentrations in the polar South ( $r = -0.48$ ). Both polar South and westerlies South regions are characterized by high background nutrient concentrations and deep mixed layers. The deepening of the MLD may dilute phytoplankton biomass and DMS, leading to the observed relationships. In the trades region (open ocean), where nutrient levels are generally low and small phytoplankton dominate, DMS concentrations are positively correlated with DIN, DIP, PAR, and Chl *a*. This suggests that in the trades, nutrient-driven higher primary production leads to higher DMS production.

Observational and modeling studies have extensively documented the distribution of DMS concentrations and sea-to-air fluxes across the global ocean (Galí et al., 2015; Jøge et al., 2025; Lana et al., 2011; Seferian et al., 2020; Simó et al., 2002). These studies indicate that DMS emissions are not solely governed by global biogeochemical cycles but also arise from complex ecological interactions, planktonic food-web dynamics, cellular physiological processes, and marine chemical transformations (Simó et al., 2002). Notably, elevated sea-to-air DMS fluxes are predominantly observed in upwelling zones, particularly in the tropical and equatorial Pacific Ocean. Furthermore, recent research highlights the influence of SST, deep-water formation, biological productivity, and thermohaline circulation on DMS flux variability (Seferian et al., 2020). Our results demonstrate that DMS concentrations exhibit regional-scale dependence on multiple environmental drivers (Fig. 5).

### Conclusions

The comparison of DMS concentrations and flux variations over the simulation periods from 1850 to 2100 (historical and SSP5-8.5 for CMIP6 ESMs) yields two key insights. Firstly, all models exhibit relative stability during the historical period. In contrast, in future simulations, two models (CNRM-ESM2-1 and MIROC-ES2L) show an increase in surface ocean DMS concentrations and flux, while the other four models (Jøge25, NorESM2-LM, UKESM1-0-LL, and ANN) show a decreasing trend in DMS concentration. Although ESMs or non-linear equations may not fully



365 elucidate the relationship between DMS and marine phytoplankton, clarifying its  
366 response to climate change is crucial.

367 Secondly, our findings suggest that DMS concentrations exhibit regional-scale  
368 dependence on multiple environmental drivers. In the trades region (open ocean),  
369 higher DMS production is primarily driven by nutrient-mediated increases in primary  
370 productivity. Conversely, in the westerlies North region, DMS concentrations display  
371 strong negative correlations with DIN, DIP, and SiO<sub>4</sub>, while showing strong positive  
372 correlations with PAR, SSS, and SST. In the westerlies South region, Chl a emerges as  
373 a key positive predictor of DMS, and MLD is negatively correlated with DMS in the  
374 polar South.

375 Our results demonstrate that variations in DMS concentration are rarely unidirectional  
376 in response to isolated changes in a single environmental parameter (Fig. 5). This  
377 highlights the complex interactions among these environmental factors, which cannot  
378 be adequately captured by a linear regression model. Future work should focus on the  
379 combined effects, using observational data to constrain models, and integrating these  
380 with ESMs to more accurately simulate DMS concentrations under different  
381 scenarios. It is also crucial to consider the potential climatic implications of changes  
382 in DMS production driven by biogeochemical factors when projecting future climate  
383 change.

#### 384 **Author Contributions**

385 W.-L. W. conceived the project. L.Y. and W.-L. W. carried out the formal analyses. Y.  
386 L. and W.-L. W. wrote and reviewed the manuscript. Both authors have given  
387 approval to the final version of the manuscript.

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#### 394 **Competing interests**

395 The authors declare that there are no competing interests.

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398 **Reference**

399

400 Aranami, K., & Tsunogai, S. (2004). Seasonal and regional comparison of oceanic and  
401 atmospheric dimethylsulfide in the northern North Pacific: Dilution effects on  
402 its concentration during winter. *Journal of Geophysical Research: Atmospheres*,  
403 109(D12). <https://doi.org/10.1029/2003jd004288>

