



Hydrometeorology and landscapes control sediment and organic matter mobility across a diverse and changing glacier-sourced river basin

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Short summary:

Rivers are valuable indicators of climate change. We used a multi-year, integrated monitoring program within a changing northern river basin to understand how river material changes across differing landscape types and runoff conditions. We show that delivery of suspended sediment and organic matter changes widely between years and within and between rivers whether draining pristine or impacted catchments. This work demonstrates challenges facing river water users under a changing climate.



Abstract

Northern landscapes are enduring ongoing impacts of anthropogenic land use and climate change. Rivers are valuable indicators of this change reflected by the timing and amount of water and terrestrial material they mobilize. Assessing the influence of a changing climate on diverse river systems is best achieved using multi-annual monitoring and replication of effort across varied tributary catchment conditions. We used this approach to monitor concentrations, catchment yields, and export of total suspended sediments (TSS) and dissolved organic carbon (DOC) of a large, diverse, glacial river network (North Saskatchewan River; NSR) in western Canada during years of extensive weather fluctuations. Though concentrations of TSS and DOC increased eastward through the NSR basin from Rocky Mountain cordillera to agriculturalized plains, catchment yields were statistically highest from cordillera regions, reflecting an eastward rain shadow. Wet conditions across the basin resulted in variable but statistically higher TSS and DOC yields compared to drought conditions. During higher water, we observed disordered, threshold-type, erosive mobilization of TSS through the basin whereas DOC increased more predictably with runoff. Variability of yields and export was substantial both within and between pristine and impacted catchments, and within the NSR mainstem illustrating the complexity of river sediment and organic matter transport at the network scale. Consequently, in a warming and wetting climate, we expect TSS and DOC transport to intensify with sediment transport being more difficult to predict compared to organic matter, which has implications for aquatic ecosystems and >1.5M people who depend on the NSR for drinking water.



1. Introduction

Northern latitudes have reported considerable human-driven climate and landscape changes post industrial revolution. Air temperatures have increased there at twice the global rate recently (Gulev et al., 2021) and extreme heat events have become more frequent (Simpkins, 2017; La Sorte et al., 2021; Vogel et al., 2019). Severe meteorological events in the form of high intensity rainfalls and droughts are also becoming more common (Walsh et al., 2020). Consequently, northern regions are losing glacial ice and snow cover (Marzeion et al., 2014; Mudryk et al., 2020), experiencing acute high and low water events (Pomeroy et al., 2016; Yin et al., 2018), and enduring more frequent and devastating wildfires (Hanes et al., 2018). The effect of these climate-driven shifts on regional hydrology and subsequent interaction with human-impacted landscapes (Brandt et al., 2013; Hirsh-Pearson et al., 2022) has resulted in marked changes in the transport of material from landscapes to fluvial systems (Lu et al., 2013).

Export of material by rivers is a key metric to identify and assess landscape response(s) to changes in climate and anthropogenic activities (Allan, 2004). This understanding is driven by the strong coupling between hydrometeorological conditions and the ability of rivers to mobilize landscape material, as modified by its availability. Notably, total suspended sediment (TSS) and dissolved organic carbon (DOC) are ubiquitous, landscape-derived materials found in rivers that are mobilized during runoff events and are sourced from erosion and flushing of upland soils, wetlands, riparian areas, and channel stores. Sediment transport typically increases in rivers as runoff and erosion regimes become more energetic in warming and wetting climates, particularly where sediment availability is increased as a result of anthropogenic activities or changes in vegetation cover (Whitehead et al., 2009). Flushing of agricultural soils and urban areas is known to drive nutrient, DOC, and contaminant delivery downstream (St-Hilaire et al., 2016). At the other extreme, heat and drought can induce runoff shortages and low river flows that produce sediment- and organic-poor waters dominated by groundwater inputs (Mosley, 2015). Similarly, runoff through intact landscapes and wetlands can modify surface water quality by storing water and promoting physical and biogeochemical removal of nutrients, DOC, and contaminants (Knox et al., 2008; Neary



81 et al., 2009). However, hydrometeorological conditions and the mobilization of
82 landscape material can vary significantly across temporal and spatial scales making it
83 challenging to use rivers as indicators of change, especially where monitoring data may
84 be sparse or unrepresentative (Orwin et al., 2010; Shi et al., 2017).

