



1 **Seasonal Interplay of Water Mass Mixing and Nutrient Dynamics**
2 **in an Arctic Fjord: A Case Study of Kongsfjorden, Svalbard**

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19 **Abstract**

20 This study examined seasonal variations in water mass structure and nutrient dynamics in
21 Kongsfjorden, a high Arctic fjord where water mass composition varies seasonally due to
22 mixing among Atlantic Water, Polar Surface Water, and glacial meltwater. In spring, the
23 dominance of Modified Atlantic Water (MAW) facilitated active vertical mixing, leading to
24 relatively high, uniform nutrient concentrations throughout the water column. In summer, the
25 enhanced influence of glacial meltwater and warmer Polar Surface Waters (PSWw) resulted in
26 strong surface stratification and significant nutrient depletion in the upper layer. To disentangle
27 the effects of physical mixing from biological consumption, theoretical nutrient concentrations
28 were calculated based on a four-component water mass mixing model. The positive differences
29 between theoretical and observed concentrations ($\Delta\text{Nutrient}$) were indicative of significant
30 biological uptake, which accounted for substantial nutrient reductions in observed surface
31 concentrations from spring to summer: approximately $69 \pm 18\%$ for nitrate, $74 \pm 15\%$ for
32 phosphate, and $47 \pm 18\%$ for silicate. Crucially, $\Delta\text{Nutrient}$ values served as a 'biogeochemical
33 memory,' reflecting the cumulative net biological consumption since the spring bloom rather
34 than just instantaneous phytoplankton biomass. These biological processes also altered nutrient
35 stoichiometry, causing the surface nitrogen-to-phosphorus (N/P) ratio to increase from 15.0 in
36 spring to 18.8 in summer, indicating a shift in nutrient limitation patterns. Consequently,
37 summer surface waters transitioned toward potential co-limitation, with concentrations of
38 phosphate ($\sim 0.13 \pm 0.07 \mu\text{M}$) and silicate ($\sim 1.66 \pm 0.39 \mu\text{M}$) approaching their respective
39 limitation thresholds. These findings highlight a clear seasonal transition from a physically
40 controlled, nutrient-replete spring to a biologically regulated, nutrient-limited summer. This
41 understanding is crucial for predicting how Arctic fjord ecosystems, and their primary



42 productivity, will respond to ongoing Atlantification and increased freshwater input under
43 climate change.

44 **1. Introduction**

45 The Arctic marine ecosystem, which is characterized by unique and dynamic environmental
46 conditions, is governed by the complex interaction of physical, chemical, and biological factors.
47 Within this system, nutrient availability, which is primarily controlled by ocean currents,
48 riverine discharge, and atmospheric deposition, plays a fundamental role in maintaining
49 biological productivity and ecological health (Duarte et al., 2012; Tovar-Sánchez et al., 2010).
50 These nutrients are particularly vital for primary production, which is the foundation of the
51 Arctic marine food web. Ocean currents, notably Atlantic and Pacific inflows, transport
52 essential nutrients into the Arctic Ocean, thus influencing regional primary productivity
53 (Carmack et al., 2006; Codispoti et al., 2013; Torres-Valdés et al., 2013). As a result, seasonal
54 fluctuations in sea ice, solar radiation, and water column stratification drive nutrient dynamics
55 and productivity cycles (Arrigo et al., 2017). In particular, during spring and summer, increased
56 sunlight and meltwater often promote stratification and phytoplankton blooms (Leu et al., 2015;
57 Tremblay et al., 2015).

58 Arctic fjords such as Kongsfjorden in Svalbard are useful areas for assessing nutrient cycling
59 processes due to the interactions between advected ocean currents (e.g., warm, saline Atlantic
60 Water) and local water masses (Cottier et al., 2005; Svendsen et al., 2002). The inflow of
61 nutrient-rich Atlantic Water can significantly affect the nutrient supply and productivity in
62 fjords, leading to complex spatiotemporal variability (Carmack et al., 2006). Understanding
63 these dynamics, especially before and after blooms, is therefore essential for predicting how



64 Arctic fjord ecosystems respond to environmental changes. This is because seasonal shifts in
65 nutrient availability and plankton community structure strongly influence the region's
66 fundamental biogeochemical processes (Tremblay et al., 2015; Vonnahme et al., 2022; Singh
67 et al., 2020). Water mass mixing significantly influences nutrient distribution in Arctic fjords
68 (Randelhoff et al., 2018; Hodal et al., 2012; Tamelander et al., 2013; Rysgaard et al., 1999).
69 While AW inflow can enhance productivity by supplying nutrients (Carmack et al., 2006;
70 Torres-Valdés et al., 2013), a quantitative understanding of how physical mixing and biological
71 processes separately contribute to seasonal nutrient depletion remains a key knowledge gap.
72 Disentangling these effects is critical for accurately assessing the biological drivers of
73 productivity.

74 The present study addresses these gaps by examining seasonal (spring/summer) variation in
75 water mass mixing and nutrient dynamics in Kongsfjorden. Specifically, we employ a nutrient
76 anomaly approach (Δ Nutrient) derived from a water mass mixing model to quantify the net
77 biological impact on the nutrient inventory. Furthermore, we intend to determine the impact of
78 these seasonal mixing patterns, notably the active vertical mixing characteristic of spring and
79 the enhanced stratification observed in summer on nutrient concentrations. A key aspect of our
80 research will be to explore whether differences between theoretical nutrient concentrations,
81 derived from mixing models, and actual observed nutrient levels can be effectively used to
82 discern the influence of biological processes. Specifically, this study tests the hypothesis that
83 the difference between theoretical (mixing-derived) and observed nutrient concentrations can
84 effectively quantify the cumulative influence of biological processes. By comparing these
85 observed and theoretical nutrient levels, this study will assess the relative influence of physical
86 mixing versus biological processes. Ultimately, this research aims to provide crucial baseline



87 data for understanding how Arctic marine ecosystems respond to climate change, particularly
88 in the context of warming-induced alterations to water masses and mixing dynamics within
89 sensitive fjord environments.

90 **2. Materials and Methods**

91 **2.1 Study Sites**

92 Kongsfjorden, an Arctic fjord situated on the west coast of Spitsbergen, Svalbard, was used
93 as the primary study site (Fig. 1). This fjord is approximately 20 km in length and varies in
94 width from 4 to 10 km, reaching a maximum depth of approximately 300 m near its mouth.
95 The hydrography in Kongsfjorden is characterized by significant freshwater input from several
96 tidewater glaciers, a process that is more intense during the summer melt season. Furthermore,
97 the fjord is influenced by the advection of relatively warm and saline AW transported via the
98 West Spitsbergen Current and by the presence of colder, fresher waters of Arctic origin.

