

1 **Drivers and vertical CO₂ flux balances in a Sahelian agroforestry system: Insights from**
2 **high frequency measurements**

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28 **Highlights:**

- 29
- Long-term high frequency CO₂ flux measurements using automated dynamic chambers in
30 a Sahelian *F. albida* parkland.
 - Empirical gap-filling and flux partitioning methods validated against Eddy Covariance
31 GPP.
 - Fluxes peaked during the rainy season both at a distance from trees in full sun (FS) and
32 under tree canopies (Sh), driven mainly by soil moisture and leaf area.
 - *F. albida* trees enhance CO₂ fluxes under canopies ("fertile island" effect) and account for
33 ~23% of annual ecosystem GPP.
- 34
35
36

37 **ABSTRACT:**

38 Agroforestry systems — combining trees with crops and/or livestock — are increasingly
39 promoted as sustainable and climate-resilient land-use strategies. Despite their widespread
40 presence in the Sahel, experimental data on their potential as carbon sinks are scarce. This study
41 presents a full-year, high-frequency dataset of CO₂ fluxes in a Sahelian agro-silvo-pastoral
42 parkland dominated by *Faidherbia albida*, located in Senegal's groundnut basin. CO₂ fluxes were
43 continuously measured using automated dynamic chambers, allowing the quantification of soil
44 and crop respiration (R_{ch}), gross primary production (GPP_{ch}), and net carbon exchange (FCO_{2ch})
45 under both full sun and shaded (under tree canopies) environments. Seasonal patterns of CO₂
46 fluxes were similar in both environments, with peaks during the rainy season. R_{ch} and GPP_{ch}
47 were significantly higher under tree canopies, indicating a 'fertile island' effect. CO₂ flux variability
48 was primarily driven by soil moisture and leaf area index. Chamber-based GPP estimates closely
49 matched those from Eddy Covariance measurements. On an annual scale, *F. albida* trees
50 contributed approximately 23% of total ecosystem GPP, with a carbon use efficiency of 0.48. Net
51 annual vertical CO₂ exchange was estimated at -1.4 ± 0.46 and -1.8 ± 0.17 Mg C-CO₂ ha⁻¹ using
52 chamber and Eddy Covariance methods, respectively. These findings underscore the role of *F.*
53 *albida*-based agroforestry systems as effective carbon sinks in Sahelian landscapes, supporting
54 their potential contribution to climate change mitigation.

55 **Keywords:** Sahelian agro-silvo-pastoral systems, CO₂ fluxes, automated dynamic chambers, Eddy
56 Covariance, 'fertile island effect', carbon balances.

57 1. Introduction

58 Plant photosynthesis and respiration —both autotrophic (plant) and heterotrophic (microbial)—
59 are fundamental processes driving carbon dioxide (CO₂) fluxes in terrestrial ecosystems
60 (Lambers et al., 2008; Raich et al., 2014; Reichle, 2020). Accurate quantification of these processes
61 is critical for assessing ecosystem carbon (C) sink potential (Baldocchi, 2020), particularly for
62 informing climate-smart land management strategies.

63 To capture these processes at the ecosystem scale, the Eddy Covariance (EC) technique has
64 emerged as a transformative method, enabling continuous and high-frequency CO₂ flux
65 measurements (Baldocchi, 2003, 2008). The EC technique quantifies CO₂ exchanges between
66 ecosystems and the atmosphere by correlating fluctuations in vertical wind velocity with
67 simultaneous variations in CO₂ concentrations, providing a direct and non-invasive estimate of
68 CO₂ fluxes (Baldocchi, 2003). Extensive EC networks in Europe (Stojanović et al., 2024), Asia (Yu
69 et al., 2011), and the Americas (Chu et al., 2021) have significantly advanced our understanding
70 of the global C cycle. In contrast, sub-Saharan Africa remains critically underrepresented
71 (Bombelli et al., 2009; Houghton & Hackler, 2006; Williams et al., 2007). Although some studies
72 have used EC (Ardö et al., 2008; Brümmer et al., 2008; Merbold et al., 2009; Tagesson et al., 2016),
73 static chambers (Assouma et al., 2017; Owusu et al., 2024; Rosenstock et al., 2016; Wachiye et al.,
74 2020), or modeling approaches (Agbohessou et al., 2023, 2024; Delon et al., 2019; Rahimi et al.,
75 2021), they remain sparse and methodologically heterogeneous, limiting comparability and
76 regional C budget integration.

77 Among these underrepresented landscapes, agroforestry systems in the Sahel— particularly
78 agro-silvo-pastoral systems (ASPS) that combine trees, crops, and livestock— are increasingly
79 promoted for sustainable land management and climate resilience (Cardinael et al., 2021; Gupta
80 et al., 2023; Mbow et al., 2014; Stetter & Sauer, 2024). However, the structural and functional
81 heterogeneity of these systems poses significant challenges for accurately quantifying and
82 upscaling C fluxes. *Faidherbia albida*, a keystone agroforestry tree species in these ASPs (Leroux
83 et al., 2022; Lu et al., 2022), is of particular interest due to its reverse phenology, capacity to
84 enhance soil fertility and crop yields (Bayala et al., 2020; Roupsard et al., 2020; Sileshi et al., 2016;
85 2020). Yet, its functional role in modulating both the magnitude and seasonal dynamics of CO₂
86 fluxes remains poorly understood.

87 Addressing this knowledge gap requires integrated approaches capable of capturing both
88 aggregate and component-specific CO₂ fluxes. While EC remains the gold standard method for CO₂
89 flux measurements at the landscape scale (Baldocchi, 2003), it captures net ecosystem exchange
90 (NEE) as an aggregate signal, without separating the contributions from individual compartments
91 such as soil, crops, and trees. This limits its utility for disentangling processes and attributing
92 sources in heterogeneous systems like ASPs. Automated dynamic chambers provide a valuable

93 complement to EC, as they enable continuous, high-frequency measurements at finer scales and
94 at the level of specific ecosystem components. This approach facilitates component-specific
95 quantification of CO₂ fluxes, particularly from soil and crop compartments (Luo & Zhou, 2006;
96 Denmead, 2008; Zaman et al., 2021). When combined with EC, this dual-method approach
97 strengthens source attribution and improves the partitioning of fluxes across complex
98 agroforestry landscapes.

99 This study presents one of the first integrated quantification of CO₂ fluxes in a Sahelian ASPS
100 dominated by *F. albida*, combining EC and automated dynamic chambers.

101 Specifically, the study aims to (1) conduct year-round, high-frequency *in situ* CO₂ flux
102 measurements from soil and crops using automated dynamic chambers; (2) partition the net CO₂
103 fluxes (FCO₂ch) into respiration (Rch) and photosynthesis (GPPch); (3) investigate the
104 environmental drivers of fluxes and the spatial variability linked to tree presence; and (4)
105 compare chamber-based flux estimates with ecosystem-scale measurements derived from the EC
106 method.

107 Based on these objectives, we hypothesize that (1) Rch and GPPch are higher under the canopy of
108 *F. albida* than in full sun, (2) soil moisture is the main environmental factor directly controlling
109 both Rch and GPPch, (3) when extrapolated to the field scale, the chamber-based method provides
110 seasonal dynamics of respiration and photosynthesis fluxes comparable to those derived from EC
111 technique.

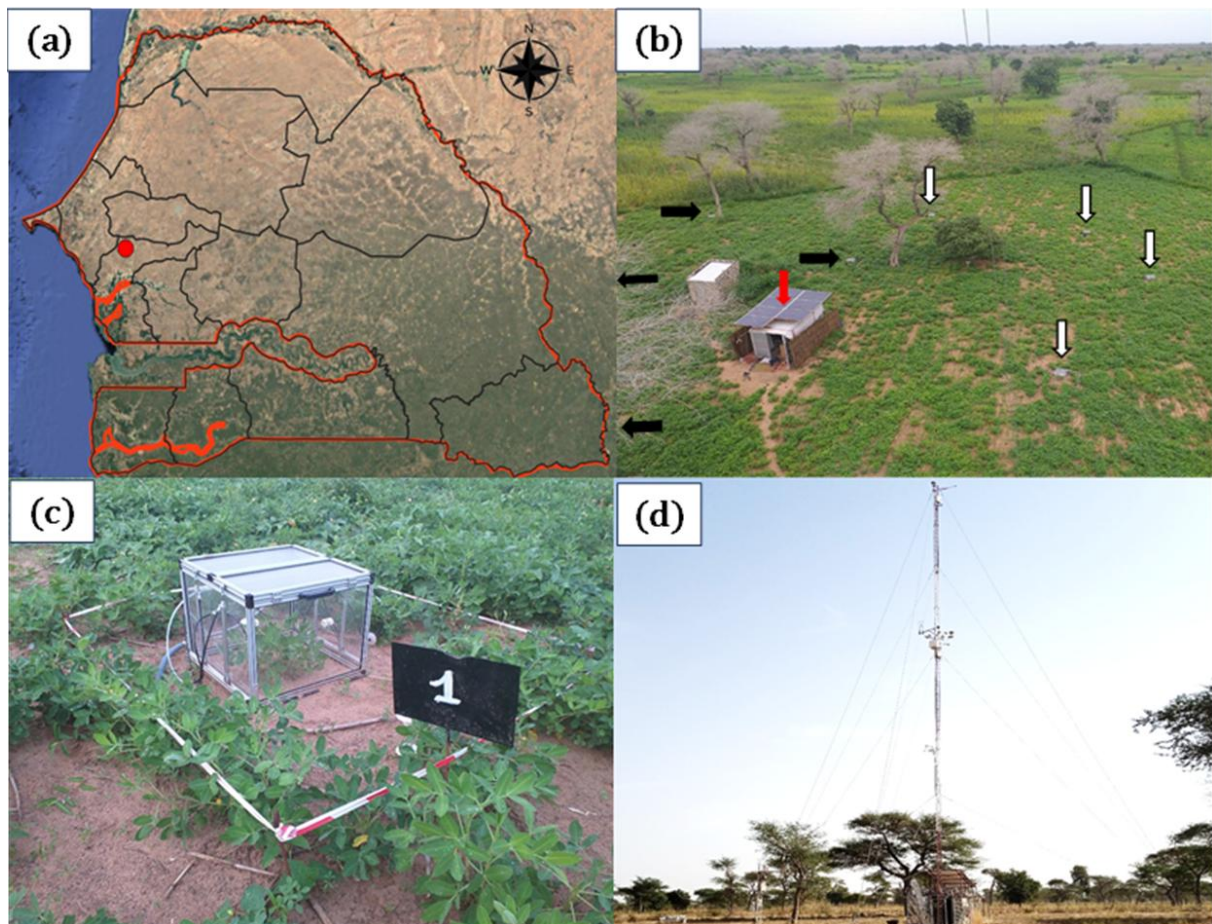
112 2. Materials and methods

113 2.1. Site description

114 The study was conducted in the agroforestry parkland of Sob village (Niakhar municipality, Fatick
115 region), located in the groundnut basin of Senegal, within the Sahelo-Sudanian climatic zone of
116 West Africa (Fig. 1). The climate is characterized by a long dry season (8–9 months) with high
117 temperatures and strong diurnal variations, and a short rainy season from late June to mid-
118 October (Delaunay et al., 2018).

119 Soils are locally known as “*Dior*” and classified as Arenosols (IUSS Working Group WRB, 2022).
120 The topsoil has low organic matter (<1%) and phosphorus (<3 mg kg⁻¹), a sandy texture (>85%
121 sand), and an acidic pH (Malou et al., 2021; Siegwart et al., 2022). Rainfed agriculture
122 predominates. The main cropping system includes pearl millet (*Pennisetum glaucum L.*) and
123 groundnut (*Arachis hypogaea L.*) in biennial rotation, with occasional intercropping of cowpea
124 (*Vigna unguiculata L.*).

125 The site hosts the 'Faidherbia Flux' station (14°29'44.916"N; 16°27'12.851"W; FLUXNET ID: SN-
126 Nkr), a long-term research platform for monitoring ecosystem services in agroforestry systems.
127 It is dominated by *F. albida*, a nitrogen-fixing, reverse-phenology tree with deep roots accessing
128 groundwater (Roupsard et al., 1999). The tree density is ~13 trees ha⁻¹, with canopies covering
129 ~10% of the soil surface (Roupsard et al., 2020). The EC tower is installed at 20 m height,
130 approximately 12.5 m above the canopy. The study field is a typical 'bush field', characterized by
131 low soil fertility, no mineral fertilization, and off-site export of crop residues and manure (Malou
132 et al., 2021).