85 Long-term environmental monitoring programs are key tools for investigating the
86 impacts of multiple stressors on fluvial systems (Lovett et al., 2007) and can address
87 deficits common in shorter-term studies. For example, understanding climate impacts
88 on rivers requires consistent monitoring to account for sometimes considerable inter-
89 and intra-annual variability in hydrometeorological conditions. Different landscapes have
90 varied and often complex influences on river water quality (Maloney and Weller, 2011),
91 thus monitoring programs with sampling and replication of diverse conditions is
92 beneficial (Lintern et al., 2018a). Further, using a river network monitoring design
93 provides a means to differentiate contributions of landscape material via tributary
94 catchments from in-river sources (i.e., banks, channels) of major rivers downstream
95 (Chalov and Ivanov, 2023). Monitoring programs that include these features are rare but
96 present an opportunity to better understand how climate variability and change across
97 diverse landscapes can impact major rivers and water users.

98 Here, we assess how wide fluctuations in hydrometeorological conditions of a
99 diverse and changing northern river basin impact the downstream delivery of key water
100 quality, aquatic ecosystem health, and drinking water treatment parameters (i.e., TSS,
101 DOC). The North Saskatchewan River (NSR) in western Canada is sourced from the
102 glaciated Canadian Rocky Mountains and transects mountain bedrock, forest,
103 agricultural, and urbanized landscapes. In addition to ongoing changes experienced by
104 Rocky Mountain rivers (Goode et al., 2012; Rood et al., 2008), the NSR basin has
105 recently endured hydrometeorological extremes, ranging from high runoff years to
106 historic heat and drought conditions. These wide-ranging conditions may become more
107 common in the future and thus provide a window to better understand the response of
108 northern rivers to climate extremes. We used data from two monitoring programs for this
109 study. The WaterSHED (Water: Saskatchewan Headwaters Edmonton Downstream)
110 Monitoring Program is a collaborative tributary catchment monitoring network across the
111 NSR basin initiated in 2019 by a drinking water treatment utility, local governments, and



112 a non-government organization (Emmerton et al., 2023). The program was designed to
113 balance monitoring across different catchment conditions within the NSR basin. In
114 addition, long-term water quality and flow monitoring data were used from stations on
115 the NSR mainstem, operated by the governments of Alberta and Canada. We expected
116 differences between how TSS (i.e., particulates) and DOC (i.e., organic material)
117 mobilized through the river network, but also differences in how each mobilized between
118 catchment types and different hydrometeorological conditions.

119 **2. Methods**

120 **2.1 Setting**

121 The upper NSR basin in Alberta, Canada is a large, diverse, continental basin
122 draining ~40,000 km² of headwater landscapes and high prairie (Figure 1). The NSR
123 originates at the Saskatchewan Glacier in Banff National Park in the Rockies. As the
124 river travels east, it passes through cordillera, forested foothills, mixed forests,
125 agriculturalized plains, and an urban centre (Edmonton; population: 1.4M; Statistics
126 Canada, 2023) that sources drinking water from the NSR. Mean annual discharge of the
127 NSR at Edmonton is ~140 to 360 m³ s⁻¹. Surficial geology across the basin progresses
128 from exposed bedrock in the Rockies to heterogeneous glacial tills in foothills and
129 plains. This till has been shaped by climate, topography, and biogeochemistry since
130 deglaciation ca. 14,000 years ago, resulting in sparse and thin luvisols in cordillera and
131 foothills regions to thicker, organic- and nutrient-rich chernozemic soils through
132 mixedwood forests and plains. Precipitation is low across the basin due to the Rocky
133 Mountain rain shadow; ranging from ~600 mm yr⁻¹ in the west basin to ~350 mm yr⁻¹ in
134 the east. Water yields, accordingly, are highest in the western cordillera and foothills
135 regions relative to the lower-relief, easterly portions of the basin. The largest western
136 tributaries of the NSR include the Brazeau, Ram, and Clearwater rivers, while smaller
137 creeks and rivers in the central and eastern basin include the Sturgeon and Redwater
138 rivers (Figure 1). Two reservoirs exist in the west basin on the NSR (Abraham Lake-
139 Bighorn Dam) and the Brazeau River (Brazeau Reservoir-Brazeau Dam).



