



methane emissions in China

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18 **Abstract**

19 Methane (CH₄), the second most important anthropogenic greenhouse gas, significantly impacts global
20 warming. As the world's largest anthropogenic CH₄ emitter, China faces challenges in accurately
21 estimating its emissions. Top-down methods often suffer from coarse resolution, limited data
22 constraints, and result discrepancies. Here, we developed the Regional Methane Assimilation System
23 (RegGCAS-CH₄) based on the WRF-CMAQ model and the EnKF algorithm. By assimilating
24 extensive TROPOMI column-averaged dry CH₄ mixing ratio (XCH₄) retrievals, we conducted high-
25 resolution nested inversions to quantify daily CH₄ emissions across China and Shanxi Province in
26 2022. Nationally, posterior CH₄ emissions were 45.1 TgCH₄·yr⁻¹, 36.5% lower than the EDGAR
27 estimates, with the largest reductions in the coal and waste sectors. In North China, emissions
28 decreased most significantly, mainly attributed to the coal and enteric fermentation sectors. Posterior
29 emissions in coal-reliant Shanxi Province decreased by 46.3%. Sporadic emission increases were
30 detected in major coal-producing cities but were missed by the coarse-resolution inversion. Monthly
31 emissions exhibited a winter-low, summer-high pattern, with the rice cultivation and waste sectors
32 showing higher seasonal increases than those in EDGAR. The inversion significantly improved XCH₄
33 and surface CH₄ concentration simulations, reducing emission uncertainty. Compared to other bottom-
34 up/top-down estimates, our results were the lowest, primarily because the high-resolution inversion
35 better captured local emission hotspots. Sensitivity tests underscored the importance of nested
36 inversions in reducing the influence of boundary condition uncertainties on emission estimates. This
37 study provides robust CH₄ emission estimates for China, crucial for understanding the CH₄ budget and
38 informing climate mitigation strategies.

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42 **Keywords** CH₄ emissions, Emission inversion, Methane assimilation system, China, Data assimilation

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45 **1. Introduction**

46 Methane (CH₄) ranks as the second most significant anthropogenic greenhouse gas and plays a crucial
47 role in global warming. Its global warming potential is 28–29.8 times that of carbon dioxide (CO₂)
48 over a 100-year time scale and contributes approximately 17% to the total radiative forcing of
49 greenhouse gases since the industrial era (Forster et al., 2021; Saunio et al., 2025). With a relatively
50 short steady state atmospheric budget lifetime of slightly over 9 years (Prather, 2007), CH₄ is a key
51 target for rapid climate change mitigation efforts (Tu et al., 2024), as emphasized by the Global
52 Methane Pledge, which aims to reduce global anthropogenic CH₄ emissions by 30% below 2020 levels
53 by 2030 (GMP, 2023). China, as the world's largest CH₄ emitter, accounting for around 14%–22% of
54 global anthropogenic CH₄ emissions, faces a complex challenge (Zhang et al., 2022; Janssens-
55 Maenhout et al., 2019). China's CH₄ emissions originate from diverse sources, with coal mining and
56 rice cultivation being particularly prominent (Nisbet, 2023; Lin et al., 2021). Comprehending the
57 spatiotemporal distribution and trends of China's CH₄ emissions is of utmost importance for
58 formulating effective climate policies and fulfilling international climate commitments.

59 The bottom-up approaches, which rely on activity data and emission factors, have been widely used to
60 estimate China's CH₄ emissions. However, accurately estimating these emissions remains a significant
61 challenge. Existing inventories are often characterized by large uncertainties in both magnitude and
62 sectoral attribution, with differences between various bottom-up estimates reaching up to 40–60% for
63 China (Saunio et al., 2025). Among the various sources, estimates of China's coal mine CH₄ emissions
64 can range from 14–28 Tg·yr⁻¹ (Sheng et al., 2019). This wide disparity leads to much uncertainty in
65 the bottom-up estimates. There are multiple reasons for the uncertainties, especially the lack of
66 comprehensive data on emission sources, especially for small-scale and sporadic emitters, and the use
67 of complex and perhaps inappropriate emission factors in different sectors (Stavert et al., 2022). For
68 instance, in coal mining, the emission factors vary significantly depending on the mining method
69 (underground vs. surface mining), geological conditions, and the quality of coal. More generally, due
70 to the use of a spatial proxy approach for the spatial allocation of the total emissions, existing global
71 inventories exhibit significant spatial discrepancies when compared with high-resolution local
72 measurements, resulting in misattribution of emissions (Qin et al., 2024).



73 In contrast, top-down inversion methods, that utilize satellite and surface observations in conjunction
74 with atmospheric transport models, have the potential to provide more accurate and comprehensive
75 estimates. Compared with ground-based inversions, satellite-based atmospheric inversions, such as
76 those using column-averaged dry CH_4 mixing ratios (XCH_4) data from the Greenhouse Gases
77 Observing Satellite (GOSAT) or the TROPospheric Monitoring Instrument (TROPOMI), can offer
78 valuable insights into the spatial distribution of emissions (Lu et al., 2023). Studies utilizing satellite
79 observations have quantified CH_4 emissions from various sources, including oil, gas, and coal mining
80 sectors at the global, regional, and key source scales (Nesser et al., 2024; Bai et al., 2024; Zhang et al.,
81 2021). For instance, Maasakkers et al. (2019) investigated the contribution of different regions to the
82 global CH_4 budget with GOSAT XCH_4 data, identifying areas with significant emissions and sinks.
83 Pandey et al. (2019) used TROPOMI XCH_4 observations to reveal extreme CH_4 leakage from a natural
84 gas well blowout, demonstrating the instrument's ability to detect both large-scale and short-term
85 emission events.

86 The Methane Emission Control Action Plan released by China explicitly states the exploration and
87 implementation of research on atmospheric methane emission inversion models, and strengthens the
88 verification of emission inventories by inversion data (MECAP, 2023). A number of inversions
89 relevant to China have been conducted with satellite observations. Using inverse analysis of 2019
90 TROPOMI XCH_4 data, Chen et al. (2022) quantified CH_4 emissions across China and attributed
91 contributions to specific sectors. Zhang et al. (2022) estimated China's CH_4 emissions from 2010 to
92 2017 by combining satellite and surface observations, revealing complex linkages between emission
93 trends and associated policy drivers. However, the trends of China's CH_4 emissions quantified by the
94 top-down approach (Sheng et al., 2021; Miller et al., 2019) are contrary to those estimated by the
95 bottom-up approach, which is mainly due to the unclear quantification of emissions from the coal
96 mining sector (Sheng et al., 2019; Liu et al., 2021).

97 Currently, global-scale CH_4 assimilation systems are widely applied, such as CarbonTracker- CH_4 in
98 the United States (Bruhwiler et al., 2014), CAMS in Europe (Agustí-Panareda et al., 2023), NTFVAR
99 in Japan (Wang et al., 2019), LMDz-SACS-CIF in France (Thanwerdas et al., 2022), and GONGGA-
100 CH_4 in China (Zhao et al., 2024). However, significant knowledge gaps remain in accurately estimating
101 CH_4 emissions at the regional scale. There are relatively few existing regional CH_4 assimilation



102 systems, such as the ICON-ART-CTDAS (Steiner et al., 2024) and CarbonTracker Europe-CH₄
103 (Tsuruta et al., 2017) in Europe, and the IMI in the United States (Varon et al., 2022). Additionally,
104 existing regional inversions often use global atmospheric transport models with relatively coarse
105 resolutions on the annual or monthly average scale, lacking high-resolution spatial and temporal
106 coverage to capture local characteristics and short-term variations in emissions (Chen et al., 2022). Qu
107 et al. (2021) highlighted significant challenges in separating rice and coal emissions over southeast
108 China due to coarse grid resolution. High-resolution estimates are crucial for understanding the
109 detailed distribution of emissions, especially in regions with heterogeneous source landscapes. For
110 instance, in Shanxi Province, a major coal producing region, the traditional low-resolution models may
111 fail to capture the emissions from numerous small-scale coal mines. Additionally, the inversion results
112 of global and regional systems still show significant differences, mainly due to different observation
113 data, inversion methods, transport models, and resolutions (Kou et al., 2025; Chen et al., 2022; Liang
114 et al., 2023). Another challenge is that regional models need to consider the impact of boundary fields,
115 especially for long-lived species. Current studies usually directly derive regional boundary fields from
116 global simulation or analysis fields, which still contain large errors (Zhang et al., 2022; Kou et al.,
117 2025). Consequently, China's contribution to the global CH₄ budget remains unclear.

118 In this study, we aimed to address these limitations by developing a Regional Methane Assimilation
119 System (RegGCAS-CH₄) based on the Weather Research and Forecasting-Community Multiscale Air
120 Quality (WRF-CMAQ) model and the Ensemble Kalman Filter (EnKF) algorithm (Evensen, 1994).
121 This system enabled the assimilation of a large volume of TROPOMI XCH₄ data to achieve nested
122 inversions of daily CH₄ emissions with high spatial-temporal resolution. We conducted a
123 comprehensive analysis of the spatial characteristics and monthly variations of China's CH₄ emissions
124 in 2022. In particular, we focused on the nested inversion analysis of CH₄ emissions in Shanxi
125 Province, China.

