



Global quantification of the eco-hydrological co-benefits of soil carbon sequestration

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Abstract. Soil carbon sequestration is an important strategy for climate change mitigation with several co-benefits, including increased water holding capacity and infiltration. However, a global-scale quantification of hydrological co-benefits for water availability to plants is still lacking. In this study, we investigate the effect of soil carbon sequestration on hydrology and water resources by conducting experiments with the Community Terrestrial Systems Model (CTSM). Using global experiments with spatially explicit soil organic carbon (SOC), we apply various carbon sequestration scenarios, including one aligned with the ‘4 per 1000’ initiative, to investigate the effect on soil moisture and soil water balance variables with a focus on cropland regions. Our results show that soil organic carbon redistributes water within the soil profile, retaining moisture in the rooting zone and limiting percolation into deeper layers, which is particularly pronounced in arid regions with sandy soils. Under a scenario with a uniform SOC increase of gC kg^{-1} soil, globally averaged total global soil liquid water content increases by 4 mm in the first 30 cm. Carbon sequestration also redistributes the soil water balance, with global mean reductions in surface runoff (–1 mm), subsurface runoff (–0.6 mm), and an increase in evapotranspiration (+2 mm), contributing to improved vegetation productivity. Water stress is modestly reduced across most regions, though effects vary spatially. Although the hydrological impacts of soil carbon sequestration are generally small in magnitude, they are consistent and systematic. The relative changes following realistic and policy-relevant SOC enhancement scenarios, such as those under the 4 per 1000 initiative, are limited due to the modest carbon additions involved. Nevertheless, these changes offer measurable eco-hydrological co-benefits that may support both climate mitigation and ecosystem resilience, particularly in water-limited environments.



1 Introduction

Soil carbon sequestration is widely recognized as a key strategy for climate change mitigation, through storing carbon from anthropogenic CO₂ in soils (Amelung et al., 2020; Paustian et al., 2016; Lal, 2004). It leverages sustainable land management practices, such as cover cropping, mulching, conservation tillage, organic manure application, and agroforestry, to increase the soil's capacity to capture and retain carbon (Paustian et al., 2016; Bossio et al., 2020). Soil organic carbon (SOC) stocks have declined globally due to land use and land cover changes, resulting in an estimated loss of approximately 133 PgC (Sanderman et al., 2017). Regions experiencing the highest SOC losses are typically cropland and grazing areas. By implication, these soils hold the highest potential for restoration through sequestration (Sanderman et al., 2017). Global estimates of carbon sequestration potential in croplands range from 29 to 65 PgC over a period of at least 20 years (Zomer et al., 2017; Padarian et al., 2022).

Soil carbon sequestration through sustainable practices is widely recognized as a nature-based solution for climate mitigation, intended to complement efforts to reduce greenhouse gas emissions (Paustian et al., 2016; Minasny et al., 2017; Chabbi et al., 2017). Soil carbon sequestration has been endorsed by the international 4 per 1000 initiative (4p1000), launched at COP21 in Paris, which advocates for increasing global SOC stocks by 0.4% per year to contribute to offsetting greenhouse gas emissions (Rumpel et al., 2018, 2020). The initiative has highlighted the potential of soil carbon storage for climate mitigation while also initiating discussions about the complexities of its implementation, including soil carbon dynamics, regional variability, socio-economic factors, and the importance of complementing sequestration efforts with reductions in fossil fuel emissions (de Vries, 2018; Lal, 2016; Rumpel et al., 2020).

In addition to its mitigation potential, soil carbon sequestration provides adaptation potential through various benefits such as improving soil health, reducing erosion, and enhancing biodiversity (Chabbi et al., 2017; Amelung et al., 2020; Bossio et al., 2020). Additionally, biogeophysical effects may arise through the influence of SOC on soil hydraulic properties. SOC stabilizes soil aggregates and increases porosity, thereby enhancing water retention, infiltration, and overall water holding capacity (Bowling et al., 2020; Arenas-Calle et al., 2021). In cropland soils, enhanced carbon sequestration has the potential to mitigate drought-induced impacts such as delaying the drought onset because of wetter soils (Turek et al., 2023), or increase drought resilience, thereby increasing crop yields (Iizumi and Wagai, 2019; Kane et al., 2021).

The extent to which SOC influences soil hydraulic properties depends on factors such as soil texture and climate conditions. Coarse-textured soils are more effective at retaining water for plant use, as they typically have lower wilting points, while soils with high clay content may retain more total water but have less plant-available water due to higher wilting points (Minasny and McBratney, 2018; Williams et al., 2016). SOC benefits tend to be more apparent in dry regions and in soils with low water holding capacity and large interannual fluctuations in climate conditions, such as temperature and precipitation (Iizumi and Wagai, 2019; Williams et al., 2016). Through its control on soil water retention, SOC should also moderate plant water avail-



ability and thus affect all biogeophysical and biogeochemical processes at the land-atmosphere interface. However, despite the widely recognized increase in the water holding capacity of soils, the effect of SOC on available water capacity to plants is less clear. A meta-analysis of 60 studies by Minasny and McBratney (2018) found that a 1% mass increase (10 gC kg^{-1} soil) in SOC corresponds to only a modest 1.16 mm water per 100 mm soil in available water capacity, suggesting that the influence of SOC on plant-available water may be limited.

The effects of SOC changes on soil moisture have mostly been investigated using local-scale empirical or modeling studies (e.g. Jordán et al., 2010; Turek et al., 2023), meta-analyses (Minasny and McBratney, 2018) and regional or global statistical analyses (e.g. Iizumi and Wagai, 2019; Kane et al., 2021). However, insights from global or regional land modeling experiments remains sparse. McDermid et al. (2022) used the NASA Goddard Institute for Space Studies (GISS) ModelE, an Earth System Model, to explore the impacts of SOC loss from soil degradation on soil hydraulic properties and moisture retention. Their findings indicated that reductions in SOC lead to lower porosity across agricultural lands, causing reductions in total soil water content at regional scales. Other studies with land and climate models investigate changes in soil carbon from a biogeochemical and carbon cycle perspective for model evaluation and process understanding (Luo et al., 2016; Ito et al., 2020).

Despite these advances, a global quantification of the effects of policy-relevant soil carbon sequestration on soil moisture and related water balance variables is still lacking. Such a comprehensive analysis helps identify the hydrological co-benefits of soil carbon sequestration. Global-scale assessments are particularly relevant to inform policy initiatives, such as the 4 per 1000 initiative, and to enable comparisons across regions with varying soil textures, hydro-climatic conditions, and vegetation cover.

In this study, we address this gap by using the Community Terrestrial Systems Model (CTSM) to identify the hydrological co-benefits of soil carbon sequestration under policy-relevant scenarios. Thereby, we analyze the impact of carbon sequestration on soil hydrology and the implications for soil hydraulic properties and soil water balance related variables including soil evaporation, vegetation evapotranspiration, runoff and subsurface drainage. To this end, we use scenarios representing high- and low-end sequestration rates alongside a scenario aligned with the 4 per 1000 initiative.

2 Material and Methods

2.1 Model description

The Community Terrestrial Systems Model (CTSM version 5.2) is an advanced land model that simulates physical, chemical, and biological processes in terrestrial ecosystems and climate across varying spatial and temporal scales (?). Land surface heterogeneity is captured through a nested subgrid hierarchy, where each grid cell consists of multiple land units representing lakes, urban areas, crops, glaciers, and vegetated areas. Each land unit includes one or more columns that define the state



85 variables for soil energy and water content. Columns host patches that represent distinct Plant Functional Types (PFTs) or bare
ground for vegetated units, and different crop functional types for cropland units. In total, 16 PFTs are defined, each varying in
physiology and structure. In the default CTSM configuration, PFTs may share a soil column, competing for water and energy.
In our set-up however, we conducted additional simulations where each PFT resides on its own soil column to investigate the
impact of plant competition for water on carbon sequestration scenarios. This setup isolates the effects of transpiration and root
90 water uptake specific to each PFT.

When the crop model is inactive, crops are represented by one irrigated and one unirrigated unmanaged C3 crop, treated
similarly to C3 grass (Lawrence et al., 2018). Effects of land management on soil carbon, such as harvesting, tillage and erosion
are not included in this modeling setup. Land unit distributions are prescribed using the default land cover dataset from the
95 Land Use Harmonization Project 2 (LUH2; ?, appendix Fig. A2).

The default soil column configuration in CTSM includes 25 layers with a total soil depth of about 50 m. The soil is hydrologically active in the top 20 layers (down to 8.6 m depth). Soil layer depths and thicknesses follow the default CTSM parameterization (Table A1). The soil depth to bedrock is spatially variable and prescribed using data from (Pelletier et al.,
100 2016, , Appendix Fig. A1), which constrains the number of hydrologically active layers to the maximum soil depth for each grid cell. One dimensional water flow is described, following the (Clapp and Hornberger, 1978) parametrization, by Darcy's law and Richards equation with the hydraulic conductivity (k , mm s^{-1}) and soil matric potential (ψ , mm), which are both varying per soil layer with volumetric soil water (θ , $\text{mm}^3 \text{mm}^{-3}$), soil texture and soil organic matter (Lawrence et al., 2018).

105 Soil texture and soil organic matter are prescribed based on spatial datasets of clay and sand percentages and organic matter density (kg OM m^{-3}). We use the WISE30sec dataset (Batjes, 2016), a harmonized soil profile database based on the World Inventory of Soil property Estimates (WISE), which provides global soil property estimates up to 2 m depth at a 30-arcsecond resolution. Percentages of clay and sand from WISE30sec are used as direct input to CTSM (Fig 1b,c), while soil organic carbon (SOC , gC kg^{-1} soil) from WISE30sec is converted to organic matter density (OM, kg OM m^{-3}) using the formula:

$$110 \quad \text{OM} = \text{SOC} \cdot \rho_{\text{soil}} \cdot (100 - f_{\text{coarse}}) / 100 \cdot 1/0.58 \quad (1)$$

using the bulk density (ρ_{soil} , g cm^{-3}) and fraction of coarse fragments (f_{coarse} , volumetric, %), and 1 g OM is equivalent to the van Bemmelen conversion factor of 0.58 gC (Fig. 1a). By using SOC and texture values from the same dataset, we ensure consistency across soil characteristics. Within CTSM, organic matter is further converted to an organic fraction using a maximum density of 130 kg m^{-3} corresponding to 100%, corresponding to the standard organic matter density of peat soils.

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The volumetric water content at saturation (i.e., porosity) for soil layer i ($\theta_{\text{sat},i}$, $\text{mm}^3 \text{mm}^{-3}$) is computed as a weighted average of the organic and mineral soil components:



$$\theta_{\text{sat},i} = (1 - f_{\text{om},i}) \cdot \theta_{\text{sat,min},i} + f_{\text{om},i} \cdot \theta_{\text{sat,om}} \quad (2)$$

where $f_{\text{om},i}$ is the soil organic matter fraction, $\theta_{\text{sat,om}}$ ($\text{mm}^3 \text{mm}^{-3}$), is the porosity of organic matter (set to 0.9; Letts et al., 2000), and $\theta_{\text{sat,min},i}$ ($\text{mm}^3 \text{mm}^{-3}$) the porosity of the mineral fraction, which depends on the sand fraction ($f_{\text{sand},i}$) as:

$$\theta_{\text{sat,min},i} = 0.489 - 0.00126 \cdot f_{\text{sand},i} \quad (3)$$

The volumetric water content at the permanent wilting point (θ_{wp} , $\text{mm}^3 \text{mm}^{-3}$) is calculated from the soil water retention function:

$$\theta_{\text{wp}} = \theta_{\text{sat}} \cdot \left(\frac{\psi_{\text{wp}}}{\psi_{\text{sat}}} \right)^{-1/B} \quad (4)$$

where $\psi_{\text{wp}} = -150000$ mm, the soil matric potential at wilting point, B the Clapp and Hornberger (1978) exponent, and ψ_{sat} (mm) the saturated soil matric potential. The exponent B s determined as a weighted average of the organic value (fixed at 2.7) and a mineral value dependent on the clay content ($B_{\text{min},i} = 2.91 + 0.159 \cdot f_{\text{clay},i}$). The saturated matric potential ψ_{sat} is similarly calculated as weighted average, with the organic component fixed at -10.3 mm and the mineral component defined as $\psi_{\text{sat,min},i} = -10.0 \cdot 10^{(1.88 - 0.0131 \cdot f_{\text{sand},i})}$. For a complete description of the parametrization, the reader is referred to the CTSM technical documentation (Lawrence et al., 2018). Since the fraction organic matter and the percent sand are prescribed and constant in time throughout a simulation, the volumetric water content at saturation, the field capacity and the wilting point are invariant in time.

In CTSM, plant transpiration processes, including root water uptake and stomatal conductance, are simulated using the plant hydraulic stress module of Kennedy et al. (2019). This parametrisation models water transport through vegetation using prognostic vegetation water potential at the root, stem, and leaf levels. Leaf water potential serves as the basis for stomatal conductance water stress, replacing soil potential, while hydraulic root water uptake is simulated using root water potential, replacing the previous transpiration partitioning function (Kennedy et al., 2019).

2.2 Experimental design

We conduct land-only simulations for the present day climate using a regular horizontal grid at 0.5° by 0.5° resolution. All simulations are conducted with CTSM version 5.2, using prescribed vegetation phenology based on MODIS satellite observations (IHistCLM51SP compset; ?). As the biogeochemistry module is not activated, there is no soil carbon dynamics modeled and soil organic matter is constant throughout the simulation. Atmospheric forcing is prescribed by the Global Soil Wetness Project (GSWP3; <http://hydro.iis.u-tokyo.ac.jp/GSWP3/>; see also ?), a reanalysis product with 0.5° global resolution. GSWP3 provides three-hourly bias-adjusted meteorological variables based on the dynamically downscaled 20th century reanalysis



(version 2) of the NCEP model (Compo et al., 2011).

