



1 **Atmospheric chemical processing dictates aerosol aluminum solubility: insights from**
2 **field measurement at two locations in northern China**

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23 **Abstract**

24 Deposition of mineral dust aerosol into open oceans significantly impacts marine
25 biogeochemistry and primary production, and the deposition rates can be constrained using
26 widely measured dissolved aluminum (Al) in surface seawater as a tracer. However, aerosol
27 Al solubility, a critical parameter used in this method, remains highly uncertain. This work
28 investigated seasonal variations of aerosol Al solubility for supermicron and submicron
29 particles at two locations (Xi'an and Qingdao) in northern China. Aerosol Al solubility was
30 found to be very low at Xi'an and much higher at Qingdao. Furthermore, seasonal variability
31 of Al solubility, its correlation with relative abundance of sulfate and nitrate, and its
32 dependence on relative humidity (RH), are all different at the two locations. We suggest that
33 all the features observed for aerosol Al solubility at the two locations can be well explained by
34 the effects of atmospheric chemical processing. Mineral dust transported to Xi'an (an inland
35 city in Northwest China) was still not significantly aged and thus chemical processing had little
36 effects on aerosol Al solubility. After arriving at Qingdao (a coastal city in the Northwest
37 Pacific), mineral dust was substantially aged by chemical processing, leading to substantial
38 enhancement in aerosol Al solubility. Our work further reveals that aerosol liquid water and
39 acidity play vital roles in the dissolution of aerosol Al by atmospheric chemical processing.
40 We suggest that spatial variation of aerosol Al solubility should be taken into account so that
41 oceanic dust deposition can be better constrained using dissolved Al concentrations in surface
42 seawater.



43 **1. Introduction**

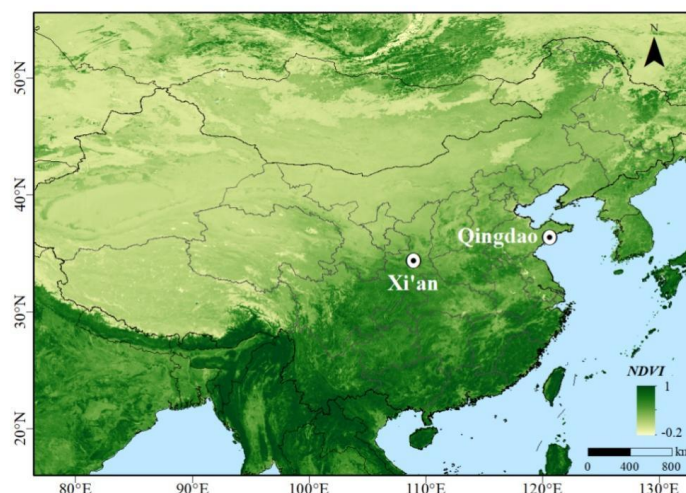
44 As an important type of tropospheric aerosols, mineral dust aerosol significantly impacts
45 atmosphere chemistry, climate, and ecological systems (Jickells et al., 2005; Tang et al., 2016;
46 Kok et al., 2023). After long-range transport, deposition of mineral dust into the oceans is a
47 major external source of several nutrient and toxic elements for surface seawater (Moore et al.,
48 2013; Westberry et al., 2023), impacting primary production and biogeochemical cycles in the
49 oceans and having further feedback on the climate system (Mahowald, 2011; Jiang et al., 2024).
50 The deposition flux of desert dust aerosol into the oceans should be accurately estimated before
51 we can assess its impacts on marine biogeochemistry in a reliable manner (Schulz et al., 2012;
52 Anderson et al., 2016). Previous studies used several different methods to estimate dust
53 deposition fluxes and found large discrepancies (Huneeus et al., 2011; Anderson et al., 2016).

54 Deposition of desert dust aerosol is the dominant source of dissolved aluminum (Al) in
55 the surface water of open oceans, and dissolved Al is generally considered to be chemically
56 and biologically inactive in seawater. As a result, dissolved Al concentrations in surface
57 seawater could be used to calculate dust deposition flux into the oceans (Measures and Brown,
58 1996; Measures and Vink, 2000), and the fractional solubility of aerosol Al (the fraction of
59 aerosol Al that can be dissolved) is one of the key parameters used in this method. Previous
60 studies which used this method to estimate dust depositions fluxes (Han et al., 2008; Measures
61 et al., 2010; Grand et al., 2015; Benaltabet et al., 2022) usually assumed uniform Al solubility
62 values in the range of 1.5-5%. However, field measurements found that aerosol Al solubility
63 could vary by more than an order of magnitude (Baker et al., 2006; Buck et al., 2013), and
64 thereby using a uniform aerosol Al solubility value could lead to large uncertainties in



65 estimated dust deposition fluxes (Han et al., 2008; Xu and Weber, 2021). As a result, it is
66 important to understand the spatiotemporal variations of aerosol Al solubility and elucidate the
67 processes and mechanisms which drive such variations.

68 The initial Al solubility is generally low (typically $<1.5\%$) for dust particles over desert
69 regions (Shi et al., 2011; Agnati et al., 2014; Li et al., 2022), and field studies found that
70 aerosol Al solubility in the troposphere could be much higher and showed wide variability. For
71 example, Al solubility ranged from 0.2-15.9% for total suspended particles (TSP) over the
72 Pacific (Buck et al., 2013), and were in the range of 3-78% over the Atlantic (Buck et al., 2010;
73 Chance et al., 2015). Some studies (Measures et al., 2010; Sakata et al., 2023) found good
74 correlations between dissolved aerosol Al (or Al solubility) and acid species in aerosol particles,
75 and thus suggested that chemical processes in the atmosphere could substantially enhance
76 aerosol Al solubility; furthermore, Li et al. (2017) found that Al solubility was significantly
77 increased during cloud events when cloud processing enhanced the formation of secondary
78 inorganic ions (mainly sulfate and nitrate) and thus increased the acidity of cloud droplets.
79 However, Yang et al. (2023) found no correlations between Al solubility and the concentrations
80 of aerosol acidic species, and concluded that the effect of acid processing on Al solubility was
81 negligible. Aerosol Al solubility over the Atlantic appeared to be higher for air masses from
82 Europe than those from the Saharan region (Baker et al., 2006; López-García et al., 2017), and
83 some studies (Paris et al., 2010; López-García et al., 2017) hypothesized that this could be
84 potentially explained by the influence of anthropogenic aerosol Al if it had higher solubility
85 than desert dust. In summary, although aerosol Al solubility in the atmosphere was explored by
86 previous studies, the governing processes and environmental factors remain poorly understood.



