Sea Breeze-Driven Daytime Vertical Distributions of Air Pollutants and Photochemical Implications in an Island Environment

Bohai Li[1], Shanshan Wang[1,2], Zhiwen Jiang[1], Yuhao Yan[1], Sanbao Zhang[1], Ruibin Xue[1], Yuhan Shi[1], Chuanqi Gu[1], Jian Xu[3], Bin Zhou[1,2,4]

[1]Shanghai Key Laboratory of Atmospheric Particle Pollution and Prevention (LAP[3]), Department of Environmental Science and Engineering, Fudan University, Shanghai, 200438, China.
[2]Institute of Eco-Chongming (IEC), Shanghai, 2021516̶2̶, China.
[3]National Space Science Center, Chinese Academy of Sciences, 100190 Beijing.
[4]Institute of Atmospheric Sciences, Fudan University, Shanghai, 200438, China.

Correspondence to: Shanshan Wang (shanshanwang@fudan.edu.cn)

Abstract. Atmospheric pollutants in island and coastal environments are profoundly modulated by sea breezes (SB) and typhoons. Understanding the characteristics of pollutants and their photochemical indicators under different airflow regimes is crucial for effective pollution control. Utilizing Multi-Axis Differential Optical Absorption Spectroscopy (MAX-DOAS) and a sea-land breeze objective identification algorithm, we reveal distinct spatiotemporal patterns in NO_2, HCHO, CHOCHO, and associated photochemistry under varying airflow patterns in a rural coastal area of Hainan, China during the summer of 2024. Non-sea breeze days (NSBDs) showed higher pollutant levels with broader vertical distribution range under conducive meteorological conditions. Conversely, on sea breeze days (SBDs), SB limits pollutant dispersion and its cooling effect suppresses ozone formation. Furthermore, SB also enhances transport of NO_2 and biogenic volatile organic compounds (BVOC) below 300 m, influencing ozone formation sensitivity (OFS) throughout SB phases. Typhoons effectively scavenge pollutants via strong winds and precipitation but also facilitate mid-to-high altitude BVOC transport and vertical dispersion of surface pollutants. Photochemical indicator analysis (HCHO/NO_2 [FNR] and CHOCHO/NO_2 [GNR]) indicates that the VOC-limited regimes persist even at high altitudes during typhoons. Elevated FNR and GNR thresholds suggest that existing OFS classification thresholds are inadequate for low-NO_2 tropical coastal rural areas, underscoring the need for region-specific assessments. Given the BVOC-dominated environment and additional inputs from SB and typhoons, GNR proves more reliable than FNR for OFS determination. This study emphasizes the necessity of integrating local meteorology and

environment conditions in O_3 control strategies, providing a scientific basis for pollution mitigation in tropical coastal regions prone to SB and typhoons.

1 Introduction

Vertical gradients in atmospheric temperature, humidity, and solar radiation, among others, establish stratified photochemical environments, while atmospheric dynamics drive spatiotemporal heterogeneity in pollutant distribution, collectively rendering tropospheric chemistry evolution exceptionally complex (Cooke et al., 2010; Gordon, 1987; Jiang et al., 2008; Mason, 2013; Wang et al., 2024). As typical photochemical precursors, nitrogen dioxide (NO_2), formaldehyde (HCHO), and glyoxal (CHOCHO) exhibit unique vertical distributions in the troposphere. Vertical profiles of these species have been reported using gradient observations from various measurement techniques such as tethered balloons, unmanned aerial vehicles, towers, remote sensing techniques and aircraft (Benish et al., 2020; Chang et al., 2025; Glaser et al., 2003; Song et al., 2024; Vo et al., 2018; Z. Wang et al., 2025). Among these, Multi-Axis Differential Optical Absorption Spectroscopy (MAX-DOAS) is a cost-effective, automated remote sensing technique capable of observing vertical changes in atmospheric trace gas concentrations over long periods of time. This method has been extensively applied worldwide for atmospheric composition monitoring (Chang et al., 2025; Franco et al., 2015; Jiang et al., 2025; Wagner et al., 2011; Z. Wang et al., 2025).

Previous studies reported that well-established retrieval frameworks for NO_2 and HCHO have shown good agreement with satellite and in situ measurements, demonstrating its robustness for long-term and multi-environment observations (Chan et al., 2020; Jiang et al., 2025; Kumar et al., 2020; Martin et al., 2004; Schreier et al., 2020; Schreier et al., 2016; Wagner et al., 2011; Zhang et al., 2022). CHOCHO is challenging to detect due to its weak absorption and spectral interferences, yet MAX-DOAS remains one of the few ground-based techniques capable of reliable retrievals. Successful measurements in both urban and biogenic regions have advanced understanding of CHOCHO photochemistry and have been used to validate satellite products (Hong et al., 2024; Hoque et al., 2018; Lerot et al., 2021; Liu et al., 2021; Schreier et al., 2020; Zhang et al., 2025). Overall, MAX-DOAS demonstrates robust and reliable performance in monitoring NO_2, HCHO, and CHOCHO, providing strong methodological support for the data quality in this study.

NO_2 concentrations typically decrease exponentially with altitude due to the fact that it originates mainly from surface-level

50 anthropogenic emissions, whereas photochemical oxidation of long-lived volatile organic compounds (VOCs) at high altitudes

sometimes leads to higher distribution heights for HCHO and CHOCHO (Hong et al., 2022; Kumar et al., 2020; Liu et al.,

2023). The ratio of CHOCHO to HCHO (R_{GF}) serves as a robust indicator for distinguishing VOCs sources due to their

different sources and similar sinks (Chan Miller et al., 2014; Vrekoussis et al., 2010; Xing et al., 2020; Zhang et al., 2021).

Observations reveal altitude-dependent characteristics of R_{GF}, with lower values at low altitudes reflecting dominant biogenic

55 VOCs (BVOC) contributions, while elevated R_{GF} at higher altitudes highlights anthropogenic VOCs (AVOC) (Hong et al.,

2022; Xing et al., 2020). The HCHO to NO_2 ratio (FNR) is widely used to assess ozone (O_3) formation sensitivity (OFS) by

classifying regimes into VOC-limited, transition, and NO_x-limited ~~regimes~~ (Jiang et al., 2025; Liu et al., 2021; Z. Wang et al.,

2025). However, recent studies propose the CHOCHO as a more reliable indicator because it is more directly linked to

secondary formation through the VOCs oxidation than HCHO (Hong et al., 2024; Liu et al., 2021). Therefore, comprehensive

60 systematic investigation of vertical distribution patterns of these species is critical for refining OFS assessments across

atmospheric layers, optimizing chemical transport models, and informing targeted air quality management strategies.

Coastal atmospheric environments are significantly influenced by sea-land breeze (SLB) circulation and also typhoons in some

regions (e.g., the western North Pacific). SLB, a mesoscale phenomenon driven by thermal contrast, exhibits a diurnal

alternation of onshore sea breezes (SB) and offshore land breezes (LB), recirculating pollutants between marine and terrestrial

65 zones (Gangoiti et al., 2002; Martins et al., 2012; Zhao et al., 2022). SB enhances pollutant accumulation through reduced

boundary layer height (BLH) and suppressed vertical dispersion (Puygrenier et al., 2005), while its moisture influx alters

temperature fields and photochemical conditions, complicating secondary pollutant (e.g., ~~ozone (O_3)~~, secondary aerosols)

~~distribution~~formation (Gangoiti et al., 2002; Xu et al., 2021). Typhoons are large-scale weather systems that can cleanse

pollutants via strong winds and precipitation (Roux et al., 2020; Yang et al., 2012), yet their peripheral subsidence zones with

70 reduced cloud cover intensify solar radiation, triggering severe surface O_3~~ozone~~ pollution (Deng et al., 2019; Wu et al., 2013).

Additionally, typhoon-enhanced winds promote cross-regional O_3 transport (Qu et al., 2021; Wang et al., 2024). Both air

currents critically modulate pollutant vertical structures and photochemical environments across atmospheric layers.

Nevertheless, research gaps persist regarding vertical distributions of NO_2, HCHO, and CHOCHO under SB and typhoon

conditions, particularly in island where complex topography and unique atmospheric environments may cause these pollutants

75   to behave differently than over continental or coastal regions.

Hainan Island is ~~located~~away far from mainland China and lies approximately 20 km from the nearest point in Guangdong Province. The island spans approximately 32,900 km^2 and supports a population of 10.48 million people concentrated in coastal cities (e.g., Haikou, Sanya, and Danzhou), while also maintaining dense vegetation with over 60% forest coverage featuring unique tropical ecosystems (Hainan Provincial Bureau of Statistics, 2024).~~and~~ ~~A~~adjacent to the South China Sea, ~~where~~ its

80   favorable topography and geographic position render the island ~~it~~ frequently influenced by SB and typhoons (Fu et al., 2023; Liang and Wang, 2017; Zhang et al., 2014). Given the ~~superior~~ good air quality—mean NO$_2$ (7.34 μg m^{-3}) and O$_3$ (52.80 μg m^{-3}) concentrations below the National Ambient Air Quality Standard (NAAQS) Grade II and urban VOCs (13.1 ppbv) remaining significantly lower than megacities like Beijing and Shanghai—it has earned designation as a National Ecological Civilization Pilot Zone of China (Xu et al., 2024). Notably, several studies have shown that typhoons and the East Asian

85   monsoon can transport pollutants such as O$_3$ and its precursors to the island via regional-scale advection, thereby inducing pollution events (Fu et al., 2023; Xu et al., 2024; Zhan et al., 2023). However, the specific effects of SB and typhoons on O$_3$ precursors over Hainan Island are ~~still unclear~~not investigated yet.

By combining ground-based MAX-DOAS measurements with meteorological reanalysis data, we investigated the spatiotemporal distributions of NO$_2$, HCHO, CHOCHO, and their photochemical indicators under various air current patterns

90   (ACPs) in the coastal area of Hainan. This study aims to elucidate the transport mechanisms of pollutants in complex atmospheric environments and their response characteristics to local photochemical processes, thereby supporting air pollution control and environmental management in coastal regions.

2 Data and Methods

2.1 Measurement of trace gases profile

95   This study was conducted in Hainan National Field Science Observation and Research Observatory for Space Weather (19.53°N, 109.14°E) at Fuke village (FK), Danzhou City, Hainan Province, during the summer of 2024 (June 1 to August 31). The observation site (~~A~~altitude approx. 100 m above sea level) lies within a topographically complex region bordered to the

northwest by the ~~Beibu~~ Gulf of Tonkin and to the southeast by the Wuzhi Mountain ~~massif (18.9°N, 109.7°E)~~ (Fig. 1a). During the observation period, the observation site was directly affected due to the passage of Typhoon "Prapiroon" from south-east to north-west across Hainan Island. The national expressway G98 passes about 1,600 m north-west of FK (Fig. 1b). The MAX-DOAS instrument comprises a scanning telescope, a stepper motor controlling the zenith observation angle, an Ocean Optics QE65 Pro spectrometer, and a data processing computer (Fig. 1c). Data acquisition, with a temporal resolution of 30 seconds for individual spectrum, is conducted daily from 06:00 to 19:00 local time (LT), with dark current signals automatically extracted from nightly background measurements. A full scan sequence of different elevation angles takes approximately 5 minutes to complete.

Spectral data were analyzed using the QDOAS software (http://uv-vis.aeronomie.be/software/QDOAS/), which retrieves differential slant column densities (DSCD) of trace gases by fitting their characteristic absorption bands with reference cross-sections and background corrections. Both ultraviolet (UV) and visible (VIS) channels were employed for oxygen dimer (O_4), NO_2, HCHO and CHOCHO DSCD retrievals, with configurations detailed in Table S1. Zenith spectra during each scan cycle served as Fraunhofer reference spectra (FRS) to roughly eliminate stratospheric contributions. Measured spectra underwent minor shift and stretch corrections to mitigate instrument instability effects. Quality control excluded retrievals with Root Mean Square (RMS) > 0.001 ~~or~~ and DSCD error to DSCD ratios exceeding 10% (30% for HCHO, 50% for CHOCHO). Observations before 07:00 and after 18:00 were excluded due to insufficient light intensity. Detection limits (DDL) for different elevation angles were calculated following S1 (Supplement~~s~~). NO_2 data all exceeded DDL, while HCHO and CHOCHO exhibited significant proportions of sub-DDL values (particularly CHOCHO at high elevations), attributable to their low concentration levels. To preserve temporal continuity, sub-DDL results were retained rather than discarded in subsequent analyses. Typical DOAS fitting results for four gases are illustrated in Fig. S1.

This study employed the HEIPRO algorithm (Heidelberg Profile, developed by IUP Heidelberg) based on optimal estimation methodology (OEM) to retrieve vertical profiles of tropospheric trace gases (Frieß et al., 2006; 2011). The algorithm incorporates the SCIATRAN radiative transfer model (RTM) as an orthorectification mode (Rozanov et al., 2002). The retrieval strategy comprises two sequential steps: (1) aerosol extinction vertical profiles are derived from measured O_4 DSCD, followed by (2) trace gas profile retrievals using the aerosol profiles as inputs in RTM. Vertical grids for trace gases span 0.0–

3.0 km altitude with 100 m resolution, and retrievals were performed at 10-minute intervals. Exponential decay functions served as a priori profiles for all species with error S_a of 100% and 500 m covariance correlation length. Fixed aerosol parameters included: 0.3 km^{-1} extinction coefficient, 0.1 surface albedo, 0.95 single scattering albedo, and 0.72 asymmetry parameter at 360 nm (Henyey and Greenstein, 1941; Zhang et al., 2021). Surface concentrations of NO_2, HCHO, and CHOCHO were initialized at 2, 5, and 0.2 ppbv, respectively. Representative retrieved profiles are illustrated in Fig. S2. To ensure that further analyses were based on reliable data, the absolute deviation thresholds between measured and modelled DSCD (optical densities, OD) values were set to 1×10^{-3}, 2×10^{15}, 7×10^{15} and 4×10^{14} molec cm^{-2} for O_4 OD, NO_2, HCHO and CHOCHO DSCD, respectively, in addition to retaining only those results with a degree of freedom of the signal (DoF) > 1.0. Filtered data showed strong model-measurement agreement ($R^2 > 0.95$), confirming retrieval accuracy (Fig. S3).

In addition, hourly surface O_3 mass concentrations (μg m^{-3}) at the reference state (298 K, 1,013 hPa) were recorded at the DongPo School monitoring station (DPS, 19.52°N, 109.56°E) using a Thermo Scientific 49i O_3 analyzer were converted to surface VMRs (ppbv) for the analysis of photochemical indictors.

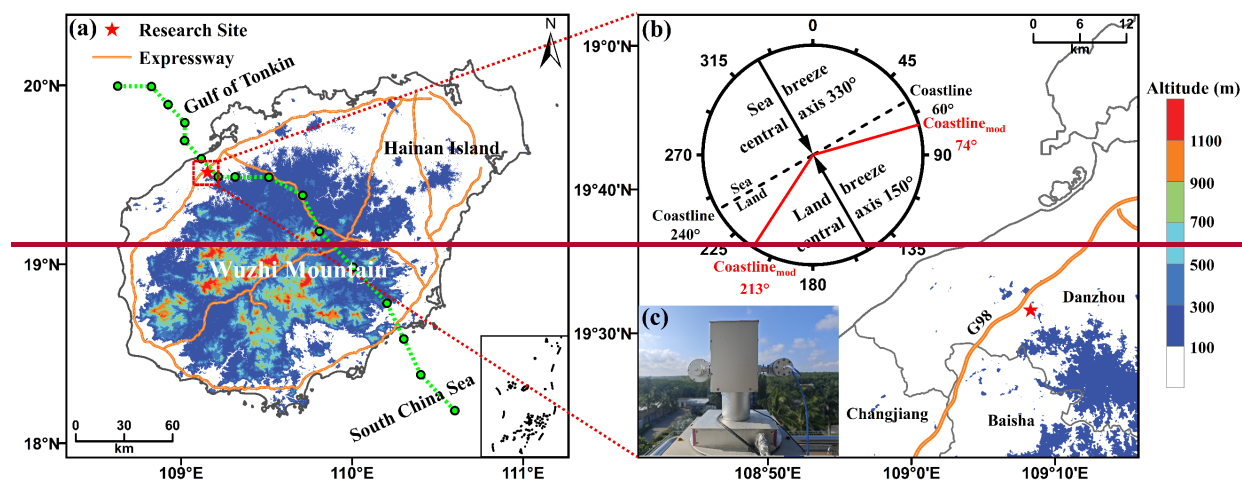
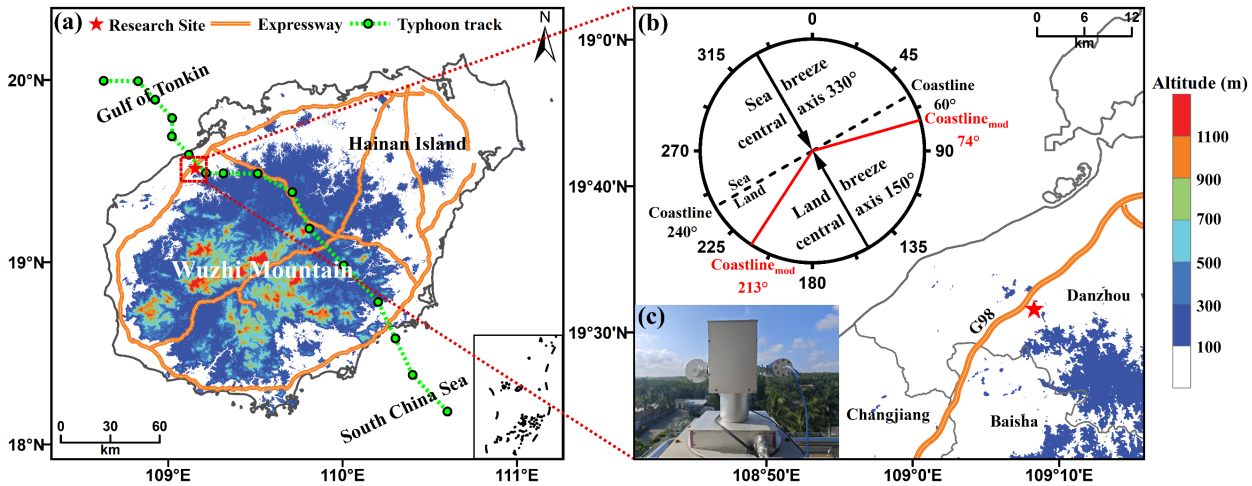



Figure 1. Overview of the study area and MAX-DOAS. (a) Topographic map of the area around the geopotential height and Multi-Axis Differential Optical Absorption Spectroscopy (MAX-DOAS) measurement station (19.53°N, 109.14°E) in Hainan Province, marked by a red five-pointed star, and the highway network is indicated by a~~n~~ ~~yellow~~ orange line. The track of the typhoon "Prapiroon" is shown by the green dotted line; (b) Topographic map of the area around the observation site, where a highway (G98) passes about 1,600 m to the northwest of the observation site. Compass charts for recording the direction of the sea breezes (SB) and land breezes (LB) are given in the upper left. The central axes of the SB and LB directions are shown by black arrows, and the red solid lines represent the modified coastline (Coastline_mod); and (c) MAX-DOAS instrument.

