



1 TROPOMI NO₂ for urban and polluted areas globally from 2019 to 2 2024

3 Daniel E. Huber¹, Gaige H. Kerr¹, M. Omar Nawaz², Sara Runkel³, Susan C. Anenberg¹, Daniel L.
4 Goldberg¹

5 ¹Department Environmental and Occupational Health, Milken Institute School of Public Health, George Washington
6 University, Washington, DC, USA

7 ²School of Earth and Environmental Science, Cardiff University, Cardiff, United Kingdom

8 ³National Center for Atmospheric Research, Boulder, CO, USA

9 *Correspondence to:* Daniel E. Huber (daniel.huber@gwu.edu)

10 **Abstract.** We present a global assessment of space-based urban nitrogen dioxide (NO₂) observation trends from 2019 to 2024
11 using annual and monthly mean tropospheric vertical column densities (VCDs) from the TROPOspheric Monitoring
12 Instrument (TROPOMI). Across 11,500 cities defined by the Global Human Settlement Layer-Settlement Model (GHS-
13 SMOD), we find population-weighted annual mean urban NO₂ VCDs declined between 2019 and 2024 in Asian (-17%),
14 European (-13%), and North American (-4%) cities, with seasonal decomposition indicating that most of the annual changes
15 are driven by wintertime concentration decreases. South American (-2%) cities exhibited lesser population-weighted changes
16 on average, while African (+3%) cities experienced a gradual increase in NO₂. Over this timeframe, Tehran had the largest
17 NO₂ VCDs ($>30 \times 10^{15}$ molecules cm⁻²) and Seoul had the largest reduction (-40%). We further identify changes in NO₂ near
18 fossil fuel operations and note conflict-related changes in NO₂, highlighting the responsiveness of satellite NO₂ to certain
19 societal disruptions. We then calculate NO₂ VCD urban enhancements (VCD_{ENH}) by removing background concentrations
20 from urban signatures and compare VCD_{ENH} to changes in nitrogen oxide (NO_x) emissions from the Emissions Database for
21 Global Atmospheric Research (EDGARv8.1), to highlight regions with potential inventory discrepancies. We find VCD_{ENH}
22 and EDGARv8.1 NO_x change at a similar rate from year to year in Europe and North America, with worse agreement in the
23 Global South. This work demonstrates the value in space-based remote sensing being an accountability agent for air pollution
24 emissions on a global scale and to identify changes in NO₂ in otherwise unmonitored regions.

25 1 Introduction

26 Nitrogen dioxide (NO₂) is a harmful air pollutant that originates from both anthropogenic and natural emissions sources,
27 including fossil fuel combustion, biomass burning, lightning, and soils (Dix et al., 2020; Jin et al., 2021; Schuman & Huntrieser,
28 2007; Huber et al., 2024), with fossil fuel combustion accounting for ~45% of total global nitrogen oxide emissions (Song et
29 al., 2021). Only a small amount of NO₂ is emitted from these sources directly, with nitric oxide (NO) being the primary
30 emissions product that quickly cycles to NO₂ in the presence of oxidants such as ozone (O₃) or peroxy radicals (HO₂ or RO₂).



31 The summed concentrations of NO and NO₂ are referred to as nitrogen oxides (NO_x = NO + NO₂), as the concentrations of
32 NO and NO₂ are inherently linked. NO₂ is more commonly targeted by regulatory measures than NO, as it constitutes the
33 majority of atmospheric NO_x concentrations and is linked to increased morbidity and mortality from long-term exposure,
34 particularly within urban environments (Chen et al., 2024). While NO_x is commonly associated with health risks, the direct
35 association between NO_x exposure and adverse health outcomes remains uncertain (Anenberg et al., 2022). Despite this, NO_x
36 contributes to known harmful secondary pollutants, including O₃ and fine particulate matter.

37 NO₂ concentrations are measured using: (1) in-situ monitoring, e.g. chemiluminescence analyzers at the surface, or (2) remote
38 sensing instrumentation leveraging the unique spectral properties of NO₂, that absorbs light most efficiently in the visible
39 portions (405 – 465 nm) of the electromagnetic spectrum (Lamsal et al., 2015). The latter method relies upon spectrometers
40 detecting in the UV-Visible spectral range to infer NO₂ vertical column densities (VCDs), defined as the summed concentration
41 of NO₂ in a column from the surface to an upper limit of the atmosphere, with the tropopause often used as the upper limit.
42 Spectrometers have been used to measure NO₂ VCDs from ground-level directed upward, from aircraft directed downward, or
43 from space-based satellites directed downward, including from the TROPOspheric Monitoring Instrument (TROPOMI)
44 onboard the Sentinel-5P satellite (Herman et al., 2009; Fishman et al., 2012; Veeffkind et al., 2012).

45 The earliest space-based spectrometers detecting NO₂ were flown on low-earth polar orbiting satellites, and were launched
46 within the mid-1990s to mid-2000s. These include the Global Ozone Monitoring Experiment (GOME; Burrows et al., 1999)
47 and GOME-2 satellites, the SCanning Imaging Absorption spectroMeter for Atmospheric CHartographY (SCIAMACHY;
48 Bovensmann et al., 1999) and the Ozone Monitoring Instrument (OMI; Levelt et al., 2006). The data collected using these
49 instruments provided unique insight into atmospheric chemistry and composition across the globe, including in mostly
50 unmonitored regions. OMI, launched in 2004, provided NO₂ VCDs at a spatial resolution of 13 x 24 km² at nadir and has
51 remained operable for more than two decades at the time this was written, providing a valuable long-term record of NO₂
52 globally. OMI remained the highest resolution space-based NO₂ product until TROPOMI launched in 2017, which ultimately
53 provided NO₂ VCDs at a spatial resolution of 3.5 x 5.5 km² at nadir. Observations at this resolution facilitated the evaluation
54 of satellite NO₂ at previously unprecedented spatial scales, including at the intra-urban level (Goldberg et al., 2021; Goldberg
55 et al., 2024).

56 NO₂ trends have been characterized in urban and broader environments using space-based instruments. Earlier satellite studies
57 used the GOME and SCHIAMACY satellites to identify increasing NO₂ VCD trends in China from the mid-1990s to the mid-
58 2000s (Richter et al., 2005; Stavrakou et al., 2008; Van der A et al., 2008), driven primarily by economic growth and
59 industrialization. Later studies, incorporating OMI observations, highlighted further increases in China through the early
60 2010s, with VCDs and satellite-inferred surface concentrations steadily declining since (Miyazaki et al., 2017; Jiang et al.,
61 2022). Europe has exhibited steady NO₂ VCD declines since the start of the satellite NO₂ record (Richter et al., 2005; Krotkov
62 et al., 2016; Duncan et al., 2016), driven largely by the implementation of various emissions control technologies. In the United
63 States, NO₂ concentrations increased through roughly 2005, then decreased substantially through the early to mid-2010s



64 (Lamsal et al., 2015), with VCD decreases more gradual since, in part due to an increased influence from regional background
65 NO₂ levels (Jiang et al., 2018; Dang et al., 2023). Additionally, numerous studies have highlighted the influence that the
66 COVID-19 pandemic had on NO₂ globally, with most regions exhibiting broad NO₂ decreases in 2020 during numerous
67 lockdowns and subsequent, regionally-distinct rebounds in emissions (Lonsdale et al., 2023; Fisher et al., 2024).

68 Satellite studies have been used to characterize trends within the urban environment specifically, using different methods to
69 characterize the urban extent. Geddes et al. (2016) used GOME, SCIAMACHY and GOME-2 oversampled to a 0.1° x 0.1°
70 grid to highlight NO₂ VCD trends globally, as well as in select urban areas, with the urban region defined as the surrounding
71 ~ 200 km x 200 km. Fioletov et al. (2022) and Fioletov et al. (2025) used urban density from the Gridded Population of the
72 World (SEDAC, 2017) as a proxy for the extent of the urban environment to identify changes in urban NO_x emissions.
73 Anenberg et al. (2022) used urban boundaries provided from the 2019 version of the Global Human Settlement Layer-
74 Settlement model (GHS-SMOD) to evaluate NO₂ trends from 2000 – 2019 using surface NO₂ estimates derived from
75 TROPOMI NO₂ and a land-use regression model.

76 Here, we use TROPOMI tropospheric NO₂ VCDs to quantify general NO₂ trends globally from 2019 to 2024, with a particular
77 focus on urban areas. The urban boundaries are defined by the 2023 version of GHS-SMOD, which provides urban cluster
78 boundaries for all urban regions globally. We evaluate urban NO₂ trends against a NO_x emissions database, and evaluate the
79 influence of different seasons on annual trends. We additionally note changes in select oil, gas, and other mining regions,
80 which exhibit the largest changes globally outside of urban areas. This study represents the first detailed global-scale analysis
81 of urban TROPOMI NO₂ trends from 2019 to 2024. Our findings illustrate how NO₂ responded to specific societal events
82 during this timeframe, such as the impact of clean air policies, population migration away from urban areas due to war, the
83 increased demand for fossil fuels and rare-Earth minerals, and the emergence and waning of a global pandemic. Furthermore,
84 by directly linking observed NO₂ urban enhancements with NO_x emission inventory data from the updated EDGARv8.1, our
85 work provides valuable insights into regions where emissions inventories align closely with observations, as well as areas
86 exhibiting potential inventory discrepancies. This work underscores the critical value of satellite-derived NO₂ as a tool for
87 urban air quality assessment and emissions management.

88 **2 Data and Methods**

89 **2.1 Global Human Settlement Layer Urban Cluster Boundaries**

90 The Global Human Settlement Layer-Settlement Model (GHS-SMOD; Schiavina et al., 2023) is a dataset developed by the
91 Joint Research Centre of the European Commission containing spatial boundaries and population estimates for all urban areas
92 globally with a population of at least 50,000, which can be used to subset gridded or spatially-disaggregated data for any built-
93 up area on Earth. GHS-SMOD uses satellite remote sensing to identify the spatial extent and boundaries of all cohesive built-
94 up areas globally at a spatial resolution of 1 km², with each separate, cohesive built-up area referred to as an “urban cluster”.



95 In this study, we use the terms “urban cluster” and “city” interchangeably, although we note that GHS-SMOD urban clusters
96 do not necessarily align with administrative city boundaries. The 2023 version of GHS-SMOD provides boundaries for
97 approximately 11,500 urban clusters, along with population estimates for the year 2020 (Fig. S1). We note that GHS-SMOD
98 urban clusters do not reflect the traditional boundaries of individual cities as we may understand them, and as such, GHS-
99 SMOD urban clusters can span multiple cities, regions or even countries. For example, the urban cluster encompassing San
100 Diego, California includes the city of San Diego, but also the adjacent surrounding suburbs, as well as the entirety of Tijuana,
101 Mexico (Fig. S2). In such cases, attribution of an urban cluster to one particular city is not possible.

102 We use the GHS-SMOD boundaries to subset monthly- and annually-averaged satellite NO₂ column concentration data for all
103 urban clusters, as described in Section 2.2.1.

104 2.2 TROPOMI NO₂ vertical column densities

105 The TROPOspheric Monitoring Instrument (TROPOMI) is a pushbroom spectrometer on board the Sentinel-5P satellite
106 traveling in low earth orbit, with approximately one overpass each afternoon (Veefkind et al., 2012). Launched in October,
107 2017, TROPOMI detects radiation in spectral bands ranging from the ultraviolet to shortwave infrared to infer concentrations
108 of various atmospheric constituents, including nitrogen dioxide (NO₂), which is best inferred from the near-UV and visible
109 portions of the spectrum. We use Level 3 monthly- and annually-averaged TROPOMI tropospheric NO₂ vertical column
110 densities (VCDs) on a 0.1° global grid (Goldberg, 2024), which were created by oversampling daily Level 2 TROPOMI NO₂
111 VCDs, with a spatial resolution of 3.5 x 5.5 km² at nadir, derived from version 2.4+ of the European Space Agency retrieval
112 algorithm (van Geffen et al., 2022). The TROPOMI NO₂ data used in this study span six full calendar years from January 2019
113 to December 2024 (Fig. 1); we use the RPRO version from 1 January 2019 – 25 July 2022 and the OFFL version from 26 July
114 2022 – 31 December 2024. On 7 September 2024 there was an update of the surface reflectivity assumptions and on 16
115 November 2024 there was an update to the cloud retrieval, both of which induce a small positive step change in the data, but
116 likely does not meaningfully affect the 2024 annual average.

117

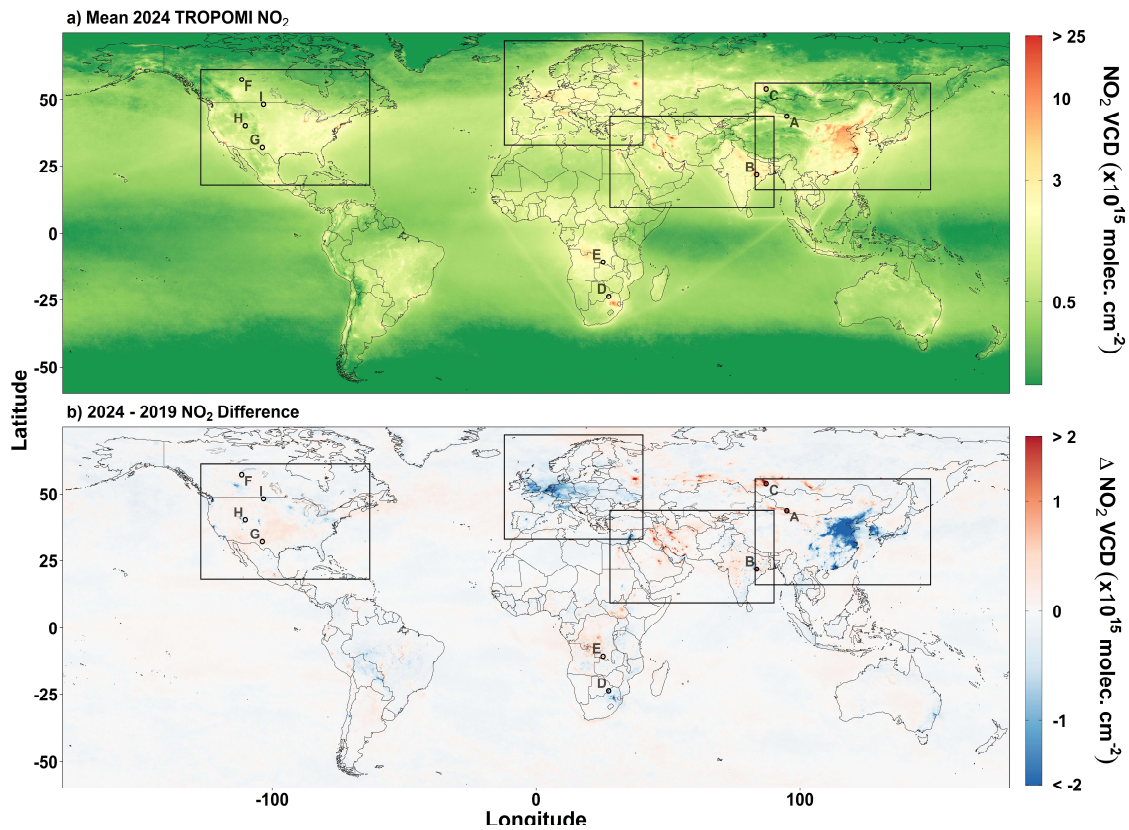


Figure 1: (a) Global 2024 annual average NO₂ VCDs colored on a log-scale and (b) the difference in VCD from 2019 to 2024 colored on a symmetric log-scale. Points labeled A-I correspond with locations of oil, gas and mining operations highlighted in Fig. 3. Boxes indicate select focus regions in Section 3.