126 The novelty of our study lies in the high spatial resolution emission inversions, which allows for
127 capturing the fine-scale features of CH₄ emissions that are often missed by global inversion models.
128 Secondly, by assimilating a larger amount of TROPOMI XCH₄ observations, which feature high
129 spatial resolution, wide spatial coverage, and high-frequency retrievals (~ 100 times more
130 observations than GOSAT) (Qu et al., 2021), we can incorporate the latest and more representative



131 information on atmospheric CH₄ concentrations to update the daily emissions. This high temporal
132 resolution is essential for understanding the short-term fluctuations in emissions and their response to
133 various factors. Notably, we have pre-updated the global boundary fields, which effectively mitigates
134 the impact of boundary condition errors on regional emission inversions. This is a crucial step that has
135 not been emphasized or implemented in most previous studies. By comparing our results with the
136 currently widely used inventories, we aim to provide a more accurate and detailed understanding of
137 China's CH₄ emissions, which is crucial for formulating effective climate change mitigation strategies.

138 **2. Method and data**

139 **2.1 Data assimilation system**

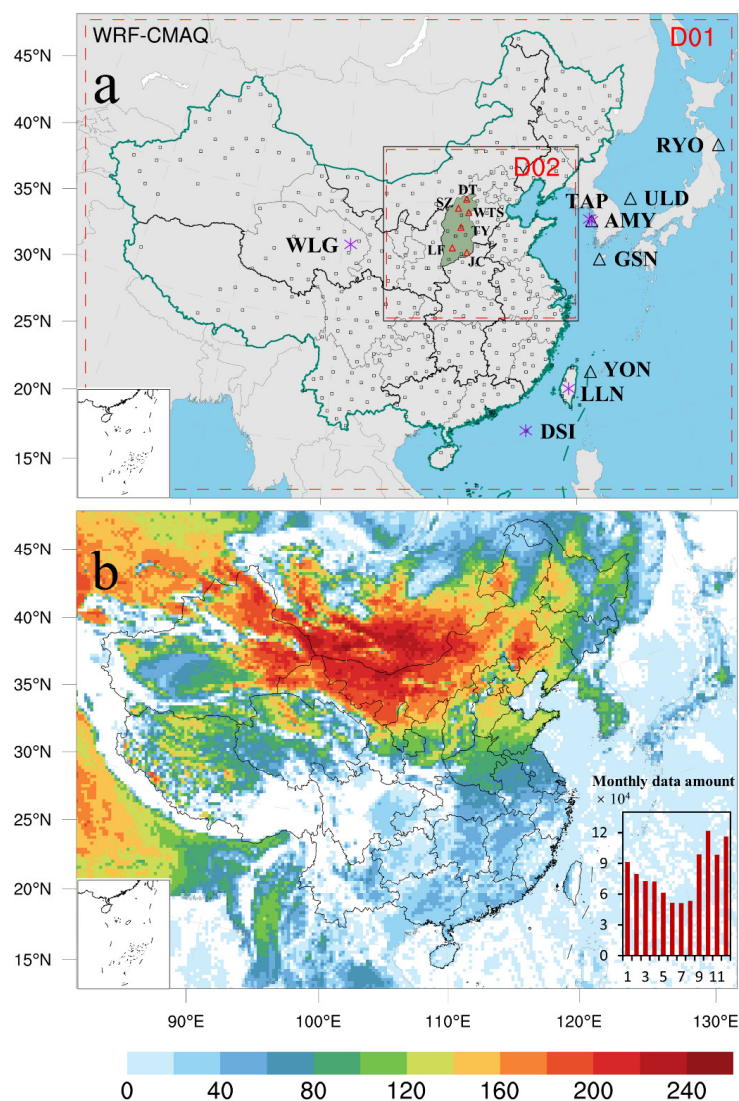
140 The RegGCAS-CH₄ system was extended from RegGCAS (Zhang et al., 2024), which was developed
141 by Feng et al. (2023) based on the Regional multi-Air Pollutant Assimilation System (RAPAS) (Feng
142 et al., 2020) and Global Carbon Assimilation System (GCAS), version 2 (Jiang et al., 2021).
143 RegGCAS-CH₄ includes a regional chemical transport model (CTM) and an ensemble square root
144 filter (EnSRF) assimilation module (Whitaker and Hamill, 2002), which are employed to simulate
145 atmospheric compositions and infer anthropogenic emissions, respectively. The basic framework is
146 almost identical to that in Zhang et al. (2024), in which we inferred the fossil fuel CO₂ emission using
147 surface CO₂ observations. In this study, the system was extended to high-resolution nested inversion
148 for CH₄ emissions, which can optimize emissions from the outer to inner domain and reduce the
149 influence of inaccurate boundary conditions on the inversion of the inner area (Feng et al., 2022).
150 Moreover, the assimilation framework was updated to optimize CH₄ emissions by assimilating the
151 TROPOMI CH₄ column retrievals. The RegGCAS-CH₄ performed a “two-step” inversion scheme in
152 each data assimilation (DA) window. First, the prior emissions were optimized using the available
153 atmospheric observations. Then, the optimized emissions were input back into the CTM to generate
154 the initial fields for the next assimilation window and boundary conditions for the inner domain. Thus,
155 the uncertainties in boundary conditions for the estimates of emissions in the inner domain were also
156 reduced. Simultaneously, the optimized emissions were transferred to the next window to serve as prior
157 emissions. This “two-step” scheme facilitates error propagation and iterative emission optimization,
158 which can ensure the mass conservation of the system and effectively enhance the stability and



159 consistency of the emission updates (Feng et al., 2024).

160 **2.1.1 Atmospheric transport model**

161 The Weather Research and Forecast (WRF v4.0) model (Skamarock and Klemp, 2008) and the
162 Community Multiscale Air Quality Modeling System (CMAQ v5.0.2) (Byun and Schere, 2006) were
163 applied to simulate meteorological conditions and atmospheric chemistry, respectively. In this study,
164 the CMAQ model employed two-nested simulations. The outer domain (D01), which covered the
165 whole mainland of China with a grid of 225×165 cells, and the inner domain (D02), which covered
166 the Shanxi Province and surrounding areas with a grid of 195×174 cells, had grid spacings of 27 and
167 9 km, respectively (Figure 1). There were 20 levels on the sigma–pressure coordinates extending from
168 the surface to 100 Pa. To account for the rapid expansion of urbanization, we updated the underlying
169 surface information for urban and built-up land using the MODIS Land Cover Type Product
170 (MCD12C1) Version 6.1 of 2022. The detailed configuration of WRF-CMAQ is shown in Table S1.



171

172 **Figure 1** (a) Nested inversion domain and (b) number of TROPOMI XCH₄ retrievals during 2022. The
 173 red dashed frame depicts the CMAQ modeling domain; black squares represent the surface
 174 meteorological measurement sites; red triangles represent the six *in-situ* CH₄ measurement sites in
 175 Shanxi Province; purple asterisks and black triangles represent the flask and *in-situ* CH₄ measurement
 176 sites, respectively. The TROPOMI observations that fall within the same model grid are processed and
 177 counted as one super-observations. The subfigure in panel (b) shows the total number of super-
 178 observations for each month within the study area.



179 The meteorological initial and lateral boundary conditions were obtained from the Final (FNL)
180 Operational Global Analysis data of the National Center for Environmental Prediction (NCEP) with a
181 $1^\circ \times 1^\circ$ resolution at 6-h intervals. The chemical lateral boundary conditions for the outer domain were
182 derived from the CAMS global inversion-optimized CH_4 concentrations with a $1^\circ \times 1^\circ$ resolution at
183 6-hour intervals (Bergamaschi et al., 2013), while those for the inner domain were obtained from the
184 forward simulation of the outer domain with optimized CH_4 emissions. In the first DA window, the
185 chemical initial conditions were also extracted from the CAMS, whereas in subsequent windows, they
186 were derived through forward simulation using optimized emissions from the previous window. Since
187 the time required for CH_4 to be transported in the study area is much shorter than its lifetime, to reduce
188 the computational cost to the lowest possible level, we deactivated all the atmospheric chemical
189 reaction processes (Chen et al., 2022; Kou et al., 2025) and incorporated a set of CH_4 tracers variables
190 specifically for conducting ensemble simulations in the CMAQ, enabling all CH_4 concentration sets
191 to be obtained from a single simulation.

192 It is of utmost importance to eliminate biases in boundary conditions, as any such biases have the
193 potential to cascade through the system. We found that the boundary conditions extracted from the
194 CAMS global fields still had considerable biases over East Asia (see Section 4). Thus, we calculated a
195 scaling factor for each grid of CAMS fields (50°E – 160°E , 0° – 70°N) against TROPOMI XCH_4
196 retrievals, and then applied these factors to correct the biases in the boundary conditions of CAMS
197 (Figure S1). Based on each pair of CAMS and TROPOMI data, we first calculated the column
198 concentration of the CAMS reanalysis field. Then, the column concentrations of CAMS and
199 TROPOMI were respectively smoothed in the longitudinal and latitudinal directions with a radius of
200 4° , and in terms of time using a 4-day window. Subsequently, the latitudinal average values were
201 calculated, along with the average column concentrations biases between CAMS and TROPOMI. After
202 that, linear interpolation was applied to fill in the missing values of the biases in the longitudinal
203 direction. For the biases of the grids where there were missing values at the same latitude, it was
204 assumed that they were consistent with the average biases of the non-missing values at that latitude.
205 Finally, grid-by-grid bias was calculated and correction was carried out for the original CAMS CH_4
206 concentrations.