404 Bock, J., Michou, M., Nabat, P., Abe, M., Mulcahy, J. P., Olivié, D. J. L., Schwinger, J.,  
405 Suntharalingam, P., Tjiputra, J., van Hulten, M., Watanabe, M., Yool, A., &  
406 Séférian, R. (2021). Evaluation of ocean dimethylsulfide concentration and  
407 emission in CMIP6 models. *Biogeosciences*, 18(12), 3823-3860.  
408 <https://doi.org/10.5194/bg-18-3823-2021>

409 Bopp, L., Aumont, O., Belviso, S., & Monfray, P. (2003). Potential impact of climate  
410 change on marine dimethyl sulfide  
411 emissions. *Tellus B: Chemical and Physical Meteorology*, 55(1).  
412 <https://doi.org/10.3402/tellusb.v55i1.16359>

413 Breiman, L. (2001). Random Forests. *Machine Learning*, 45, 5-32.

414 Charlson, R. J., Lovelock, J. E., Andreae, M. O., & Warren, S. (1987). Oceanic  
415 phytoplankton atmospheric sulphur cloud albedo and climate. *Nature*,  
416 326(6114), 655-661. <https://doi.org/10.1038/326655a0>

417 del Valle, D. A., Kieber, D. J., John, B., & Kiene, R. P. (2007). Light-stimulated

418 production of dissolved DMSO by a particle-associated process in the Ross Sea,  
419 Antarctica. *Limnology and Oceanography*, 52(6), 2456-2466.  
420 <https://doi.org/10.4319/lo.2007.52.6.2456>

421 Derevianko, G. J., Deutsch, C., & Hall, A. (2009). On the relationship between ocean  
422 DMS and solar radiation. *Geophysical Research Letters*, 36(17).  
423 <https://doi.org/10.1029/2009gl039412>

424 Gabric, A. J., Qu, B., Rotstayn, L., & Shephard, J. (2013). Global simulations of the  
425 impact on contemporary climate of a perturbation to the sea-to-air flux of  
426 dimethylsulfide. *Australian Meteorological and Oceanographic Journal*, 63(3),  
427 365-376. <https://doi.org/10.22499/2.6303.002>

428 Galí, M., Devred, E., Levasseur, M., Royer, S.-J., & Babin, M. (2015). A remote sensing  
429 algorithm for planktonic dimethylsulfoniopropionate (DMSP) and an analysis  
430 of global patterns. *Remote Sensing of Environment*, 171, 171-184.  
431 <https://doi.org/10.1016/j.rse.2015.10.012>

432 Galí, M., Saló, V., Almeda, R., Calbet, A., & Simó, R. (2011). Stimulation of gross  
433 dimethylsulfide (DMS) production by solar radiation. *Geophysical Research*  
434 *Letters*, 38(15). <https://doi.org/10.1029/2011gl048051>

435 Galí, M., & Simó, R. (2015). A meta-analysis of oceanic DMS and DMSP cycling  
436 processes: Disentangling the summer paradox. *Global Biogeochemical Cycles*,



