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27    **Key Points:**

- 28        • Proximal northward-recurving typhoons are the most likely to induce ozone  
29        pollution.
- 30        • The northward typhoon will cause ozone to increase by 0.3~12.3ppbv in  
31        vertical height.
- 32        • The contribution rate of transboundary layer transport under the influence of  
33        typhoon to the ozone in the boundary layer can reach 16%.

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### 39 **Abstract**

40       Ozone (O<sub>3</sub>) pollution has emerged as one of the core challenges in atmospheric  
41 environmental governance in China, particularly in Guangdong Province. As a crucial  
42 weather system during East Asian summers, typhoons exert profound influences on O<sub>3</sub>  
43 formation, accumulation, and transboundary transport through variations in their  
44 tracks and intensities. This study examined 237 historical typhoons occurring in China's  
45 coastal waters between 2013-2023, classifying them into three distinct trajectory  
46 types using k-means clustering: westward-moving typhoons (Type 1), Distant  
47 northward-recurving typhoons (Type2) and Proximal northward-recurving typhoons  
48 (Type3). By integrating ground-based observations, reanalysis data, and WRF-CMAQ  
49 model simulations to investigate the mechanisms through which typhoon tracks affect  
50 ozone pollution in Guangdong Province. The results demonstrate that for Guangdong  
51 Province, proximal northward-recurving typhoons induce more extreme  
52 meteorological conditions compared to westward-moving and distant northward-  
53 moving typhoons. Backward trajectory analysis reveals that northward-moving  
54 typhoons significantly enhance vertical downward transport of upper-level ozone,  
55 increasing ozone vertical gradients in Guangdong Province, with concentration  
56 enhancements of 2.5–11.6 ppbv (Type 2) and 0.3–12.3 ppbv (Type 3). The analysis of  
57 consecutive northward-moving typhoons' impact on ozone pollution in Guangdong  
58 Province reveals that surface photochemical reactions served as the dominant factor,  
59 while vertical downward transport of upper-level ozone acted as a secondary  
60 contributor. During this event, vertical transport contributed up to 39.9 ppbv to near-  
61 surface (100 m) ozone concentrations, with cross-boundary-layer transport accounting  
62 for up to 16% of boundary layer ozone concentrations, demonstrating that typhoon-  
63 induced vertical transport significantly enhances boundary layer ozone levels and  
64 consequently worsens surface pollution.



## 65 Plain Language Summary

66 It is well established that typhoon tracks exert significant impacts on ozone  
67 pollution. However, current research predominantly focuses on individual typhoon  
68 case studies or isolated meteorological factors, leaving a gap in comparative analyses  
69 of the mechanisms associated with different typhoon pathways. This study categorizes  
70 the trajectories of 237 typhoons that occurred over the western Pacific Ocean,  
71 specifically investigating the influence mechanisms of westward-moving typhoons  
72 (Type1), distant northward-moving typhoons (Type2), and Proximal northward-  
73 recurving typhoons (Type3) on ozone pollution in Guangdong Province. The results  
74 demonstrate that close-in northward-moving typhoons induce the most favorable  
75 conditions for ozone formation and the least favorable atmospheric dispersion  
76 conditions in Guangdong, thereby promoting ozone pollution. Additionally,  
77 northward-moving typhoons facilitate the subsidence of high-latitude, high-  
78 concentration ozone into the boundary layer, leading to elevated ozone levels. Finally,  
79 consecutive northward-moving typhoons trigger widespread and persistent ozone  
80 pollution across Guangdong. During this process, cross-boundary-layer transport via  
81 vertical motion contributes up to 16% of the ozone concentration within the boundary  
82 layer, underscoring the substantial impact of northward-moving typhoons on  
83 boundary-layer ozone through vertical transport mechanisms.

## 84 1 Introduction

85 Ozone ( $O_3$ ) pollution has become one of the core challenges in atmospheric  
86 environmental governance in China, particularly in the Pearl River Delta region. As a  
87 typical secondary pollutant, the formation of  $O_3$  is dually regulated by precursor  
88 emissions ( $NO_x$  and VOCs) and meteorological conditions. (Dou et al., 2024; Gong et  
89 al., 2025; Qiu et al., 2025; Yang et al., 2019). In recent years, despite continuous  
90 strengthening of anthropogenic emission control measures, the increasing frequency  
91 of extreme weather events has significantly amplified the complexity of  $O_3$  pollution.



92 (Chen et al., 2022a; Lu et al., 2024; Wan et al., 2022; Wang et al., 2024a). Among  
93 these factors, typhoons, as a crucial weather system during the East Asian summer,  
94 exert profound impacts on ozone ( $O_3$ ) formation, accumulation, and transboundary  
95 transport through their track and intensity variations, which significantly modify  
96 regional meteorological conditions (e.g., temperature, humidity, wind speed) and  
97 atmospheric transport processes. (Chen et al., 2021; Qu et al., 2021; Shen et al., 2023;  
98 Wang et al., 2022a).

99 The peripheral subsidence flows of typhoons frequently induce high  
100 temperatures, low humidity, and stagnant weather conditions, which enhances  
101 photochemical reactions while suppressing pollutant dispersion, consequently leads  
102 to localized  $O_3$  accumulation. (Chen et al., 2022b). Simultaneously, the heat stagnation  
103 induced by typhoon conditions favors biogenic emissions, with isoprene  
104 concentrations potentially doubling, leading to BVOCs contributing up to 10 ppbv to  
105  $O_3$  formation (Xu et al., 2023). Under the combined influence of the western Pacific  
106 subtropical high and typhoon peripheral circulation, tropical cyclones facilitate the  
107 downward transport of ozone-rich stratospheric air into the lower troposphere. This  
108 process leads to the formation of elevated ozone concentrations in the middle  
109 boundary layer. Subsequently, the downward transport of residual layer pollutants  
110 significantly contributes to the accumulation of ground-level pollutant concentrations  
111 during morning hours (Chen et al., 2021; Zhan et al., 2020; Chen et al., 2022c). The  
112 approach of typhoons significantly enhances both biogenic emissions and  
113 transboundary ozone transport, with observed increases reaching 78.0% and 22.5%,  
114 respectively. The abundance of precursors coupled with intensified photochemical  
115 reactions more than doubles ozone formation efficiency.(Wang et al., 2022a).  
116 Regarding the effect of vertical transport on ozone pollution under typhoon weather,  
117 some researchers attribute observed ozone increases primarily to enhanced surface-  
118 level photochemical activity (Huang et al., 2021; Jiang et al., 2024; Wang et al., 2025).  
119 In this view, upper-level subsidence flows primarily suppress tropospheric ozone



120 dispersion without making significant positive contributions to ozone transport (Ding  
121 et al., 2023; Li et al., 2022; Ouyang et al., 2022). Other studies have documented cases  
122 where typhoon-induced vertical mixing facilitates downward transport of elevated  
123 ozone layers (e.g., through stratospheric intrusions), generating measurable surface  
124 ozone enhancements of 10-15 ppbv (Chen et al., 2021).

125 Current research demonstrates that typhoon impacts on ozone pollution exhibit  
126 significant path dependence. Westward-moving typhoons induce increased net ozone  
127 production in the Pearl River Delta (PRD) core region prior to landfall, followed by a  
128 rapid decline to near-zero levels on the landfall day(Ding et al., 2023). During their  
129 initial development stage, typhoons enhance ozone production through subsidence-  
130 induced meteorological conditions. However, as they approach landfall, associated  
131 heavy precipitation and strong winds effectively scavenge pollutants, leading to  
132 negative ozone anomalies over the Yangtze River Delta region. These anomalies extend  
133 vertically, with maximum ozone reductions of 14-18 ppbv observed at 5 km  
134 altitude(Chen et al., 2021). When typhoons track northward across the Taiwan Strait  
135 through the low-latitude western Pacific, they trigger sequential regional ozone  
136 pollution episodes in both the Yangtze River Delta (YRD) and Pearl River Delta (PRD)  
137 regions(Wang et al., 2022b). The northerly peripheral circulation of such typhoons  
138 transports precursors from North China and the Yangtze River Delta (YRD) southward,  
139 which, when superimposed with local emissions, triggers abrupt ozone concentration  
140 increases(Shen et al., 2023). Successive northward-moving typhoons can elevate O<sub>3</sub>  
141 concentrations by 30% across eastern China while prolonging pollution duration(Wang  
142 et al., 2024b). Furthermore, the interaction between typhoons and the subtropical  
143 high can form a compound weather system, which exacerbates O<sub>3</sub> pollution intensity  
144 and prolongs its duration(Gao et al., 2020; Han et al., 2020a; Qin et al., 2020). However,  
145 current research predominantly focuses on individual typhoon cases or isolated  
146 meteorological factors(Kumar et al., 2023; Li et al., 2023a; Zhan et al., 2020), leaving  
147 significant gaps in comparative analyses of mechanisms associated with different



148 typhoon tracks. Key unresolved questions include: How do typhoons with distinct  
149 paths differentially modulate meteorological conditions and regional transport? How  
150 do large-scale circulation changes induced by varying typhoon tracks influence the  
151 vertical distribution of ozone? These questions demand systematic investigation to  
152 advance our understanding of typhoon-O<sub>3</sub> interactions.