207



2.1.2 EnKF assimilation algorithm

The EnKF is based on the Monte Carlo approach, which uses a stochastic ensemble of model states to approximate the probability distribution of the true state. It has been widely employed for updating the model state by incorporating observational data to minimize the difference between the model-simulated and observed values (Evensen, 1994). The ensemble square root filter (EnSRF) approach, introduced by Whitaker and Hamill (2002), was used to constrain the CH₄ emissions in this study.

The EnSRF process commences with the initialization of an ensemble of model states, which are generated by applying a Gaussian perturbation with an average value of zero and the standard deviation of the uncertainty to a state vector $\mathbf{X}^b$. The ensemble-estimated background error covariance matrix $\mathbf{P}^b$ is then calculated as:

$$\mathbf{P}^b = \frac{1}{N-1} \sum_{i=1}^N (\mathbf{X}_i^b - \bar{\mathbf{X}}^b) (\mathbf{X}_i^b - \bar{\mathbf{X}}^b)^T \quad (1)$$

where N is the ensemble size; $\mathbf{X}_i^b$ represents the i th sampling; $\bar{\mathbf{X}}^b$ represents the mean of the ensemble samples. $\mathbf{P}^b$ plays a pivotal role in determining how the model state will be adjusted based on new observations.

During the forecast step, each ensemble member is advanced in time using the WRF-CMAQ model. As the model runs, uncertainties in emissions can lead to errors in CH₄ concentrations, and thus the response relationships of the concentration ensembles to the emission ensembles are obtained. In the analysis step, observational data $\mathbf{y}$ are incorporated to update the analyzed state. The ensemble mean of the analyzed state $\bar{\mathbf{X}}^a$ is regarded as the best estimate of emissions, which is obtained through the following equations:

$$\bar{\mathbf{X}}^a = \bar{\mathbf{X}}^b + \mathbf{K}(\mathbf{y} - \mathbf{H}\bar{\mathbf{X}}^b) \quad (2)$$

$$\mathbf{K} = \mathbf{P}^b \mathbf{H}^T (\mathbf{H} \mathbf{P}^b \mathbf{H}^T + \mathbf{R})^{-1} \quad (3)$$

where $\mathbf{R}$ is an error covariance matrix; $\mathbf{K}$ is the Kalman gain matrix, estimated from the ensemble simulations and determining the relative contributions of observation and background to analysis. The state vector was defined as $\mathbf{X}^b = (\mathbf{Ea}^T, \mathbf{Ep}^T)^T$, where $\mathbf{Ea}$ and $\mathbf{Ep}$ represent the vectors of CH₄ emissions for the area and power plant sources, respectively. Area sources included the daily total



emissions from the enteric fermentation & manure, landfills & waste, rice cultivation, coal mining, Oil & gas, industry, transport sources, etc. The updated emissions are then used as the new initial states for the next forecast step, creating a cycle of assimilation that gradually refines the estimate of CH₄ emissions.

The observation operator **H** maps the model state to the observation space. In the context of CH₄, **H** is configured to first horizontally geo-locate simulated CH₄ concentrations to match the TROPOMI XCH₄ retrievals. Subsequently, it remaps the sub-column concentrations from the 20-layer CMAQ vertical grid to the 12-layer TROPOMI vertical grid by totally or partially allocating CMAQ layers to TROPOMI layers based on pressure edges (Varon et al., 2022). Finally, the column average dry-air mixing ratio XCH_4_s can be obtained by applying the TROPOMI column averaging kernel of each layer a_i to sub-columns:

$$XCH_4_s = VCH_4_s / VAIR_{dry,i} \quad (4)$$

$$VCH_4_s = VCH_4_p + \sum_{i=1}^n a_i (XCH_{4s,i} \Delta VAIR_{dry,i} - \Delta VCH_{4p,i}) \quad (5)$$

where n is the number of retrieval layers; $\Delta VCH_{4p,i}$ and VCH_4_p represent the prior CH₄ column in retrieval layer i and total CH₄ column; $XCH_{4s,i}$ is the simulated dry air mixing ratio of CH₄ in retrieval layer i ; $\Delta VAIR_{dry,i}$ and $VAIR_{dry,i}$ represent the dry air column in retrieval layer i and total dry air column, respectively, provided along with the TROPOMI product.

In this study, the DA window was set to 1 d, meaning that daily TROPOMI XCH₄ retrievals were utilized as emission constraints. The ensemble size was set to 50 to fully characterize the system uncertainties and inversion accuracy. To address the problem of spurious long-distance correlations in the EnKF (Houtekamer and Mitchell, 2001), we applied covariance localization using the Gaspari and Cohn function, which is a piecewise continuous fifth-order polynomial approximation of a normal distribution (Miyazaki et al., 2017; Gaspari and Cohn, 1999). Taking into account the transmission distance of CH₄ within one assimilation window, the localization scale was set to 300 km. By setting the covariance to be smaller the farther away from the observations, we could reduce the spurious influence of remote observations on the local analysis.

2.2 Prior emissions and uncertainties

The prior anthropogenic CH₄ emissions were taken from the Emission Database for Global



262 Atmospheric Research version 8.0 (EDGAR v8), which offers detailed monthly gridded CH₄
263 emissions at 0.1°×0.1° from various anthropogenic sources (Crippa et al., 2024). The daily emissions,
264 obtained by uniformly allocating the aggregated monthly emission inventory, were directly utilized as
265 the initial estimate in the RegGCAS-CH₄. For natural CH₄ emissions, we utilized the ensemble average
266 of 18 emissions from the WetCHARTs v1.3.1 inventory (0.5° × 0.5°, monthly) in 2019 for wetland
267 emissions (Bloom et al., 2021), the Global Fire Emissions Database (GFED v4.1s, 0.25° × 0.25°, three-
268 hourly) in 2022 for biomass burning emissions (van Wees et al., 2022), and CAMS global emissions
269 inventory (CAMS-GLOB-TERM v1.1, 0.5° × 0.5°, monthly) in 2000 for termite emissions (Jamali et
270 al., 2011). Additionally, CH₄ sinks resulting from soil absorption were derived from datasets simulated
271 by Soil Methanotrophy Model (MeMo v1.0, 1° × 1°, yearly) in 2020 (Murguia-Flores et al., 2018).

272 Given the model error compensation and the relatively comparable emission uncertainties from one
273 day to the next, we applied an identical uncertainty of 40% to each emission grid at every DA window.
274 Since the RegGCAS-CH₄ adopts a ‘two-step’ inversion strategy and the daily posterior emissions are
275 iteratively optimized, the emission analysis is generally no longer sensitive to the prior uncertainties
276 of the original emission inventory after several assimilation windows (Zhang et al., 2024).

277 **2.3 Assimilation data and errors**

278 TROPOMI, onboard the Sentinel-5 Precursor satellite, was launched in 2017. Operating in a sun-
279 synchronous orbit with an equator local overpass time of 13:30 h, TROPOMI offers daily global
280 continuous monitoring of XCH₄. The RemoTeC full-physics algorithm is used to retrieve XCH₄ from
281 TROPOMI measurements of sunlight backscattered by Earth's surface and atmosphere in the near-
282 infrared (NIR) and shortwave-infrared (SWIR) spectral bands (Lorente et al., 2022). After 2019, the
283 spatial resolution of TROPOMI was adjusted to a remarkable 5.5×3.5 km² at nadir, enabling the
284 identification of even relatively small-scale CH₄ emission sources. We used the TROPOMI XCH₄ level
285 2 data product to estimate CH₄ emissions.

286 All TROPOMI individual pixels with a quality assurance value (qa_value) smaller than 0.5 were
287 discarded, which corresponds to high-cloud conditions or the presence of snow or ice. However, we
288 still found many unrealistic low values, especially in summer. To further minimize the impact of
289 outliers, we selected another XCH₄ product generated based on the WFMD algorithm for cross-



validation (Schneising et al., 2023). Only those pixels that were concurrently available in the TROPOMI/WFMD product and met the quality flag requirements were assimilated. Figure 1b illustrates the observational amount of TROPOMI XCH₄ in 2022 at each grid. Although the distribution of filtered data exhibits spatial nonuniformity, most grid cells in the central-northern regions with intense anthropogenic emissions have observational coverage for more than 100 days in 2022. Additionally, the monthly variation in the data amount shows fluctuations, with the peak occurring at the beginning and end of the year and the troughs around the middle. According to the latest quarterly validation report, the 1 σ spread of the relative difference between the TROPOMI and the TCCON (Wunch et al., 2011) is of the order of 0.7% for bias corrected product, which is recommended to be considered as an upper boundary of the random uncertainty of the satellite data (Lambert et al., 2024).