- 29(4), 496-515. <https://doi.org/10.1002/2014gb004940>
- Galí, M., Simó, R., Vila-Costa, M., Ruiz-González, C., Gasol, J. M., & Matrai, P. (2013). Diel patterns of oceanic dimethylsulfide (DMS) cycling: Microbial and physical drivers. *Global Biogeochemical Cycles*, 27(3), 620-636. <https://doi.org/10.1002/gbc.20047>
- Gardner, M. W., & Dorling, S. (1998). Artificial neural networks (the multilayer perceptron)—a review of applications in the atmospheric sciences. *Atmospheric Environment*, 32(14), 2627-2636. [https://doi.org/10.1016/S1352-2310\(97\)00447-0](https://doi.org/10.1016/S1352-2310(97)00447-0)
- Goddijn-Murphy, L., Woolf, D. K., & Marandino, C. (2012). Space-based retrievals of air-sea gas transfer velocities using altimeters: Calibration for dimethyl sulfide. *Journal of Geophysical Research: Oceans*, 117(C8). <https://doi.org/10.1029/2011jc007535>
- Hajima, T., Watanabe, M., Yamamoto, A., Tatebe, H., Noguchi, M. A., Abe, M., Ohgaito, R., Ito, A., Yamazaki, D., Okajima, H., Ito, A., Takata, K., Ogochi, K., Watanabe, S., & Kawamiya, M. (2020). Development of the MIROC-ES2L Earth system model and the evaluation of biogeochemical processes and feedbacks. *Geoscientific Model Development*, 13(5), 2197-2244. <https://doi.org/10.5194/gmd-13-2197-2020>
- Hamilton, D. S., Lee, L. A., Pringle, K. J., Reddington, C. L., Spracklen, D. V., & Carslaw, K. S. (2014). Occurrence of pristine aerosol environments on a polluted planet. *Proc Natl Acad Sci U S A*, 111(52), 18466-18471. <https://doi.org/10.1073/pnas.1415440111>
- Hatton, A. D., & Wilson, S. T. (2007). Particulate dimethylsulphoxide and dimethylsulphoniopropionate in phytoplankton cultures and Scottish coastal waters. *Aquatic Sciences*, 69(3), 330-340. <https://doi.org/10.1007/s00027-007-0891-4>
- Hirata, T., Hardman-Mountford, N. J., Brewin, R. J. W., Aiken, J., Barlow, R., Suzuki, K., Isada, T., Howell, E., Hashioka, T., Noguchi-Aita, M., & Yamanaka, Y. (2011). Synoptic relationships between surface Chlorophyll-a and diagnostic pigments specific to phytoplankton functional types. *Biogeosciences*, 8(2), 311-327. <https://doi.org/10.5194/bg-8-311-2011>
- Joge, S. D., Mansour, K., Simo, R., Gali, M., Steiner, N., Saiz-Lopez, A., & Mahajan, A. S. (2025). Climate warming increases global oceanic dimethyl sulfide emissions. *Proc Natl Acad Sci U S A*, 122(23), e2502077122. <https://doi.org/10.1073/pnas.2502077122>
- Kinsey, J. D., Kieber, D. J., & Neale, P. J. (2016). Effects of iron limitation and UV radiation on *Phaeocystis antarctica* growth and dimethylsulfoniopropionate, dimethylsulfoxide and acrylate concentrations. *Environmental Chemistry*, 13(2). <https://doi.org/10.1071/en14275>
- Lana, A., Simó, R., Vallina, S. M., & Dachs, J. (2011). Re-examination of global