153 As a high-frequency typhoon landing region, Guangdong Province exhibits  
154 particularly strong correlations between ozone pollution and typhoon activity  
155 (Shuping et al., 2022; Yaoyao et al., 2022). Statistical analyses reveal that over 80% of  
156 ozone exceedance days during Guangdong's summer-autumn seasons from 2015-  
157 2021 were typhoon-associated (Shen et al., 2023). Under climate change scenarios,  
158 observed trends of northward-shifting typhoon tracks and intensifying storm strength  
159 may further alter regional ozone pollution patterns (Guo and Tan, 2022). Consequently,  
160 elucidating the mechanistic links between typhoon paths and ozone pollution holds  
161 dual significance: advancing regional atmospheric multipollutant theory while  
162 providing scientific foundations for dynamic, precision-based ozone control strategies.

163 This study systematically investigates all typhoons near Guangdong Province from  
164 2013 to 2023 by integrating multi-source observational data and numerical  
165 simulations. Through comprehensive classification of typhoon tracks, we conduct in-  
166 depth analyses of the relationships between meteorological factors, circulation  
167 patterns, atmospheric transport, and three-dimensional ozone distribution under  
168 different typhoon paths. Specifically, we examine the contribution of upper-level  
169 transport to boundary layer ozone concentrations during typical typhoon events. The  
170 research aims to elucidate the differential impacts of various typhoon tracks on O<sub>3</sub>  
171 pollution in Guangdong region, thereby providing scientific support for refined air  
172 quality management strategies.

## 173 **2 Materials and Methods**

### 174 **2.1 K-means Clustering Analysis**

175 K-means represents one of the most prevalent partition-based clustering



176 methods. The algorithm categorizes  $n$  objects into  $K$  clusters based on a predefined  
177 parameter  $K$ , aiming to minimize the within-cluster sum of squares (WCSS) while  
178 maximizing the between-cluster sum of squares (BCSS). This ensures high intra-cluster  
179 similarity and low inter-cluster similarity. The K-means algorithm has been widely  
180 applied in atmospheric trajectory classification studies due to its effectiveness in  
181 identifying characteristic transport patterns (Han et al., 2020b; Yufeng et al., 2024; Zhu  
182 et al., 2023).

183 In this study, we performed two distinct clustering analyses using the K-means  
184 method: typhoon track clustering and atmospheric transport pathway clustering. For  
185 typhoon track clustering: 1. Targeted typhoon tracks over the western Pacific Ocean;  
186 2. Employed Euclidean distance metric for data point allocation; 3. Determined the  
187 optimal  $K$  value by identifying the elbow point where the rate of WCSS decrease  
188 substantially diminished; 4. Selected  $K=3$  as the optimal cluster number, yielding three  
189 distinct typhoon track types (**Fig.S3**). For atmospheric transport pathway clustering: 1.  
190 Analyzed 7-day three-dimensional backward trajectories; 2. Classified atmospheric  
191 transport channels into four categories (**Fig.S4**); 3. Implemented similar optimization  
192 procedures for cluster determination. The methodology ensures statistically robust  
193 classification of both typhoon trajectories and associated air mass transport patterns,  
194 providing a quantitative basis for subsequent ozone transport analysis.

195

## 196 2.2 HYSPLIT Trajectory Model

197 HYSPLIT is a complete system for computing simple air parcel trajectories, as well  
198 as complex transport, dispersion, chemical transformation, and deposition simulations. A common application is a back trajectory analysis to determine the origin of air masses and establish source-receptor relationships (Rolph et al., 2017; Stein et al., 2015).

202 This study employs the NOAA HYSPLIT Trajectory Model ([https://www.ready.noaa.gov/HYSPLIT\\_traj.php](https://www.ready.noaa.gov/HYSPLIT_traj.php)) to conduct backward trajectory simulations for 237 typhoon  
203





s in the Western Pacific region between 2013 and 2023. The meteorological data used is GDAS (1-degree resolution). The source location is set at 113.5°E, 23.6°N, with the backward trajectories initiated at 14:00 (local time) on the day of peak ozone pollution during each typhoon event. The backward simulation runs for 168 hours (7 days), with trajectory heights set at 500 m, 1000 m, and 2000 m above ground level (AGL).

209

### 2.3 WRF-CMAQ

The WRF-CMAQ modeling system was employed to simulate meteorological fields and ozone concentration variations during the typhoon process. The WRF (Weather Research and Forecasting) model version 3.9 was configured with the following parameterizations: Microphysics, WSM6 Scheme; Cumulus Parameterization: Grell-Freitas (GF) Scheme; Radiation: RRTMG Scheme; Boundary Layer: YSU Scheme; Surface Layer: MM5 Similarity Theory; Land Surface: Noah LSM. The large-scale meteorological fields and boundary conditions were derived from NCEP's Global 6-hourly FNL forecast data. The CMAQ (Community Multiscale Air Quality) model version 5.0.2 was implemented with the IPR (integrated process rate) analysis module. The CB05 mechanism was selected for gas-phase chemistry, while the AE6 mechanism was adopted for aerosol chemistry.

The modeling system utilized a triple-nested grid configuration (see **Fig.S1**) with Lambert conformal projection centered at 114°E, 28.5°N and two standard parallels at 15°N and 40°N. The outermost domain (D01) had a horizontal resolution of 27 km × 27 km, covering China, Southeast Asia and the western Pacific region. The intermediate domain (D02) featured a 9 km × 9 km resolution encompassing South China, while the innermost domain (D03) employed a 3 km × 3 km resolution focusing on Guangdong Province and surrounding cities. The vertical structure consisted of 14 layers with the model top set at 200 hPa. For the first and second nested domains, the air pollutant emission inventory adopted was the 0.25°×0.25° MEIC (Multi-resolution Emission Inventory for China) developed by Tsinghua University for the year 2020. For



232 the third (innermost) domain, a higher-resolution 3 km×3 km emission inventory  
233 compiled by the research team (Li et al., 2023b) was utilized. The simulation period  
234 spanned from 00:00 UTC on 24 August to 00:00 UTC on 31 August 2020.

235 In the present study, O<sub>3</sub> was used as a model pollutant to analyze the effects of  
236 atmospheric processes on the pollutants' value in deep convection events by using  
237 Integrated Process Rate (IPR) analysis. The IPR analysis in CMAQ can be used to  
238 calculate the influence of different atmospheric processes on the values of pollutants,  
239 and to quantify the importance of each process in the evolution of the pollutant  
240 value (Chen et al., 2018; Chen et al., 2022a). The processes include gas-phase chemistry  
241 (CHEM), vertical advection (ZADV), horizontal advection (HADV), vertical diffusion  
242 (VDIF), horizontal diffusion (HDIF), dry deposition (DDEP) and cloud processes (CLDS).

### 243 **3 Data**

#### 244 **3.1 Typhoon track data**

245 The typhoon track data were obtained from the CMA Best Track Dataset  
246 (tcdata.typhoon.org.cn) maintained by the Tropical Cyclone Data Center of China  
247 Meteorological Administration. This dataset provides 6-hourly positional and intensity  
248 records of tropical cyclones in the Northwest Pacific (including the South China Sea,  
249 north of the equator and west of 180°E) since 1949, covering all typhoons  
250 approaching/making landfall in China, with a spatial resolution of 0.1°×0.1° (Lu et al.,  
251 2021; Ying et al., 2014). For this study, we extracted all typhoon track data from  
252 January 1, 2013, to December 31, 2023, including temporal, geographical coordinates  
253 (longitude and latitude), and intensity information. After interpolating the data, we  
254 performed typhoon track classification using the K-means clustering method.

255

#### 256 **3.2 Ozone data**

257 The ground-level ozone monitoring data were obtained from the China National  
258 Environmental Monitoring Center (CNEMC). This dataset contains hourly  
259 concentrations of SO<sub>2</sub>, NO<sub>2</sub>, CO, O<sub>3</sub>, PM<sub>10</sub>, and PM<sub>2.5</sub> from 1,657 monitoring stations



260 across China. For this study, we extracted hourly  $O_3$  data from 105 stations within  
261 Guangdong Province (station locations are shown in **Fig.S2**). Following the "Technical  
262 Regulation on Ambient Air Quality Index (on trial)" (HJ 663-2013), we calculated the  
263 daily maximum 8-hour average ozone concentration (MDA8  $O_3$ ). Days with MDA8  $O_3$   
264 concentrations exceeding  $160 \mu\text{g}/\text{m}^3$  (approximately 75 ppbv) were identified as ozone  
265 exceedance days.