2.4 Evaluation Data

To evaluate the performance of the WRF simulations, we utilized the surface meteorological measurements of 400 stations with 3-hour intervals, including temperature at 2 m (T2), relative humidity at 2 m (RH2), and wind speed at 10 m (WS10). These measurements were obtained from the National Climate Data Center (NCDC) integrated surface database (<http://www.ncdc.noaa.gov/oa/ncdc.html>, last access: 25 August 2024). Overall, the WRF demonstrated satisfactory performance in reproducing meteorological conditions over China (Figure S2), with minimal biases of -0.4°C for T2, -4.9% for RH2, and 0.5 m/s for WS10, respectively. To evaluate the posterior CH₄ emissions, two parallel forward modeling experiments were conducted: the control experiment with prior emissions (CEP) and the validation experiment with posterior emissions (VEP). Both experiments utilized identical meteorological fields, as well as initial and boundary conditions. We evaluated: (1) the simulated XCH₄ against TROPOMI XCH₄; (2) the simulated surface CH₄ concentrations against independent ground CH₄ observations from eleven *in-situ* and five flask monitoring sites (Table S2). The *in-situ* measurements included five global sites outside China (AMY, GSN, RYO, ULD, and YON) with hourly resolution from NOAA's GLOBALVIEWplus CH₄ ObsPack v6.0 (Schuldt et al., 2023), as well as six regional sites in Shanxi Province (TY, DT, LF, SZ, JC, and WTS) equipped with Picarro G2301 analyzers for high-precision, 5-second CH₄ measurements (data were processed as daily averages to reduce random errors). The Picarro instrument, based on



319 wavelength-scanned cavity ring-down spectroscopy technology, is designated by the World
320 Meteorological Organization as the international reference instrument for CH₄ observations in
321 international comparisons. Additionally, observations from five flask sampling sites (AMY, DSI, LLN,
322 TAP, and WLG) were also obtained from ObsPack dataset, providing weekly measurements.

323 **3. Results and discussion**

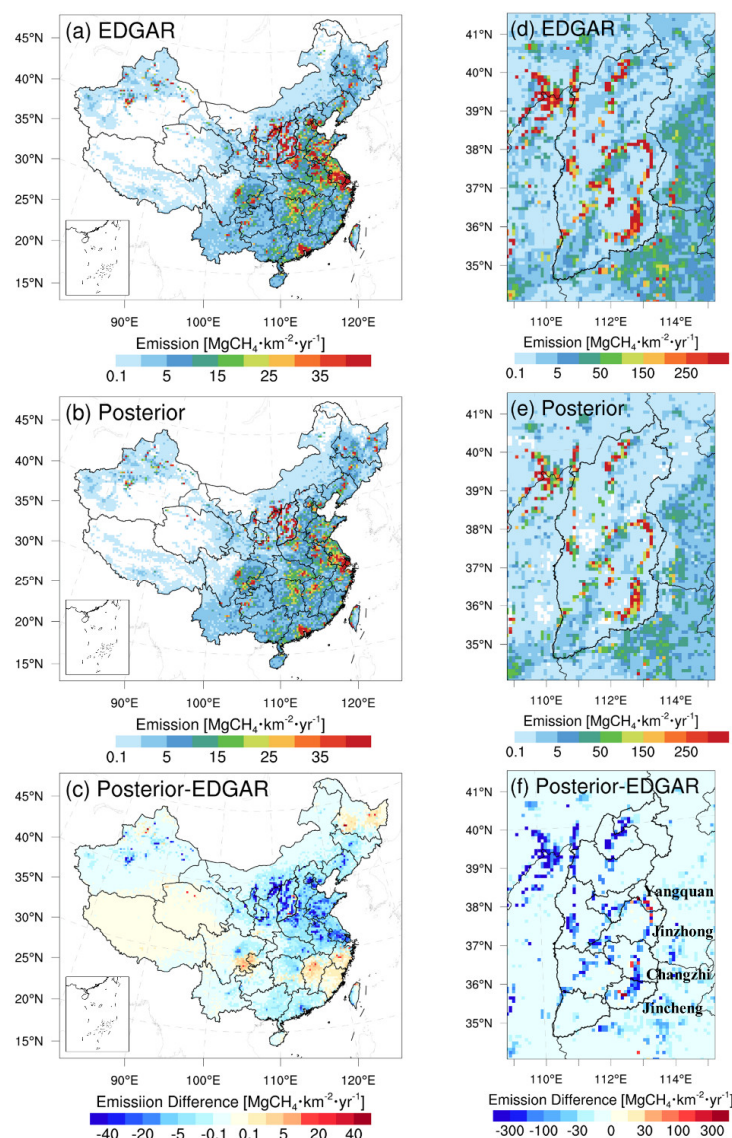
324 **3.1 Posterior CH₄ emissions**

325 Figure 2 shows the spatial distributions of the estimated annual CH₄ emissions and the differences
326 from the prior emissions (i.e., EDGAR) over China in 2022. On a national scale, the high CH₄
327 emissions were primarily concentrated in the Yangtze River Delta (YRD), Pearl River Delta (PRD),
328 North China Plain (NCP), and Shanxi Province. For the YRD and PRD characterized by high-density
329 populations, a large number of industrial plants, as well as extensive agricultural activities such as rice
330 cultivation, contributed significantly to CH₄ emissions. The NCP, with its heavy reliance on coal-based
331 energy and large-scale livestock farming, was also expected to be a major emission source. Lower
332 emissions were mainly distributed across Northwest, and Southwest China.

333 Compared with the EDGAR, the posterior emissions were generally smaller over most areas of
334 mainland China, with total anthropogenic emissions decreasing to 45.1 TgCH₄·yr⁻¹, 36.5% lower than
335 the EDGAR (71.0 TgCH₄·yr⁻¹). Additionally, previous studies have consistently shown that the
336 EDGAR inventory overestimates China's CH₄ emissions. For example, through a Bayesian inversion
337 of CH₄ and stable isotope (δ¹³C-CH₄) measurements for East Asia, Thompson et al. (2015) found that
338 posterior values decreased significantly across eastern and southern China, especially in the
339 NCP. Overall, EDGAR overestimated China's emissions by 29%. Similarly, Zhang et al. (2021)
340 conducted a global inverse analysis using GOSAT observations and revealed that EDGAR significantly
341 overestimates anthropogenic emissions in eastern China. The posterior estimate of Chinese
342 anthropogenic emissions was 30 % lower than that of EDGAR. Turner et al. (2015) also demonstrated
343 that after assimilating GOSAT observations, China's posterior CH₄ emissions were revised downward
344 by 50% relative to EDGAR. Using a regional model to assimilate TROPOMI observations, Kou et al.
345 (2025) detected decreases of varying magnitudes across nearly the entirety of China, with the exception
346 of Northwest China. Similar results have been reported by Alexe et al. (2015), Pandey et al. (2016),



347 and Maasakkers et al. (2019). Our study, using a higher-resolution regional inversion method, provided
348 more detailed emission information. Overall, our inversion results were comparable to the ensemble
349 mean of GCP ground-based inversions (Wang et al., 2025).



350

351 **Figure 2** Spatial distribution of the annual total prior emissions (top, EDGAR v8, $\text{MgCH}_4 \cdot \text{km}^{-2} \cdot \text{yr}^{-1}$),
352 posterior emissions (middle), and differences (bottom, posterior minus prior) over (a-c) Mainland
353 China and (d-f) Shanxi Province.



Table 1 shows the comparison of posterior and prior anthropogenic CH₄ emissions of the main emission sectors in China. Consistent with the previous studies, we also found that North China is the region with the most significant reduction nationwide. Almost the entire region has experienced a decrease, with a reduction of 56.9%, indicating that there are indeed substantial systematic biases in the EDGAR inventory. Among them, the coal sector contributed the most to this discrepancy, followed by enteric fermentation, with decreases of 56.2% and 64.0%, respectively (Table 1).

For Northeast China, the increase in emissions mainly occurred in Heilongjiang Province, especially in the western part, which is the base of Daqing Oilfield, China's largest oilfield. The oil and gas sectors in the entire Northeast China increased by 27.1% and 23.1%, respectively. However, emissions from other sectors decreased, resulting in an overall 11.1% reduction in emissions. Rice paddy CH₄ emissions serve as the dominant emission source in East China. However, the high spatial heterogeneity and the insufficiency of data on rice cultivation introduce large uncertainties to inventories. Additionally, different fertilization management practices in various regions, such as the use of nitrogen fertilizer versus organic fertilizer (Zhang et al., 2022), and mid-season drainage management (Lin et al., 2021), bring about considerable uncertainties in the emission factors related to rice cultivation practices. Emissions in Zhejiang, Fujian, and Jiangxi Provinces increased, mainly attributed to emissions from rice paddies. In contrast, emissions in other provinces decreased, dominated by the coal mining, leading to an overall 26.4% reduction in emissions in East China. In Central and South China, the dominant sources of emissions remain rice paddies. However, the decreases were mainly attributed to the coal mining and wastewater treatment sectors, which reduced emissions by 51.3% and 12.4%, respectively. Overall, the total emissions in these two regions decreased by 21.8% and 7.8%, respectively. In Northwest China, the reduction in emissions was mainly distributed in the hotspot areas of coal mines, which dominated the overall 55.7% decrease in emissions. In Southwest China, the increase in emissions occurred predominantly in the southern part of the Sichuan Basin, a major coal-producing region, while emissions in other regions decreased. Sheng et al. (2019) found that the EDGAR inventory lacks a significant amount of statistics on coal mines. Moreover, China has issued an energy policy of “phasing out small coal mines” (approximately 50% of which are located in the southwest with high CH₄ content) to shift production towards lower-emission areas and consolidate into large coal mines (Zhang et al., 2022; Sheng et al., 2019).