133 Fig. 1: Study area.

134 (a) geographical location of Sob, Groundnut basin, Senegal (Map data © Google Earth, 2025), (b) overview
 135 (image from the Eddy Covariance tower located in the same bush-field) of the *Faidherbia albida* parkland
 136 during the rainy season, depicting groundnut crops with bare soil in the inter-row, *F. albida* trees
 137 (defoliated during the rainy season, average height = 13m) and location of the chambers under the Shade
 138 of trees (horizontal black arrows; N=4) and in Full sun (vertical white arrows; N=4); The shelter (red
 139 arrow) with solar panels is to fit the analyser, automation and batteries (c) automatic chamber enclosing a
 140 groundnut plant (during the rainy season) or bare soil (during the dry season), (d) Eddy Covariance (EC)
 141 tower (measurement height = 20 m) during the dry season.

142 *2.2. Experimental setup*

143 *2.2.1. CO₂ flux measurements in automatic chambers*

144 Continuous net CO₂ fluxes (FCO₂ch) from soil and groundnut plants were measured over a full
145 phenological year (June 17, 2021 – June 17, 2022) using eight automated dynamic chambers
146 (50×50×50 cm), each enclosing one groundnut plant. Four chambers were installed in full sun
147 (FS), at least 20 m from trees, and four under *F. albida* canopy shade (Sh). The chambers were
148 transparent, custom-built (Duthoit et al., 2020), and installed on metal bases embedded 10 cm
149 into the soil one month prior to measurements.

150 During the rainy season (June–November), groundnut coexisted briefly with spontaneous weeds
151 until weeding (mid-July), after which chambers contained only groundnut. Post-harvest (early
152 November), chambers remained bare while surrounding plots experienced weed regrowth.

153 CO₂ concentrations were measured at 1 Hz using a Picarro G2508 gas analyser (Picarro Inc., Santa
154 Clara, CA, USA) (Fleck et al., 2013; Reum et al., 2019; Valujeva et al., 2022). A fully automated
155 system was built for sequential half-hour flux measurements alternating FS and Sh chambers
156 (Table S2). Measurement duration was 15 min per chamber in the dry season, reduced to 5 min
157 during the rainy season to limit condensation effects.

158 *2.2.2. CO₂ flux measurements by Eddy Covariance*

159 The EC system (Li-COR SMARTFLUX®, including a Gill MasterPro 3D sonic anemometer and a LI-
160 7500 RS open path CO₂ and H₂O gas analyser) was mounted at a height of 20 m on a 30m mast,
161 above *F. albida*. It continuously monitored net CO₂ exchange from the ecosystem. Raw data were
162 collected at 20 Hz frequency and post-processed from binary files using the advanced mode of the
163 EddyPro® v7.0, with standard corrections and procedures: sonic tilt correction (double rotation),
164 block averaging, covariance maximisation for time lag, and WPL correction (Webb et al., 1980).
165 Quality control followed Foken et al. (2004) and Vickers & Mahrt (1997); random uncertainty was
166 estimated per Finkelstein & Sims (2001). Spectral corrections were applied according to
167 Moncrieff et al. (1997, 2004). Footprints were computed according to Kormann and Meixner
168 (2001), using the FREddyPro R package (Xenakis, 2016), indicating a ~1 ha source area covering
169 the entire field. Gap-filling and flux partitioning were conducted using ReddyProc (Wutzler et al.,
170 2018), applying the daytime partitioning approach of Lasslop et al. (2010).

171 *2.2.3. Ancillary measurements*

172 Environmental and vegetation variables were monitored continuously throughout the study.
173 Global radiation (Rg) was estimated from photosynthetically active radiation (PAR) using a Skye
174 sensor (averaged over 30-min intervals). The normalised difference vegetation index (NDVI) of

175 crops under full sun was recorded semi-hourly by a calibrated downward-facing sensor installed
176 at 20 m height (Ponzailler et al., 2003), processed following Soudani et al. (2012), and used to
177 estimate the leaf area index (LAI) time series for groundnut, weeds, and cowpea based on end-of-
178 season field LAI measurements in six 15 m² plots (as in Rousard et al., 2020).

179 Rainfall was recorded by an automatic weather station (CR1000 with TE525MM rain gauge,
180 Campbell Scientific), and soil volumetric water content (VWC) and temperature (T_{soil} , at 6 cm
181 depth) were monitored using TOMST® TMS-4 sensors, benchmarked prior to field deployment
182 inside and outside the chambers (Wild et al., 2019). Air temperature (T_{air}) was recorded inside
183 each chamber at 15 cm above ground, all at 5-min intervals. These measurements contribute to
184 the SoilTemp global database (Lembrechts et al., 2020, 2022).

185 Groundnut development was tracked weekly by counting leaves in each chamber. Total
186 groundnut LAI (LAI_{ch}) was then derived from average single-leaf area and chamber surface.

187 A detailed description of the data used in this study is provided in Supplement (Table S1).

188 2.3. Data processing

189 2.3.1. Flux calculation

190 Net CO₂ fluxes (FCO_{2ch}, in $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$) from the chambers were calculated from the linear
191 change in CO₂ concentration over time ($\Delta C/\Delta t$; Fig. S1 and Fig. S2) using the Eq.1.

$$192 \text{FCO}_{2\text{ch}} = \left(\frac{P}{RT_k}\right) \left(\frac{V}{A}\right) \left(\frac{\Delta C}{\Delta t}\right) \quad (\text{Eq. 1})$$

193 where P is atmospheric pressure (101 325 N m⁻²), R is the ideal gas constant (8.31 N m mol⁻¹ K⁻¹),
194 T_k is air temperature inside the chamber in Kelvin, V (0.125 m³) is the total system volume
195 (chamber, tubing, analyser cavity, pump, and water trap), and A (0.25 m²) is the chamber
196 footprint. The slope $\Delta C/\Delta t$ was obtained via linear regression (Duthoit et al., 2020).

197 Mean FCO_{2ch} values were computed separately for the four replicate chambers in full sun (FS)
198 and under *F. albida* shade (Sh). By convention, negative values indicate net CO₂ uptake
199 (photosynthesis), and positive values indicate net CO₂ release (respiration).

200 2.3.2. Quality control of chamber-based CO₂ flux measurements

201 The quality of chamber-based CO₂ flux measurements was assessed using a threshold of $R^2 \geq 0.8$
202 of the linear increase in CO₂ concentration during chamber closure. The minimum detectable flux
203 (MDF) was then calculated following Nickerson (2016) (Eq.2). The MDF defines the flux detection
204 threshold, below which data are considered unreliable due to instrument sensitivity and sampling
205 constraints (Zaman et al., 2021). In this study, the MDF was $\pm 0.0004 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$.

206
$$\mathbf{MDF} = \left(\frac{A_a}{tc(\sqrt{tc/ps})} \right) \left(\frac{VP}{ART} \right) \quad (\mathbf{Eq. 2})$$

207 where A_a is the analytical precision of the Picarro analyser (0.6 ppm; Picarro Inc., 2015), tc the
 208 closure time (s), p_s the sampling frequency (1 Hz), V the chamber volume, P the atmospheric
 209 pressure (101 325 N m⁻²), A the chamber footprint, R the gas constant (8.3 N m mol⁻¹·K⁻¹), and T
 210 the air temperature in Kelvin.

211 Following this quality control, fluxes were partitioned (Section 2.3.3) and gap-filled (Section
 212 2.3.4).

213 *2.3.3. Partitioning of chamber-based CO₂ fluxes*

214 The net CO₂ fluxes (FCO₂ch), averaged from four chambers per environment (FS and Sh), were
 215 partitioned into two components according to Eq. 3 (Reichstein et al., 2005).

216
$$\mathbf{FCO_2ch} = \mathbf{Rch} + \mathbf{GPPch} \quad (\mathbf{Eq. 3})$$

217 Rch includes heterotrophic respiration (Rh) from soil and other autotrophic respiration (Ra) from
 218 groundnut plants and roots of *F. albida* (Ra Groundnut + Ra tree below-ground). Rch is always
 219 positive (Rch > 0). GPPch (Gross Primary Productivity) represents the photosynthetic CO₂ uptake
 220 by the groundnut plants and is negative during the day (GPPch < 0), and zero at night, when
 221 FCO₂ch = Rch.

222 Half-hourly FCO₂ch fluxes were partitioned as follows: (1) an Arrhenius-type function (Lloyd &
 223 Taylor, 1994) was fitted between nocturnal Rch and T_{soil} during nighttime periods, for each 5-days
 224 throughout the time series (Eq. 4). This empirical formulation is based on several key
 225 assumptions. First, the relationship between nocturnal respiration and soil temperature is
 226 assumed to follow an exponential response, reflecting the temperature sensitivity of respiration
 227 processes. Second, the model assumes temporal stability of the respiration-temperature
 228 relationship between night and day, allowing diurnal respiration to be extrapolated from fitted
 229 parameters in Eq.4 and daytime T_{soil}. Third, we assumed that no abrupt changes in substrate
 230 availability or soil moisture occur between day and night — conditions that could otherwise
 231 disrupt the temperature-respiration relationship. These assumptions are widely applied in CO₂
 232 flux partitioning approaches (Reichstein et al., 2005; Lasslop et al., 2010). (2) Diurnal Rch was
 233 estimated by applying the Lloyd & Taylor function, previously calibrated on nocturnal data, to the
 234 corresponding daytime T_{soil} measurements for each 5-day interval. (3) GPPch was subsequently
 235 derived as the residual component of the net CO₂ flux during the day, according to:

236 **nocturnal Rch** = $R_{\text{ref}} \cdot \exp \left[E_0 \left(\frac{1}{T_{\text{ref}} - T_0} - \frac{1}{T_{\text{soil}} - T_0} \right) \right]$ (Eq. 4)

237 where R_{ref} ($\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$) is a fitted parameter representing the base respiration at the
 238 reference temperature [T_{ref} (K), (set at 288.15 K)]. E_0 (K) is the temperature sensitivity (set at
 239 250 K), T_{soil} (K) the soil temperature (K), and T_0 (K) is kept constant at 231.13 K, according to
 240 Lloyd & Taylor (1994).

241 **GPPch** = **diurnal FCO₂ch** – **diurnal Rch** (Eq. 5)

242 where diurnal FCO₂ch and diurnal Rch represent the daytime net CO₂ fluxes and respiration in
 243 $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$, respectively.

244 2.3.4. Gap-filling procedure

245 Missing Rch data were gap-filled using the model derived from Eq. 4 (Lloyd & Taylor, 1994). Prior
 246 to gap-filling GPPch, raw data were standardised by LAI to reduce variability between chambers
 247 due to differences in leaf surface area (Eq. 6). A light-response model was then fitted to the
 248 standardised GPPch data, every 5-day period, to gap-fill missing values. The model is based on a
 249 rectangular hyperbolic function that describes the relationship between photosynthetic CO₂
 250 uptake and incoming global radiation (Rg) (Eq. 7). It corresponds to a Michaelis–Menten-type
 251 light-response curve, commonly used in ecosystem carbon exchange studies (Falge et al., 2001;
 252 Lasslop et al., 2010).

253 **GPPch.stand** = $\frac{\text{GPPch}}{\text{LAIch}} * \text{LAI.field}$ (Eq. 6)

254 where GPPch.stand ($\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$) is the standardised GPPch. LAI_{ch} and LAI_{field} (m^2 leaves
 255 m^{-2} soil) represent the groundnut LAI inside the chambers and the groundnut + weeds + cowpea
 256 LAI for the whole field, respectively.

257 **GPP** = $\frac{\alpha\beta Rg}{\alpha Rg + \beta}$ (Eq. 7)

258 where α ($\mu\text{mol CO}_2 \text{ J}^{-1}$) represents the light use efficiency of the groundnut plants inside the
 259 chambers, and refers to the initial slope of the light-response curve, β ($\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$) is the
 260 maximum CO₂ uptake rate by the groundnut plants at light saturation, and Rg the global radiation
 261 (W m^{-2}).

262 2.3.5. Comparing chamber-based (Ch) and Eddy Covariance (EC) methods

263 Chamber measurements were upscaled to field-level CO₂ fluxes and compared with EC-derived
 264 fluxes. Before comparison, a correction was applied (Eq. 6) to account for differences in LAI
 265 between chambers (LAI_{ch}) and the field (LAI_{field}), due to the presence of cowpea and weeds in
 266 the field but not in the weeded chambers.

267 Upscaling considered tree cover, with FS and Sh chamber fluxes weighted at 90% and 10%,
 268 respectively. Rch.stand and GPPch.stand, representing chamber-based respiration and

269 photosynthesis at field scale. These fluxes were compared, on a half-hourly basis, to EC-derived
270 Reco.EC and GPP.EC (Table S4). The November–December transition period was excluded due to
271 weed-driven uncertainties after groundnut harvest.