87

88 **Figure 1.** A map of East Asia and surrounding areas. The two locations (Xi'an and Qingdao)
89 where we collected aerosol particles are highlighted. NDVI: normalized difference vegetation
90 index provided by MODIS (Moderate Resolution Imaging Spectroradiometer).

91

92 In this work, we collected supermicron (above 1 μm) and submicron (below 1 μm) aerosol
93 particles at Xi'an and Qingdao, both located in northern China, and investigated seasonal
94 variations of aerosol Al solubility at these two locations. Taklimakan and Gobi Deserts in
95 northwestern China are two important source regions of Asian dust (Prospero et al., 2002). As
96 shown in Figure 1, Xi'an is an inland city in northwestern China, and the aging extent of dust
97 was found to be quite limited at Xi'an due to its proximity of desert regions (Wang et al., 2014;
98 Wu et al., 2017). As Asian dust is transported eastward, it passes over the North China Plain
99 where anthropogenic emission is very high, and may become much more aged when arriving
100 at Qingdao, a coastal city of the Northwest Pacific (Li et al., 2014; Pan et al., 2017). By
101 comparing aerosol Al solubility at Xi'an and Qingdao, our work can provide valuable insights



102 into how and to which extent aging processes during long-range transport can change aerosol
103 Al solubility. Dust aerosol concentrations and meteorological conditions vary remarkably at
104 different seasons in Northern China; as a result, examining its seasonal variations provides a
105 good opportunity to understand the factors which regulate aerosol Al solubility.

106 **2. Materials and methods**

107 **2.1 Sample collection**

108 Samples were collected at two cities (Xi'an and Qingdao) in northern China at four
109 different seasons during 2021-2023 (Zhang et al., 2023; Chen et al., 2024), and further details
110 can be found in the supplement (Text S1 and Table S1). In brief, supermicron ($>1\ \mu\text{m}$) and
111 submicron ($<1\ \mu\text{m}$) particles were simultaneously collected using a two-stage aerosol sampler
112 (TH-150C, Tianhong Co., China) which was operated at 100 L/min, and the sampling duration
113 was typically 23.5 hours for each pair of aerosol samples. Whatman 41 cellulose filters were
114 used for aerosol collection in our work, and they were acid-washed before being used for
115 aerosol sampling to reduce background levels (Zhang et al., 2022). A total of 126 and 106 pairs
116 of aerosol samples were collected at Xi'an and Qingdao, respectively (Zhang et al., 2023; Chen
117 et al., 2024). After collection, all the aerosol samples were stored at -20°C for further analysis.

118 In addition to aerosol particles, we also sampled atmospheric acidic and alkaline gases
119 (mainly NH_3 , HCl and HNO_3) at Qingdao, using a ChemComb 3500 Speciation Collection
120 Cartridge (Thermo Fisher Scientific, USA) at a flow rate of 10 L/min (Walters and Hastings,
121 2018; Fang et al., 2025). Gas sampling was carried out concurrently with aerosol sampling. In
122 brief, NH_3 , HNO_3 and HCl were absorbed onto the inner walls of two tandem honeycomb
123 diffusion tubes coated with proper adsorbents, and then converted into NH_4^+ , NO_3^- and Cl^- .



124 After the sampling was completed, 20 mL ultrapure water was used to rinse each tube
125 immediately, and a PTFE membrane syringe filter (0.22 μm in pore size) was used to filter the
126 solution. The solution was then frozen at -20°C for further analysis.

127 2.2 Sample analysis and aerosol acidity calculation

128 Sample pretreatment and analysis were detailed in our previous work (Zhang et al., 2022),
129 and therefore are only briefly summarized here. The first half of a filter (and only one quarter
130 of a filter for supermicron particles) was shredded and digested in a Teflon jar using a
131 microwave digestion instrument. After digestion, the Teflon jar was filled with 1% HNO_3 (20
132 mL), and a PTFE membrane syringe filter (0.22 μm in pore size) was used to filter the solution;
133 subsequently, the solution was analyzed by inductively coupled plasma-mass spectrometry
134 (ICP-MS) to determine total concentrations of individual trace elements, including Al.

135 The other half of a filter was immersed in ultrapure water (20 mL) and stirred using an
136 orbital shaking for two hours; in the next step, the solution was filtered using a PTFE membrane
137 syringe filter (0.22 μm in pore size) and divided into two parts. The first solution was acidified
138 to contain 1% HNO_3 and subsequently analyzed by ICP-MS to determine the concentrations
139 of dissolved trace elements; the second solution was analyzed by ion chromatography (IC) to
140 quantify the concentration of water-soluble cations and anions.

141 The solutions obtained from honeycomb diffusion tubes (see Section 2.1 for more details)
142 were also analyzed using IC to determine the concentrations of gaseous NH_3 , HCl and HNO_3
143 in the atmosphere. ISORROPIA-II, a widely used aerosol thermodynamic model (Fountoukis
144 and Nenes, 2007), was employed in this work to calculate the acidity of supermicron and
145 submicron particles. It was operated in the forward mode, and aerosol particles were assumed



146 to remain metastable. Input parameters included concentrations of water-soluble ions in aerosol
147 particles and gaseous NH_3 , HCl and HNO_3 , temperature and relative humidity (RH). Our
148 previous work (Fang et al., 2025) found good agreement between measured and calculated NH_3
149 partitioning coefficients at Qingdao, and as a result the method we used could well estimate
150 the acidity of supermicron and submicron particles.