266 The TROPESS Chemical Reanalysis  $O_3$  Increment 6-Hourly 3-dimensional Product  
267 V1 dataset from NASA was utilized to investigate the three-dimensional spatial  
268 distribution of ozone under typhoon conditions  
269 ([https://disc.gsfc.nasa.gov/datasets/TRPSCRO3I6H3D\\_1/summary](https://disc.gsfc.nasa.gov/datasets/TRPSCRO3I6H3D_1/summary)). The data are part  
270 of the Tropospheric Chemical Reanalysis v2 (TCR-2) for the period 2005-2021. TCR-2  
271 uses JPL's Multi-mOdel Multi-cOnstituent Chemical (MOMO-Chem) data assimilation  
272 framework that simultaneously optimizes both concentrations and emissions of  
273 multiple species from multiple satellite sensors. The data files contains a year of data  
274 at 6-hourly resolution, and a spatial resolution of  $1.125 \times 1.125$  degrees at 27 pressure  
275 levels between 1000 and 60 hPa. This study extracted data from January 1, 2013 to  
276 December 31, 2021 for spatial analysis of ozone distribution.

277

### 278 3.4 Meteorological data

279 Meteorological data from ERA5 (the fifth-generation European Mesoscale  
280 Weather Forecasting Center reanalysis of global climate and weather for the past four  
281 to seven decades) was also adopted in order to understand the pollution  
282 characteristics. The temporal resolution of the data is hourly and the spatial resolution  
283 is  $0.25^\circ \times 0.25^\circ$ . The parameters extracted herein include 2-m temperature, surface  
284 relative humidity, total cloud cover, downward UV radiation at the surface, total  
285 precipitation, mean sea level pressure, the u-component and v-component of wind at  
286 the 10m, 175hPa and 900hPa level, boundary layer height, vertical velocity at the 850  
287 hPa level, the Geopotential at the 175hPa and 900hPa level.



288 (<https://cds.climate.copernicus.eu/cdsapp#!/dataset/reanalysis-era5-single->  
289 [levels?tab=overview](https://cds.climate.copernicus.eu/cdsapp#!/dataset/reanalysis-era5-single-levels?tab=overview)).

290

291 3.5 ground-level ozone reanalysis dataset

292 The ground-level MDA8 O<sub>3</sub> concentrations across China were obtained from the  
293 China 1km High-Resolution Daily Ground-Level Ozone (O<sub>3</sub>) Dataset (2000–2023), a  
294 high-resolution (1 km) product developed by Wei et al. and hosted on the National  
295 Earth System Science Data Sharing Platform (<http://geodata.nnu.edu.cn>) (Wei et al.,  
296 2022). This dataset was generated through an ensemble learning approach combining  
297 multi-source data, including hourly O<sub>3</sub> measurements from ~940 to 1,630 monitoring  
298 stations (2013–2020) under China’s Ministry of Ecology and Environment (MEE)  
299 network, OMI/Aura total-column O<sub>3</sub> and tropospheric NO<sub>2</sub> retrievals, downward solar  
300 radiation (DSR) and surface air temperature (TEM) from ERA5 reanalysis (0.1°  
301 resolution), emissions of NO<sub>x</sub>, VOCs, and CO from MEIC inventory, land cover from  
302 MODIS, elevation from SRTM, and population density from LandScan. The subset of  
303 data from January 1, 2013, to December 31, 2023, was temporally aligned with  
304 recorded typhoon tracks to assess the spatio-temporal variability of O<sub>3</sub> during periods  
305 with distinct typhoon track types.

## 306 **4 Results**

### 307 4.1 Characteristics of ozone pollution under different typhoon paths

#### 308 4.1.1 Typhoon track clustering

309 Through k-means clustering analysis, the 237 typhoon tracks over the western  
310 Pacific from 2013 to 2023 were classified into three distinct types (**Fig1.a-c**). Type 1  
311 comprises typhoons that form in the western Pacific, move into the South China Sea,  
312 and subsequently make landfall in South China or pass through its southern maritime  
313 areas (total: 105 cases). Type 2 consists of typhoons originating from low-latitude  
314 regions of the western Pacific that approach China before recurring northward,



315 traversing Japan and Korea before returning to the western Pacific basin (total: 77  
316 cases). Type 3 represents typhoons generated in low-latitude western Pacific regions  
317 that approach China and recurve northward, ultimately making landfall in China or  
318 dissipating near Japan/Korea (total: 55 cases).

319 For clarity, these three typhoon types are respectively designated as: Type 1:  
320 Westward-moving typhoons; Type 2: Distant northward-recurving typhoons; Type 3:  
321 Proximal northward-recurving typhoons. Temporal distribution analysis (**Fig1.d**)  
322 reveals that both Type 1 and Type 2 primarily occur from July to November, with peak  
323 frequency in autumn, while Type 3 is predominantly observed from July to September,  
324 showing maximum occurrence during summer.

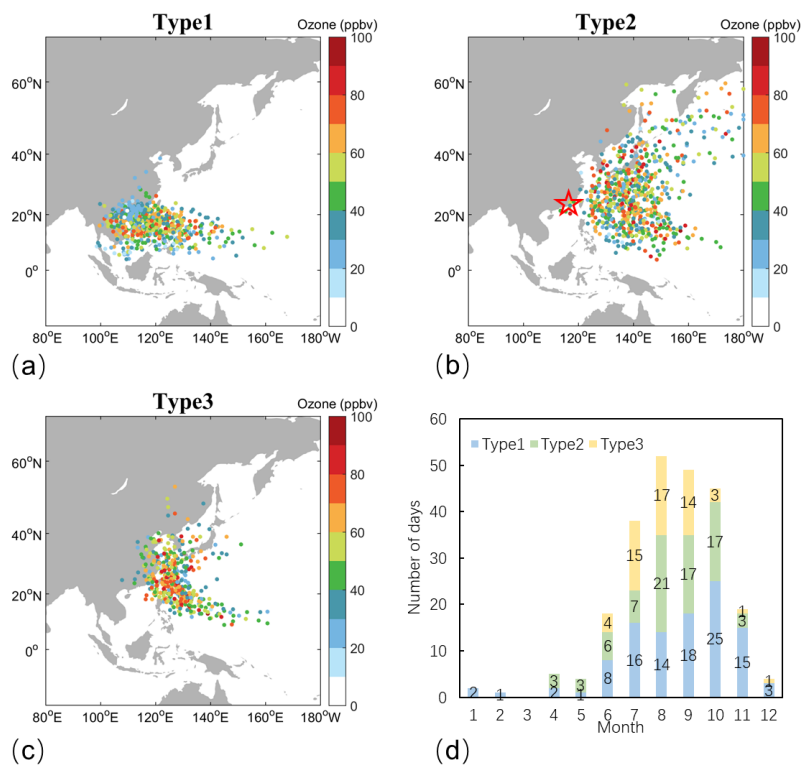
#### 325 4.1.2 Characteristics of ozone pollution

326 Figures 1a-1c illustrate the temporal evolution of maximum daily 8-hour average  
327 ozone (MDA8) concentrations in Guangdong Province in relation to typhoon track  
328 movements. From the perspective of ozone pollution characteristics, during the  
329 approach of Type 1 typhoons toward mainland China, ozone concentrations in  
330 Guangdong Province exhibit a gradual increase. If the typhoon does not make landfall,  
331 ozone concentrations remain elevated until the typhoon dissipates. However, if the  
332 typhoon makes landfall, ozone concentrations decrease rapidly due to precipitation  
333 and strong winds (**Fig. 1a**). Recent studies highlight the dual effects of typhoons on  
334 ozone: initial stages often enhance ozone through photochemical processes and  
335 stratospheric intrusions, whereas landfall phases suppress it via convective activity and  
336 precipitation (Chen et al., 2021; Li et al., 2021). Typhoons of Type2 can induce ozone  
337 concentration increases in Guangdong Province through long-distance influences, as  
338 demonstrated by cases where northward-moving typhoons beyond 40°N still triggered  
339 ozone elevation in Guangdong (**Fig.1b**). This phenomenon may be associated with  
340 large-scale transport of ozone and its precursors. Typhoons of Type3 tend to induce  
341 ozone pollution in Guangdong when approaching eastern China, with peak  
342 ozone concentrations occurring when the typhoon reaches approximately 25°N



343 latitude. Following typhoon landfall or eastward deflection, ozone  
344 concentrations decrease (**Fig. 1c**).

345 We extracted the MDA8 O<sub>3</sub> concentrations during each typhoon event and  
346 calculated Type-specific averages to examine ozone distribution patterns in  
347 Guangdong under different typhoon types (**Fig. 2**). The results demonstrate that: Type  
348 1 corresponds to MDA8 O<sub>3</sub> concentrations ranging 9.2-70.9 ppbv, with an average of  
349 20 monitoring stations exceeding standards. Type 2 shows MDA8 O<sub>3</sub> concentrations of  
350 12.2-90.3 ppbv, averaging 34 exceedance stations. Type 3 exhibits MDA8 O<sub>3</sub>  
351 concentrations of 3.3-89.7 ppbv, with 35 stations exceeding limits on average. The  
352 spatial analysis reveals that ozone hotspots for all types consistently cluster in central  
353 Guangdong, indicating similar spatial distribution patterns despite varying intensity.  
354 Type3 exhibited the highest number of non-compliant monitoring sites, while Type1  
355 showed the lowest count.