2.2.1 Quantifying average TROPOMI NO₂ VCDs for GHS-SMOD urban clusters

For each urban cluster, we subset the oversampled TROPOMI data for grid cells that are located within 0.1° of the urban cluster boundary. For most cities, this results in approximately 20-25 grid cells, depending on the extent of the individual cluster. Given that the spatial resolution of GHS-SMOD is roughly an order of magnitude finer than that of the oversampled TROPOMI data (1 km vs. 0.1°) we interpolate the subsetting TROPOMI data to the 0.01° × 0.01° resolution of GHS-SMOD using a nearest neighbor approach. We then calculate an area-weighted average of interpolated grid cells that have a grid cell center falling within the urban cluster boundary (Fig. S2). This approach allows for the portions of oversampled 0.1° × 0.1° grid cells that may not be centered within an urban cluster boundary, but that still overlap with a cluster, to be accounted for within the average NO₂ column estimate.



131 To evaluate the changes in VCDs for broader regions, e.g. countries containing multiple urban clusters, we can calculate a
132 population-weighted average VCD, taking into account varying population sizes in different urban clusters.

$$133 \quad VCD_{PW} = \frac{\sum_i (POP_i \times VCD_i)}{\sum_i (POP_i)}, \quad (1)$$

134 In Eq. 1, VCD_{PW} represents the population-weighted VCD for a given country, POP_i represents the 2020 GHS-SMOD-
135 estimated population for a given urban cluster i , and VCD_i represents the mean NO_2 VCD for i .

136 2.3 Accounting for background NO_2

137 To account for changes in upwind background NO_2 concentrations that may influence urban NO_2 VCDs, we quantify an urban
138 NO_2 enhancement.

$$139 \quad VCD_{ENH} = VCD_{UC} - VCD_{BG}, \quad (2)$$

140 In Eq. 2, VCD_{ENH} is the urban NO_2 VCD enhancement, VCD_{UC} is the NO_2 VCD within each urban cluster as described in
141 Section 2.2.1, and VCD_{BG} is the background concentration for an urban cluster. We define VCD_{BG} as the 10th percentile of
142 NO_2 VCDs extending 0.5 degrees in any direction from an UC boundary (de Gouw et al., 2020).

143 2.4 EDGAR NO_x emissions

144 We use version 8.1 of the Emissions Database for Global Atmospheric Research (EDGARv8.1; Crippa et al., 2024) to evaluate
145 NO_x emissions. EDGAR provides summed total and sector-specific NO_x emissions at $0.1^\circ \times 0.1^\circ$ spatial resolution globally.
146 EDGAR NO_x emissions include contributions from energy generation, industrial sources, transportation, residential sources
147 and agriculture. EDGAR emissions are produced using a bottom-up method that combines activity data together with sector-
148 specific emissions factors to produce gridded annual emissions. Similar to the handling of TROPOMI data (Sec. 2.21), we use
149 GHS-SMOD to quantify mean NO_x emissions for each urban cluster.

150 3 Global TROPOMI NO_2 vertical column densities from 2019 to 2024

151 The following subsections describe the NO_2 VCDs and trends in four global subregions: Asia and Oceania, Africa, Europe,
152 North and South America.

153 3.1 Asia and Oceania

154 North and East China, one of the most populated regions globally with approximately 11% of the 1000 largest GHS-SMOD
155 cities, produced the broadest continuous expanse of 2024 annual mean NO_2 VCDs at or above 5×10^{15} molecules cm^{-2} (Fig.
156 2a). Despite this, substantial decreases were observed in this region from 2019 to 2024 (Fig. 2b). While NO_2 concentrations
157 had already been decreasing in China prior to 2019 (Liu et al., 2016; de Foy et al., 2016), the decrease accelerated after the



onset of the COVID-19 pandemic, coinciding with reduced emissions during numerous lockdowns throughout the country from 2020 to 2022 (Zheng et al., 2021; Ma et al., 2023; Zhao et al., 2024). The decrease in NO₂ also coincided with general Chinese government policies directed at reducing emissions, including stricter emissions controls for industrial sources, energy generation and the transportation sector (Shi et al., 2022; Li et al., 2024). Few regions in China experienced increased VCDs, with the most notable increases occurring outside of major urban areas. The most substantial increase in VCD over China through 2024 was observed in the sparsely-populated Santanghu Basin (Fig. 3a), a region in eastern Xinjiang Province with a relatively nascent coal mining industry (Zhang et al., 2018; Liu et al., 2018). Annual mean NO₂ VCDs in the basin increased by 1.9×10^{15} molecules cm⁻², or +172%, from 2019 to 2024. The expansion of mining operations is clearly evident in visible satellite imagery (Fig. S3).

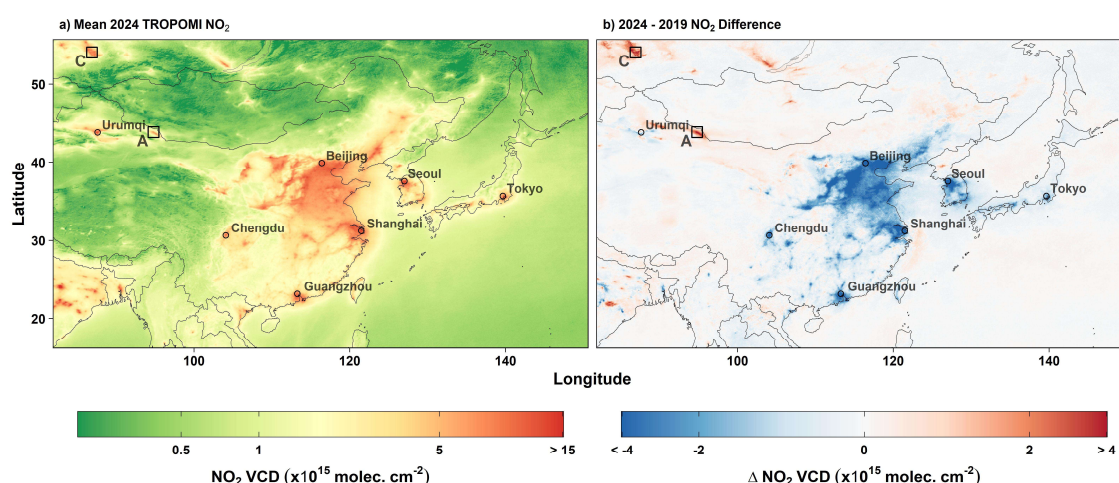


Figure 2: (a) Mean 2024 TROPOMI NO₂ VCDs and (b) relative changes in TROPOMI VCDs for from 2019 to 2024, centered on East Asia. Labeled black squares indicates the locations of mining regions highlighted in Fig. 3.

In India, elevated NO₂ near numerous coal-fired power plants and coal mines is a common feature (Panda et al., 2023), evidenced by the many apparent point sources in the 2024 annual average TROPOMI VCDs throughout the country (Fig. 4a). NO₂ VCDs increased at many of these points sources from 2019 to 2024 (Fig. 4b), suggesting an increase in emissions from energy production and use. The largest regional increase in VCD anywhere in India from 2019 to 2024 ($+2.1 \times 10^{15}$ molecules cm⁻²; +37%) was observed in the Ib Valley in northwestern Odisha state (Fig. 3b), a region with multiple surface coal mines and coal-fired power plants (Varma et al., 2015). NO₂ VCDs near numerous other coal mines and power plants throughout India exhibited changes, but NO₂ VCD increases were more prevalent than decreases. Of the major urban regions in India, the



largest decreases from 2019 to 2024 were observed in New Delhi (-1.6×10^{15} molecules cm^{-2} ; -18%) and Mumbai (-1.0×10^{15} molecules cm^{-2} ; -15%).

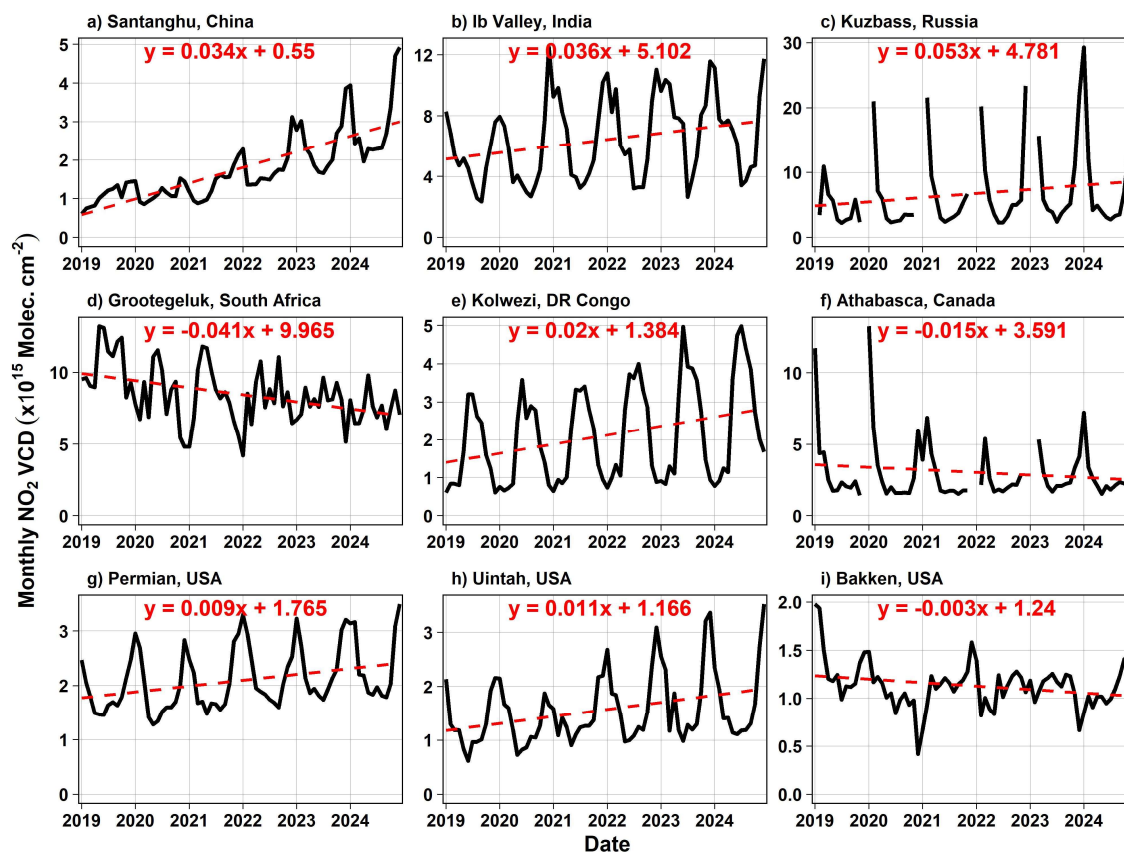


Figure 3: Monthly time series of NO₂ VCDs over select oil, gas, and other mining regions. Black lines denote monthly mean VCDs, and red lines represent trends characterized by ordinary least-squares regression for each site. The slope of each trend line represents the change in NO₂ VCD per month, with the y-intercept representing the intercept for January, 2019. Months with missing data lacked quality-assured TROPOMI observations. Note the differing y-axis extents for each panel.

Urban regions in Middle Eastern countries experienced some of the highest NO₂ VCDs globally in the TROPOMI record. Near the Iranian capital of Tehran, 2024 annual average NO₂ VCDs of individual grid cells exceeded 40×10^{15} molecules cm^{-2} (Fig. 4a), the highest urban annual average among all global cities. Much of the Middle East exhibited substantial increases in population-weighted, urban NO₂ VCDs from 2019 to 2024, most notably in regions of Saudi Arabia (+5%), Iraq (+18%), and Iran (+10%), with broad increases that extend beyond the urban environment. One of the most salient VCD decreases in the

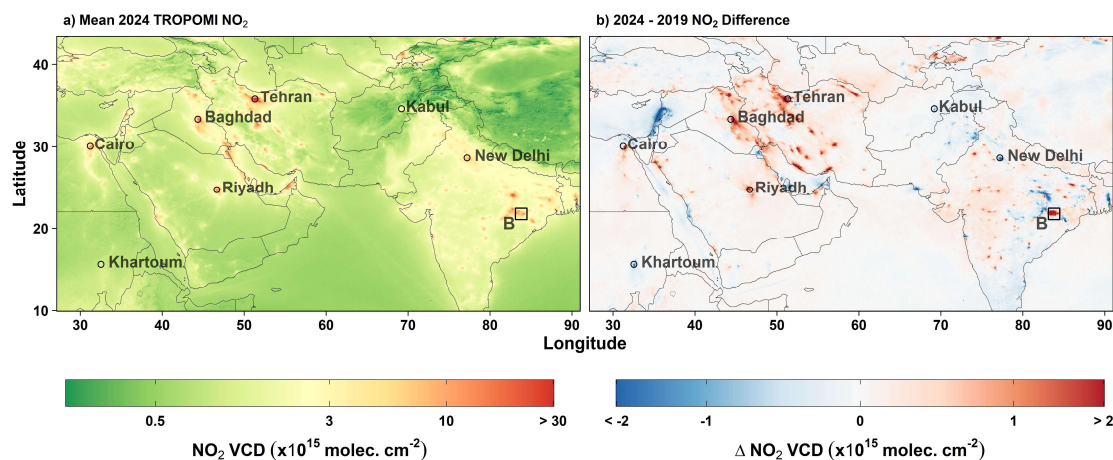


193 Middle East occurred in Lebanon (-39%), coinciding with the country's severe economic and financial crisis that began in late
194 2019 (Harake et al., 2019). VCD decreases through 2024 were particularly stark in the Lebanese capital Beirut (-6.7×10^{15}
195 molecules cm^{-2} ; -37%). Additional Middle Eastern countries that exhibited decreased urban NO_2 VCDs through 2024 include
196 much of Israel (-27%), Kuwait (-5%), Qatar (-17%), and Afghanistan (-13%).