2.2 Identification of sea and non-sea breeze days

In this study, a filtering method was adopted as the primary means of identifying days favorable to SB development. Although this approach may underestimate SB frequency, the results will not be ambiguous. Accurate identification of onshore and offshore wind direction (WD) is essential for the method's effectiveness, with the geometry of the coastline serving as the basis for determining the central axes of SB and LB (Azorin-Molina et al., 2009; Azorin-Molina et al., 2011; KiranKumar et al., 2019). ~~Accurate determination of onshore and offshore WD is critical to the identification process, and the orientation of the local coastline serves as the basis for defining the principal axes of SB and LB.~~ In the vicinity of FK, the coastline at larger scales is approximately linear, extending from 240° to 60°, with the central axes of the SB and LB oriented at 321° and 141°, respectively (Fig. 1a, and b). However, previous studies have shown that local coastline geometry and topographic influences, in combination with coastal-parallel flows driven by the Coriolis force, can induce deviations in the SB direction (Azorin-Molina et al., 2011; Miller et al., 2003). Based on these local features, the directional range of the SB at FK was revised to

span from 213° to 74° (The red solid lines on the compass map in Figure 1b). In addition, islands and peninsulas can trigger SB fronts along different segments of the coastline (Zhu et al., 2017), whereas mountainous terrain often inhibits inland propagation (Barthlott and Kirshbaum, 2013). Specifically, on Hainan Island, the central Wuzhishan mMountain range acts as a barrier to the southeastward SB front, limiting its reach into the FK. Notably, during summer, the island is predominantly influenced by the southwesterly monsoon, which frequently produces daytime background winds oriented landward or along the coast (Fig. S4). The formation pattern of these background air currents is fundamentally different from that of the SB. To enhance the contrast with sea breeze days (SBDs), dates influenced by background winds or failing to meet the SB criteria were identified. Those with anomalously elevated daytime relative humidity were excluded to minimize the influence of marine air masses. The remaining dates were classified as non-sea breeze days (NSBDs).

Figure S42 illustrates the flow of an objective recognition algorithm (ORA) for SLB containing four modules and 12 filters. Unless otherwise noted, the data used by ORA are near-surface data. Module 1 filters the large-scale circulation background unfavorable to the formation of SLB, one of the important factors influencing the formation of SLB, through the wind speed (WS) and direction (WD) at high altitude (filters 1 to 2) (Arrillaga et al., 2020). Module 2 is the core module, which contains three channels: channels 1 (filter 3 to 7-1) and 2 (filter 3 to 7-2) are based on the sudden change of WS and WD to identify the SB, and the rate of change of WS is introduced to capture the dynamics of the local wind field (filter 5-2), which can make up for the deficiencies of the traditional WS thresholding method; channel 3 rejects dates potentially affected by oceanic currents by setting a relative humidity (RH) increase threshold (filters 8-9). Module 3 excludes large-scale airflow disturbances through barometric pressure conditions (filters 10-11). Module 4 ensures that the remaining days satisfy the minimum sea-land temperature gradient (filter 12) as a prerequisite for SB establishment. The specific functions of the modules and filters are detailed in S2 (Supplements). and identification results are summarized in Table S3 and Section 3.2.

175

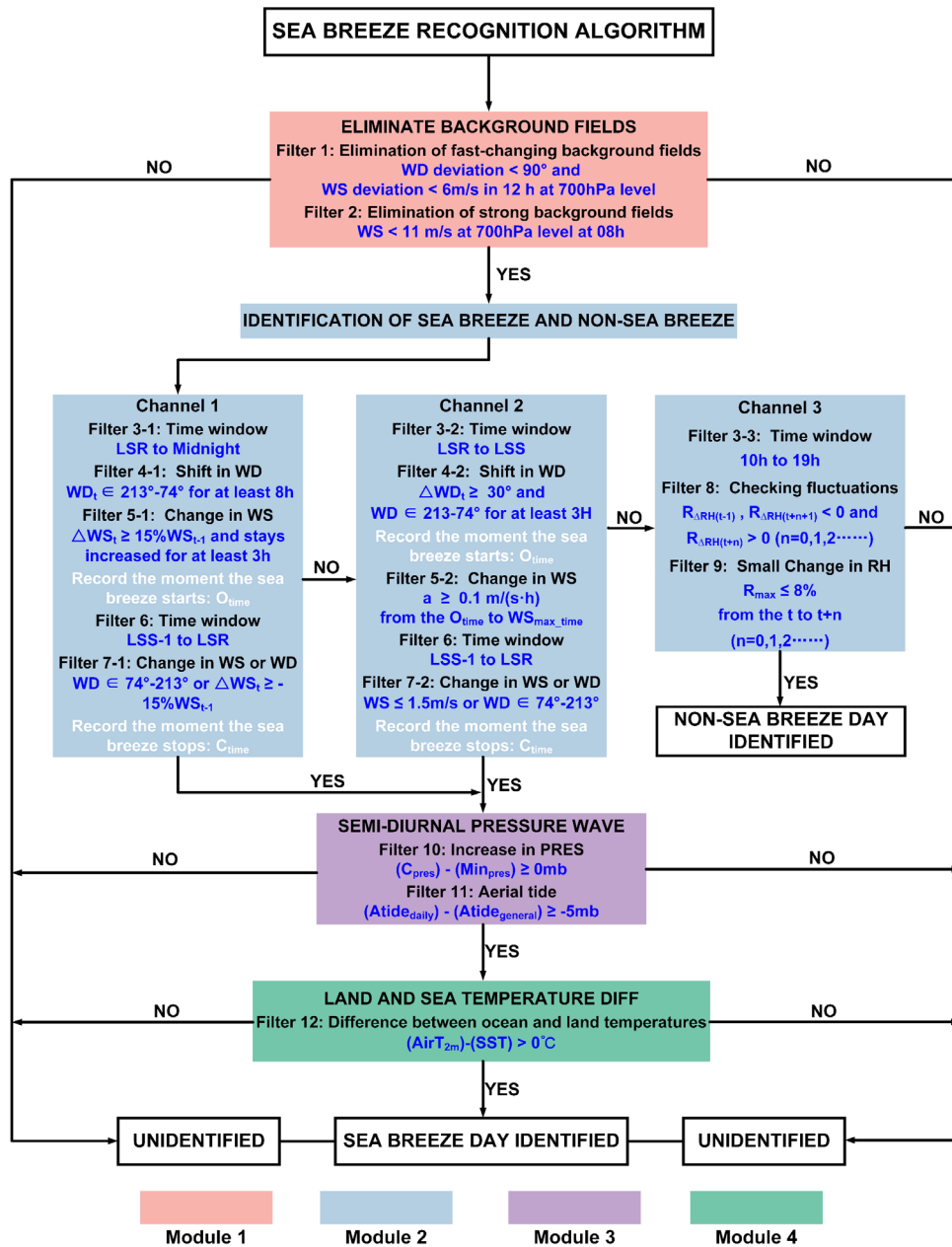



Figure 2. Flowchart of the objective recognition algorithm (ORA) for SBDs. The pink squares represent eliminate background fields, referred to as Module 1; the blue squares represent identification of sea breeze and non-sea breeze, referred to as Module 2; the purple squares represent semi-diurnal pressure wave, referred to as Module 3; the green squares represent land and sea temperature diff, or Module 4 for short.

2.3 Meteorological reanalysis and ancillary data

Meteorological data were obtained from the European Centre for Medium-Range Weather Forecasts (ECMWF) fifth-generation atmospheric reanalysis product (ERA5), incorporating three principal components: the ERA5 single-level dataset, ERA5 pressure-level data, and ERA5-Land. The single-level dataset provides hourly surface and sea-level parameters at 0.25° spatial resolution, while the pressure-level product delivers three-dimensional atmospheric profiles across 37 vertical levels from 1,000 hPa to 1 hPa, maintaining equivalent spatial and temporal resolutions. The ERA5-Land (0.1×0.1°) component—released in 2019—offers enhanced terrestrial monitoring through refined 9 km spatial resolution while preserving hourly temporal sampling. Specific meteorological variables employed in this analysis, along with their respective dataset sources are comprehensively detailed in Table S4. Notably, ERA5 lacks the surface RH data required; therefore, we computed actual and saturated water vapor pressures using the 2m dew temperature (T_d in K) and 2m temperature (T in K) from ERA5-Land, respectively, and derived RH from their ratio (Zhu et al., 2025):

$$RH = \frac{e_a(T_d)}{e_s(T)} \times 100\% \tag{1}$$

where $e_a(T_d)$ is the actual vapor pressure at dew point temperature, and $e_s(T)$ is the saturation vapor pressure at the current temperature. These can be determined by the following formula:

$$e_a(T_d) = 0.6108 \times exp\left(\frac{17.27 \times (T_d - 273.15)}{T_d - 35.86}\right) \tag{2}$$

$$e_s(T) = 0.6108 \times exp\left(\frac{17.27(T - 273.15)}{T - 35.86}\right) \tag{3}$$

Monthly emission fields of isoprene (0.25°×0.25°), anthropogenic NO_x, and VOCs (0.1°×0.1°) were also obtained from the CAMS (Copernicus Atmosphere Monitoring Service) global emission inventories (https://ads.atmosphere.copernicus.eu/datasets/cams-global-emission-inventories?tab=overview).

3 Results

3.1 Time Series of the Measurement Results

Figure 3 presents the time series of daily averaged near-surface aerosol extinction coefficient (AEC), NO_2, HCHO, CHOCHO volume mixing ratios (VMRs), corresponding column integrated aerosol optical depth (AOD) and vertical column densities (VCDs) observed by MAX-DOAS at FK, Hainan Island, China during the 2024 summer season (1 June–31 August). The

measurements reveal distinct atmospheric characteristics shaped by regional environmental conditions. The observed s~~S~~ummer aerosol loading (mean AEC and AOD measuring 0.11 ± 0.03 km^{-1} and 0.29 ± 0.1, respectively) exhibited markedly lower values compared to urban agglomerations like Beijing and Shanghai (where AOD is around 0.4~~typically exceeds 0.5~~) (Fan et al., 2025; Peng et al., 2025), ~~with mean AEC and AOD measuring 0.11 ± 0.03 km^{-1} and 0.29 ± 0.1, respectively~~and slightly higher than in coastal Thailand, where monsoonal rainfall efficiently scavenges aerosols ~~.~~ (Peengam et al., 2025). The average NO$_2$ surface VMRs (1.61 ± 0.53 ppbv) was about 3 to 5 times lower than values recorded in major Chinese megacity centers, such as the Beijing–Tianjin–Hebei region (7.62 ± 1.39 ppbv) and the Yangtze River Delta (7.45 ± 0.87 ppbv) (Lou et al., 2025; Ministry of Ecology and Environment of the People's Republic of China, 2024a, b, c), reflecting minimal local traffic and industrial emissions. Notably, HCHO and CHOCHO concentrations (4.33 ± 1.07 ppbv and 0.10 ± 0.02 ppbv, respectively) ~~reached levels~~ are comparable to values reported for summertime metropolitan regions, such as Beijing (HCHO~~ surface VMRs~~: 4.41 ppbv) and Guangzhou (CHOCHO~~ surface VMRs~~: 0.13 ppbv) (Hong et al., 2024). This likely stems from biogenic emissions from surrounding tropical ecosystems compensating for diminished anthropogenic sources. However, the observed HCHO and CHOCHO substantially higher than values reported in less anthropogenically influenced islands and coastal regions, including the coastal area of Mt. Lao, Qingdao, the Galápagos Islands, and Cape Verde (Lawson et al., 2015; Mahajan et al., 2014; Mahajan et al., 2010; Zhao et al., 2024).

~~In refined~~ At an hourly temporal resolution (Fig. S6)~~.~~, HCHO and CHOCHO exhibited significant positive correlations with elevated temperatures (> 30°C) and reduced humidity (< 60%), particularly during two enhanced photochemical periods: 12–22 June and 1–23 August, when the elevated concentrations persisted from the surface through elevated atmospheric layers. Surface VMRs and VCDs show relatively similar patterns in their time series for all species because of their predominantly low-level distribution. However, during specific episodes (e.g., August 7–10, 2024), pronounced fluctuations in vertical temperature and humidity profiles resulted in a clear decoupling between surface and column measurements for NO$_2$, HCHO, and CHOCHO, which was more pronounced in aerosols~~.~~ as indicated by the weakest correlation (R^2=0.34) between surface AEC and AOD (Fig. S7). ~~This~~ These results suggest~~s~~ that the vertical distribution pattern of pollutants may involve more complex physicochemical processes that deserve further investigation.

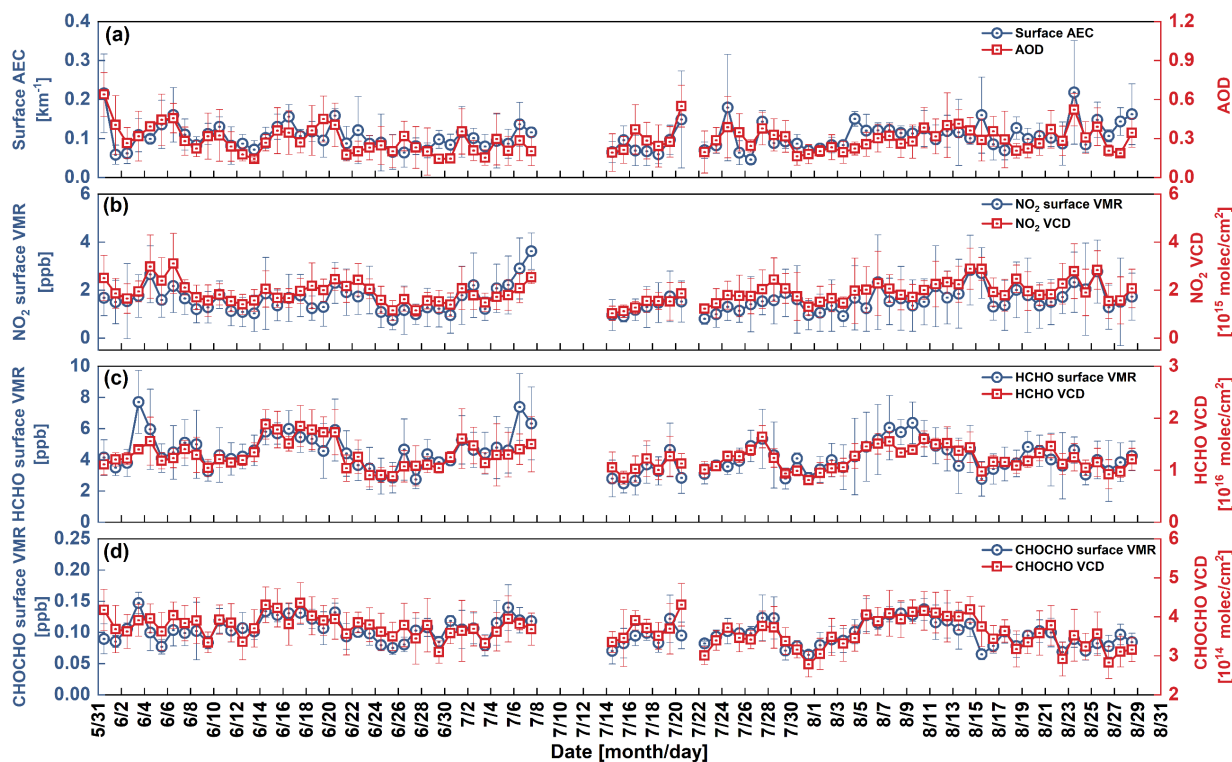



Figure 2̶3̲. Diurnal variation of pollutants at FK. Time series of daily (a) aerosol, (b) NO₂, (c) HCHO and (d) CHOCHO surface mixing ratios and vertical column densities from MAX-DOAS measurements in Hainan Island, China.