356



Figure 1. (a-c) Maximum daily 8-hour average (MDA8) ozone concentrations in Guangdong Province (marked by red pentagrams) under different typhoon tracks(Different colors of dots represent the average ozone concentration at all monitoring stations in Guangdong Province when the typhoon is at that location), and (d) the corresponding temporal distributions of typhoon occurrences for each track type.

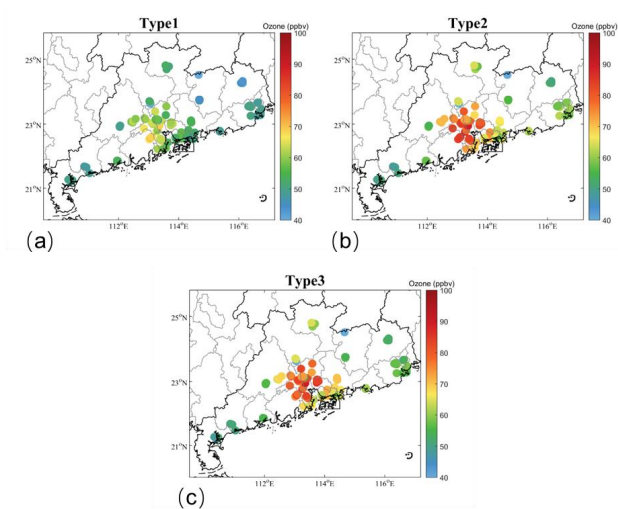


Figure. 2 Distribution of ozone pollution under different typhoon paths

#### 4.1.3 Meteorological characteristics

To investigate the influence of meteorological factors on ozone pollution in Guangdong Province under different typhoon tracks, we compared the differences in meteorological conditions between three types of typhoon weather and non-typhoon weather in Guangdong. Typhoon conditions refer to the day with the most severe pollution during the typhoon event, while non-typhoon conditions correspond to the remaining periods after excluding the entire typhoon process. The meteorological factors analyzed included surface temperature, total cloud cover, surface solar radiation, precipitation, surface relative humidity, boundary layer height, 10m wind speed, and vertical velocity at 850 hPa. All meteorological data were extracted from ERA5 at 14:00 local time for comparative analysis.

The results indicate that, compared to non-typhoon weather, typhoon weather in Guangdong is characterized by higher temperatures, stronger solar radiation, lower cloud cover, reduced precipitation, lower relative humidity, higher boundary layer



378 height, weaker surface winds, and suppressed vertical motion (**Fig.3**). The peripheral  
379 circulation of typhoons modifies the thermodynamic and dynamic structure of the  
380 boundary layer, creating an "ideal reactor" for ozone formation. Near-surface  
381 conditions of high temperatures, low humidity, and weak winds foster a stable  
382 boundary layer structure, significantly enhancing photochemical reaction rates (Ding  
383 et al., 2023). Additionally, increased solar radiation and elevated boundary layer height  
384 further expand the spatial domain for ozone production.

385 A comparison of meteorological characteristics across different typhoon track  
386 types reveals that Type 3 corresponds to what may be termed "extreme"  
387 meteorological conditions. It brings high temperature(32.4°C), high radiation  
388 (0.28MJ/m<sup>2</sup>), low cloud cover (0.47), low precipitation (0.08mm), low relative humidity  
389 (65.4%), high boundary layer height (1.03km), low wind speed(2.08m/s), and less  
390 vertical movement (-0.02pa/s) meteorological conditions, which are more likely to  
391 cause ozone pollution in Guangdong Province. Compared to non-typhoon conditions,  
392 Type 3 exhibits a temperature increase of 7.6°C, a cloud cover reduction of 0.28, a  
393 radiation intensity enhancement of 0.09 MJ/m<sup>2</sup>, and a boundary layer height elevation  
394 of 0.21 km. It demonstrates the poorest horizontal diffusion conditions, with a near-  
395 surface wind speed of 1.12 m/s lower than non-typhoon conditions. The severe ozone  
396 pollution observed in Guangdong Province results from the combined effects of strong  
397 ozone production rates and poor diffusion conditions, creating a synergistic  
398 amplification of pollution levels. The photochemical reaction conditions in Type2 are  
399 slightly weaker than those in Type3; however, reduced precipitation inhibits the wet  
400 scavenging of ozone and its precursors. Additionally, strong subsidence at the 850 hPa  
401 level not only suppresses the vertical diffusion of pollutants within the boundary layer  
402 but also transports ozone from higher altitudes downward, further increasing surface  
403 ozone concentrations. Compared to the other two typhoon types, Type1 exhibits  
404 weaker ozone formation conditions and better dispersion, resulting in the least severe  
405 ozone pollution.



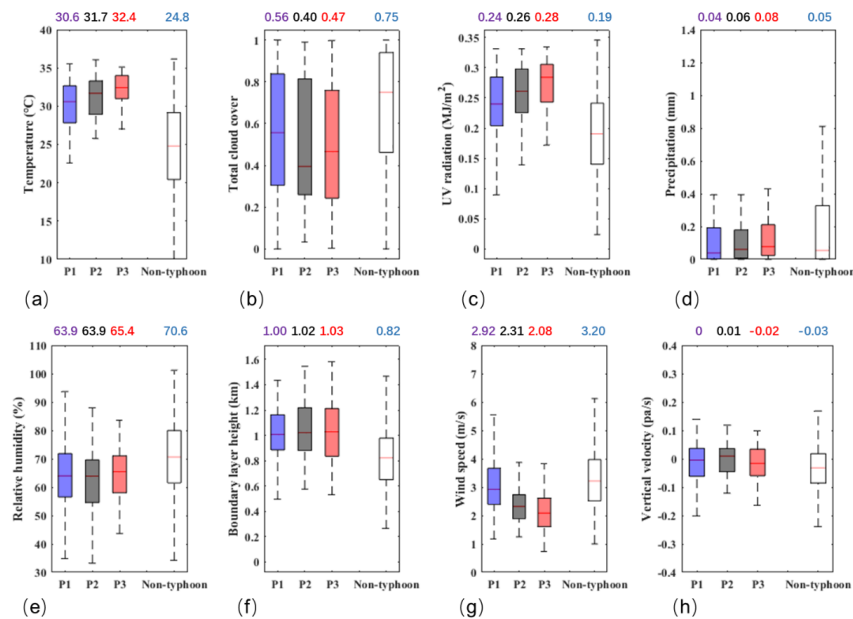


Figure 3. Comparison of meteorological conditions between typhoon and non-typhoon weather. (a–f) represent 2m temperature, total cloud cover, surface solar radiation, precipitation, relative humidity, boundary layer height, 10m wind speed, and vertical velocity at 850 hPa, respectively. P1, P2, and P3 denote three distinct typhoon tracks, while Non-typhoon refers to non-typhoon conditions. The numerical values above each boxplot indicate the median of the corresponding dataset.

## 4.2 Effect of regional transport on ozone distribution

### 4.2.1 Three-dimensional spatial distribution of ozone

The impact of typhoons on ozone extends beyond creating favorable photochemical conditions. The regional transport induced by large-scale circulation plays a pivotal role in determining ozone concentration distribution(Chen et al., 2022b; Wang et al., 2018). Typhoon tracks modify regional airflow patterns, facilitating cross-regional transport of ozone and its precursors(Chen et al., 2021). This study employs three-dimensional reanalysis O<sub>3</sub> data (2013-2021) coupled with wind fields and geopotential height to examine how typhoon-induced regional transport affects the three-dimensional spatial distribution of ozone concentrations (**Fig4-5**). When the typhoon moves northward (type 2 and type3), a high-pressure center emerges over



424 western China at the 175 hPa level, causing a southward displacement of the westerly  
425 jet. Under this circulation pattern, an ozone transport pathway is established,  
426 extending from high to low latitudes and accompanied by subsidence (**Fig. 4b, c**).  
427 Through this transport channel, stratospheric ozone with high concentrations (>75  
428 ppbv) is advected southward to approximately 20°N and descends below the 500 hPa  
429 level (**Fig. 5e, f**). Mechanistic analysis demonstrates that the combined effects of the  
430 westerly jet and mid-latitude high-pressure systems on typhoon motion create upper-  
431 tropospheric wind convergence, which enhances stratosphere-to-boundary-layer  
432 transport of ozone-rich air from the North China Plain (Meng et al., 2022). In contrast,  
433 westward-propagating typhoons (Type 1) do not generate perturbations in the  
434 westerly jet, and no pronounced southward transport or subsidence of upper-level  
435 ozone is evident (**Fig. 5a, d**).