151 **3. Results**

152 **3.1 Seasonal variations of total and dissolved aerosol Al**

153 **3.1.1 Total aerosol Al**

154 Figure 2 displays seasonal variations of total and dissolved aerosol Al at Xi'an and
155 Qingdao. At Xi'an (Figure 2a), total Al in supermicron particles showed highest concentrations
156 in spring and winter (1.54 ± 0.89 and $1.91 \pm 0.93 \mu\text{g}/\text{m}^3$) and lowest concentrations in summer
157 ($0.96 \pm 0.54 \mu\text{g}/\text{m}^3$); a similar seasonal pattern was observed for submicron particles, with total
158 Al concentrations being highest in spring and winter (4.29 ± 3.70 and $2.92 \pm 1.47 \mu\text{g}/\text{m}^3$) and
159 lowest in summer ($0.95 \pm 0.44 \mu\text{g}/\text{m}^3$). At Qingdao (Figure 2b), total Al concentrations in
160 supermicron particles were highest in spring ($1.04 \pm 1.12 \mu\text{g}/\text{m}^3$) and lowest in summer and
161 autumn (0.33 ± 0.18 and $0.31 \pm 0.12 \mu\text{g}/\text{m}^3$); similarly, for submicron particles, total Al
162 concentrations were also highest in spring ($1.88 \pm 2.51 \mu\text{g}/\text{m}^3$) and lowest in summer and
163 autumn (0.35 ± 0.22 and $0.65 \pm 0.82 \mu\text{g}/\text{m}^3$).

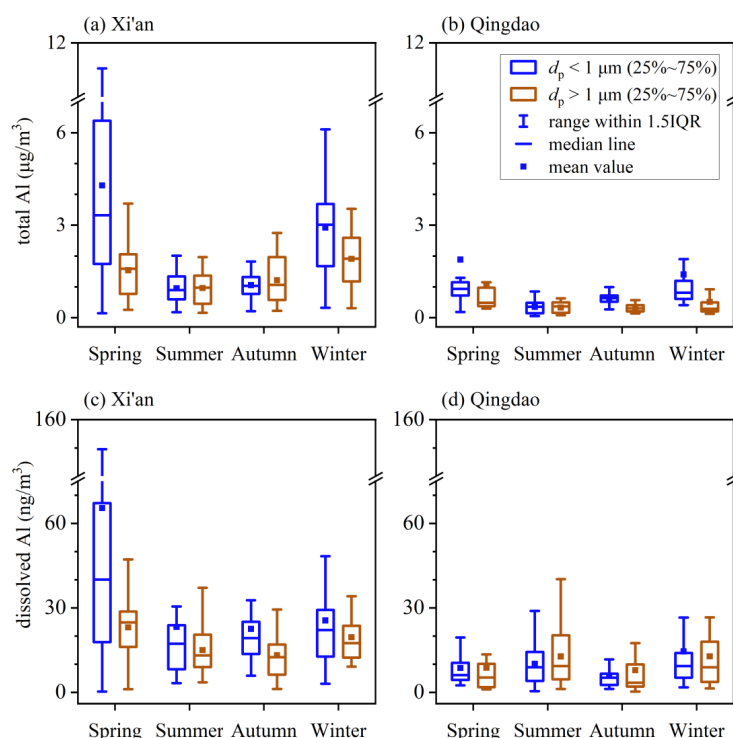


Figure 2. Seasonal variations of total and dissolved aerosol Al for submicron and supermicron particles: (a) total Al at Xi'an; (b) total Al at Qingdao; (c) dissolved Al at Xi'an; (d) dissolved Al at Qingdao.

Overall, total aerosol Al concentrations showed similar seasonal variations at Xi'an and Qingdao, being highest in spring and lowest in summer. This was consistent with previous studies carried out in other locations in East Asia, such as Huaniao Island in the East China Sea (Guo et al., 2014), northern Taiwan (Hsu et al., 2008), Hong Kong (Yang et al., 2023), and Japan (Sakata et al., 2023). In East Asia, desert dust aerosol was emitted into the atmosphere mainly in spring, leading to the increase in total aerosol Al concentrations. Lowest concentrations of total aerosol Al were observed in summer because precipitation in Northern



176 China mainly occurred in summer, leading to enhanced wet deposition of aerosol particles (Cao
177 and Cui, 2021). Furthermore, Qingdao was frequently affected by marine air masses in summer,
178 and this is also one reason why total aerosol Al concentrations were lower in summer than
179 other seasons.

180 As summarized in the supplement (Table S2), total aerosol Al concentrations exhibited
181 evident spatial variations in East Asia. As Asian dust was transported eastward to the North
182 Pacific, a clear decrease in aerosol Al concentrations was observed. Desert dust was the
183 dominant source for aerosol Al, and therefore concentrations of aerosol Al were found to be
184 very high in desert regions. For example, total Al concentrations in TSP could reach $24 \mu\text{g}/\text{m}^3$
185 over the Taklimakan Desert (Zhang et al., 2003). In our current study, annual average total Al
186 concentrations at Xi'an, an inland city close to the desert, were reported to be 1.42 ± 0.86 and
187 $2.28 \pm 2.35 \mu\text{g}/\text{m}^3$ for supermicron and submicron particles, significantly lower than that
188 observed over the Taklimakan Desert. Further decrease in total Al concentrations was observed
189 in coastal and oceanic regions. For example, our work found that the annual average total Al
190 concentrations were 0.56 ± 0.75 and $1.08 \pm 1.67 \mu\text{g}/\text{m}^3$ for supermicron and submicron particles
191 at Qingdao, lower than those at Xi'an; total Al concentrations in TSP ranged from 0.17 to 1.72
192 $\mu\text{g}/\text{m}^3$ in Hiroshima (Sakata et al., 2023), and further decreased to $1\text{--}56 \text{ ng}/\text{m}^3$ in Hawaii in the
193 central Pacific (Measures et al., 2010).

194 3.1.2 Dissolved aerosol Al

195 At Xi'an (Figure 2c), for supermicron particles, dissolved aerosol Al concentrations were
196 highest in spring ($23.1 \pm 10.9 \text{ ng}/\text{m}^3$) and lowest in summer and autumn (15.0 ± 8.7 and 13.2 ± 8.6
197 ng/m^3); for submicron particles, dissolved Al concentrations were also highest in spring



198 (65.4±79.2 ng/m³) and lowest in summer and autumn (23.2±23.4 and 22.6±20.1 ng/m³). Total
199 (Figure 2a) and dissolved aerosol Al (Figure 2c) showed similar seasonal patterns at Xi'an,
200 indicating that dissolved aerosol Al was mainly regulated by total aerosol Al.