99 **2.2. Sampling and Analytical Methods**

100 Seawater samples were collected from three discrete depths within vertical water columns
101 using a conductivity-temperature-depth (CTD) rosette system aboard the MS Teisten (April)
102 and the RV Helmer Hanssen (July) during 2023 in Svalbard. Sampling depths were adjusted
103 by season to capture key hydrographic features. In spring (April), samples were collected at 0
104 m (surface), 20 m (mid-depth), and 50 m (deep) to represent the well-mixed water column. In
105 summer (July), a more stratified sampling strategy was employed to resolve the sharp vertical
106 gradients caused by meltwater; samples were collected from 0–5 m (surface), 10–25 m (mid-



107 depth, capturing the thermocline), and 50–100 m (deep). Detailed station-specific depths are
108 provided in the caption of Figure 4. During sample collection, the salinity and temperature
109 were measured using sensors within the CTD system. Fluorescence was measured using a
110 CTD-attached fluorometer and is presented in fluorescence-derived chlorophyll-*a*
111 concentrations (mg/m³).

112 Dissolved inorganic nutrients (NO₂⁻, NO₃⁻, PO₄³⁻, and Si(OH)₄) were analyzed using a
113 nutrient autoanalyzer (New QuAatro39; SEAL Analytical, UK). For each nutrient, 50 mL of
114 seawater was filtered through 0.7 µm GF/F filters (25 mm, Whatman Inc., Florham Park, NJ,
115 USA). This filtration was conducted using acid-washed syringes, and the filtrate was collected
116 in polypropylene conical tubes, which were stored at –20°C until analysis. To ensure the
117 accuracy and precision of the nutrient analysis, certified reference materials for each nutrient
118 were run concurrently with the samples. According to the certified reference material (KANSO
119 Co., LTD), the analytical uncertainty was within 5% for nitrate (the sum of NO₂⁻ and NO₃⁻),
120 Si(OH)₄, and PO₄³⁻. Hereafter, the sum of nitrate (NO₃⁻) and nitrite (NO₂⁻) is referred to as NO_x,
121 PO₄³⁻ as phosphate, and Si(OH)₄ as silicate. This terminology is used to ensure accuracy as
122 nitrite concentrations, while minor, were not consistently negligible..

123 **2.3. Water Mass Analysis and Theoretical Nutrient Concentrations**

124 To assess the seasonal variability in the hydrographic structure of the fjord and its influence
125 on the distribution of nutrients, water mass analysis was conducted. The mixing ratios of the
126 different water masses present in Kongsfjorden were calculated using observed temperature
127 and salinity data for both spring and summer. This analytical approach was in accordance with



128 established methodologies detailed by Miller (1950) and Tomczak (1999), which require the
129 precise definition of characteristic end-member water types that contribute to the observed
130 water properties within the fjord. Nutrient concentrations for the end-members were adopted
131 from the comprehensive study of Duarte et al. (2021), which provides representative
132 background values for the water masses advected into the Svalbard region.

133 In the present study, four principal end-member water types were used in the mixing model
134 due to their characteristic presence and influence in the Arctic region and specifically in
135 Kongsfjorden: Atlantic Water (AW), Modified AW (MAW), Polar Surface Water (PSW), and
136 its warmer variant Polar Surface Water warm (PSWw). While glacial meltwater (GMW) is a
137 significant source of freshwater in summer, its direct influence was simplified and incorporated
138 into the characteristics of PSWw, which represents the warm, low-salinity surface layer. This
139 assumption is further addressed in the discussion regarding silicate dynamics. The selection of
140 these water types was consistent with previous hydrographic characterization of the region
141 (Nilsen et al., 2008; Rudels et al., 2000). AW, which is defined by its relatively high
142 temperature and salinity, originates from lower latitudes and is advected into the Arctic. MAW
143 represents AW that has undergone significant transformation through cooling, freshening, and
144 nutrient alteration following its entry and circulation within the Arctic system. PSW is
145 characterized by its cold temperatures and lower salinity, typically occupying the upper layers
146 of the water column and originating from Arctic surface processes. PSWw shares many of the
147 same general characteristics as PSW but is distinguished by notably warmer temperatures,
148 often reflecting the influence of seasonal surface heating and increased meltwater input,
149 particularly during the summer months.



150 The temperature–salinity (T-S) characteristics defining these end-members are detailed
151 in Table 1 and visually represented in Fig. 2a. These definitions were carefully established
152 based on a combination of established values from past research (e.g., Rudels et al., 2000) and
153 an examination of the observed distribution of T-S data collected during the present study. This
154 dual approach ensured that the defined end-members comprehensively and accurately covered
155 the full spectrum of water types observed in Kongsfjorden during the sampling periods.
156 Because the hydrographic properties of the deep-water masses in Kongsfjorden exhibited
157 minimal temporal variation between the spring and summer seasons, a single, consistent set of
158 T-S characteristics for each end-member was employed for water mass analysis in both the
159 spring and summer datasets, allowing for a direct comparison of seasonal shifts in their relative
160 contributions.

161 The fractional contribution of AW, MAW, PSW, and PSW_w (denoted as A, B, C, and D,
162 respectively, Table. 1) to any given water sample (P) collected within the fjord was calculated
163 using a standard four end-member mixing model (Fig. 2b). This model operates on the principle
164 of the conservative mixing of temperature and salinity (Miller, 1950). The output of this model
165 provides the fractional contributions (f_A , f_B , f_C , and f_D) of each end-member to the sampled
166 water under the fundamental constraint that the sum of these individual fractions equals unity
167 (i.e., $f_A + f_B + f_C + f_D = 1$, or 100%).

168 Theoretical nutrient concentrations ($Nutrient^*$) for each sample were calculated by
169 multiplying the fraction of each end-member water mass (defined in Table 1) by its end-
170 member nutrient concentration (Nut_x) and summing the contributions as follows:

$$171 \quad Nutrient^* = (f_A \times Nut_A) + (f_B \times Nut_B) + (f_C \times Nut_C) + (f_D \times Nut_D)$$



172 To assess the biological impact on nutrient concentrations, the difference ($\Delta\text{Nutrient}$)
173 between the theoretical and observed concentrations was calculated:

174
$$\Delta\text{Nutrient} = \text{Nutrient}^* - \text{Nutrient}_{\text{observed}}$$

175 A positive value indicated net nutrient removal beyond physical mixing, which was attributed
176 to the net biological effect, primarily biological consumption.

177 To evaluate the robustness of these calculations, we performed a sensitivity analysis to
178 quantify the uncertainty propagated from the end-member nutrient definitions. The end-
179 member concentrations for NO_x, phosphate, and silicate were varied by $\pm 10\%$, a range selected
180 as a conservative estimate of natural variability based on regional oceanographic studies (e.g.,
181 Torres-Valdés et al., 2013; Hopwood et al., 2020). The resulting range in the calculated
182 $\Delta\text{Nutrient}$ values was used to define the uncertainty of our model-derived results, which is
183 reported alongside our key quantitative findings. This assessment provides a measure of
184 confidence in our conclusions against potential variations in the end-member characteristics.