140 **2.2 Data**

141 *2.2.1 Catchment structure*

142 To quantify the influence of major landscape types on modulating different
143 hydrometeorological conditions, the physical structure of all land across the entire NSR
144 basin was delineated using the catchment structural unit (CSU) approach (Nippgen et
145 al., 2011; Orwin et al., 2022). CSUs are a series of unique numerical identifiers derived
146 by combining key structural layers that include land use/landcover, surficial geology,
147 wetlands, and slope that are known to play significant roles in how catchments respond
148 to hydrometeorological inputs (Orwin et al., 2022). CSU codes are generated at a 400
149 m² pixel scale across the basin, to provide wall-to-wall coverage. For example, a pixel
150 code of 60111 represents a dominant shrubland cover (60,000), mixed coarse-fine
151 surface geology (100), with wetlands present (10), and a slope less than 10% (1).
152 Further data details are provided in the supporting information and Table S1a.

153 *2.2.2 Hydrometeorology*

154 Mean annual air temperature and total annual precipitation data were retrieved from
155 Canadian Climate Normal meteorological stations (1981-2010) across the NSR basin
156 (Table 1; Government of Canada, 2024a). Mean annual air temperature and total
157 annual precipitation for the water years (November 1st to October 31st) 2019 to 2022
158 were also collected from the same stations. Mean annual river flow data were collected
159 from either long-term federal (Government of Canada, 2024b) or provincial
160 (Government of Alberta, 2023) monitoring stations across the NSR basin (Table 1).
161 Historic data were collected between the water years of 1981 and 2010. Mean annual
162 flow from water years 2019 to 2022 were also collected.

163 *2.2.3 Water quality*

164 Tributary monitoring occurs at the mouths of 20 WaterSHED catchments (Table 1;
165 Figure 1) and includes: 1. flow gauging that monitors water level and estimates river
166 flow every 15 minutes; and 2. flow-targeted water quality grab sampling at 30 cm depth
167 throughout the open water season (March–October) at all sites and monthly winter
168 sampling (below-ice) at larger tributary stations. Water samples are collected and
169 analyzed for general chemistry, nutrients, trace metals, contaminants, and biological



Table 1 Metrics of meteorological (Canadian Climate Normals; Government of Canada, 2024a), water quality (WaterSHED Monitoring Program, Long Term River Network; Emmerton et al., 2023), and flow (Government of Alberta, 2023; Government of Canada, 2024b) stations of the North Saskatchewan River (NSR) basin used in this study. Paired flow indicates flow monitoring occurs at same locations as water quality sampling. Stations ordered upstream to downstream through the basin. Open-water (May to October; OW); M-monthly; C-continuous.