436 As demonstrated by recent studies (Wang et al., 2022b; Yufeng et al., 2024), the  
437 peripheral circulation of western North Pacific typhoons can effectively transport  
438 ozone and its precursors from source regions (including the Yangtze River Delta, Fujian,  
439 and Anhui provinces) to Guangdong through well-organized atmospheric transport  
440 pathways. Analysis of ozone distribution at the 900 hPa level reveals that northward-  
441 moving typhoons not only induce ozone pollution in Guangdong, but also lead to  
442 elevated ozone concentrations in the Beijing-Tianjin-Hebei and Yangtze River Delta  
443 regions (**Fig. 5b,c**). During the typhoon's northward progression, the low-pressure  
444 center traverses China's eastern coastal areas, where cyclonic circulation facilitates  
445 southward transport of pollutants along the coast, ultimately impacting Guangdong  
446 Province (**Fig. 4e,f**).

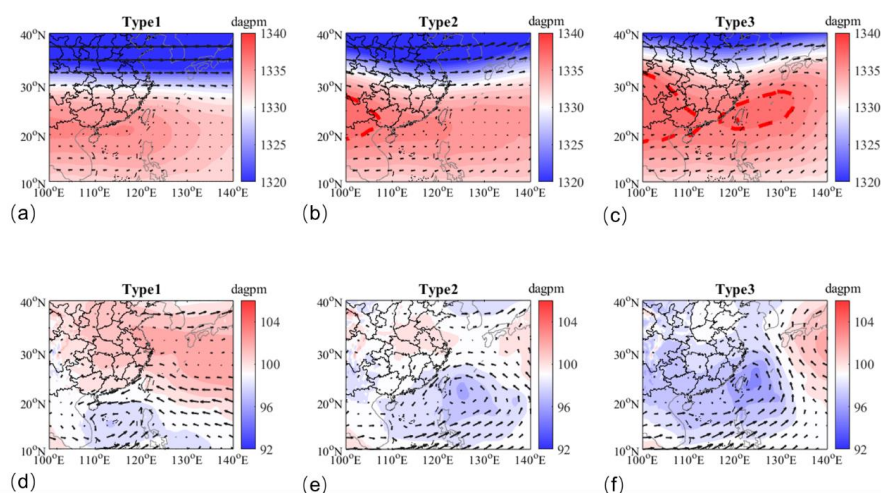


Figure 4. Comparison of circulation patterns under different typhoon tracks. (a-c) show geopotential height and wind fields at 175 hPa (upper panels) and 900 hPa (lower panels), respectively. The red curves indicate the positions of high-pressure centers.

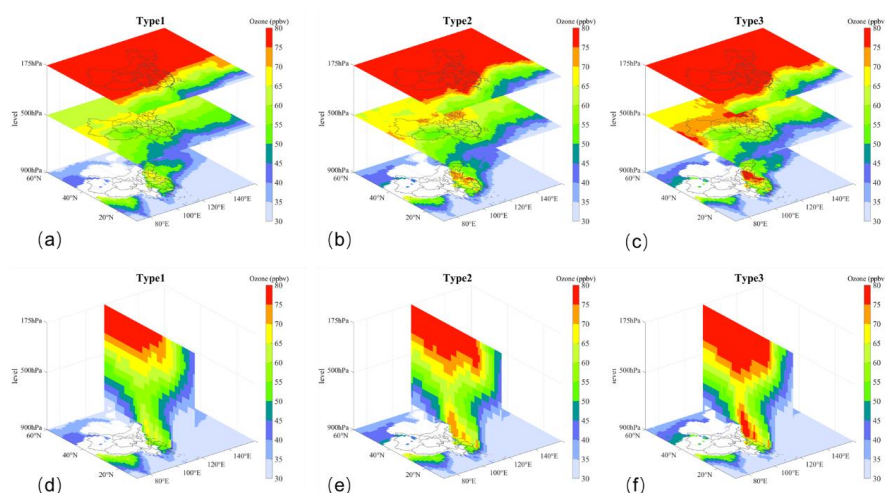


Figure 5. Three-dimensional spatial distribution of ozone under different typhoon tracks. (a-c) Horizontal ozone distributions at 900 hPa, 500 hPa, and 175 hPa for the three typhoon track types. (d-f) Horizontal ozone distributions at 900 hPa and corresponding vertical cross-sections along 114°E for each typhoon type.

**Figure.S5** presents the spatial distribution of ground-level MDA8  $O_3$  concentrations across China, as derived from the reanalysis 1 km high-resolution daily dataset, under three distinct typhoon track types (type 1, type 2, and type 3). The



460 analysis focuses on typhoon events, characterized as the date with the highest number  
461 of ground monitoring sites exceeding the  $160 \mu\text{g}/\text{m}^3$  ( $\sim 75\text{ppbv}$ ) MDA8  $\text{O}_3$  threshold  
462 during the entire typhoon track.

463 Being consistent with spatial distribution of ground monitoring  $\text{O}_3$  concentrations  
464 in section 4.1.2, here reveals significant spatial heterogeneity in  $\text{O}_3$  concentrations  
465 across typhoon track types, particularly in Guangdong Province, where the mean  
466 MDA8  $\text{O}_3$  follows the order: Type 2 ( $56.9 \text{ ppbv}$ ) > Type 3 ( $54.6 \text{ ppbv}$ ) > Type 1 ( $51.25$   
467  $\text{ppbv}$ ). This variability is attributed to differences in regional transport pathways and  
468 precursor availability. Specifically, type 2 typhoons exhibit elevated  $\text{O}_3$  levels in eastern  
469 China but reduced concentrations in northern and central regions compared to type 3.  
470 The enhanced  $\text{O}_3$  under type 2 conditions is driven by two synergistic mechanisms: (1)  
471 intensified low-tropospheric transport along China's eastern coastal region, as  
472 evidenced by atmospheric circulation patterns (**Fig. 4e**), and (2) the advection of  $\text{O}_3$ -  
473 rich air masses from northern and central China, which supply abundant precursors to  
474 Guangdong, particularly its eastern sector. Type 3 typhoons facilitate a more direct,  
475 meridional transport of  $\text{O}_3$  from northern and central China, coupled with pronounced  
476 stratospheric intrusions that enhance upper-tropospheric  $\text{O}_3$  contributions (**Fig. 5c and**  
477 **5f**). While type 2 systems lack the robust northern transport pathway observed in type  
478 3, they compensate via secondary  $\text{O}_3$  delivery through coastal advection, which  
479 subsequently propagates inland. This dual transport mechanism culminates in the  
480 highest  $\text{O}_3$  concentrations in Guangdong, especially the eastern and coastal part,  
481 during type 2 events.

482 Collectively, integrating atmospheric dynamics (**Fig.4**), three dimensional  
483 evolution of  $\text{O}_3$  (**Fig.5**), and ground-level  $\text{O}_3$  distributions (**Fig.S5**), underscores the  
484 critical role of typhoon-track-dependent transport pathways in modulating regional  $\text{O}_3$   
485 pollution. These highlight the necessity of considering multi-scale meteorological  
486 processes in air quality forecasting and quantifying their contributions to  $\text{O}_3$   
487 concentrations across different vertical levels.



488 To further investigate typhoon-induced ozone variations, spatial ozone  
489 concentration differences between typhoon conditions and non-typhoon conditions  
490 (June-November) were calculated (**Fig.6**). The June-November period was selected to  
491 eliminate seasonal influences. The results indicate that northward-moving typhoons  
492 (Type 2 and Type 3) can induce substantial ozone increases throughout the vertical  
493 column (200-900 hPa) (**Fig. 6b,c**). At the central point (113.23°E, 23.16°N), ozone  
494 concentration changes ranged between 2.5-11.6 ppbv (Type 2) and 0.3-12.3 ppbv  
495 (Type 3). In contrast, Type 1 did not cause significant high-altitude ozone increases,  
496 with central point ozone concentration changes ranging from -3.2 to 0.99 ppbv. Studies  
497 indicate that when gravity waves break in the upper troposphere and lower  
498 stratosphere on the western side of typhoon centers, intense turbulence occurs,  
499 leading to stratosphere-troposphere exchange (STE) (Huang et al., 2024). Subsequently,  
500 typhoons approaching landfall significantly enhance cross-regional ozone transport  
501 from North China to South China through STE (Wang et al., 2024c). This suggests that  
502 after Types 2 and 3 typhoons move northward, their cyclonic circulations transport  
503 high-concentration ozone from the tropopause to lower latitudes and altitudes  
504 through STE, causing significant changes in ozone vertical distribution and increased  
505 ozone concentrations within the boundary layer.