Figure 3̶4̲ illustrates the average diurnal variation and daily profiles of aerosol, NO₂, HCHO, CHOCHO, temperature and RH during the campaign. Aerosols, HCHO, and CHOCHO exhibited vertical distribution patterns that closely tracked the diurnal variation of BLH. Their vertical distributions peaked synchronously with maximum BLH at noon, followed by gradual subsidence to lower altitudes during afternoon planetary boundary layer (PBL) contraction, though residual amounts persisted in upper layers. The peak of the AEC occurs at the top of the boundary layer almost throughout the observing period due to the aerosols from different sources consistently suppressing the daytime PBL (Xin et al., 2023). Notably, the aerosol vertical distribution bears a remarkable similarity to the RH̲,̶ ̶w̶i̶t̶h̶ ̶a̶ ̶c̶o̶r̶r̶e̶s̶p̶o̶n̶d̶i̶n̶g̶ ̶h̶u̶m̶i̶d̶i̶t̶y̶ ̶m̶a̶x̶i̶m̶u̶m̶ ̶o̶b̶s̶e̶r̶v̶e̶d̶ ̶s̶i̶m̶u̶l̶t̶a̶n̶e̶o̶u̶s̶l̶y̶ ̶a̶t̶ ̶t̶h̶e̶ ̶t̶o̶p̶ ̶o̶f̶ ̶t̶h̶e̶ ̶P̶B̶L̶ ̶t̶h̶r̶o̶u̶g̶h̶o̶u̶t̶ ̶t̶h̶e̶ ̶d̶i̶u̶r̶n̶a̶l̶ ̶c̶y̶c̶l̶e̲ The Pearson correlation between AEC and RH remains positive above 0.5 km, reaching a maximum of 0.86 around 1 km, indicating a strong positive relationship at this altitude (Fig. S8)̲. This phenomenon reflects aerosol hygroscopic growth characteristics, particularly the water vapor sensitivity of fine particulate matter, which

enhances light extinction at these altitudes (Liu et al., 2020). The occasional observed deviation between surface AEC and total AOD variations (Fig. 3~~2~~a) may also be explained by this altitude-dependent hygroscopic behavior.

The diurnal NO_2 cycle exhibited a characteristic bimodal pattern, with primary peaks in morning (08:00 LT, VMRs = 2.67 ppbv) and evening (17:00 LT, VMRs = 1.66 ppbv) aligned with traffic rush hours, reflecting dominant vehicular emission sources. Notably, a secondary minor peak emerged at 14:00 LT, potentially driven by regional transport mediated by local circulation patterns (e.g., SB systems) (Fig. 6~~7~~b). Vertically, NO_2 predominantly resides below 300 m. The high~~early peaks of~~ HCHO and CHOCHO morning levels are attributed to the oxidation of residual VOCs by OH radicals ~~in the previous night~~ during daytime and the rerelease from the aerosols (Liu and Wang, 2024; Xing et al., 2020). In contrast, the delayed HCHO peak (10:00 LT, VMRs = 5.14 ppbv) can be attributed to the photochemical oxidation of early morning traffic emitted VOCs under increasing solar radiation.~~suggests the existence of additional emission sources beyond this mechanism, which are likely linked to morning traffic emissions.~~ CHOCHO surface VMRs ~~(0.12 ppbv)~~ exhibited a distinct peak (0.12 ppbv) between 12:00–13:00 LT, potentially linked to photooxidation of VOCs—particularly alkenes and aromatics—and elevated temperatures that intensified photochemical activity and biogenic isoprene emissions (Duncan et al., 2010; Hendrick et al., 2014). However, ~~A~~ a slow decreasing trend regarding the HCHO surface VMRs was still detected in the concurrent period, most likely due to the stronger vertical convection after midday, transferring the secondary HCHO production to higher altitudes (< 800 m) and thus diluting the surface concentration. In contrast, CHOCHO ~~has a shorter atmospheric residence time (~1.3 h) and thus~~ remains confined to lower altitudes (< 500 m) (Fig. 4d, left), mainly because its weaker secondary formation at higher altitudes compared to HCHO and stronger uptake by humid aerosols that depletes its abundance (Fortems-Cheiney et al., 2012; Liao et al., 2025; Sander, 2015; Waxman et al., 2015; Wolfe et al., 2019). The late afternoon rebound in HCHO and CHOCHO concentrations may be due to the combing effects of increased anthropogenic emissions of traffic and cooking activities, as well as reduced atmospheric mixing ~~and lower photolysis rates as~~with temperatures drop and solar radiation decreases (Fig. S9).

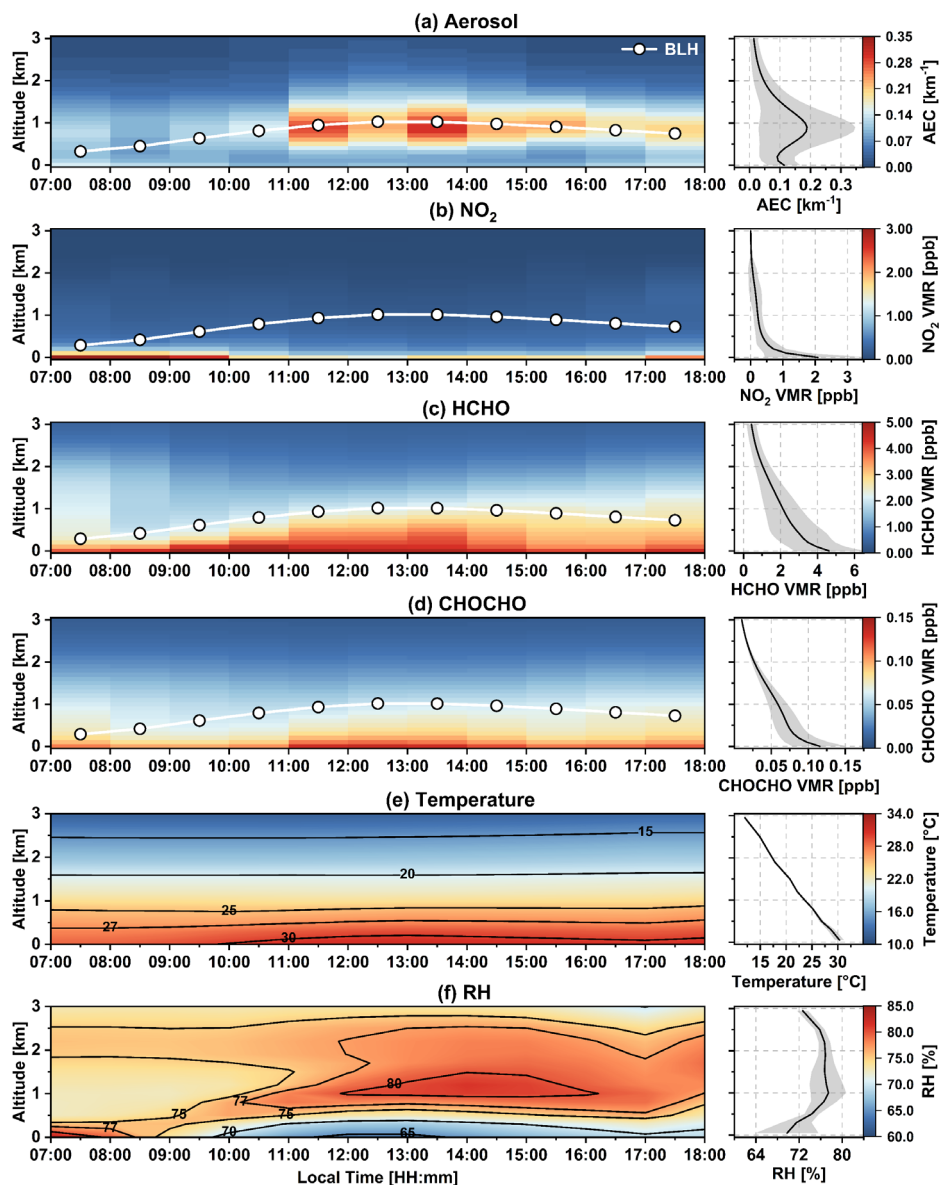



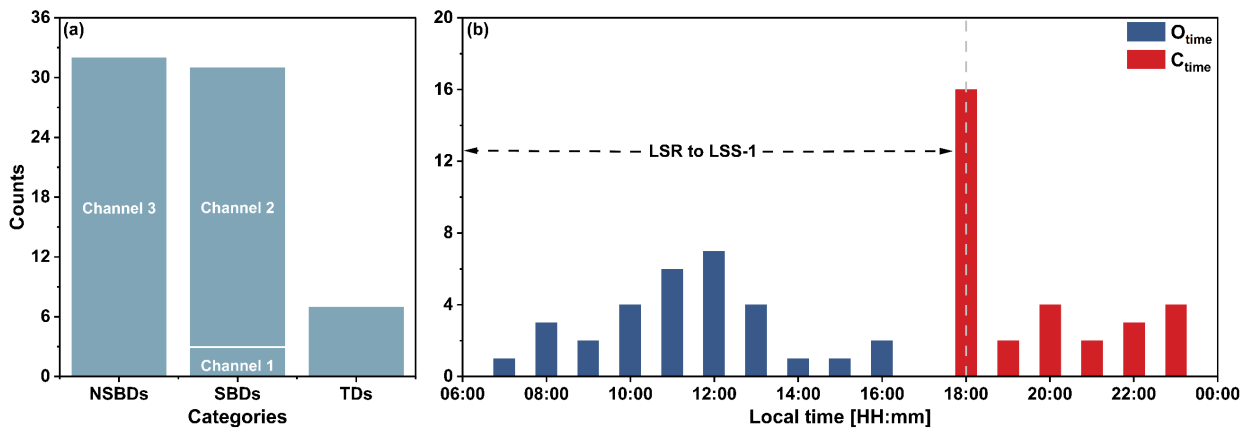
Figure 34. Diurnal variations in vertical profiles of pollutants and meteorological factors at FK. (a) AEC, (b) NO₂, (c) HCHO, and (d) CHOCHO measured by MAX-DOAS; (e) Temperature and (f) relative humidity (RH) from ERA5 reanalysis. Boundary layer height (BLH) cycles are indicated, while mean vertical profiles displayed in right panels.

3.2 Air current patterns impacted meteorological parameters and trace gases

During the campaign, a unique opportunity arose to study the impact of Typhoon "Prapiroon", which traversed Hainan from southeast to northwest on July 21–23, 2024, and Iits intense disturbances to the local meteorology continued for one week

270 after landfall (Fig. 1a and ~~5~~6). Although the seven typhoon-influenced days (TDs) may be not sufficient sample from a statistical perspective, they represent a coherent case of a recurring meteorological phenomenon—typhoons following a characteristic east-to-west track across the South China Sea. This consistent pathway results in systematic advection of specific air masses into the study region. So ~~We~~ we treated these seven ~~Typhoon days~~ (TDs) as an independent case against NSBDs and SBDs identified by the ORA (Table S~~3~~4). Days influenced by synoptic-scale systems or hybrid weather patterns were

275 excluded to ensure sample representativeness. Notably, most of the SBDs occur consecutively~~,~~. This can be explained by~~reflecting the regularity of~~ the ~~local~~ mesoscale circulation~~,~~, which usually occur under stable atmospheric conditions dominated by persistent high pressure. Figure ~~4~~5 presents the occurrence frequency of NSBDs, SBDs, and TDs, as well as the frequency distribution of SB onset and cessation times (O_{time} and C_{time}). The frequency of NSBDs and SBDs was 32 days (34.8% of the total observation period) and 31 days (33.7%), respectively. The study site exhibited 6.3% lower SB frequency compared

280 to Hainan's regional average (Zhang et al., 2014). SB initiation predominantly occurred after 10:00 LT, peaking at 10:00–13:00 LT (n = 21), likely driven by rapid solar heating in late morning (Parker, 2015). Cessation events clustered post-sunset, with a maximum occurrence at 18:00 LT (n = 16), consistent with the restoration of thermal equilibrium following land surface cooling.

285 **Figure ~~4~~5. Statistics of the results of objective recognition algorithm for different air currents. (a) the frequency of occurrence of NSBDs, SBDs and TDs, as well as (b) the frequency distribution of SB onset and cessation times (O_{time} and C_{time}). The time ranges for local sunrise (LSR) and one hour before local sunset (LSS) are marked with dotted arrows.**

Afterwards, the mean diurnal variations of each meteorological parameter: WS, WD, temperature, RH, BLH, ~~and~~ UV radiation

flux (UVF) and vertical velocity (ω) in three ACPs are summarized in Figure 46. It should be noted that the composite analysis of SBDs excludes dates selected through Channel 1 methodology due to systematic discrepancies in meteorological characteristics compared to Channel 2-derived selections (Fig. S107). This exclusion is maintained in all subsequent analyses unless otherwise specified. Typhoon events show higher WS than NSBDs, but their diurnal variations are similar. The prevailing sSouthwesterly winds persist underare caused by the typhoon's peripheral airflow over the area of measurement, whereas NSBDs feature daytime southwest winds shifting to southerly at night. The weaker WS of SB (peak 1.89 m s^{-1} at 14:00 LT) likely results from urban/vegetation friction and climate impacts of warming and urbanization (Fu et al., 2011; Guo et al., 2011; Li et al., 2018). Observational data reveal a distinct diurnal cycle in SBDs: after sunrise, WS decreased markedly until reaching a trough at 10:00 LT, followed the WD veered from southeast (149.7°) to south-southeast (170.3°), indicating dissipation of the onshore winds. SB onset at 11:00 LT brought a 41.3° clockwise shift to southwest and rising WS. Maximum SB intensity occurred at 15:00 LT with northwesterly flow (289.3°), then winds gradually weakened and shifted offshore. A sharp transition (17:00–18:00 LT) featured a 97.2° anticyclonic shift from southwest (222.3°) to southeast (125.0°), with minimal wind speed (0.37 m s^{-1}), signaling SB termination and LB onset.

Compared to SB and typhoon conditions, NSBDs exhibit higher surface temperatures and UVF coefficients, likely attributable to reduced cloud cover and enhanced solar irradiance. The well-documented SB thermal mitigation effect manifests itself in the transport of water vapor inland by advancing oceanic breeze fronts (Yan and Anthes, 1988), producing discernible surface cooling and driving a premature rise and anomalous increase in RH following the SB onset (13:00–17:00 LT) (Fig. 65c,d). Typhoon conditions exhibit thermal minimums due to combined radiative shielding from cyclonic circulation and extensive cloud coverage, while maintaining elevated humidity levels (mean RH: 82.28 ± 3.49%) through continuous moisture advection. BLH displays distinct variability across ACPs: NSBDs show rapid BLH growth peaking at 1,215.72 m (13:00 LT) from enhanced thermal convection, whereas TDs achieve the highest mean BLH (694.07 ± 122.07 m) through persistent vertical wind shear and turbulent mixing (NSBDs: 619.26 ± 363.76 m, SBDs: 385.31 ± 281.80 m). Consistent with previous findings, SB conditions exhibit suppressed PBL development, attributable to periodic advection of stable marine air masses that inhibit atmospheric destabilization (Li et al., 2025; Puygrenier et al., 2005). As shown in Fig. 6(g) to (i), the differences in BLH further modulate vertical motion under the three ACPs. The positive values of vertical velocity (ω) indicate downward airflow

movement. Upward motion prevails below 2 km on NSBDs and TDs, associated with enhanced convection and typhoon-induced ascent, respectively. In contrast, SBDs feature weak morning uplift that transitions to subsidence as the SB intensifies, with the strongest downward motion ($\omega \approx 0.05$ m s^{-1}) near 1.5 km, suggesting that the SB further restricts pollutant dispersion in FK.