The simulation workflow begins with a 60-year spin-up control simulation to ensure steady-state conditions in the soil water compartments. All subsequent simulations branch off from this control and span the period 1985–2014. The first ten years are
150 treated as an additional spin-up, leaving 20 years (1995–2014) for analysis. Land use is kept constant at the year 2000 in all simulations.

The main land cover types targeted for carbon sequestration are cropland and grassland. Consequently, the focus is on soil column variables specific to the crop fraction of each grid cell (Fig. A2). As crop columns are subdivided in an irrigated and
155 non-irrigated fraction, the soil column is not shared by other PFTs. The set-up assumes a generic C3 crop.

2.3 Soil carbon sequestration scenarios

The effect of carbon sequestration on soil hydrology is assessed by comparing a control simulation with present-day, fixed SOC content to three distinct scenarios representing the soil carbon content after 20 years of active sequestration (Fig. 1d, table
160 1). A 20-year period reflects a commonly cited saturation point after which a new equilibrium in SOC for the upper soil layers is reached (Zomer et al., 2017). In each scenario, carbon increases are applied to the top 30cm of the soil column. This depth is consistent with prior studies (Zomer et al., 2017; Padarian et al., 2022) and the 4p1000 initiative, as this layer holds the majority of soil carbon and is influenced by management practices (Batjes, 1996). The input data implicitly assume historical carbon loss through cultivation since the onset of agricultural practices, allowing for sequestration to reach a new equilibrium.

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The first two scenarios assume uniform carbon sequestration applied globally to all cropland grid cells (Table 1). These amounts are based on estimates by Zomer et al. (2017), who estimated a realistic SOC sequestration potential based on the SOC of the SoilGrids 250 m dataset and the scenarios of Sommer and Bossio (2014). The High scenario assumes a total absolute increase of 0.55%, or 5.5 gC kg⁻¹ soil in SOC over 20 years relative to present-day levels, representing a high sequestration
170 potential for cropland. The Medium scenario applies a total absolute increase of 0.27% or 2.7 gC kg⁻¹ soil after 20 years, representing medium potential for cropland and high potential for grassland. For both scenarios, the increase is applied to all grid cells worldwide. The third scenario, 4p1000, is based on the 4 per 1000 initiative and assumes an annual increase of 0.4%. Over a period of 20 years, this results in a total increase of 8% relative to current SOC (SOC) levels, under the assumption that the increase does not compound annually. So while the first two scenarios represent a spatially invariant, rather small but
175 attainable increases everywhere, the 4p1000 scenario results in spatially variable increases consistent to the pattern of present-day SOC, with high SOC changes in high-latitude regions and lower values in mid to low latitudes (Fig. A3). In all scenarios, sequestration is applied to the soil organic carbon (SOC, gC kg⁻¹), based on the organic matter (OM) input layer (equation 1), for the top five soil layers (0–32 cm), approximating the 30 cm sequestration depth assumption.



Table 1. Soil carbon sequestration scenarios. The values of the High and Medium scenarios correspond to changes in SOC after 20 years relative to present-day SOC values, while the values of the 4p1000 scenario correspond to that relative to present-day SOC.

Scenario	Δ SOC (%) after 20 years	Reference
High	+ 5.5 gC kg ⁻¹ soil	Zomer et al. (2017); Sommer and Bossio (2014)
Medium	+ 2.7 gC kg ⁻¹ soil	Zomer et al. (2017); Sommer and Bossio (2014)
4p1000	+ 8 % of present-day SOC (gC kg soil ⁻¹)	Minasny et al. (2017); Rumpel et al. (2020)

180 The input SOC map, based on the WISE30sec and aggregated to the horizontal resolution of 0.5° by 0.5° used in the simulation, is validated by comparing the mean SOC in the top 30 cm of cropland soil with values from literature. The mean SOC in cropland topsoils amounts to 68 t ha⁻¹ and is derived using the SOC (gC kg⁻¹) and soil bulk density (g cm⁻³) of the CTSM input derived from WISE30sec and a soil depth of 30 cm. This value aligns with recent studies: Padarian et al. (2022) reports 1.3 gC kg soil⁻¹, corresponding to 59.1 t ha⁻¹ derived using a neural network based on global climate, land cover and soil datasets.

185 Zomer et al. (2017) cites 82 t ha⁻¹, based on the Soils Grid 250 m global database. Additionally, Zomer et al. (2017) estimates total cropland SOC at 131.81 PgC, while our CTSM input map yields 108.42 PgC using the same cropland mask.

Four experiments are conducted, which differ in the soil organic matter input provided. The control scenario (CTL) uses the default WISE30sec SOC map, while the High, Medium, and 4p1000 scenarios use modified soil organic matter input maps.

190 2.4 Analysis

The effects of carbon sequestration on soil hydrology are quantified by comparing climatological differences between the scenarios and CTL simulations across various soil moisture-related variables including water holding capacity, saturated fraction, volumetric and total water content. The water holding capacity is defined as the plant available water, calculated as the difference between field capacity (θ_{fc} , mm³ mm⁻³) and wilting point (θ_{wp} , mm³ mm⁻³).

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The saturated fraction (θ_s , mm³ mm⁻³) represents the extent to which the soil is saturated and is calculated as:

$$\theta_s = \frac{\theta}{\theta_{sat}} \quad (5)$$

where θ mm³ mm⁻³ is the actual volumetric water content, and θ_{sat} (mm³ mm⁻³) is the volumetric water content at saturation.

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Annual water stress (m) is determined by calculating the daily difference between the wilting point (θ_{wp}) and the water content (θ) for days where $\theta < \theta_{wp}$. This difference is accumulated over the year (Equation 6) and summed across the first seven soil layers ($d = 0.68$ m), representing the upper 60 cm of the soil corresponding to the soil depth to which irrigation



would be applied. This definition of water stress is stringent, as stomatal closure typically occurs prior to reaching the wilting
205 point.

$$\text{water stress} = \sum_{i=1}^7 \sum_{\text{day}=1}^{365} (\theta_{wp,\text{day},i} - \theta_{\text{day},i}) \cdot d \quad \theta_{\text{day},i} < \theta_{wp,\text{day},i} \quad (6)$$

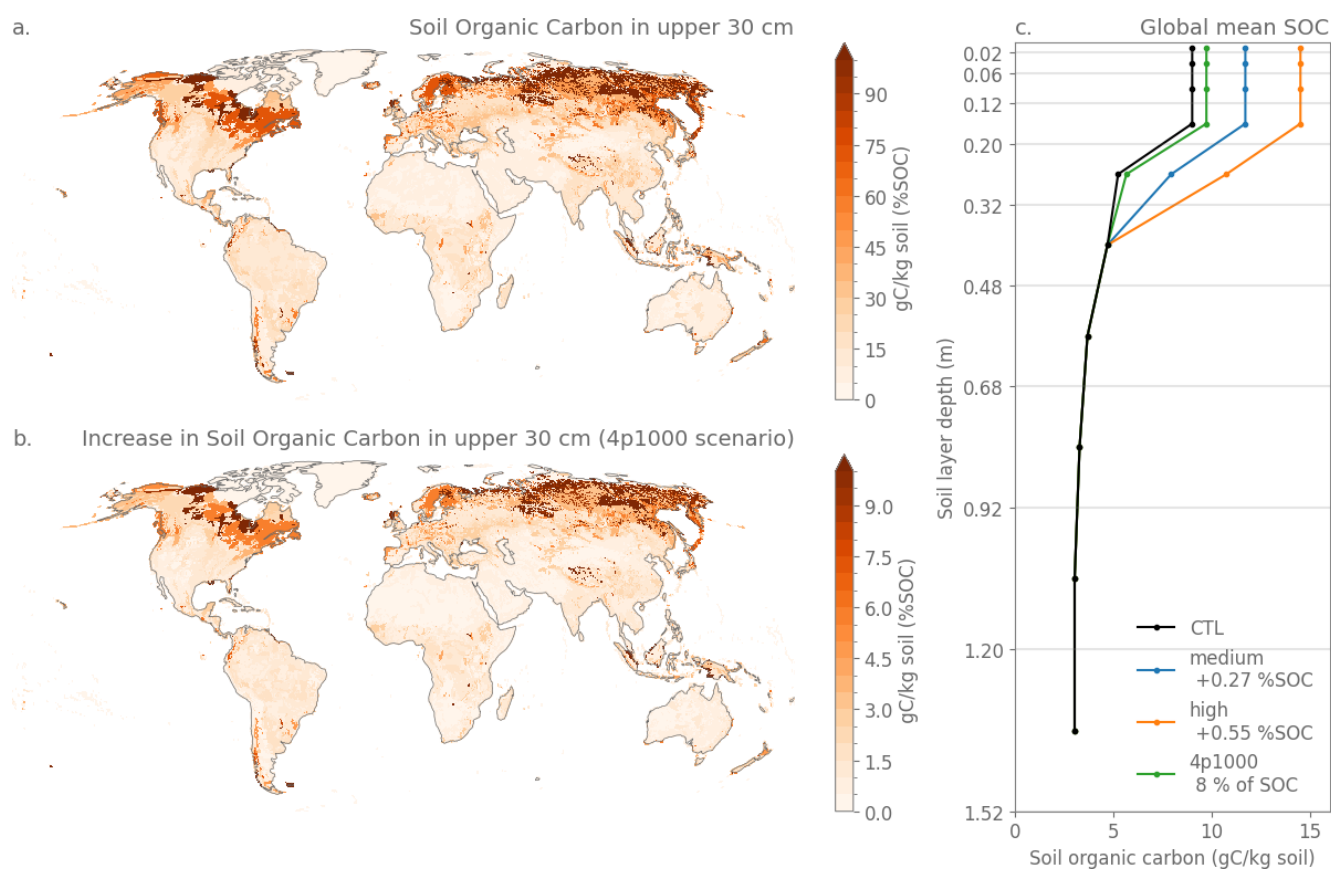


Figure 1. Soil Organic Carbon (SOC), sand and clay percentage and different soil carbon sequestration scenarios used as model input
Spatial distribution of SOC in the top 30 cm of soil based on the WISE30sec dataset (a), weighted sand percentage (b) and clay percentage, both based on the WISE30sec dataset. Vertical profiles of global mean SOC in the control simulation (CTL) and the three sequestration scenarios: medium, high, and 4p1000 (c).



3 Results

3.1 Impact of soil carbon sequestration on water holding capacity, saturated fraction and soil water content

The impact of carbon sequestration is measured through changes in water holding capacity, saturated fraction, and volumetric water content of the High scenario compared to CTL (Fig. 2). Water holding capacity in each simulation remains constant over time, as defined by the model formulation. Spatially, however, it increases globally, with a global mean volumetric rise of 0.002 m³ m⁻³, or 2 mm water per 100 mm soil (Fig. 2a). This pattern is driven by a consistent increase in soil moisture at field capacity, strongly influenced by the soil organic carbon (SOC) fraction (Appendix Fig. A4b) and reflects the model's representation of the improved soil porosity from added organic carbon. While wilting point soil moisture also increases, this effect is less widespread and less pronounced (Appendix Fig. A4a). Changes in the saturated fraction exhibit a more heterogeneous spatial pattern and are generally of much lower magnitude, with areas showing substantial decreases (Fig. 2b). These decreases occur when porosity, or soil moisture content at saturation (θ_{sat}), increases more than the actual volumetric soil moisture (θ). Volumetric water content generally increases when weighted over the soil layers, with a global mean increase of 0.002 m³ m⁻³ (Fig. 2c). Overall, the changes are small and consistent in the High scenario, which involves a small forcing of a 5.5 gC kg⁻¹ soil (0.55%) increase in SOC.

With the soil's increased capacity to retain water and rising volumetric water content in most regions, a key question emerges: Is this additional water accessible to plants, which would reduce water stress and thereby potentially alleviate pressures on irrigation water abstraction and unsustainable use? Given the importance of this issue, and the potential of soil carbon sequestration for croplands, the remainder of our analysis focuses on the crop fraction of the grid cell. Cropland resides on its own soil column, which elucidates the competing effects of different PFTs with different rooting depths.