- 478 emerging patterns of ocean DMS concentration. *Biogeochemistry*, 110(1-3),  
479 173-182. <https://doi.org/10.1007/s10533-011-9677-9>
- 480 Liss, P. S., & Merlivat, L. (1986). *Air-sea gas exchange rates: Introduction and*  
481 *synthesis*.
- 482 Liss, P. S., & Slater, P. G. (1974). Flux of Gases across the Air-Sea Interface. *Nature*,  
483 247, 181-184.
- 484 Longhurst, A. (1998). *Ecological Geography of the Sea* (2nd ed. ed.). Acad. Press.
- 485 McNabb, B. J., & Tortell, P. D. (2022). Improved prediction of dimethyl sulfide (DMS)  
486 distributions in the northeast subarctic Pacific using machine-learning  
487 algorithms. *Biogeosciences*, 19(6), 1705-1721. [https://doi.org/10.5194/bg-19-](https://doi.org/10.5194/bg-19-1705-2022)  
488 [1705-2022](https://doi.org/10.5194/bg-19-1705-2022)
- 489 Michou, M., Nabat, P., Saint-Martin, D., Bock, J., Decharme, B., Mallet, M., Roehrig,  
490 R., Séférian, R., Sénési, S., & Voldoire, A. (2020). Present-Day and Historical  
491 Aerosol and Ozone Characteristics in CNRM CMIP6 Simulations. *Journal of*  
492 *Advances in Modeling Earth Systems*, 12(1).  
493 <https://doi.org/10.1029/2019ms001816>
- 494 Mulcahy, J. P., Johnson, C., Jones, C. G., Povey, A. C., Scott, C. E., Sellar, A., Turnock,  
495 S. T., Woodhouse, M. T., Abraham, N. L., Andrews, M. B., Bellouin, N., Browse,  
496 J., Carslaw, K. S., Dalvi, M., Folberth, G. A., Glover, M., Grosvenor, D. P.,  
497 Hardacre, C., Hill, R.,... Yool, A. (2020). Description and evaluation of aerosol  
498 in UKESM1 and HadGEM3-GC3.1 CMIP6 historical simulations.  
499 *Geoscientific Model Development*, 13(12), 6383-6423.  
500 <https://doi.org/10.5194/gmd-13-6383-2020>
- 501 Nemcek, N., Ianson, D., & Tortell, P. D. (2008). A high-resolution survey of DMS,  
502 CO<sub>2</sub>, and O<sub>2</sub>/Ar distributions in productive coastal waters. *Global*  
503 *Biogeochemical Cycles*, 22(2). <https://doi.org/10.1029/2006gb002879>
- 504 Quinn, P. K., & Bates, T. S. (2011). The case against climate regulation via oceanic  
505 phytoplankton sulphur emissions. *Nature*, 480(7375), 51-56.  
506 <https://doi.org/10.1038/nature10580>
- 507 Seferian, R., Berthet, S., Yool, A., Palmieri, J., Bopp, L., Tagliabue, A., Kwiatkowski,  
508 L., Aumont, O., Christian, J., Dunne, J., Gehlen, M., Ilyina, T., John, J. G., Li,  
509 H., Long, M. C., Luo, J. Y., Nakano, H., Romanou, A., Schwinger,  
510 J.,... Yamamoto, A. (2020). Tracking Improvement in Simulated Marine  
511 Biogeochemistry Between CMIP5 and CMIP6. *Curr Clim Change Rep*, 6(3),  
512 95-119. <https://doi.org/10.1007/s40641-020-00160-0>
- 513 Simó, R., Archer, S. D., Pedrós-Alió, C., Gilpin, L., & Stelfox-Widdicombe, C. E.  
514 (2002). Coupled dynamics of dimethylsulfoniopropionate and dimethylsulfide  
515 cycling and the microbial food web in surface waters of the North Atlantic.  
516 *Limnology and Oceanography*, 47(1), 53-61.  
517 <https://doi.org/10.4319/lo.2002.47.1.0053>
- 518 Simó, R., & Dachs, J. (2002). Global ocean emission of dimethylsulfide predicted from