185 **2.4. Statistical analysis**

186 All statistical analyses were performed using SPSS ver. 19 (IBM Corp., Armonk, NY, USA).
187 Prior to hypothesis testing, the normality of the data was assessed using the Shapiro–Wilk test.
188 Depending on the results of the normality test, either independent samples t-tests (for normally
189 distributed variables) or Mann–Whitney U tests (for non-normally distributed variables) were
190 applied to compare differences between groups. A significance level of $p < 0.05$ was used for
191 all tests.



192 **3. Results and Discussion**

193 **3.1. Seasonal Variation in Hydrography and Observed Dissolved Inorganic Nutrient** 194 **Levels**

195 Kongsfjorden exhibited distinct seasonal hydrographic conditions during the study period
196 (Fig. 3). Water temperatures in the fjord ranged from a minimum of -0.86°C to a maximum of
197 6.88°C (Fig. 3a), and salinity ranged from a minimum of 28.05 to a maximum of 34.93 (Fig.
198 3b). The spring season was characterized by lower temperatures, with a mean temperature of
199 $0.16 \pm 0.56^{\circ}\text{C}$, and relatively high and uniform salinity, averaging 35.67 ± 0.28 . In contrast,
200 summer had significantly warmer waters (mean: $3.56 \pm 1.49^{\circ}\text{C}$) and markedly lower and more
201 variable salinity (mean: 33.03 ± 1.92). These hydrographic changes were primarily driven by
202 seasonal increases in solar radiation, sea ice meltwater, and glacial freshwater input, which
203 collectively enhanced the vertical stratification of the water column.

204 Consistent with these hydrographic shifts, the levels of dissolved inorganic nutrients also
205 exhibited strong seasonal patterns. The NO_x concentration varied from $0.67 \mu\text{M}$ to $10.41 \mu\text{M}$
206 (Fig. 3c). During spring, the mean surface nitrate level was $7.10 \pm 1.83 \mu\text{M}$. In summer,
207 however, mean surface nitrate concentrations decreased significantly to $2.20 \pm 1.15 \mu\text{M}$,
208 representing an approximate $69 \pm 18\%$ reduction from spring levels. While surface nitrate was
209 depleted, concentrations in deeper water remained high, resulting in a stronger vertical gradient
210 in summer compared to that in spring. This suggests that active vertical mixing replenished
211 surface nutrients in spring, whereas reduced mixing and significant biological uptake occurred
212 during the summer period.



213 Phosphate concentrations ranged from 0.07 μM to 0.70 μM (Fig. 3d). The spring surface
214 mean was $0.50 \pm 0.12 \mu\text{M}$, declining considerably to $0.13 \pm 0.07 \mu\text{M}$ during summer, a
215 reduction of approximately $74 \pm 15\%$. Notably, summer phosphate concentrations often fell
216 below the 0.20 μM threshold commonly regarded as limiting for phytoplankton growth in
217 Arctic waters (Tremblay et al., 2015). Thus, there was a strong likelihood of phosphate
218 limitation during this period, particularly given that phosphate declined at a greater rate than
219 nitrate from spring to summer.

220 Silicate concentrations ranged from 0.83 to 4.45 μM (Fig. 3e). The mean surface
221 concentration was $3.11 \pm 0.72 \mu\text{M}$ in spring, decreasing to $1.66 \pm 0.39 \mu\text{M}$ in summer,
222 representing a $47 \pm 18\%$ reduction. The summer surface silicate concentration approached the
223 2 μM threshold frequently cited as indicative of potential silicate limitation for diatom growth
224 (Egge & Aksnes, 1992). In some samples, the summer surface silicate concentrations were
225 higher than expected despite biological uptake, likely due to the influence of glacial meltwater
226 enriched in silicate via bedrock erosion (Hawkings et al., 2017).

227 Statistical analysis confirmed that the seasonal differences observed for all three nutrients
228 were significant ($p < 0.05$ for all comparisons). These observed nutrient patterns in
229 Kongsfjorden were largely consistent with findings from previous studies in the same location
230 (e.g., Leu et al., 2015). However, the background nutrient levels observed in this study were
231 generally higher than those reported for some other Arctic regions, such as Young Sound,
232 Greenland (Rysgaard et al., 1999), a difference attributable to the stronger and more direct
233 influence of nutrient-rich AW in the Svalbard region. Spatial differences were also apparent
234 within Kongsfjorden; in particular, stations with higher contributions from PSWw exhibited



235 more pronounced summer surface nutrient depletion, particularly for phosphate, which had
236 mean concentrations as low as $0.08 \pm 0.03 \mu\text{M}$. (This observation will be further discussed in
237 the context of nutrient limitation in Section 3.4). This likely reflects the influence of glacial
238 meltwater input and enhanced stratification associated with PSWw-dominated surface layers.

239 **3.2. Seasonal Characteristics of Water Masses and Theoretical Nutrient Concentrations**

240 The four-component end-member mixing model revealed distinct seasonal distributions of
241 water masses within Kongsfjorden (Fig. 4). Overall, MAW, with a mean contribution of $52 \pm$
242 29% , and AW ($20 \pm 16\%$) were the dominant water masses influencing the fjord. These water
243 masses are recognized as the primary sources of inorganic nutrients in this system. The
244 contributions of PSW ($14 \pm 13\%$) and PSWw ($14 \pm 14\%$) were lower on average, though their
245 influence varied considerably with season and depth.

246 During the spring season, the proportion of MAW was generally higher throughout the water
247 column than in summer, suggesting the active mixing of the inflowing AW and the resident
248 PSW. This mixing is facilitated by physical and chemical processes in the Arctic Ocean that
249 promote the formation of MAW (Rudels et al., 2004), resulting in a relatively uniform vertical
250 distribution of water masses from the surface to the deep layers of the fjord. In contrast, the
251 summer season was characterized by a marked shift in the water mass composition. The surface
252 layer (0-30 m) had a considerably higher proportion of PSWw ($33 \pm 25\%$) and PSW ($19 \pm$
253 16%), which was primarily associated with seasonal sea ice meltwater, surface warming, and
254 freshwater-induced stratification. However, the deep layer (>50 m) remained dominated by
255 AW ($36 \pm 3\%$) and MAW ($53 \pm 3\%$). This vertical stratification limited the vertical exchange
256 of water and nutrients between the surface and deep layers.



257 The observed water mass distribution patterns were broadly consistent with previous
258 descriptions of Kongsfjorden by Svendsen et al. (2002) and Cottier et al. (2005). However, the
259 proportion of MAW observed in this study was substantially higher than reported in some
260 earlier studies, which may reflect the ongoing process of Atlantification, which is the enhanced
261 penetration of Atlantic-origin waters into the Arctic Ocean (Polyakov et al., 2017), or be the
262 result of long-term changes in the Arctic hydrography and climate. This trend suggests that
263 future warming could further intensify the influence of warm, saline Atlantic-origin waters,
264 fundamentally altering the fjord's stratification and nutrient supply regimes. Additionally,
265 continued glacier melting driven by regional warming is expected to increase the volume of
266 PSWw, thus intensifying surface stratification in the future.