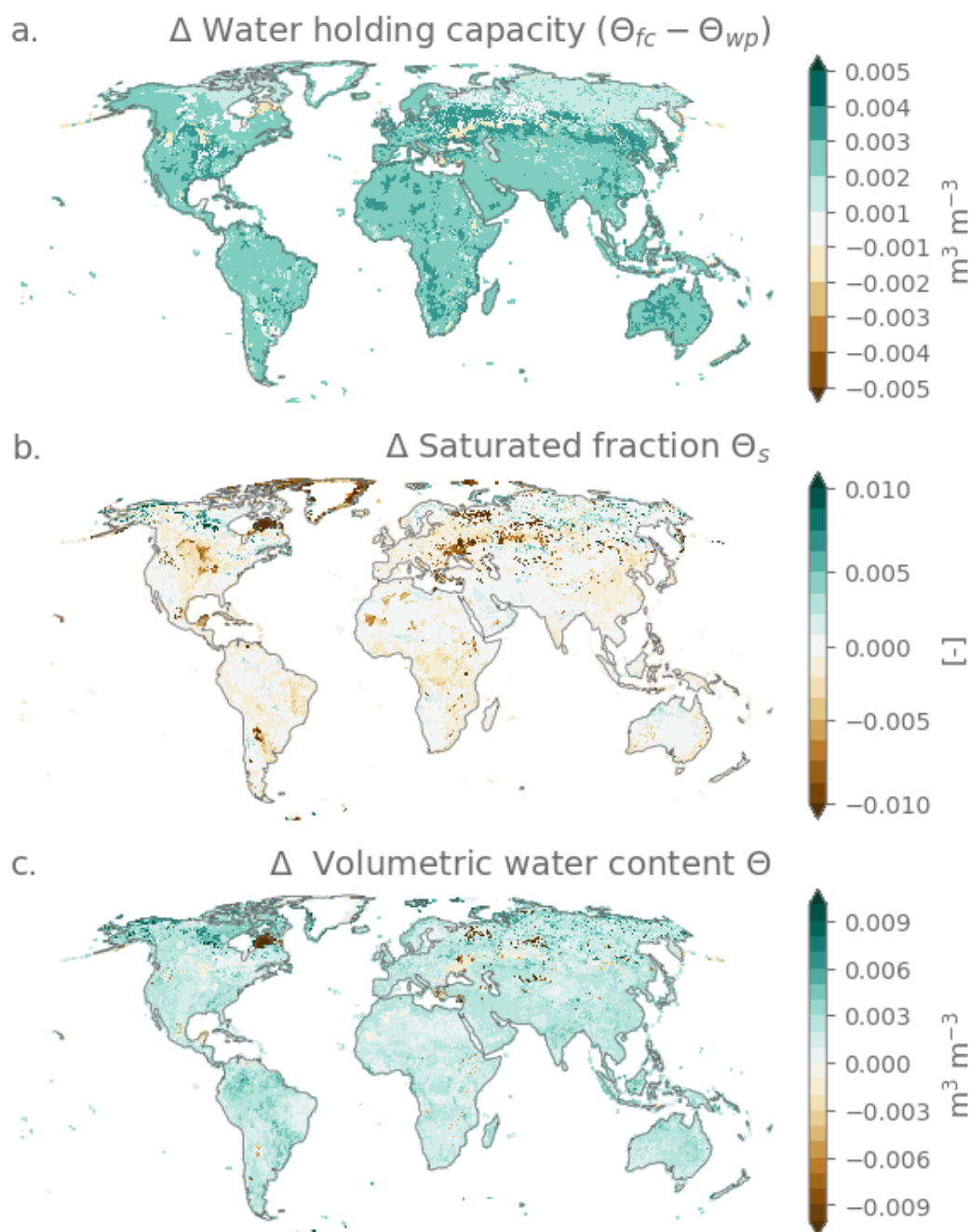


Figure 2. Effect of soil carbon sequestration on the water holding capacity, saturated fraction and volumetric water content. Difference in the High and CTL scenarios for all land grid cells in (a) the soil columns water holding capacity (field capacity - wilting point); (b) saturated fraction; and (c) volumetric water content; all weighted averages over the first 10 soil layers of CTSM.

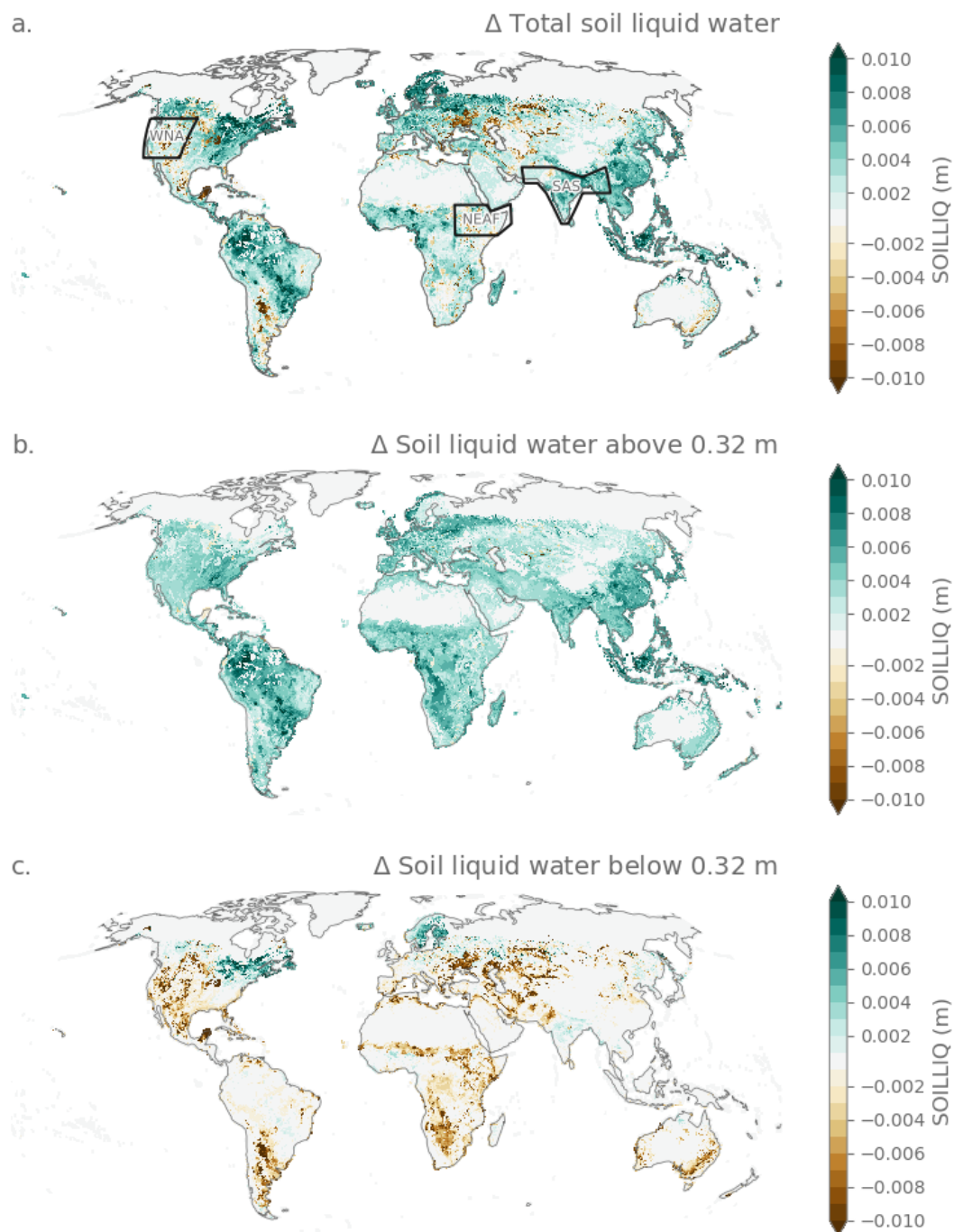


Figure 3. Effect of soil carbon sequestration scenario on soil liquid water content. Difference in High and CTL scenarios for the column hosting crop fraction for total soil water content over the whole soil column. Black contours refer to the regions showed in Fig. A7 (a.), liquid water content above 32 cm, corresponding to the first five soil layers in CTSM (b) and liquid water content below 32 cm (c).



The effect on mean total water availability and total soil liquid water content shows an increase of 2 mm averaged globally, and varies by region (Fig. 3a). In areas such as the Amazon, the eastern United States and Canada, Scandinavia, Western Europe, West and Central Africa, Madagascar, and East and Southeast Asia, carbon sequestration leads to an increase in total soil liquid water content. Conversely, a marked decrease in soil liquid water content is observed in regions such as the western United States, the Argentinian Pampas, parts of Southern Africa (spanning Namibia, Angola, Botswana, and South Africa), the Sahel, areas of East Africa, portions of the Eurasian continent, northeastern China, and Australia. These contrasting responses result from differences in soil liquid water content above and below a depth of 32 cm.

In the upper 32 cm of the soil, where carbon sequestration is applied, mean liquid water content consistently increases across all crop regions, with a global average increase of +4 mm (Fig. 3b). In contrast, the deeper soil layers below 32 cm exhibit diverging patterns (Fig. 3c, global average decrease of -1 mm). In most cropland areas, liquid water content in these deeper layers decreases, particularly where total water content is reduced throughout the entire soil column. However, some regions—such as eastern Canada, Scandinavia, and parts of India and Southeast Asia—show increases in deeper layer water content. These increases may be linked to ice melting within the model or regional soil characteristics. The soil depth in these layers is determined by the soil depth map (Appendix Fig. A1), which defines maximum depths for each location. Areas with strong declines in water content tend to correspond with more arid or sandy regions (Fig. 1b).

To better assess regional differences, we examine volumetric water content (θ) and the saturated fraction (θ_s) in three IPCC AR6 regions: Western North America (WNA), Northeastern Africa (NEAF), and South Asia (SAS) (Fig. 3a; Iturbide et al., 2020). WNA, a dry region with extensive irrigation and a rainy season from November to March, has 50% sand and 20% clay (Fig. 1b,c). NEAF, covering the arid Horn of Africa and wetter Ethiopian Highlands, has high sand (55–60%) and low clay (<20%) fractions, with a dry season from December to March. Both regions show no clear increase in total water content due to declines below 32 cm (Fig. 3a, c). SAS, a major irrigated region with a monsoon season from April to September, has 30% sand and 60% clay, leading to a slight increase in soil liquid water below 32 cm (Fig. 3c).

Seasonal variations in the influence of soil carbon sequestration on soil liquid water for these regions are examined through vertical profiles of changes in volumetric water content and the saturated fraction (Fig. 4). Volumetric water content is highest between 20 and 32 cm—the deepest layer where carbon sequestration is applied (Fig. 4a-c). Across all three regions, water content increases most during the rainy season and persists at greater depths beyond the wet periods. In contrast, a slight sequestration-induced drying occurs in the surface layers during the dry season, with the deepest drying observed in Western North America. Below the carbon-sequestered layers, changes in water content are minimal, with slight drying or no change observed. This pattern suggests that increased SOC in the upper layers enhances water retention in the upper layers, thereby reducing percolation and limiting the downward movement of moisture through soil texture effects.



The reduction in saturated fraction is most pronounced in the upper layers and diminishes with depth, indicating lower surface soil saturation (Fig. 4). Relative increases in volumetric water content reach up to 10% at the deepest sequestration layer, with minimal seasonal variation compared to absolute values (Appendix Fig. A7). This suggests that carbon sequestration amplifies soil water seasonality and promotes water percolation into deeper layers. Especially in the surface layers, the increase in soil water storage capacity (saturation) is not fully matched by the actual increase in water content from percolation, resulting in less saturated surface soil layers.

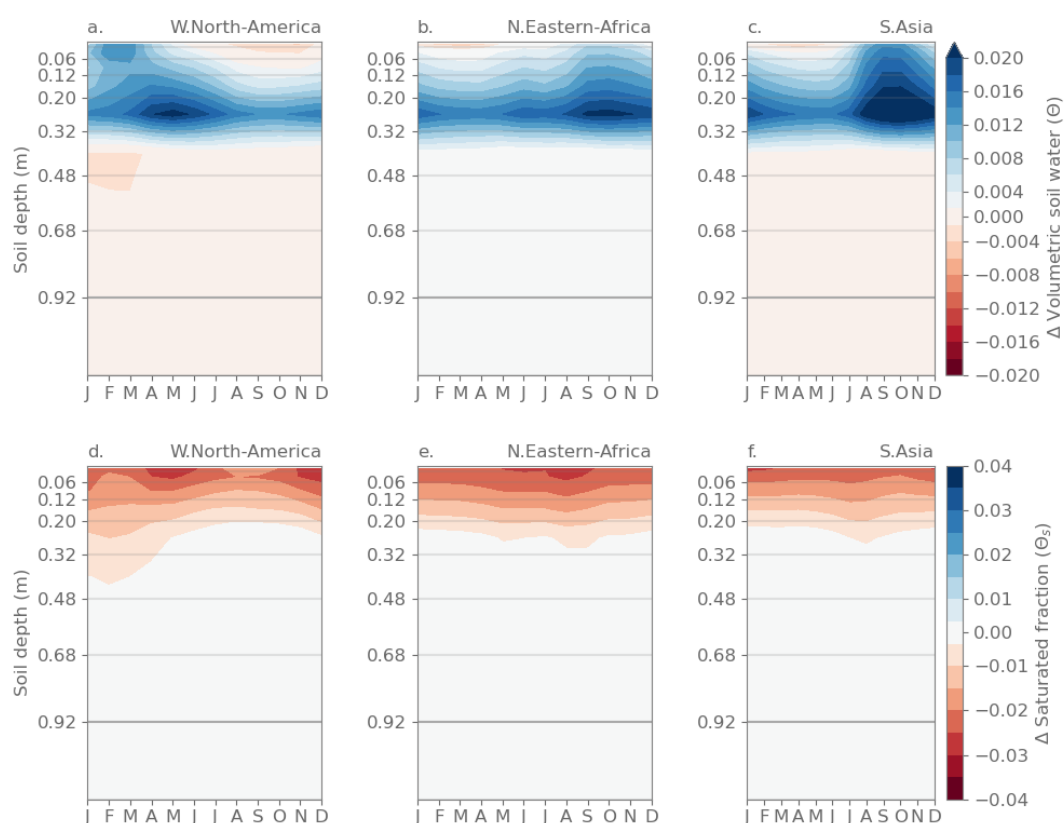


Figure 4. Seasonal effect of soil carbon sequestration on volumetric soil water and saturated fraction for different regions and soil depths. (a,b,c) Difference in High and CTL scenarios for volumetric soil water θ ($mm^3 mm^{-3}$) and (d,e,f) saturated fraction θ_s ($mm^3 mm^{-3}$) across seasons and soil depth for three regions: (a,d) Western North-America, (d,e) North-Eastern Africa and (c) South Asia, averaged over their cropland grid cells. The y-axis indicates the different soil layers.