201 As shown in Figure 2d, the average dissolved aerosol Al concentrations were 8.8±10.8,
202 12.8±11.1, 7.9±10.5 and 12.8±12.9 ng/m³ for supermicron particles at Qingdao in spring,
203 summer, autumn, and winter, respectively, and 8.7±5.8, 10.2±8.2, 6.0±4.8 and 14.5±15.2 ng/m³
204 for submicron particles. Dissolved aerosol Al concentrations were highest in summer and
205 winter and lowest in autumn for both supermicron and submicron particles. In contrast to Xi'an,
206 total and dissolved aerosol Al at Qingdao showed different seasonal patterns (Figures 2b and
207 2d); for example, total Al concentrations were lowest in summer at Qingdao when dissolved
208 Al concentrations were highest. This indicates that dissolved aerosol Al at Qingdao was not
209 only regulated by total aerosol Al but also affected by other factors such as atmospheric aging
210 processes.

211 Compared to Xi'an, dissolved Al concentrations at Qingdao were lower across all the four
212 seasons, mainly because total Al concentrations were much lower at Qingdao (Tables S3-S4 in
213 the supplement). As shown in Figure 2, similar seasonal patterns were observed at two
214 locations for total aerosol Al, but dissolved aerosol Al showed very different seasonality; this
215 suggests that seasonal patterns of aerosol Al solubility were different at Xi'an and Qingdao, as
216 presented in Section 3.2.

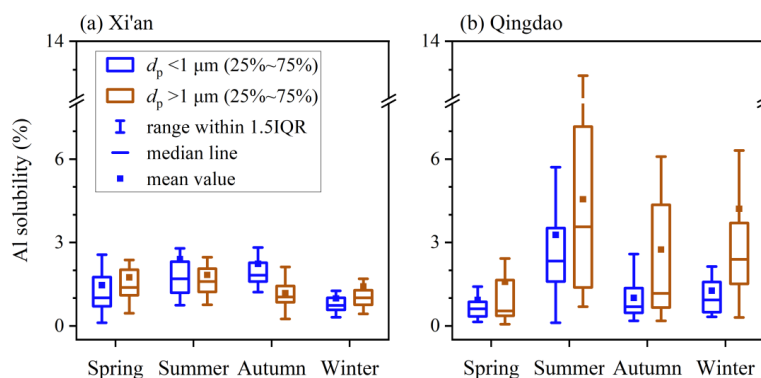
217 **3.2 Fractional solubility of aerosol Al**

218 **3.2.1 Seasonal variations of Al solubility**

219 Figure 3 displays aerosol Al solubility in different seasons at Xi'an and Qingdao. The



220 median solubilities of aerosol Al were determined to be 1.38%, 1.59%, 1.04% and 1.01% for
221 supermicron particles at Xi'an in spring, summer, autumn and winter, respectively, and 1.01%,
222 1.69%, 1.82% and 0.74% for submicron particles. Aerosol Al solubilities were generally low
223 for the four seasons at Xi'an, showing no apparent variation with seasons (Figure 3a). In
224 contrast, aerosol Al solubilities exhibited distinct seasonal variability at Qingdao (Figure 3b),
225 and the median Al solubilities were highest in summer (3.56% and 2.33%) and lowest in spring
226 (0.54% and 0.61%) for both supermicron and submicron particles.



227 **Figure 3.** Seasonal variations of aerosol Al solubility for submicron and supermicron particles
228 at (a) Xi'an and (b) Qingdao.
229

230

231 In three seasons (summer, autumn and winter), aerosol Al solubility at Qingdao was
232 significantly higher than that at Xi'an (Figure 3, Table S5). This is likely because Xi'an is close
233 to major deserts in northwestern China and thus the aging extent of mineral dust transported to
234 Xi'an was rather limited (Wang et al., 2014; Wu et al., 2017). On the contrary, Qingdao is much
235 farther from deserts and consequently mineral dust aerosol which arrived at Qingdao after long-
236 distance transport was substantially aged (Trochikine et al., 2003; Takahashi et al., 2011; Jeong,



237 2020), thereby leading to enhanced dissolution of aerosol Al and thus the increase in Al
238 solubility.

239 On the other hand, no significant difference in aerosol Al solubility was observed between
240 Xi'an and Qingdao in spring, with median aerosol Al solubilities being <1.4% for supermicron
241 and submicron particles (Figure 3). This agrees with a previous study (Hsu et al., 2010) which
242 found that aerosol Al solubility was very low (average: ~0.7%) in spring even over the East
243 China Sea. Furthermore, similar to what we observed in spring at Xi'an and Qingdao, Al
244 solubility was found to be low (<1.5%) for surface soil particles collected from deserts (Shi et
245 al., 2011; Aghnatiev et al., 2014; Li et al., 2022). Overall, our work implies that in spring when
246 Asian dust occurred most frequently, mineral dust particles arriving at Qingdao after long-
247 distance transport did not show significant increase in Al solubility.

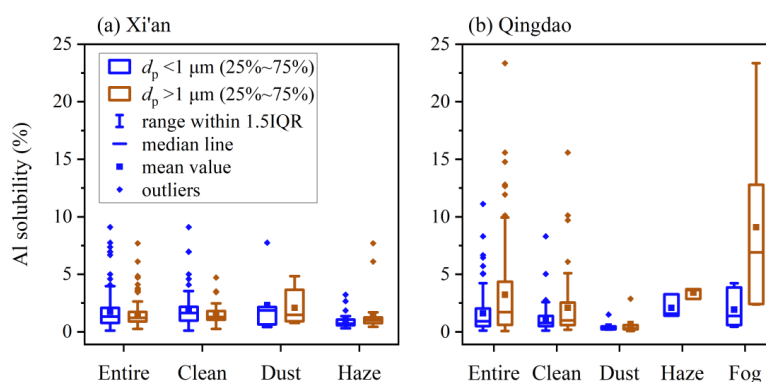
248 **3.2.2 Al solubility under different weather conditions**

249 We encountered four representative weather conditions (i.e. clean, dust, haze and fog days)
250 during our sampling at Xi'an and Qingdao, and investigated aerosol Al solubility under
251 different weather conditions (Figure 4, Tables S6-S7).