267 **3.3. Biological Impact on Nutrient Concentrations: Differences between Observed and** 268 **Theoretical Concentrations (Δ Nutrient)**

269 To assess the biological influence on nutrient dynamics, the observed nutrient concentrations
270 were compared to theoretical values derived from end-member mixing (Fig. 5). The difference
271 represents net nutrient removal that exceeds a level that can be explained by physical mixing
272 alone. A positive Δ Nutrient value indicates that observed concentrations are lower than
273 expected from conservative mixing, thus suggesting biological uptake or transformation. With
274 few exceptions, observed nutrient concentrations were significantly lower than theoretical
275 values ($p < 0.05$ for all three nutrients), resulting in consistently positive Δ Nutrient values. This
276 provides strong evidence for substantial nutrient removal in Kongsfjorden beyond what can be
277 accounted for by physical advection and mixing, with phytoplankton uptake the most likely
278 mechanism.



279 Seasonal and depth-related comparisons of Δ Nutrient values highlight the extent of this
280 biological influence. Δ NO_x increased from spring (mean surface: $3.13 \pm 1.64 \mu\text{M}$; mean deep:
281 $2.66 \pm 2.26 \mu\text{M}$) to summer (mean surface: $5.76 \pm 1.99 \mu\text{M}$; mean deep: $7.03 \pm 0.75 \mu\text{M}$). A
282 similar trend was observed for Δ Phosphate, rising from spring (mean surface: $0.25 \pm 0.10 \mu\text{M}$;
283 mean deep: $0.20 \pm 0.10 \mu\text{M}$) to summer (mean surface: $0.48 \pm 0.11 \mu\text{M}$; mean deep: $0.48 \pm$
284 $0.04 \mu\text{M}$). Δ Silicate also increased between seasons, from spring mean (surface: 1.65 ± 0.66
285 μM ; mean deep: $1.27 \pm 0.77 \mu\text{M}$) to summer (mean surface: $2.14 \pm 0.99 \mu\text{M}$; mean deep: 3.28
286 $\pm 0.25 \mu\text{M}$). These consistently larger summer Δ Nutrient values strongly indicate enhanced
287 biological uptake during the stratified summer, representing the cumulative effect of nutrient
288 consumption that occurred since the spring bloom.

289 In spring, slightly higher surface Δ Nutrient values imply active phytoplankton uptake in the
290 surface layer, potentially supported by vertical nutrient replenishment from underlying waters.
291 During summer, the increase in Δ NO_x and Δ Silicate at depth relative to the surface points to
292 pronounced nutrient depletion in surface waters and subsequent export of organic matter. These
293 elevated values at depth likely stem from sustained biological uptake beneath the surface or
294 downward transport of nutrient-depleted waters, with limited remineralization en route,
295 resulting in deep nutrient concentrations lower than predicted by conservative mixing. The
296 modestly higher Δ Phosphate at the surface may reflect suppressed phosphate uptake under
297 nitrogen or silicate limitation, or additional phosphate input from glacial meltwater
298 insufficiently captured in the PSWw end-member (Hawkings et al., 2016).

299 Seasonal variability in phytoplankton activity was further examined through the relationship
300 between chlorophyll-*a* and Δ Nutrient (Fig. 6). In spring, the absence of significant correlations



301 ($r^2 < 0.04$) suggests that sampling preceded the main phytoplankton bloom, as supported by
302 elevated background nutrient levels. In contrast, summer data revealed weak but significant
303 negative correlations most notably for nitrate ($r^2 = 0.15$, $p < 0.05$) (Fig. 6d) and silicate ($r^2 =$
304 0.39 , $p < 0.01$) (Fig. 6f), indicative of biological drawdown, particularly by diatoms. These
305 observations are consistent with post-bloom conditions (Egge & Aksnes, 1992; Hodal et al.,
306 2012) and align with the seasonal rise in surface N/P ratios (from 14.99 to 18.80), suggestive
307 of NO_x depletion following diatom-dominated productivity (Leu et al., 2011).

308 The relationship between salinity and Δ Nutrient exhibited clear seasonal contrasts (Fig. 7).
309 In spring, all Δ Nutrient values showed weak negative correlations with salinity—for instance,
310 Δ Phosphate ($r^2 = 0.14$) (Fig. 7b) suggesting a reduced influence of high-salinity, nutrient-rich
311 AW and MAW on biological drawdown. During summer, NO_x ($r^2 = 0.29$) (Fig. 7d) and
312 phosphate ($r^2 = 0.43$) (Fig. 7e) continued to show positive correlations, though NO_x remained
313 weak. By contrast, silicate was inversely related to salinity ($r^2 = 0.94$) (Fig. 7f), suggesting an
314 additional input from glacial meltwater associated with PSWw. This influx may obscure the
315 biological drawdown signal typically expected for silicate.

316 The weak summer correlation between Δ Nutrient and chlorophyll-*a* is a critical finding, as
317 it implies a temporal decoupling between cumulative nutrient consumption and instantaneous
318 phytoplankton biomass. The consistently high Δ Nutrient values observed in summer serve as
319 a ‘biogeochemical memory’, reflecting the legacy of intense nutrient uptake during the
320 preceding spring bloom. In contrast, the lower and more variable chlorophyll-*a* concentrations
321 likely represent a snapshot of a post-bloom community, where phytoplankton biomass has
322 been diminished by factors such as grazing and sinking. Therefore, this study demonstrates that



323 Δ Nutrient is not merely a proxy for concurrent biological activity but rather a
324 powerful integrated indicator that quantifies the total impact of seasonal biological processes
325 on the nutrient inventory.

326 Interpretation of Δ Silicate warrants particular attention, given the influence of non-
327 conservative silicate input from glacial meltwater, which was not included as a separate end-
328 member in our model. The strong negative correlation between observed silicate and salinity
329 in summer ($r^2 = 0.94$, Fig. 7f) supports the presence of glacially sourced silicate (Hawkings et
330 al., 2017). As a result, the calculated Δ Silicate values likely underestimate the true biological
331 drawdown. Therefore, the values presented here should be considered a conservative estimate
332 of silicate consumption. Despite this limitation, elevated Δ Silicate values during summer
333 indicate that diatom-driven uptake was substantial, drawing down not only the silicate initially
334 present in water masses but also the ongoing input from meltwater. Ambient concentrations
335 approached potential limitation levels ($\sim 2 \mu\text{M}$), underscoring the scale of biological
336 consumption. Together, the underestimated yet high Δ Silicate and the approach toward limiting
337 concentrations provide strong evidence that biological processes, rather than physical mixing
338 alone, regulate silicate dynamics in Kongsfjorden during the summer season.