272 During the rainy season (*F. albida* leafless), GPP.EC represented ground vegetation (groundnut,
273 cowpea, weeds), while Reco.EC included autotrophic respiration from all vegetation (including
274 trees), and heterotrophic respiration (Reco.EC = R_a tree below-ground + R_a tree above-ground
275 + R_a groundnut + R_a cowpea + R_a weeds + R_h). Rch.stand could not be fully upscaled to the field
276 due to uncertainty in its partitioning between R_a and R_h . Rch.stand accounted only for R_a tree
277 below-ground, R_a groundnut, and R_h .

278 In the dry season (leafy trees, bare soil), GPP.EC reflected tree photosynthesis only (GPP tree),
279 while GPPch.stand was nil. Reco.EC included R_a tree (above- and below-ground) and R_h .
280 Rch.stand, measured on bare soil, represented only R_a tree below-ground + R_h .

281 *2.3.6. Contribution of trees to full ecosystem respiration and photosynthesis*

282 During the dry season, when the trees (*F. albida*) maintained their foliage, a comparison between
283 chamber and EC measurements allowed for the estimation of the contribution of the above-
284 ground tree compartments to total ecosystem respiration (Table S4). Based on this estimate, total
285 tree respiration (R_a tree) was then calculated under the assumption that the tree root systems
286 (R_a tree below-ground) represent $\frac{1}{3}$ of the above-ground biomass (Jackson et al. 1996).

287 Given the GPP measured during the dry season was equivalent to GPP of trees (GPP trees) from
288 EC measurements, the carbon use efficiency of the trees (CUE tree) was then calculated (Table
289 S4). The resulting CUE value was assessed to determine whether it approximated the typical value
290 of 0.5, which is often used as a default in ecosystem models (Zhou et al., 2019; 2020).

291 *2.3.7. Net annual vertical C balance at the ASPS scale*

292 The net C balance of CO₂ fluxes in a yearly basis was estimated for chambers and EC measurements
293 in Mg C-CO₂ ha⁻¹. The chambers CO₂ flux balances were obtained by calculating the annual sum
294 of the net CO₂ flux measurements and then weighting with the tree cover rate (10% for the Sh,
295 90% for the FS). These annual balances for the field are considered apparent representing vertical
296 CO₂ exchanges only, as they do not account for the biomass exported from the field after the
297 harvest, the decomposition of which therefore escaped both the chambers and the EC.
298 Additionally, the inputs and the outputs of faecal matter resulting from livestock wandering
299 during the dry season were not quantified and are therefore neglected. The objective here is to
300 compare two approaches at different scales using vertical net C balances, rather than to provide
301 an absolute C budget which would also include horizontal transfers of carbon.

302 2.4. Statistical analyses

303 Statistical analyses were performed using the R software (R. Core Team, 2023). To compare the
304 mean values of climatic parameters between the FS and Sh situations, a non-parametric Mann-
305 Whitney test was used when both the normality (`shapiro.test`) and the homogeneity of the
306 variance (Levene Test, R package 'Car'; Fox et al., 2023) were not confirmed. This approach was
307 similarly applied to compare the seasonal dynamics of CO₂ fluxes between FS and Sh, as well as
308 between the chamber-based and Eddy Covariance (EC) methods. Means and standard deviations
309 were computed using the 'skim' function from the R package 'skimr' (Waring et al., 2022).

310 Respiration (R_{ch}) (Eq. 4) and GPP (GPP_{ch}) models (Eq. 7) were fitted using non-linear least
311 squares regression, implemented in the library in R 'nls.multstart' (Padfield et al., 2025). For the
312 GPP_{ch} model, parameters α and β with non-significant p-values were removed, and then the
313 remaining values were interpolated and smoothed using a 'spline' function from the 'zoo' library
314 in R (Zeileis et al., 2024). Ordinary least-square linear regressions were fitted between the
315 measured and the modeled values derived from. Model performance of Eq. 4 and Eq. 7 was
316 evaluated by fitting ordinary least-square linear regressions between the measured and the
317 modeled values using R², root mean square error (RMSE), and the bias metrics. Given that the
318 primary objective of these equations was to accurately reproduce the seasonal dynamics of the
319 CO₂ fluxes to fill gaps in data, particular emphasis was placed on R², with a higher value reflecting
320 a better fit of the model to the measurements.

321 Correlation analysis was conducted between chamber CO₂ fluxes (FCO_{2ch}, R_{ch}, GPP_{ch}) and soil
322 temperature (T_{soil}, °C), air temperature (T_{air}, °C), VWC, the leaf area index of groundnut plants in
323 the chambers (LAI_{ch}), and the fitted parameters for respiration — R_{ref} — and photosynthesis —
324 α and β . This analysis was performed using the 'cor.test' function from the 'stats' package in R
325 (Lüdecke et al., 2021), applying the Spearman method.

326 The threshold of the daily mean soil temperature (T_{soil}, °C) at which the cumulative daily
327 respiration (R_{ch}, g C-CO₂ m⁻² d⁻¹) began to decline was determined using segmented regression
328 from the R package 'segmented' (Muggeo, 2003). The associated uncertainty (standard error) of
329 this estimate was evaluated through a bootstrap procedure.

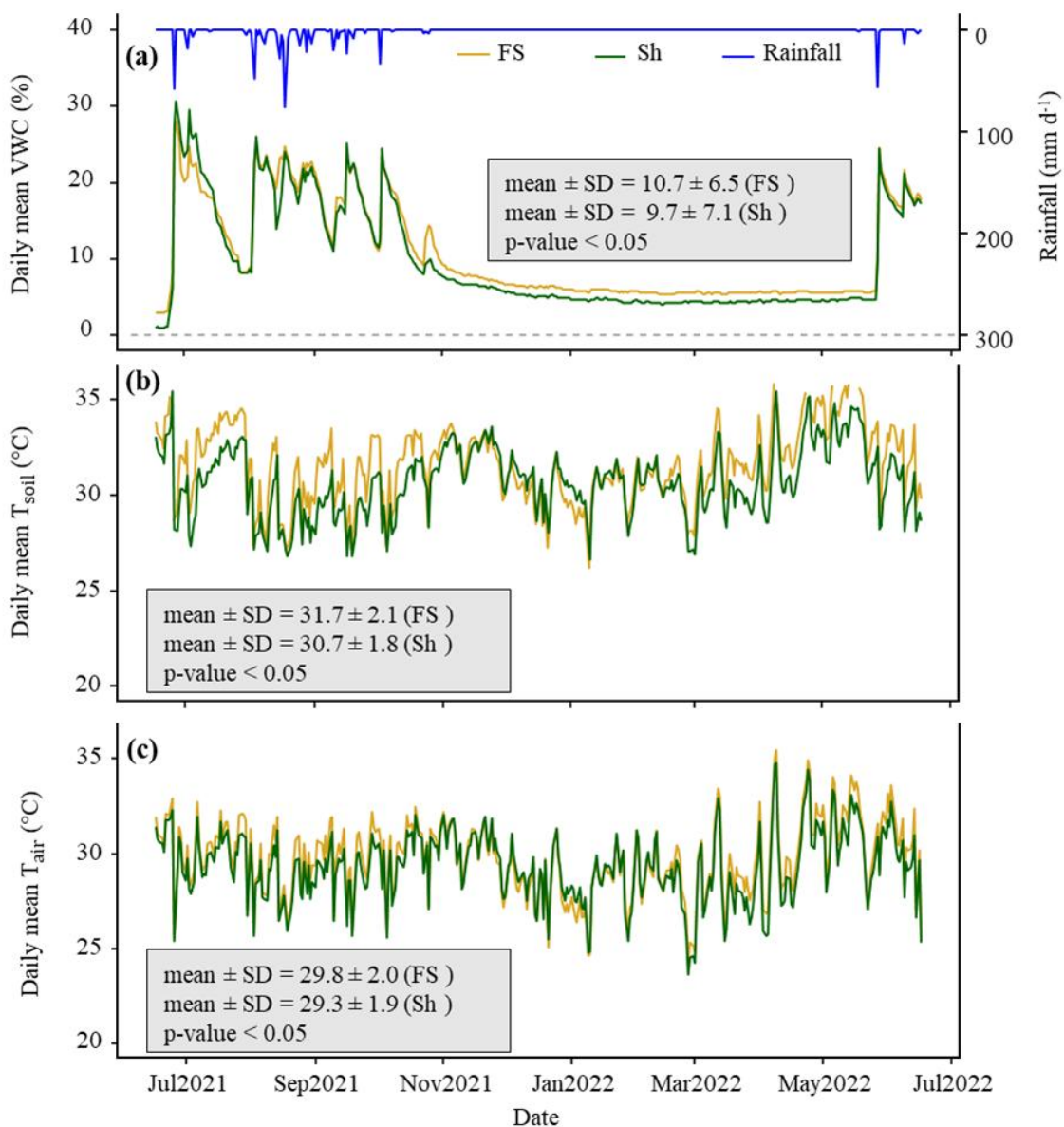
330 The standard error of the total annual flux was estimated using the error propagation method.
331 This calculation considered the mean standard deviation of daily fluxes (g C-CO₂ d⁻¹) and the
332 effective number of measurement days (365). For each FS and Sh condition, the mean daily
333 standard deviation was multiplied by the square root of 365 to obtain the annual standard error.
334 The resulting values were then weighted by 90% for FS and 10% for Sh to derive the overall
335 standard error of the annual flux sum, which was subsequently converted to Mg C-CO₂ ha⁻¹.

336 3. Results

337 3.1. Microclimatic conditions

338 During the experiment, the cumulative rainfall was 550 mm, which was representative of the
339 interannual average. Precipitations were lowest in July and highest between August and
340 September, a period that typically corresponds to the peak of the rainy season (Fig. 2a). Global
341 radiation ranged between 5.8 and 32.4 MJ m⁻² d⁻¹ (data not shown). The daily mean VWC in the
342 chambers showed significant variation, ranging from 1% at the end of the dry season to a
343 maximum of 30% during the rainy season (Fig. 2a). While VWC was similar during the rainy
344 season, it remained consistently higher in FS than in Sh throughout the dry season ($p < 0.05$),
345 which was unexpected. However, it should be noted that the last rain of October 2021 recharged
346 the FS chambers more effectively, likely due to foliage rainfall interception by *F. albida* which had
347 just put on leaves at that time, potentially explaining this discrepancy in VWC.

348 Within the chamber, the daily mean T_{soil} ranged from 26°C in April to 37.5°C at the end of the dry
349 season (Fig. 2b), while T_{air} varied between 23.7°C and 35.5°C (Fig. 2c). However, during
350 instantaneous daily peaks, T_{soil} could exceed 45°C in May (data not shown). As expected, both daily
351 mean T_{soil} and T_{air} were significantly higher in FS compared to Sh situations ($p < 0.05$), with T_{soil}
352 and T_{air} averaging respectively 1°C and 0.5°C lower under the tree canopy.



353 Fig. 2: One-year time series of daily average microclimatic parameters measured inside the
 354 chambers.

355 (a) volumetric soil water content (VWC) at a depth of 6 cm (%). (b) soil temperature (T_{soil}) at a depth of 6
 356 cm ($^{\circ}\text{C}$), (c) air temperature (T_{air}) at a height of 15 cm ($^{\circ}\text{C}$). The blue line depicts the daily rainfall (mm d^{-1})
 357 throughout the year. FS: Full sun chambers; Sh: Shaded chambers. Mean and SD represent respectively the
 358 mean value and the standard deviation. The p-value indicates the probability associated with the statistical
 359 test, assessing the differences in means between FS and Sh with the significance level α set to 0.05.