197 Nearly all urban regions in eastern Russia (Siberia) exhibited increased NO_2 VCDs, as did regions coinciding with known
198 mining operations. In the Kuzbass Region of Siberia, one of Russia's largest coal mining regions, annual mean VCDs increased
199 by 2.4×10^{15} molecules cm^{-2} from 2019 to 2024, representing a 58% increase (Fig. 3c). A previous study identified a correlation
200 between space-based NO_2 observations and regional coal production in the Kuzbass region (Labzovskii et al., 2022), providing
201 relevant context for the observed VCD increases.

202 Other notable changes in NO_2 VCD in Asia include extensive decreases throughout Japan (-22%), South Korea (-39%),
203 Thailand (-7%), Pakistan (-14%) and Australia (-14%). Urban increases were observed in much of Central Asia, including
204 Turkmenistan (+21%), Kazakhstan (+22%) Mongolia (+75%).

205



206

207 **Figure 4:** Same as Fig. 2, but centered on the Middle East. Labeled black squares indicate the locations of mining regions highlighted
208 in Fig. 3.

209 3.2 Africa

210 Johannesburg, South Africa and the surrounding region exhibited the largest NO_2 VCD for the African continent in 2024 (Fig.
211 S4). Numerous surface coal mines and coal-fired power plants, particularly to the east of Johannesburg, contribute to the
212 region's NO_2 signature (Shikwambana et al., 2020). Despite these elevated NO_2 levels, 2024 mean NO_2 VCDs in the city of
213 Johannesburg were 8% lower than in 2019. Northwest of Johannesburg in Limpopo Province, mining operations at the



214 Grooteegeluk surface coal mine, together with two adjacent power plants (Faure et al., 2010; Shikwambana et al., 2020), produce
215 one of the largest NO₂ point sources in Africa, despite annual mean NO₂ VCDs at the site decreasing by 3.5×10^{15} molecules
216 cm⁻² from 2019 to 2024, or a decrease of 32% (Fig. 3d). The Cairo, Egypt urban region, in Northern Africa, represents the
217 second largest urban NO₂ signature in Africa in 2024. The 2024 annual mean NO₂ VCD in Cairo was 9.4×10^{15} molecules cm⁻²,
218 and elevated VCDs extend along the Nile River south of Cairo, as well as north into the Nile River Delta. Cairo exhibited
219 increased VCDs from 2019 to 2024 (+8%), as did regions immediately adjacent to the Nile River, while regions north into the
220 Nile River Delta exhibited decreased NO₂ VCDs.

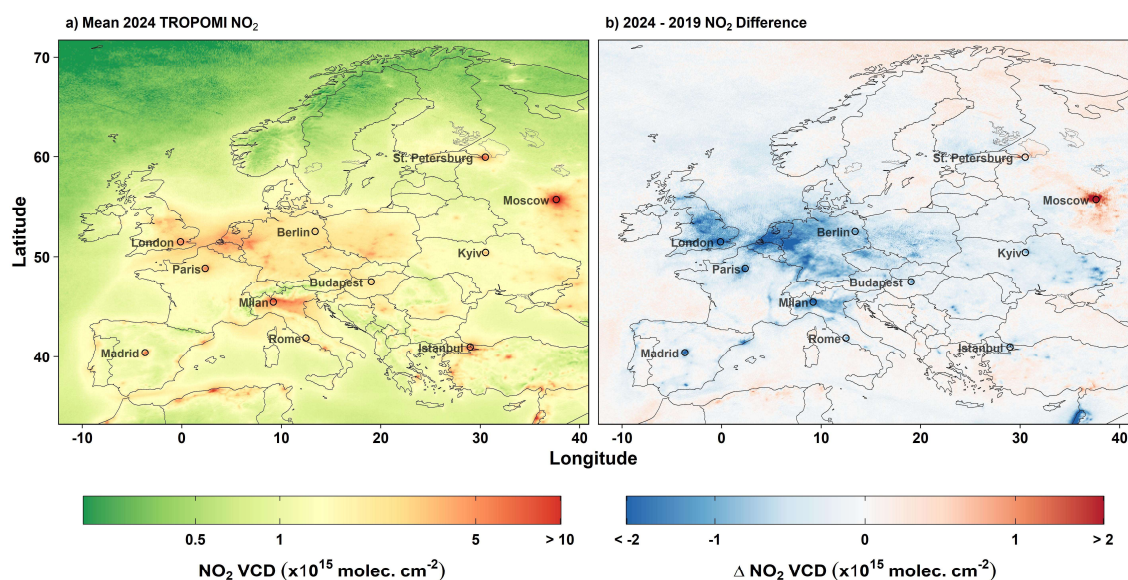
221 In the Sudanese capital of Khartoum, NO₂ VCDs started decreasing in 2023, coinciding with the onset of conflict within Sudan
222 (Guo et al., 2023). This resulted in annual mean VCDs decreasing by 58% from 2019 to 2024 (Fig. S5). In a mining region
223 known as the Copperbelt in the south of the Democratic Republic of the Congo (DRC), broad NO₂ VCD increases were
224 observed, including at a large surface mine near Kolwezi. VCDs at the Kolwezi mine increased by 1.4×10^{15} molecules cm⁻²
225 from 2019 to 2024, or an increase of 64% (Fig. 3e). Numerous surface mines exist in the region, with most observing increases
226 in NO_x emissions from mining operations in recent years (Martínez-Alonso et al., 2023). Throughout the remainder of Africa,
227 moderate VCD enhancements were observed near most urban centers, with mean VCDs near most cities typically at or below
228 4×10^{15} molecules cm⁻² (Fig. 1a). Along the African Mediterranean coast, most urban areas showed increased NO₂ VCDs
229 through 2024. Other national capitals and major cities exhibited increased VCDs, including Abidjan, Ivory Coast (+41%);
230 Addis Ababa, Ethiopia (+34%); Kinshasa, DRC (+20%); and Dakar, Senegal (+15%).

231 3.3 Europe

232 NO₂ VCDs in Europe were largest in urban areas, with the largest 2024 mean VCD occurring in Moscow, Russia (15.5×10^{15}
233 molecules cm⁻²) (Fig. 5a). Broad enhanced 2024 annual mean VCDs exceeding 4×10^{15} molecules cm⁻² were observed in a
234 region encompassing Belgium, the Netherlands and western portions of Germany, with values exceeding 5×10^{15} molecules
235 cm⁻² in the Po River Valley of northern Italy.

236 From 2019 to 2024, decreases in NO₂ VCD occurred in 61% of all urban clusters in Europe. All cities with a population greater
237 than 1,000,000 experienced decreases, with the exception of Moscow (+29%) and other cities of western Russia, which
238 experienced increases (Fig. 5b). The broad decreases across large Europe cities are likely due to a combination of (1) continued
239 decreased emissions trends that accelerated during the COVID-19 pandemic, (2) continued transition to alternative energy
240 sources following the start of the Russia-Ukraine war in 2022 and (3) existing policies implemented within the EU (Matthias
241 et al., 2021; Rokicki et al., 2023; Cifuentes-Faura, 2022). These policies include the European Green Deal and European
242 Climate Law, which promote zero-emission vehicles, stricter vehicle emissions targets and updated industrial emissions
243 regulations.

244



245

246 **Figure 5: Same as Fig. 2, but centered on Europe.**

247 3.4 North America and South America

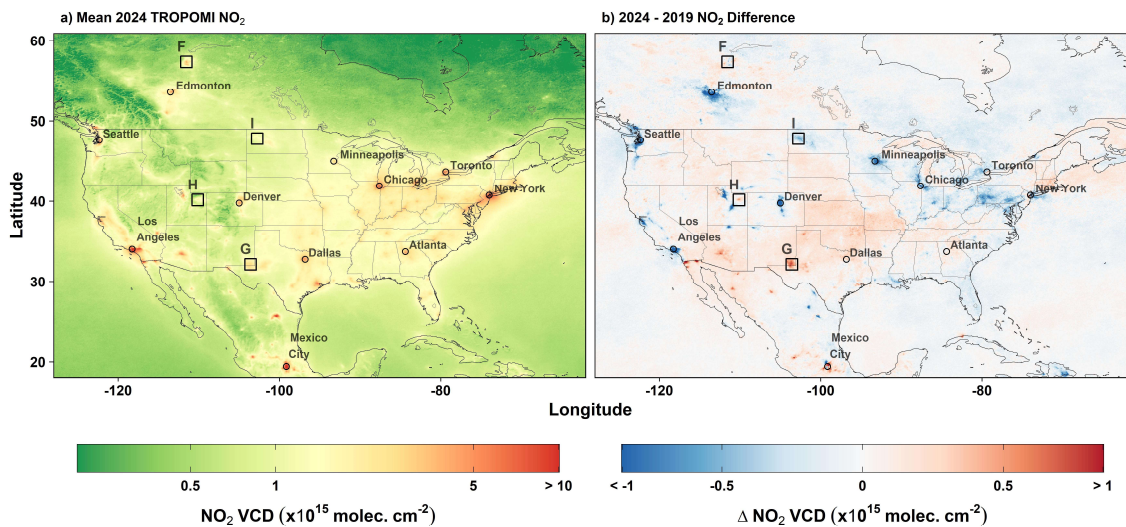
248 Throughout North America, 2024 annual mean NO₂ VCDs were largest in urban regions, including Los Angeles (7.4×10^{15}
 249 molecules cm⁻²), New York (7.0×10^{15} molecules cm⁻²), Chicago (5.0×10^{15} molecules cm⁻²), Mexico City (11.3×10^{15}
 250 molecules cm⁻²) and Toronto (4.3×10^{15} molecules cm⁻²), as well as near fossil fuel-fired power plant and mining operations
 251 (Fig. 6a). A majority of cities in the U.S. and Canada exhibited decreased or unchanged NO₂ VCDs (Fig. 6b), with notable
 252 exceptions being Phoenix, Arizona (+10%) and Dallas, Texas (+6%), which experienced increases (Fig. S6).

253 In Canada, the largest VCD decreases were observed in Alberta Province in and around Edmonton (-19%). In the U.S., aside
 254 from decreases in urban environments, the largest changes were observed in remote areas near power plants, e.g. near the
 255 decommissioned Navajo Generating Station in northern Arizona, the Four Corners Generating Station in northern New
 256 Mexico, and the Hunter and Huntington Power Plants in central Utah (Goldberg et al., 2021). Oil, gas, and coal mining
 257 operations influenced regional VCD changes as well, with annual mean NO₂ VCDs decreasing from 2019 to 2024 in the
 258 Athabasca oil sands (+1%) in Northern Alberta (Fig. 3f), increases in the Permian (+29%) and Uintah (+35%) Basins in the
 259 southwestern U.S. (Fig. 3g-h), and decreases in the Bakken (-16%) in North Dakota (Fig. 3i). Apparent within the U.S. is a
 260 slight increase in background concentrations in rural regions, particularly in the Central and Western U.S. It is unclear if this
 261 is due to an extension of the NO₂ lifetime due to decreasing VOCs and O₃ over this 6-year period (e.g., Laughner & Cohen
 262 2019) or due to increased NO_x emissions in rural areas or both. Further work should investigate this.



263 In Mexico, Central America and the Caribbean, the largest VCDs are observed near Mexico City (11.3×10^{15} molecules cm^{-2}) and Monterrey, Mexico (7.7×10^{15} molecules cm^{-2}), with numerous other urban signatures. The largest increases were
265 observed at sites in Northern Mexico, including Mexicali (+31%) and Hermosillo (+32%) as well as a handful of regions with
266 decreased VCDs in northern Mexico, including Monterrey (-9%). VCDs also decreased near the capital city of Santo Domingo,
267 Dominican Republic (-28%), and increased near Havana, Cuba (+39%).

268 In South America, the largest VCDs are observed near Lima, Peru (6.3×10^{15} molecules cm^{-2}); Santiago, Chile (9.7×10^{15}
269 molecules cm^{-2}); and Sao Paulo, Brazil (7.3×10^{15} molecules cm^{-2}) (Fig. S7). Regions near Santiago experienced some of the
270 largest decreases in VCD in South America (-19%), while Quito, Ecuador experienced the largest increase (+86%).



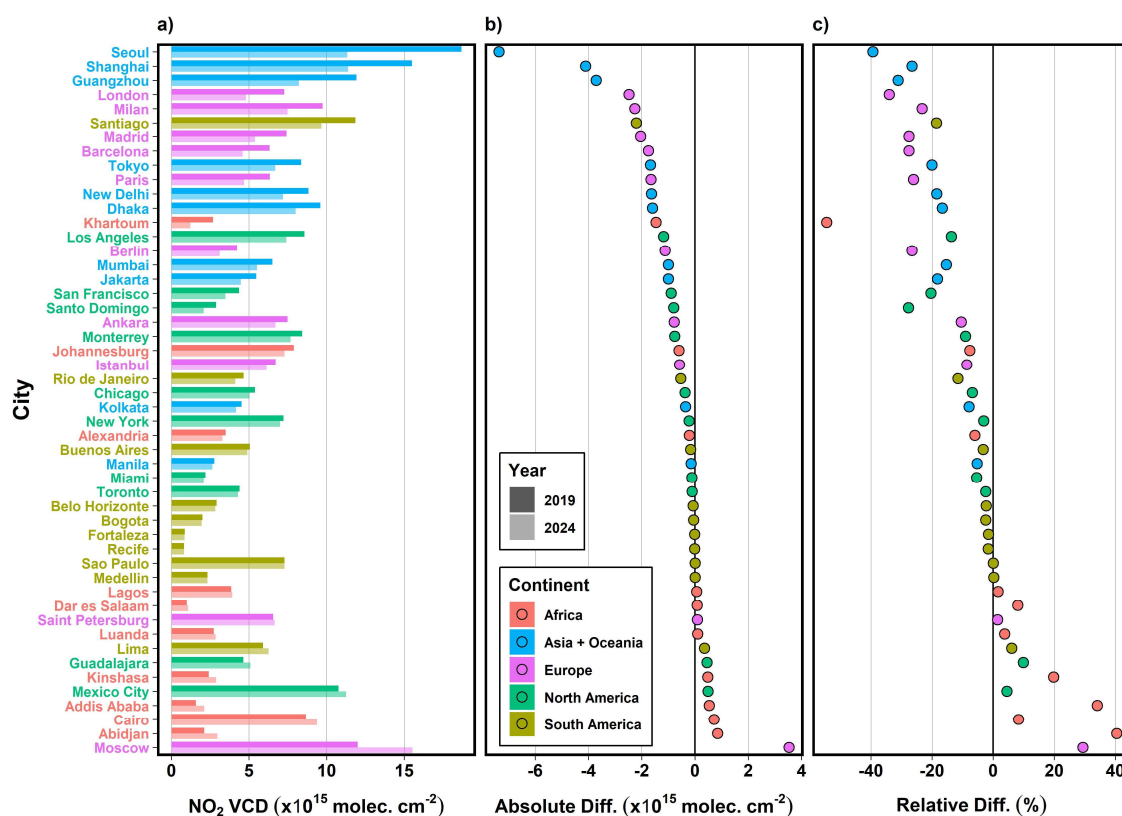
272
273 **Figure 6:** Same as Fig. 2, but for North America. Squares and numbers represent select oil and gas regions highlighted in Fig. 3.