339 **3.4. Seasonal Shift in Nutrient Limitation Patterns**

340 The potential for nutrient limitations on phytoplankton growth in Kongsfjorden was
341 evaluated using the N/P ratio and the absolute concentrations of key nutrients (Fig. 8; see also
342 Section 3.1). During spring, the mean surface N/P ratio was 15.0 ± 2.7 , while that in the mean
343 deep-water was 13.8 ± 2.1 . These values were slightly below or close to the canonical Redfield
344 ratio of 16:1, suggesting that phytoplankton growth was not strongly limited by either nitrogen



345 or phosphorus during this period. If any trend was present, it may have leaned toward mild
346 nitrogen limitation. The relatively uniform N/P ratios with increasing depth also indicated
347 effective vertical mixing in spring.

348 In contrast, the summer season exhibited a pronounced increase in the surface N/P ratio,
349 averaging 18.8 ± 7.0 and exceeding the Redfield ratio. This shift strongly suggests a transition
350 toward phosphorus limitation in surface waters. The deep-layer N/P ratio remained lower
351 (mean: 13.2 ± 3.8), resulting in a marked vertical divergence. This contrast emphasizes the role
352 of enhanced summer stratification in establishing distinct biogeochemical regimes in the
353 surface and deep layers.

354 The possibility of phosphorus limitation in summer surface waters was supported by the
355 absolute phosphate concentrations observed during this period. The mean summer surface
356 phosphate concentration ($0.13 \pm 0.07 \mu\text{M}$) fell below the commonly used $0.2 \mu\text{M}$ threshold
357 indicating phosphorus limitation for Arctic phytoplankton (Tremblay et al., 2015).
358 Concurrently, the mean surface silicate concentration during summer was $1.66 \pm 0.39 \mu\text{M}$,
359 approaching the $2 \mu\text{M}$ threshold commonly associated with potential silicate limitation for
360 diatom growth (Egge & Aksnes, 1992). This nutrient regime, characterized by low phosphate
361 ($< 0.2 \mu\text{M}$) and low silicate ($< 2 \mu\text{M}$), likely imposed significant selective pressure on the
362 phytoplankton community, potentially favoring the dominance of small flagellates, which are
363 more competitive under nutrient-depleted conditions, particularly phosphorus limitation, over
364 diatoms (Degerlund & Eilertsen, 2010; Larsen et al., 2004; Egge & Aksnes, 1992).

365 These observed seasonal shifts in nutrient limitation patterns were closely linked to
366 phytoplankton community succession. In Arctic waters, spring diatom blooms typically deplete



large amounts of NO_x and silicate. Following these blooms, summer conditions, which are marked by stratification and altered nutrient ratios, may favor the dominance of other phytoplankton groups, including nitrogen-fixing microalgae or small species with distinct nutrient uptake strategies (Leu et al., 2011; Sakshaug, 2004). The observed increase in the surface N/P ratio from spring to summer supports this interpretation, as it indicates a faster depletion of phosphate relative to NO_x following the spring diatom bloom. This pattern is consistent with the known consequences of intense spring diatom blooms in Arctic fjords. While these blooms consume large amounts of NO_x and silicate, the post-bloom summer conditions, characterized by stratified and nutrient-depleted surface waters, often lead to a shift toward phosphorus limitation, as observed in our study. This succession favors smaller phytoplankton with distinct uptake strategies (Hodal et al., 2012; Leu et al., 2011).

4. Conclusion

The present study highlighted significant seasonal differences in water mass mixing and nutrient dynamics in Kongsfjorden, Svalbard. Spring conditions were dominated by MAW and active vertical mixing, resulting in relatively high and uniform nutrient concentrations, with N/P ratios near the Redfield ratio, suggesting nutrient-replete conditions. In contrast, summer featured increased surface freshening from PSW and PSW_w due to meltwater input, leading to strong stratification. This physical structure, together with enhanced biological uptake reflected by large Δ Nutrient values, led to substantial reductions in observed surface concentrations of NO_x (~69%), phosphate (~74%), and silicate (~47%) compared to spring. The Δ Nutrient metric effectively captured the cumulative biological drawdown over the season, acting as a 'biogeochemical memory' that is decoupled from instantaneous biomass. As a result, summer



389 surface waters shifted toward the potential co-limitation of phosphorus (N/P ~18.8; phosphate
390 ~0.13 ± 0.07 μM) and silicate (~1.66 ± 0.39 μM). These results suggest a seasonal transition
391 from a well-mixed, nutrient-rich spring regime to a stratified, nutrient-limited summer system
392 driven by biological processes. Understanding these dynamics is essential for predicting how
393 Arctic fjord ecosystems may respond to ongoing climate change, which is expected to affect
394 the water mass structure, meltwater input, and stratification, thus altering nutrient cycling and
395 primary productivity.



396 **Author contributions**

397 HK wrote the original draft and performed visualization and formal analysis. THK was
398 involved in the review and editing of the manuscript, funding acquisition, and
399 conceptualization. DH contributed to the review and editing of the manuscript and conducted
400 the investigation. SRP contributed to the review and editing of the manuscript and provided
401 critical input for the conceptualization of the study.

402



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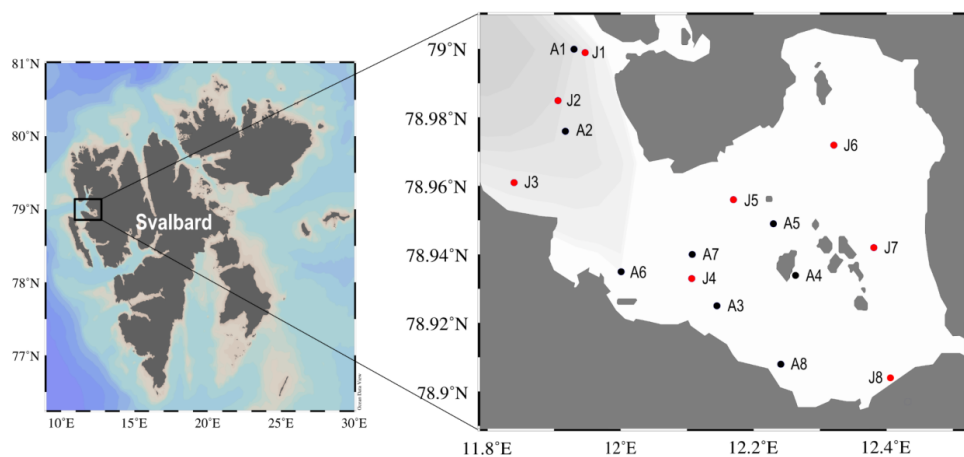
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519 **Figure 1.** Map of the study area in Kongsfjorden, which is located on the west coast of
520 Spitsbergen, Svalbard. Sampling stations from the spring (April 2023) and summer (July 2023)
521 cruises are shown. Black circles represent spring and red circles represent summer.