- 519 biogeophysical data. *Global Biogeochemical Cycles*, 16(4).  
 520 <https://doi.org/10.1029/2001gb001829>
- 521 Simó, R., & Vila-Costa, M. (2006). Ubiquity of algal dimethylsulfoxide in the surface  
 522 ocean: Geographic and temporal distribution patterns. *Marine Chemistry*,  
 523 100(1-2), 136-146. <https://doi.org/10.1016/j.marchem.2005.11.006>
- 524 Six, K. D., Kloster, S., Ilyina, T., Archer, S. D., Zhang, K., & Maier-Reimer, E. (2013).  
 525 Global warming amplified by reduced sulphur fluxes as a result of ocean  
 526 acidification. *Nature Climate Change*, 3(11), 975-978.  
 527 <https://doi.org/10.1038/nclimate1981>
- 528 Spiese, C. E., & Tatarkov, E. A. (2014). Dimethylsulfoxide reduction activity is linked  
 529 to nutrient stress in *Thalassiosira pseudonana* NCMA 1335. *Marine Ecology*  
 530 *Progress Series*, 507, 31-38. <https://doi.org/10.3354/meps10842>
- 531 Stefels, J. (2000). Physiological aspects of the production and conversion of DMSP in  
 532 marine algae and higher plants. *Journal of Sea Research*, 43(3), 183-197.  
 533 [https://doi.org/10.1016/S1385-1101\(00\)00030-7](https://doi.org/10.1016/S1385-1101(00)00030-7)
- 534 Tjiputra, J. F., Schwinger, J., Bentsen, M., Morée, A. L., Gao, S., Bethke, I., Heinze, C.,  
 535 Goris, N., Gupta, A., He, Y.-C., Olivié, D., Seland, Ø., & Schulz, M. (2020).  
 536 Ocean biogeochemistry in the Norwegian Earth System Model version 2  
 537 (NorESM2). *Geoscientific Model Development*, 13(5), 2393-2431.  
 538 <https://doi.org/10.5194/gmd-13-2393-2020>
- 539 Vance, T. R., Davidson, A. T., Thomson, P. G., Levasseur, M., Lizotte, M., Curran, M.  
 540 A. J., & Jones, G. B. (2013). Rapid DMSP production by an Antarctic  
 541 phytoplankton community exposed to natural surface irradiances in late spring.  
 542 *Aquatic Microbial Ecology*, 71(2), 117-129. <https://doi.org/10.3354/ame01670>
- 543 Wang, S., Maltrud, M. E., Burrows, S. M., Elliott, S. M., & Cameron-Smith, P. (2018).  
 544 Impacts of Shifts in Phytoplankton Community on Clouds and Climate via the  
 545 Sulfur Cycle. *Global Biogeochemical Cycles*, 32(6), 1005-1026.  
 546 <https://doi.org/10.1029/2017gb005862>
- 547 Wanninkhof, R. (2014). Relationship between wind speed and gas exchange over the  
 548 ocean revisited. *Limnology and Oceanography: Methods*, 12(6), 351-362.  
 549 <https://doi.org/10.4319/lom.2014.12.351>
- 550 Watanabe, Y. W., Yoshinari, H., Sakamoto, A., Nakano, Y., Kasamatsu, N., Midorikawa,  
 551 T., & Ono, T. (2007). Reconstruction of sea surface dimethylsulfide in the North  
 552 Pacific during 1970s to 2000s. *Marine Chemistry*, 103(3-4), 347-358.  
 553 <https://doi.org/10.1016/j.marchem.2006.10.004>
- 554 Webb, A. L., Malin, G., Hopkins, F. E., Ho, K. L., Riebesell, U., Schulz, K. G., Larsen,  
 555 A., & Liss, P. S. (2016). Ocean acidification has different effects on the  
 556 production of dimethylsulfide and dimethylsulfoniopropionate measured in  
 557 cultures of *Emiliania huxleyi* and a mesocosm study: a comparison of laboratory  
 558 monocultures and community interactions. *Environmental Chemistry*, 13(2).  
 559 <https://doi.org/10.1071/en14268>
- 560 Yool, A., Popova, E. E., & Anderson, T. R. (2013). MEDUSA-2.0: an intermediate  
 561 complexity biogeochemical model of the marine carbon cycle for climate



562 change and ocean acidification studies. *Geoscientific Model Development*, 6(5),  
563 1767-1811. <https://doi.org/10.5194/gmd-6-1767-2013>  
564 Zhao, J., Zhang, Y., Bie, S., Bilsback, K. R., Pierce, J. R., & Chen, Y. (2024). Changes  
565 in global DMS production driven by increased CO<sub>2</sub> levels and its impact on  
566 radiative forcing. *npj Climate and Atmospheric Science*, 7(1).  
567 <https://doi.org/10.1038/s41612-024-00563-y>  
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571 **Table 1.** Summary of DMS trends for historical and future scenarios at different  
572 durations from 1850 to 2100 under the high emission scenario (SSP5.85). The trends  
573 were calculated according to the method described by Joge et al. (2025) using a  
574 bootstrap approach. The unit is % per decade, which indicates the relative changes in  
575 DMS concentration compared to the initial year of each period.

	Trend $\pm$ SD % decade <sup>-1</sup>				
Model	1850-1900	1900-1950	1950-2014	2015-2050	2050-2100
<b>CNRM-ESE2-1</b>	-0.06 $\pm$ 0.01	0.02 $\pm$ 0.01	0.06 $\pm$ 0.01	0.41 $\pm$ 0.03	1.19 $\pm$ 0.04
<b>MIROC-ES2L</b>	-0.01 $\pm$ 0.02	0.15 $\pm$ 0.02	0.43 $\pm$ 0.02	0.82 $\pm$ 0.04	0.69 $\pm$ 0.02
<b>NorESM2-MM</b>	0.02 $\pm$ 0.04	-0.09 $\pm$ 0.03	0.06 $\pm$ 0.03	-0.69 $\pm$ 0.13	-1.01 $\pm$ 0.09
<b>UKESM1-0-LL</b>	0.04 $\pm$ 0.03	-0.09 $\pm$ 0.03	-0.18 $\pm$ 0.03	-1.51 $\pm$ 0.08	-1.26 $\pm$ 0.06
<b>Joge25</b>	0.07 $\pm$ 0.02	0.01 $\pm$ 0.02	0.06 $\pm$ 0.02	-0.46 $\pm$ 0.03	-0.58 $\pm$ 0.02
<b>this study</b>	0.14 $\pm$ 0.08	-0.12 $\pm$ 0.09	-0.45 $\pm$ 0.06	-0.40 $\pm$ 0.13	-0.89 $\pm$ 0.08