360 *3.2. Modeling the chamber-based total respiration (R_{ch}) and photosynthesis (GPP_{ch})*

361 *3.2.1. Dynamics of reference respiration, light use efficiency, and maximum CO₂ uptake rate at*
362 *light saturation (R_{ref}, α , and β)*

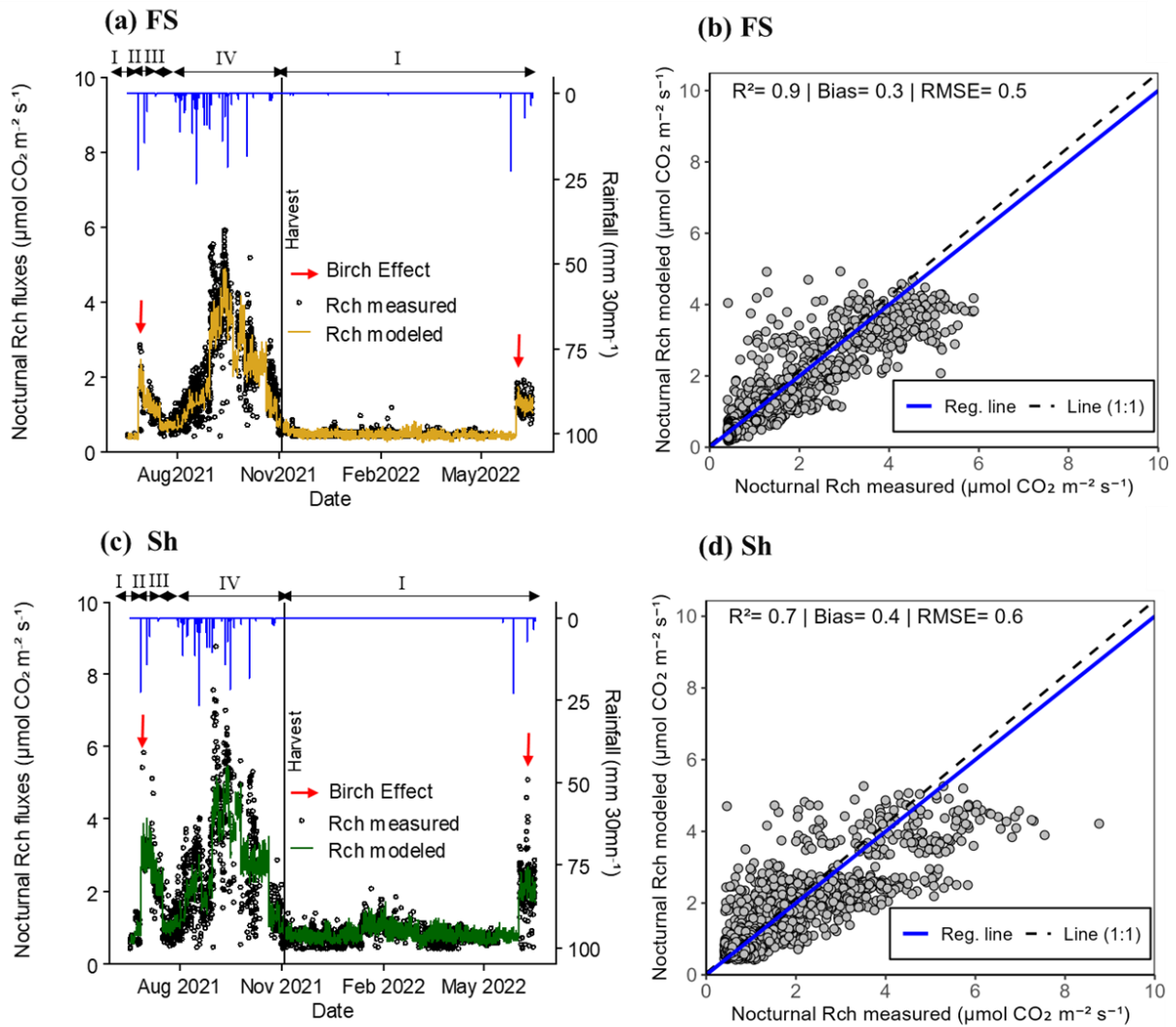
363 The reference respiration (R_{ref}) showed comparable seasonal dynamics both at a distance from
364 the trees (FS) and under the tree canopies (Sh) (Fig. S4). In both situations, R_{ref} showed strong
365 variability during the rainy season, peaking in September 2021 at 2.4 $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ for FS and
366 2.9 $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ for Sh (Table S3). In contrast, during the dry season — from November 3,
367 2021 (after harvest) until the onset of the following rainy season (June 2022) — R_{ref} values
368 dropped both for FS and Sh, averaging $0.3 \pm 0.5 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ for FS and $0.5 \pm 0.6 \mu\text{mol CO}_2$
369 $\text{m}^{-2} \text{ s}^{-1}$ for Sh. This represents a reduction by a factor of 8 for FS and 6 for Sh compared to the
370 rainy season. The mean annual R_{ref} values were significantly higher under Sh than in FS, with value
371 approximately 1.5 times greater (Table S3).

372 Regarding GPP in chambers, the light use efficiency (α) and the maximum CO₂ uptake by
373 groundnut plants in the chambers (β), also reached their maximum during the peak of the rainy
374 season (Fig. S5, a and b). The maximum value of α reached 0.2 $\mu\text{mol CO}_2 \text{ J}^{-1}$ in FS and 0.3 μmol
375 $\text{CO}_2 \text{ J}^{-1}$ in Sh (Table S3). Similarly, the maximum values of optimum CO₂ uptake rate at light
376 saturation (β) were 40.2 $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ for FS and 42.8 $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ for Sh (Table S3). In
377 the dry season, when photosynthetic activity ceased in the chambers, both α and β were assumed
378 to be nil (Fig. S5, a and b). On average, α and β were significantly higher in Sh than in FS, by a factor
379 of 1.7 and 1.2, respectively (Table S3). We noted that the decline in photosynthetic activity of the
380 groundnut crop occurred earlier and rapidly at a distance from the trees (FS), as reflected by the
381 sharply observed recession of α and β in FS.

382 *3.2.2. Dynamics of nocturnal respiration in chambers*

383 The averaged nocturnal respiration (nocturnal R_{ch}) calculated from the measurements across
384 each treatment (FS and Sh), showed similar seasonal patterns (Fig. 3, a and c). Following the first
385 rains, R_{ch} values increased dramatically, with a nocturnal 'Birch effect' — a sudden pulse of CO₂
386 release following soil rewetting — observed to be more pronounced under Sh compared to FS,
387 approximately by a factor of 2. At the peak of the rainy season (September), the maximum
388 nocturnal R_{ch} values reached approximately 6.0 $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ in FS and 9.0 $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$
389 in Sh (Fig. 3, a and c). Thereafter, nocturnal R_{ch} declined well before the groundnut harvest along
390 with the rainfall spacing and the groundnut crop senescence (data not shown). During the dry
391 season nocturnal R_{ch} continued to decrease, with maximum values around 1.0 $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$
392 in FS and 2.0 $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ in Sh (Fig. 3, a and c).

393 The modeled nocturnal Rch values closely matched the measured nocturnal Rch values (mean
394 across four chambers per treatment), as indicated by the model performance metrics ($R^2 = 0.9$,
395 with bias and RMSE values of 0.3 and 0.5 $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$, respectively, for FS; $R^2 = 0.7$, with bias
396 and RMSE values of 0.4 and 0.6 $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$, respectively, for Sh) (Fig. 3, b and d). Similarly,
397 the daily mean modeled values also fitted well with the measured values, with FS showing (mean
398 \pm standard deviation) $0.9 \pm 0.9 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ (modeled) and $1.2 \pm 1.2 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$
399 (measured), while Sh recorded $1.4 \pm 0.9 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ (modeled) and $1.5 \pm 1.2 \mu\text{mol CO}_2 \text{ m}^{-2}$
400 s^{-1} (measured). Given the close match between the measured and modeled values, the fitted
401 model parameters were used subsequently to fill data gaps and estimate diurnal Rch values, as
402 presented in Fig. 4, a and c.



403 Fig. 3: Dynamics of instantaneous nocturnal CO₂ fluxes in chambers in Full sun (FS; a and b) and
 404 Shaded (Sh; c and d) environments (data filtered based on R² of the CO₂ variation over the time
 405 of chamber closure and Minimum Detectable Flux, Eq.2).

406 (a) and (c): measured nocturnal respiration in chambers (Rch: black dots; average of measurements in 4
 407 chambers per location) vs. modeled (coloured line). The vertical black line indicates the harvest date of
 408 groundnuts inside the chambers. The red arrows indicate the 'Birch' effect and the blue line represents the
 409 rainfall (mm 30mn⁻¹). Roman numerals (above the black arrows) refer to vegetation conditions prevailing
 410 inside the chambers, i.e. (I) bare soil, (II) weeds, (III) weeds + groundnuts, and (IV) groundnuts only.

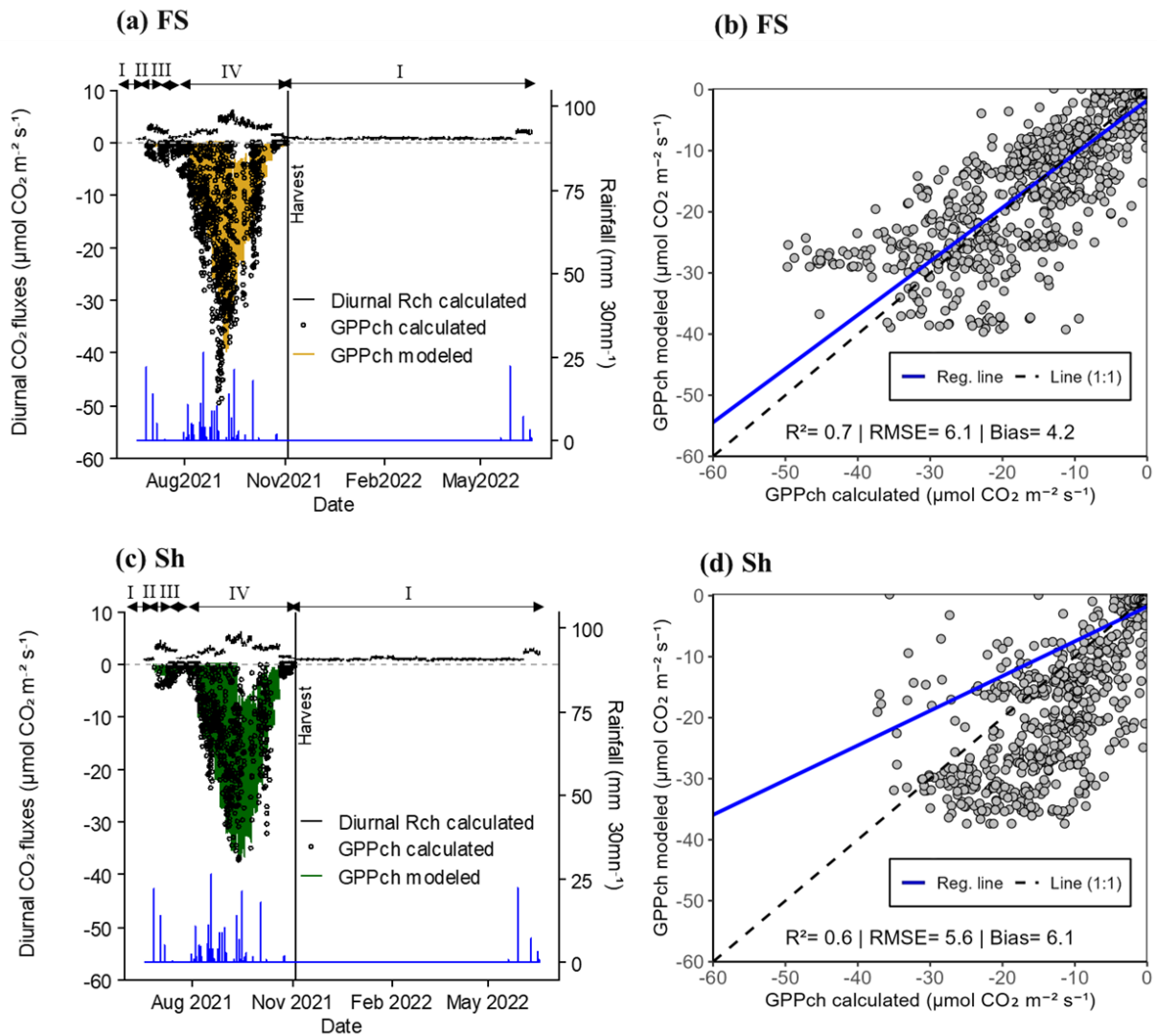
411 (b) and (d): scatter plot between measured and modeled nocturnal Rch. The solid blue line indicates the
 412 regression line and the dashed black one the (1:1) line RMSE and bias are expressed as fluxes (in μmol CO₂
 413 m⁻² s⁻¹). Each point represents the mean value from 4 chambers within the FS or Sh environments.

414 *3.2.3. Dynamics of daytime fluxes in chambers*

415 The measured GPP_{ch} stand, as well as GPP modeled with Eq. 6, showed similar seasonal dynamics
416 in FS and Sh (Fig. 4, a and c). The fluxes peaked during the rainy season (Fig. 4a and c), coinciding
417 with periods of vigorous vegetative growth characterised by a high leaf area index (LAI_{ch}) of
418 groundnut plants within the chambers (Fig. S3). The maximum calculated and standardised
419 GPP_{ch} values reached $-50 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ for FS and $-37 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ for Sh. As expected,
420 these fluxes were nil during the dry season when the soil was bare (Fig. 4, a and c).