3.2 Effect on evapotranspiration

To assess whether the increased water content in upper soil layers improves plant water availability, we examine its impact on evapotranspiration components: soil evaporation, vegetation transpiration, vegetation evaporation, and the total evapotran-



270 spiration (Fig. 5). Carbon sequestration has a clear spatial effect on soil evaporation, but its impact is smaller compared to
vegetation transpiration (Fig. 5a). Soil evaporation increases in more sandy regions like the Monte Desert in Argentina, West-
ern Southern Africa, Western North America, as well as the Sahel and parts of the Eurasian Steppe. In contrast, areas with
higher clay fractions, such as the Ethiopian Highlands, parts of India and Northwestern Australia, show reductions in soil
evaporation. Sequestration seems to reduce soil evaporation in clay-rich regions while enhancing it in sandy ones. The increase
275 in soil evaporation appears to reduce the water available for transpiration. This is somewhat unexpected, given that evaporation
primarily draws from near-surface moisture, whereas transpiration accesses water from deeper soil layers.

Carbon sequestration generally boosts vegetation transpiration, suggesting that plants have more access to water (Fig. 5b).
This effect is particularly notable in clay-rich regions, such as in the tropical rain forests, India, and Southeast Asia. In con-
280 trast, the impact on vegetation evaporation is minimal (Fig. 5c). As our simulations do not include a coupled atmosphere,
no direct atmospheric feedback is expected. However, because 2-meter temperature is a diagnostic variable in CTSM, a small
temperature-related feedback is observed in vegetation evaporation rates. The overall change in evapotranspiration is dominated
by the increased vegetation transpiration, and to a lesser extend soil evaporation (Fig. 5d). The overall change in evapotran-
spiration is primarily driven by increased vegetation transpiration, with a smaller contribution from enhanced soil evaporation
285 and amounts up to a global average increase of +2 mm (Fig. 5d).

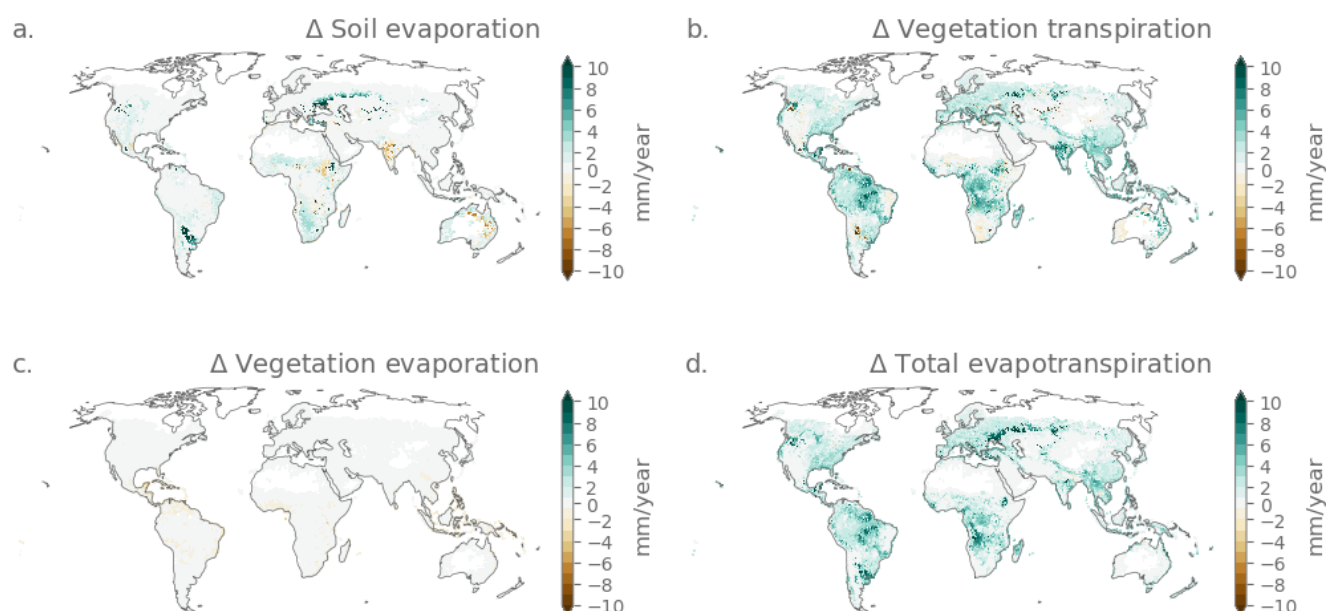


Figure 5. Effect of soil carbon sequestration on evapotranspiration. Difference in High and Control scenarios for the column hosting crop fraction for (a.) soil evaporation, (b.) vegetation transpiration, (c.) vegetation evaporation and (d.) total evapotranspiration



3.3 Effect on water stress

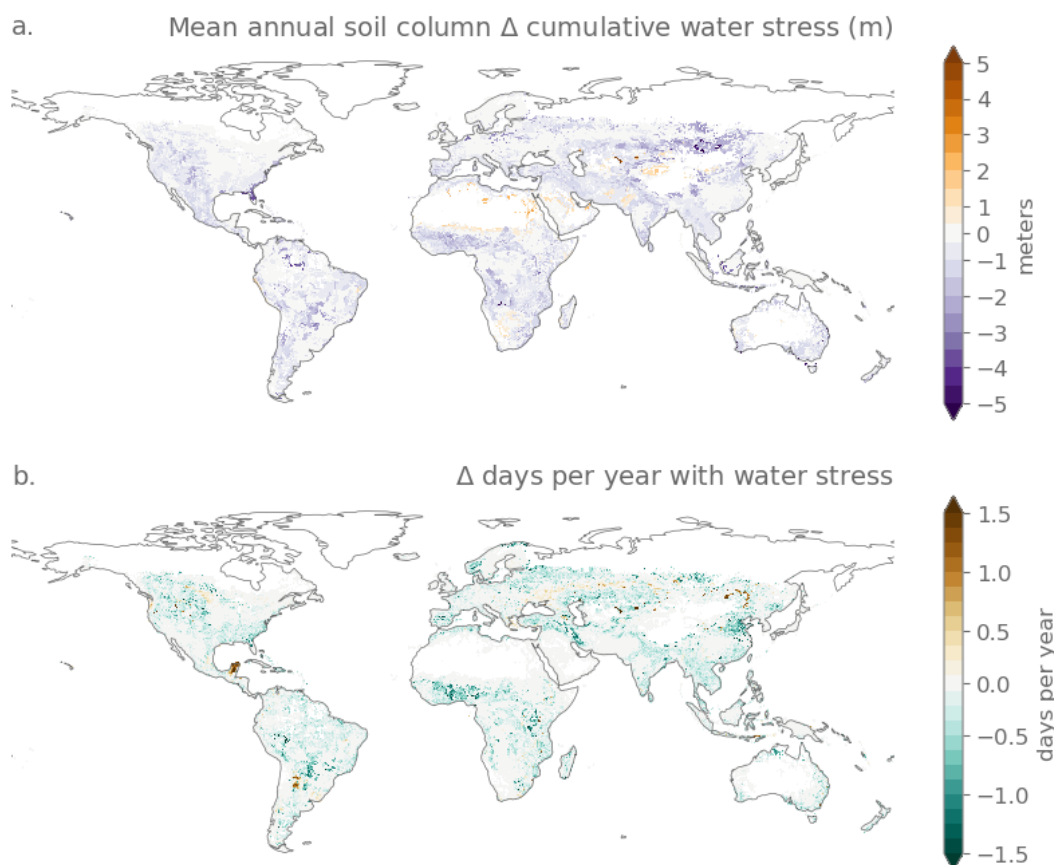


Figure 6. Effect of soil carbon sequestration scenario on water stress. (a.) Difference of annual cumulative water stress in the soil layers above 60 cm and (b.) the mean number of days per year with water stress in the High and Control scenarios for the column hosting crop fraction.

Annual water stress decreases across most regions, particularly in Western North America, parts of South America, the Sahel, Western Southern Africa, South Asia and the Middle East, which correspond to areas with high sand contents. The decrease in water stress reaches several meters per year in some areas (Fig. 6a). These areas, which are primarily arid (desert and steppe), show a general decrease in the number of days with water stress. Still, the reduction remains modest, with a maximum decrease of 1.5 days (Fig. 6b). However, the water stress definition used here captures only extreme conditions, which may lead to an underestimation and therefore conservative estimates of the actual frequency and severity of water stress experienced by vegetation. While soil carbon sequestration appears to reduce water stress for croplands and lower the number of stress days per year, the effect is limited by the small forcing of the High scenario. Nonetheless, it holds potential for reducing irrigation



295 demands. It is however not possible to quantify the direct reduction in irrigation needs due to the way irrigation is parametrized in the CTSM, using a threshold soil moisture value that also changes with increased soil carbon.

3.4 Implications for runoff

Soil carbon sequestration reduces surface runoff in most regions (Fig. 7a, -1 mm globally averaged), with pronounced decreases observed in Scandinavia, Central Europe, northeastern Canada and the western Amazon. This reduction suggests improved infiltration rates (appendix Fig. A16). Subsurface drainage, which refers to water exiting the bottom of the grid cell, also decreases in most areas, except in regions like northeastern Canada, Scandinavia and Central Europe (Fig. 7b, -0.6 mm globally averaged). The simulated reduction in drainage is attributed to a lower saturated fraction across the soil column (section 3.1, Fig. 2), leading to reduced water tables and subsurface runoff. This result aligns with the model's representation of subsurface drainage ?. Overall, these findings suggest that increased soil carbon in CTSM generally enhances the soil's ability to retain water.

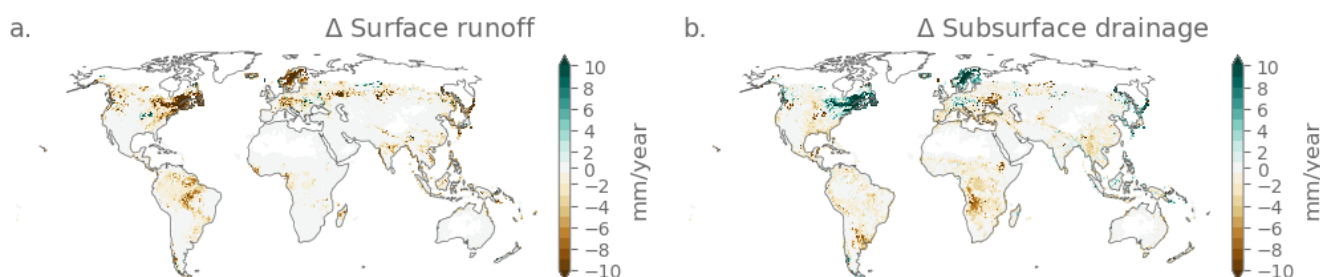


Figure 7. Effect of soil carbon sequestration on runoff. (a.) Difference in the High and CTL scenarios for the column hosting crop fraction for surface runoff and (b.) subsurface drainage.

3.5 Differences between scenarios

The soil water balance responses described above correspond to the High sequestration scenario. The Medium sequestration scenario exhibits similar spatial patterns but with smaller magnitude differences (appendix Fig. A6, A9, A11, A13 and A15). The 4p1000 scenario, where soil carbon sequestration is relative to existing carbon stocks (appendix Fig. 1a), shows distinct spatial patterns, which are reflected in the hydrological responses (appendix Fig. A5, A8, A10, A12 and A14).

Comparing volumetric water content across the three selected regions (as shown in Fig. 3a) shows compensatory effects, with increased soil moisture above 32 cm and decreases below (Fig. 8). This pattern is most pronounced in Western North America and Northeastern Africa, where both the High and Medium sequestration scenarios result in a net soil water increase. However, in the 4p1000 scenario, the decline below 32 cm outweighs the gains above, leading to an overall reduction in volumetric water



content. In South Asia, decreases below 32 cm are minimal, and in the 4p1000 scenario, a slight increase is observed, resulting in a strong overall increase in soil water in the full column (Fig. 8).

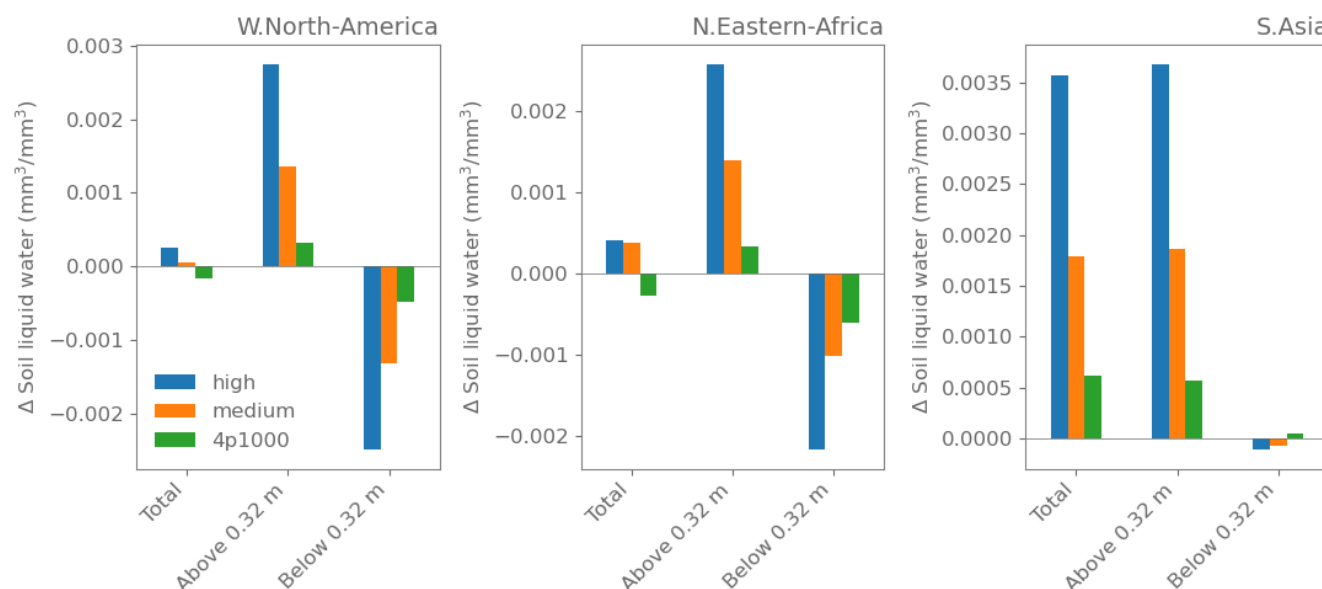


Figure 8. Effect of soil carbon sequestration following different scenarios on volumetric soil water content. Difference in soil moisture under the High, Medium, and 4p1000 scenarios, spatially averaged over Western North America, Northeastern Africa, and South Asia.