274 4 Urban-level NO₂ VCD trends

275 Using the method outlined in Section 2.2.1, the GHS-SMOD urban cluster boundaries are used to determine mean TROPOMI
276 NO₂ concentrations for all urban clusters globally with a minimum population of 50,000. Looking at VCD changes from 2019
277 to 2024 in the 50 cities representing the ten most populous urban clusters on each continent, with Asia and Oceania considered
278 jointly, East Asian cities represent four and European cities represent five of the ten largest VCD decreases (Fig. 7a). Seoul
279 experienced the greatest reduction in NO₂ VCD of any of these 50 cities, with annual average levels from 2019 to 2024
280 decreasing by 7.4×10^{15} molecules cm^{-2} (Fig. 7b), or nearly -40% (Fig. 7c). London, England produced the greatest NO₂ VCD



281 decrease of the ten most populous European cities, with a mean decrease of 2.5×10^{15} molecules cm^{-2} (Fig. 7b), or -34%. This
 282 decrease occurred alongside the introduction of the city's ultra-low emission zone introduced in 2019 and expanded in 2023,
 283 which has been shown to decrease local NO_2 concentrations (Hajmohammadi and Heydecker, 2022).
 284 Large South American cities generally experienced minimal changes in NO_2 VCD, with relative changes typically less than
 285 $\pm 5\%$ (Fig. 7c). The most notable exception is Santiago, Chile, which experienced a mean VCD decrease of nearly 20% from
 286 2019 to 2024. The largest North American cities mostly experienced moderate VCD decreases, with the largest absolute
 287 decreases occurring in Los Angeles (-13.7%) and the San Francisco Bay Area (-20.4%), and largest increases occurring in the
 288 Mexican cities of Guadalajara (+9.9%) and Mexico City (+4.6%). Chicago and New York City, two of the largest cities in the
 289 U.S., also experienced decreases, though less pronounced.



290

291 **Figure 7: (a) NO_2 VCD in 2019 (dark bars) and 2024 (light bars) for the 10 most populous urban clusters on each continent, based**
 292 **on GHS-SMOD populations. (b) Absolute difference in NO_2 VCD for each city from 2019 to 2024. (c) Relative percent change in**
 293 **VCD from 2019 to 2024. Colors correspond to the respective continent for each city. Cities are ordered by magnitude of absolute**
 294 **VCD decrease.**



295 Most of the largest African cities experienced increased NO₂ VCDs from 2019 to 2024, with Abidjan, Ivory Coast experiencing
296 the largest urban increase of 0.85×10^{15} molecules cm⁻² (Fig. 7b), or an increase of 40.5% (Fig. 7c). Additional notable African
297 increases are Cairo, Egypt (+8.3%), Addis Ababa, Ethiopia (+34.1%) and Kinshasa, DR Congo (+19.9%). The largest decrease
298 on the African continent was observed in the Sudanese capital of Khartoum, which experienced an average decrease of $1.46 \times$
299 10^{15} molecules cm⁻² (Fig. 7b) or a decrease of 54.5% (Fig. 7c). These strong VCD decreases in Khartoum coincide with conflict
300 in the country, causing large portions of the city to be displaced, impacting NO₂ concentrations (see Sec. 3.2).

301 Of the cities presented in Fig. 7, the three largest absolute decreases between 2019 and 2024 were in the East Asian cities of
302 Seoul, South Korea (Fig. 8a), Shanghai, China (Fig. 8b) and Guangzhou, China (Fig. 8c). Seoul experienced decreases greater
303 than 7×10^{15} molecules cm⁻² from 2019 to 2024, largely due to effective policies implemented by the South Korean government
304 since the early 2000s to reduce local emissions, as well as trends in emissions that began following the COVID-19 pandemic
305 (Ho et al., 2021; Seo et al. 2021). The observed annual decreases in these East Asian cities were primarily driven by decreases
306 during the winter months (Fig. 7d). European cities also experienced some of the largest decreases in VCD, with the three
307 largest decreases occurring in London, UK (-34%); Milan, Italy (-23%); and Madrid, Spain (-28%) (Fig. S8). Three cities with
308 notable increases include Moscow, Russia (+29%), Baghdad, Iraq (+17%) and Riyadh, Saudi Arabia (+13%) (Fig. S9).
309 Moscow experienced the largest NO₂ VCD increase of any GHS-SMOD city through 2024, with a mean increase of 3.5×10^{15}
310 molecules cm⁻² (Fig. 7b). The increasing trend in Moscow accelerated in early 2022 (Fig. S9), following the onset of the
311 Russia-Ukraine war in Ukraine, when monthly mean NO₂ VCDs for March reached 59×10^{15} molecules cm⁻² (see Sec. 3.3).

312

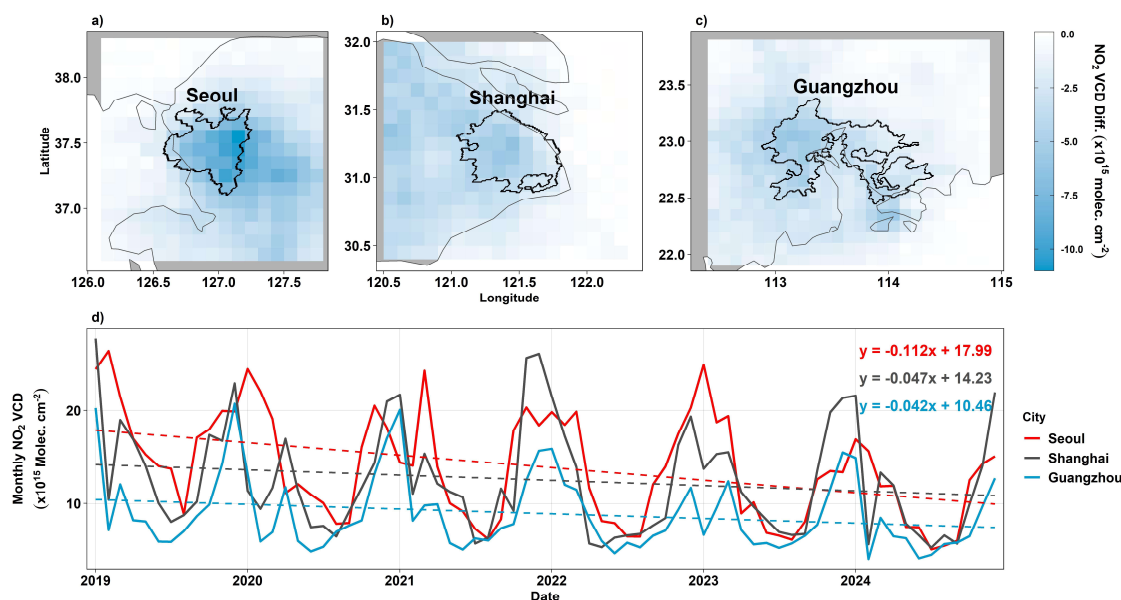


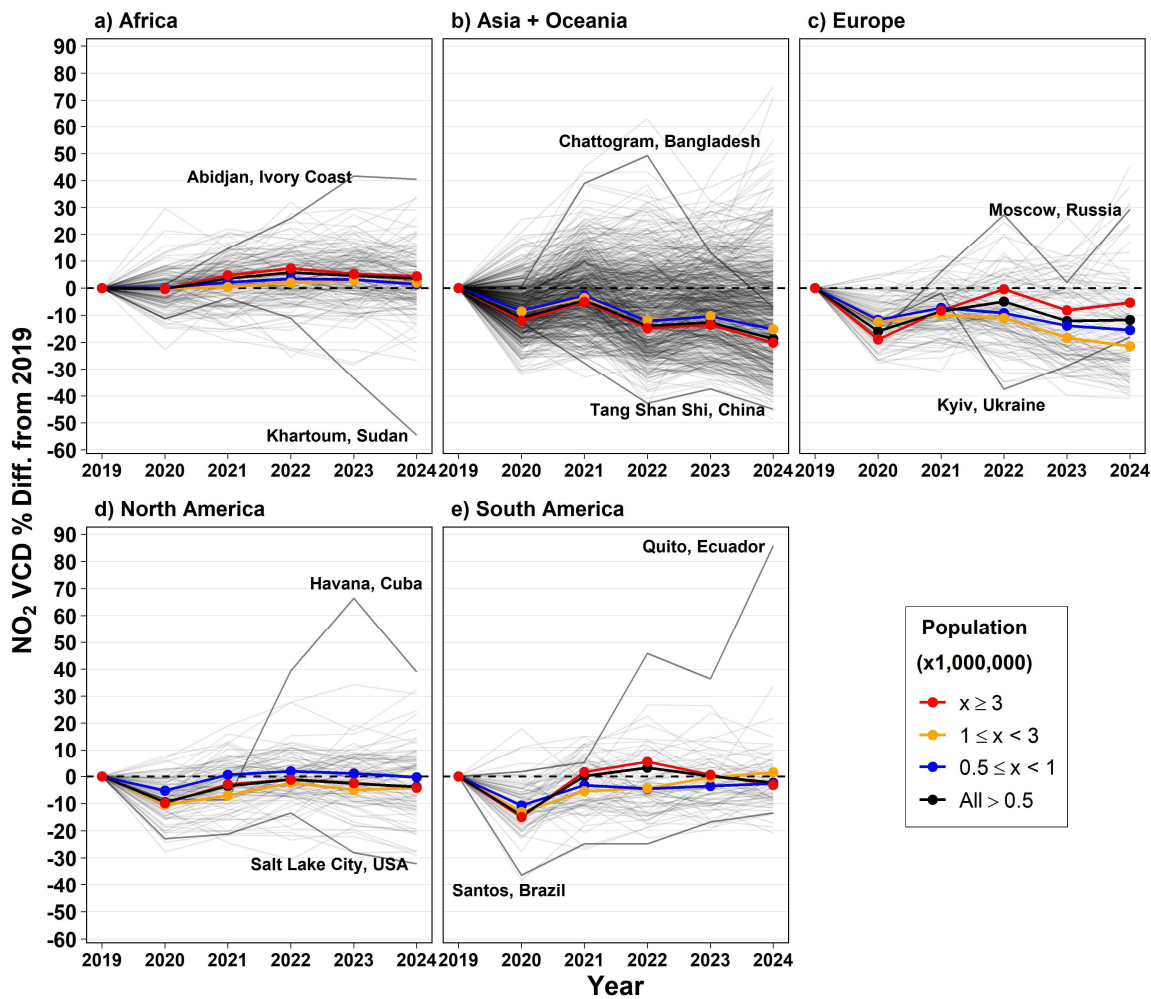
Figure 8: Absolute change in mean annual NO₂ VCD from 2019 to 2024 for three East Asian cities: (a) Seoul, South Korea, (b) Shanghai, China and (c) Guangzhou, China. Colors in panels a–c show magnitude of VCD change, thin lines show national borders or coastlines, and thick lines show the GHS-SMOD urban boundary. (d) Solid lines show monthly mean TROPOMI NO₂ VCD from 01/2019 through 12/2024, colored by city. Dashed lines and equations show ordinary least-squares regression trends, with the slope representing the change in NO₂ VCD per month, and the y-intercept representing the intercept for January, 2019.

5 Aggregated trends in urban TROPOMI NO₂

TROPOMI NO₂ changes from 2019 to 2024 for larger urban clusters (i.e. clusters with a population greater than 500,000) are shown in Fig. 9. This represents approximately 1000 of the most populated urban clusters, or just over 9% of all urban clusters in the GHS-SMOD dataset. 15.2% of these cities are in Africa, 57.8% are in Asia and Oceania, 10.9% are in Europe, 9.1% are in North America and 6.9% are in South America. On average, annual mean NO₂ VCDs are characterized by a decrease of about 10% from 2019 to 2020; previous work has attributed such decreases to the COVID-19 pandemic (Cooper et al., 2022). Africa was the exception to these 2020 decreases (Fig. 9a), which saw average VCDs largely unchanged for that year. Through 2024, African cities experienced a gradual increase in VCD, with larger cities exhibiting a larger fractional increase in NO₂ (Fig. 9a). The largest percent increase occurred in Abidjan, the capital city of Ivory Coast, which experienced an increase in NO₂ VCD of more than 40% from 2019 through 2024. Khartoum, Sudan experienced the largest percent decrease of any large African City, with mean VCDs decreasing by nearly 60% through 2024, with most of that decrease accelerating in the Spring of 2023 (Fig. S5). This decrease coincides with the onset of conflict within Sudan, which heavily impacted the capital city of Khartoum, leading to the displacement of much of the Khartoum population outside the city (Guo et al., 2023).



332



333

334 **Figure 9:** Percent change in annual mean TROPOMI tropospheric NO₂ vertical column densities (VCD) for individual GHS-SMOD
335 urban clusters with a population of at least 500,000 (gray lines), relative to 2019 values. Population-weighted percent change is shown
336 for urban clusters with a population between 500,000 and 1 million (blue), between one and three million (yellow), greater than three
337 million (red) and all clusters with a population greater than 500,000 (black). Results are separated by continent for (a) Africa, (b)
338 Asia and Oceania, (c) Europe, (d) North America and (e) South America.

339 Asian cities, representing a majority of all urban clusters globally, experienced an average population-weighted NO₂ VCD
340 decrease of approximately 19% from 2019 to 2024 (Fig. 9b). Urban clusters with a population greater than 3 million
341 experienced the largest decreases, with mean VCDs in those cities decreasing by 20% through 2024, with the majority of these



decreases occurring in Chinese cities. One notable decrease in Asia occurred in the Chinese city of Tangshan Shi, located to the east of the Chinese capital of Beijing, which experienced an NO₂ VCD decrease of nearly 45% from 2019 to 2024. The largest increase in Asia through 2024 occurred in the Mongolian capital of Ulaanbaatar, where VCDs have increased by more than 70%. Numerous Bangladeshi cities, including Chattogram, experienced substantially increased VCDs from 2020 through 2022, with VCDs decreasing again by 2024 to the near 2019 levels (Fig. S10). Tehran, Iran by far has the largest annual average VCD in the TROPOMI tropospheric NO₂ record for all GHS-SMOD cities, with annual mean values remaining above 30×10^{15} molecules cm⁻² throughout the entirety of the TROPOMI record (Fig. S11).