383 Additionally, the wastewater treatment sector achieved significant emission reductions, driving a 1.7%
384 overall decrease in Southwest China.

385 **Table 1** Comparison of posterior and prior anthropogenic CH₄ emissions (GgCH₄·yr⁻¹) of the main
386 emission sectors in the 7 major regions of China. Waste includes wastewater treatment, solid waste
387 landfills, and solid waste incineration.

		Coal	Gas	Rice	Waste	Livestock	Building	Mature
North	Prior	13841.1	98.4	169.1	2462.1	3882.9	385.5	859.5
	Posterior	6068.4	46.2	116.4	1238.1	1396.4	206.7	279.6
Northeast	Prior	771.9	82.6	879.6	1427.9	251.9	268.4	36.8
	Posterior	665.0	101.8	821.2	1259.6	237.7	243.3	34.1
East	Prior	2257.3	54.4	5100.3	6681.4	907.9	732.9	200.5
	Posterior	1114.5	45.1	4592.4	4750.9	512.7	576.2	114.4
Central	Prior	846.5	32.4	3139.5	1882.9	434.2	399.4	91.2
	Posterior	412.2	25.7	2862.6	1324.4	343.3	314.1	75.2
South	Prior	15.0	20.5	2186.4	1622.5	280.2	254.6	44.0
	Posterior	13.8	18.9	2080.6	1421.3	269.9	235.0	42.2
Northwest	Prior	8188.3	1083.4	143.3	970.5	746.2	226.7	76.4
	Posterior	2899.1	645.8	120.7	593.6	534.3	164.4	39.6
Southwest	Prior	531.3	76.5	2046.4	1124.0	667.0	356.3	107.1
	Posterior	505.6	77.2	2063.0	1060.7	663.7	352.1	107.2

388 China accounts for as high as 69% of the global mitigation potential in coal mining in 2030 (EPA,
389 2019). Shanxi Province, where 94% of the emissions come from the coal mining sector, counts for
390 nearly one-third of the country's total CH₄ emissions. Therefore, we conducted a focused analysis on
391 the emissions in Shanxi Province with high-resolution (9 km) inversion, which can better diagnose the
392 spatial characteristics of emissions caused by coal mining (Figure 2d-f). The optimized emissions
393 showed a decrease in the vast majority of areas in Shanxi Province. Compared with the EDGAR
394 inventory (9.0 TgCH₄·yr⁻¹), the posterior emissions decreased by 46.3% to 4.8 TgCH₄·yr⁻¹. The
395 overestimation of coal mining emissions may be due to the fact that the standard IPCC emission factors
396 used by EDGAR are too high for Chinese coal mines (Lin et al., 2021). Additionally, the average



emission factors in the north are lower than those in other regions (Gao et al., 2021). The use of a uniform regional emission factor by EDGAR further exacerbates the overestimation of emissions (Shi et al., 2025). Moreover, in the past decade, the extraction and utilization of coal mine CH₄ in China have been largely improved (Lu et al., 2021), but the recovery of coal mine CH₄ is not adequately considered in the EDGAR inventory. Finally, the CH₄ emission intensity of surface mining is ten times lower than that of underground mining. However, surface mining is overlooked in the EDGAR inventory (Gao et al., 2020). In Yangquan, Jincheng, Changzhi, and Jinzhong cities (Figure 2), some increased emission hotspots caused by the coal mining activities could be found. Coincidentally, these cities are the main coal-producing areas in Shanxi Province. Overall, the emissions in these cities have still decreased by 3.3%, 4.2%, 29.5%, and 32.3%, respectively (Figure S3). Tu et al. (2024) utilized TROPOMI observations and implemented a wind-assigned anomaly method to quantify the CH₄ emissions from coal mines in Yangquan, Changzhi, and Jincheng. Compared to EDGAR, emissions decreased by 56.5%, 40.5%, and 65.0%, respectively—a larger reduction than our findings.

3.2 Monthly variations

Figure 3 illustrates the comparison of the monthly variations in prior and posterior emissions both in China and Shanxi province. Nationally, influenced by CH₄ emissions from paddy fields, both the prior and posterior emissions exhibited the characteristic of being low in winter and high in summer. However, the posterior emissions for each month were lower than the prior emissions. The monthly posterior emissions ranged from 2.3 to 7.5 Tg·month⁻¹, with a modification rate ranging from -30.7% to 5.6% compared to the prior emissions (4.8–7.6 Tg·month⁻¹). The month with the smallest difference occurred in August, mainly because the underestimation of CH₄ emissions from paddy fields compensated for the overestimation of emissions from coal mining (Figure 3c). Additionally, due to the weather conditions such as clouds and rain in summer, the amount of TROPOMI data was significantly smaller than that in other seasons (Figure 1b), which might lead to insufficient constraints on emissions. Regarding Shanxi Province, even with an agnostic flat monthly prior (Figure 3b), our estimates generated a monthly variation, and emissions in all months were lower than the prior values. We also detected a decrease and subsequent increase in emissions that correspond to the Spring Festival in February. A noticeable recovery of production capacity in the following two months was also evident. Overall, although relatively large uncertainties were introduced by the amount of



426 observations, the seasonal variation of monthly posterior emissions could be roughly captured.

427 To gain a deeper understanding of the posterior emissions, the assimilated estimations were allocated
428 to different emission sectors, based on the monthly prior proportions at the model grid points. The
429 sectoral patterns offer insights into the underlying factors influencing China's emission changes.
430 Consequently, we further concentrated on interpreting the emissions from the coal, gas, rice cultivation,
431 waste, livestock, building, and manure management sectors, which are the most significant sectors in
432 China (Table S3).

433 It could be found that the reduction in emissions each month in China was mainly dominated by the
434 coal sector, with an overall annual reduction of 55.9%, indicating a high level of uncertainty in the
435 prior emissions. Moreover, the gas, livestock, and manure management sectors also showed varying
436 degrees of reduction in different months, with overall decreases of 34.0%, 45.1%, and 51.5%,
437 respectively. However, for the rice cultivation sector, the posterior emissions were lower than the prior
438 emissions in the first half of the year, while the situation was reversed in the second half of the
439 year. Summer and autumn are the heading and flowering stages for both single-cropping late rice and
440 double-cropping late rice across the eastern, central, and southern China. During these periods, the
441 emissions from paddy fields tended to be higher than in other seasons. Overall, the posterior CH₄
442 emissions from paddy fields have slightly decreased by 4.7%. A similar situation was observed in the
443 waste sector. Previous studies have shown that significant CH₄ production in wastewater is more likely
444 to occur because methanogens become more active as the temperature rises (Hu et al., 2023). Therefore,
445 a significant increase in the posterior emissions compared to the prior emissions could be observed in
446 August. For the building sector, due to winter heating or hot water supply, the emissions in winter were
447 approximately three times higher than those in summer. After optimization, the emissions in winter
448 and spring significantly decreased, while in some months of summer and autumn, the decrease was
449 less pronounced, and there was even a slight increase. The inter-monthly difference in emissions
450 decreased, but the monthly variation remained significant. Overall, the posterior emissions have
451 decreased by 25.4%. The changes in emissions in Shanxi Province were mainly dominated by the coal
452 mining sector. Compared to EDGAR, the monthly reduction rate ranged from 12.6% in August to 67.3%
453 in February. For other sectors, there was little overall monthly variation.

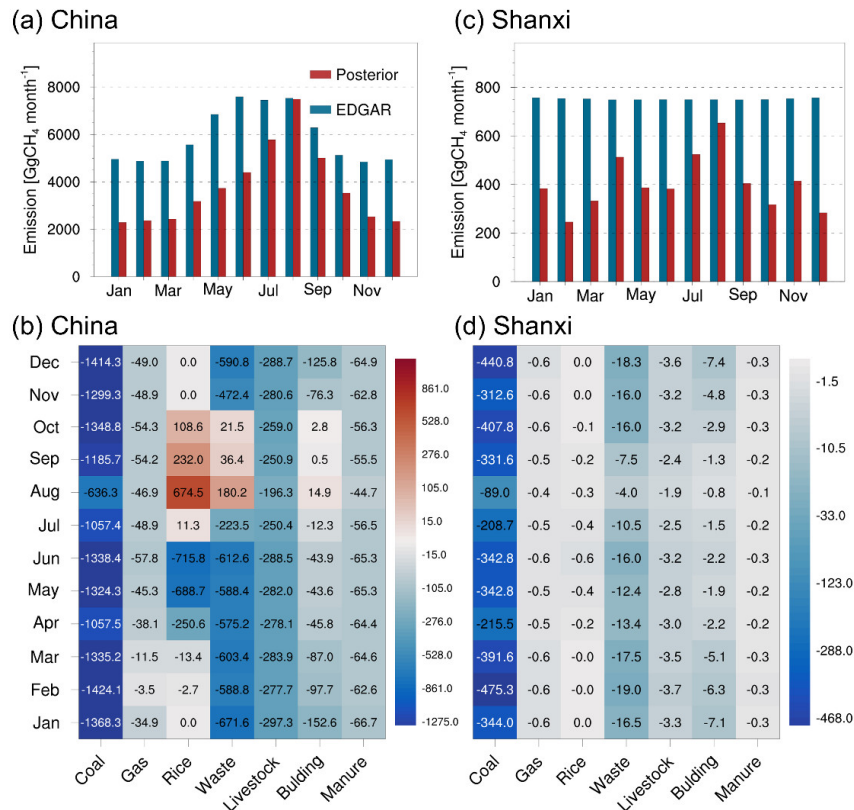


Figure 3 Comparison of the monthly variations and the sectoral differences ($\text{GgCH}_4 \cdot \text{mon}^{-1}$) in prior and posterior CH_4 emissions over (a, b) China and (c, d) Shanxi Province.