4 Discussion

Our simulations indicate that carbon sequestration enhances soil moisture in the upper layers where it is applied, effectively retaining water near the surface, limiting percolation to deeper layers and also leading to a reduction in surface runoff (Fig. 9). The increased soil moisture supports higher evapotranspiration rates, primarily driven by more vegetation transpiration and soil evaporation. Finally, as water storage capacity increases more than the actual water retained, saturation levels decline, resulting in reduced subsurface drainage from the grid cell. Soil organic carbon thus redistributes the partitioning of the available water from precipitation from less surface and subsurface runoff to more evapotranspiration.

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Similar to the findings of Minasny and McBratney (2018), we observe a positive effect of increased soil organic carbon (SOC) on available water capacity, defined as the difference between field capacity and the wilting point. While Minasny and McBratney (2018) report that a 1% mass increase in SOC leads to a 1.16% volumetric increase in available water capacity, our results suggest a slightly stronger response: in the high SOC scenario (corresponding to a 0.55% mass increase), we find a global average increase in water holding capacity of approximately 2% (Fig. 2, Section 3.1). The accompanying increase in volumetric water content of 2% globally averaged, further supports the conclusion that carbon sequestration can enhance water

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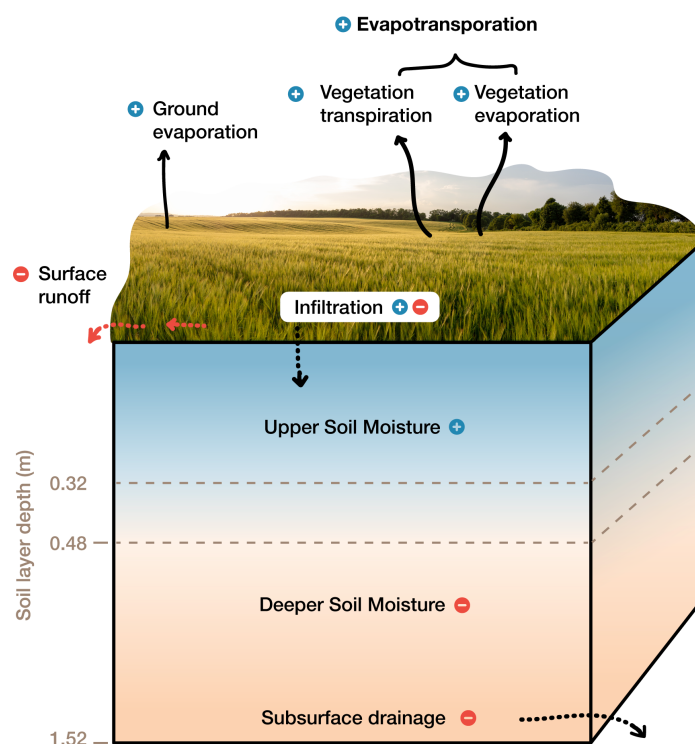


Figure 9. Influence of soil organic carbon on water balance components. Conceptual illustration showing changes in the soil column and hydrological processes due to increased soil organic carbon.

availability, albeit to a limited extent.

While our simulations provide insights at large scales, global land models have inherent limitations in capturing local soil hydrological processes and their feedbacks with the atmosphere. In CTSM, as for other global land models, soil hydraulic functions like water retention and hydraulic conductivity are parametrized based on pedotransfer functions, which heavily rely on the input soil texture maps and organic matter input. These pedotransfer functions in the model might not be directly suited designed for sensitivity experiments with changing SOC, as they are based on empirical relationships. Furthermore, in these parameterizations, the role of soil structure through biological activity, such as the formation of biopores and soil aggregates, is generally not considered (Fatichi et al., 2020). As soil processes associated with carbon sequestration, such as no-till practices, compaction, increased soil microfauna, and other management techniques like cover cropping, are not fully captured, the model may not adequately represent their effects on soil water retention (Minasny and McBratney, 2018). Additionally,



the coarse horizontal resolution of 0.5° by 0.5° and the generalized parametrization conceal local heterogeneity, potentially overlooking important regional or site-specific processes.

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Here, we focused on croplands to apply carbon sequestration. However, next to croplands, other agricultural land such as meadows and pastures where herbaceous forage crops are grown, provide potential to store carbon (Sommer and Bossio, 2014; Tessema et al., 2020). Bai and Cotrufo (2022) provides estimates of the global potential for SOC sequestration through grass-land restoration, with mean theoretical, realistic, and achievable capacities of 10.2, 6.8, and 3.4 billion t CO₂ equivalents per year, respectively. Converted to SOC after 20 years, these values correspond to 55.6 PgC, 37.0 PgC and 18.6 PgC, respectively. Compared to estimated range of 29 to 65 PgC realistic storage potential in croplands, these values are slightly lower (Zomer et al., 2017; Padarian et al., 2022). Nevertheless, grassland areas could also provide eco-hydrological co-benefits, with magnitudes similar to the Medium scenario of croplands.

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This study focused on mean changes as a first step in assessing the hydrological co-benefits of soil carbon sequestration. However, its potential becomes clearer when considering its role in mitigating drought risk and reducing irrigation water demand. This is particularly relevant in the context of climate change, as drought frequency is projected to increase (?). To better quantify these benefits, future studies should account for precipitation anomalies and evolving drought conditions rather than relying solely on mean changes. Additionally, our approach, which prescribes vegetation phenology, does not capture soil carbon dynamics, vegetation responses or land-atmosphere feedbacks. To fully assess the impact of soil carbon sequestration under different climate change scenarios, future research should employ dynamic vegetation models coupled to the atmosphere.

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5 Conclusions

This study presents a sensitivity experiment using CTSM to assess the eco-hydrological co-benefits of soil carbon sequestration under policy-relevant scenarios. By prescribing atmospheric conditions and vegetation phenology, we isolate the direct effects of SOC on soil water dynamics.

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Our simulation experiments with CTSM show that soil carbon sequestration enhances the soil's water-holding capacity, improving its ability to retain moisture. Across most regions, total soil water content increases, primarily due to higher moisture levels in the upper 30 cm, where sequestration is applied, while deeper layers often experience slight reductions—particularly in sandy and arid regions. However, the increase in soil liquid water is insufficient to fully offset the rise in water-holding capacity, leading to less saturated upper layers. Despite this, simulations indicate increased vegetation transpiration, suggesting greater water availability for plant uptake, especially in clay-rich soils, which indicates that the partitioning of the available precipitation shifts to less surface and subsurface runoff and more evapotranspiration. This effect is particularly relevant for annual water stress, which is consistently reduced. However, the overall impact remains limited, reflecting the relatively small but realistic forcing applied in the three scenarios.

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Our findings demonstrate the value of global land surface models for identifying regional differences in eco-hydrological co-benefits of SOC sequestration, supporting locally relevant climate mitigation and adaptation strategies. Future work could build on this work by incorporating multiple land surface models with varying soil hydrological parameterizations and conducting coupled simulations to explore potential atmospheric feedbacks, particularly through changes in evapotranspiration. As interest
380 in soil carbon sequestration restoration grows (Amelung et al., 2020), its potential to provide co-benefits by increasing plant available water availability should be carefully considered.

Code and data availability. The WISE30sec dataset described in Batjes (2016) is available at <https://data.isric.org/geonetwork/srv/api/records/dc7b283a-8f19-45e1-aaed-e9bd515119bc>. CTSM is available through the following repository: <https://github.com/ESCOMP/CTSM>. The scripts used in this study are available at: https://github.com/Ivanderkelen/Vanderkelen_etal_2025_BG with the DOI: <https://doi.org/10.5281/zenodo.15561256>. Finally, the created input data, ancillary data and model output are available on Zenodo at <https://doi.org/10.5281/zenodo.15552986>
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The soil layers are subdivided into hydrologically active and non-active layers. We tested the hydrologically active layers to see what it is going to

Author contributions. IV and ELD designed the study. IV performed the analysis and wrote the manuscript. SS and DL provided scientific
390 input on CTSM responses and source code. IV wrote the manuscript with major contributions from MED, BDS and MK and input from all other authors. All authors critically revised the draft and gave final approval for publication.

Competing interests. Some authors are members of the editorial board of journal BG.

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395 under project ID s1207. We thank Dr. Petra Sieber for helping with the porting of CTSM on Piz Daint and Daria Vuistiner for the design of Figure 9.



Table A1. Soil layers in CTSM (Lawrence et al., 2018)

Layer	Node (m)	Thickness (m)	Depth (m)
1	0.01	0.02	0.02
2	0.04	0.04	0.06
3	0.09	0.06	0.12
4	0.16	0.08	0.2
5	0.26	0.12	0.32
6	0.4	0.16	0.48
7	0.58	0.2	0.68
8	0.8	0.24	0.92
9	1.06	0.28	1.2
10	1.36	0.32	1.52
11	1.7	0.36	1.88
12	2.08	0.4	2.28
13	2.5	0.44	2.72
14	2.99	0.54	3.26
15	3.58	0.64	3.9
16	4.27	0.74	4.64
17	5.06	0.84	5.48
18	5.95	0.94	6.42
19	6.94	1.04	7.46
20	8.03	1.14	8.6
21	9.795	2.39	10.99
22	13.328	4.676	15.666
23	19.483	7.635	23.301
24	28.871	11.14	34.441
25	41.998	15.115	49.556

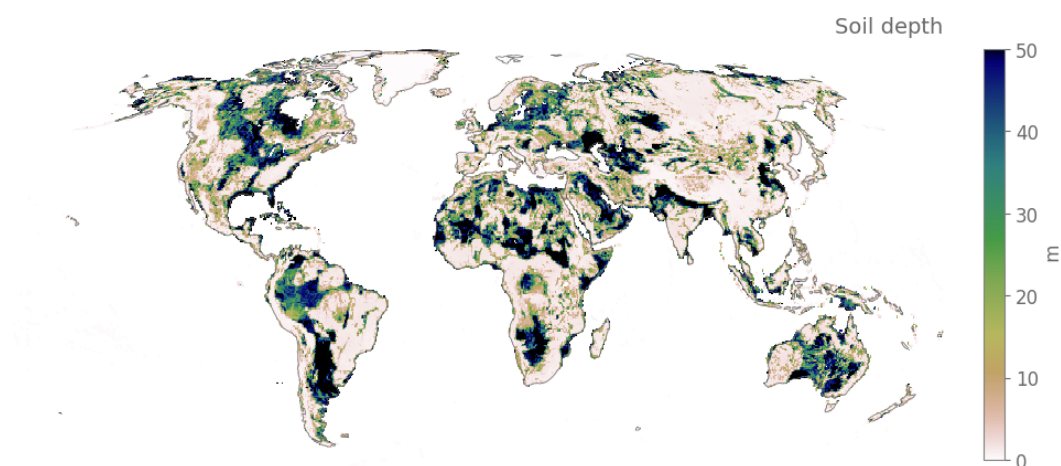


Figure A1. Soil depth map used as input in CTSM5.1

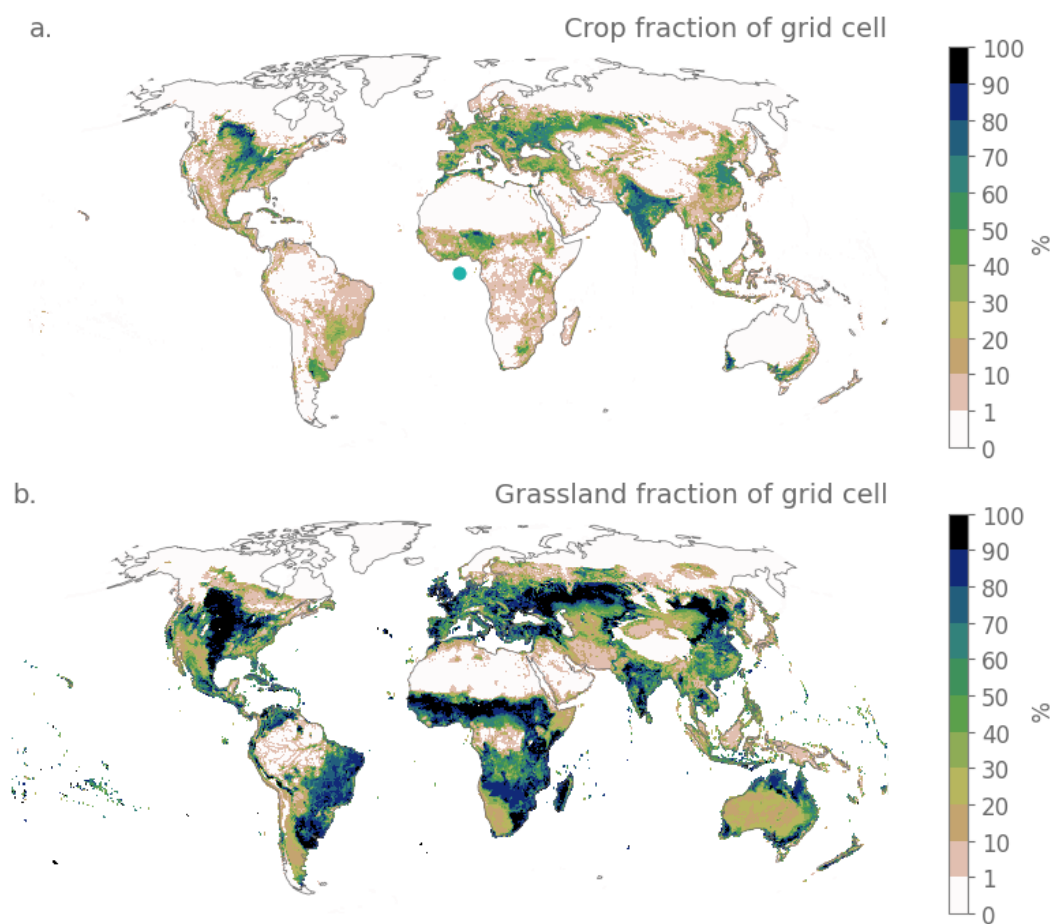


Figure A2. (a.) Crop fraction and (b.) grassland fraction of the grid cells as prescribed by the surface datasets in CTSM5.1.