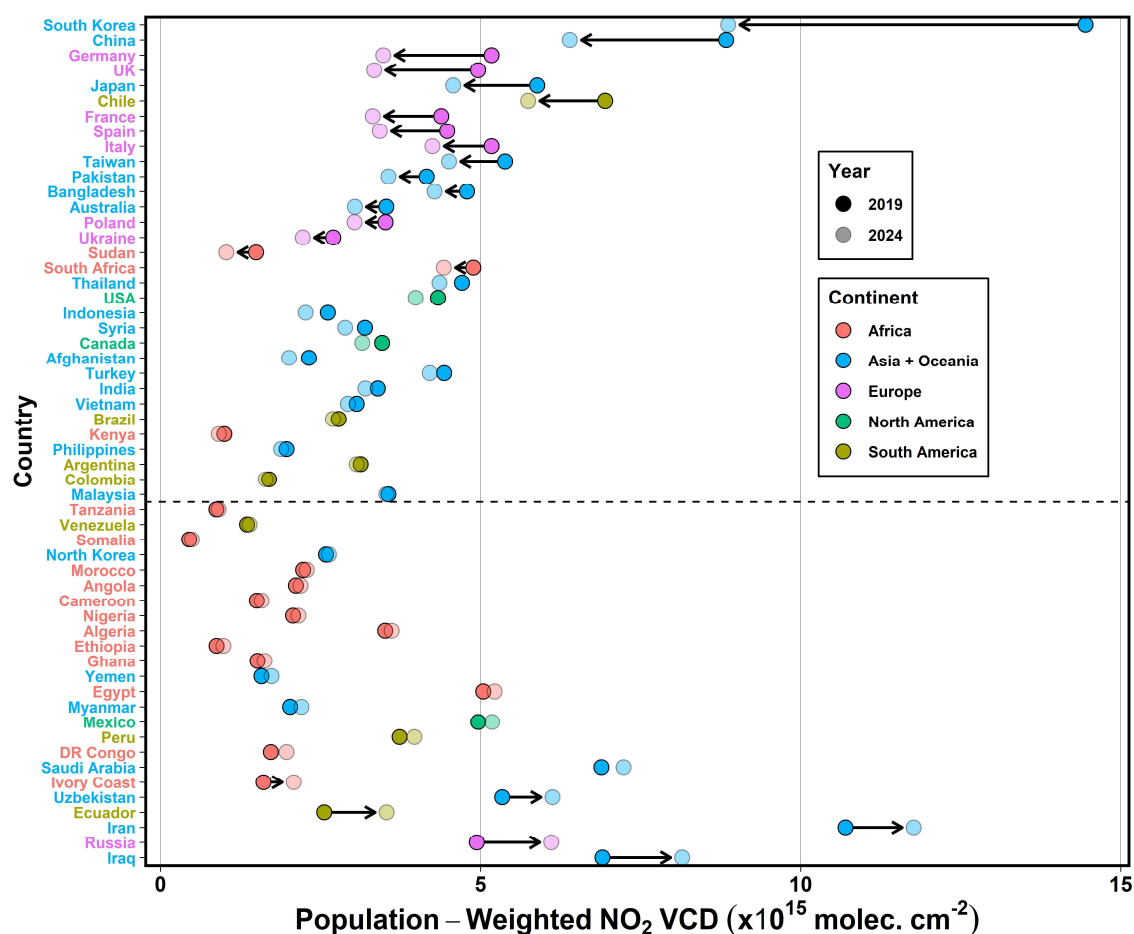
The impact of the COVID-19 pandemic on NO₂ VCDs is particularly stark in Asia, due to the multiple waves of COVID-19 related lockdowns and closures in China, leading to reduced NO₂ levels. Initial lockdowns in 2020 led to widespread VCD decreases in China, which were followed by a rebound in levels in 2021 (Fig. S12). A resurgence of the virus in 2022 led to multiple further lockdowns throughout the year, some lasting for months (Zheng et al., 2021; Zhao et al., 2024), that ultimately resulted in reduced VCDs. Chinese cities continued to experience decreased NO₂ VCDs in 2023 and 2024, in large part due to effective emissions reduction policies (Li et al., 2024).

Column NO₂ in European cities experienced the most pronounced decrease in column NO₂ of any continent in 2020, with larger cities with a population greater than three million experiencing a nearly 20% reduction in population-weighted VCD (Fig. 9c). NO₂ VCDs rebounded marginally in 2021 and 2022, followed by decreases into 2023 and 2024, although decreases are more pronounced when only analyzing cities in the 27 member countries of the European Union, as of 2024 (Fig. S13). One notable feature within the European annual average VCDs are the contrasting VCD trends in Russian and Ukrainian cities in 2022, at the onset of the Russia-Ukraine War (Fig. S14). In the Ukrainian capital of Kyiv, annual VCDs dropped nearly 40% in 2022 relative to 2019, coinciding with a large portion of the city fleeing due to conflict in and near the city. To contrast this, VCDs increased nearly 30% in the Russian capital of Moscow during the same period. Following 2022, VCDs in Kyiv increased steadily, while in Moscow, concentrations decreased in 2023 then increased sharply again in 2024.

In North America, most cities experienced a decrease in annual NO₂ VCD of less than 10% in 2020, with concentrations generally rebounding to 2019 levels by 2024 (Fig. 9d). Havana, Cuba was a notable exception of North American cities, with VCDs increasing by nearly 70% through 2023 relative to 2019, with a slight decrease in 2024. Cities in the western U.S., such as Salt Lake City and Denver experienced some of the largest percent decreases on the continent, decreasing by approximately 30% through 2024. In South America, most cities experienced a 10% VCD decrease in 2020, with mean concentrations rebounding to 2019 values by 2021 and remaining around those levels through 2024 (Fig. 9e). One notable exception is Quito, Ecuador, which experienced a VCD increase of over 85% through 2024. Santos, Brazil, an active port town southeast of São Paulo, experienced one of the largest VCD decreases in South America, with a 35% decrease in VCDs from 2019 to 2020, followed by sustained, gradual annual increases through 2024.



373 Aggregating the NO₂ VCD changes to the country level and taking into account the population of each urban cluster (Eq. 1),
374 the urban and population-weighted NO₂ VCD decreases for countries with an urban population of at least nine million were
375 largest in the East Asia, including South Korea, China and Japan, as well as countries of Western and Central Europe (Fig.
376 10). Urban population-weighted VCD decreases in South Korea were particularly pronounced, with concentrations decreasing
377 by 5.6×10^{15} molecules cm⁻² from 2019 to 2024. Germany experienced the largest VCD decrease in Europe through 2024,
378 with a decrease of 1.7×10^{15} molecules cm⁻². Chile saw the largest decrease in urban NO₂ VCD of any South American country,
379 due in large part to VCD decreases in the capital city of Santiago.



380

381 Figure 10: Urban population-weighted NO₂ VCD changes from 2019 to 2024 for the 56 countries with an urban population of at
382 least nine million, based on urban cluster populations provided from GHS-SMOD. Countries are ordered by the magnitude of VCD
383 decrease and colored by continent. Darker points represent 2019 VCDs and lighter points represent 2024 VCDs. Arrows indicate



the direction of VCD change from 2019 to 2024. The dashed line separates countries that experienced population-weighted VCD decreases (above line) and increases (below line).

Of countries with increased VCDs and an urban population greater than nine million, about half were in Africa. The majority of increases in these African countries were relatively minor, with country-level urban VCDs typically increasing by less than 0.25×10^{15} molecules cm^{-2} through 2024 (Fig. 10). Middle Eastern and Central Asian countries experienced some of the largest urban VCD increases, with Iraq experiencing the largest increase of any larger country ($+1.2 \times 10^{15}$ molecules cm^{-2}). Of the most-populous European countries, Russia was the only country to experience increased population-weighted NO_2 VCDs through 2024 ($+1.16 \times 10^{15}$ molecules cm^{-2}).

We further identify notable changes in countries with a GHS-SMOD urban population less than nine million and therefore excluded from Fig. 10. Less-populated countries in Africa saw either increasing or little-changed NO_2 VCDs from 2019 to 2024 (Fig. 11). In Asia, Mongolia experienced an increase of 2.05×10^{15} molecules cm^{-2} through 2024, or an increase of 75%, the largest population-weighted percent increase of any Asian country. We note that Mongolia has just three GHS-SMOD urban clusters, two of which are located in or near the capital city of Ulaanbaatar, where the bulk of the country-level increases were observed. In Sri Lanka, VCDs decreased by 0.5×10^{15} molecules cm^{-2} , or a decrease of 27%, one of the larger decreases for Asian countries. In Europe, both Belgium and the Netherlands experienced VCD decreases through 2024 that exceed 30%.

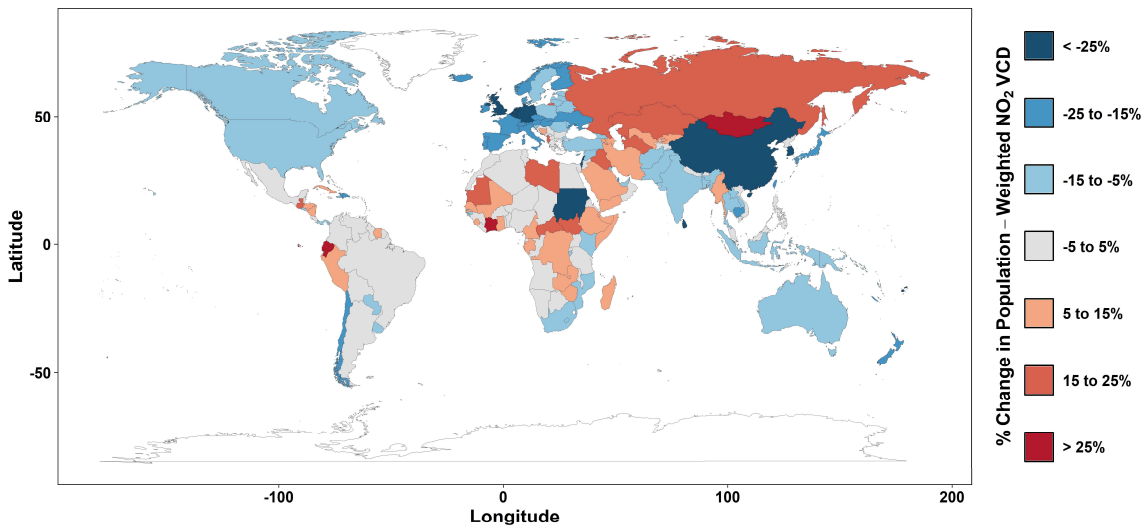


Figure 11: Spatial representation of the percent change in urban population-weighted NO_2 VCD for all countries globally from 2019 to 2024, binned by the magnitude of percent change. Increases greater than 5% are shown in reds, decreases less than 5% in blues, and countries with a change between -5% and 5% in gray.



405 **6 Influence of background NO₂ and seasonal variability on urban NO₂**

406 Urban NO₂ concentrations are not only influenced by local emissions, but also by advection of upwind pollutants into the urban
407 boundary. We account for the role that upwind background concentrations may play in urban NO₂ concentrations by identifying
408 changes in the urban enhancement of NO₂ (VCD_{ENH}), represented by the difference between NO₂ VCDs in the urban cluster
409 and the background. By removing the background concentrations, VCD_{ENH} more closely represents the portion of the urban
410 VCD that is primarily a result of local, urban emission sources.

411 In Africa, moderate-sized cities with a population between one and three million experienced the smallest relative increase in
412 VCD_{ENH} through 2024 (+4.9%), while smaller and larger cities experienced larger increases (Fig. 12a). Notably, African cities
413 on average did not experience decreased VCD_{ENH} in 2020 at the onset of the COVID-19 pandemic, a distinct feature observed
414 on all other continents. Regardless of population size, African cities experienced an average VCD_{ENH} increase of +6% through
415 2024. In Asia and Oceania, cities experienced sustained decreases in VCD_{ENH} regardless of the city population, with a mean
416 decrease of -22.7%, although larger cities experienced more pronounced decreases (Fig. 12b). In contrast, changes in VCD_{ENH}
417 in European cities largely depended on the population of the city, with smaller (-17.6%) and moderate-sized (-26.7%) cities
418 exhibiting the largest decreases, while larger cities (-5.3%) experienced lesser decreases on average (Fig. 12c). In North
419 America, smaller cities between 500,000 and 1 million saw a -8% decrease in VCD_{ENH} in 2020, but quickly rebounded in 2021
420 to near 2019 levels, which were sustained through 2024 (Fig. 12d). Moderate and large North American cities also rebounded
421 following the dip in 2020, however VCD_{ENH} remained approximately 7.5% below 2019 levels by 2024. In South America,
422 cities experienced a VCD_{ENH} decrease of 16% on average in 2020, with concentrations in cities of all populations rebounding
423 to near 2019 levels by 2024 (Fig. 12e).