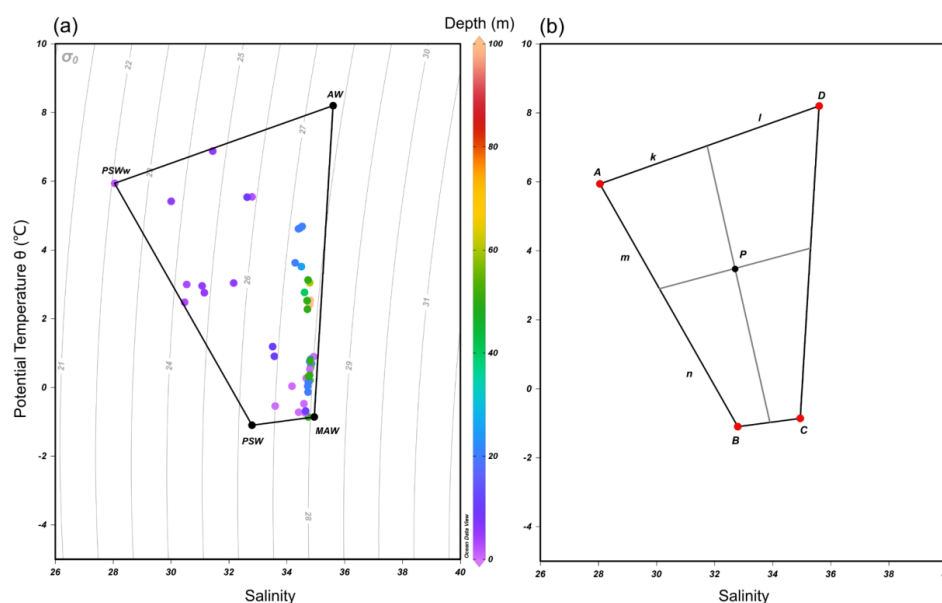
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524 **Figure 2.** (a) Temperature–salinity (T–S) diagram showing the four end-member water types
525 used in this study: Atlantic Water (AW), Modified Atlantic Water (MAW), Polar Surface Water
526 (PSW), and warm Polar Surface Water (PSWw). These end-members were defined based on
527 previously published criteria (e.g., Rudels et al., 2000) and supported by hydrographic data
528 collected during the cruises (Table 1). (b) Conceptual diagram of the four-end-member mixing
529 framework. Point P denotes an arbitrary water parcel in T–S space. Its location relative to the
530 end-members was used to estimate fractional contributions (f_A , f_B , f_C , and f_D), with the sum
531 constrained to unity ($f_A + f_B + f_C + f_D = 1$).

532



533



Figure 3. Vertical profiles of (a) temperature ($^{\circ}\text{C}$), (b) salinity, and (c) nitrate (μM), (d) phosphate (μM), and (e) silicate levels (μM) in Kongsfjorden. Black circles indicate spring data; red circles indicate summer data. Data represent measurements from multiple stations and depths.

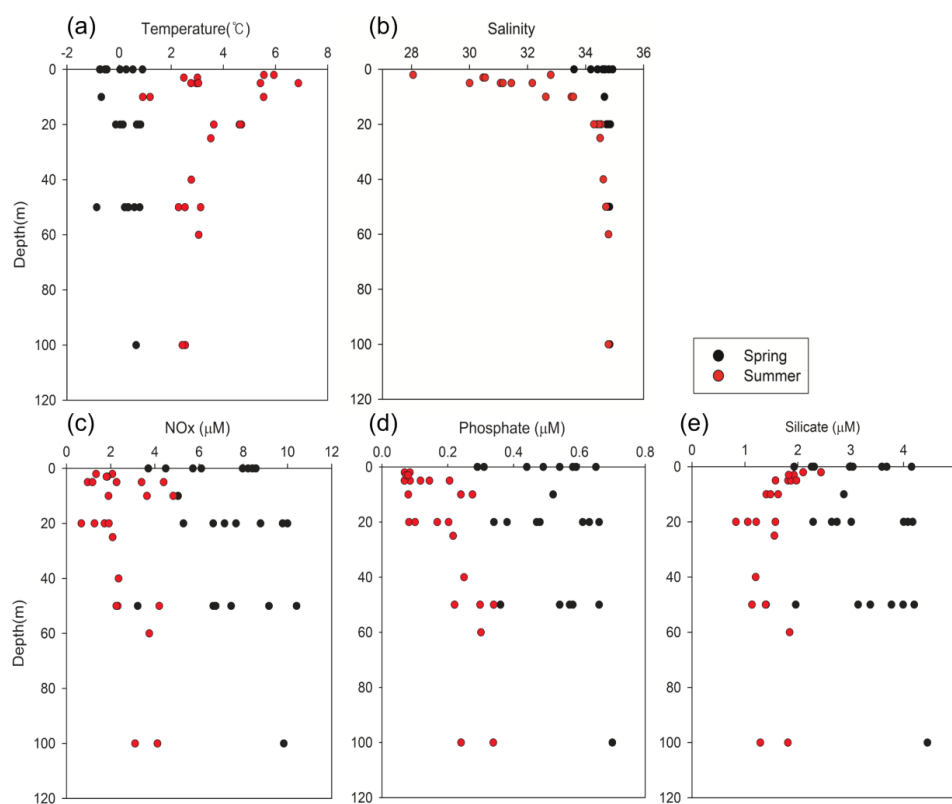




Figure 4. Relative contributions of the four end-member water masses (AW, MAW, PSW, and PSWw) in Kongsfjorden during (a) spring and (b) summer based on the four-component mixing model. Labels on the x-axis indicate the sampling station followed by the relative sampling depth: S (Surface), M (Mid-depth), and D (Deep). For the spring cruise (a), S, M, and D samples were typically collected at 0 m, 20 m, and 50 m, respectively (except for station A2, where D was 100 m). For the summer cruise (b), sampling depths varied by station, with S samples from 0-5 m, M from 10-25 m, and D from 50-100 m for all stations except J1. At station J1, S, M, and D samples were collected at 2 m, 5 m, and 20 m, respectively.