421 The modeled GPP_{ch} values closely followed the same trends as the calculated values, although
422 model performance was slightly better for FS ($R^2 = 0.7$ with bias and RMSE values of 4.2 and 6.1
423 $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$, respectively) compared to Sh ($R^2 = 0.6$ with bias and RMSE values of 6.1 and
424 $5.6 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$, respectively) (Fig. 4, b and d).

425 The calculated diurnal respiration values (diurnal R_{ch} calculated) for FS and Sh revealed a 'Birch
426 effect' similar to that observed during the night, though slightly more pronounced under Sh by a
427 factor of 1.2. Diurnal R_{ch} values increased significantly during the rainy season, reaching a
428 maximum of $6.0 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ for both FS and Sh (Fig. 4, a and c). In the dry season, on bare
429 soil, these values declined, with maximum respiration reaching only $0.5 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ for both
430 situations (FS and Sh) (Fig. 4, a and c).



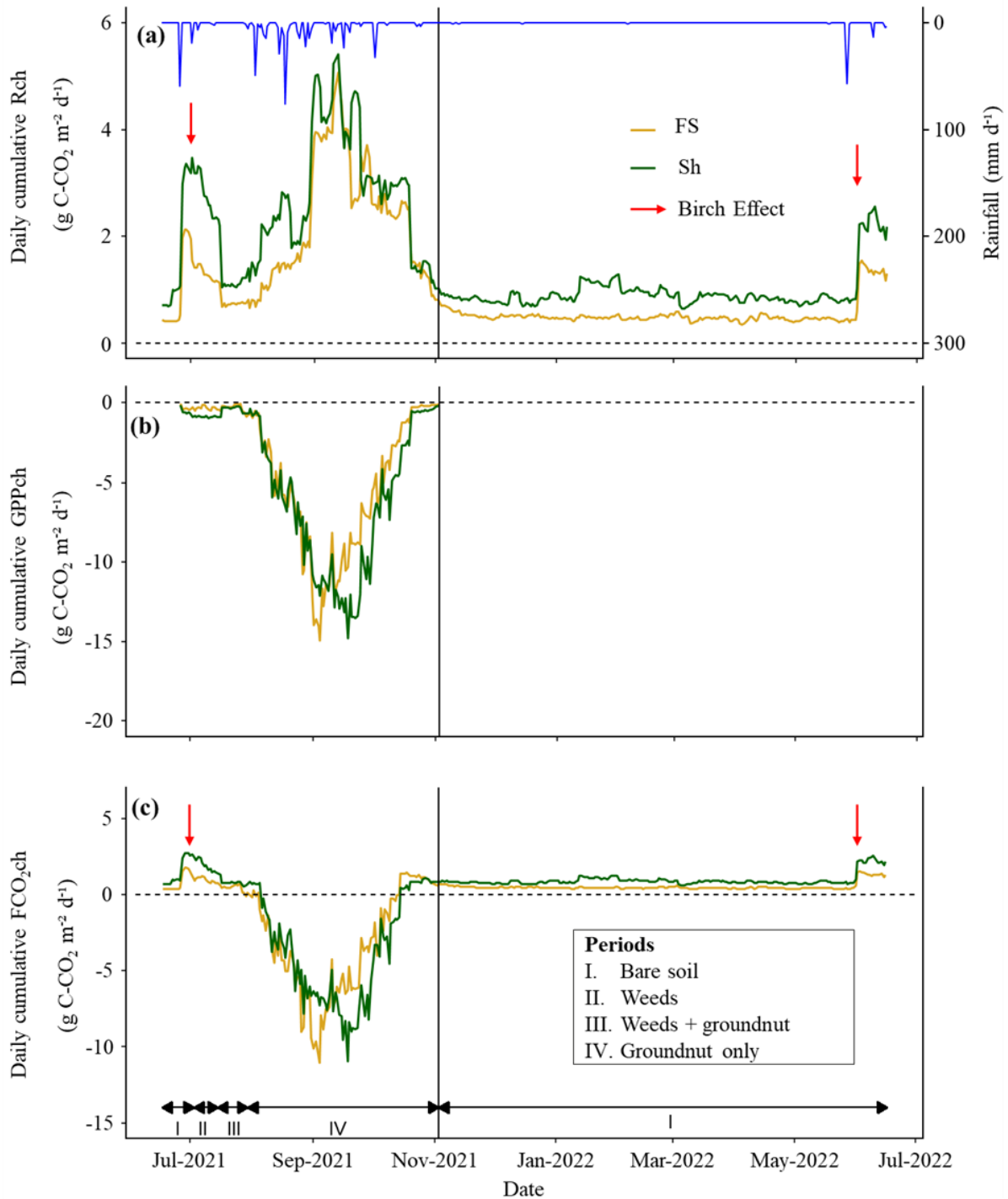
431 Fig. 4: Dynamics of instantaneous diurnal CO₂ fluxes in chambers in Full sun (FS; a and b) and
 432 Shaded (Sh; b and d) environments (filtered based on R² of the CO₂ variation over the time closure
 433 in FS and Sh and Minimum Detectable Flux, Eq.2).

434 (a) and (c): non-gap-filled diurnal Rch calculated (black line, positive values; average of measurements in
 435 4 chambers per location) and GPPch calculated from Eq.5 then standardised for LAI (black dots, negative
 436 values) and modeled (coloured line, negative values). The vertical black line indicates the harvest date of
 437 groundnuts inside the chambers and the blue line represents the rainfall (mm 30mn⁻¹). Roman numerals
 438 (above the black arrows) refer to conditions prevailing inside the chambers, i.e., (I) bare soil, (II) weeds,
 439 (III) weeds + groundnuts, and (IV) groundnuts.

440 (b) and (d): scatter plot between calculated and modeled GPPch. The solid blue line indicates the regression
 441 line and the dashed black one the (1:1) line. RMSE and bias are expressed as fluxes (in μmol CO₂ m⁻² s⁻¹).
 442 Each point represents the mean value from 4 chambers within the FS or Sh environments.

443 *3.3. Dynamics of daily cumulative CO₂ fluxes in chambers*

444 The seasonality of daily cumulative of GPPch stand showed similar dynamics between FS and Sh,
445 with higher variability during the rainy season than during the dry season (Fig. 5). Daily total Rch
446 peaked during the rainy season at 5.1 g C-CO₂ m⁻² d⁻¹ for FS and 5.4 g C-CO₂ m⁻² d⁻¹ for Sh, while
447 the maximum GPPch stand values were comparable at around -15.0 g C-CO₂ m⁻² d⁻¹ for both FS
448 and Sh (Table 1; Fig. S7, a, b, c, and d). In the dry season, Rch decreased (Fig. 5), averaging 0.5 g C-
449 CO₂ m⁻² d⁻¹ for FS and 1.0 g C-CO₂ m⁻² d⁻¹ for Sh. GPPch declined well before harvest (senescence)
450 and remained nil during the dry season (Fig. 5). During the rainy season FCO₂ch peaked in
451 absolute value at around 11.0 g C-CO₂ m⁻² d⁻¹ for FS and Sh (Fig. 5) (Table 1; Fig. S7, e and f),
452 while FCO₂ch values were the same as Rch during the dry season. In absolute terms, the mean Rch
453 and GPPch were significantly higher under Sh as compared to FS, by factors of 1.3 and 1.2,
454 respectively. Conversely, the mean FCO₂ch was significantly higher in absolute value under FS
455 (0.4 g C-CO₂ m⁻² d⁻¹) than under Sh (0.2 g C-CO₂ m⁻² d⁻¹) (Table 1).
456 The annual cumulative Rch values were 392.8 g C-CO₂ m⁻² for FS and 574.5 g C-CO₂ m⁻² for Sh.
457 The GPPch fluxes reached -539.5 g C-CO₂ m⁻² for FS and -632.6 g C-CO₂ m⁻² for Sh. The net annual
458 cumulative C exchange (FCO₂ch) was -146.7 g C-CO₂ m⁻² in FS and -58.1 g C-CO₂ m⁻² in Sh.



459 Fig. 5: Seasonal dynamics of daily gap-filled cumulative fluxes (in gC-CO₂ m⁻² d⁻¹) in chambers.

460 (a) soil+crop respiration (Rch), (b) photosynthesis (GPPch, standardised for LAI) and (c) net CO₂ exchange
 461 (FCO₂ch). The yellow and green solid lines compare the FS and Sh environments, respectively. The vertical
 462 black line indicates the harvest date of groundnuts inside the chambers. The blue line depicts the daily
 463 cumulative rainfall (mm d⁻¹) throughout the rainy season, and the red arrow indicates the 'Birch'
 464 effect. Roman numerals (above the black arrows) in (a) and (c) refer to the prevailing conditions inside the
 465 chambers: (I) bare soil, (II) weeds, (III) weeds + groundnuts, (IV) groundnuts.

466 Table 1: Comparison of daily cumulative and gap-filled chamber CO₂ fluxes (Rch, GPPch
 467 standardised for LAI, and FCO₂ch in g C-CO₂ m⁻²) in the FS and Sh condition.

468

	Annual sum	Daily Mean \pm SD	Min	Max	Mann-Whitney test
(g C-CO ₂ m ⁻²)	.yr ⁻¹	.d ⁻¹	.d ⁻¹	.d ⁻¹	
Rch					
FS	392.8	1.1 \pm 0.9	0.4	5.1	
Sh	574.5	1.6 \pm 1.1	0.6	5.4	*
GPPch					
FS	-539.5	-4.1 \pm 4.3	< -0.1	-14.9	
Sh	-632.6	-4.8 \pm 4.6	< -0.1	-14.8	*
FCO₂ch					
FS	-146.7	-0.4 \pm 2.4	-11.0	1.8	
Sh	-58.1	-0.2 \pm 2.7	-10.9	2.8	*

469 Annual sum corresponds to the annual cumulative fluxes (g C-CO₂ m⁻² yr⁻¹). Mean, SD, Min, and Max
 470 represent respectively the mean, standard deviation, minimum, and maximum values at the daily scale (g
 471 C-CO₂ m⁻² d⁻¹). Asterisks (*) indicate the p-values from the Mann-Whitney test, used to assess differences in
 472 mean between FS and Sh (p < 0.05). Positive values indicate CO₂ emissions, while negative values represent
 473 CO₂ uptake.

474 *3.4. Drivers of daily respiration and photosynthesis in chambers*

475 The chamber-based daily cumulative respiration (Rch) and GPPch showed significant and positive
476 correlations with the leaf area index (LAIch), both at a distance from the trees (FS) and under the
477 trees (Sh) (Table 2). The influence of LAIch on GPPch was stronger ($r = 0.86$ for FS and Sh) than
478 its influence on Rch ($r = 0.60$ for FS; $r = 0.69$ for Sh). Soil VWC was also positively correlated with
479 Rch and GPPch, both in FS and Sh. However, the influence of soil VWC on Rch was stronger under
480 Sh compared to FS, while its influence on GPPch was similar in both situations (FS and Sh). Soil
481 temperature showed weak negative correlations with Rch (in FS and Sh) and with GPPch (only in
482 Sh). Finally, no significant correlations were found between T_{air} , and any of the CO₂ fluxes (Table
483 2).

484 Table 2: Spearman correlation matrix based on daily cumulative and gap-filled CO₂ fluxes from full
 485 year chamber measurements (g C-CO₂ m⁻² d⁻¹) with microclimatic parameters.

Parameters	Condition	r-coef. (Rch)	p (Rch)	r-coef. (GPPch)	p (GPPch)
T _{soil}	FS	-0.25 ***	7.47 x 10 ⁻⁴	ns	1.18 x 10 ⁻³
	Sh	-0.28 ***	9.69 x 10 ⁻¹⁴	-0.38 ***	2.88 x 10 ⁻⁷
T _{air}	FS	ns	0.22	ns	0.35
	Sh	ns	0.98	ns	0.15
VWC	FS	0.51 ***	3.00 x 10 ⁻³⁴	0.75 **	6.73 x 10 ⁻³
	Sh	0.78 ***	1.29 x 10 ⁻⁶⁶	0.75*	0.02
LAIch	FS	0.60 ***	1.11 x 10 ⁻⁶¹	0.86 ***	2.23 x 10 ⁻⁸
	Sh	0.69 ***	6.08 x 10 ⁻⁶⁹	0.86 ***	2.11 x 10 ⁻¹²

486 Spearman correlation coefficients (r-coef.) between daily cumulative and gap-filled CO₂ flux components
 487 (Rch and GPPch, with GPPch in absolute terms) and daily mean microclimatic parameters in full sun (FS)
 488 and shaded chambers (Sh). T_{soil} (°C) is the daily mean soil temperature at 6 cm depth, T_{air} (°C) the daily
 489 mean air temperature at 15 cm height, VWC (%) the daily mean volumetric water content (VWC, %), and
 490 LAIch (m⁻² leaf m⁻² soil) the chamber leaf area index value for a given day. Letter p represents the p-value
 491 and significance levels are indicated by (***) p<0.001; ** p<0.01; * p<0.05); ns p > 0.05.