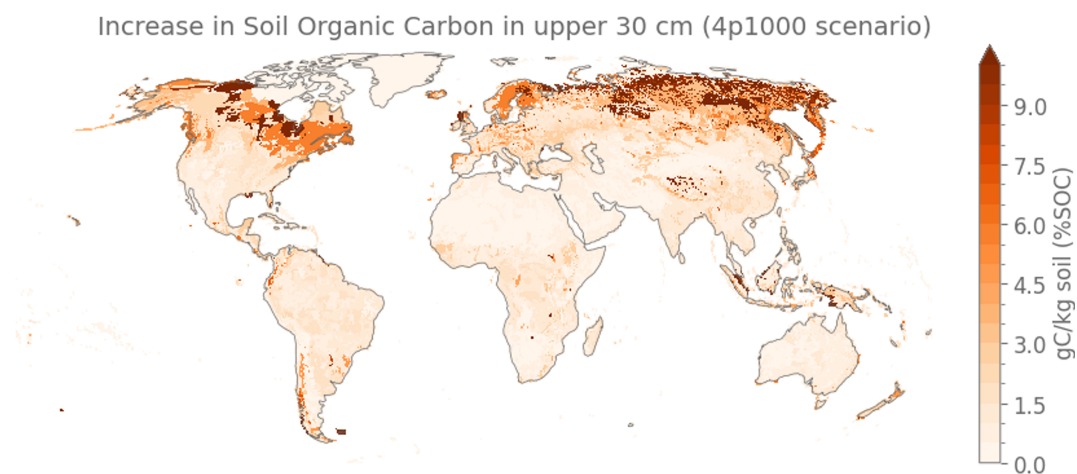


Figure A3. Increase in SOC following the 4p1000 scenario, assuming a 0.4% annual increase of the current carbon stocks over a period of 20 years

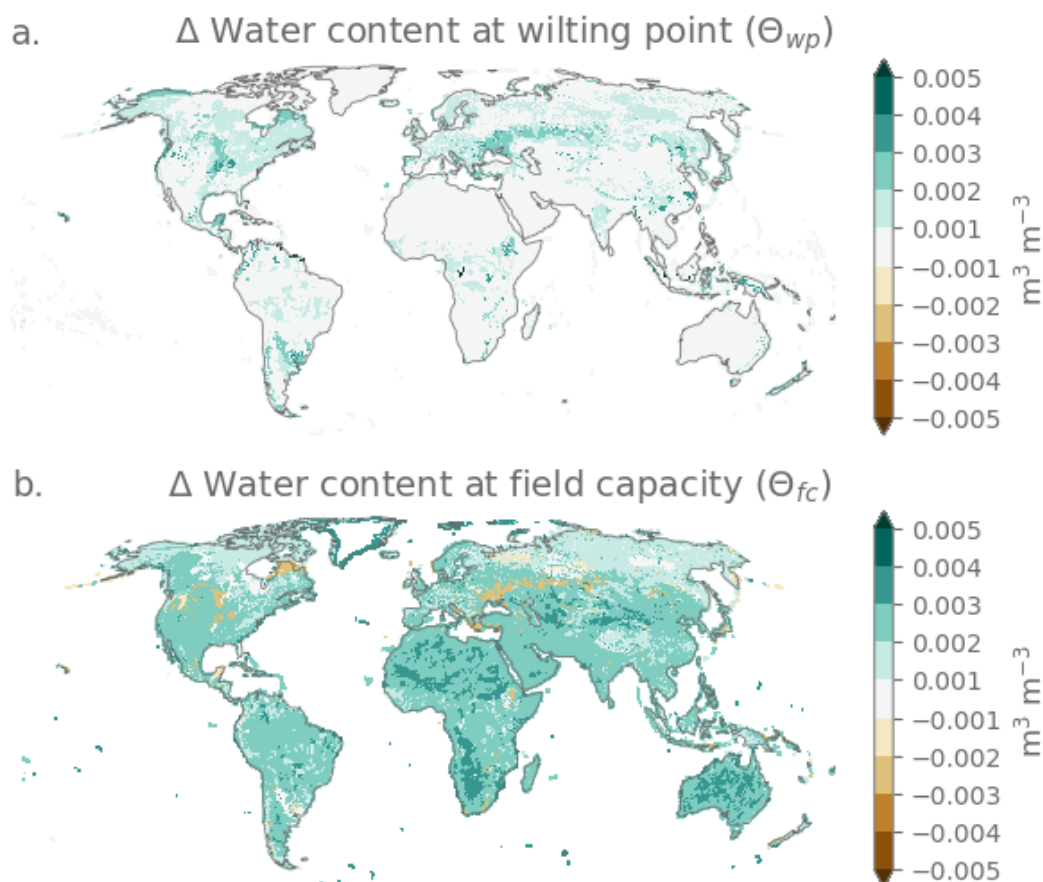


Figure A4. Effect of soil carbon sequestration on the water content at wilting point and field capacity. (a.) Difference in the High and CTL scenarios for all land grid cells in the soil columns volumetric water content at permanent wilting point and (b.) field capacity, all weighted averages over the first 10 soil layers of CTSM.

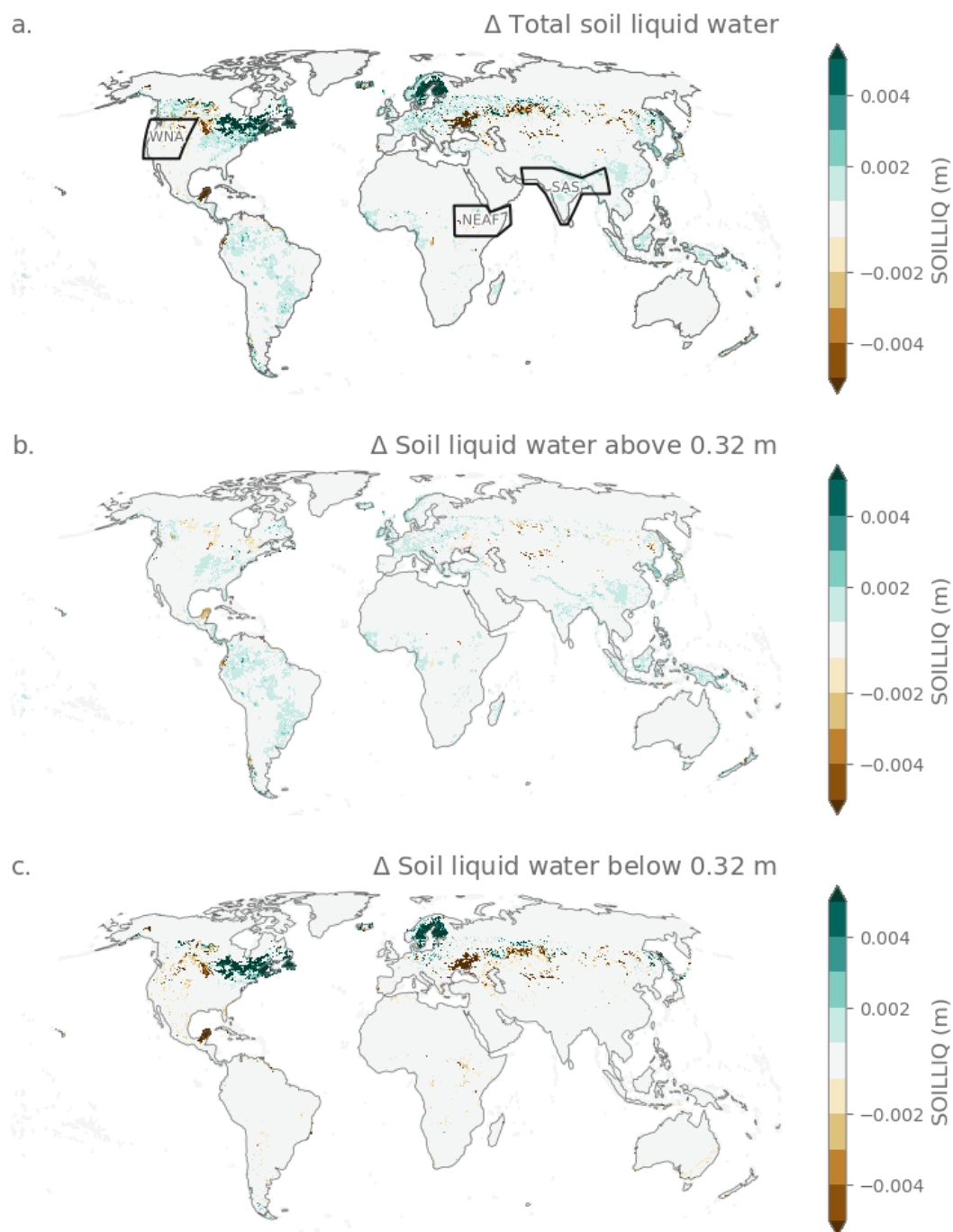


Figure A5. As Fig. 3, but for the 4p1000 scenario.

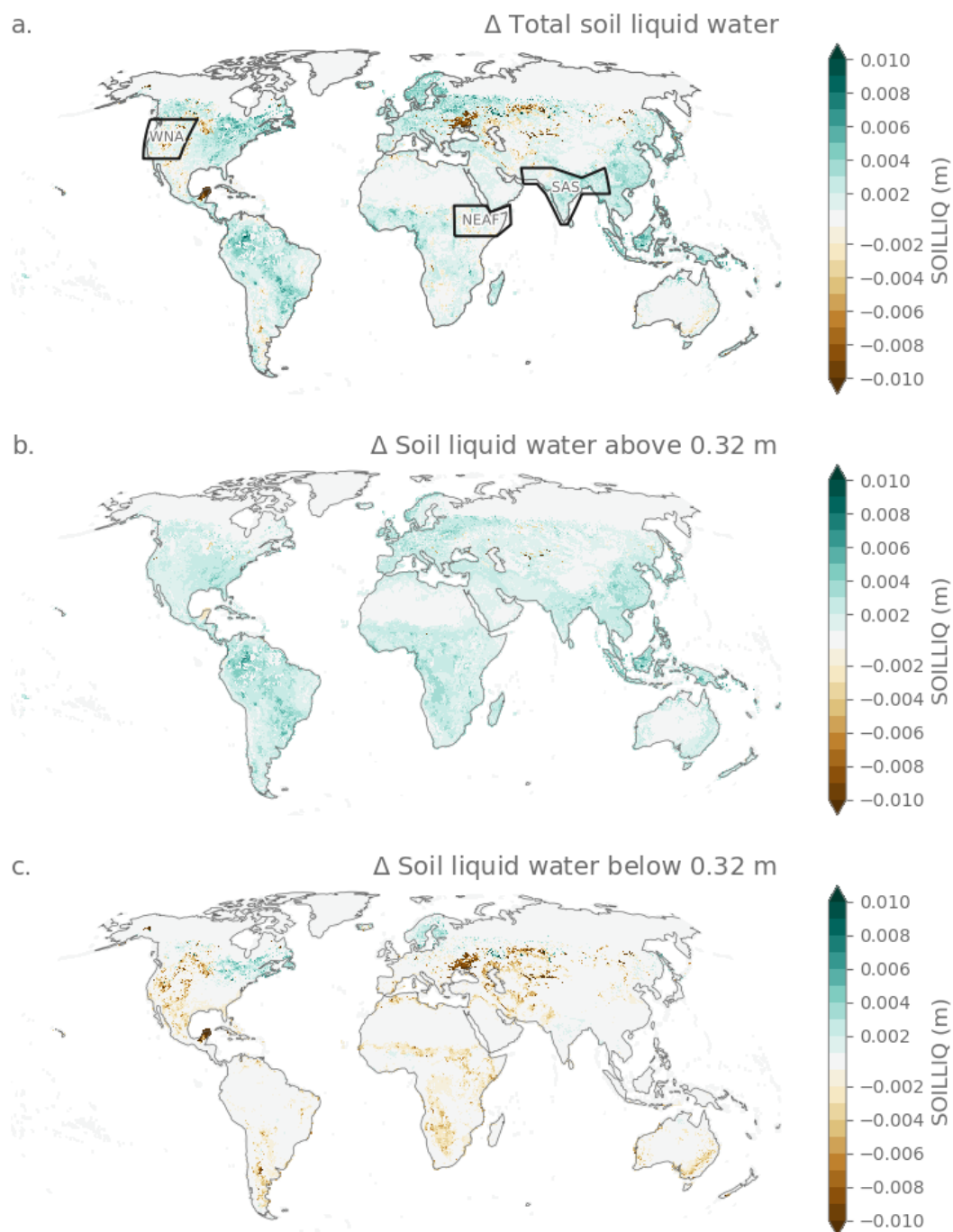


Figure A6. As Fig. A6, but for the medium scenario.