424

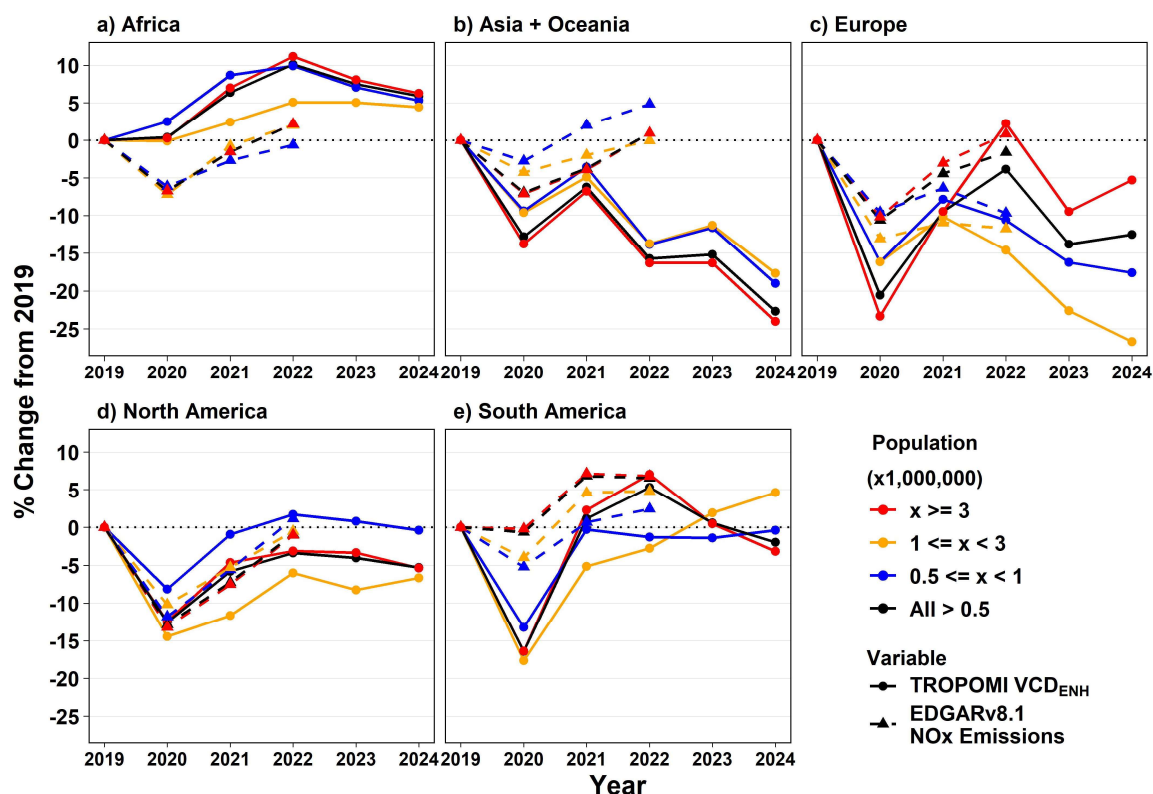


Figure 12: Percent change in population-weighted TROPOMI NO₂ VCD urban enhancements (solid lines; 2019 - 2024) and EDGARv8.1 NO_x emissions (dashed lines; 2019 - 2022) for GHS-SMOD urban clusters, relative to 2019 levels. Colors represent urban cluster population range, and results are separated by continent (a-e).

Assuming that the percent change in VCD_{ENH} relative to a baseline year can be attributed to changes in NO_x emissions within each urban cluster, we additionally evaluate changes in EDGAR NO_x emissions averaged for each continent, with emissions estimates available through 2022. In African cities, a mean difference of -7.7% was exhibited between the percent changes in EDGAR NO_x emissions and VCD_{ENH} relative to 2019, indicating a potential underestimate in EDGAR NO_x emissions for this period (Fig. 12a). Cities in Asia and Oceania experienced VCD_{ENH} that tracked relatively well with EDGAR NO_x emissions from 2019 to 2021, with a mean difference of +4.2% between emissions and VCD_{ENH}. However, emissions showed further increases in 2022, while VCD_{ENH} exhibited a sharp decrease for that year. This resulted in a percent difference of +16.7% between emissions and VCD_{ENH} in 2022 relative to 2019 levels (Fig. 12b). The 2022 VCD_{ENH} decrease coincided with large lockdowns in China related to the COVID-19 pandemic, suggesting that EDGAR emissions may not reflect emissions



439 decreases during that lockdown period. In Europe and North America, EDGAR NO_x emissions and VCD_{ENH} exhibited a
440 similar change relative to 2019 levels through 2022, with a mean difference of +5.7% and +0.2%, respectively, suggesting
441 more accurate EDGAR NO_x emissions for cities on those continents (Fig. 12c,d). In South America, the mean percent change
442 relative to 2019 was +7.6% higher for EDGAR NO_x emissions than VCD_{ENH} (Fig. 12e); however, EDGAR emissions do
443 correlate with changes in VCD_{ENH}, e.g. a flat or slower increase from 2021 to 2022 for both VCD_{ENH} and EDGAR. The better
444 agreement in Europe and North America than other continents could be due to a higher availability of observational constraints
445 on emissions, leading to more accurate changes in emissions from year to year.

446 To identify the impact that different seasons may have on annual trends, we evaluate changes in urban population-weighted
447 NO₂ VCDs for May – September and November – March. In African cities (Fig. 13a), mean VCDs increased by 0.1×10^{15}
448 molecules cm⁻² during November – March through 2024, with little to no change occurring on average during May –
449 September. In Asian (Fig. 13b) and North American cities (Fig. 13d), the bulk of the observed annual decreases through 2024
450 occurred during the winter months, with average winter decreases of -1.8×10^{15} molecules cm⁻² and -0.5×10^{15} molecules cm⁻², respectively. Despite the generally larger absolute changes during winter months in Asia and Oceania, the relative percent
451 changes for the summer and winter months exhibited more similar behavior (Fig. S15). In European cities, population-weighted
452 VCDs decreased by -0.4×10^{15} molecules cm⁻² (-10%) through 2024 during the summer months, while winter month changes
453 remained negligible, despite a sharp increase in winter-time levels in 2022 during the onset of the Russia-Ukraine war (Fig.
454 13c). Seasonal changes impacted South American cities less than cities on other continents through 2024 (Fig. 13e), with mean
455 winter and summer VCDs both changing by less than 0.3×10^{15} molecules cm⁻² through 2024.

457

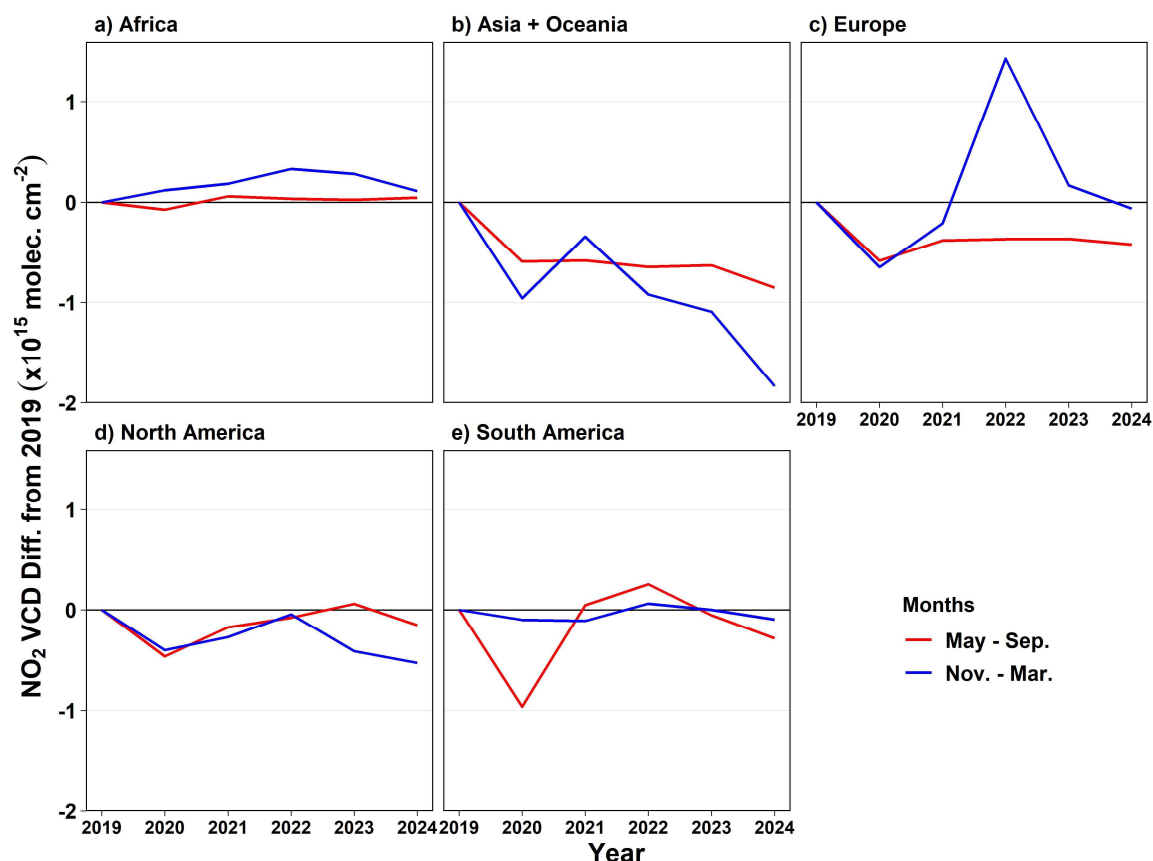


Figure 13: Change in population-weighted urban NO₂ VCDs averaged for GHS-SMOD urban clusters from 2019 to 2024 for May – September (red lines) and November – March (blue lines) in (a) Africa, (b) Asia and Oceania, (c) Europe, (d) North America and (e) South America.

7 Conclusions

We present a global analysis of urban TROPOMI tropospheric NO₂ trends from 2019 to 2024 using GHS-SMOD-defined urban boundaries, encompassing more than 11,500 cities. Our results reveal widespread decreases in NO₂ across cities in Asia and Oceania (-17% on average), Europe (-17%), and North America (-5%), with particularly strong reductions in cities including Seoul (-40%), Guangzhou (-30%), and London, England (-34%). These decreases generally reflect a combination of long-term emissions control policies and economic incentives, indicating policies to tackle NO₂ pollution have broadly worked. COVID-19 induced reductions in activity often caused a temporary NO₂ reduction but is unlikely to have caused much of the long-term changes between 2019 and 2024. Conversely, urban NO₂ in Africa has gradually increased over the same period,



470 with Abidjan (+40%) and Addis Ababa (+35%) leading the continent's upward trend. With the exception of May-September
471 in 2020, South America exhibited little mean VCD change from 2019 to 2024, with being Santiago (-19%) being a notable
472 exception. Population-weighted NO₂ VCDs increased in countries in the Middle East and much of Africa, highlighting a
473 potential degradation in air quality in regions of the world that lack extensive ground-level monitoring.

474 Evaluating annual changes in TROPOMI NO₂ urban enhancements (VCD_{ENH})—the difference between mean urban and
475 background VCDs—against changes in EDGAR NO_x emissions, we show that changes in VCD_{ENH} scales best with changes
476 in EDGAR NO_x in European and North American cities, with mean percent differences of +5.7% and +0.2% relative to 2019
477 levels, respectively, and scale worse in other parts of the globe, revealing potential discrepancies in emissions inventories. This
478 mismatch is particularly evident in African (-7.7%) and Asian (+8.3%) cities, and may stem from rapidly evolving emission
479 sources or limitations in the EDGAR bottom-up inventory methods. Similar discrepancies in emissions inventories in the
480 Global South have been reported in previous studies (Ahn et al., 2023), suggesting a systematic emissions underestimation in
481 regions where unmonitored emissions activity may be significant.

482 In most regions, VCD trends from 2019 to 2024 were driven by changes during the colder months (November – March). This
483 was most pronounced in Asian cities, where mean cold season VCDs decreased by -1.2×10^{15} molecules cm⁻² (-18%) on from
484 2019 to 2024, compared with warm season VCD decreases of -0.5×10^{15} molecules cm⁻² (-13%). Large changes in NO₂ were
485 not confined to urban regions alone. We identified localized increases near fossil fuel and other mining operations, including
486 in the Santanghu Basin in China (+172%), the Permian (+19%) and Uintah (+35%) Basins in the U.S., and the Copperbelt
487 region of the DRC (+64%), signaling expanding industrial activity. In Khartoum and Kyiv, conflict and displacement drove
488 sharp reductions in NO₂, demonstrating the utility of satellite data in detecting societal disruptions.

489 Several limitations of this work should be noted. First, satellite NO₂ column densities may not always reflect surface-level NO₂
490 concentrations, particularly in regions with vertically elevated sources. In urban areas dominated by surface-based
491 transportation emissions, NO₂ VCDs are likely more representative of surface exposure. However, in areas with tall-stack
492 sources, such as power plants, NO₂ columns may be decoupled from near-surface levels. Second, we assume static city
493 boundaries defined by the 2023 version of GHS-SMOD, with population estimates from 2020. This is likely a reasonable
494 approximation for urbanized regions in Europe and North America, where built-up area changes are slow, but may introduce
495 uncertainty in rapidly urbanizing regions of Africa and Asia over a six-year period. Future analyses could incorporate time-
496 varying urban boundaries to address this.

497 Taken together, these results demonstrate the utility of high-resolution satellite instruments for characterizing both broad
498 regional trends and localized pollution changes, and linking with anthropogenically induced factors such as urban growth,
499 industrial expansion, policy interventions, and conflict. This highlights potential in using TROPOMI observations as an
500 accountability agent to determine how local changes in human activities affect local and global air pollution. As the TROPOMI
501 record lengthens and newer, geostationary satellites come online and begin to detect changes in atmospheric composition,



502 continued space-based monitoring will be essential for improving our understanding of atmospheric composition and chemistry
503 around the globe.

504 **Data Availability.**

505 The level 3 annual and monthly average TROPOMI NO₂ VCDs are available at 10.5067/ACADNS5UBWPQ and
506 <https://doi.org/10.5067/KKPPL39PEIGE>, respectively. The GHS-SMOD urban boundaries can be downloaded from
507 <https://human-settlement.emergency.copernicus.eu/download.php?ds=smod>. The EDGARDv8.1 NO_x emissions can be
508 downloaded from https://edgar.jrc.ec.europa.eu/dataset_ap81.

509 **Supplement.**

510 The supplement contains additional figures related to the study, including: S1 All GHS-SMOD urban clusters. S2 Data
511 disaggregation example. S3 Satellite view of surface mines. S4 Spatial plot of African NO₂. S5 Khartoum NO₂ time series. S6
512 NO₂ increases in three U.S. cities. S7 Spatial plot of South American NO₂. S8 Annual mean NO₂ VCDs for Bangladeshi cities.
513 S9 Annual mean NO₂ in Iran. S10 Annual mean VCDs in Chinese cities. S11 Annual mean NO₂ changes in the European
514 Union. S12 Annual mean NO₂ changes in Russian and Ukrainian cities. S13 Seasonal NO₂ changes by continent.

515 **Author Contribution.**

516 D.H. and D.G. contributed to the project design. D.G. processed and provided the annually- and monthly-averaged NO₂ vertical
517 column densities. All authors edited the manuscript.

518 **Competing Interests.**

519 The authors declare that they have no conflict of interest.

520 **Acknowledgements.**

521 This work was supported by National Aeronautics and Space Administration (NASA) Health and Air Quality Applied Sciences
522 Team (HAQAST) grant #80NSSC21K0511 and NASA Aura Atmospheric Composition Modeling and Analysis Program
523 (ACMAP) grant #80NSSC23K1002.