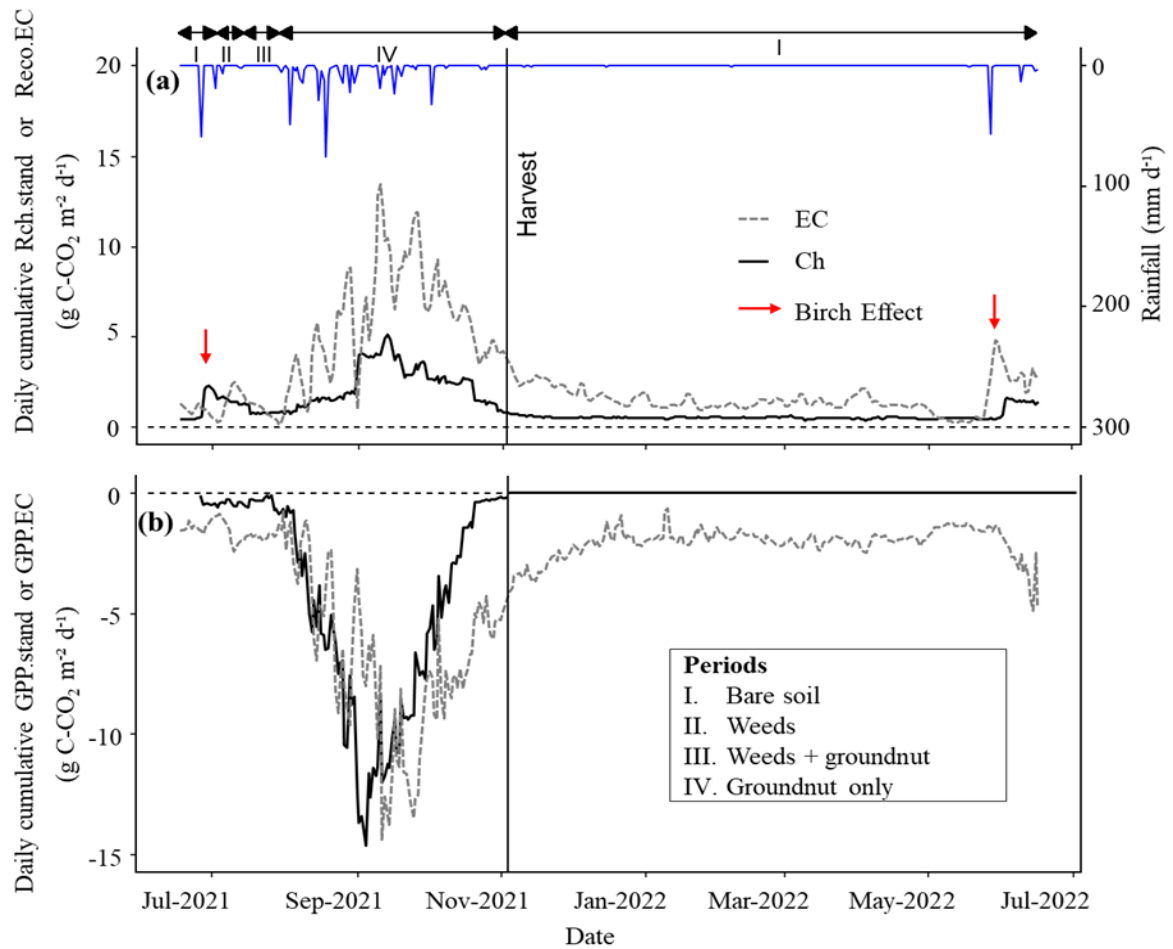
492 3.5. Comparison of respiration and GPP measurements between chambers (Ch) and Eddy
493 Covariance (EC) methods

494 The chamber-based daily total CO₂ fluxes, gap-filled and weighted according tree cover were
495 compared with the fluxes obtained using the EC method (Fig. 6).

496 During the rainy season, both total respiration and GPP showed comparable dynamics between
497 the two methods, with synchronised peaks and higher variability compared to the dry season (Fig.
498 6). The maximum value of Reco.EC, peaked at 13.5 g C-CO₂ m⁻² d⁻¹ (Table 3). The initial value of
499 Rch.stand was comparable to Reco.EC but peaked only at 5.1 g C-CO₂ m⁻² d⁻¹ (Table 3), meaning
500 a third of the peak of Reco.EC. The maximum GPP, was -14.3 g C-CO₂ m⁻² d⁻¹ and -14.6 g C-CO₂
501 m⁻² d⁻¹ for GPP.EC and GPPch.stand, respectively (Table 3). This indicates that the LAI-based
502 standardisation and upscaling approach were realistic, at least up to the peak of groundnut
503 growth.

504 On average, Reco.EC was significantly higher than Rch.stand, by a factor of 2.3. GPP.EC was also
505 significantly higher than GPPch.stand, but only by a factor of 1.2 (Table 3).

506 During the dry season, Reco.EC and Rch.stand gradually decreased. The values for Reco.EC
507 remained higher than for Rch.stand, which was fairly consistent with the contribution of the Ra
508 tree above-ground compartment, even if this difference seemed to disappear at the end of the dry
509 season (Fig. 6). The measured 'Birch effect' was highest for Rch.stand in 2021, but was the
510 opposite in 2022 due to a system failure at the beginning of the rainy season. The maximum value
511 of GPP.EC reached -2.4 g C-CO₂ m⁻² d⁻¹ when the trees were at their maximum of foliage, after
512 harvest and while weeds were still present in the field. However, after the harvest, chamber
513 photosynthesis (GPPch.stand) was nil (Table 3).



514 Fig 6: Comparing the seasonal dynamics of CO₂ fluxes between Eddy Covariance (EC)
 515 measurements and upscaled chamber measurements (ch. stand).

516 (a) represent the seasonal dynamics of soil + crop respiration (Rch. stand) and ecosystem respiration
 517 (Reco. EC) and (b) photosynthesis (GPP. stand and GPP. EC). The black and dashed grey lines show Ch and
 518 EC seasonal dynamics, respectively. The vertical black line indicates the harvest date of groundnuts inside
 519 the chambers. The blue line depicts the daily cumulative rainfall (mm d⁻¹), and the red arrow indicates the
 520 'Birch' effect. Roman numerals (above the black arrows) refer to conditions prevailing inside the
 521 chambers: (I) bare soil, (II) weeds, (III) weeds + groundnuts, (IV) groundnuts.

522 Table 3: Comparison of gap-filled CO₂ fluxes between Eddy Covariance (EC) and upscaled chamber (Ch.stand) measurements, by season (rainy or dry).

	Rainy season				Dry season			
	Daily Mean ± SD .d ⁻¹	Min .d ⁻¹	Max .d ⁻¹	Mann-Whitney test	Daily Mean ± SD .d ⁻¹	Min .d ⁻¹	Max .d ⁻¹	Mann-Whitney test
Reco.EC or Rch.stand								
EC	4.6 ± 3.2	0.2	13.5	*	1.2 ± 0.4	0.3	2.1	*
Ch.stand	2.0 ± 1.1	0.5	5.1		0.5 ± 0.04	0.4	0.6	
GPP.EC or GPPch.stand								
EC	-5.1 ± 3.6	-0.7	-14.3	*	-1.7 ± 0.3	-0.6	-2.4	
Ch.stand	-4.2 ± 4.3	<-0.1	-14.6		0	0	0	-

523 Mean, SD, Min, and Max represent the daily mean fluxes, standard deviation, minimum, and maximum values, respectively (g C- CO₂ m⁻² d⁻¹). The Asterisks (*) indicate
524 the p-values from the Mann-Whitney test, used to assess differences in mean between EC and Ch. Positive values indicate CO₂ emissions, while negative values
525 represent CO₂ uptake.

526 *3.6. The contribution of F. albida to Reco and GPP*

527 During the dry season, the cumulative contribution of *F. albida* to ecosystem respiration (Ra tree)
528 was 139.6 g C-CO₂ m⁻². This represent 14% of the total annual cumulative Reco, which was
529 estimated at 1000.0 g C-CO₂ m⁻² (Table S4). The contribution of trees (GPP tree) to total annual
530 GPP in absolute term was 270.2 g C-CO₂ m⁻², equivalent to ~23% of the total annual cumulative
531 GPP of the ecosystem measured by EC (1180.0 g C-CO₂ m⁻²) (Table S4).
532 The ratio between these two components (Ra tree / GPP tree) in absolute terms was 0.52,
533 reflecting a carbon use efficiency (CUE) of 0.48 (Table S4).

534 *3.7. Annual vertical CO₂ balances at the field-scale*

535 The upscaled chamber-based annual cumulative total respiration flux (Rch.stand) was estimated
536 to be 4.1 ± 0.18 Mg C-CO₂ ha⁻¹ (Table 4). In comparison, the annual Reco.EC was 10.0 ± 0.49 Mg
537 C-CO₂ ha⁻¹ (Table 4), more than two times larger than Rch.stand.
538 The upscaled GPPch.stand reached an annual cumulative value of -5.5 ± 0.83 Mg C-CO₂ ha⁻¹,
539 whereas the annual cumulative GPP.EC was -11.8 ± 0.53 Mg C-CO₂ ha⁻¹ (Table 4).
540 The net annual vertical C balance, based on both methods, was estimated at -1.4 ± 0.46 Mg C-CO₂
541 ha⁻¹ for chambers (FCO₂ch.stand) and -1.8 ± 0.17 Mg C-CO₂ ha⁻¹ for Eddy Covariance (NEE.EC)
542 (Table 4).

543 Table 4: Annual budget of CO₂ fluxes based on Eddy Covariance (EC) and upscaled chamber
 544 methods (Ch.stand).

	Annual sum (Mg C-CO ₂ ha ⁻¹)	Std error (Mg C-CO ₂ ha ⁻¹)
Reco.EC or Rch.stand		
EC	10.0	0.49
Ch.stand	4.1	0.18
GPP.EC or GPPch.stand		
EC	-11.8	0.53
Ch.stand	-5.5	0.83
NEE.EC or FCO₂ch.stand		
EC	-1.8	0.17
Ch.stand	-1.4	0.46

545 Annual sum corresponds to the annual cumulative fluxes for full year measurements (Mg C-CO₂ ha⁻¹). EC
 546 refers to fluxes measured by the Eddy Covariance method, and Ch refers to the fluxes measured by
 547 chambers, which are then upscaled to the whole field. Rch.stand represents the chamber respiration, while
 548 Reco.EC denotes the ecosystem respiration according to the EC method. GPP.EC and GPPch.stand are the
 549 gross primary production or photosynthesis flux, measured by EC and Ch methods, respectively. NEE.EC
 550 and FCO₂ch.stand represent the net ecosystem exchange for EC and Ch, respectively. The associated
 551 standard error is denoted as Std error (Mg C-CO₂ ha⁻¹). Positive values indicate CO₂ emissions, while
 552 negative values represent CO₂ uptake.

553 4. Discussion

554 4.1. Soil respiration modeling and limitations

555 In this study the Lloyd and Taylor (1994) model, based on a modified Arrhenius-type formulation,
556 was used to model nocturnal soil respiration fluxes for estimating daytime respiration. Unlike the
557 classical Arrhenius equation, this model includes the $(T_{\text{soil}} - T_0)$ term in the denominator of the
558 exponential expression (Eq. 4), which inherently limits the effects of high temperatures by
559 progressively reducing the temperature sensitivity of soil respiration as temperatures rise above
560 a given threshold. This structural feature produces a flattening of the respiration-temperature
561 relationship at elevated temperatures, thereby preventing the overestimations (Lloyd and Taylor,
562 1994).

563 The Lloyd and Taylor model has successfully been widely applied, primarily in boreal and
564 temperate ecosystems (Lasslop et al., 2010; Reichstein et al., 2003), and relies on the assumption
565 of comparable thermal conditions between daytime and nighttime periods (Juszczak et al., 2012).
566 In our study, instantaneous soil temperatures ranged from 20.7 to 45.8 °C during the day and from
567 22.1 to 45.0 °C at night, indicating largely overlapping thermal ranges between the two periods.
568 Model parameters were recalibrated using five-day fixed windows, which provided sufficient
569 temporal resolution while capturing seasonal dynamics of soil respiration.