506

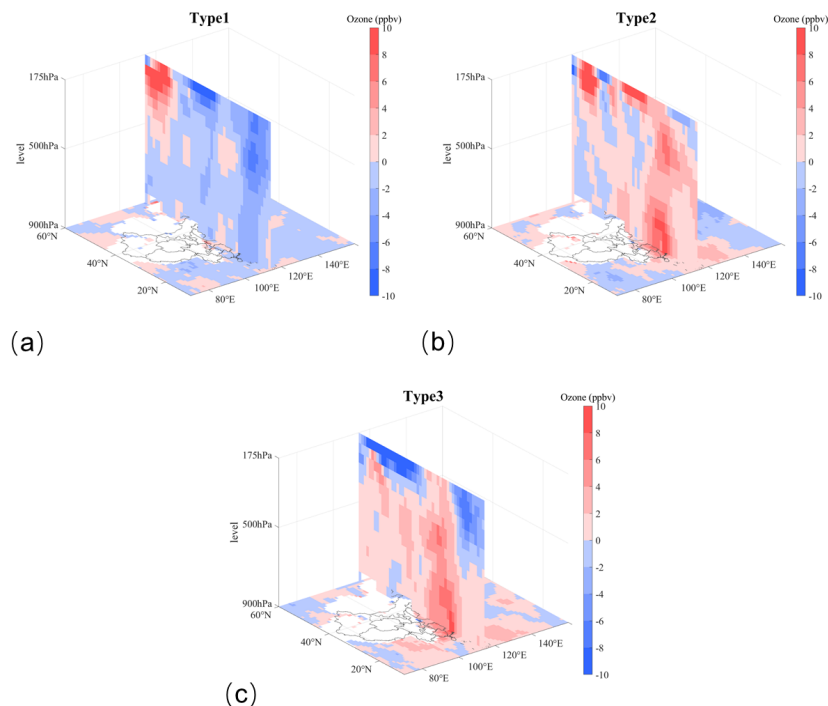


Figure 6. Ozone concentration changes induced by different typhoon types (a-c: horizontal distribution changes at 900 hPa and vertical cross-section changes along 114°E for each typhoon track type respectively).

#### 4.2.2 Boundary layer ozone

To investigate the ozone transport pathways within the boundary layer over Guangdong Province under typhoon conditions, and to examine the differences in ozone sources associated with distinct typhoon tracks, this study conducted HYSPLIT backward trajectory analysis for 237 typhoon events. The analysis focused on 7-day air mass origins at 500m altitude over central Guangdong (**Fig.S6**). For each typhoon type, cluster analysis of air mass origins was performed. After K-value screening, the air mass origins were classified into four trajectory clusters (**Fig.7**). **Table S1** presents statistics for each trajectory type, including: (1) The percentage of different trajectories, (2) mean ozone concentrations along trajectories, and (3) corresponding surface ozone concentrations.



523 Under Type 1 conditions, air masses in the target area mainly originated from  
524 within the boundary layer, accounting for 60.8%, with air transported from the South  
525 China Sea below 841 meters to Guangdong Province (Traj\_4). Analysis of the  
526 subtropical high's influence shows that under this typhoon type, Guangdong  
527 experienced the highest surface pressure and was closest to the subtropical high  
528 (**Fig.S7**). Research indicates that under the influence of the subtropical high, O<sub>3</sub>  
529 pollution is primarily affected by local emissions (Chen et al., 2024). This aligns with  
530 Traj\_4's characteristics of short transport distance and low altitude. The other three  
531 trajectories originated from northwest China (Traj\_1, 8.1%), western China (Traj\_2,  
532 13.5%), and central China (Traj\_3, 17.6%) respectively (**Fig.7a**). The trajectory with the  
533 highest surface ozone concentration was Traj\_2, which descended from 3794.1 meters  
534 with an average ozone concentration of 50.3 ppbv along the trajectory(**Fig.7d**,  
535 **Table.S1**). Under Type 2 conditions, nearly half of the air masses in the target area  
536 originated from northwest China (Traj\_1, 21.7% and Traj\_2, 23.9%), while the other  
537 half came from the South China Sea region (Traj\_3, 26.1% and Traj\_4, 28.9%)(**Fig.7b**).  
538 Among these, Traj\_1 and Traj\_2 air masses descended from above 2000m, whereas  
539 Traj\_3 and Traj\_4 air masses were transported within the boundary layer(**Fig.7b**). The  
540 trajectory with the highest surface concentration was Traj\_1, which descended from  
541 2646 meters with an average ozone concentration of 61.9 ppbv along the trajectory  
542 (**Fig.7e, Table.S1**). Under Type 3 conditions, Traj\_1 carried high-concentration ozone  
543 (>75 ppbv) from high-altitude (6356m) over high-latitude areas through North China  
544 to the target region, corresponding to the highest surface ozone concentration (15.2%  
545 proportion) (**Fig.7f, Table.S1**). The other three trajectories originated from central  
546 China (Traj\_2, 18.2%) and the South China Sea region (Traj\_3, 30.3% and Traj\_4,  
547 36.4%)(**Fig.7c**).

548 A comparative analysis of air mass trajectories from different directions  
549 demonstrates that marine air masses originating from the South China Sea are  
550 characterized by lower altitudes and extended residence time over Guangdong



Province, thereby constituting local ozone pollution sources. Conversely, continental  
air masses exhibit longer transport pathways and higher altitudes, representing  
regional ozone transport sources. Quantitative analysis reveals that the proportional  
contributions of local pollution sources under different typhoon tracks are 60.8%,  
55.0%, and 66.4%, respectively. Analysis of long-range transport trajectories reveals  
that different typhoon types can respectively deliver ozone from maximum altitudes  
of 7,468 meters (~380 hPa), 8,927 meters (~320 hPa), and 9,980 meters (~250 hPa)  
into the boundary layer. Type 2 and Type 3 exhibit significantly greater proportion from  
upper-level air mass transport (23.9% and 15.2% respectively) compared to Type 1.  
These typhoons can transport ozone from altitudes down into the boundary layer.  
Combined with the high ozone concentrations along the atmospheric transport  
pathways, this results in boundary-layer ozone increases of 10.7 ppbv and 12.3 ppbv  
for these two types, respectively (**Fig6.b-c**).



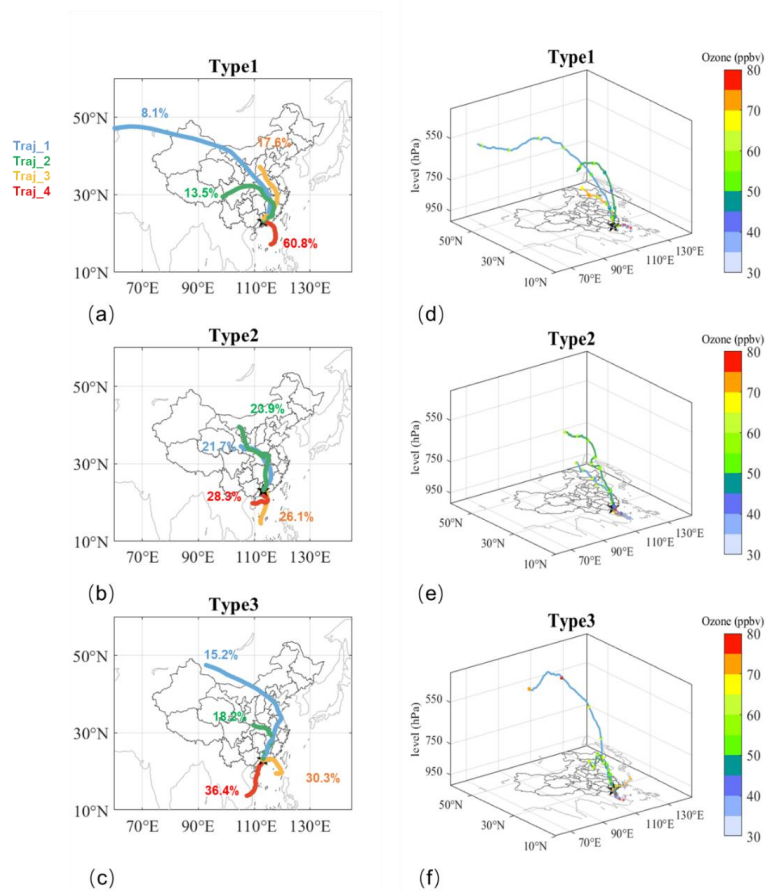


Figure 7. Comparison of boundary-layer air mass trajectory sources under different typhoon tracks: (a-c) two-dimensional views with trajectory percentages indicated numerically; (d-f) three-dimensional views showing ozone concentrations (ppbv) along trajectories (marked by colored points); target regions are denoted by black pentagrams on maps.

### 4.3 Contribution of typhoons to the vertical transport of ozone

During the period from August 21 to September 6, 2020, the consecutive occurrence of two northward-moving typhoons (Bavi and Maysak) triggered prolonged ozone pollution episodes in the Beijing-Tianjin-Hebei and Yangtze River Delta regions, with over 50% of monitoring stations exceeding ozone standards (Cong et al., 2024; Hu et al., 2024). Our study reveals that Guangdong Province similarly experienced extended ozone pollution episodes, particularly between August 28-30



578 and September 1-3, when more than 40 out of 105 monitoring stations (38.1%)  
579 recorded exceedances. The most severe pollution occurred on August 29, with 57  
580 stations (54.3%) exceeding standards and an average MDA8 ozone concentration of  
581 80.6 ppbv (**Fig.8a-b**). Backward trajectory analysis for August 29 identified a 7-day  
582 vertical transport pathway from upper levels to the boundary layer, suggesting  
583 potential downward mixing of high-ozone air masses (**Fig.8c**). This section examines  
584 the period from August 24 to August 31, 2020, employing the WRF-CMAQ model to  
585 simulate the spatial distribution of ozone. Integrated Process Rate (IPR) analysis is  
586 applied to investigate the formation mechanisms of surface ozone pollution in  
587 Guangdong Province under the influence of consecutive northward-moving typhoons,  
588 with a quantitative assessment of the impact of vertical transport on ozone  
589 concentrations within the planetary boundary layer.