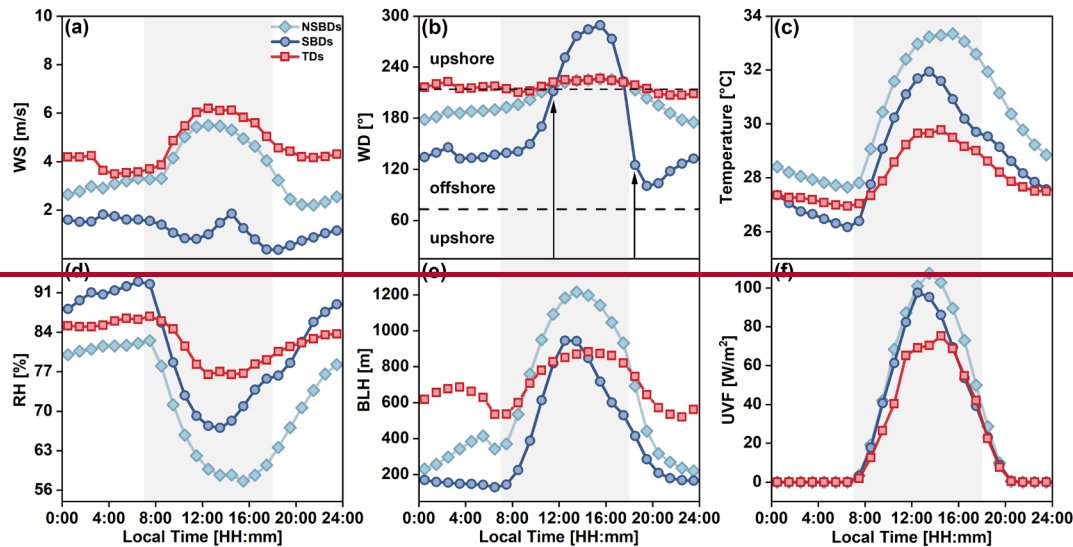
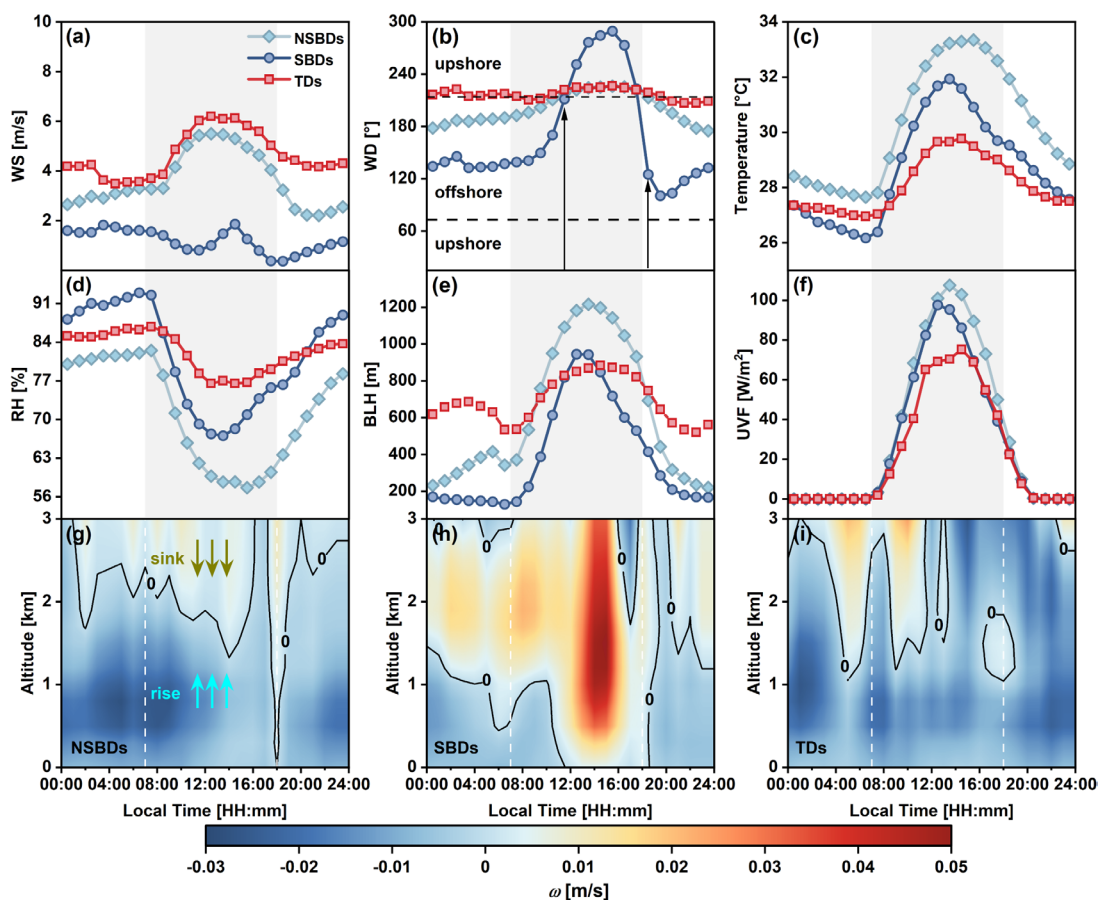



Figure 5̶6. Diurnal variations of meteorological factors under different air currents. Mean diurnal variations of (a) WS, (b) WD, (c) temperature, (d) RH, (e) BLH,̶ ̶a̶n̶d̶ (f) UV radiation flux (UVF) and (g-i) vertical velocity (ω) under three air current patterns (ACPs): NSBDs, SBDs, and TDs. The gray shaded area represents the observation period of MAX-DOAS. The two black dotted lines define the range of WD for onshore and offshore winds. Critical transition points for SB initiation and cessation are marked with arrows in (b). The blue and brown arrows represent rising and sinking air currents, respectively.

Accompanied with meteorological parameters, it can be expected that the vertical distribution of NO$_2$, HCHO and CHOCHO would be also varied under different ACPs. As shown in Fig.̶ ̶ure 76̶, surface NO$_2$ exhibit typical bimodal diurnal variations (morning/evening peaks) during both NSBDs and TDs, in w̶h̶i̶c̶h̶ ̶i̶s̶ contrast to the distinct ̶N̶O̶$_2$̶ accumulation pattern during SBDs, w̶i̶t̶h̶ where concentrations r̶i̶s̶i̶n̶g̶ rise sharply from 12:00 LT,̶ and peak̶i̶n̶g̶ at 15:00 LT (2.10 ppbv) coinciding d̶u̶r̶i̶n̶g̶ ̶S̶B̶D̶s̶,̶ ̶t̶e̶m̶p̶o̶r̶a̶l̶l̶y̶ ̶c̶o̶u̶p̶l̶e̶d̶ with SB front progression and representing a 69.6% overall increase.̶ ̶T̶h̶e̶r̶e̶ ̶w̶a̶s̶ ̶a̶ ̶6̶9̶.̶6̶%̶ ̶i̶n̶c̶r̶e̶a̶s̶e̶ ̶i̶n̶ ̶N̶O̶$_2$̶ ̶t̶h̶r̶o̶u̶g̶h̶o̶u̶t̶,̶ ̶w̶h̶i̶c̶h̶ ̶r̶e̶s̶u̶l̶t̶s̶ ̶i̶n̶ ̶t̶h̶e̶ ̶S̶B̶D̶s̶ ̶(̶1̶.̶7̶3̶ ̶±̶ ̶0̶.̶3̶0̶ ̶p̶p̶b̶v̶)̶ ̶e̶x̶h̶i̶b̶i̶t̶i̶n̶g̶ ̶t̶h̶e̶ ̶l̶a̶r̶g̶e̶s̶t̶ ̶m̶e̶a̶n̶ ̶s̶u̶r̶f̶a̶c̶e̶ ̶V̶M̶R̶s̶ ̶c̶o̶m̶p̶a̶r̶e̶d̶ ̶t̶o̶ ̶t̶h̶e̶

~~NSBDs (1.52 ± 0.82 ppbv) and TDs (1.23 ± 0.59 ppbv).~~ This synchroni~~z~~sation suggests that pollutants from the onshore WD (particularly the northwest and northeast) are transported locally by the SB and accumulate due to the weak WS (Fig. S~~8~~11). Both HCHO and CHOCHO average surface concentrations exhibited similar gradient patterns along with ACPs: NSBDs ~~(HCHO: 4.84 ± 0.67 ppbv, CHOCHO: 0.114 ± 0.017 ppbv)~~ > SBDs ~~(HCHO: 4.26 ± 0.75 ppbv, CHOCHO: 0.108 ± 0.017 ppbv)~~ > TDs ~~(HCHO: 3.74 ± 0.52 ppbv, CHOCHO: 0.0995 ± 0.010 ppbv)~~, which is closely related to the differences in ~~their~~ meteorological factors (WS, temperature, RH and UVF) (Fig. ~~5~~6). Diurnal profiles of surface HCHO and CHOCHO generally followed ~~on NSBDs and TDs were consistent with~~ their average patterns throughout the summer, though~~. In contrast,~~ SBDs induce a unique bimodal HCHO distribution ~~with primary (5.10 ppbv) and secondary (5.08 ppbv) maxima at 10:00 and 13:00 LT.~~ and a delayed CHOCHO peak~~s occurred at 12:00 LT (0.15 ppbv) on~~ compared with NSBDs and TDs, ~~but shifted to 13:00 LT (0.13 ppbv) on SBDs. This delayed peak is temporally synchronized with the secondary peak of HCHO, suggesting that the SB~~which may be attributed to sea-breeze circulation transports VOCs-rich air masses from inland and enhanced secondary formation of ~~to the local area, whose photo-oxidation further affects the~~ HCHO and CHOCHO~~concentrations at FK. J. Wang et al. (2025) similarly found that SB brings in air masses from boats, vegetation, and marine organisms, resulting in high levels of carbonyls in coastal areas in the spring.(J. Wang et al., 2025)~~

Vertical profiling analysis reveals distinct stratification characteristics of pollutants across three ACPs. For NO_2, maximum concentrations remain consistently confined below 300 m altitude across all patterns. ~~It is important to note that the horizontal transport of the SB results in overall higher NO_2 concentrations within 0-300m after the O_{time} (Fig. 6b).~~ Similar to the surface concentrations, the vertical gradients of HCHO and CHOCHO followed NSBDs > SBDs > TDs sequence, clustering at 800/500/300 m and 400 m/300 m/surface, respectively. Thermodynamic vertical transport mechanisms significantly influence pollutant distribution. Intense midday thermal plume convection causes HCHO and CHOCHO to disperse upward. As temperatures drop, convective activity decreases, but residual pollutants remain at higher altitudes and continue vertical transport through atmospheric motions. During marine breeze episodes, the thermal internal boundary layer (TIBL, ~200 m thick) forms as cold marine air flows over the warm land surface, restricting pollutant dispersion (Arrillaga et al., 2016; Cuxart et al., 2014; Talbot et al., 2007). Observations revealed that pollutant concentrations were significantly elevated within the TIBL during SB penetration, with CHOCHO being particularly persistent. Interestingly, pollutant concentrations aloft

decreased markedly around 15:00 LT, particularly for HCHO, likely due to the combined effects of strong subsidence and a concurrent shift in upper-level WD that weakened pollutant transport (Fig. 6h and S12). ~~the WD at altitudes above 500 m shifted to southerly after 15:00 LT on SBDs, which weakened pollutant transport and subsequently led to significant reductions in NO~2~, HCHO, and CHOCHO concentrations at elevated levels, especially HCHO (Fig. 6f and S9).~~

360   Surface VMRs to VCDs ratios (R_{SV}), quantifying altitudinal distribution patterns (Fig. ~~7~~6d,h,i). Elevated R_{SV} values indicate enhanced near-surface accumulation, whereas lower values suggest predominant vertical dispersion to higher altitudes. Following SB incursion, the universal elevation of R_{SV} across species suggests that the onshore winds primarily affect pollutants in the lower atmosphere. Notably, the significantly more pronounced R_{SV} enhancement observed in CHOCHO relative to HCHO suggests that marine-transported VOCs prefer to produce CHOCHO rather than HCHO. During typhoon

365   events, persistent updrafts transport pollutants to higher altitudes, resulting in the lowest mean R_{SV} values, particularly for HCHO, which remains present between 1 and 3 km. Overall, favorable meteorological conditions on NSBDs promote photochemistry and vertical mixing, while SB transport and strong subsidence enhances near-surface pollution levels. Typhoons dilute pollutants but lift them to higher altitudes. ~~In addition, the average R~SV~ (NO~2~: 0.71 ± 0.090; HCHO: 3.07 ± 0.36; CHOCHO: 0.03 ± 0.0026) of all species were the least during the typhoon period, indicating that the vertical transport~~

370   ~~induced by the typhoon significantly increased the height of pollutant distribution, especially for HCHO, which persisted at 1–3 km (Fig. 6g).~~

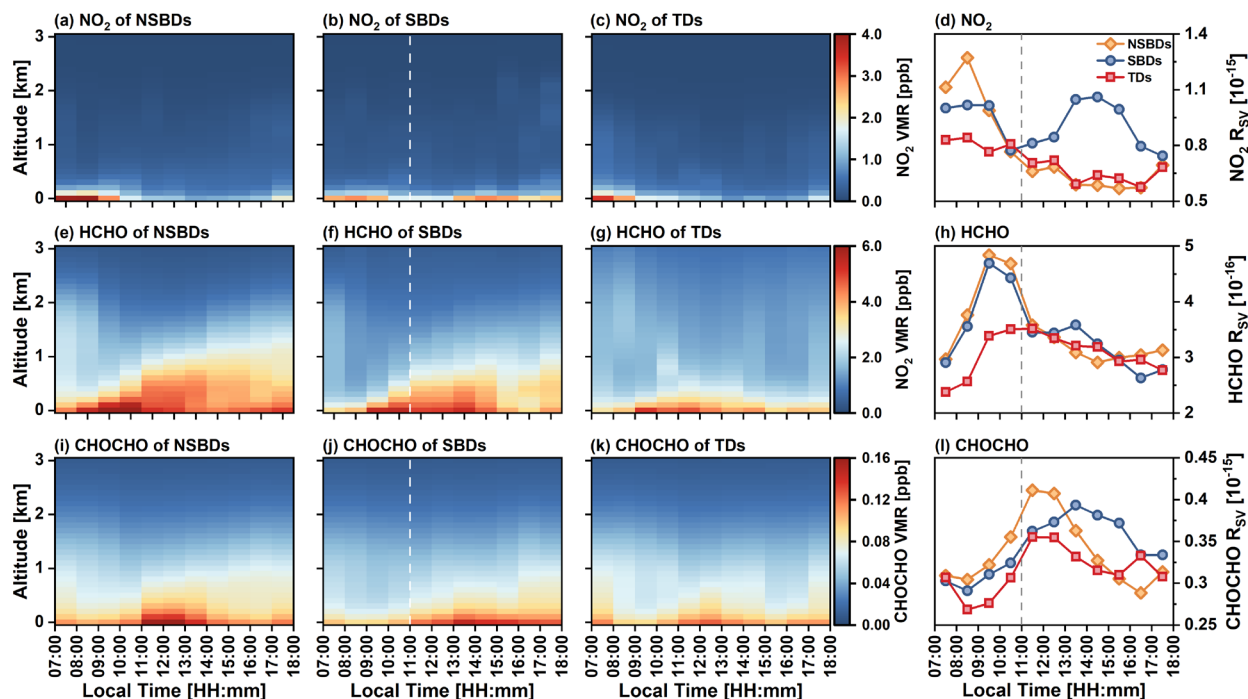



Figure 67. Diurnal variations in vertical distributions of pollutants under different air currents. Mean diurnal vertical distribution of (a-c) NO_2, (e-g) HCHO, and (i-k) CHOCHO on NSBDs, SBDs, and TDs. Diurnal variations of the ratios of surface VMRs and VCDs (R_{VS}) for (d) NO_2, (h) HCHO and (l) CHOCHO are shown in the far right column. The O_{time} of the SB is indicated in the SBDs and R_{SV} panel by a white dashed line and a grey dotted line, respectively.

3.3 Ozone and photochemical indicators under different air currents

To investigate the influence of different ACPs on summertime O_3 concentration and photochemical formation, K-means clustering was to characterize diurnal pattern of O_3 concentration (S3 (Supplements)). Figure 87 shows the daily profiles of O_3 and its precursors (NO_2 and HCHO from MAX-DOAS) and meteorological parameters (temperature, RH and total precipitation (TP)) for the five categories during the observation period. ~~The morning minimum concentration is defined as the background level, while T~~the difference between the midday peak and the morning minimum represents net O_3 increment (ΔO_3). The statistics on the number of days, background levels, ΔO_3, NSBDs and SBDs for each category are shown in Table 1S5. Five O_3 clusters exhibit different concentration characteristics: Cluster 1 (moderate background/low ΔO_3), Cluster 2 (low background/high ΔO_3), Cluster 3 (low background/extreme ΔO_3), Cluster 4 (high background/low ΔO_3), and Cluster 5 (low background/moderate-high ΔO_3). Moreover, Cluster 1 and Cluster 5 exhibit a high prevalence of SBDs, accounting for 64.7%

and 83.3%, respectively, while Clusters 3 and 4 are predominantly composed of NSBDs, at 71.4% and 86.7%. Cluster 2 shows an equal distribution of 50% each.

Table 1. After dividing the daily O_3 changes into five clusters based on the K-means method, the number of days, the O_3 background level, the net O_3 increment (ΔO_3), and the number of NSBDs and SBDs for each cluster are as follows.

Cluster	days	background level	ΔO_3	NSBDs	SBDs
Cluster1	17	15.31 ppbv	9.50 ppbv	6	11
Cluster2	12	9.04 ppbv	23.43 ppbv	6	6
Cluster3	7	9.31 ppbv	35.29 ppbv	5	2
Cluster4	15	22.23 ppbv	10.00 ppbv	13	2
Cluster5	12	7.36 ppbv	15.81 ppbv	2	10

O_3 concentrations are influenced by meteorological factors such as temperature, RH, solar radiation and wind. In general, strong solar radiation, high surface air temperatures and low RH favor the photochemical production of O_3, leading to increased O_3 concentrations (Seinfeld and Pandis, 2016). Specifically, in Clusters 1 and 5, a higher proportion of SBDs results in lower temperatures and elevated RH, leading to significantly reduced overall O_3 levels (20.1 ± 3.3 and 15.3 ± 5.3 ppbv). In contrast, Cluster 3 and 4, dominated by NSBDs, experience higher temperatures and lower RH, which favor O_3 formation, thereby yielding higher O_3 concentrations (25.4 ± 10.7 and 27.3 ± 3.9 ppbv). Specific daily variations can also be seen, following the arrival of marine air, the decrease in O_3 concentration in Clusters 1 and 5 is more pronounced than that in Cluster 3 and 4, occurring nearly in tandem with decreases in temperature and increases in RH, suggesting that the meteorological shifts induced by the SB are generally unfavorable for O_3 formation (Fig. 8~~7~~a,b,c). The correlations between ΔO_3 and both change in temperature (ΔT, defined as the difference between the daytime maximum and the morning minimum) and change in RH (ΔRH, defined as the difference between the morning maximum and the daytime minimum) further corroborate this conclusion (Fig. S14~~1~~). Compared to NSBDs, ΔT exhibited a weaker positive correlation ($R^2 = 0.24$) with ΔO_3 during marine breeze periods while maintaining a consistent negative correlation ($R^2 = 0.38$) with ΔRH, suggesting that SB arrival suppressed photochemical O_3 production.