252 At Xi'an, no significant difference in Al solubility was observed during clean, haze, and
253 dust days (Figure 4a, Table S6), with median values in the range of 1.01-1.47% for supermicron
254 particles and 0.72-1.86% for submicron particles. Al solubility was found to be <1.2% for three
255 desert dust samples (Luochuan loess, Arizona test dust, and dust collected during a dust storm
256 in Xinjiang) (Li et al., 2022), and ranged from 0.47% to 1.42% for aerosol particles generated
257 using soil samples from Saharan desert (Shi et al., 2011). Compared to desert dust in source
258 regions, Al solubility was not higher under different weather conditions at Xi'an. In addition,



259 although emission and accumulation of anthropogenic pollutants was greatly enhanced during
260 haze days at Xi'an (An et al., 2019; Cao and Cui, 2021), there was no obvious increase in
261 aerosol Al solubility, indicating that the effects of anthropogenic emissions on aerosol Al
262 solubility was limited at Xi'an. Therefore, one may conclude that aerosol Al solubility at Xi'an
263 was not different from initial Al solubility of desert dust.



264 **Figure 4.** Aerosol Al solubility under different weather conditions for submicron and
265 supermicron particles: (a) Xi'an, (b) Qingdao.

267
268 Being different to Xi'an, aerosol Al solubility at Qingdao shows remarkable variations
269 under different weather conditions (Figure 4b, Table S7). Median Al solubilities were
270 determined to be 0.31% and 0.24% for supermicron and submicron particles during dust days,
271 significantly lower than these on clean days (0.99% and 0.77%, respectively). This is probably
272 because higher wind speeds during dust events hindered the accumulation of atmospheric
273 pollutants and shortened the transport time to Qingdao, and thus limiting the aging of mineral
274 dust aerosol. This explanation is supported by a recent study (Zhang et al., 2024) which found
275 that the aging extent of dust particles in Japan was much lower during fast-moving dust events



276 than slow-moving dust events. Moreover, large amounts of alkaline components (such as
277 carbonates) which were emitted to the atmosphere during dust days neutralized acid species
278 and therefore inhibited acid-promoted dissolution of aerosol trace elements (Zhi et al., 2025).

279 Figure 4b also suggests that aerosol Al solubilities were much higher during haze and fog
280 days at Qingdao, when compared to clean days. Highest Al solubilities were observed during
281 fog days, with median values being 6.90% for supermicron particles and 1.38% for submicron
282 particles, followed by haze days (3.64% and 1.58%, respectively). This is very likely due to
283 enhanced chemical processing during haze and fog periods (Shi et al., 2020; Shang et al., 2024),
284 and especially during fog days the large increase in RH cause huge increase in aerosol liquid
285 water, therefore greatly promoting aqueous reactions and Al dissolution.

286 In summary, aerosol Al solubility at Xi'an was low in general, and did not show much
287 variability in different seasons or under different weather conditions. Compared to Xi'an,
288 aerosol Al solubility was higher at Qingdao; furthermore, it was higher in the other three
289 seasons than in spring, and much higher for haze and fog days than dust days. These results
290 imply that atmospheric aging had little effects on aerosol Al solubility at Xi'an but could
291 significantly increase aerosol Al solubility at Qingdao, as further elaborated in Section 4.

292 **4. Discussion**

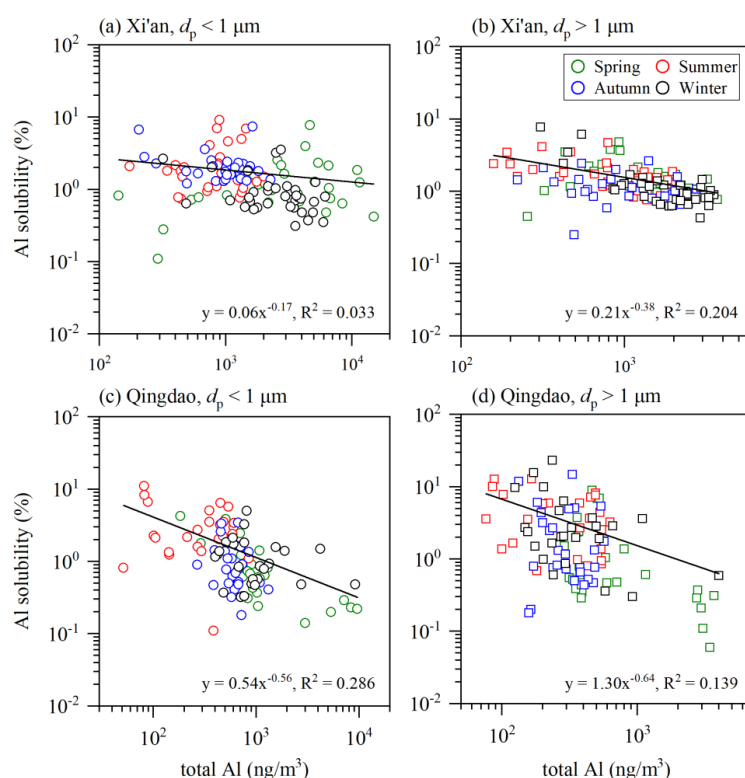
293 As shown in Figure 5, our work observed the inverse dependence of aerosol Al solubility
294 on total Al concentrations at both Xi'an and Qingdao, given by Eq. (1):

$$295 \quad f_s(\text{Al}) = a \times [\text{Al}]^{-b} \quad (1)$$

296 where $f_s(\text{Al})$ is aerosol Al solubility (%) and $[\text{Al}]$ is total Al concentration (ng/m^3). Such
297 relationship was also reported in some previous studies (Jickells et al., 2016; Shelley et al.,



298 2018; Baker et al., 2020; Shelley et al., 2025). Baker and Jickells (2006) suggested that such
299 inverse relationship was due to that larger particles have higher deposition velocities and lower
300 Al solubility: aerosol Al concentrations decrease during transport in the atmosphere due to
301 deposition, with deposition being faster for larger particles; as a result, aerosol particles will be
302 enriched with smaller particles with higher Al solubility. However, Shi et al. (2011) found no
303 significant change in Al solubility with particle size for desert dust samples, and therefore put
304 the explanation proposed by Baker and Jickells (2006) into doubt.



305
306 **Figure 5.** Aerosol Al solubility versus total aerosol Al concentrations: (a) submicron particles
307 at Xi'an, (b) supermicron particles at Xi'an, (c) submicron particles at Qingdao, (d)
308 supermicron particles at Qingdao.