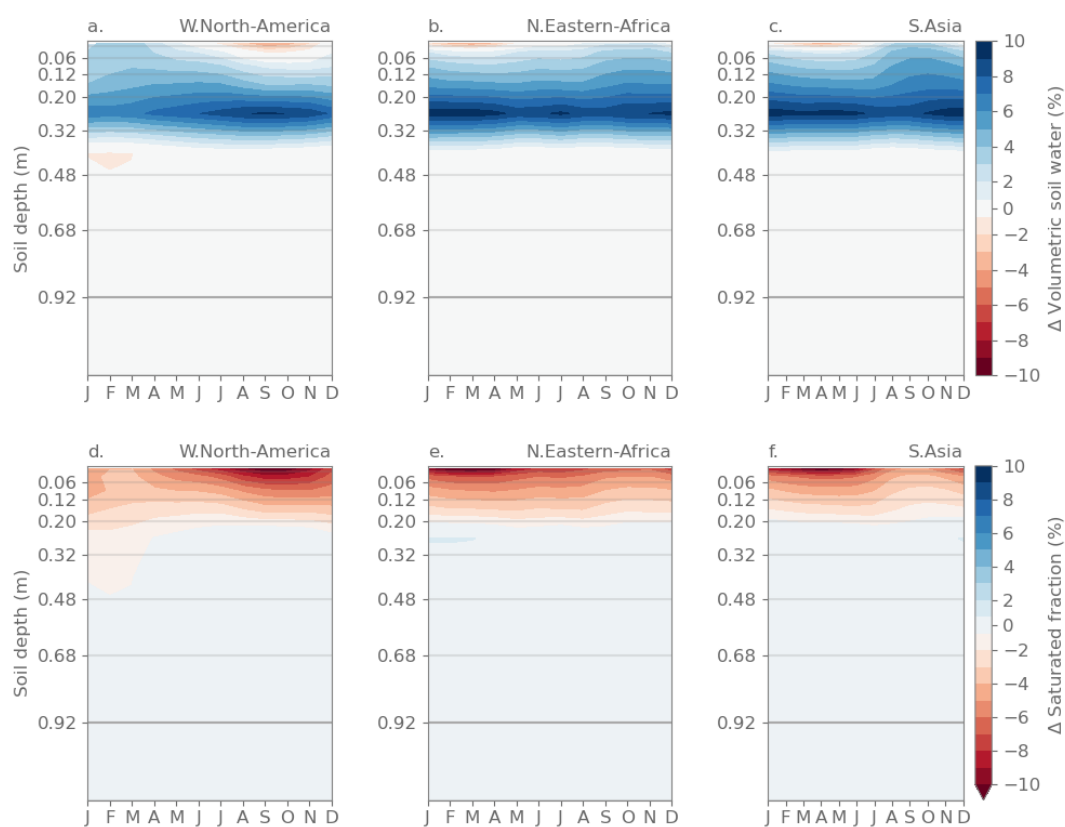


Figure A7. As fig. 4, but expressed as relative changes.

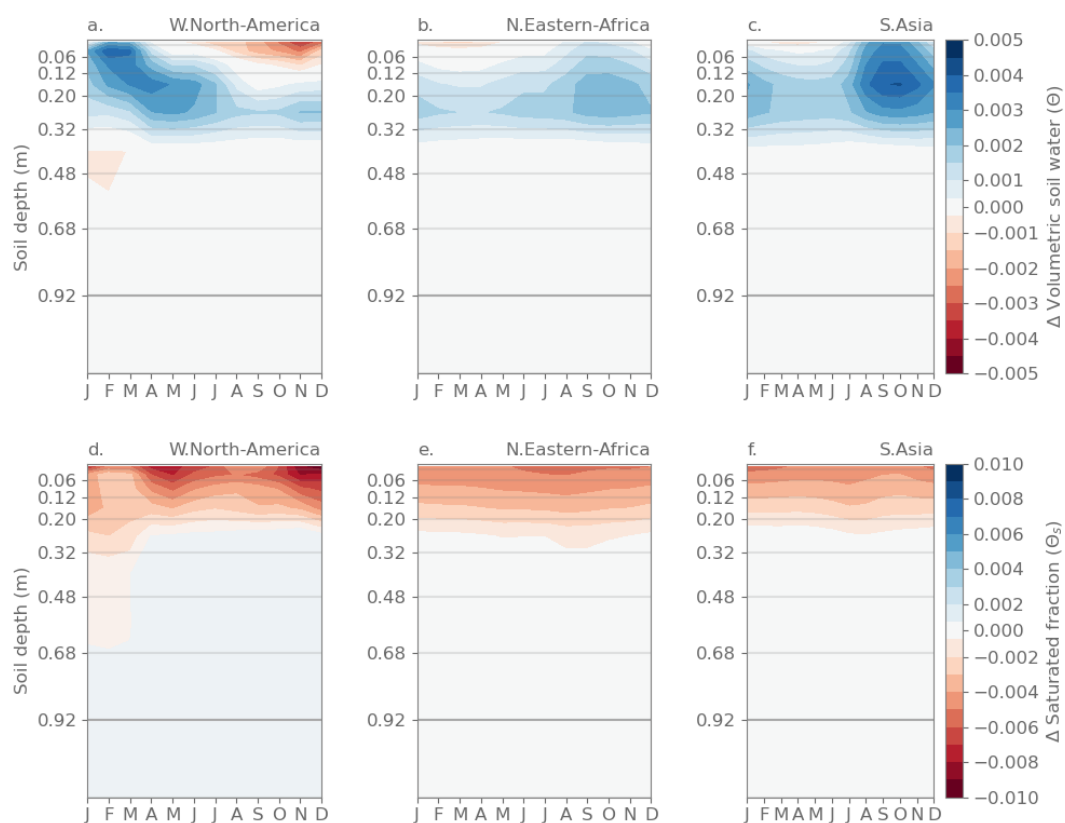


Figure A8. As Fig. 4, but for the 4p1000 scenario.

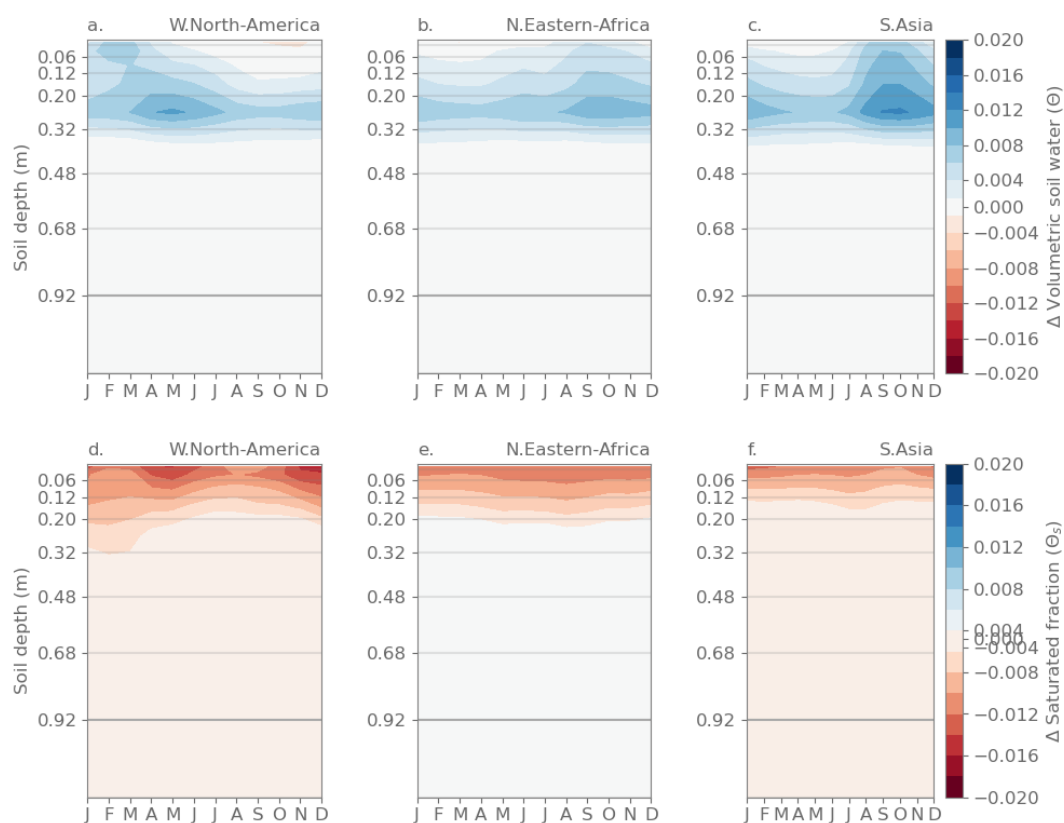


Figure A9. As Fig. 4, but for the medium scenario.