590 The WRF-CMAQ model was used to simulate ozone variations in Guangdong  
591 Province from August 24 to 31, 2020, with evaluation results showing excellent  
592 performance (**Fig.8d**). For all 105 monitoring stations across the province, the  
593 correlation coefficient between observed and simulated ozone concentrations  
594 reached 0.88 ( $p < 0.01$ ), with a root mean square error (RMSE) of 8.41 ppbv. Focusing  
595 on the Sanshui station (112.8°E, 23.15°N), which exhibited both high ozone levels and  
596 a clear increasing trend, the correlation coefficient was 0.82 ( $p < 0.01$ ) with an RMSE of  
597 3.29 ppbv. These results demonstrate that the WRF-CMAQ model successfully  
598 captured the spatial distribution and temporal evolution of this ozone pollution event  
599 in Guangdong, with statistical metrics meeting operational air quality modeling  
600 standards. The model's strong performance, particularly in reproducing both regional  
601 patterns and local pollution trends, provides reliable support for subsequent analysis  
602 of ozone formation mechanisms under typhoon conditions.

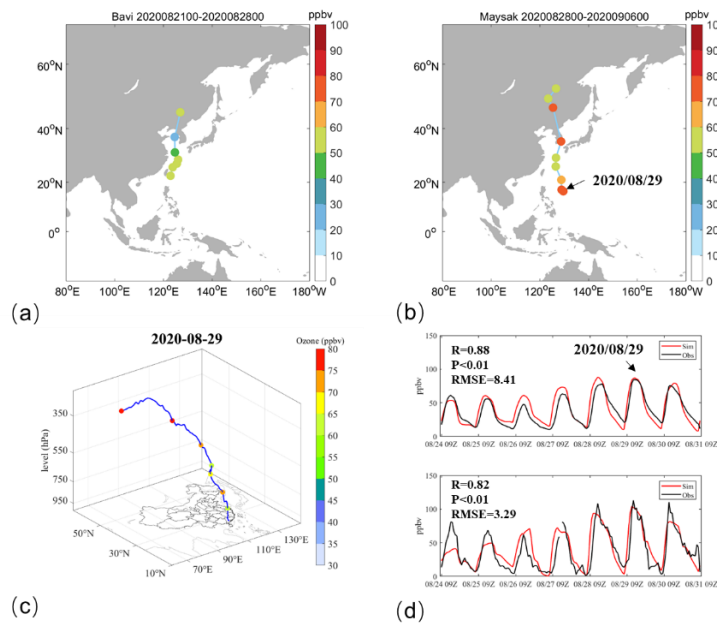


Figure 8. Consecutive northward-moving typhoon tracks, backward trajectories, and ozone variations. (a-b) Typhoon paths with corresponding MDA8 ozone concentrations in Guangdong province; (c) Backward trajectories at 1300 LST on August 29, 2020; (d) WRF-CMAQ simulated ozone variations (upper panel: average across 105 Guangdong monitoring stations; lower panel: Foshan Sanshui station (112.8°E, 23.15°N) observations, with red lines indicating simulated values and black lines representing monitored concentrations).

From August 24 to 27, Typhoon Bavi was located along the eastern coastal region of China, moving northward before gradually dissipating. From August 28 to 31, Typhoon Maysak emerged in the South China Sea and progressively approached the Chinese mainland. During this period, we analyzed variations in surface ozone concentrations and their vertical distribution under the influence of these consecutive northward-moving typhoons, based on model simulation results (Fig.9). The results show that the variation in surface ozone distribution can be divided into two stages: The first stage occurred under the influence of Typhoon Bavi, when surface ozone concentrations rapidly increased in the Beijing-Tianjin-Hebei region, Yangtze River Delta, and Pearl River Delta, and was rapidly transported to southwestern China by circulation. The second stage occurred under the influence of Typhoon Maysak, when

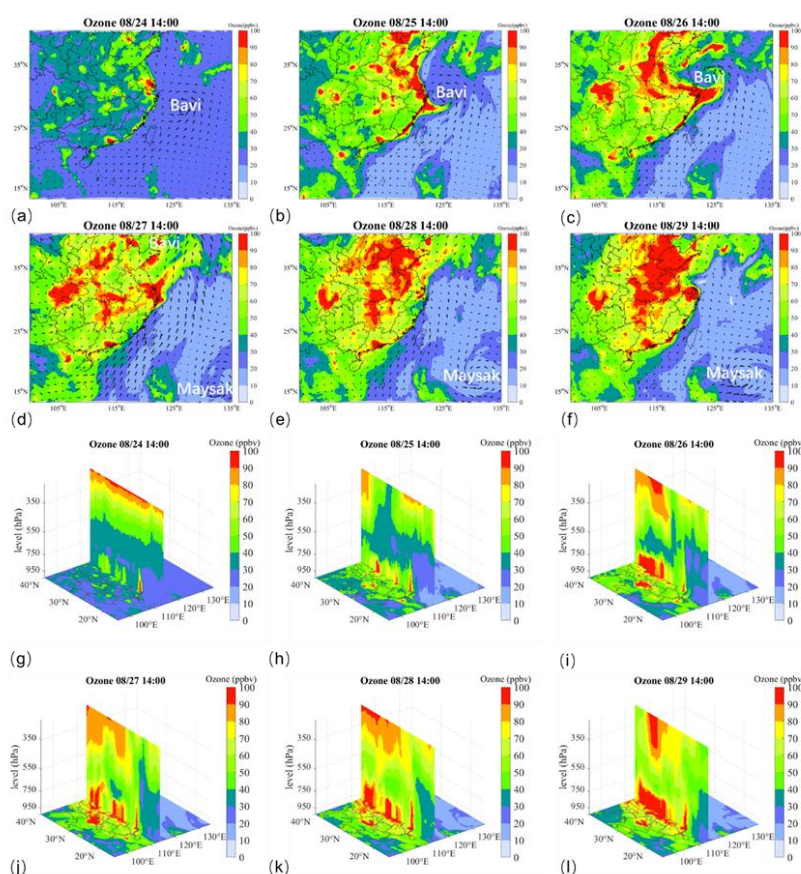


622 ozone concentrations continued to rise across most regions of China. Compared with  
623 the first stage, horizontal ozone transport was not significant during the second stage  
624 (**Fig. 9a-f**). In the vertical dimension, the consecutive northward-moving typhoons  
625 triggered a sustained downward transport process of ozone. Beginning on August 25,  
626 downward ozone transport was observed in the upper atmosphere between 35°N and  
627 40°N. From August 26 to 29, the zone of ozone subsidence gradually expanded  
628 southward, leading to a significant increase in ozone concentrations over Guangdong  
629 Province (**Fig. 9g-l**).

630 The IPR process analysis results elucidate the impacts of photochemical reactions  
631 and atmospheric transport on ozone concentration variations during this event (**Figs.**  
632 **S8-S9**). The photochemical reactions correspond to the CHEM contribution in the  
633 process analysis. The atmospheric transport represents the combined contributions of  
634 horizontal diffusion (HDIF), horizontal advection (HADV), vertical diffusion (VDIF), and  
635 vertical advection (ZADV) in the process analysis. The results indicate that the increase  
636 in surface ozone was primarily driven by photochemical reactions. During the period  
637 dominated by Typhoon Bavi (August 24-27), photochemical reactions intensified  
638 rapidly over Guangdong Province, contributing more than 30 ppbv to surface ozone  
639 concentrations in the central region (**Figs. S8a-d**). Under the influence of Typhoon  
640 Maysak (August 28-29), the positive contribution from photochemical reactions was  
641 slightly lower than in the previous phase, but still exceeded 16 ppbv in the central  
642 Guangdong region (**Figs. S8e-f**). The contribution of atmospheric transport varied  
643 significantly across different altitudes, exhibiting predominantly negative effects below  
644 850 hPa and positive effects above 850 hPa. Vertical cross-sections of daily mean  
645 atmospheric transport contributions reveal a gradual southward transport of ozone  
646 from higher to lower latitudes. However, its positive contribution to ozone  
647 concentrations was substantially lower than that of photochemical reactions, with  
648 daily mean contributions remaining below 4.5 ppbv (**Fig. S9**). The downward transport  
649 of upper-level ozone inhibited vertical diffusion of surface ozone while simultaneously



650 transporting high-concentration ozone downward into the boundary layer, further  
651 intensifying ozone pollution levels. In summary, during this ozone pollution event  
652 caused by consecutive northward-moving typhoons: Chemical processes were the  
653 main cause of surface ozone pollution in Guangdong Province, Atmospheric transport  
654 was a secondary contributing factor.