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579 **Table 2.** Summary of DMS flux trends for historical and future scenarios at different  
580 durations from 1850 to 2100 under the high emission scenario (SSP5.85). The trends  
581 are calculated according to the method described by Joge et al. (2025) using a  
582 bootstrap method. The unit is % per decade, which indicates the relative changes in  
583 DMS flux compared to the initial year of each period.

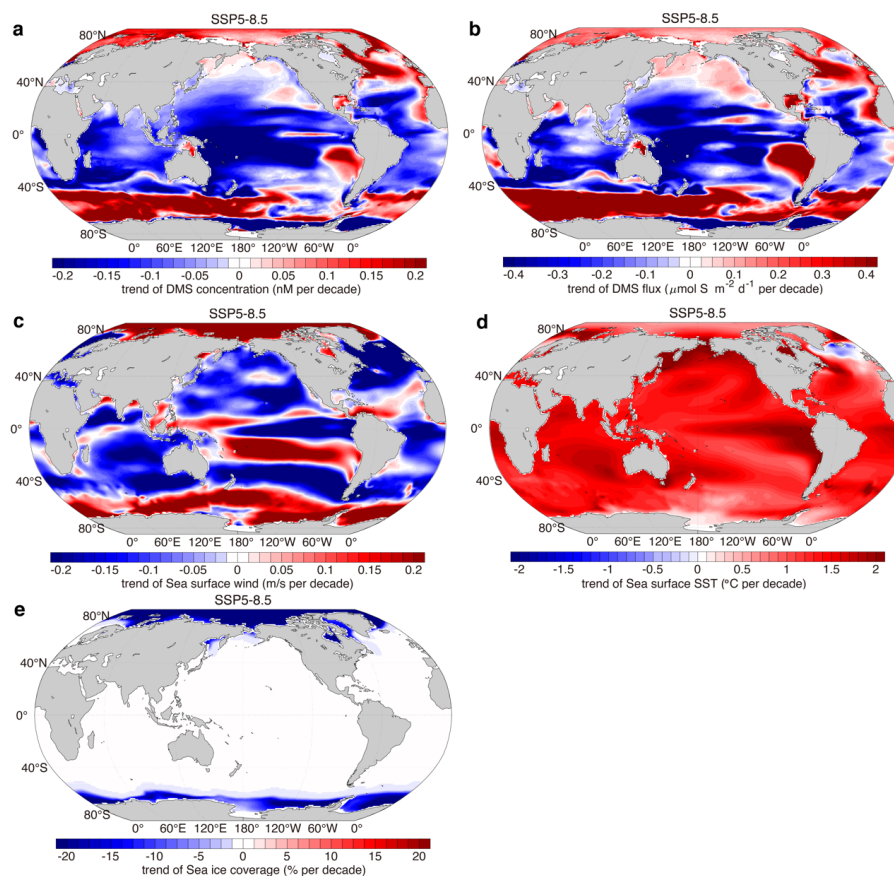
	Trend $\pm$ SD % decade <sup>-1</sup>				
Model	1850-1900	1900-1950	1950-2014	2015-2050	2050-2100
<b>CNRM-ESE2-1</b>	-0.04 $\pm$ 0.05	0.03 $\pm$ 0.04	0.37 $\pm$ 0.03	0.68 $\pm$ 0.13	1.32 $\pm$ 0.07
<b>MIROC-ES2L</b>	-0.005 $\pm$ 0.05	0.18 $\pm$ 0.03	0.33 $\pm$ 0.04	1.16 $\pm$ 0.06	1.45 $\pm$ 0.03
<b>NorESM2-MM</b>	-0.009 $\pm$ 0.06	-0.13 $\pm$ 0.04	0.06 $\pm$ 0.04	-0.41 $\pm$ 0.13	-0.51 $\pm$ 0.08
<b>UKESM1-0-LL</b>	0.03 $\pm$ 0.04	-0.06 $\pm$ 0.03	0.17 $\pm$ 0.02	-0.71 $\pm$ 0.09	0.03 $\pm$ 0.06
<b>Joge25</b>	0.04 $\pm$ 0.03	0.08 $\pm$ 0.02	0.26 $\pm$ 0.02	0.16 $\pm$ 0.03	0.37 $\pm$ 0.03
<b>this study</b>	0.01 $\pm$ 0.09	-0.04 $\pm$ 0.07	-0.07 $\pm$ 0.06	0.51 $\pm$ 0.16	-0.37 $\pm$ 0.11