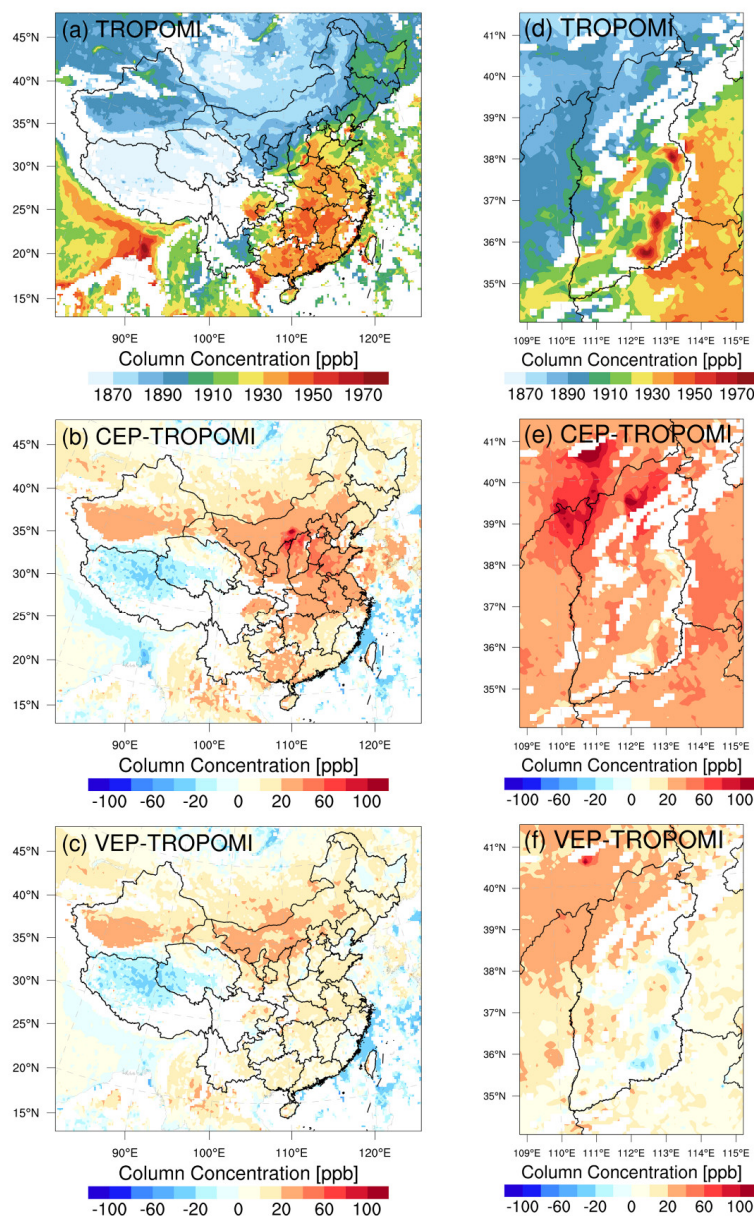
3.3 Evaluation of posterior emission estimates

Figure 4 shows the spatial distribution of XCH_4 in the posterior simulation, as well as the validation of XCH_4 simulated by prior and posterior emissions with TROPOMI observations. It could be observed that relatively large XCH_4 were present over eastern China, which was driven by coal mining in the northern part and rice paddy fields in the southern part (Zhang et al., 2023). The simulation using prior emissions significantly overestimated the XCH_4 concentration in China, especially in the NCP. The maximum overestimation exceeded 100 ppb, which was consistent with the overestimated emissions in these regions (Figure 2). After the inversion optimization, the simulated XCH_4 showed better spatial distribution consistency with the TROPOMI observations. The vast majority of the biases were within 20 ppb, and the national average biases decreased from 14.7 ppb to 7.9 ppb, representing



467 a 46.0% reduction. In Shanxi Province, high emissions were mainly concentrated in the coal-rich
468 southeastern region, with an annual average maximum exceeding 1970 ppb, which was difficult to
469 reproduce the characteristics of local high concentrations at a coarse resolution. The prior simulation
470 showed obvious overestimation across the entire region, particularly in northern Shanxi. Due to the
471 high spatial resolution adopted, the model could take into account more practical factors, and thus
472 capture the spatiotemporal variations of emission sources and transport processes more precisely. As
473 a result, the agreement between the posterior simulations and the TROPOMI observations was
474 remarkably enhanced. The average deviation decreased from 37.9 ppb to 11.9 ppb, with a reduction of
475 68.7%. These changes not only demonstrate the effectiveness of reducing uncertainties in optimizing
476 emission sources but also imply that there are indeed significant uncertainties in the prior emission
477 inventory.

478



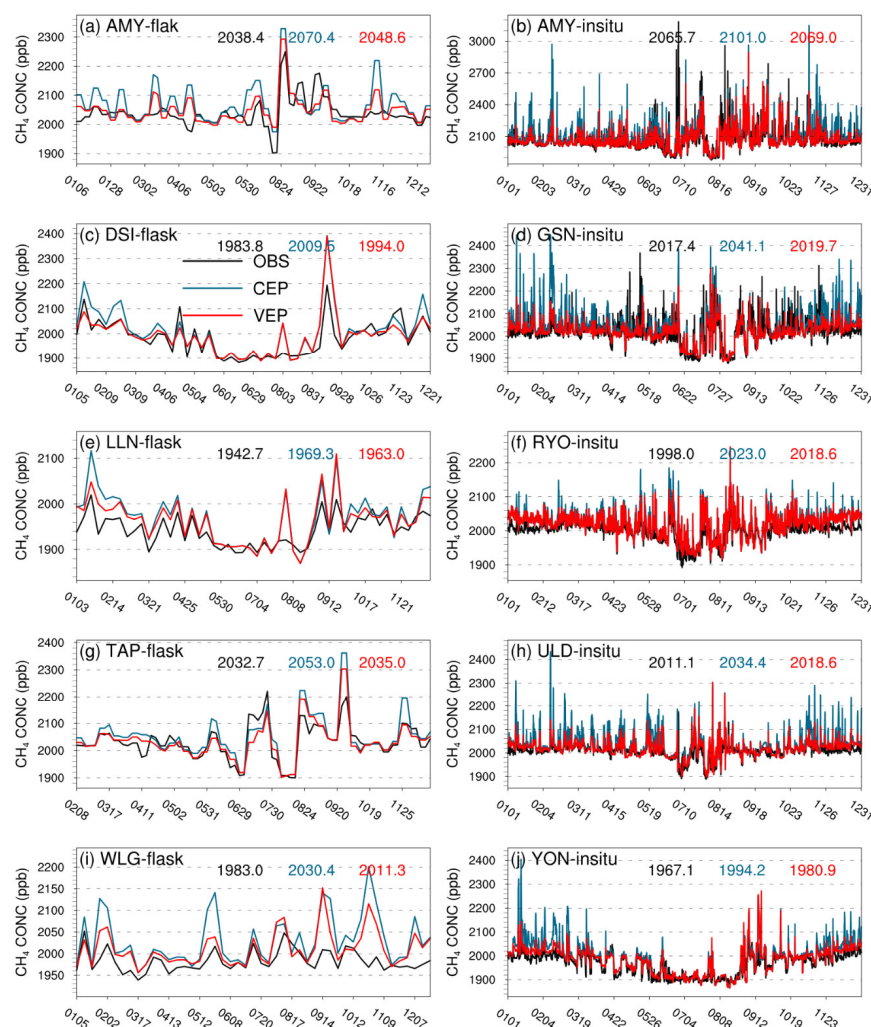
479

480 **Figure 4** Comparison of simulated XCH₄ (ppb) from prior and posterior emissions with TROPOMI
 481 observations over (a-c) China and (d-f) Shanxi Province. (a, d) Spatial distribution of XCH₄ observed
 482 by TROPOMI. (b, e) Differences between prior simulations and observations. (c, f) Differences
 483 between posterior simulations and observations.