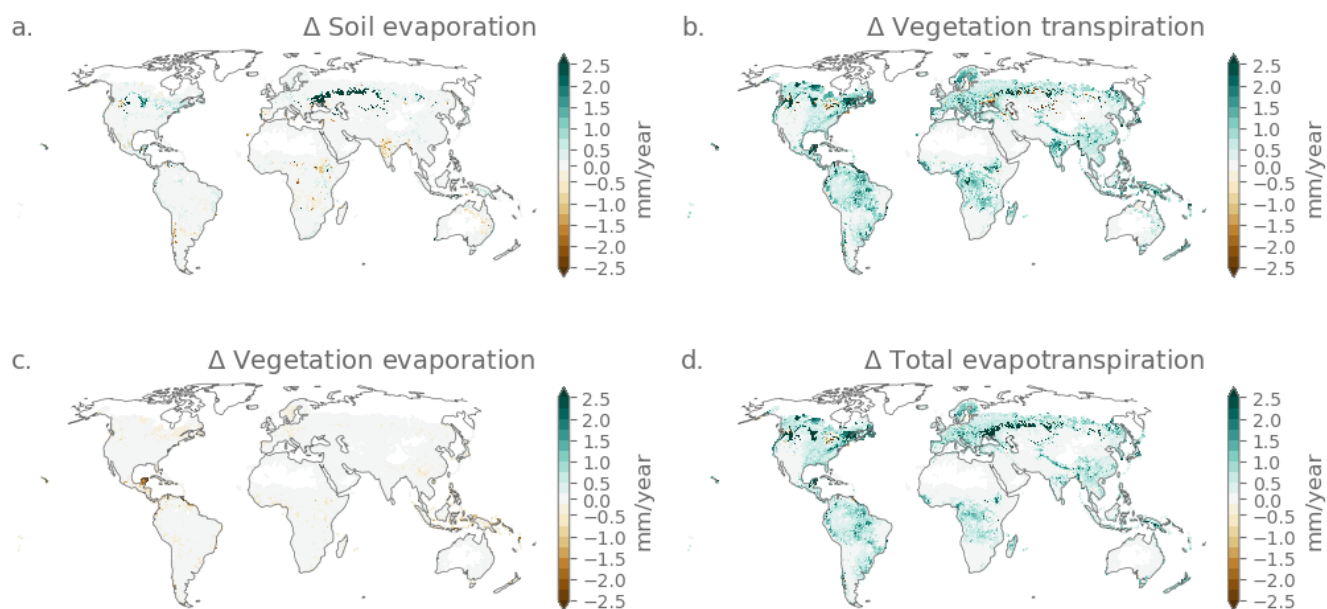


Figure A10. As Fig. 5, but for the 4p1000 scenario

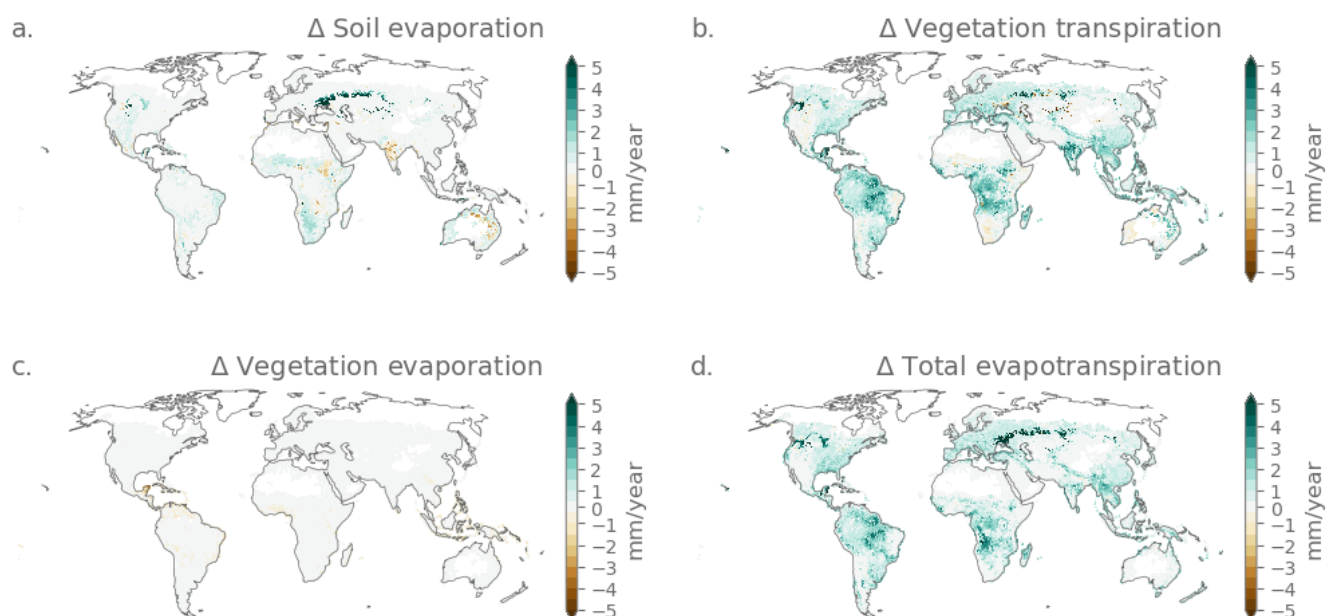


Figure A11. As Fig. 5, but for the medium scenario

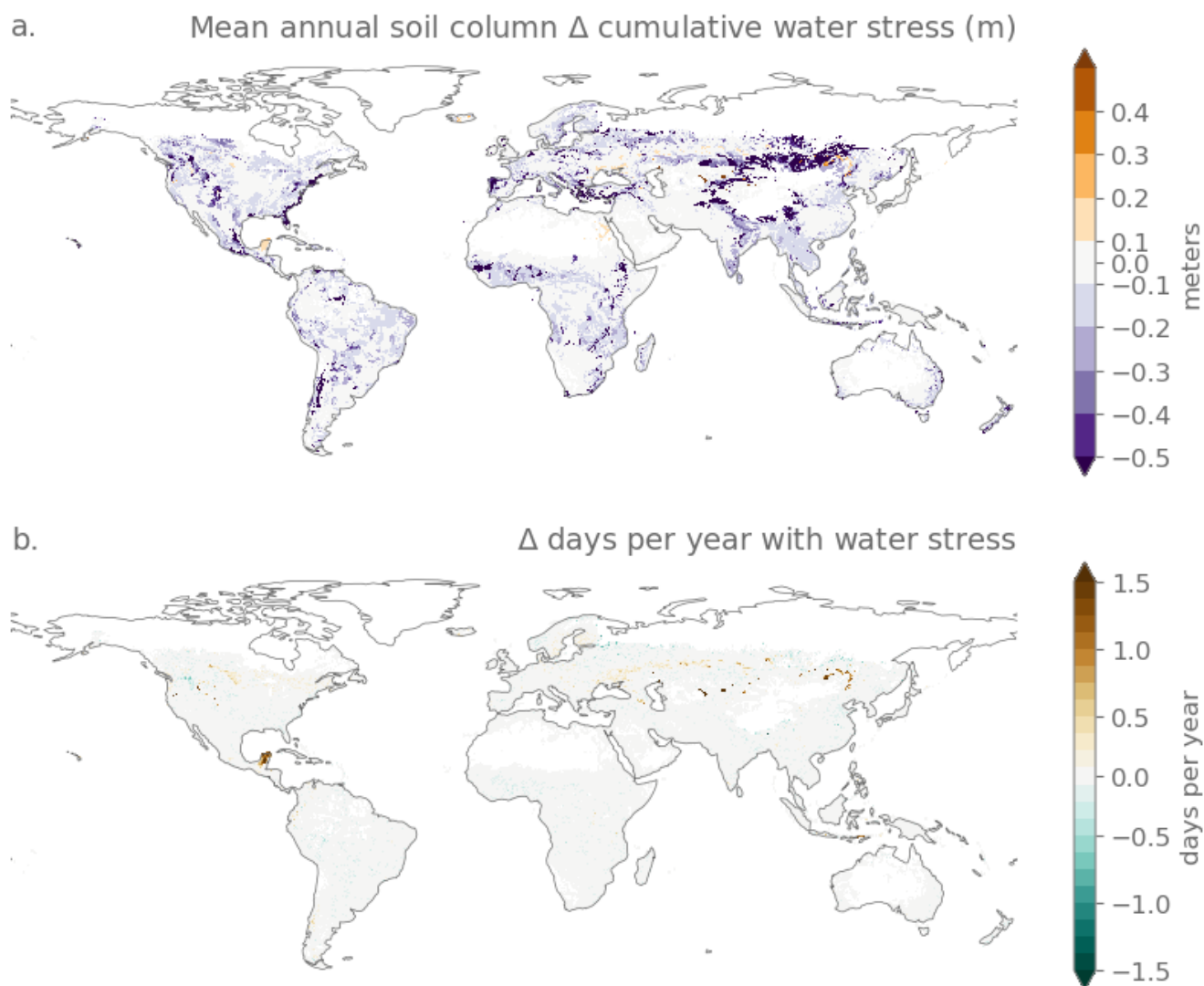


Figure A12. As Fig. 6, but for the 4p1000 scenario

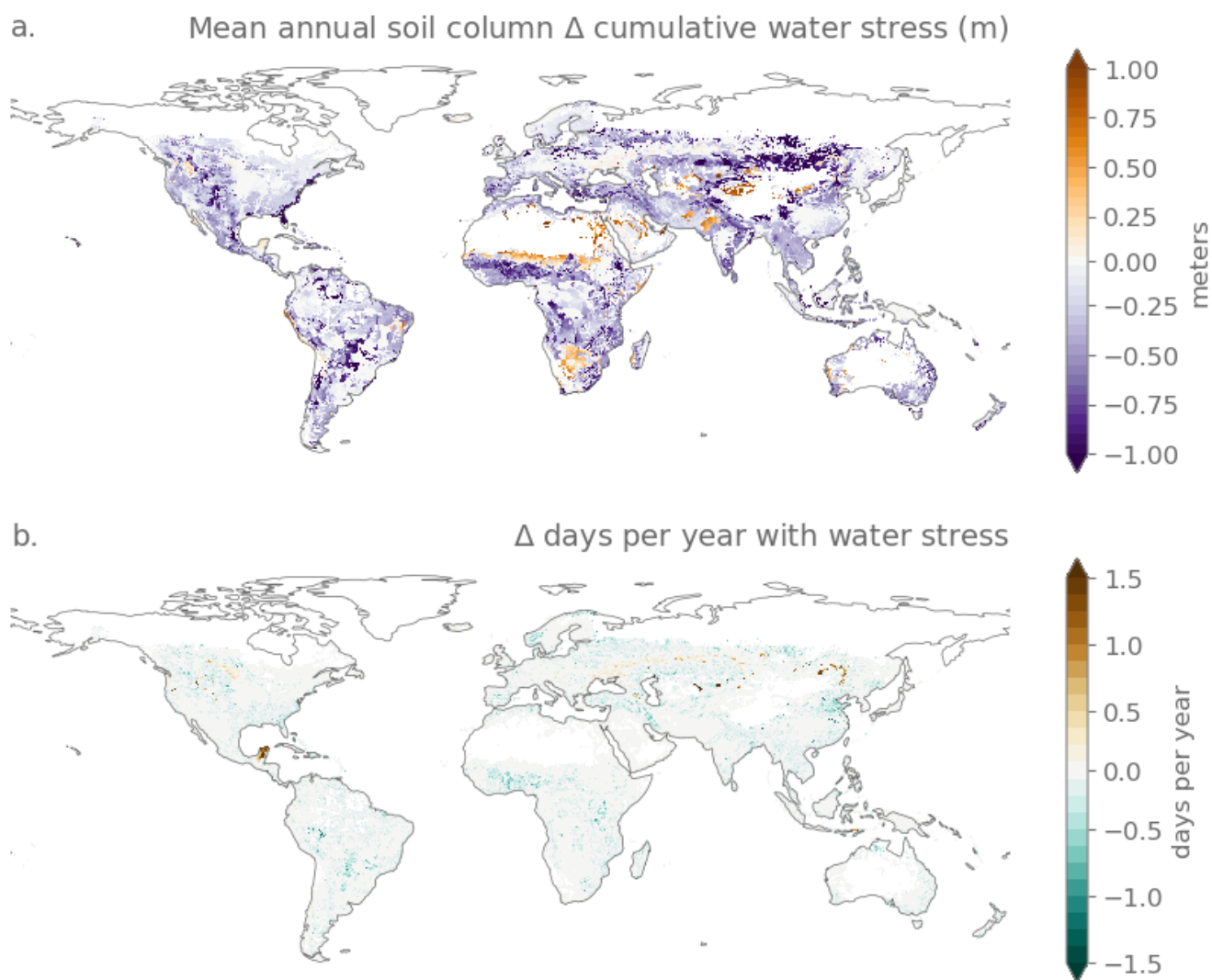


Figure A13. As Fig. 6, but for the medium scenario

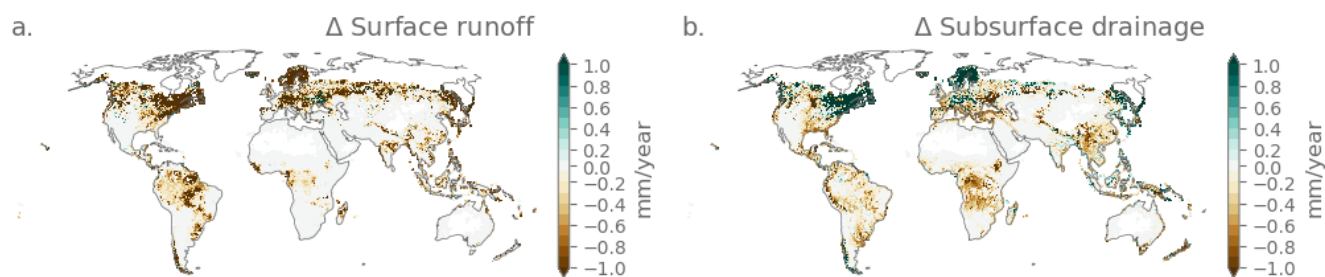


Figure A14. As Fig. 7, but for the 4p1000 scenario

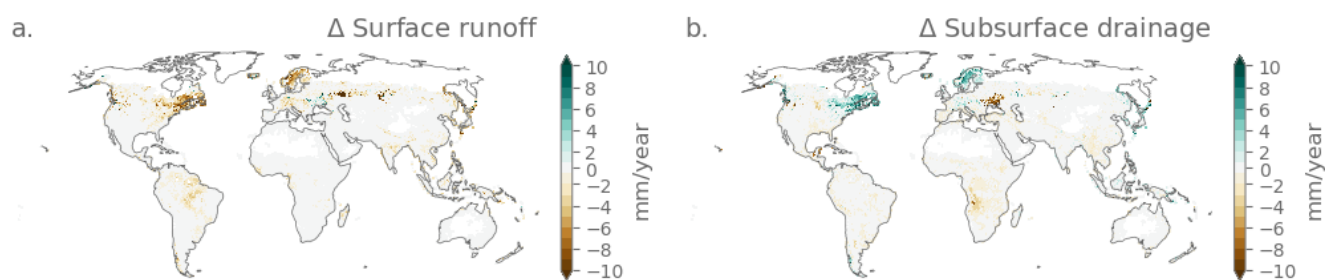


Figure A15. As Fig. 7, but for the medium scenario

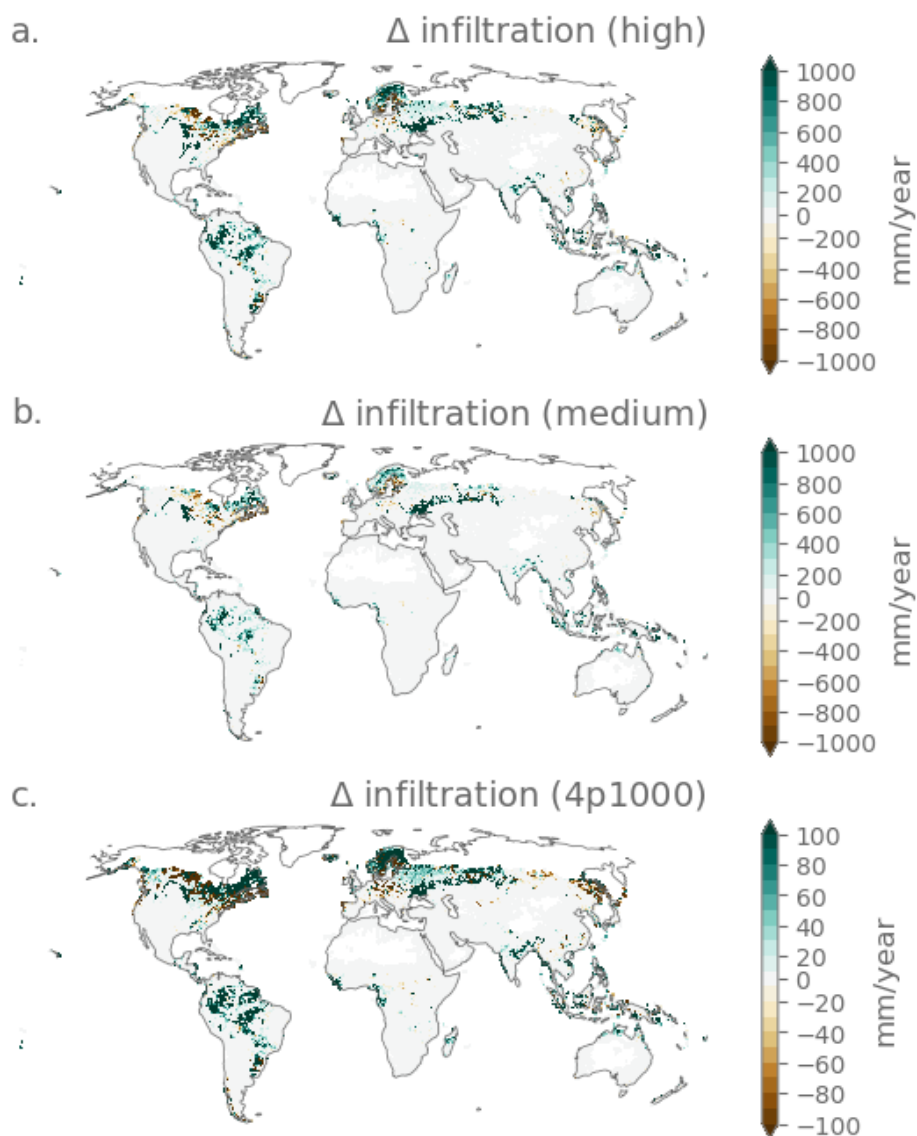


Figure A16. Effect of soil carbon sequestration on infiltration following the (a.) High, (b.) Medium and (c.) 4p1000 scenarios.



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