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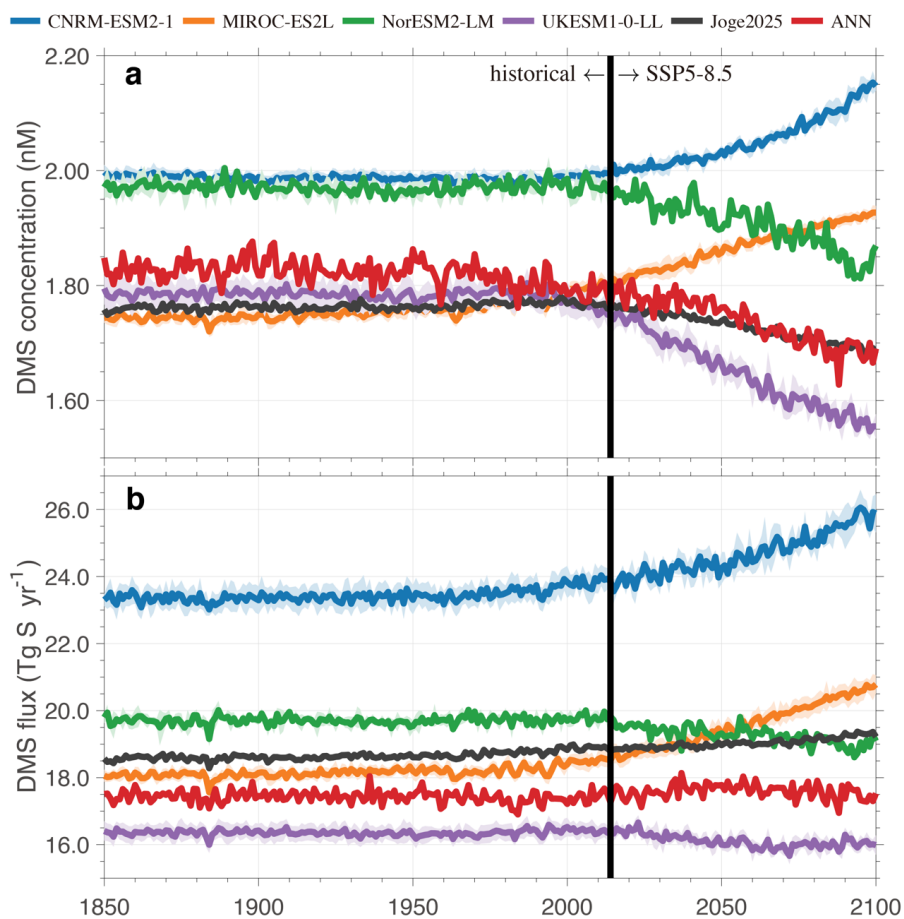
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**Figure 1.** Changing trends of concentration and flux DMS with assimilatory parameters from 2015 to 2100 under SSP5–8.5 simulation. **a**, trend of DMS concentration. **b**, trends of DMS flux. **c**, trend of sea surface wind speed. **d**, trend of sea surface temperature. **e**, trend of sea ice coverage.