484 We further conducted an independent evaluation of the posterior estimate by comparing it with 10 *in-*
485 *situ* and flask monitoring sites from the CH₄ ObsPack v6.0 database. Figure 5 shows the time series
486 comparison of the CEP and VEP experiments with the observations. The evaluation statistics for all
487 sites are presented in Table S4. There was a significant overestimation in the CEP experiment
488 throughout most of the study period. Although most of the sites are located in the eastern regions
489 outside China, influenced by the atmospheric circulation, the westerlies prevail in most parts of
490 China. Thus, the optimized emission information can be effectively reflected in the concentrations
491 observed downwind. The optimized simulation in VEP experiments was more consistent with the
492 observations. Among them, the AMY flask site showed the largest reduction in bias, with the average
493 bias decreasing from 35.3 ppb to 3.3 ppb, a decrease of 90.7%. For WLG, the only ObsPack site
494 available in mainland China, the average bias decreased from 47.4 ppb to 28.3 ppb, a reduction of
495 40.3%. This indicates that the optimized emissions can significantly improve the CH₄ simulation,
496 whether in areas close to the source or downwind of the source. Overall, the average bias of the 10
497 sites decreased by 58.6%, from 28.6 ppb to 11.9 ppb, the root mean square error (RMSE) decreased
498 by 28.9%, from 56.2 ppb to 39.9 ppb, and the correlation coefficient (CORR) increased from 0.72 to
499 0.75. These evaluation results demonstrate that the inversion effectively reduces the uncertainties of
500 prior emission inventory.

501



502

503 **Figure 5** Time series comparison of surface CH₄ concentrations (ppb) from prior (CEP) and posterior
 504 (VEP) emission simulations with observations from 5 flask sites and 5 *in-situ* sites within the ObsPack
 505 dataset. The black, blue, and red values represent the averaged observations, prior simulations, and
 506 posterior simulations, respectively.

507 We also carried out an independent validation using six high-precision *in-situ* observation sites within
 508 Shanxi Province (Figure 1a). Table 2 shows the bias, RMSE, and CORR of the CEP and VEP
 509 simulations against these surface observations. Except for the LF site, where the VEP experiment
 510 exhibited an underestimation, the VEP experiment demonstrated varying degrees of improvement at



the other five sites. Notably, the bias reduction at the DT and WTS sites exceeded 93.3%. Especially for the WTS site, a high mountain site with an altitude of over 2208 m, the bias was decreased to 8.3 ppb. On average across all sites, the bias significantly decreased by 96.0%, from 535.1 ppb to 21.4 ppb. For the RMSE, the most significant decreases were observed at the DT and SZ sites, both exceeding 73%. Overall, the RMSE reduction ranged from 13.5% to 79.0%. Additionally, the CORR of the VEP experiment increased to 0.48–0.63. In addition, we evaluated the CH₄ concentration simulation for the afternoon, during which the model generally demonstrates better boundary layer simulation performance, with overall lower bias and higher CORR (Table S5). These results further confirm that the RegGCAS-CH₄ system can effectively capture the characteristics of high-resolution CH₄ emission changes and improve the accuracy of concentration simulations.

Table 2 Statistics comparing the daily average CH₄ concentrations (ppb) from the simulations with prior (CEP) and posterior (VEP) emissions against six independent surface *in-situ* observation sites in Shanxi Province, respectively. The numbers under the site names represent the number of valid observations.

Site Name	Mean Obs.	Mean Sim.		BIAS		RMSE		CORR	
		CEP	VEP	CEP	VEP	CEP	VEP	CEP	VEP
TY (351)	2595.5	2771.0	2353.2	175.4	-242.3	408.7	353.6	0.46	0.51
DT (286)	2122.0	2441.1	2139.2	319.1	17.2	489.0	102.6	0.49	0.58
LF (277)	2395.1	2397.2	2197.9	2.0	-197.3	186.8	257.3	0.45	0.35
SZ (359)	2148.8	4560.5	2782.4	2411.7	633.6	2778.1	745.1	0.44	0.48
JC (362)	2446.0	2624.5	2355.2	178.4	-90.9	385.0	356.3	0.52	0.52
WTS (361)	2064.0	2188.0	2072.2	124.1	8.3	162.6	59.1	0.49	0.63

* BIAS, mean bias; RMSE, root mean square error; CORR, correlation coefficient

4. Discussion

To further explore the characteristics of our posterior emissions and offer valuable guidance for the

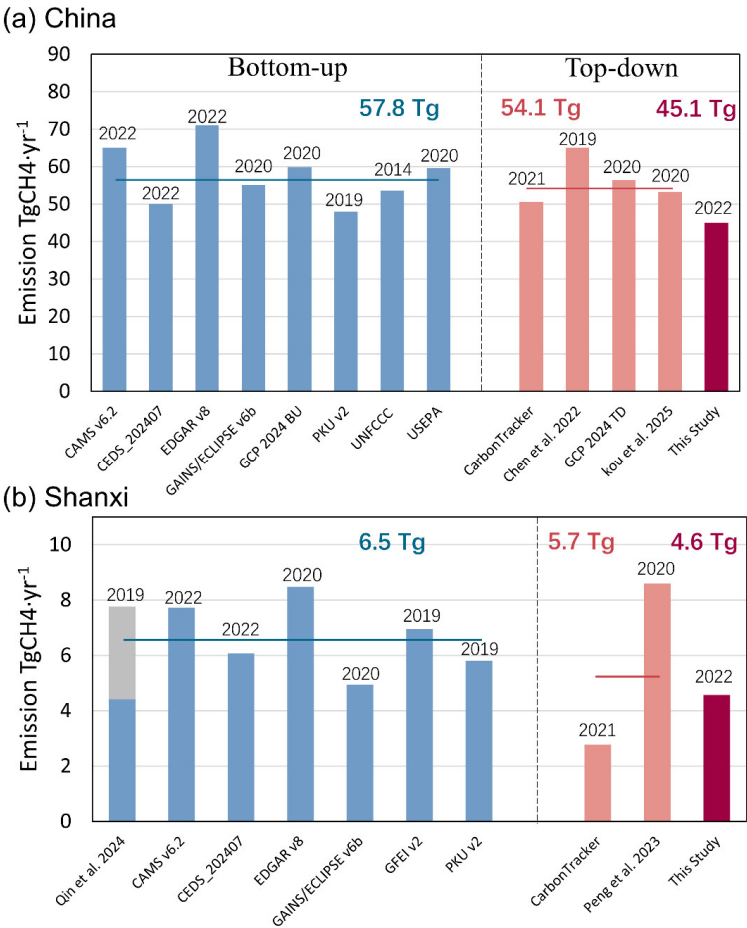


528 refinement of bottom-up inventory in China, we conducted a comparison analysis with the latest 8
529 bottom-up inventories and 4 top-down emission estimates (Figure 6). Specifically, apart from EDGAR
530 v8, the bottom-up inventories were sourced from the Copernicus Atmosphere Monitoring Service
531 (CAMS v6.2) (Soulie et al., 2024), Community Emissions Data System (CEDS_202407) (Hoesly et
532 al., 2018), Evaluating the Climate and Air Quality Impacts of Short-Lived Pollutants inventory
533 (ECLIPSE v6b) from the GAINS model (Stohl et al., 2015), Global Carbon Project (GCP 2024)
534 (Saunois et al., 2025), Peking University (PKU v2) (Liu et al., 2021), United Nations Framework
535 Convention on Climate Change (UNFCCC, 2020), and the United States Environmental Protection
536 Agency (EPA, 2019). In addition, we also obtained the Global Fuel Exploitation Inventory (GFEI
537 v2) (Scarpelli et al., 2022) to assess the emissions from the coal mining sector in Shanxi Province. The
538 top-down emissions were mainly sourced from the CarbonTracker-CH₄ (Oh et al., 2023) and the
539 research results of Chen et al. (2022), Kou et al. (2025), and Peng et al. (2023).

540 Typically, the top-down estimates showed lower emissions than the bottom-up inventories. Although
541 our posterior estimates were the lowest across all datasets, they closely matched CEDS, PKU, and
542 CarbonTracker-CH₄ and comparable to the ensemble mean of GCP ground-based inversions (Wang et
543 al., 2025). Overall, our posterior emissions were 22.0% lower than the average of bottom-up
544 inventories and 16.6% lower than the previous top-down estimates. The lower emissions in this study
545 were predominantly driven by the downward revision of coal emissions. Overestimated emission
546 factors and the difficulty in tracking the spatial distribution of coal mines due to mine closures and
547 regional transfers remain significant obstacles to the assessment of coal mine emissions (Gao et al.,
548 2021). Our estimate was lower than the previous top-down studies (53–65 Tg yr⁻¹), likely because
549 those previous studies were conducted at much coarser resolutions (0.3°–2.5° versus 27 km) or with
550 much sparser observations (*in-situ* ground measurements and GOSAT versus TROPOMI). For Shanxi
551 Province, compared with the bottom-up inventories, our results were close to those of GAINS and fell
552 within the uncertainty range of Qin et al. (2024), which set multiple groups of emission factors
553 according to coal types, mining methods, and geological structures. There were substantial disparities
554 among the top-down inversions. Notably, for CarbonTracker-CH₄, despite having minimal divergence
555 at the national scale when contrasted with other estimations, the values for Shanxi Province were
556 markedly lower than those of other estimates. This could be attributed to the insufficient assimilation