524



525 References

- 526 Ahn, D. Y., Goldberg, D. L., Coombes, T., Kleiman, G., and Anenberg, S. C.: CO₂ emissions from C40 cities: citywide
527 emission inventories and comparisons with global gridded emission datasets, *Environ. Res. Lett.*, 18, 034032,
528 <https://doi.org/10.1088/1748-9326/ACBB91>, 2023.
- 529 Anenberg, S. C., Mohegh, A., Goldberg, D. L., Kerr, G. H., Brauer, M., Burkart, K., Hystad, P., Larkin, A., Wozniak, S., and
530 Lamsal, L.: Long-term trends in urban NO₂ concentrations and associated paediatric asthma incidence: estimates from
531 global datasets, *Lancet Planet. Heal.*, 6, e49–e58, [https://doi.org/10.1016/S2542-5196\(21\)00255-](https://doi.org/10.1016/S2542-5196(21)00255-2/ATTACHMENT/7B9BB37F-5DEA-41D7-A9CC-9AFFA3989D6A/MMC1.PDF)
532 [2/ATTACHMENT/7B9BB37F-5DEA-41D7-A9CC-9AFFA3989D6A/MMC1.PDF](https://doi.org/10.1016/S2542-5196(21)00255-2/ATTACHMENT/7B9BB37F-5DEA-41D7-A9CC-9AFFA3989D6A/MMC1.PDF), 2022.
- 533 Bovensmann, H., Burrows, J. P., Buchwitz, M., Frerick, J., Noël, S., Rozanov, V. V., Chance, K. V., and Goede, A. P. H.:
534 SCIAMACHY: Mission Objectives and Measurement Modes, *J. Atmos. Sci.*, 56, 127–150, [https://doi.org/10.1175/1520-](https://doi.org/10.1175/1520-0469(1999)056<0127:SMOAMM>2.0.CO;2)
535 [0469\(1999\)056<0127:SMOAMM>2.0.CO;2](https://doi.org/10.1175/1520-0469(1999)056<0127:SMOAMM>2.0.CO;2), 1999.
- 536 Burrows, J. P., Weber, M., Buchwitz, M., Rozanov, V., Ladstätter-Weissenmayer, A., Richter, A., Debeek, R., Hoogen, R.,
537 Bramstedt, K., Eichmann, K. U., Eisinger, M., and Perner, D.: The Global Ozone Monitoring Experiment (GOME):
538 Mission Concept and First Scientific Results, *J. Atmos. Sci.*, 56, 151–175, [https://doi.org/10.1175/1520-](https://doi.org/10.1175/1520-0469(1999)056<0151:TGOMEG>2.0.CO;2)
539 [0469\(1999\)056<0151:TGOMEG>2.0.CO;2](https://doi.org/10.1175/1520-0469(1999)056<0151:TGOMEG>2.0.CO;2), 1999.
- 540 Chen, X., Qi, L., Li, S., and Duan, X.: Long-term NO₂ exposure and mortality: A comprehensive meta-analysis, *Environ.*
541 *Pollut.*, 341, 122971, <https://doi.org/10.1016/J.ENVPOL.2023.122971>, 2024.
- 542 Cifuentes-Faura, J.: European Union policies and their role in combating climate change over the years, *Air Qual. Atmos.*
543 *Heal.*, 15, 1333–1340, <https://doi.org/10.1007/S11869-022-01156-5/FIGURES/1>, 2022.
- 544 Cooper, M. J., Martin, R. V., Hammer, M. S., Levelt, P. F., Veefkind, P., Lamsal, L. N., Krotkov, N. A., Brook, J. R., and
545 McLinden, C. A.: Global fine-scale changes in ambient NO₂ during COVID-19 lockdowns, *Nat.* 2022 6017893, 601,
546 380–387, <https://doi.org/10.1038/s41586-021-04229-0>, 2022.
- 547 Crippa, M., Guizzardi, D., Pagani, F., Schiavina, M., Melchiorri, M., Pisoni, E., Graziosi, F., Muntean, M., Maes, J., Dijkstra,
548 L., Damme, M. Van, Clarisse, L., and Coheur, P.: Insights into the spatial distribution of global, national, and subnational
549 greenhouse gas emissions in the Emissions Database for Global Atmospheric Research (EDGAR v8.0), *Earth Syst. Sci.*
550 *Data*, 16, 2811–2830, <https://doi.org/10.5194/essd-16-2811-2024>, 2024.
- 551 Cui, R. Y., Hultman, N., Cui, D., McJeon, H., Yu, S., Edwards, M. R., Sen, A., Song, K., Bowman, C., Clarke, L., Kang, J.,
552 Lou, J., Yang, F., Yuan, J., Zhang, W., and Zhu, M.: A plant-by-plant strategy for high-ambition coal power phaseout in
553 China, *Nat. Commun.* 2021 121, 12, 1–10, <https://doi.org/10.1038/s41467-021-21786-0>, 2021.



- 554 Dang, R., Jacob, D. J., Shah, V., Eastham, S. D., Fritz, T. M., Mickley, L. J., Liu, T., Wang, Y., and Wang, J.: Background
555 nitrogen dioxide (NO₂) over the United States and its implications for satellite observations and trends: effects of nitrate
556 photolysis, aircraft, and open fires, *Atmos. Chem. Phys.*, 23, 6271–6284, <https://doi.org/10.5194/ACP-23-6271-2023>,
557 2023.
- 558 Dix, B., de Bruin, J., Roosenbrand, E., Vlemmix, T., Francoeur, C., Gorchov-Negron, A., McDonald, B., Zhizhin, M., Elvidge,
559 C., Veefkind, P., Levelt, P., and de Gouw, J.: Nitrogen Oxide Emissions from U.S. Oil and Gas Production: Recent Trends
560 and Source Attribution, *Geophys. Res. Lett.*, 47, e2019GL085866, <https://doi.org/10.1029/2019GL085866>, 2020.
- 561 Duncan, B. N., Lamsal, L. N., Thompson, A. M., Yoshida, Y., Lu, Z., Streets, D. G., Hurwitz, M. M., and Pickering, K. E.: A
562 space-based, high-resolution view of notable changes in urban NO_x pollution around the world (2005–2014), *J. Geophys.*
563 *Res.*, 121, 976–996, <https://doi.org/10.1002/2015JD024121>, 2016.
- 564 Faure, K., Willis, J. P., and Dreyer, J. C.: The grootegeeluk formation in the Waterberg Coalfield, South Africa: facies,
565 palaeoenvironment and thermal history — evidence from organic and clastic matter, *Int. J. Coal Geol.*, 29, 147–186,
566 [https://doi.org/10.1016/0166-5162\(95\)00029-1](https://doi.org/10.1016/0166-5162(95)00029-1), 1996.
- 567 Fioletov, V., McLinden, C. A., Griffin, D., Krotkov, N., Liu, F., and Eskes, H.: Quantifying urban, industrial, and background
568 changes in NO₂ during the COVID-19 lockdown period based on TROPOMI satellite observations, *Atmos. Chem. Phys.*,
569 22, 4201–4236, <https://doi.org/10.5194/acp-22-4201-2022>, 2022.
- 570 Fioletov, V., McLinden, C. A., Griffin, D., Zhao, X., and Eskes, H.: Global seasonal urban, industrial, and background NO₂
571 estimated from TROPOMI satellite observations, *Atmos. Chem. Phys.*, 25, 575–596, [https://doi.org/10.5194/acp-25-575-](https://doi.org/10.5194/acp-25-575-2025)
572 2025, 2025.
- 573 Fisher, B. L., Lamsal, L. N., Fasnacht, Z., Oman, L. D., Joiner, J., Krotkov, N. A., Choi, S., Qin, W., and Yang, E. S.: Revised
574 estimates of NO₂ reductions during the COVID-19 lockdowns using updated TROPOMI NO₂ retrievals and model
575 simulations, *Atmos. Environ.*, 326, 120459, <https://doi.org/10.1016/J.ATMOENV.2024.120459>, 2024.
- 576 Fishman, J., Iraci, L. T., Al-Saadi, J., Chance, K., Chavez, F., Chin, M., Coble, P., Davis, C., DiGiacomo, P. M., Edwards, D.,
577 Eldering, A., Goes, J., Herman, J., Hu, C., Jacob, D. J., Jordan, C., Kawa, S. R., Key, R., Liu, X., Lohrenz, S., Mannino,
578 A., Natraj, V., Neil, D., Neu, J., Newchurch, M., Pickering, K., Salisbury, J., Sosik, H., Subramaniam, A., Tzortziou, M.,
579 Wang, J., and Wang, M.: The United States’ Next Generation of Atmospheric Composition and Coastal Ecosystem
580 Measurements: NASA’s Geostationary Coastal and Air Pollution Events (GEO-CAPE) Mission, *Bull. Am. Meteorol.*
581 *Soc.*, 93, 1547–1566, <https://doi.org/10.1175/BAMS-D-11-00201.1>, 2012.
- 582 De Foy, B., Lu, Z., and Streets, D. G.: Satellite NO₂ retrievals suggest China has exceeded its NO_x reduction goals from the
583 twelfth Five-Year Plan, *Sci. Reports* 2016 61, 6, 1–9, <https://doi.org/10.1038/srep35912>, 2016.



- 584 Geddes, J. A., Martin, R. V., Boys, B. L., and van Donkelaar, A.: Long-term trends worldwide in ambient NO₂ concentrations
585 inferred from satellite observations, *Environ. Health Perspect.*, 124, 281–289,
586 https://doi.org/10.1289/EHP.1409567/SUPPL_FILE/EHP.1409567.S001.ACCO.PDF, 2016.
- 587 Van Geffen, J., Eskes, H., Compernelle, S., Pinardi, G., Verhoelst, T., Lambert, J.-C., Snee, M., Linden, M. Ter, Ludewig,
588 A., Folkert Boersma, K., and Pepijn Veefkind, J.: Sentinel-5P TROPOMI NO₂ retrieval: impact of version v2.2
589 improvements and comparisons with OMI and ground-based data, *Atmos. Meas. Tech.*, 15, 2037–2060,
590 <https://doi.org/10.5194/amt-15-2037-2022>, 2022.
- 591 Goldberg, D.: HAQAST Sentinel-5P TROPOMI Nitrogen Dioxide (NO₂) GLOBAL Monthly Level 3 0.1 x 0.1 Degree
592 Gridded Data Version 2.4 (HAQ_TROPOMI_NO2_GLOBAL_M_L3) at GES DISC,
593 <https://doi.org/10.5067/KKPPL39PEIGE>, 15 March 2024.
- 594 Goldberg, D. L., Anenberg, S. C., Kerr, G. H., Moheg, A., Lu, Z., and Streets, D. G.: TROPOMI NO₂ in the United States:
595 A Detailed Look at the Annual Averages, Weekly Cycles, Effects of Temperature, and Correlation With Surface NO₂
596 Concentrations, *Earth's Futur.*, 9, <https://doi.org/10.1029/2020EF001665>, 2021.
- 597 Goldberg, D.L.: HAQAST Sentinel-5P TROPOMI Nitrogen Dioxide (NO₂) GLOBAL Monthly Level 3 0.1 x 0.1 Degree
598 Gridded Data Version 2.4 (HAQ_TROPOMI_NO2_GLOBAL_M_L3) at GES DISC, Goddard Earth Sciences Data and
599 Information Services Center (GES DISC), <https://doi.org/10.5067/KKPPL39PEIGE>, 2024.
- 600 Goldberg, D. L., Tao, M., Kerr, G. H., Ma, S., Tong, D. Q., Fiore, A. M., Dickens, A. F., Adelman, Z. E., and Anenberg, S.
601 C.: Evaluating the spatial patterns of U.S. urban NO_x emissions using TROPOMI NO₂, *Remote Sens. Environ.*, 300,
602 <https://doi.org/10.1016/j.rse.2023.113917>, 2024.
- 603 de Gouw, J. A., Veefkind, J. P., Roosenbrand, E., Dix, B., Lin, J. C., Landgraf, J., and Levelt, P. F.: Daily Satellite Observations
604 of Methane from Oil and Gas Production Regions in the United States, *Sci. Rep.*, 10, 1–10,
605 [https://doi.org/10.1038/S41598-020-57678-](https://doi.org/10.1038/S41598-020-57678-4)
606 [4](https://doi.org/10.1038/S41598-020-57678-4);TECHMETA=134;SUBJMETA=106,169,172,35,704,824;KWRD=ATMOSPHERIC+CHEMISTRY, 2020.
- 607 Guo, Z., Abushama, H., Siddig, K., Kirui, O. K., Abay, K., and You, L.: Monitoring Indicators of Economic Activities in
608 Sudan Amidst Ongoing Conflict Using Satellite Data, *Def. Peace Econ.*,
609 <https://doi.org/10.1080/10242694.2023.2290474>;PAGE:STRING:ARTICLE/CHAPTER, 2023.
- 610 Hajmohammadi, H. and Heydecker, B.: Evaluation of air quality effects of the London ultra-low emission zone by state-space
611 modelling, *Atmos. Pollut. Res.*, 13, 101514, <https://doi.org/10.1016/J.APR.2022.101514>, 2022.
- 612 Harake, W., Jamali, I., and Abou Hamde, N.: Lebanon Economic Monitor : Lebanon Sinking (to the Top 3), Washington, D.C.,
613 <https://doi.org/http://documents.worldbank.org/curated/en/394741622469174252>, 2021.