Clusters 1 and 4 exhibit relatively low daytime ΔO_3 because both start from elevated nocturnal O_3 baselines. In Cluster 1, possible compensation by onshore winds sustained moderate O_3 levels despite heavy nighttime precipitation. Lower nocturnal RH in Cluster 4 inhibited plant stomatal uptake and diminishes dry deposition efficiency, allowing O_3 to be accumulated

(Kavassalis and Murphy, 2017). In contrast, Cluster 5 shows higher ΔO_3 than Clusters 1 and 4, despite its worse meteorological conditions, for two synergistic reasons: first, intense precipitation and high RH suppress the nocturnal O_3 baseline, and second, daytime SB transport delivers greater NO_2 precursor loads, enhancing photochemical O_3 production and amplifying the daytime peak. Cluster 3 features a sharp afternoon increase in O_3—reaching 44.6 ppbv at 18:00 LT—likely driven by elevated midday HCHO concentrations boosting radical production, combined with a weakening NO_2 titration sink as NO_2 levels fall toward evening (Fig. 8~~7~~a,e,f). The contribution of ACPs to pollutant was quantified using the cumulative ratio (R_{SN}) of pollutant concentrations between SBDs and NSBDs across five clusters (Fig. S15~~2~~). R_{SN} and cumulative R_{SN} approaching 1 or 5 indicate equivalent contributions, while values exceeding or falling below these thresholds demonstrate SBDs or NSBDs dominance, respectively. During 11:00–15:00 LT, O_3 R_{SN} for Clusters 1 and 5 (characterized by low temperature, high humidity and high TP) remained below 1, whereas Clusters 2 and 4 (high temperature, low humidity, very low TP) frequently exceeded 1. Furthermore, O_3 cumulative R_{SN} gradually rose after the O_{time}, marginally surpassing 5 between 14:00–16:00 LT— considerably weaker than NO_2. These phenomena suggest that, under favorable meteorological conditions (Clusters 2 and 4), SB-transported precursors (especially NO_2) bolster O_3 formation, whereas under the typically induced adverse conditions (Clusters 1 and 5), SB prefers to inhibit O_3 production.

410

415

420

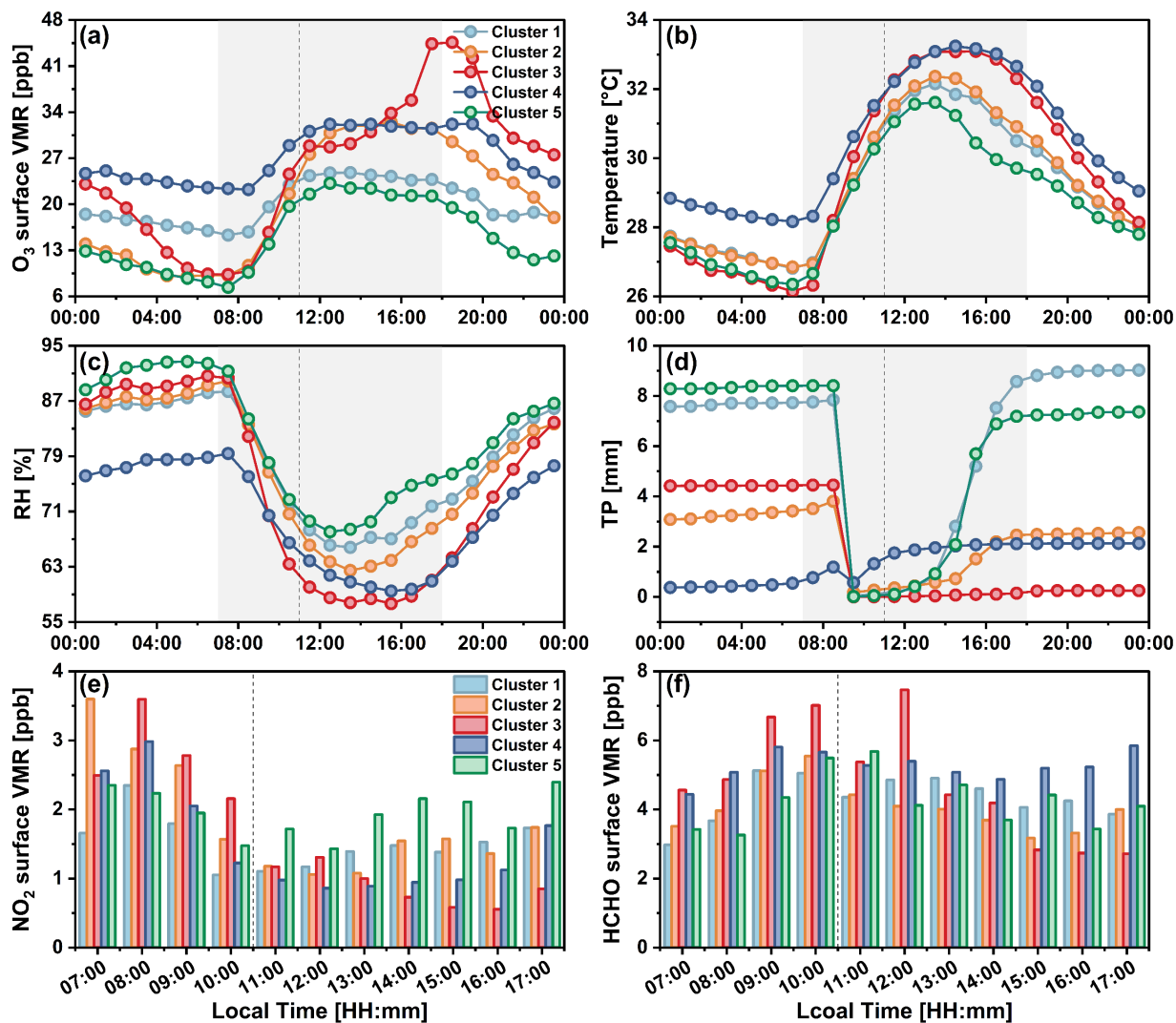



Figure 78. Diurnal variations of O₃, meteorological factors and O₃ precursors in different categories. Mean diurnal cycles of (a) O₃, (b) temperature, (c) RH, (d) total precipitation (TP), (e) NO₂ and (f) HCHO for five categories. MAX-DOAS instrument observation times are indicated by gray shaded areas. The black dotted line indicates when the SB started.

Figure 98 illustrates the diurnal variations in the vertical distributions of R_{GF}, FNR, and GNR under different ACPs, as well as the average characteristics of the vertical profiles. The average surface R_{GF} for three ACPs were 0.024 ± 0.0044, 0.0267 ± 0.0056, and 0.0271 ± 0.0044, all remaining below the 0.04 threshold even during peak daytime concentrations, matching observational records from rural environments (Lee et al., 1995; Moortgat et al., 2002). On NSBDs, enhanced solar radiation

at midday increased CHOCHO concentrations, leading to a simultaneous increase in R_{GF} at the surface (peak 0.03 at 12:00 LT) and in the altitude layer below 300 m. Owing to the distinct vertical distributions of HCHO and CHOCHO, R_{GF} values observed above 300 m decline markedly, and the low-R_{GF} region ($R_{GF} < 0.025$) progressively migrates upward over time (Fig. 76e,i and 98a). The surface R_{GF} exhibits a markedly different pattern on SBDs. Following the onset of the SB, R_{GF} increases steadily until a surface maximum of 0.038 is reached at 16:00, which again proves that the SB transports more VOCs that tend to produce CHOCHO. Aircraft-based photochemical tracer studies indicated that under low-NO_x conditions, isoprene peroxyl radicals (ISOPO2) rapidly isomerize to form CHOCHO (Chan Miller et al., 2017). Given the relatively low local NO_2 levels, it is reasonable to infer that the VOCs introduced by the SB are BVOC (Especially isoprene). The upper R_{GF} also shown a similar upward trend to the near-surface, but is limited to below 300m, suggesting that oceanic air advection transports BVOC mainly in this altitude range (Fig. 98b).

Previous studies have identified regions with biogenic emissions typically showing lower R_{GF} values (< 0.04), while higher R_{GF} (> 0.04) are associated with strong AVOC emissions (Chan Miller et al., 2014; Chen et al., 2023). As shown in Fig. 98d, all ACPs demonstrated mean R_{GF} values < 0.04 across altitudes, indicating BVOC dominance in the FK, which can be supported by the bottom-up emission maps from CAMS reanalysis data (Fig. S16). It shows higher emission flux of isoprene compared to anthropogenic NO_x and VOCs—especially considering that other BVOC species were not included here, e.g., monoterpenes and sesquiterpenes. On NSBDs, R_{GF} progressively increased with altitude, suggesting higher-altitude VOCs preferentially form CHOCHO over HCHO, consistent with previous findings (Hong et al., 2024; Zhang et al., 2021). The overall higher R_{GF} at 0–2,000 m on SBDs before 10:00 LT may be due to a combination of southerly onshore winds and mountain winds in the morning, which bring the BVOC that has accumulated at high altitude during the night from the mountains to the local areameasurement site. The 08:00 LT R_{GF} values are high throughout the vertical column, peaking at 700 m (0.042), especially below 1,000 m where 54.54% of the R_{GF} exceeds 0.04, suggesting that AVOC may be contributing more here. Due to the transport of BVOC emitted from the vegetation of Wuzhi Mountain by "Prapiroon" to FK, the R_{GF} remained elevated at most heights (especially at the mid-levels) under typhoon conditions, with a peak of 0.038 (at the 800 m). Both SB and typhoon scenarios amplified mid-altitude R_{GF}, demonstrating distinct vertical profiles from NSBDs.

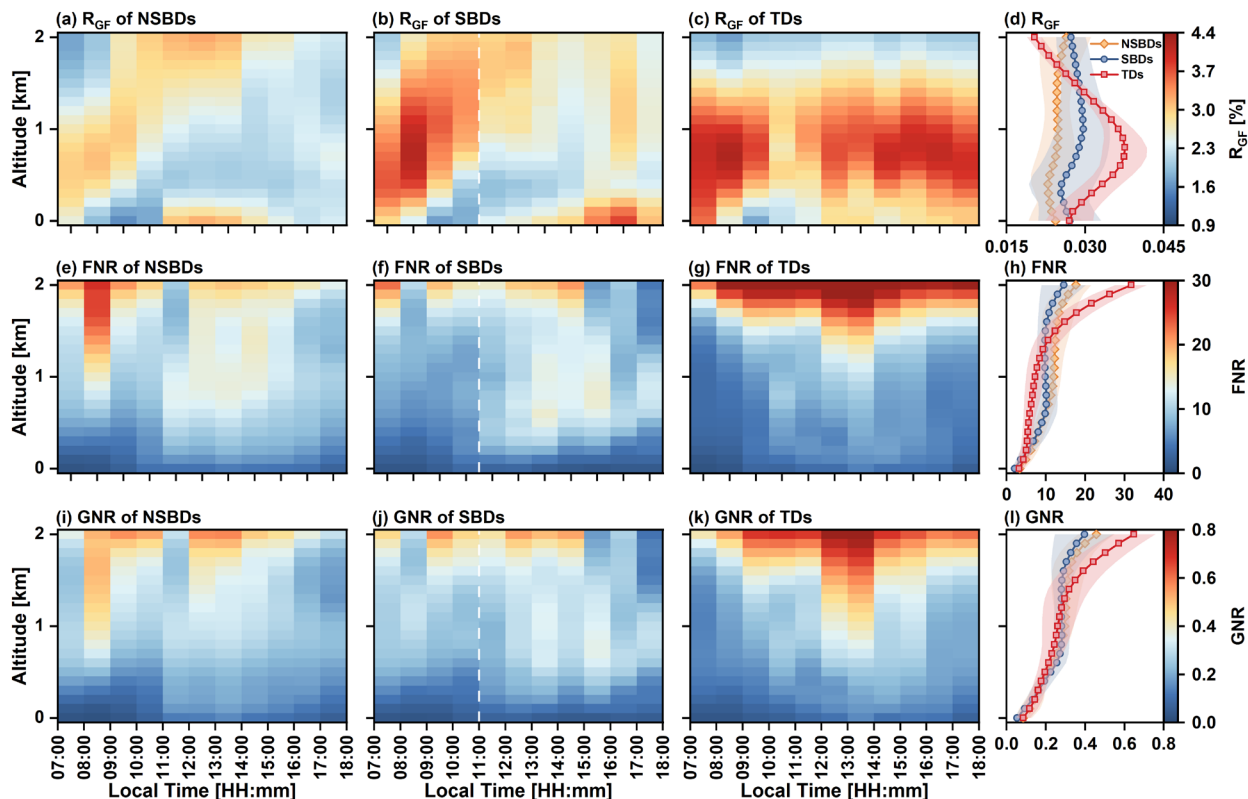



Figure 89. Diurnal variations in vertical distributions of Photochemical indicators. Mean diurnal cycles in the vertical distribution of (a-c) R_{GF}, (e-g) FNR, and (i-k) GNR on NSBDs, SBDs, and TDs. Vertical profiles of (d) R_{GF}, (h) FNR and (l) GNR are shown on the far right, with the shaded area indicating the standard deviation. The O_{time} of the SB is indicated by a white dashed line in the SBDs panel.

Different ACPs have a strong influence on the spatial and temporal distribution of O_3 precursors, especially in SBDs and TDs, where the influx of external precursors can lead to anomalous distributions of pollutant concentrations in the near-surface and upper atmosphere, which may alter the regional O_3 production mechanism (Fig. 7<s>6</s> and 9<s>8</s>). The determination of FNR and GNR thresholds is critical for discerning this regime. In this study, a third-order polynomial fit was employed to elucidate the relationship between FNR, GNR, and O_3 concentration (Jin et al., 2020; Ren et al., 2022). Figure 10<s>9</s> shows the strong correlations between O_3 levels and both FNR ($R^2 = 0.94$) and GNR ($R^2 = 0.84$) indicate robust polynomial model performance. We treat the transition state as spanning the first 1% of the indicator range of the O_3 peak, and end up with local near-surface transition zones of FNR and GNR of 6.33–8.90 and 0.16–0.24, respectively. it should be noted that the fitted curves show only

a slight downward trend when the FNR (7.52) and GNR (0.19) are larger than the indicator values with peak O_3 concentrations, and are based on the sparse data, introducing considerable uncertainty. Nevertheless, it is certain that the local thresholds for FNR and GNR are much higher than those reported in previous studies (Liu et al., 2021; Wang et al., 2021). This suggests that

470 the critical importance of localized evaluation for FNR and GNR under conditions of extremely low NO_2 concentrations.

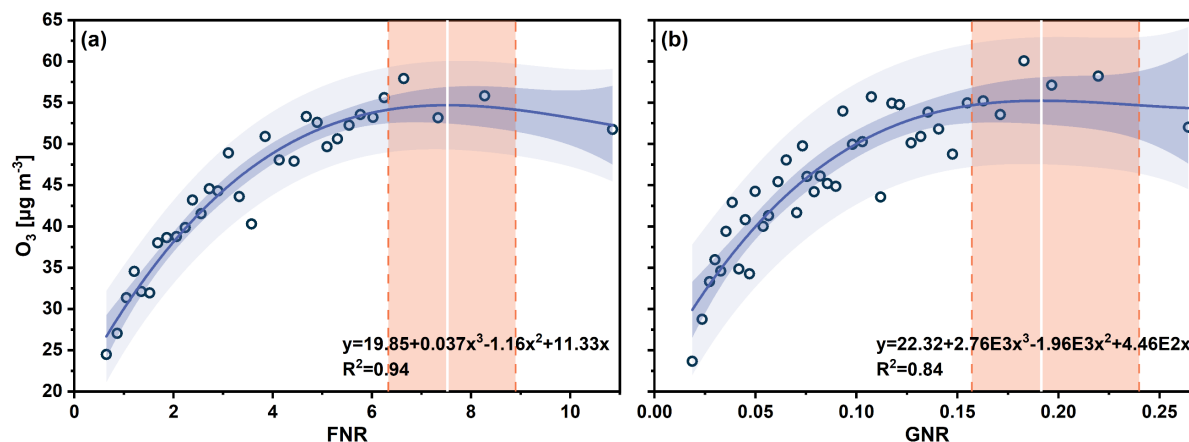



Figure 9̶10. Diagnosing the O_3 formation sensitivity. The relationship between local O_3 surface concentrations and (a) FNR and (b) GNR values was fitted based on MAX-DOAS observations. The blue solid line is the fitted third-order polynomial curve, and the shading indicates the 95% confidence level and 95% prediction band, respectively. The white solid line indicates the peak of the
475 fitted curve and the pink vertical shading indicates the range above the top 1% of the fitted curve (transition regime).