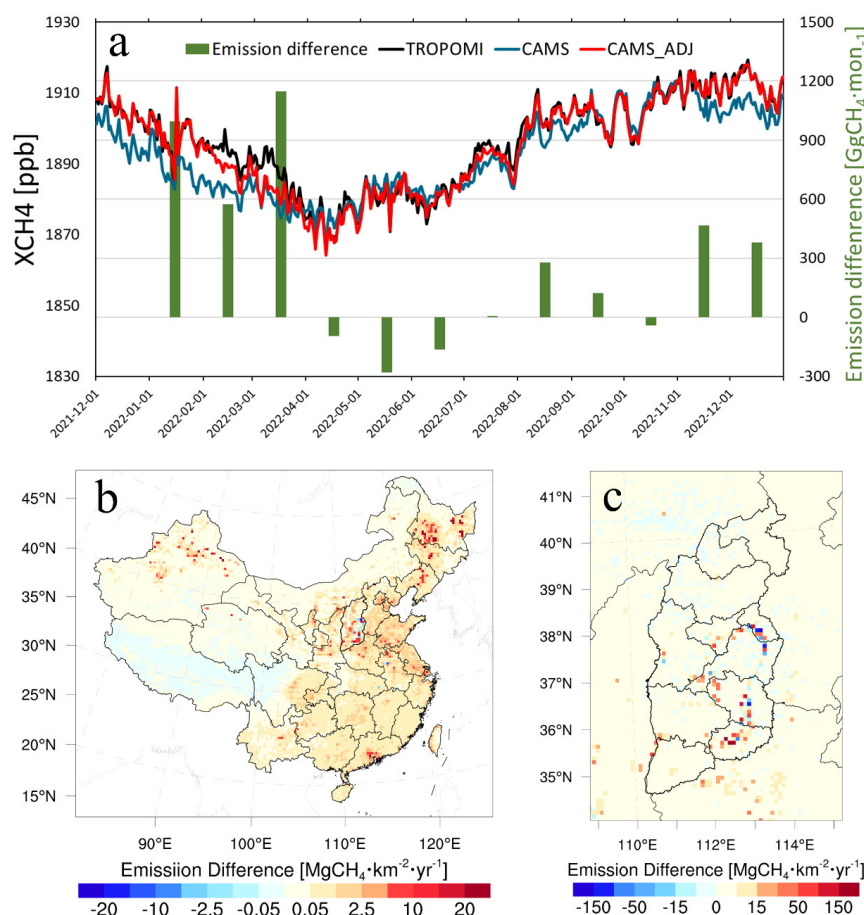
557 of China's surface observations and systematic biases in the spatial distribution of prior emissions.
558 Overall, the coal-mining emissions in this study were 29.2% lower than the bottom-up inventories and
559 19.3% lower than the previous top-down inversions.



560
561 **Figure 6** Total anthropogenic CH₄ emissions (TgCH₄·yr⁻¹) in China (a) and CH₄ emissions from the
562 coal mining sector in Shanxi Province (b). GCP 2024 BU and GCP 2024 TD represent the bottom-up
563 and top-down ensemble means included in the dataset, respectively. The grey bars in panel b represent
564 the range of CH₄ emissions estimated by Qin et al. (2024) using four different methods. The latest year
565 in the inventory is marked above the bars.
566 Uncertainties in boundary conditions constitute a significant source of error in regional inversion.
567 Despite the optimization of concentration fields in the CAMS, significant biases remained evident



(Figure 7a). To address this issue, we implemented additional constraints on the CAMS concentration fields using TROPOMI XCH₄ observations, resulting in a notably improved agreement with the observations. Specifically, the average bias decreased from -3.7 ppb to -1.0 ppb, representing a reduction of 73.7%. A sensitivity experiment (SENS) was further conducted, where the unadjusted CAMS global fields were extracted as boundary conditions to invert anthropogenic CH₄ emissions, aiming to evaluate the impact of boundary condition uncertainties on regional emission inversions. Compared with the base experiment (BASE), the largest discrepancies in monthly variations were observed during the winter months, indicating significant overestimations. The emission differences across different months closely aligned with the concentration differences, suggesting that underestimations in concentrations prompted more substantial emission adjustments for compensation. Regarding the spatial distribution, due to the relative long lifetime of CH₄, emission changes caused by boundary condition biases were not confined to regional boundaries, such as in Northeast and Northwest China, but were also observable throughout the entire region. In the outer national inversion, the average CH₄ emissions increased by 5.1%, while in the inner nested inversion, the increase was 4.3%. This highlights the necessity of expanding the inversion scope to mitigate the influence of boundary condition errors on the inversion of central regions. Furthermore, by comparing the difference in CH₄ emissions over the inner domain between two outer national inversions with that between two inner nested regional inversions (BASE and SENS), we found that the two inner regional inversions exhibited a 23.1% smaller discrepancy. This indicates that adopting a nested inversion approach is essential, which can further reduce boundary errors during inner inversion through outer optimization, thereby enhancing the robustness of the inversion process.



589

590 **Figure 7** Differences in XCH₄ column concentrations (ppb) between TROPOMI-adjusted and
 591 unadjusted CAMS fields, along with the induced monthly posterior CH₄ emission differences (SENS
 592 - BASE) over China (a); differences in two posterior emissions (MgCH₄·km⁻²·yr⁻¹) derived from
 593 unadjusted and adjusted boundaries over China (b) and Shanxi Province (c)

594 Our inversion results are generally lower than previous emission estimates. By comparing the
 595 differences in emissions from the inner region at different resolutions, we found that in both the BASE
 596 and SENS experiments, emissions inverted in the inner region (9 km) were typically 7.6% and 8.4%
 597 lower than those in the outer region (27 km), respectively. This discrepancy likely arises because
 598 higher-resolution simulations excel at capturing localized emission hotspots that lead to elevated
 599 concentration values. Therefore, compared to other coarse-resolution inversions, this study tends to



adjust emissions downward to better align with observations. Additionally, China's 2022 COVID-19 restrictions, the most stringent since 2020, may be another factor driving lower emissions due to production shutdowns and home quarantine. Peng et al. (2022) combined bottom-up and top-down approaches to quantify CH₄ source changes, revealing a 1.2 Tg·CH₄ yr⁻¹ reduction in anthropogenic CH₄ emissions in 2020 compared to 2019. Taking oil and gas sector as an example, an IEA report stated that global CH₄ emissions from oil and gas operations decreased by approximately 10% year-on-year in 2020 (IEA, 2021), primarily attributed to reduced oil and gas production during the pandemic (Thorpe et al., 2023).

5 Summary and conclusions

In this study, we developed a Regional Methane Assimilation System (RegGCAS-CH₄) based on the WRF-CMAQ model and EnKF algorithm to perform high-resolution nested inversions of China's CH₄ emissions using TROPOMI XCH₄ retrievals. Taking the EDGAR v8 as prior emissions, the daily anthropogenic CH₄ emissions over China in 2022 were inferred.

Using TROPOMI XCH₄ as constraints, we revealed significant overestimations in the prior emissions. Nationally, the posterior anthropogenic CH₄ emissions in 2022 were estimated to be 45.1 TgCH₄·yr⁻¹, 36.5% lower than the EDGAR inventory, mainly driven by the decreases in the coal and waste sectors. North China experienced the most substantial reduction, with a 56.9% decrease, mainly due to downward revisions in the coal and enteric fermentation sectors. In Shanxi Province, influenced by dominant sources of coal mining, the posterior emissions decreased by 46.3% compared to the EDGAR inventory, with value of 4.8 TgCH₄·yr⁻¹. The monthly variation of posterior emissions showed a pattern of being low in winter and high in summer. Except for the rice cultivation and waste sectors, which showed some increases in certain months related to seasonal growth stages and temperature-driven methanogen activity, respectively, other sectors exhibited varying degrees of reduction. Observation-constrained emissions significantly improved the performance of simulations regarding XCH₄ columns, as well as surface CH₄ in both the entire region and Shanxi Province, reducing biases by 46.0%, 58.6%, and 96.0%, respectively. This highlights the effectiveness of RegGCAS-CH₄ in reducing uncertainty in CH₄ emissions. Sensitivity inversions highlight the importance of high-resolution satellite-based nested inversions in accurately estimating CH₄ emissions, especially in reducing the impact of boundary condition errors on emission inversions.



629 Top-down inversions usually result in lower emissions compared to those constructed by bottom-up
630 methods. When comparing our posterior emissions with other inventories, our estimates were the
631 lowest among both bottom-up and top-down estimations. On average, they were 22.0% lower than
632 bottom-up inventories and 16.6% lower than previous top-down estimates. The high-resolution
633 simulations employed in this study are capable of more accurately simulating local high concentration
634 values, and thus, the reduction in emissions is greater than that in previous studies. The overestimation
635 of China's CH₄ emissions may have led to an overestimation of the country's climate change mitigation
636 burden. The more accurate emission estimates presented here not only enhance our understanding of
637 the CH₄ budget but also contribute to more effective global climate change mitigation efforts.

638

639 **Data availability**

640 The TROPOMI XCH₄ data is publicly available for download at the Copernicus Data Space Ecosystem
641 (<https://browser.dataspace.copernicus.eu/>, last access: August 5, 2024). The observations used for
642 evaluation and the posterior emissions and simulations produced in this study can be accessed at
643 <https://doi.org/10.5281/zenodo.15602944> (Feng, 2025).

644

645 **Author contribution**

646 SF, FJ and YZ conceived and designed the research. SF developed the RegGCAS-CH₄ system,
647 analyzed data, and prepared the paper. SW evaluated the CH₄ simulation in Shanxi Province. HC, HZ,
648 SB, HW, and WJ reviewed and commented on the paper.

649

650 **Competing interests**

651 At least one of the (co-)authors is a member of the editorial board of Atmospheric Chemistry and
652 Physics.

653

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663

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