655  
656 Figure 9. Temporal evolution of (a–f) horizontal distributions of surface ozone and (g–l) vertical  
657 distributions (along 114°E cross-section) of ozone from 1400 LST 24 August to 1400 LST 29 August  
658 2020.

659 To quantitatively analyze the contribution of vertical transport to ozone  
660 concentrations within the boundary layer, we employed the IPR (Integrated Process  
661 Rate) analysis method to decompose ozone sources and sinks across the study area. A  
662 detailed analysis was conducted using results from the Sanshui station at 100m and



663 1300m altitudes (**Fig.10a-b**). Subsequently, we calculated the contribution rate of  
664 cross-boundary-layer vertical transport to ozone concentrations in the boundary layer  
665 at each time point (**Fig.10c**) using the following formula (Chen et al., 2022a):

666 
$$Transport\ flux = (IPR_{v,pbl} \times Z_{pbl}) \div (\sum_{j=1}^{pbl} O_{3,j} \times Z_j) * 100\%$$

667 where  $IPR_{v,pbl}$  indicates the IPR value corresponding to vertical transport  
668 (VDIF+ZADV) on the Boundary layer height. that is, the change in the values of  
669 pollutants caused by the vertical diffusion,  $Z_{pbl}$  represents the height of the layer in  
670 the model that is close to the height of the boundary layer.  $O_{3,j}$  indicates the ozone  
671 concentration in layer  $j$ ,  $Z_j$  represents the height of  $j$  layer.

672 Detailed analysis of process contributions at different heights within the  
673 boundary layer shows that while near-surface atmospheric transport exhibited  
674 negative contributions to daily mean ozone concentrations, the decomposition of  
675 individual processes at 100m height revealed positive contributions from vertical  
676 diffusion (VDIF) during 0900-1100 LST on 29 August, with magnitudes of 39.9 ppbv,  
677 26.4 ppbv, and 12.3 ppbv respectively (**Fig. 10a**). Further analysis of process  
678 contributions at 1300m height reveals distinct positive signals from vertical transport  
679 during the morning hours of both 28 and 29 August (**Fig. 10b**). This confirms that  
680 upper-level ozone can be transported into the boundary layer, thereby influencing  
681 ozone concentrations within the boundary layer. Calculation of cross-boundary-layer  
682 vertical transport contributions revealed six distinct ozone transport events during this  
683 consecutive northward-moving typhoon episode, occurring on 24, 25, 27, 28, 29, and  
684 30 August. The maximum contribution rate to ozone concentrations within the  
685 boundary layer reached 16%.

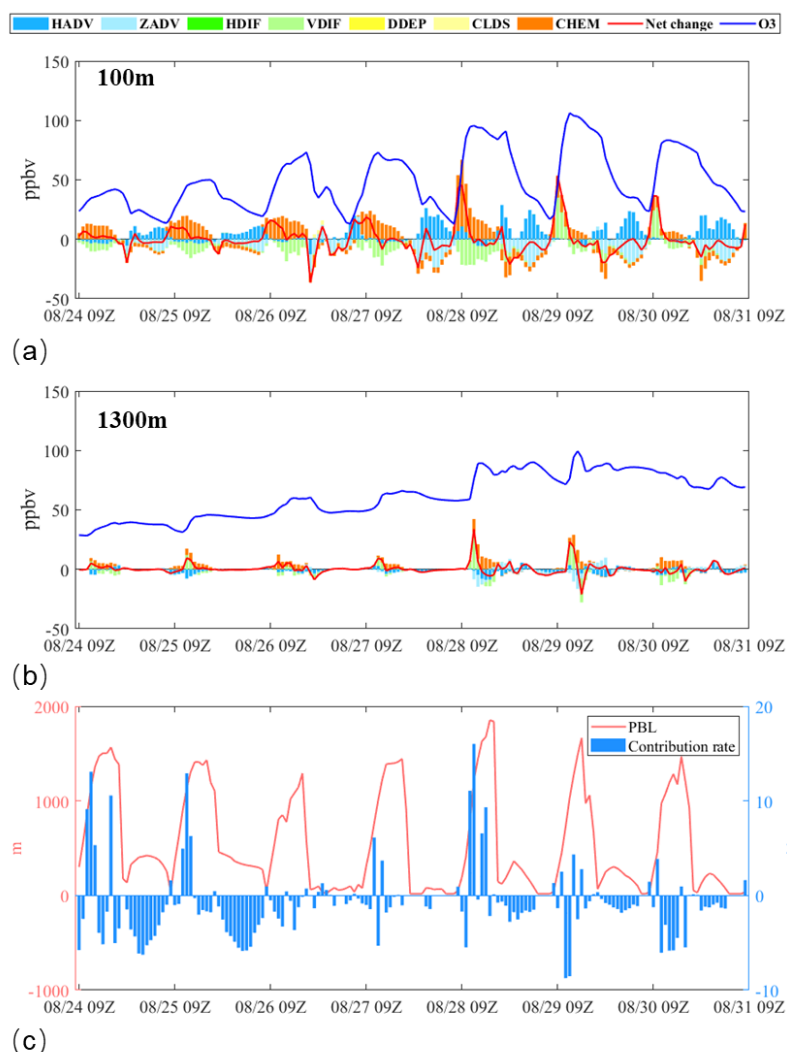


Figure 10. Process contributions to ozone concentrations at 100m and 1300m altitudes, and cross-boundary-layer vertical transport contribution rates. (a-b: Contributions from horizontal diffusion (HDIF), horizontal advection (HADV), vertical diffusion (VDIF), vertical advection (ZADV), chemical processes (CHEM), dry deposition (DDEP), and cloud processes (CLDS). c: Red lines indicate net ozone change, while blue lines show ozone concentration variations. )

## 5 Conclusions

This study systematically investigated the mechanisms by which different typhoon tracks influence ozone pollution in Guangdong Province through meteorological factors, atmospheric circulation patterns, transport trajectories, and





696 vertical transport contributions, based on 237 typhoons in China's adjacent waters  
697 from 2013-2023. The key findings are:

- 698 1. Historical typhoons were classified into three types using the K-MEANS  
699 clustering method: westward-moving typhoons (Type 1), distant northward-  
700 moving typhoons (Type 2), and proximal northward-recurving typhoons (Type  
701 3). Among these, near-track northward-moving typhoons are more likely to  
702 induce ozone pollution in Guangdong Province due to their more extreme  
703 meteorological conditions, including higher temperatures, stronger solar  
704 radiation, lower cloud cover, reduced precipitation, decreased relative  
705 humidity, elevated boundary layer height, weaker surface winds, and  
706 suppressed vertical motion.
- 707 2. Under the influence of northward-moving typhoons (type2 and type3), an  
708 upper-level anticyclonic center forms near the tropopause height in mid-  
709 latitudes, causing the westerly jet stream to shift southward. This process  
710 triggers the subsidence of high-concentration ozone from the upper  
711 troposphere, accompanied by pole-to-equator transport. Comparative  
712 analysis between typhoon and non-typhoon conditions reveals that both  
713 types of northward-moving typhoons induce significant ozone enhancement  
714 throughout the vertical column, with increases ranging from 2.5 to 11.6 ppbv  
715 (Type 2) and 0.3 to 12.3 ppbv (Type 3).
- 716 3. For Type 1 typhoons, the associated ozone pollution is primarily controlled by  
717 the radiative high-pressure system, with significant contributions from local  
718 pollution sources. In contrast, Type 2 and Type 3 typhoons exhibit the highest  
719 proportions of upper-level transport trajectories (23.9% and 15.2%,  
720 respectively), capable of delivering air masses from as high as 9,980 m (~250  
721 hPa) into the boundary layer. Coupled with the elevated ozone concentrations  
722 along these transport pathways, these mechanisms result in ozone  
723 enhancements of 10.7 ppbv and 12.3 ppbv at boundary layer altitudes for





724 Type 2 and Type 3, respectively.

725 4. Under the influence of two consecutive northward-moving typhoons from  
726 August 21 to September 6, 2020, Guangdong Province experienced a  
727 prolonged ozone pollution episode. On August 29, ozone exceedance was  
728 observed at 54.3% of monitoring stations. The primary cause of this ozone  
729 pollution event was enhanced photochemical production, with secondary  
730 contributions from upper-level ozone transport. Process analysis revealed  
731 that during 09:00-11:00 LST on August 29, the positive contributions of near-  
732 surface vertical transport to ozone concentrations were 39.9 ppbv, 26.4 ppbv,  
733 and 12.3 ppbv, respectively. During this typhoon event, cross-boundary-layer  
734 transport via vertical mixing contributed up to 16% of the ozone  
735 concentration within the boundary layer.

736

737

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