309

310 Aerosol Fe solubility was also frequently observed to increase with the decrease in total
311 Fe concentrations (Sedwick et al., 2007; Mahowald et al., 2018; Meskhidze et al., 2019), and
312 one possible reason is the influence of anthropogenic aerosol Fe (Sholkovitz et al., 2009; Ito
313 and Shi, 2016) with higher solubility than desert dust (Schroth et al., 2009; Fu et al., 2012; Ito
314 et al., 2021). Nevertheless, being different from aerosol Fe, aerosol Al stems predominantly
315 from desert dust, with little contribution from anthropogenic sources; furthermore, Al solubility
316 was measured to be $0.4 \pm 0.6\%$ for coal fly ash (Li et al., 2022), an important type of
317 anthropogenic aerosols, not higher than that for desert dust ($0.8 \pm 0.4\%$). Therefore, we suggest
318 that anthropogenic emission may not be able to explain the inverse dependence of Al aerosol
319 solubility on total Al concentrations.

320 We argue that chemical processing in the atmosphere can very well explain such inverse
321 dependence. Total aerosol Al concentrations decrease with transport due to deposition, while
322 reactions with acidic gases (such as SO₂ and NO_x) can significantly enhance the dissolution of
323 aerosol Al (Jickells et al., 2016). Figure 5 shows that the inverse dependence of Al solubility
324 on total Al concentration was more pronounced at Qingdao, with the slopes (*b* values) much
325 larger than those obtained at Xi'an. This is because compared to Xi'an, Qingdao is more distant
326 from deserts and therefore dust aerosol is expected to be more aged at Qingdao. It also further
327 supports the vital role chemical aging plays in regulating aerosol Al solubility,

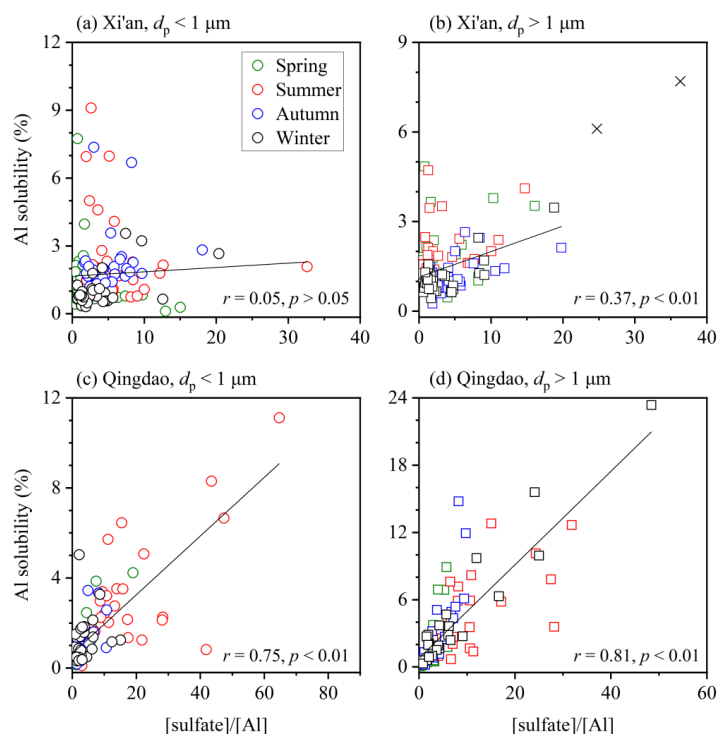
328 **4.1 Effects of acid processing and the role of RH**

329 **4.1.1 Effects of acid processing**



330 Laboratory experiments found that the amount of Al dissolved from minerals would
331 increase with the decrease in solution pH (Amram and Ganor, 2005; Bibi et al., 2011, 2014;
332 Cappelli et al., 2018), and some field measurements also suggested that acid processing in the
333 atmosphere could lead to significant increase in aerosol Al solubility (Measures et al., 2010;
334 Sakata et al., 2023). In this work, we examined the relationship between aerosol Al solubility
335 and the relative abundance of acidic species ([sulfate]/[Al] and [nitrate]/[Al]) at Xi'an and
336 Qingdao. It should be noted that non-sea-salts sulfate (Virkkula et al., 2006), instead of sulfate,
337 was used at Qingdao because it is a coastal city and heavily impacted by sea spray aerosol.

338 At Xi'an, aerosol Al solubility showed no significant correlation with [sulfate]/[Al] or
339 [nitrate]/[Al] for either supermicron or submicron particles ($r < 0.4$, Figure 6 and S1),
340 indicating that acid processing did not enhance aerosol Al solubility. Enhancement of aerosol
341 trace element solubility by acid processing requires internal mixing of acid species with mineral
342 dust particles (Baker and Croot, 2010). Previous studies (Wang et al., 2014; Wu et al., 2017)
343 suggested that mineral dust particles observed at Xi'an which is close to deserts largely
344 remained externally mixed with acid species, and thus aerosol Al solubility was not
345 significantly enhanced by acid processing at Xi'an.



346

347 **Figure 6.** Aerosol Al solubility versus [sulfate]/[Al]: (a) submicron particles at Xi'an, (b)
348 supermicron particles at Xi'an, (c) submicron particles at Qingdao, (d) supermicron particles
349 at Qingdao. Data represented by crosses are not included in fitting.

350

351 On the contrary, Figure 6 shows that aerosol Al solubility at Qingdao was well correlated
352 with [sulfate]/[Al] ($r > 0.7, p < 0.01$), implying that acid-promoted dissolution significantly
353 enhanced Al solubility. We also found that correlations of Al solubility with [sulfate]/[Al] was
354 better than those with [nitrate]/[Al] (Figures 6 and S1, Table S8), in line with a previous study
355 (Sakata et al., 2023) which found aerosol Al solubility at Hiroshima, southern Japan, to be
356 correlated with [sulfate]/[Al] but not with [nitrate]/[Al]. This may imply that chemical



processing by sulfate was more important than nitrate for Al solubility enhancement via acid processing, likely because aluminosilicate dust particles tend to react preferentially with SO₂ and H₂SO₄ while nitrogen oxides react mainly with carbonate particles (Sullivan et al., 2007; Fitzgerald et al., 2015). Furthermore, our work reveals better correlations between Al solubility and [sulfate]/[Al] for supermicron particles than submicron particles (Figure 6), indicating that the effect of acid processing on Al solubility was more important in supermicron particles.