Following the OFS thresholds determined, mean surface FNR and GNR for NSBDs (3.92 ± 1.43 and 0.096 ± 0.044), SBDs (2.53 ± 0.82 and 0.065 ± 0.016), and TDs (3.62 ± 1.33 and 0.097 ± 0.037) indicated that the local near-surface was VOC-limited regimes throughout the day. The FNR and GNR under different ACPs revealed pronounced spatiotemporal variations (Fig. 9̶8e-l). On NSBDs and TDs, two indicators follow a synchronized diurnal cycle, rising in the morning to a maximum at
480 13:00 LT before declining in the afternoon. In contrast, both ratios on SBDs peak synchronously at 12:00 LT (FNR = 3.89, GNR = 0.098), and were then rapidly reduced by the continuously transported NO_2 during the developmental phase of the SB (12:00–15:00 LT), thus strengthening VOC-limited regimes. With the attenuation of SB, reduced external transport and photochemical depletion led to a decline in NO_2 levels, FNR and GNR gradually increase, resulting in a slight shift of the OFS toward the transition regimes, but remained VOC-limited regimes. A mechanism similar to the dual regulation of O_3 production
485 by near-surface SBs still applies at lower atmospheric levels (< 300 m), but it is complicated by the weakening of the NO_2

effect of SB transport with increasing altitude as well as by the heterogeneity of meteorological factors in the vertical direction. The FNR and GNR above 2.5 km on TDs were much higher than those of NSBDs and SBDs throughout the day.

Since the effect of primary emissions on NO_x concentrations decreases with increasing altitude, the sensitivity of O_3 formation to NO_x may increase, leading to an overall increase in FNR and GNR with altitude for all ACPs (Fig. 9~~8~~h,l). Specifically, less primary emissions and wind dilution on TDs~~typhoon days~~ resulted in an exponential increase in both metrics with altitude. In particular, above 2.5 km, the FNR and GNR are much larger than those of NSBDs and SBDs throughout the day (Fig. 9~~8~~g,k). The SB inhibited the vertical transport of pollutants, resulting in a reduced sensitivity of the NO_2 concentration response at high altitude to anthropogenic emissions, so that above 600 m, the FNR is lower on SBDs than on NSBDs, and the difference progressively widens with increasing altitude. This vertical difference feature is weaker in GNR, which is attributed to the fact that CHOCHO is mainly dominated by the oxidation pathway of VOCs rather than the transport process, compared to HCHO. Below 900 m, the photo-oxidation of VOCs and its more significant vertical diffusion advantage over NO_2 together drive the sharp increase in FNR and GNR with altitude. However, from 900 to 1,500 m, both of them show a stabilizing trend, reflecting the synergistic effects of VOCs and NO_x in this altitude range. At higher altitudes, both metrics again showed an increasing trend due to further attenuation of NO_x concentrations.

To further investigate the distribution of OFS across altitudes under different meteorological patterns, we grouped the two metrics below 1.2 km altitude based on local BLH levels and assumed that the same OFS thresholds apply at high altitude and near the ground. Figure 11~~0~~ shows the contributions of the VOC-limited, transition, and NO_x-limited regimes in each altitude layer on NSBDs, SBDs and TDs. Overall, surface OFS is predominantly VOC-limited regime (contribution >70%), while at higher elevations its influence~~the VOC contribution~~ decreases and the transitional and NO_x-limited regime become dominant. On SBDs, the contribution of FNR-based transition increases from the 600–800 m layer upward, a trend not seen with the GNR indicator, likely due to differing vertical sensitivities of the two indices to precursors. Below 400 m, the SB inputs more NO_2, resulting in a contribution of more than 50% to the VOC--limited regime in the 0–200 m and 200–400 m layers. ~~Across all meteorological patterns, GNR indicates a high transition regime contribution—over 40% in the 200–400 m and 400–600 m layers. During typhoons, strong turbulence in the boundary layer transports NO_2 upward, thus attenuating the VOC-limited regime at all levels to some extent. With the FNR indicator, all layers except the 1,000–1,200 m layer are VOC-dominated;~~

however, at the GNR metric, substantial BVOC input results in a predominantly transition and NO$_x$-limited regimes above 200 m. Notably, GNR identifies transition and NO$_x$-limited regimes more often than FNR. On SBDs, it shows that layers above ~400 m are largely NO$_x$-limited. On TDs, FNR indicates VOC-limitation for most layers except 1000–1200 m, whereas GNR shows transition or NO$_x$-limited regimes above ~200 m. An interesting feature is that both metrics exhibit VOC-limitation aloft on TDs, likely because upper-level winds align with the SB pathway and transport upstream NO$_x$ into the region. Such behavior shows that WD may be a key factor driving OFS transitions across different ACPs (Fig. S12). SB and typhoon events transport BVOC and increase VOCs supply, which theoretically shifts OFS toward NO$_x$-limited or transition regimes—a shift GNR captures more clearly, while FNR is less sensitive. Prior work shows that elevated primary HCHO can raise FNR independent of photochemistry (Liu et al., 2021). These findings suggest that (1) GNR may be more suitable as a photochemical indicator of OFS in BVOC-enhanced or BVOC-dominated regions than FNR; (2) reduction of VOCs emissions (especially BVOC) can effectively control the local surface O$_3$ pollution; (3) NO$_2$ transport barriers need to be set up below the low elevation in response to the SB, to inhibit the disruption of exogenous pollutants to photochemical equilibrium of this layer; and (4) The formation process of O$_3$ pollution due to regional and vertical transport of precursors under extreme weather is more complicated, and needs to be combined with regional joint prevention and control and vertical pollution control measures.

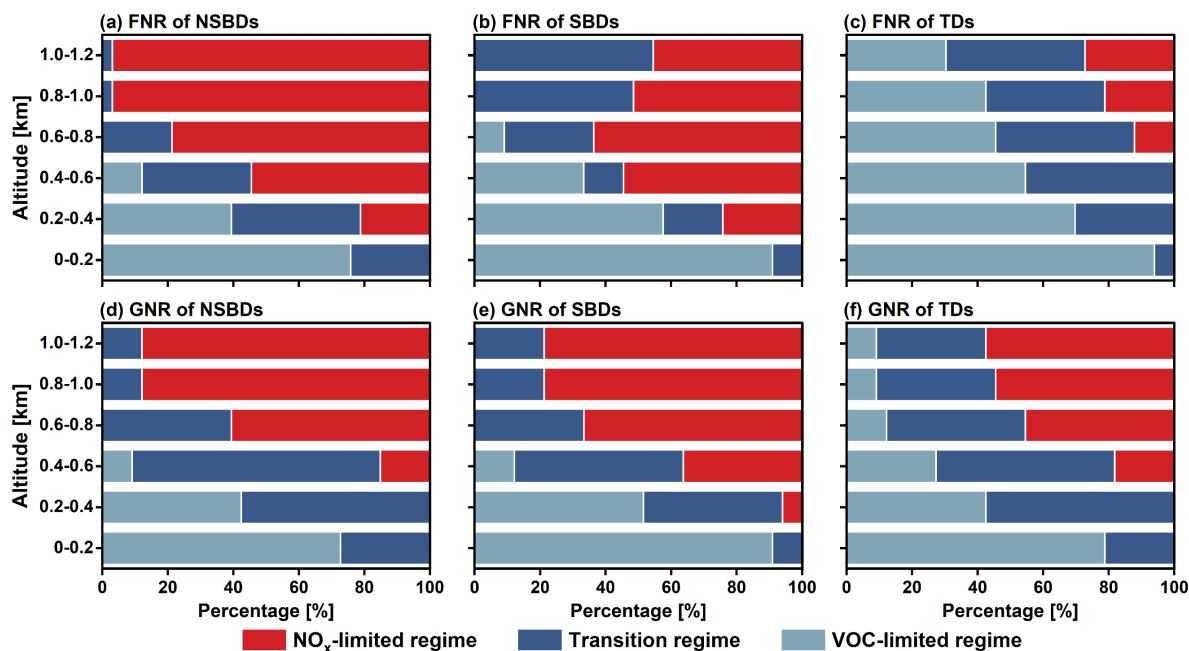



Figure ~~10~~11. Vertical distributions of ozone formation sensitivity under different air currents. The contribution of VOC-limited, transition, and NOₓ-limited regimes for each layer below 1.2 km on (a, d) NSBDs, (b, e) SBDs and (c, f) TDs at the observation site.

4 Discussion

This study employed MAX-DOAS observations to characterize the vertical distribution of daytime aerosols, NO_2, HCHO, and CHOCHO in a coastal rural area of an island environment. The aerosol AOD (0.29 ± 0.1) and NO_2 surface VMRs (1.61 ± 0.53 ppbv) were substantially lower than those typically observed in urban settings, whereas the HCHO and CHOCHO surface VMRs (4.33 ± 1.07 and 0.10 ± 0.02 ppbv) were comparable to those in several megacities (Hong et al., 2024). HCHO and CHOCHO were closely related to high temperatures and low RH, but due to dilution caused by strong vertical convection, surface HCHO did not show signs of significant elevated midday concentrations similar to those of CHOCHO. The hygroscopic effect of the aerosol caused its vertical distribution pattern to be consistent with the RH. NO_2, HCHO and CHOCHO were concentrated below 300 m, 800 m and 500 m, respectively. Notably, pollutants remaining at higher elevations in the afternoon were transported further upward.

The observational findings provide critical insights into the characteristics of pollutants on the NSBDs, SBDs and TDs. In this study, an ORA was established for selecting the first two ACPs. The results demonstrate that the occurrence frequencies of

SBDs and NSBDs during local summer reach 33.7% and 34.8%, respectively, with SB typically initiating at 11:00 LT and dissipating by 18:00 LT. Indeed, The ORA can be broadly applied, as parameter tuning and threshold adjustment allow for the selection of circulation patterns that correspond to local characteristics. Nevertheless, precise delineation of NSBDs remains challenging in complex island environments under monsoon influence, as transient cloud cover may introduce anomalies in

545   RH variation indices even on valid NSBDs. Combining the ORA with MAX-DOAS observations, we analyzed the spatial and temporal distributions of pollutants and photochemical indicators (R_{GF}, FNR and GNR) under different ACPs. The overall low R_{GF} values (< 0.04) for all ACPs indicated predominant local BVOC contributions. Elevated concentrations (e.g., O_3) and vertical distribution (e.g., HCHO, CHOCHO) of pollutants are observed on NSBDs, driven by highly favorable meteorological conditions that enhance VOCs photochemical oxidation and vertical dispersion, necessitating stringent controls on precursor

550   emissions and transport process under this pattern. SB introduced moist air masses that cooled surface temperatures and its vertical structure reduced BLH, suppressing pollutant generation and dispersion. Notably, SB intrusion enhanced NO_2 and BVOC transport below 300 m, which may enhance O_3 production under suitable conditions. Typhoons minimized pollutant concentrations through strong dilution by wind and precipitation, yet simultaneously elevated mid-altitude BVOC via regional transport and promoted vertical dispersion through intense turbulence.

555   Vertical heterogeneity in OFS under distinct ACPs reveals complex dependence of O_3 formation on airflow-driven precursors. Local surface O_3 formation showed VOC-sensitive characteristics based on FNR and GNR, transitioning to NO_x-limited regime with altitude, but VOC-limited regime still contributed to OFS at high altitude on TDs. Additional BVOC inputs from SB and typhoons and strong local biological discharges suggest that GNR may be a more appropriate indicator of OFS. It should be noted that our determined FNR and GNR thresholds significantly exceeded literature values, suggesting existing

560   classification criteria may inadequately represent tropical coastal rural environments with low NO_2 baseline concentrations. Overall, abatement of VOCs is an effective means to manage local O_3, while the basic strategy of controlling the emission of NO_x as well as VOCs from the SB may help to reduce local photochemical pollution. It should be noted that we cannot accurately determine the OFS in the vertical direction because the DPS only provides surface O_3 concentration data and carry over the surface OFS thresholds to other altitudes introduces a large uncertainty. Future work should aim to develop local FNR

565   and GNR thresholds in the vertical direction for more accurate assessments.

This study reveals several summertime physical and chemical processes that are characteristic of Hainan Island, so the quantitative values reported here—such as pollutant concentrations, ORA parameter ranges and FNR/GNR thresholds—are shaped by the island's unique land–sea configuration, monsoon regime, and emission environment and should be treated as site-specific. However, the vertical stratification mechanisms linked to species lifetimes and atmospheric dynamics, the thermodynamic and compositional effects of SB intrusions, the typhoon-related scavenging, redistribution, and uplift, and the diagnostic value of different photochemical indicators under different ACPs are processes that commonly occur across many coastal and island settings. Thus, while specific numerical thresholds may vary, the underlying conceptual processes and the ORA framework offer a transferable basis for interpreting similar coastal phenomena, particularly in regions with comparable climatic and geographic conditions.

Data available. The MAX-DOAS datasets generated in this study are available from the corresponding authors (S.W.) upon reasonable request for research purposes. O_3 data can be downloaded from the https://www.mee.gov.cn, ERA5 pressure-level data can be downloaded from https://cds.climate.copernicus.eu/datasets/reanalysis-era5-pressure-levels?tab=overview, ERA5 single-level data can be downloaded from https://cds.climate.copernicus.eu/datasets/reanalysis-era5-single-levels?tab=overview, ERA5-Land data can be downloaded from https://cds.climate.copernicus.eu/datasets/reanalysis-era5-land?tab=overview.

Author Contributions. B. L., S. Z., and Z. J. set up and operated the MAX-DOAS instrument, and B. L. analyzed the data with the help of S. Z., Y. S., and Z. J.; B. L. wrote the manuscript, S. W., B. Z., S. Z., and R. X. provided comments, and S. W. conceived, supervised, reviewed and edited the paper. All authors participated in the discussion of the results and revision of the manuscript.

Competing interests. The authors declare that they have no competing interests.

Acknowledgements. This study has been supported by National Natural Science Foundation of China (42375089, W2411028),

National Key Research and Development Program of China (2023YFC3709401 and 2023YFC3709400), and National Civilian Space Infrastructure Project (Grant No. Y5BZ31AC60). The authors would like to thank the Chinese Ministry of Ecology and Environment for providing the O_3 data and to the developers of the ERA5 data for providing free and Open-access materials. The authors would like to thank the Institute of Remote Sensing/Institute of Environmental Physics, University of Bremen, Germany, for providing the SCIATRAN radiative transfer model used in this paper.

Financial support. This study has been supported by National Natural Science Foundation of China (42375089, W2411028), National Key Research and Development Program of China (2023YFC3709401 and 2023YFC3709400), and National Civilian Space Infrastructure Project (Grant No. Y5BZ31AC60).

References

Arrillaga, J. A., Jiménez, P., Vilà-Guerau de Arellano, J., Jiménez, M. A., Román-Cascón, C., Sastre, M., and Yagüe, C.: Analyzing the Synoptic-, Meso- and Local- Scale Involved in Sea Breeze Formation and Frontal Characteristics, J. Geophys. Res.: Atmos., 125, https://doi.org/10.1029/2019jd031302, 2020.

Arrillaga, J. A., Yagüe, C., Sastre, M., and Román-Cascón, C.: A characterisation of sea-breeze events in the eastern Cantabrian coast (Spain) from observational data and WRF simulations, Atmos. Res., 181, 265-280, https://doi.org/10.1016/j.atmosres.2016.06.021, 2016.

Azorin-Molina, C., Connell, B. H., and Baena-Calatrava, R.: Sea-Breeze Convergence Zones from AVHRR over the Iberian Mediterranean Area and the Isle of Mallorca, Spain, J. Appl. Meteorol. Climatol., 48, 2069-2085, https://doi.org/10.1175/2009JAMC2141.1, 2009.

Azorin-Molina, C., Tijm, S., and Chen, D.: Development of selection algorithms and databases for sea breeze studies, Theor. Appl. Climatol., 106, 531-546, https://doi.org/10.1007/s00704-011-0454-4, 2011.

Barthlott, C. and Kirshbaum, D. J.: Sensitivity of deep convection to terrain forcing over Mediterranean islands, Q. J. R. Meteorolog. Soc., 139, 1762-1779, https://doi.org/10.1002/qj.2089, 2013.

Benish, S. E., He, H., Ren, X., Roberts, S. J., Salawitch, R. J., Li, Z., Wang, F., Wang, Y., Zhang, F., Shao, M., Lu, S., and

615   Dickerson, R. R.: Measurement report: Aircraft observations of ozone, nitrogen oxides, and volatile organic compounds over Hebei Province, China, Atmos. Chem. Phys., 20, 14523-14545, https://doi.org/10.5194/acp-20-14523-2020, 2020.

Chan, K. L., Wiegner, M., van Geffen, J., De Smedt, I., Alberti, C., Cheng, Z., Ye, S., and Wenig, M.: MAX-DOAS measurements of tropospheric NO_2 and HCHO in Munich and the comparison to OMI and TROPOMI satellite observations, Atmos. Meas. Tech., 13, 4499-4520, https://doi.org/10.5194/amt-13-4499-2020, 2020.

620   Chan Miller, C., Gonzalez Abad, G., Wang, H., Liu, X., Kurosu, T., Jacob, D. J., and Chance, K.: Glyoxal retrieval from the Ozone Monitoring Instrument, Atmos. Meas. Tech., 7, 3891-3907, https://doi.org/10.5194/amt-7-3891-2014, 2014.