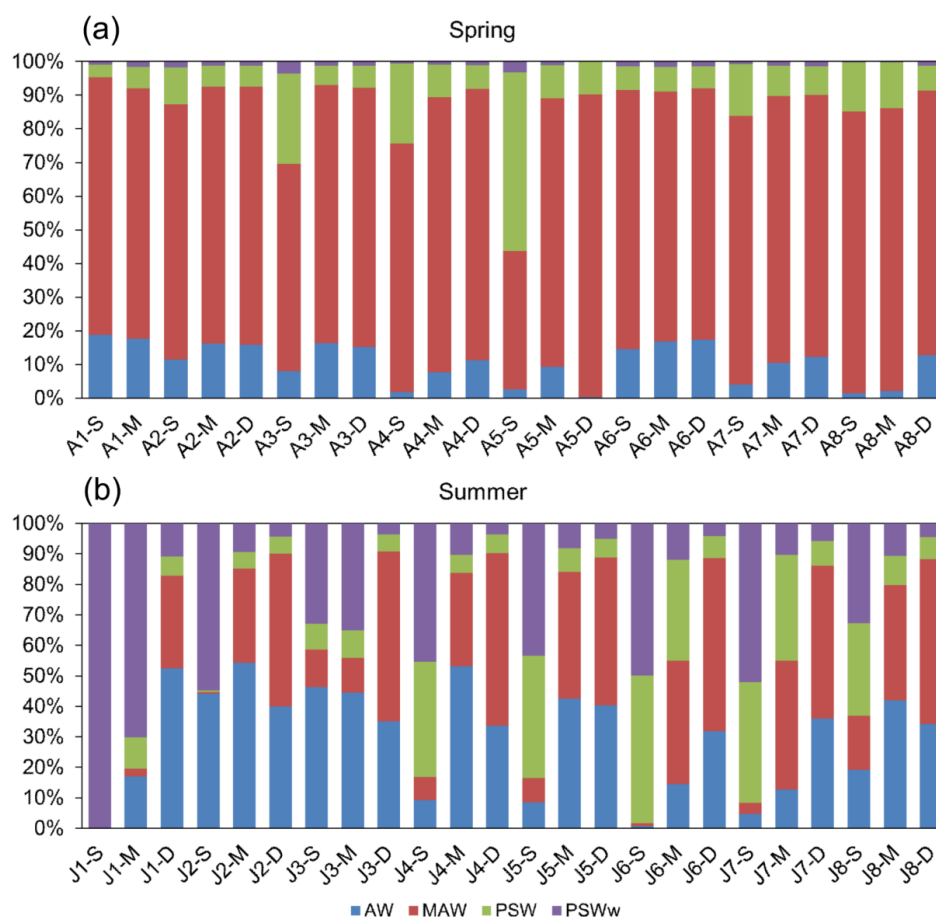
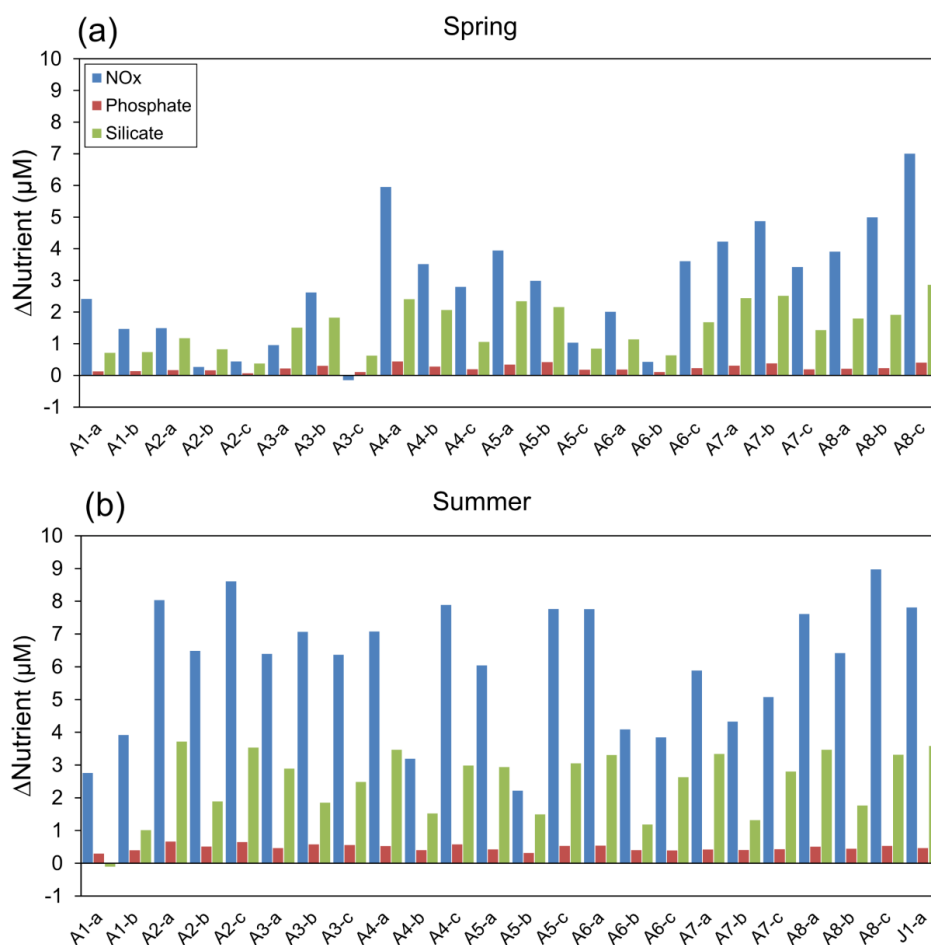




Figure 5. Differences between theoretical (mixing-derived) and observed nutrient concentrations ($\Delta\text{Nutrient} = \text{Theoretical} - \text{Observed}$; μM) during (a) spring and (b) summer. Bars represent $\Delta\text{Nutrient}$ values for nitrate, phosphate, and silicate, as indicated in the legend.