570 This study represents one of the first applications of the Lloyd and Taylor model in a Sahelian
571 semi-arid context. While Arrhenius based models are known to potentially overestimate fluxes
572 under extreme temperatures due to physiological limitations, over the range of temperatures
573 observed in this study, the modeled soil respiration was not overestimated (Fig. S6, a and b). Thus,
574 the model used in this study appears to provide a realistic representation of soil respiration under
575 local conditions. However, this conclusion is site-specific and should not be interpreted as a
576 general validation of temperature-based models across all semi-arid environments. Such models
577 should be systematically validated with respect to temperature to ensure their reliability.

578 4.2. Seasonality and drivers of chamber-based CO₂ fluxes

579 In our agroforestry context, seasonal variability in CO₂ fluxes closely followed rainfall dynamics,
580 peaking during the wet season and declining sharply in the dry season, consistent with soil
581 moisture depletion and crop senescence. This pattern is typical of semi-arid ecosystems (Ago et
582 al., 2016a; Brümmer et al., 2008; Guillen-Cruz et al., 2023; Macharia et al., 2020; Mosongo et al.,
583 2022; Wieckowski et al., 2024).

584 Respiration and photosynthesis were primarily driven by soil moisture and LAI, reflecting the
585 system's sensitivity to water availability and crop dynamics. Soil moisture enhanced both
586 processes by stimulating microbial activity and supporting plant growth (Borken et al., 2002;

587 Conant et al., 2004; Merbold et al., 2009; Yu et al., 2020; Zhao et al., 2016). The stronger correlation
588 between soil moisture and respiration under *F. albida* canopy (Sh: $r = 0.78$) compared to full sun
589 (FS: $r = 0.51$) suggests greater microbial sensitivity to moisture beneath trees. This likely reflects
590 enhanced substrate availability, resulting in stronger post-rainfall respiration pulses (Meisner et
591 al., 2015) and supporting the 'fertile island' effect, where trees improve local soil conditions
592 (Eldridge et al., 2024). Photosynthetic capacity also responded to soil moisture, as shown by
593 positive correlations with LAI and key physiological traits such as light use efficiency (α) and
594 maximum CO₂ uptake rate (β) (Gonsamo et al., 2019; Qiu et al., 2023; Zhang et al., 2024).
595 In contrast, the influence of soil temperature (T_{soil}) on respiration was weakly negative in both FS
596 and Sh, indicating a thermal threshold beyond which respiration declined —estimated at 32 ± 1.5
597 °C in FS and 29.5 ± 1.9 °C in Sh (Fig. S9, a and b), similar to findings in Eastern Ghana (Owusu et
598 al., 2024). This inhibition likely results from decreased enzymatic and microbial activity under
599 combined heat and water stress (Liu et al., 2018; Richardson et al., 2012). In semi-arid regions,
600 soil respiration often becomes decoupled from temperature due to seasonal moisture constraints
601 (Jia et al., 2020; Tucker & Reed, 2016; Warren, 2014), with microbial activity limited during dry
602 periods despite favourable temperatures. This decoupling helps explain the weak or absent
603 correlation between T_{soil} and soil moisture (Fig. S8, b), particularly under Sh ($r = -0.28$).
604 Management practices such as organic inputs can also modulate these dynamics, adding further
605 variability to soil respiration responses (Meena et al., 2020; Oyonarte et al., 2012; Rong et al.,
606 2015; Xue & Tang, 2018).

607 4.3. Magnitude of chamber-based total CO₂ respiration fluxes

608 Mean total soil respiration values were consistent with those reported in other low-input
609 agricultural systems across sub-Saharan Africa (Mapanda et al., 2010; Pelster et al., 2017;
610 Rosenstock et al., 2016). In full sun (FS), the mean respiration (1.0 ± 0.9 g C-CO₂ m⁻² d⁻¹) closely
611 matched values measured by Wachiye et al. (2020) in a semi-arid Kenyan field at 1158 m altitude
612 (1.1 ± 0.1 g C-CO₂ m⁻² d⁻¹). This similarity likely reflects comparable environmental conditions,
613 including moderate rainfall (~ 550 mm yr⁻¹) and low soil organic carbon and nitrogen contents
614 (<1%) in the 0–20 cm layer of sandy soil. In contrast, respiration under *F. albida* canopy (Sh: 1.6
615 ± 1.1 g C-CO₂ m⁻² d⁻¹) was higher, likely due to additional autotrophic respiration from tree roots
616 and greater organic inputs beneath the canopy. Nonetheless, this flux remains close to values
617 observed in low-input sorghum fields on sandy loam soils in eastern Ghana (1.7 ± 1.1 g C-CO₂ m⁻²
618 d⁻¹), despite higher rainfall (950–1000 mm yr⁻¹) in that region (Owusu et al., 2024).
619 Cumulative annual respiration fluxes fell within the range reported for Sahelian croplands (250–
620 450 g C-CO₂ m⁻²) (Brümmer et al., 2009) and other sub-Saharan African agricultural systems (Kim
621 et al., 2016). The cumulative flux under tree cover is similar to that measured in cassava fields in

622 eastern Tanzania ($440 \text{ g C-CO}_2 \text{ m}^{-2} \text{ yr}^{-1}$), despite the latter receiving higher rainfall ($\sim 1115 \text{ mm}$
623 yr^{-1}) (Rosenstock et al., 2016). This convergence may stem from comparable soil fertility
624 constraints, with low soil organic carbon (1–1.7%) and nitrogen contents ($<0.5\%$). In contrast,
625 the slightly lower cumulative flux in FS may reflect less favourable microclimatic conditions—
626 such as elevated soil temperatures and increased aridity away from tree cover—limiting
627 microbial activity (see Section 4.1).

628 Across sub-Saharan Africa, soil respiration fluxes based on static chamber measurements show
629 high spatial variability, largely shaped by climate and land use. For example, Owusu et al. (2024)
630 found higher respiration in woodlands ($3.8 \pm 0.8 \text{ g C-CO}_2 \text{ m}^{-2} \text{ d}^{-1}$) and grazed areas (2.7 ± 1.7)
631 than in croplands (1.7 ± 1.1) in humid eastern Ghana. This gradient was linked to differences in
632 soil moisture and organic matter. Similarly, Rosenstock et al. (2016) reported much higher fluxes
633 in highland pastures in Kenya ($3.8\text{--}4.4 \text{ g C-CO}_2 \text{ m}^{-2} \text{ d}^{-1}$) compared to cultivated fields in eastern
634 Tanzania (1.2 ± 0.2), highlighting the role of vegetation cover and soil fertility.

635 4.4. Effect of trees on chamber-based soil respiration and photosynthesis

636 A notable increase in respiration and photosynthesis fluxes was observed under *F. albida* trees
637 (Sh) compared at a distance from trees (FS). This increase may indicate the potential role of *F.*
638 *albida* in modulating CO_2 exchange dynamics (Rch and GPPch) within this agro-silvo-pastoral
639 system. These results are consistent with preliminary findings from similar environments
640 (Duthoit et al., 2020).

641 Numerous studies have investigated the effect of tree species on greenhouse gas fluxes,
642 particularly CO_2 , revealing significant variations across different ecological contexts (Bréchet et
643 al., 2021, 2025; Klaus et al., 2024; Mazza et al., 2021; Ramesh et al., 2013; Rheault et al., 2024).
644 However, the underlying mechanisms by which trees influence these dynamics are not yet fully
645 understood.

646 In general, agroforestry systems have been well-documented for their ability to provide a range
647 of ecosystem services (e.g., Assefa et al., 2024; Bado et al., 2021; Kuyah et al., 2019; Rolo et al.,
648 2023). Specifically, *Faidherbia*-based agroforestry systems may play a crucial role in regulating
649 CO_2 exchanges between the soil and atmosphere. *F. albida*-based agroforestry systems are
650 recognized for enhancing both soil organic and mineral fertility (Bayala et al., 2020; Dilla et al.,
651 2019; Sileshi, 2016; Sileshi et al., 2020; Stephen et al., 2020), mainly through litter accumulation
652 and direct inputs from livestock excreta under their canopies. Additionally, the extensive root
653 system of *F. albida* trees helps concentrate mineral nutrients, contributing to the formation of a
654 'fertile island' effect under the trees (Siegwart et al., 2022; Eldridge et al., 2024). Moreover, *F.*
655 *albida* improve water infiltration (Diongue et al., 2023; Faye et al., 2020; Sarr et al., 2023), enhance
656 soil moisture retention (Clermont-Dauphin et al., 2023) and contribute to reduced soil

657 temperatures (de Carvalho et al., 2021; Lopes et al., 2024; Sida et al., 2018). These changes foster
658 a more favourable environment for soil microbial activity and crop development (Diack et al.,
659 2024; Diene et al., 2024; Leroux et al., 2020; Roupsard et al., 2020) under the trees compared to
660 open areas. Consequently, this likely explains the stronger effect of soil moisture and the leaf area
661 index of groundnuts on Rch under the trees, resulting in higher total respiration (Table 2). For
662 photosynthesis, the effect of these parameters was similar in both FS and Sh (Table 2). However,
663 the significantly higher intensity of GPPch under Sh can be explained by greater light use efficiency
664 (α) and a higher maximum CO₂ uptake rate at light saturation (β) in this shaded environment. In
665 agroforestry systems, light use efficiency can at least partially mitigate the reduction in
666 photosynthetically active radiation under tree canopies (Charbonnier et al., 2017).
667 Similar results have been observed in different climatic conditions and ecosystems. Gomes et al.
668 (2016) investigated soil respiration using mobile chambers (LI-8100-102 model) under trees in
669 coffee-based agroforestry systems and in the open areas in Minas Gerais, Brazil. These studies
670 were conducted with agroecological management practices, such as weeding, intercropping maize
671 between coffee rows, and mulching. The agroforestry systems exhibited lower air and soil
672 temperatures (at 5 and 10 cm depth) and higher air and soil humidity compared to the open areas
673 (Gomes et al., 2016). These authors observed greater spatial variability in soil respiration in
674 agroforestry system (34.1%) compared to the open areas (24.2%). This variability was mainly
675 linked with fluctuations in labile carbon and total nitrogen, reflecting more favourable soil
676 microclimate for microbial activity in agroforestry system. In contrast, soil temperature (10 cm
677 depth) accounted for most of the variability observed in the open areas, where the absence of tree
678 canopy resulted in high soil temperatures and low soil moisture (Gomes et al., 2016). Likewise,
679 Haren et al. (2010) reported 38% higher soil respiration near large trees (DBH > 35 cm) in clay-
680 rich Amazonian forests compared to open sites. Interestingly, the magnitude of CO₂ fluxes was
681 independent of tree species, indicating that canopy effects may outweigh species-specific traits in
682 some contexts. In our study, *F. albida*'s influence on CO₂ fluxes aligns with this general pattern
683 observed in tropical agroforestry. However, the mechanisms linking individual tree species to
684 microbial and physicochemical drivers of CO₂ dynamics remain insufficiently understood and
685 warrant further investigation (Jevon et al., 2023).

686 4.5. Birch Effect

687 A rapid increase in soil respiration was observed following the first rainfall events, particularly
688 under *F. albida*. This phenomenon can be attributed to the lower bulk density of the soil under the
689 trees (Clermont-Dauphin et al., 2023; Siegwart et al., 2023), which potentially lead to CO₂
690 accumulation during the dry season due to higher soil organic matter (SOM) (Siegwart et al.,
691 2023). Additionally, the sensitivity of microbial communities to subtle variations in soil moisture,

692 compounded by the tree effect, may further explain this phenomenon, as outlined in Sections 4.1
693 and 4.3. This phenomenon, known as the 'Birch effect' (Birch, 1958), has been reported across
694 various semi-arid ecosystems in sub-Saharan Africa (Ago et al., 2016b; Fan et al., 2015;
695 Wieckowski et al., 2024), as well as other semi-arid ecosystems globally (Roby et al., 2022; Yan et
696 al., 2014; Yu et al., 2020). In these contexts, the 'Birch effect' may result from the displacement of
697 soil gas phases by the piston effect generated during water infiltration (Singh et al., 2023).
698 Furthermore, microbial communities in semi-arid environments adopt osmoregulatory
699 mechanisms to withstand water deficit (Warren, 2014), which is particularly pronounced during
700 the dry season. This phenomenon reduces soil microbial metabolism (Schimel et al., 2007). Upon
701 rapid soil rewetting, especially after prolonged dry periods, soil microbial metabolism process is
702 swiftly reactivated, leading to a transient pulse in respiration and a CO₂ release (Barnard et al.,
703 2020; Kim et al., 2012; Manzoni et al., 2020; Vargas et al., 2018). Isotopic signatures of soil
704 respiration provide evidence supporting the hypothesis that these pulses result from the rapid
705 mineralisation of necromass or osmolytes excreted by microorganisms under drought stress
706 (Schimel et al., 2007; Unger et al., 2010). Additional factors may amplify the 'Birch effect'. For
707 instance, drying-rewetting cycles can induce physical disruption of soil aggregates, enhance
708 oxygen penetration and thereby expose previously protected organic matter to microbial
709 decomposition (Rabbi et al., 2024). This increases substrate availability and subsequently boosts
710 soil respiration fluxes.