4.1.2 The role of RH

Relative humidity (RH) is a vital factor influencing liquid water contents and phase state of aerosol particles and thus their secondary chemistry. When RH increased >60%, the phase state of aerosol particles in northern China changed from semisolid to liquid (Liu et al., 2017; Sun et al., 2018; Song et al., 2022), leading to large increase in aerosol liquid water content and thereby potentially affecting aerosol Al solubility.

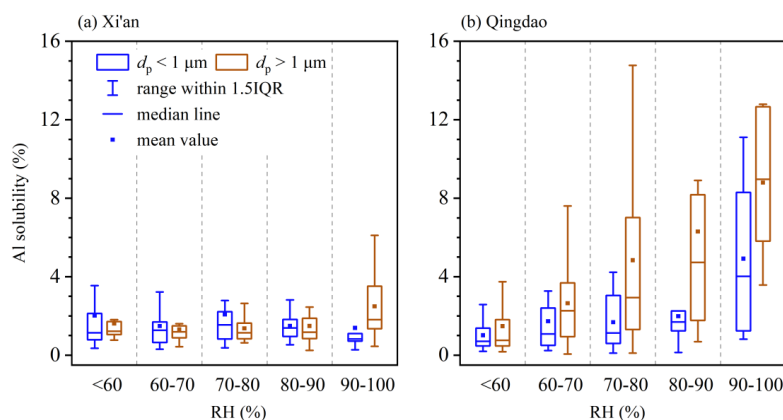


Figure 7. Aerosol Al solubility at different relative humidity (RH) for submicron and supermicron particles: (a) Xi'an, (b) Qingdao.



373 We observed no apparent variation of aerosol Al solubility with RH at Xi'an (Figure 7a).
374 When RH was <60%, median Al solubilities for supermicron and submicron particles were
375 1.22% and 1.14%, respectively; when RH increased >90%, the median Al solubilities were
376 determined to be 1.82% and 0.82%, showing no obvious increase when compared to those at
377 <60% RH. This again may imply that chemical processing had very limited impact on aerosol
378 Al solubility at Xi'an.

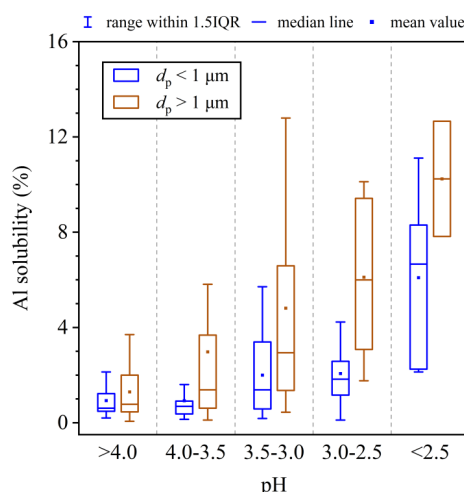
379 In contrast, RH played an important role in regulating aerosol Al solubility at Qingdao.
380 As shown in Figure 7b, for supermicron particles, the median Al solubility was only 0.76% at
381 <60% RH, and gradually increased to 4.73% at 80-90% RH, and abruptly increased to 8.87%
382 at >90% RH. For submicron particles, median Al solubility was <1% at <60% RH, and further
383 increase in RH to 80-90% did not lead to large changes in Al solubility; nevertheless, when RH
384 exceeded 90%, the median Al solubility was remarkably increased to 4.02%, much higher than
385 those observed when RH was < 90%.

386 **4.2 Effects of aerosol acidity on aerosol Al solubility at Qingdao**

387 Figure 8 shows the dependence of aerosol Al solubility on aerosol acidity (represented by
388 pH) at Qingdao (we did not measure NH₃ at Xi'an and thus could not estimate the aerosol
389 acidity in a reliable manner). For supermicron particles, the median Al solubility was only 0.99%
390 when aerosol pH was >4.0, and gradually increased to 10.24% as aerosol pH was decreased to
391 <2.5. For submicron particles, the median Al solubility was only 0.69% when pH was >4.0,
392 increased slightly with the decrease in pH when pH was in the range of 2.5-4.0, and then
393 increased greatly to 6.09% when pH was decreased to <2.5. In addition, aerosol acidity at
394 Qingdao was highest in summer and lowest in spring (Chen et al., 2024), consistent with the



395 seasonal variation of aerosol Al solubility, further supporting the importance of aerosol acidity
396 in regulating Al solubility.



397

398 **Figure 8.** Aerosol Al solubility corresponding to different aerosol acidity for submicron and
399 supermicron particles in Qingdao.

400

401 As shown in Figure S2, aerosol Al solubility was generally <2% when aerosol acidity was
402 low ($\text{pH} > 4.0$), and higher Al solubility (>2%) was usually observed for samples with high RH
403 and high acidity ($\text{pH} < 4.0$), again underscoring the roles of aerosol acidity (and RH). However,
404 some samples exhibited low Al solubility although the corresponding RH and aerosol acidity
405 were both higher, and such phenomenon was more pronounced for submicron particles. This
406 is very likely linked with aerosol mixing state (Riemer et al., 2019). Aerosol Al solubility and
407 acidity used in our work are both the average properties of an aerosol sample which contains
408 numerous particles, while in reality the two properties will have large particle-to-particle
409 variations. For a given aerosol sample, it can happen that particles with high acidity may



410 contain very little Al while particles with low acidity are enriched in Al; in this case, high
411 acidity do not promote Al solubility for this sample.

412 **4.3 Size-dependence of aerosol Al solubility**

413 At Xi'an, no significant difference in aerosol Al solubility was found between
414 supermicron and submicron particles across all the four seasons (Figure 3a). This is because
415 the aging extent of dust particles was rather limited at Xi'an (Wang et al., 2014; Wu et al., 2017)
416 and Al solubility does not vary with particle size for fresh dust particles (Shi et al., 2011). At
417 Qingdao, aerosol Al solubility showed no significant difference between supermicron and
418 submicron particles in spring, because the aging extent of dust arriving at Qingdao was also
419 limited in spring when Asian dust occurred most frequently. However, in the other three
420 seasons, Al solubility was significantly higher for supermicron particles than submicron
421 particles at Qingdao, and the ratios of median Al solubility in supermicron particles to that in
422 submicron particles were found to be 1.53, 1.70 and 2.57 in summer, autumn and winter,
423 respectively. Similar to our observation at Qingdao, Li et al. (2017) found that aerosol Al
424 solubility was significantly higher for TSP (14-28%) than PM_{2.5} (2-23%) at the summit of
425 Mount Heng, southern China.