Meteorology					
Station	Latitude, Longitude	Monitoring frequency	Canadian Climate Normal	Study period	Elevation (m asl)
Nordegg RS	52.500, -116.050	C	1981-2010	2018-2022	1320.1
Entwistle	53.600, -114.983	C	1981-2010	2018-2022	780.3
Edmonton Stony Plain	53.548, -114.108	C	1981-2010	2018-2022	766.3
Edmonton International A	53.317, -113.583	C	1981-2010	2018-2022	723.3
Elk Island National Park	53.683, -112.868	C	1981-2010	2018-2022	716.2
Water Quality			Flow		
Station	Latitude, Longitude	Monitoring frequency (# samples)	Catchment Area (km ²)	Paired flow?	Monitoring frequency
NSR at Whirlpool Pt.	52.001, -116.471	M (54)	1,920	Yes	C
NSR at Saunders Camp.	52.454, -115.759	M (47)	5,089	No ^a	C
Ram R.	52.368, -115.420	M (55)	1,881	Yes	C
NSR u/s Clearwater R.	52.348, -114.982	M (45)	7,715	No ^b	C
Clearwater R.	52.253, -114.856	M (55)	3,221	Yes	C
Baptiste R.	52.664, -115.076	M (53)	1,358	Yes	C
Nordegg R.	52.820, -115.513	M (54)	865	Yes	C
Brazeau R.	52.913, -115.364	M (51)	5,658	Yes	C
Rose Ck.	53.052, -115.051	M-OW (52)	657	Yes ^c	C-OW
Modeste Ck.	53.248, -114.706	M-OW (40)	1,178	Yes	C-OW
Tomahawk Ck.	53.352, -114.660	M-OW (36)	190	Yes ^d	C-OW
Strawberry Ck.	53.311, -114.052	M-OW (51)	589	Yes	C-OW
NSR at Devon	53.369, -113.751	M (50)	26,362	No ^e	C-OW
Whitemud Ck.	53.484, -113.555	M-OW (39)	1,086	Yes	C-OW
Sturgeon R.	53.833, -113.283	M-OW (37)	3,330	Yes	C-OW
Redwater R.	53.897, -113.000	M-OW (37)	1,602	Yes	C-OW
NSR at Pakan Bridge	53.991, -112.476	M (51)	39,333	Yes ^f	C

Notes: ^aNSR at Saunders Camp. uses flow from upstream NSR at Bighorn Dam (05DC010) and uses water yield from Ram R. (05DC006) to estimate additive flow from ungauged drainage areas; ^bNSR u/s Clearwater R. uses flow from downstream NSR near Rocky Mountain House (05DC001) and removes flow from Clearwater R. at Dovercourt (05DB006) and Prairie Ck. near Rocky Mountain House (05DB002); ^cNewly installed paired station in 2019 without a long-term record. Long-term record available upstream at Rose Creek at Alder Flats (05DE007); ^dNewly installed paired station in 2019 without long-term record. Long-term record available upstream at Tomahawk Creek at Tomahawk (05DE009); ^eNSR at Devon uses flow from downstream NSR at Edmonton (05DF001) and removes flow from Whitemud Ck. at Edmonton (05DF009); ^fNSR at Pakan Bridge uses paired flow after Nov. 6/20, otherwise uses flow from upstream NSR at Edmonton (05DF009), adds flow from Sturgeon R. (05EA001), Redwater R. (05EC005), and Waskatenau Ck. (05EC002) and uses water yield from Redwater R. (05EC005) to estimate additive flow from ungauged drainage areas. Station information can be found from Water Survey of Canada (Government of Canada, 2024b) and Alberta River Basins (Government of Alberta, 2023).