711 The magnitude of the 'Birch effect' is modulated by the severity and duration of drought. Thus, at
712 our study site, given the 8- to 9-month-long dry season, the 'Birch effect' is particularly intense.
713 Indeed, extended drought periods promote greater accumulation of microbial necromass and
714 intensify hypo-osmotic stress responses upon rewetting (Singh et al., 2023).

715 *4.6. Comparing EC and chamber-based methods*

716 Results revealed high seasonal variability, with higher values during the rainy season compared
717 to the dry season. This seasonal pattern aligns with findings from studies in the Sahel using the
718 EC method for CO₂ flux measurements (Brümmer et al., 2008; Tagesson et al., 2015; Agbohessou
719 et al., 2023, Wieckowski et al., 2024). Comparable patterns have also been documented at the
720 ecosystem scale in other semi-arid environments (Ago et al., 2014; Archibald et al., 2009; Ardö et
721 al., 2008; Jia et al., 2020; Quansah et al., 2015; Williams et al., 2009; Zhang, Bi, et al., 2024).

722 Several comparative studies between chamber and EC methods have reported both congruent
723 and divergent CO₂ flux estimates (Bastviken et al., 2022; Poyda et al., 2017; Riederer et al., 2014;
724 Tang et al., 2008; Wang et al., 2010). In the present study, ecosystem respiration fluxes during the
725 rainy season exhibited notable discrepancies measurements between EC (Reco.EC) and upscaled
726 chamber-based (Rch.stand). This is attributable to differences in the flux components captured

727 by each method. Specifically, Reco.EC included respiration from below- and above-ground tree
728 parts, crops (groundnuts and cowpeas), weeds, and soil, whereas Rch.stand accounted only
729 respiration from below-ground tree, groundnut crop, and soil. Therefore, as expected, Reco.EC
730 (4.6 ± 3.2 g C-CO₂ m⁻² d⁻¹) were significantly higher than Rch.stand (2.0 ± 1.1 g C-CO₂ m⁻² d⁻¹).
731 For chamber-based GPP measurements, values were standardised (GPP-stand) by the field leaf
732 area index (LAI.field). This allowed it to improve comparability with GPP.EC when trees were
733 leafless in the rainy season. In both cases, GPP accounted only for crops (groundnut and cowpea)
734 and weeds, as trees were non-photosynthetic in the rainy season. Despite this standardisation,
735 GPP.EC values (-5.1 ± 3.6 g C-CO₂ m⁻² d⁻¹) were significantly higher than GPPch.stand values (-4.2
736 ± 4.3 g C-CO₂ m⁻² d⁻¹). However, no divergence was observed in August, and the intensity of the
737 peak of GPP in September was similar in both methods, but from the onset of groundnut
738 senescence, when weeds became the dominant photosynthetic contributors. Thus, during the
739 groundnut growth season, with leafless *F. albida* trees and almost no weeds, GPP measurements
740 from EC and chambers generate closely comparable results. Therefore, this provides an initial
741 form of cross-validation between the two methods. It is important to note that the EC method
742 integrates CO₂ fluxes over a larger spatial scale, encompassing all ecosystem components
743 (Baldocchi, 2003), while the chamber method captures fluxes on a smaller scale (i.e., at the 0.25
744 m² scale). This scale disparity can introduce uncertainties when upscaling chamber-based fluxes
745 to the field, as vegetation composition within chambers does not represent the EC footprint
746 average vegetation. This makes upscaling chamber-based measurements challenging.
747 Nevertheless, the standardisation we applied on chamber photosynthesis by LAI has been
748 relatively successful.

749 During the dry season, Reco.EC included respiration from below- and above-ground tree parts
750 (with leaves) and bare soil, while Rch.stand measured only below-ground tree and bare soil
751 respiration. Consequently, the difference between Reco.EC and Rch.stand was solely attributable
752 to above-ground tree respiration (Ra tree above-ground). In terms of GPP, chamber
753 measurements were nil, whereas GPP.EC reflected only GPP trees.

754 The transition period, characterised by groundnut senescence, tree leaf regrowth, and weed
755 proliferation, introduced further complexity, amplifying method-specific discrepancies. Rch.stand
756 measurements facilitated the estimation of tree contribution to Reco.EC (Ra tree) and the
757 verification of the consistency for EC results in terms of carbon use efficiency (CUE), estimated
758 here at 0.48. This value indicates that nearly 50% of the carbon captured by trees is allocated to
759 biomass. The CUE estimate here is well comparable to the global average across diverse
760 ecosystems, climates, and management practices (0.49 ± 0.14) (Tang et al., 2019). Similar CUE
761 values have been reported for semi-arid grasslands (0.46 ± 0.10), but our value is notably lower
762 than those documented for wetlands (0.61 ± 0.13) (Tang et al., 2019). Overall, these findings

763 reinforce the plausibility of our assumptions regarding the compartment contributions to Reco.EC
764 and Rch.stand, thereby providing a second cross-validation of the EC-Ch comparison. However,
765 despite a frequently assumed CUE of 0.5 in models, global estimates span a broad range (0.20 to
766 0.82), depending on ecosystem type and management practices (DeLucia et al., 2007; Tang et al.,
767 2019). This underscores the importance of refining carbon flux models to better represent the
768 biophysical processes governing CO₂ exchange in semi-arid agroforestry systems. The combined
769 use of EC and chamber methodologies offers a comprehensive perspective on ecosystem-scale CO₂
770 flux dynamics, advancing the understanding of carbon cycling in these environments.

771 4.7. *Net annual vertical carbon balance*

772 The net annual carbon balance was quantified at -1.4 ± 0.46 Mg C-CO₂ ha⁻¹ with the chamber
773 method and -1.8 ± 0.17 Mg C-CO₂ ha⁻¹ by the Eddy Covariance (EC), indicating that the studied
774 agro-silvo-pastoral system functions as a net carbon sink. These findings corroborate the system
775 potential role in mitigating greenhouse gas emissions, consistent with previous studies reporting
776 vertical CO₂ flux balances in semi-arid ecosystems (Rahimi et al., 2021; Tagesson et al., 2015;
777 Agbohessou et al., 2023, Wieckowski et al., 2024).

778 The estimated net C exchange balance is close to the reported mean for Sahelian ecosystems (-1.6
779 ± 0.5 Mg C-CO₂ ha⁻¹; Tagesson et al., 2016). The EC-based net C exchange balance (-1.8 ± 0.17 Mg
780 C-CO₂ ha⁻¹) is also similar to the value of -1.9 ± 0.4 Mg C-CO₂ ha⁻¹ reported for semi-arid savannas
781 of northeastern Benin, despite higher annual rainfall (1495 mm; Ago et al., 2016b). Furthermore,
782 our EC estimate is close to the average net C exchange reported for West African terrestrial
783 ecosystems (-2.0 ± 1.5 Mg C-CO₂ ha⁻¹; Ago et al., 2016a).

784 However, estimates from Tagesson et al. (2015) (-2.7 ± 0.07 Mg C-CO₂ ha⁻¹) for a semi-arid
785 savannah in Dahra, Senegal, located between the 300 mm and 400 mm isohyets, were
786 comparatively higher. This is potentially attributable to specific characteristics of that specific
787 savannah site, such as herbaceous vegetation cover during the rainy season, the presence of
788 evergreen trees, and land management practices linked to pastoral livestock activities (Tagesson
789 et al., 2016).

790 The net annual C balance estimates presented in this study are, in fact, representing vertical fluxes
791 only, given that they exclude organic matter (OM) imports and, more critically, exports,
792 introducing uncertainties. Notably, the export of crop residues and direct inputs from animal
793 excreta —particularly significant in ‘bush fields’ during the dry season — were not accounted for.
794 In our case of ‘bush field’, crop residues are exported to feed livestock, while livestock faeces are
795 collected for use as fuel or manure in ‘home fields’. Such practices may lead to a significant soil
796 organic carbon stocks depletion (Malou et al, 2021), potentially diminishing the net C budget (-

797 $1.4 \pm 0.46 \text{ Mg C-CO}_2 \text{ ha}^{-1}$) over time and shifting the system closer to carbon neutrality (Assouma
798 et al., 2019).

799 These results should be contextualized within the broader framework of climate change and semi-
800 arid ecosystem management. Although agro-silvo-pastoral systems can function as apparent
801 annual carbon sinks, they remain highly sensitive to interannual rainfall variability and escalating
802 anthropogenic pressures. Sustainable management practices, particularly regarding C
803 inputs/outputs from the system regarding crop harvest, residues exportation, and cattle free
804 manuring, must be taken into account to confirm the capacity of the system to act as effective a
805 carbon sink.

806 *4.8. Limitations of the study*

807 This study benefited from the inverse phenology of *F. albida*, allowing for direct comparison
808 between chamber-based GPP (GPPch.stand) and ecosystem-level GPP (GPP.EC) during the
809 leafless period of the trees. However, the system spatial heterogeneity —common in
810 agroforestry— posed challenges for accurately partitioning CO₂ fluxes among trees, crops, and
811 soil. A key limitation was the development of weeds during the late rainy season, which
812 complicated the attribution of fluxes, particularly during the transitional period. Additionally,
813 while GPPch was successfully standardised by LAI for upscaling, this was not feasible for
814 respiration. Respiration integrates both autotrophic and heterotrophic components, which
815 respond to different drivers and are not directly linked to LAI, limiting the precision of upscaled
816 Rch.

817 Future improvements should aim to separately quantify respiration sources —tree roots, crops,
818 and microbial (heterotrophic) respiration— and account explicitly for the weed layer, to refine
819 flux partitioning in such complex agroforestry systems.

820 Furthermore, the present study constitutes only an intermediate step delivering a first integrated
821 estimate of the main vertical CO₂ exchanges (photosynthesis, respiration, and net ecosystem
822 exchange) as a base for a forthcoming paper that will present a more comprehensive carbon
823 budget of the ecosystem. Establishing such a carbon budget would require substantial additional
824 data acquisition and poses considerable methodological challenges. In particular, quantifying
825 carbon inputs/outputs associated with free-ranging livestock grazing would be difficult to achieve
826 with acceptable accuracy. It must also be recognised that the system is in a dynamic, non-steady
827 state, characterised by marked inter-annual variability as well as periods of carbon storage and
828 release, which are difficult to constrain empirically except through modeling.

829 **Conclusion**

830 This study demonstrates the successful application of automated dynamic chambers to quantify
831 CO₂ fluxes in a Sahelian agroforestry system dominated by *F. albida*. The continuous, high-
832 frequency measurements captured key seasonal dynamics and short-lived events (e.g., Birch
833 effect), providing a more accurate assessment of carbon exchange than traditional intermittent
834 sampling.

835 By integrating crop and soil components and applying dynamic partitioning models, the study
836 quantified both respiration and photosynthesis fluxes at fine temporal resolution. The results
837 revealed a clear 'fertile island' effect under tree canopies, with higher respiration and
838 photosynthetic activity, and highlighted the significant contribution of *F. albida* trees to annual
839 carbon uptake.

840 The consistency between chamber- and eddy covariance-based estimates reinforces the
841 robustness of the methodology. Overall, this work underscores the role of *F. albida*-based
842 agroforestry systems in the dynamic of C exchanges in semi-arid environments, offering valuable
843 insights for carbon accounting and sustainable land management in the Sahel.

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859 **Author contribution: CRediT**

860 **Seydina Mohamad BA:** Conducting in situ experiments, collecting and processing data, writing-
861 original draft, review and editing. **Olivier Roupsard:** Designing experimental apparatus and
862 methodology, writing, review and editing. **Lydie Chapuis-Lardy:** Designing methodology, writing,
863 review and editing. **Yélognissè Agbohessou:** Processing data, review and editing. **Fred Bouvery:**
864 Designing chambers and connection to the instrument, review and editing. **Maxime Duthoit:**
865 Designing experimental set and methodology, review and editing. **Aleksander Wieckowski:**
866 Review and editing. **Mohamed Habibou Assouma:** Review and editing. **Espoir Gaglo:** Processing
867 data, review and editing. **Claire Delon:** review and editing. **Torbern Tagesson:** Designing
868 methodology, review, and editing. **Bienvenu Sambou:** Review and editing. **Dominique Serça:**
869 Designing methodology, writing, review and editing.

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