Chan Miller, C., Jacob, D. J., Marais, E. A., Yu, K., Travis, K. R., Kim, P. S., Fisher, J. A., Zhu, L., Wolfe, G. M., Hanisco, T. F., Keutsch, F. N., Kaiser, J., Min, K. E., Brown, S. S., Washenfelder, R. A., González Abad, G., and Chance, K.: Glyoxal yield from isoprene oxidation and relation to formaldehyde: chemical mechanism, constraints from SENEX aircraft
625   observations, and interpretation of OMI satellite data, Atmos. Chem. Phys., 17, 8725-8738, https://doi.org/10.5194/acp-17-8725-2017, 2017.

Chang, B., Liu, H., Zhang, C., Xing, C., Tan, W., and Liu, C.: Relating satellite NO_2 tropospheric columns to near-surface concentrations: implications from ground-based MAX-DOAS NO_2 vertical profile observations, npj Clim. Atmos. Sci., 8, https://doi.org/10.1038/s41612-024-00891-z, 2025.

630   ~~Chauhan, B. V. S., Smallbone, K. L., Berg, M., and Wyche, K. P.: The temporal evolution of HCHO and changes in atmospheric composition in the southeast of the United Kingdom, Case Stud. Chem. Environ. Eng., 11, 101092, https://doi.org/10.1016/j.cscee.2024.101092, 2025.~~

Chen, Y., Liu, C., Su, W., Hu, Q., Zhang, C., Liu, H., and Yin, H.: Identification of volatile organic compound emissions from anthropogenic and biogenic sources based on satellite observation of formaldehyde and glyoxal, Sci. Total Environ., 859,
635   159997, https://doi.org/10.1016/j.scitotenv.2022.159997, 2023.

~~Cheng, T., Xu, C., Duan, J., Wang, Y., Leng, C., Tao, J., Che, H., He, Q., Wu, Y., Zhang, R., Li, X., Chen, J., Kong, L., and Yu, X.: Seasonal variation and difference of aerosol optical properties in columnar and surface atmospheres over Shanghai, Atmos. Environ., 123, 315-326, https://doi.org/10.1016/j.atmosenv.2015.05.029, 2015.~~

Cooke, M. C., Utembe, S. R., Carbajo, P. G., Archibald, A. T., Orr-Ewing, A. J., Jenkin, M. E., Derwent, R. G., Lary, D. J., and

Shallcross, D. E.: Impacts of formaldehyde photolysis rates on tropospheric chemistry, Atmos. Sci. Lett., 11, 33-38, https://doi.org/10.1002/asl.251, 2010.

Cuxart, J., Jiménez, M. A., Telišman Prtenjak, M., and Grisogono, B.: Study of a Sea-Breeze Case through Momentum, Temperature, and Turbulence Budgets, J. Appl. Meteorol. Climatol., 53, 2589-2609, https://doi.org/10.1175/jamc-d-14-0007.1, 2014.

Deng, T., Wang, T., Wang, S., Zou, Y., Yin, C., Li, F., Liu, L., Wang, N., Song, L., Wu, C., and Wu, D.: Impact of typhoon periphery on high ozone and high aerosol pollution in the Pearl River Delta region, Sci. Total Environ., 668, 617-630, https://doi.org/10.1016/j.scitotenv.2019.02.450, 2019.

Duncan, B. N., Yoshida, Y., Olson, J. R., Sillman, S., Martin, R. V., Lamsal, L., Hu, Y., Pickering, K. E., Retscher, C., Allen, D. J., and Crawford, J. H.: Application of OMI observations to a space-based indicator of NO_x and VOC controls on surface ozone formation, Atmos. Environ., 44, 2213-2223, https://doi.org/10.1016/j.atmosenv.2010.03.010, 2010.

Fan, C., de Leeuw, G., Yan, X., Dong, J., Kang, H., Fang, C., Li, Z., and Zhang, Y.: Evolution of aerosol optical depth over China in 2010–2024: increasing importance of meteorological influences, Atmos. Chem. Phys., 25, 11951-11973, https://doi.org/10.5194/acp-25-11951-2025, 2025.

Fortems-Cheiney, A., Chevallier, F., Pison, I., Bousquet, P., Saunois, M., Szopa, S., Cressot, C., Kurosu, T. P., Chance, K., and Fried, A.: The formaldehyde budget as seen by a global-scale multi-constraint and multi-species inversion system, Atmos. Chem. Phys., 12, 6699-6721, https://doi.org/10.5194/acp-12-6699-2012, 2012.

Franco, B., Hendrick, F., Van Roozendael, M., Mueller, J. F., Stavrakou, T., Marais, E. A., Bovy, B., Bader, W., Fayt, C., Hermans, C., Lejeune, B., Pinardi, G., Servais, C., and Mahieu, E.: Retrievals of formaldehyde from ground-based FTIR and MAX-DOAS observations at the Jungfraujoch station and comparisons with GEOS-Chem and IMAGES model simulations, Atmos. Meas. Tech., 8, 1733-1756, https://doi.org/10.5194/amt-8-1733-2015, 2015.

Frieß, U., Monks, P. S., Remedios, J. J., Rozanov, A., Sinreich, R., Wagner, T., and Platt, U.: MAX-DOAS O_4 measurements: A new technique to derive information on atmospheric aerosols: 2. Modeling studies, J. Geophys. Res.: Atmos., 111, https://doi.org/10.1029/2005JD006618, 2006.

Frieß, U., Sihler, H., Sander, R., Pöhler, D., Yilmaz, S., and Platt, U.: The vertical distribution of BrO and aerosols in the Arctic:

665   Measurements by active and passive differential optical absorption spectroscopy, J. Geophys. Res.: Atmos., 116, https://doi.org/10.1029/2011JD015938, 2011.

Fu, C., Dan, L., Tong, J., and Xu, W.: Influence of Typhoon Nangka Process on Ozone Pollution in Hainan Island, Huanjing Kexue, 44, 2481-2491, https://doi.org/10.13227/j.hjkx.202206123, 2023.

Fu, G., Yu, J., Zhang, Y., Hu, S., Ouyang, R., and Liu, W.: Temporal variation of wind speed in China for 1961-2007, Theor.
670   Appl. Climatol., 104, 313-324, https://doi.org/10.1007/s00704-010-0348-x, 2011.

Gangoiti, G., Alonso, L., Navazo, M., Albizuri, A., Perez-Landa, G., Matabuena, M., Valdenebro, V., Maruri, M., Antonio García, J., and Millán, M. M.: Regional transport of pollutants over the Bay of Biscay: analysis of an ozone episode under a blocking anticyclone in west-central Europe, Atmos. Environ., 36, 1349-1361, https://doi.org/10.1016/S1352-2310(01)00536-2, 2002.

675   Glaser, K., Vogt, U., Baumbach, G., Volz-Thomas, A., and Geiss, H.: Vertical profiles of O_3, NO_2, NO_x, VOC, and meteorological parameters during the Berlin Ozone Experiment (BERLIOZ) campaign, Journal of Geophysical Research-Atmospheres, 108, https://doi.org/10.1029/2002jd002475, 2003.

Gordon, G. E.: Atmospheric science: atmospheric chemistry and atmospheric chemistry and physics of air pollution, Science (New York, N.Y.), 235, 1263b-1264b, https://doi.org/10.1126/science.235.4793.1263b, 1987.

680   Guo, H., Xu, M., and Hu, Q.: Changes in near-surface wind speed in China: 1969-2005, Int. J. Climatol., 31, 349-358, https://doi.org/10.1002/joc.2091, 2011.

Hainan Provincial Bureau of Statistics (Ed.) Hainan Statistical Yearbook 2024, China Statistics Press, Beijing, https://stats.hainan.gov.cn/tjj/tjsu/ndsj/2024/202412/P020250116308974141111.pdf, 2024.

Hendrick, F., Müller, J. F., Clémer, K., Wang, P., De Mazière, M., Fayt, C., Gielen, C., Hermans, C., Ma, J. Z., Pinardi, G.,
685   Stavrakou, T., Vlemmix, T., and Van Roozendael, M.: Four years of ground-based MAX-DOAS observations of HONO and NO_2 in the Beijing area, Atmos. Chem. Phys., 14, 765-781, https://doi.org/10.5194/acp-14-765-2014, 2014.

Henyey, L. G. and Greenstein, J. L.: Diffuse radiation in the galaxy, Astrophys. J., 93, 70-83, https://doi.org/10.1086/144246, 1941.

Hong, Q., Liu, C., Hu, Q., Zhang, Y., Xing, C., Ou, J., Tan, W., Liu, H., Huang, X., and Wu, Z.: Vertical distribution and

temporal evolution of formaldehyde and glyoxal derived from MAX-DOAS observations: The indicative role of VOC sources, J Environ Sci (China), 122, 92-104, https://doi.org/10.1016/j.jes.2021.09.025, 2022.

Hong, Q., Xing, J., Xing, C., Yang, B., Su, W., Chen, Y., Zhang, C., Zhu, Y., and Liu, C.: Investigating vertical distributions and photochemical indications of formaldehyde, glyoxal, and NO_2 from MAX-DOAS observations in four typical cities of China, Sci. Total Environ., 954, https://doi.org/10.1016/j.scitotenv.2024.176447, 2024.

Hoque, H. M. S., Irie, H., and Damiani, A.: First MAX-DOAS Observations of Formaldehyde and Glyoxal in Phimai, Thailand, J. Geophys. Res.: Atmos., 123, 9957-9975, https://doi.org/10.1029/2018JD028480, 2018.

Jiang, F., Wang, T., Wang, T., Xie, M., and Zhao, H.: Numerical modeling of a continuous photochemical pollution episode in Hong Kong using WRF–chem, Atmos. Environ., 42, 8717-8727, https://doi.org/10.1016/j.atmosenv.2008.08.034, 2008.

Jiang, Z., Wang, S., Yan, Y., Zhang, S., Xue, R., Gu, C., Zhu, J., Liu, J., and Zhou, B.: Constructing the 3D Spatial Distribution of the $HCHO/NO_2$ Ratio via Satellite Observation and Machine Learning Model, Environ. Sci. Technol., 59, 4047-4058, https://doi.org/10.1021/acs.est.4c12362, 2025.

Jin, X., Fiore, A., Boersma, K. F., Smedt, I. D., and Valin, L.: Inferring Changes in Summertime Surface Ozone–NO_x–VOC Chemistry over U.S. Urban Areas from Two Decades of Satellite and Ground-Based Observations, Environ. Sci. Technol., 54, 6518-6529, https://doi.org/10.1021/acs.est.9b07785, 2020.

Kavassalis, S. C. and Murphy, J. G.: Understanding ozone-meteorology correlations: A role for dry deposition, Geophys. Res. Lett., 44, 2922-2931, https://doi.org/10.1002/2016gl071791, 2017.

KiranKumar, N. V. P., Jagadeesh, K., Niranjan, K., and Rajeev, K.: Seasonal variations of sea breeze and its effect on the spectral behaviour of surface layer winds in the coastal zone near Visakhapatnam, India, J. Atmos. Sol. Terr. Phys., 186, 1-7, https://doi.org/10.1016/j.jastp.2019.01.013, 2019.

Kumar, V., Beirle, S., Dörner, S., Mishra, A. K., Donner, S., Wang, Y., Sinha, V., and Wagner, T.: Long-term MAX-DOAS measurements of NO_2, HCHO, and aerosols and evaluation of corresponding satellite data products over Mohali in the Indo-Gangetic Plain, Atmos. Chem. Phys., 20, 14183-14235, https://doi.org/10.5194/acp-20-14183-2020, 2020.

Lawson, S. J., Selleck, P. W., Galbally, I. E., Keywood, M. D., Harvey, M. J., Lerot, C., Helmig, D., and Ristovski, Z.: Seasonal in situ observations of glyoxal and methylglyoxal over the temperate oceans of the Southern Hemisphere, Atmos. Chem.

Phys., 15, 223-240, https://doi.org/10.5194/acp-15-223-2015, 2015.

Lee, Y.-N., Zhou, X., and Hallock, K.: Atmospheric carbonyl compounds at a rural southeastern United States site, J. Geophys. Res.: Atmos., 100, 25933-25944, https://doi.org/10.1029/95JD02605, 1995.

Lerot, C., Hendrick, F., Van Roozendael, M., Alvarado, L. M. A., Richter, A., De Smedt, I., Theys, N., Vlietinck, J., Yu, H., Van Gent, J., Stavrakou, T., Müller, J. F., Valks, P., Loyola, D., Irie, H., Kumar, V., Wagner, T., Schreier, S. F., Sinha, V., Wang, T., Wang, P., and Retscher, C.: Glyoxal tropospheric column retrievals from TROPOMI – multi-satellite intercomparison and ground-based validation, Atmos. Meas. Tech., 14, 7775-7807, https://doi.org/10.5194/amt-14-7775-2021, 2021.

Li, H., Gong, W., Liu, B., Ma, Y., Jin, S., Wang, W., Fan, R., Jiang, S., Wang, Y., and Tong, Z.: Sea Breeze-Driven Variations in Planetary Boundary Layer Height over Barrow: Insights from Meteorological and Lidar Observations, Remote Sens., 17, 1633, https://www.mdpi.com/2072-4292/17/9/1633, 2025.

Li, Z., Song, L., Ma, H., Xiao, J., Wang, K., and Chen, L.: Observed surface wind speed declining induced by urbanization in East China, Clim. Dyn., 50, 735-749, https://doi.org/10.1007/s00382-017-3637-6, 2018.

Liang, Z. and Wang, D.: Sea breeze and precipitation over Hainan Island, Q. J. R. Meteorolog. Soc., 143, 137-151, https://doi.org/10.1002/qj.2952, 2017.

Liao, J., Wolfe, G. M., Kotsakis, A. E., Nicely, J. M., St. Clair, J. M., Hanisco, T. F., González Abad, G., Nowlan, C. R., Ayazpour, Z., De Smedt, I., Apel, E. C., and Hornbrook, R. S.: Validation of formaldehyde products from three satellite retrievals (OMI SAO, OMPS-NPP SAO, and OMI BIRA) in the marine atmosphere with four seasons of Atmospheric Tomography Mission (ATom) aircraft observations, Atmos. Meas. Tech., 18, 1-16, https://doi.org/10.5194/amt-18-1-2025, 2025.

Liu, G. and Wang, Y.: Vertical distribution characteristics and potential sources of atmospheric pollutants in the North China Plain basing on the MAX-DOAS measurement, Environ. Sci. Eur., 36, 78, https://doi.org/10.1186/s12302-024-00902-z, 2024.

Liu, J., Li, X., Tan, Z., Wang, W., Yang, Y., Zhu, Y., Yang, S., Song, M., Chen, S., Wang, H., Lu, K., Zeng, L., and Zhang, Y.: Assessing the Ratios of Formaldehyde and Glyoxal to NO_2 as Indicators of O_3–NO_x–VOC Sensitivity, Environ. Sci.

740   Technol., 55, 10935-10945, https://doi.org/10.1021/acs.est.0c07506, 2021.

Liu, Q., Liu, D., Gao, Q., Tian, P., Wang, F., Zhao, D., Bi, K., Wu, Y., Ding, S., Hu, K., Zhang, J., Ding, D., and Zhao, C.: Vertical characteristics of aerosol hygroscopicity and impacts on optical properties over the North China Plain during winter, Atmos. Chem. Phys., 20, 3931-3944, https://doi.org/10.5194/acp-20-3931-2020, 2020.

Liu, S., Cheng, S., Ma, J., Xu, X., Lv, J., Jin, J., Guo, J., Yu, D., and Dai, X.: MAX-DOAS Measurements of Tropospheric

745   NO_2 and HCHO Vertical Profiles at the Longfengshan Regional Background Station in Northeastern China, Sensors, 23, https://doi.org/10.3390/s23063269, 2023.

Lou, Y., Chen, Y., Chen, X., and Li, R.: Spatiotemporal estimates of anthropogenic NO_x emissions across China during 2015-2022 using a deep learning model, J. Hazard. Mater., 487, https://doi.org/10.1016/j.jhazmat.2025.137308, 2025.

Mahajan, A. S., Prados-Roman, C., Hay, T. D., Lampel, J., Pöhler, D., Groβmann, K., Tschritter, J., Frieß, U., Platt, U., Johnston,

750   P., Kreher, K., Wittrock, F., Burrows, J. P., Plane, J. M. C., and Saiz-Lopez, A.: Glyoxal observations in the global marine boundary layer, J. Geophys. Res.: Atmos., 119, 6160-6169, https://doi.org/10.1002/2013JD021388, 2014.

Mahajan, A. S., Whalley, L. K., Kozlova, E., Oetjen, H., Mendez, L., Furneaux, K. L., Goddard, A., Heard, D. E., Plane, J. M. C., and Saiz-Lopez, A.: DOAS observations of formaldehyde and its impact on the HOx balance in the tropical Atlantic marine boundary layer, J. Atmos. Chem., 66, 167-178, https://doi.org/10.1007/s10874-011-9200-7, 2010.

755 Martin, R. V., Parrish, D. D., Ryerson, T. B., Nicks Jr., D. K., Chance, K., Kurosu, T. P., Jacob, D. J., Sturges, E. D., Fried, A., and Wert, B. P.: Evaluation of GOME satellite measurements of tropospheric NO_2 and HCHO using regional data from aircraft campaigns in the southeastern United States, J. Geophys. Res.: Atmos., 109, https://doi.org/10.1029/2004JD004869, 2004.