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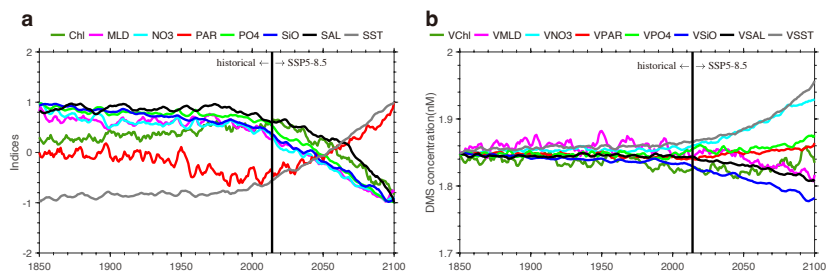
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595 **Figure 2.** Time series of mean annual global area-weighted surface ocean DMS  
596 concentration and DMS flux over 1850-2100 (CMIP6 historical and SSP5-8.5  
597 simulations). **a.** DMS concentration (nM). **b.** DMS flux (Tg S yr<sup>-1</sup>)

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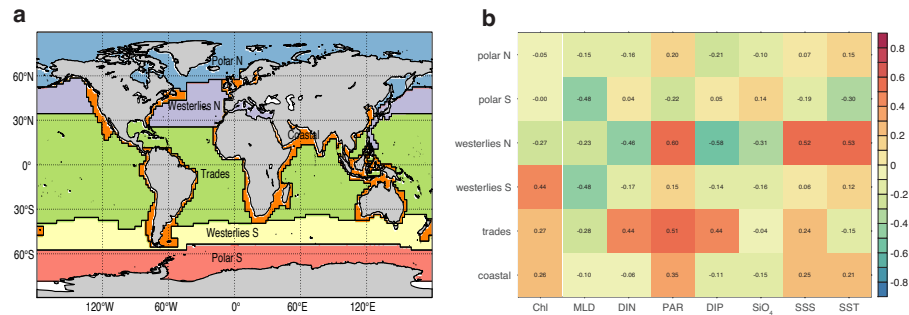
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**Figure 3.** Time series of input variables and DMS concentrations of sensitivity tests over 1850–2100. **a**, Time series of eight input environmental variables normalized to (-1,1). **b**, Time series of mean annual global area-weighted DMS concentrations of eight sensitivity tests.

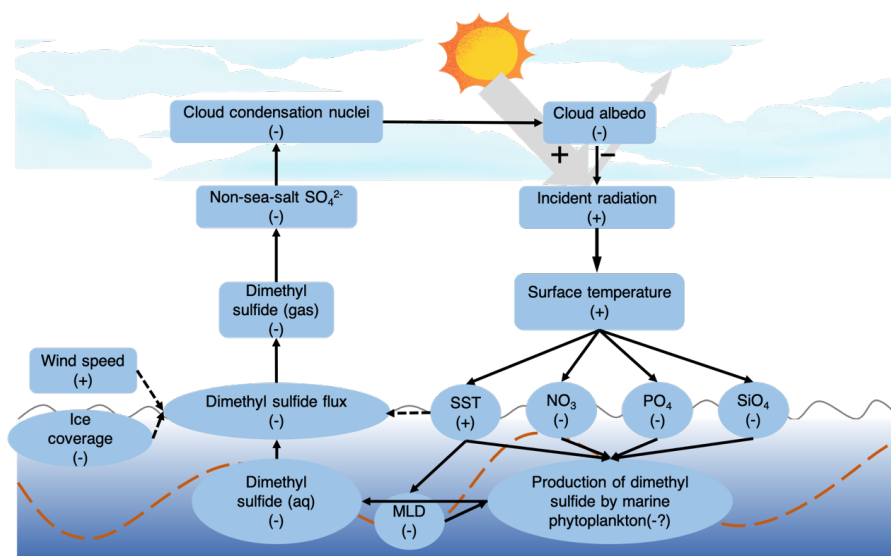




**Figure 4.** Correlation of DMS concentrations with environmental variables in six main regions. **a**, Six oceanic regions that were separated based on Longhurst's biomes (Longhurst, 1998). **b**, Correlation of DMS concentrations with eight input environmental variables in six oceanic regions.



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**Figure 5.** Modified diagram of the climate feedback loop of DMS.

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