- 614 Herman, J., Cede, A., Spinei, E., Mount, G., Tzortziou, M., and Abuhassan, N.: NO₂ column amounts from ground-based
615 Pandora and MFDOAS spectrometers using the direct-sun DOAS technique: Intercomparisons and application to OMI
616 validation, *J. Geophys. Res. Atmos.*, 114, <https://doi.org/10.1029/2009JD011848>, 2009.
- 617 Hersbach, H., Bell, B., Berrisford, P., Hirahara, S., Horányi, A., Muñoz-Sabater, J., Nicolas, J., Peubey, C., Radu, R., Schepers,
618 D., Simmons, A., Soci, C., Abdalla, S., Abellan, X., Balsamo, G., Bechtold, P., Biavati, G., Bidlot, J., Bonavita, M., De
619 Chiara, G., Dahlgren, P., Dee, D., Diamantakis, M., Dragani, R., Flemming, J., Forbes, R., Fuentes, M., Geer, A.,
620 Haimberger, L., Healy, S., Hogan, R. J., Hólm, E., Janisková, M., Keeley, S., Laloyaux, P., Lopez, P., Lupu, C., Radnoti,
621 G., de Rosnay, P., Rozum, I., Vamborg, F., Villaume, S., and Thépaut, J. N.: The ERA5 global reanalysis, *Q. J. R.
622 Meteorol. Soc.*, 146, 1999–2049, <https://doi.org/10.1002/QJ.3803>, 2020.
- 623 Ho, C. H., Heo, J. W., Chang, M., Choi, W., Kim, J., Kim, S. W., and Oh, H. R.: Regulatory measures significantly reduced
624 air-pollutant concentrations in Seoul, Korea, *Atmos. Pollut. Res.*, 12, 101098,
625 <https://doi.org/10.1016/J.APR.2021.101098>, 2021.
- 626 Huber, D. E., Kort, E. A., and Steiner, A. L.: Soil Moisture, Soil NO_x and Regional Air Quality in the Agricultural Central
627 United States, *J. Geophys. Res. Atmos.*, 129, e2024JD041015, <https://doi.org/10.1029/2024JD041015>, 2024.
- 628 Jiang, Z., McDonald, B. C., Worden, H., Worden, J. R., Miyazaki, K., Qu, Z., Henze, D. K., Jones, D. B. A., Arellano, A. F.,
629 Fischer, E. V., Zhu, L., and Folkert Boersma, K.: Unexpected slowdown of US pollutant emission reduction in the past
630 decade, *Proc. Natl. Acad. Sci. U. S. A.*, 115, 5099–5104, [https://doi.org/10.1073/PNAS.1801191115/-](https://doi.org/10.1073/PNAS.1801191115/-/DCSUPPLEMENTAL)
631 [/DCSUPPLEMENTAL](https://doi.org/10.1073/PNAS.1801191115/-/DCSUPPLEMENTAL), 2018.
- 632 Jiang, Z., Zhu, R., Miyazaki, K., McDonald, B. C., Klimont, Z., Zheng, B., Boersma, K. F., Zhang, Q., Worden, H., Worden,
633 J. R., Henze, D. K., Jones, D. B. A., Denier van der Gon, H. A. C., and Eskes, H.: Decadal Variabilities in Tropospheric
634 Nitrogen Oxides Over United States, Europe, and China, *J. Geophys. Res. Atmos.*, 127, e2021JD035872,
635 <https://doi.org/10.1029/2021JD035872>, 2022.
- 636 Jin, X., Zhu, Q., Cohen, R. C., and Xiaomeng, J.: Direct estimates of biomass burning NO_x emissions and lifetimes using
637 daily observations from TROPOMI, *Atmos. Chem. Phys.*, 21, 15569–15587, <https://doi.org/10.5194/acp-21-15569-2021>,
638 2021.
- 639 Kim, E. J., Holloway, T., Kokandakar, A., Harkey, M., Elkins, S., Goldberg, D. L., and Heck, C.: A Comparison of Regression
640 Methods for Inferring Near-Surface NO₂ With Satellite Data, *J. Geophys. Res. Atmos.*, 129, e2024JD040906,
641 <https://doi.org/10.1029/2024JD040906>, 2024.
- 642 Krotkov, N. A., McLinden, C. A., Li, C., Lamsal, L. N., Celarier, E. A., Marchenko, S. V., Swartz, W. H., Bucsela, E. J., Joiner,
643 J., Duncan, B. N., Folkert Boersma, K., Pepijn Veefkind, J., Levelt, P. F., Fioletov, V. E., Dickerson, R. R., He, H., Lu,



- 644 Z., and Streets, D. G.: Aura OMI observations of regional SO₂ and NO₂ pollution changes from 2005 to 2015, *Atmos.*
645 *Chem. Phys.*, 16, 4605–4629, <https://doi.org/10.5194/acp-16-4605-2016>, 2016.
- 646 Labzovskii, L. D., Belikov, D. A., and Damiani, A.: Spaceborne NO₂ observations are sensitive to coal mining and processing
647 in the largest coal basin of Russia, *Sci. Rep.*, 12, 1–11, [https://doi.org/10.1038/S41598-022-16850-](https://doi.org/10.1038/S41598-022-16850-8)
648 [8](https://doi.org/10.1038/S41598-022-16850-8);SUBJMETA=106,172,35,4081,704;KWRD=ATMOSPHERIC+SCIENCE,ENVIRONMENTAL+IMPACT,ENVIRO
649 NMENTAL+SCIENCES, 2022.
- 650 Lamsal, L. N., Duncan, B. N., Yoshida, Y., Krotkov, N. A., Pickering, K. E., Streets, D. G., and Lu, Z.: U.S. NO₂ trends
651 (2005–2013): EPA Air Quality System (AQS) data versus improved observations from the Ozone Monitoring Instrument
652 (OMI), *Atmos. Environ.*, 110, 130–143, <https://doi.org/10.1016/J.ATMOSENV.2015.03.055>, 2015.
- 653 Laughner, J. L. and Cohen, R. C.: Direct observation of changing NO_x lifetime in North American cities, *Science* (80-.), 366,
654 723–727, <https://doi.org/10.1126/science.aax6832>, 2019.
- 655 Levelt, P. F., Van Den Oord, G. H. J., Dobber, M. R., Mälikki, A., Visser, H., De Vries, J., Stammes, P., Lundell, J. O. V., and
656 Saari, H.: The ozone monitoring instrument, *IEEE Trans. Geosci. Remote Sens.*, 44, 1093–1100,
657 <https://doi.org/10.1109/TGRS.2006.872333>, 2006.
- 658 Li, H., Zheng, B., Lei, Y., Hauglustaine, D., Chen, C., Lin, X., Zhang, Y., Zhang, Q., and He, K.: Trends and drivers of
659 anthropogenic NO_x emissions in China since 2020, *Environ. Sci. Ecotechnology*, 21, 100425,
660 <https://doi.org/10.1016/J.ESE.2024.100425>, 2024.
- 661 Liu, B., Fu, X., Bechtel, A., Gross, D., Sachsenhofer, R. F., and Li, X.: Middle Permian environmental changes and shale oil
662 potential evidenced by high-resolution organic petrology, geochemistry and mineral composition of the sediments in the
663 Santanghu Basin, Northwest China, *Int. J. Coal Geol.*, 185, 119–137, <https://doi.org/10.1016/J.COAL.2017.11.015>, 2018.
- 664 Liu, F., Zhang, Q., Van Der A, R. J., Zheng, B., Tong, D., Yan, L., Zheng, Y., and He, K.: Recent reduction in NO_x emissions
665 over China: synthesis of satellite observations and emission inventories, *Environ. Res. Lett.*, 11, 114002,
666 <https://doi.org/10.1088/1748-9326/11/11/114002>, 2016.
- 667 Lonsdale, C. R. and Sun, K.: Nitrogen oxides emissions from selected cities in North America, Europe, and East Asia observed
668 by the TROPOspheric Monitoring Instrument (TROPOMI) before and after the COVID-19 pandemic, *Atmos. Chem.*
669 *Phys.*, 23, 8727–8748, <https://doi.org/10.5194/acp-23-8727-2023>, 2023.
- 670 Ma, Q., Wang, J., Xiong, M., and Zhu, L.: Air Quality Index (AQI) Did Not Improve during the COVID-19 Lockdown in
671 Shanghai, China, in 2022, Based on Ground and TROPOMI Observations, *Remote Sens.*, 15, 1295,
672 <https://doi.org/10.3390/RS15051295/S1>, 2023.



- 673 Martínez-Alonso, S., Veeffkind, J. P., Dix, B., Gaubert, B., Theys, N., Granier, C., Soulié, A., Darras, S., Eskes, H., Tang, W.,
674 Worden, H., de Gouw, J., and Levelt, P. F.: S-5P/TROPOMI-Derived NO_x Emissions From Copper/Cobalt Mining and
675 Other Industrial Activities in the Copperbelt (Democratic Republic of Congo and Zambia), *Geophys. Res. Lett.*, 50,
676 e2023GL104109, <https://doi.org/10.1029/2023GL104109>, 2023.
- 677 Matthias, V., Quante, M., Arndt, J. A., Badeke, R., Fink, L., Petrik, R., Feldner, J., Schwarzkopf, D., Link, E. M., Ramacher,
678 M. O. P., and Wedemann, R.: The role of emission reductions and the meteorological situation for air quality
679 improvements during the COVID-19 lockdown period in central Europe, *Atmos. Chem. Phys.*, 21, 13931–13971,
680 <https://doi.org/10.5194/ACP-21-13931-2021>, 2021.
- 681 Miyazaki, K., Eskes, H., Sudo, K., Folkert Boersma, K., Bowman, K., and Kanaya, Y.: Decadal changes in global surface NO_x
682 emissions from multi-constituent satellite data assimilation, *Atmos. Chem. Phys.*, 17, 807–837,
683 <https://doi.org/10.5194/ACP-17-807-2017>, 2017.
- 684 Panda, M. R., Tyagi, A., Dhanya, C. T., Verma, A., and Swain, A.: Vulnerability assessment of thermal power plants in India
685 under water stress conditions, *Energy*, 276, 127553, <https://doi.org/10.1016/J.ENERGY.2023.127553>, 2023.
- 686 Richter, A., Burrows, J. P., Nüß, H., Granier, C., and Niemeier, U.: Increase in tropospheric nitrogen dioxide over China
687 observed from space, *Nature*, 437, 129–132, <https://doi.org/10.1038/nature04092>, 2005.
- 688 Rokicki, T., Bórawski, P., and Szeberényi, A.: The Impact of the 2020–2022 Crises on EU Countries’ Independence from
689 Energy Imports, Particularly from Russia, *Energies* 2023, Vol. 16, Page 6629, 16, 6629,
690 <https://doi.org/10.3390/EN16186629>, 2023.
- 691 Schiavina M., Melchiorri M. & Pesaresi M.: GHS-SMOD R2023A - GHS settlement layers, application of the Degree of
692 Urbanisation methodology (stage I) to GHS-POP R2023A and GHS-BUILT-S R2023A, multitemporal (1975-2030).
693 European Commission, Joint Research Centre (JRC). [https://doi.org/10.2905/A0DF7A6F-49DE-46EA-9BDE-](https://doi.org/10.2905/A0DF7A6F-49DE-46EA-9BDE-563437A6E2BA)
694 [563437A6E2BA](https://doi.org/10.2905/A0DF7A6F-49DE-46EA-9BDE-563437A6E2BA), 2023.
- 695 Schumann, U. and Huntrieser, H.: The global lightning-induced nitrogen oxides source, *Atmos. Chem. Phys.*, 7, 3823–3907,
696 2007.
- 697 SEDAC: The Gridded Population of the World (GPW), <https://sedac.ciesin.columbia.edu/data/collection/gpw-v4> (last access:
698 25 June 2024), 2017.
- 699 Seo, S., Kim, S. W., Kim, K. M., Lamsal, L. N., and Jin, H.: Reductions in NO₂ concentrations in Seoul, South Korea detected
700 from space and ground-based monitors prior to and during the COVID-19 pandemic, *Environ. Res. Commun.*, 3, 051005,
701 <https://doi.org/10.1088/2515-7620/ABED92>, 2021.



- 702 Shi, Q., Zheng, B., Zheng, Y., Tong, D., Liu, Y., Ma, H., Hong, C., Geng, G., Guan, D., He, K., and Zhang, Q.: Co-benefits
703 of CO₂ emission reduction from China's clean air actions between 2013-2020, *Nat. Commun.* 2022 131, 13, 1–8,
704 <https://doi.org/10.1038/s41467-022-32656-8>, 2022.
- 705 Shikwambana, L., Mhangara, P., and Mbatha, N.: Trend analysis and first time observations of sulphur dioxide and nitrogen
706 dioxide in South Africa using TROPOMI/Sentinel-5 P data, *Int. J. Appl. Earth Obs. Geoinf.*, 91, 102130,
707 <https://doi.org/10.1016/J.JAG.2020.102130>, 2020.
- 708 Song, W., Liu, X. Y., Hu, C. C., Chen, G. Y., Liu, X. J., Walters, W. W., Michalski, G., and Liu, C. Q.: Important contributions
709 of non-fossil fuel nitrogen oxides emissions, *Nat. Commun.* 2021 121, 12, 1–7, [https://doi.org/10.1038/s41467-020-](https://doi.org/10.1038/s41467-020-20356-0)
710 20356-0, 2021.
- 711 Stavrakou, T., Müller, J.-F., Boersma, K. F., De Smedt, I., and van der A, R. J.: Assessing the distribution and growth rates of
712 NO_x emission sources by inverting a 10-year record of NO₂ satellite columns, *Geophys. Res. Lett.*, 35,
713 <https://doi.org/10.1029/2008GL033521>, 2008.
- 714 VanDerA, R. J., Eskes, H. J., Boersma, K. F., van Noije, T. P. C., Van Roozendael, M., De Smedt, I., Peters, D. H. M. U., and
715 Meijer, E. W.: Trends, seasonal variability and dominant NO_x source derived from a ten year record of NO₂ measured
716 from space, *J. Geophys. Res. Atmos.*, 113, 1–12, <https://doi.org/10.1029/2007JD009021>, 2008.
- 717 Varma, A. K., Biswal, S., Hazra, B., Mendhe, V. A., Misra, S., Samad, S. K., Singh, B. D., Dayal, A. M., and Mani, D.:
718 Petrographic characteristics and methane sorption dynamics of coal and shaly-coal samples from Ib Valley Basin, Odisha,
719 India, *Int. J. Coal Geol.*, 141–142, 51–62, <https://doi.org/10.1016/J.COAL.2015.03.005>, 2015.
- 720 Veefkind, J. P., Aben, I., McMullan, K., Förster, H., de Vries, J., Otter, G., Claas, J., Eskes, H. J., de Haan, J. F., Kleipool, Q.,
721 van Weele, M., Hasekamp, O., Hoogeveen, R., Landgraf, J., Snel, R., Tol, P., Ingmann, P., Voors, R., Kruizinga, B., Vink,
722 R., Visser, H., and Levelt, P. F.: TROPOMI on the ESA Sentinel-5 Precursor: A GMES mission for global observations
723 of the atmospheric composition for climate, air quality and ozone layer applications, *Remote Sens. Environ.*, 120, 70–83,
724 <https://doi.org/10.1016/J.RSE.2011.09.027>, 2012.
- 725 Zhang, Y., Tian, J., Feng, S., Yang, F., and Lu, X.: The occurrence modes and geologic origins of arsenic in coal from
726 Santanghu Coalfield, Xinjiang, *J. Geochemical Explor.*, 186, 225–234, <https://doi.org/10.1016/J.GEXPLO.2017.12.006>,
727 2018.
- 728 Zhao, X., Li, X. X., Xin, R., Zhang, Y., and Liu, C. H.: Impact of Lockdowns on Air Pollution: Case Studies of Two Periods
729 in 2022 in Guangzhou, China, *Atmos.* 2024, Vol. 15, Page 1144, 15, 1144, <https://doi.org/10.3390/ATMOS15091144>,
730 2024.



731 Zheng, B., Zhang, Q., Geng, G., Chen, C., Shi, Q., Cui, M., Lei, Y., and He, K.: Changes in China's anthropogenic emissions
732 and air quality during the COVID-19 pandemic in 2020, *Earth Syst. Sci. Data*, 13, 2895–2907,
733 <https://doi.org/10.5194/ESSD-13-2895-2021>, 2021.

734

735