Martins, D. K., Stauffer, R. M., Thompson, A. M., Knepp, T. N., and Pippin, M.: Surface ozone at a coastal suburban site in

760   2009 and 2010: Relationships to chemical and meteorological processes, J. Geophys. Res.: Atmos., 117, https://doi.org/10.1029/2011JD016828, 2012.

Mason, R. S.: Atmospheric Photochemistry, in: Applied Photochemistry, edited by: Evans, R. C., Douglas, P., and Burrow, H. D., Springer Netherlands, Dordrecht, 217-246, http://doi.org/10.1007/978-90-481-3830-2_5, 2013.

Miller, S. T. K., Keim, B. D., Talbot, R. W., and Mao, H.: Sea breeze: Structure, forecasting, and impacts, Rev. Geophys., 41,

765   https://doi.org/10.1029/2003rg000124, 2003.

Ministry of Ecology and Environment of the People's Republic of China: National Urban Air Quality Report for August 2024,
        https://www.mee.gov.cn/hjzl/dqhj/cskqzlzkyb/202409/W020240924585303726990.pdf, 2024a.

Ministry of Ecology and Environment of the People's Republic of China: National Urban Air Quality Report for July 2024,
        https://www.mee.gov.cn/hjzl/dqhj/cskqzlzkyb/202408/W020240823817374806303.pdf, 2024b.

770   Ministry of Ecology and Environment of the People's Republic of China: National Urban Air Quality Report for June 2024,
        https://www.mee.gov.cn/hjzl/dqhj/cskqzlzkyb/202407/W020240725571411086492.pdf, 2024c.

Moortgat, G. K., Grossmann, D., Boddenberg, A., Dallmann, G., Ligon, A. P., Turner, W. V., Gäb, S., Slemr, F., Wieprecht, W.,
        Acker, K., Kibler, M., Schlomski, S., and Bächmann, K.: Hydrogen Peroxide, Organic Peroxides and Higher Carbonyl
        Compounds   Determined   during   the   BERLIOZ   Campaign,   J.   Atmos.   Chem.,   42,   443-463,
775   https://doi.org/10.1023/A:1015743013107, 2002.

Parker, D. J.: MESOSCALE METEOROLOGY | Overview, in: Encyclopedia of Atmospheric Sciences (Second Edition),
        edited by: North, G. R., Pyle, J., and Zhang, F., Academic Press, Oxford, 316-322, http://doi.org/10.1016/B978-0-12-
        382225-3.00478-3, 2015.

Peengam, S., Pariyothon, J., Chanalert, W., Laiwarin, P., and Pattarapanitchai, S.: Characteristics of aerosol optical depth and
780   Ångström exponent indicated aerosol types over a tropical urban and coastal areas in Thailand, Air Quality, Atmosphere
        & Health, https://doi.org/10.1007/s11869-025-01853-x, 2025.

Peng, S., Zheng, Y., Li, L., Gui, K., Zhu, J., Liu, S., Zhang, H., Zhao, H., Che, H., and Zhang, X.: Long-term variations in
        aerosol optical properties in Beijing: insights from diurnal and nocturnal continuous measurements, Atmos. Environ., 360,
        121416, https://doi.org/10.1016/j.atmosenv.2025.121416, 2025.

785   Puygrenier, V., Lohou, F., Campistron, B., Saïd, F., Pigeon, G., Bénech, B., and Serça, D.: Investigation on the fine structure
        of sea-breeze during ESCOMPTE experiment, Atmos. Res., 74, 329-353, https://doi.org/10.1016/j.atmosres.2004.06.011,
        2005.

Qu, K., Wang, X., Yan, Y., Shen, J., Xiao, T., Dong, H., Zeng, L., and Zhang, Y.: A comparative study to reveal the influence
        of typhoons on the transport, production and accumulation of O_3 in the Pearl River Delta, China, Atmos. Chem. Phys.,

790   21, 11593-11612, https://doi.org/10.5194/acp-21-11593-2021, 2021.

Ren, J., Guo, F., and Xie, S.: Diagnosing ozone–NO_x–VOC sensitivity and revealing causes of ozone increases in China based on 2013–2021 satellite retrievals, Atmos. Chem. Phys., 22, 15035-15047, https://doi.org/10.5194/acp-22-15035-2022, 2022.

Roux, F., Clark, H., Wang, K. Y., Rohs, S., Sauvage, B., and Nédélec, P.: The influence of typhoons on atmospheric composition
795   deduced from IAGOS measurements over Taipei, Atmos. Chem. Phys., 20, 3945-3963, https://doi.org/10.5194/acp-20-3945-2020, 2020.

Rozanov, V. V., Buchwitz, M., Eichmann, K. U., de Beek, R., and Burrows, J. P.: Sciatran - a new radiative transfer model for geophysical applications in the 240–2400 NM spectral region: the pseudo-spherical version, Adv. Space Res., 29, 1831-1835, https://doi.org/10.1016/S0273-1177(02)00095-9, 2002.

800   Sander, R.: Compilation of Henry's law constants (version 4.0) for water as solvent, Atmos. Chem. Phys., 15, 4399-4981, https://doi.org/10.5194/acp-15-4399-2015, 2015.

Schreier, S. F., Richter, A., Peters, E., Ostendorf, M., Schmalwieser, A. W., Weihs, P., and Burrows, J. P.: Dual ground-based MAX-DOAS observations in Vienna, Austria: Evaluation of horizontal and temporal NO2, HCHO, and CHOCHO distributions and comparison with independent data sets, Atmos. Environ.: X, 5, 100059,
805   https://doi.org/10.1016/j.aeaoa.2019.100059, 2020.

Schreier, S. F., Richter, A., Wittrock, F., and Burrows, J. P.: Estimates of free-tropospheric NO2 and HCHO mixing ratios derived from high-altitude mountain MAX-DOAS observations at midlatitudes and in the tropics, Atmos. Chem. Phys., 16, 2803-2817, https://doi.org/10.5194/acp-16-2803-2016, 2016.

Seinfeld, J. H. and Pandis, S. N.: Atmospheric Chemistry and Physics: From Air Pollution to Climate Change, 3rd, Wiley,
810   2016.

Song, X., Li, X.-B., Yuan, B., He, X., Chen, Y., Wang, S., Huangfu, Y., Peng, Y., Zhang, C., Liu, A., Yang, H., Liu, C., Li, J., and Shao, M.: Elucidating key factors in regulating budgets of ozone and its precursors in atmospheric boundary layer, npj Clim. Atmos. Sci., 7, https://doi.org/10.1038/s41612-024-00818-8, 2024.

Talbot, C., Augustin, P., Leroy, C., Willart, V., Delbarre, H., and Khomenko, G.: Impact of a sea breeze on the boundary-layer

dynamics and the atmospheric stratification in a coastal area of the North Sea, Boundary Layer Meteorol., 125, 133-154, https://doi.org/10.1007/s10546-007-9185-6, 2007.

Vo, T.-D.-H., Lin, C., Weng, C.-E., Yuan, C. S., Lee, C.-W., Hung, C.-H., Xuan-Thanh, B., Lo, K.-C., and Lin, J.-X.: Vertical stratification of volatile organic compounds and their photochemical product formation potential in an industrial urban area, J. Environ. Manage., 217, 327-336, https://doi.org/10.1016/j.jenvman.2018.03.101, 2018.

Vrekoussis, M., Wittrock, F., Richter, A., and Burrows, J. P.: GOME-2 observations of oxygenated VOCs: what can we learn from the ratio glyoxal to formaldehyde on a global scale?, Atmos. Chem. Phys., 10, 10145-10160, https://doi.org/10.5194/acp-10-10145-2010, 2010.

Wagner, T., Beirle, S., Brauers, T., Deutschmann, T., Friess, U., Hak, C., Halla, J. D., Heue, K. P., Junkermann, W., Li, X., Platt, U., and Pundt-Gruber, I.: Inversion of tropospheric profiles of aerosol extinction and HCHO and NO_2 mixing ratios from MAX-DOAS observations in Milano during the summer of 2003 and comparison with independent data sets, Atmos. Meas. Tech., 4, 2685-2715, https://doi.org/10.5194/amt-4-2685-2011, 2011.

Wang, J., Li, O., Zhang, P., Yang, X., Yadav, R., Chen, S., Liu, Y., Zhu, C., Ren, Y., and Mellouki, A.: Atmospheric chemistry of the coastal area is influenced by the convergence between the inland and marine air: Insight into carbonyl compounds, J. Environ. Sci., 154, 859-870, https://doi.org/10.1016/j.jes.2024.06.019, 2025.

Wang, J., Wang, P., Tian, C., Gao, M., Cheng, T., and Mei, W.: Consecutive Northward Super Typhoons Induced Extreme Ozone Pollution Events in Eastern China, npj Clim. Atmos. Sci., 7, https://doi.org/10.1038/s41612-024-00786-z, 2024.

Wang, W., van der A, R., Ding, J., van Weele, M., and Cheng, T.: Spatial and temporal changes of the ozone sensitivity in China based on satellite and ground-based observations, Atmos. Chem. Phys., 21, 7253-7269, https://doi.org/10.5194/acp-21-7253-2021, 2021.

Wang, Z., Zhang, H., Shi, C., Ji, X., Zhu, Y., Xia, C., Sun, X., Zhang, M., Lin, X., Yan, S., Zhou, Y., Xing, C., Chen, Y., and Liu, C.: Vertical and spatial differences in ozone formation sensitivities under different ozone pollution levels in eastern Chinese cities, npj Clim. Atmos. Sci., 8, https://doi.org/10.1038/s41612-024-00855-3, 2025.

Wittrock, F., Richter, A., Oetjen, H., Burrows, J. P., Kanakidou, M., Myriokefalitakis, S., Volkamer, R., Beirle, S., Platt, U., and Wagner, T.: Simultaneous global observations of glyoxal and formaldehyde from space, Geophys. Res. Lett., 33,

840   https://doi.org/10.1029/2006gl026310, 2006.

Waxman, E. M., Elm, J., Kurtén, T., Mikkelsen, K. V., Ziemann, P. J., and Volkamer, R.: Glyoxal and Methylglyoxal Setschenow Salting Constants in Sulfate, Nitrate, and Chloride Solutions: Measurements and Gibbs Energies, Environ. Sci. Technol., 49, 11500-11508, https://doi.org/10.1021/acs.est.5b02782, 2015.

Wolfe, G. M., Nicely, J. M., St. Clair, J. M., Hanisco, T. F., Liao, J., Oman, L. D., Brune, W. B., Miller, D., Thames, A.,
845   González Abad, G., Ryerson, T. B., Thompson, C. R., Peischl, J., McKain, K., Sweeney, C., Wennberg, P. O., Kim, M., Crounse, J. D., Hall, S. R., Ullmann, K., Diskin, G., Bui, P., Chang, C., and Dean-Day, J.: Mapping hydroxyl variability throughout the global remote troposphere via synthesis of airborne and satellite formaldehyde observations, Proceedings of the National Academy of Sciences, 116, 11171-11180, https://doi.org/10.1073/pnas.1821661116, 2019.

Wu, M., Wu, D., Fan, Q., Wang, B. M., Li, H. W., and Fan, S. J.: Observational studies of the meteorological characteristics
850   associated with poor air quality over the Pearl River Delta in China, Atmos. Chem. Phys., 13, 10755-10766, https://doi.org/10.5194/acp-13-10755-2013, 2013.

Xin, J., Ma, Y., Zhao, D., Gong, C., Ren, X., Tang, G., Xia, X., Wang, Z., Cao, J., de Arellano, J. V.-G., and Martin, S. T.: The feedback effects of aerosols from different sources on the urban boundary layer in Beijing China, Environ. Pollut., 325, 121440, https://doi.org/10.1016/j.envpol.2023.121440, 2023.

855   Xing, C., Liu, C., Hu, Q., Fu, Q., Lin, H., Wang, S., Su, W., Wang, W., Javed, Z., and Liu, J.: Identifying the wintertime sources of volatile organic compounds (VOCs) from MAX-DOAS measured formaldehyde and glyoxal in Chongqing, southwest China, Sci. Total Environ., 715, 136258, https://doi.org/10.1016/j.scitotenv.2019.136258, 2020.

Xu, J., Huang, X., Wang, N., Li, Y., and Ding, A.: Understanding ozone pollution in the Yangtze River Delta of eastern China from the perspective of diurnal cycles, Sci. Total Environ., 752, 141928, https://doi.org/10.1016/j.scitotenv.2020.141928,
860   2021.

Xu, W., Xing, Q., Pan, L., Wang, Z., Cao, X., Yan, W., Xie, W., Meng, X., and Wu, X.: Characterization, source apportionment, and risk assessment of ambient volatile organic compounds in urban and background regions of Hainan Island, China, Atmos. Environ., 316, 120167, https://doi.org/10.1016/j.atmosenv.2023.120167, 2024.

Yan, H. and Anthes, R. A.: THE EFFECT OF VARIATIONS IN SURFACE MOISTURE ON MESOSCALE CIRCULATIONS,

865   Mon. Weather Rev., 116, 192-208, https://doi.org/10.1175/1520-0493(1988)116, 1988.

Yang, J. X., Lau, A. K. H., Fung, J. C. H., Zhou, W., and Wenig, M.: An air pollution episode and its formation mechanism during the tropical cyclone Nuri's landfall in a coastal city of south China, Atmos. Environ., 54, 746-753, https://doi.org/10.1016/j.atmosenv.2011.12.023, 2012.

Yang, Y., Zhou, Y., Wang, H., Li, M., Li, H., Wang, P., Yue, X., Li, K., Zhu, J., and Liao, H.: Meteorological characteristics of
870   extreme ozone pollution events in China and their future predictions, Atmos. Chem. Phys., 24, 1177-1191, https://doi.org/10.5194/acp-24-1177-2024, 2024.

Yu, Y., Zhou, H., Zhao, Z., Chang, Y., Wu, D., Li, Z., Wang, F., Fang, M., and Zhou, X.: Spatiotemporal Distribution, Meteorological Influence, and Potential Sources of Air Pollution over Hainan Island, China, Atmosphere, 15, 1336, https://www.mdpi.com/2073-4433/15/11/1336, 2024.

875 Zhan, J., Zheng, F., Xie, R., Liu, J., Chu, B., Ma, J., Xie, D., Meng, X., Huang, Q., He, H., and Liu, Y.: The role of NO_x in Co-occurrence of O_3 and $PM_{2.5}$ pollution driven by wintertime east Asian monsoon in Hainan, J. Environ. Manage., 345, 118645, https://doi.org/10.1016/j.jenvman.2023.118645, 2023.

Zhang, R., Wang, S., Zhang, S., Xue, R., Zhu, J., and Zhou, B.: MAX-DOAS observation in the midlatitude marine boundary layer: Influences of typhoon forced air mass, J. Environ. Sci., 120, 63-73, https://doi.org/10.1016/j.jes.2021.12.010, 2022.

880 Zhang, S., Wang, S., Zhang, R., Guo, Y., Yan, Y., Ding, Z., and Zhou, B.: Investigating the Sources of Formaldehyde and Corresponding Photochemical Indications at a Suburb Site in Shanghai From MAX-DOAS Measurements, J. Geophys. Res.: Atmos., 126, https://doi.org/10.1029/2020jd033351, 2021.

Zhang, Y., Man, X., Zhang, S., Liu, L., Kong, F., Feng, T., and Liu, R.: Ground-based MAX-DOAS observations of formaldehyde and glyoxal in Xishuangbanna, China, J. Environ. Sci., 152, 328-339,
885   https://doi.org/10.1016/j.jes.2024.04.036, 2025.

Zhang, Z.-z., Cao, C.-x., Song, Y., Kang, L., and Cai, X.-h.: STATISTICAL CHARACTERISTICS AND NUMERICAL SIMULATION OF SEA-LAND BREEZES IN HAINAN ISLAND, J. Trop. Meteorol., 20, 267-278, https://jtm.itmm.org.cn/en/article/id/20140309, 2014.

Zhao, D., Xin, J., Wang, W., Jia, D., Wang, Z., Xiao, H., Liu, C., Zhou, J., Tong, L., Ma, Y., Wen, T., Wu, F., and Wang, L.:

890   Effects of the sea-land breeze on coastal ozone pollution in the Yangtze River Delta, China, Sci. Total Environ., 807, https://doi.org/10.1016/j.scitotenv.2021.150306, 2022.

Zhao, M., Shen, H., Zhang, J., Liu, Y., Sun, Y., Wang, X., Dong, C., Zhu, Y., Li, H., Shan, Y., Mu, J., Zhong, X., Tang, J., Guo, M., Wang, W., and Xue, L.: Carbonyl Compounds Regulate Atmospheric Oxidation Capacity and Particulate Sulfur Chemistry in the Coastal Atmosphere, Environ. Sci. Technol., 58, 17334-17343, https://doi.org/10.1021/acs.est.4c03947, 895   2024.

Zhu, L., Meng, Z., Zhang, F., and Markowski, P. M.: The influence of sea- and land-breeze circulations on the diurnal variability in precipitation over a tropical island, Atmos. Chem. Phys., 17, 13213-13232, https://doi.org/10.5194/acp-17-13213-2017, 2017.

Zhu, M., Zhu, D., Huang, M., Gong, D., Li, S., Xia, Y., Lin, H., and Altan, O.: Assessing the Impact of Climate Change on the 900   Landscape Stability in the Mediterranean World Heritage Site Based on Multi-Sourced Remote Sensing Data: A Case Study of the Causses and Cévennes, France, Remote Sens., 17, 203, https://www.mdpi.com/2072-4292/17/2/203, 2025.