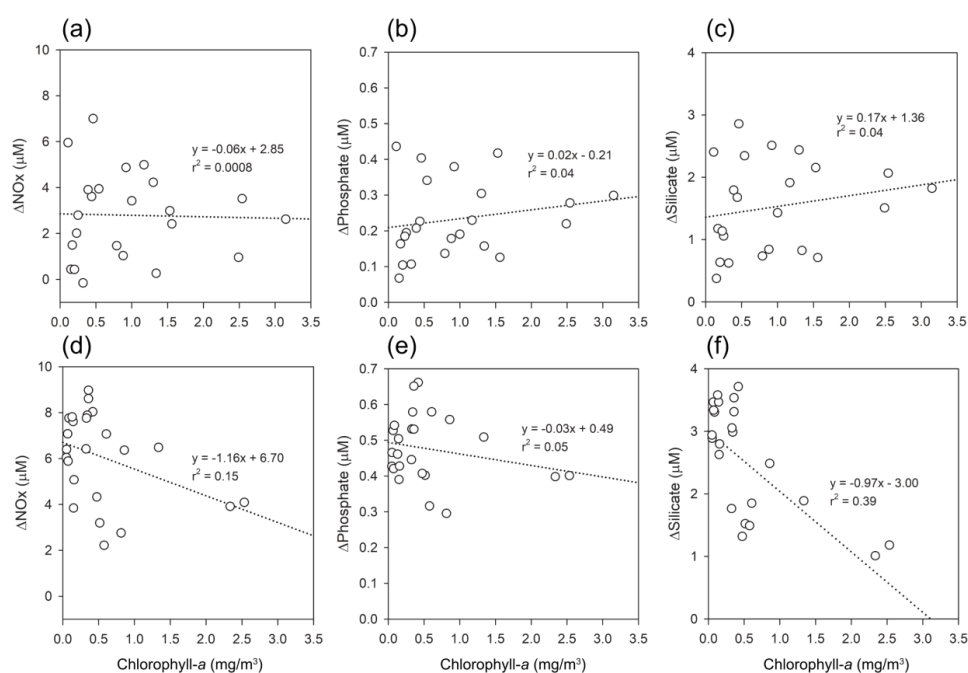


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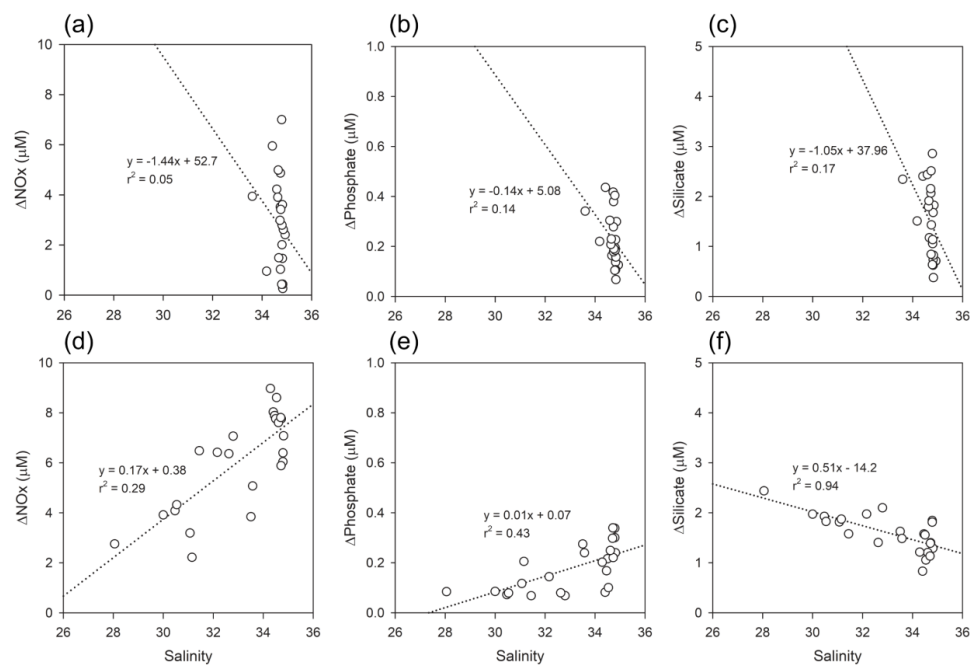
Figure 6. Relationships between chlorophyll-*a* (mg/m³) and Δ Nutrient (μ M) in Kongsfjorden: (a–c) spring, (d–f) summer. Regression lines and r^2 values are shown for each panel.



556



557 **Figure 7.** Relationships between salinity and Δ Nutrient (μM) in Kongsfjorden: (a–c) spring,
558 (d–f) summer. Regression lines and r^2 values are shown for each panel.



559



Figure 8. Vertical profiles of the nitrogen-to-phosphorus (N/P) molar ratio in Kongsfjorden. White circles represent spring values; black circles represent summer. The vertical dashed line indicates the Redfield ratio (16:1).

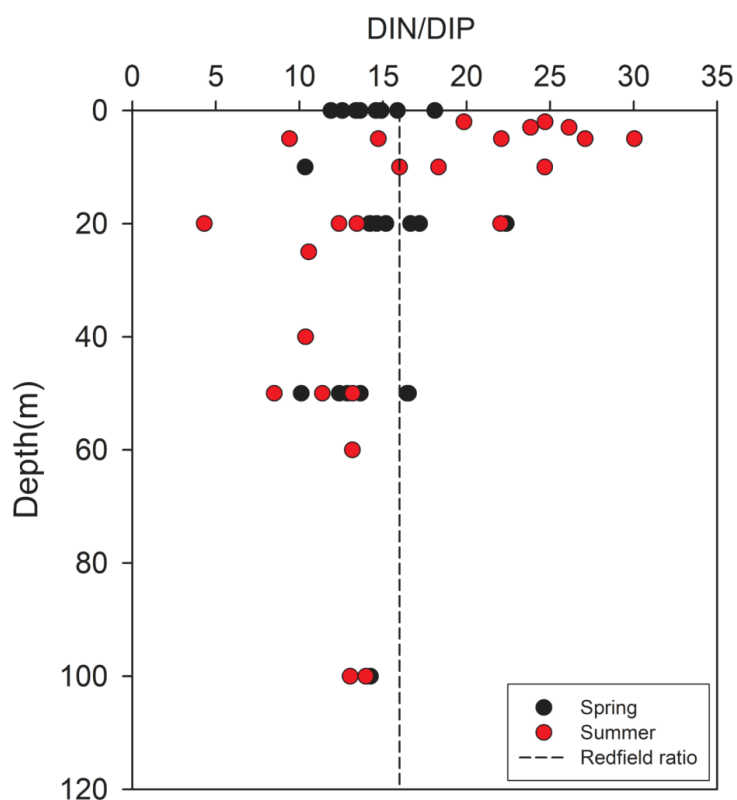




Table 1. Temperature (°C), salinity, and nutrient concentrations (nitrate, phosphate, and silicate; μM) for the four end-member water types: Atlantic Water (AW), Modified Atlantic Water (MAW), Polar Surface Water (PSW), and warm Polar Surface Water (PSWw). The temperature-salinity (T-S) definitions were adopted from Rudels et al. (2000), with σ_0 represents the potential density anomaly referenced to 0 dbar. determined based on the characteristics of the most representative samples collected in this study (identified at the vertices of the T-S diagram in Fig. 2a). Nutrient values for each water mass are based on literature values from Duarte et al. (2021).

End member	Water mass	NOx (μM)	Phosphate (μM)	Silicate (μM)	Temperature (°C)	Salinity	Reference (Rudels et al. 2000)
A	Atlantic Water (AW)	10.66	0.82	4.86	8.2	35.6	$27.70 < \sigma_0 < 27.97, T > 2^\circ\text{C}, \text{ or } 27.97 < \sigma_0, \text{ and } \sigma_{0.5} < 30.444, T > 0^\circ\text{C}$
B	Modified Atlantic Water (MAW)	10.55	0.78	4.94	-0.86	34.95	$27.70 < \sigma_0 < 27.97, T < 0^\circ\text{C}, S < 34.676 + 0.232 \cdot T, \text{ or } 27.97 < \sigma_0, \text{ and } \sigma_{0.5} < 30.444, T > 0^\circ\text{C}$
C	Polar Surface water (PSW)	6.91	0.56	3.85	-1.1	32.8	$27.70 > \sigma_0, T < 0^\circ\text{C}$
D	Polar Surface water warm (PSWw)	4.83	0.38	2.33	5.94	28.05	$27.70 > \sigma_0, T > 0^\circ\text{C}$

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