426 On the other hand, a few other studies (Baker et al., 2020; Hsieh et al., 2023; Sakata et al.,
427 2023; Yang et al., 2023) found that aerosol Al solubility was higher in fine particles than coarse
428 particles. For example, aerosol Al solubility was found to increase with the decrease in particle
429 size over the tropical eastern Atlantic (Baker et al., 2020), being ~10.31% for particles in the
430 size of 0.36-0.61 μm and 0.43-4.53% for particles above 0.61 μm . At Hiroshima, southern
431 Japan, aerosol Al solubility was reported to be $8.82 \pm 6.48\%$ for fine particles ($< 1.3 \mu\text{m}$), more



432 than two times larger than that ($3.25 \pm 3.41\%$) for coarse particles ($>1.3 \mu\text{m}$) (Sakata et al., 2023).
433 Baker and Jickells (2006) suggested that this is because fine particles have larger surface-to-
434 volume ratios and thus facilitate Al dissolution via acid processing. Hsieh et al. (2023) found
435 aerosol Al solubility to be 38% for fine particles ($0.57\text{--}1.0 \mu\text{m}$) but only 0.37% for coarse
436 particles ($>7.3 \mu\text{m}$) over the East China Sea, and suggested that the observed size-dependence
437 could be explained by the enrichment of anthropogenic Al (which has higher solubility than
438 dust Al) in fine particles. However, aerosol Al originates predominantly from desert dust, with
439 little contributions from anthropogenic sources (Taylor and McLennan, 1985; Mahowald et al.,
440 2018), and fractional solubility of anthropogenic Al was not necessarily higher than desert dust
441 (Li et al., 2022).

442 As discussed above, there is not clear yet how and why aerosol Al solubility varies with
443 particle size. Such discrepancy is at least partly because different leaching protocols were used
444 in previous studies to extract dissolved aerosol Al and thereby Al solubility obtained in
445 different studies was not directly comparable (Meskhidze et al., 2019; Li et al., 2023; Li et al.,
446 2024). Furthermore, mechanistic insights can be obtained by laboratory experiments which
447 examine the size dependence of the solubility and dissolution kinetics of Al for mineral dust
448 particles under atmospherically relevant conditions.

449 **5. Conclusions and atmospheric implications**

450 Deposition of mineral dust aerosol is a major external source of several nutrient and toxic
451 elements for surface water in open oceans, and thus have significant impacts on marine
452 biogeochemistry; however, previous studies which estimated dust deposition flux into the
453 oceans reveals large discrepancies. Aerosol Al solubility, which is a critical parameter in using



454 dissolved Al concentrations in surface seawater as a tracer to constrain dust deposition flux,
455 remains poorly understood. In this work, we investigated seasonal variations of aerosol Al
456 solubility for supermicron ($>1\ \mu\text{m}$) and submicron ($<1\ \mu\text{m}$) aerosol particles at Xi'an and
457 Qingdao, both located in northern China, in attempt to elucidate the processes and mechanisms
458 which govern the variation of aerosol Al solubility in the atmosphere.

459 At Xi'an, aerosol Al solubility was low in general for both supermicron and submicron
460 particles, showing no obvious variability in different seasons or under different weather
461 conditions. This implies that chemical processing did not significantly enhance aerosol Al
462 solubility at Xi'an, as it is an inland city close to major deserts in northwestern China and thus
463 the aging extent of mineral dust particles arriving at Xi'an was quite limited. Compared to
464 Xi'an, aerosol Al solubility was higher at Qingdao, a coastal city in northern China;
465 furthermore, Al solubility was higher in the other three seasons than in spring, and much higher
466 for haze- and especially fog-impacted days than dust days. This indicates that chemical
467 processing substantially increased aerosol Al solubility at Qingdao.

468 Aerosol Al solubility at Xi'an showed no significant correlation with relative abundance
469 of sulfate or nitrate, and did not vary apparently with RH; in contrast, Al solubility at Qingdao
470 was well correlated with relative abundance of sulfate and nitrate, and increased with RH. This
471 further supports that chemical processing had little impacts on aerosol Al solubility at Xi'an
472 (because the aging extent of mineral dust aerosol at Xi'an is very limited) but significantly
473 increased aerosol Al solubility at Qingdao (because mineral dust particles transported to
474 Qingdao were substantially aged). Moreover, for both supermicron and submicron particles,
475 Al solubility at Qingdao was found to increase with aerosol acidity (in addition to RH),



476 underscoring the vital role of aerosol liquid water and acidity in enhancing Al dissolution via
477 chemical aging.

478 Our comprehensive investigation of aerosol Al solubility at two locations in Northern
479 China suggests that atmospheric chemical processing dictates aerosol Al solubility. As a result,
480 aerosol Al solubility is expected to spatially variable, depending on the extent of chemical
481 processing. For example, we found that aerosol Al solubility is higher at Qingdao than Xi'an
482 in general, and expect it to increase further as mineral dust aerosol is further transported
483 eastward to the Pacific. As a result, when leveraging dissolved Al concentrations in surface
484 seawater as a tracer to estimate deposition flux of desert dust aerosol into open oceans,
485 considering the spatial distribution of aerosol Al solubility, instead of using a uniform value on
486 the global scale, can help us better constrain the oceanic deposition flux of desert dust.

487

488 **Author contribution.**

489 **Tianyu Zhang:** Formal analysis, Investigation, Writing - Original Draft, Writing -
490 Review & Editing; **Yizhu Chen:** Formal analysis, Investigation, Writing - Original Draft;
491 **Huanhuan Zhang:** Investigation; **Lei Liu:** Writing - Review & Editing; **Chengpeng Huang:**
492 Investigation; **Zhengyang Fang:** Investigation; **Yifan Zhang:** Investigation; **Fu Wang:**
493 Resources; **Lan Luo:** Resources; **Guohua Zhang:** Writing - Review & Editing; **Xinming**
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495 - Original Draft, Writing - Review & Editing.

496 **Competing interests.**

497 The authors declare that they have no conflict of interest.



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