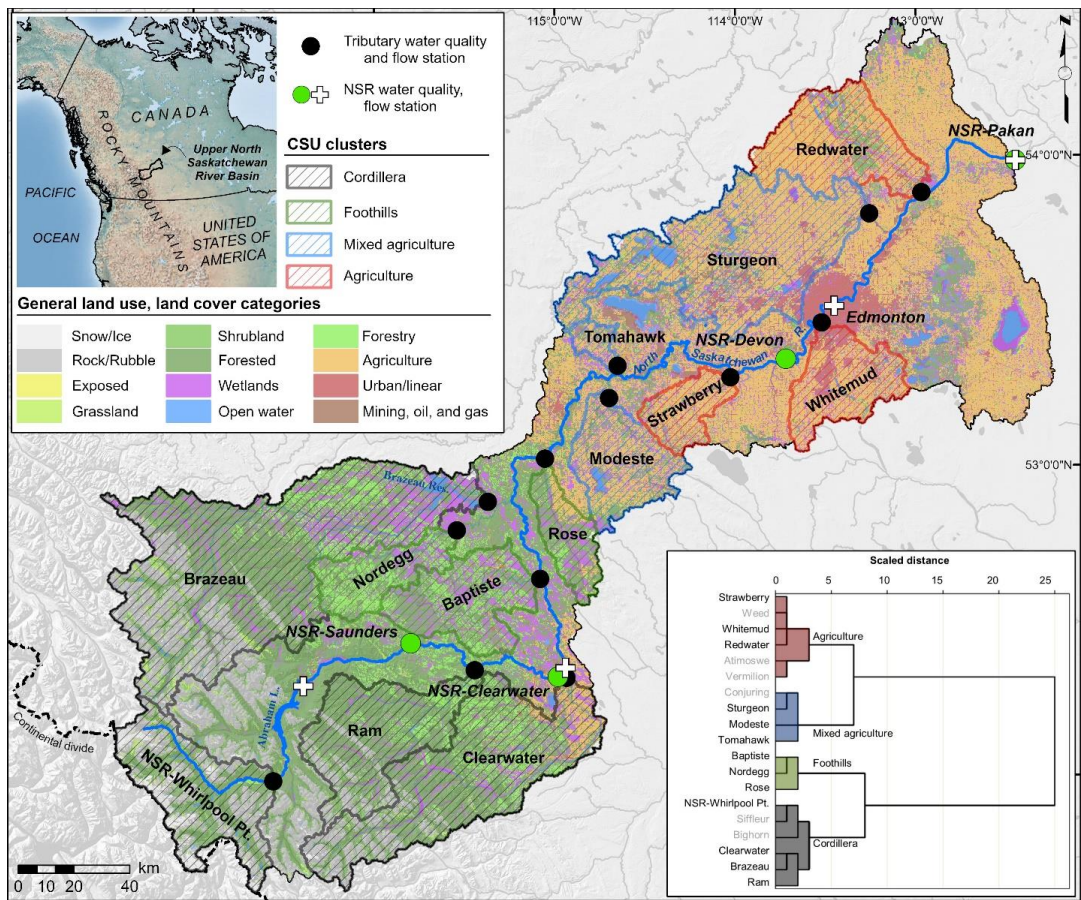


Figure 1 The upper North Saskatchewan River (NSR) basin in Alberta, Canada showing tributary catchment and NSR mainstem water quality and flow stations. Catchment structural units (CSU; see methods) showing general landscape cover and land use across the basin is shown and monitored catchments are classified using hierarchical clustering analysis (lower inset; see Methods). Greyed catchment names contributed to the landscape clustering analysis but were not included in this study due to insufficient flow data. Of note, though NSR Whirlpool Point is on the NSR mainstem, it is viewed in this study as a cordillera tributary as the river's origin is part of this catchment. Additional data from Shaded Relief (http://www.shadedrelief.com/world_relief/home.html).

measures at each station (Emmerton et al., 2023). Four mainstem stations of the NSR are monitored for water quality but are not necessarily co-located with flow stations. Therefore, flow was estimated at water quality stations using records from the nearest flow station and local tributary water yields applied to remaining ungauged catchment area between stations (Table 1; Figure 1). Sample collection and laboratory analysis



206 methods under this program are the same as those under WaterSHED, though sample
207 collections at NSR mainstem stations are at a monthly frequency only.

208 Only TSS and DOC were used as water quality parameters for numerical
209 assessments, for reasons described in the following section. TSS was measured using
210 the gravimetric approach (SM 23 2540Cm; Rice et al., 2017) with 2 μm filtration and a
211 reported detection limit of 1 mg L^{-1} . Colorimetry via an infrared autoanalyzer (SM 22
212 5310Bm; Rice et al., 2017) was used to determine DOC concentrations in water
213 samples at a detection limit of 0.5 mg L^{-1} after 0.45 μm filtration. All WaterSHED and
214 NSR mainstem data are publicly available (Government of Alberta, 2024) and are
215 assessed for quality control and assurance (Lacey et al., 2022).

216 **2.3 Numerical assessments**

217 *2.3.1 Catchment structure*

218 As the CSU codes are generated in a wall-to-wall coverage across the NSR basin,
219 numeric summaries and distribution of CSU codes can be generated within catchment
220 areas. We used ArcGIS (v10.7.1) HydroTools and a 100 m interpolated digital elevation
221 model to delineate areas of each of the 20 WaterSHED catchments covering most of
222 the NSR basin. Individual CSU codes in each catchment were summed by area to
223 quantify a percent catchment coverage of each code. To classify different landscape
224 types across NSR basin, we produced a matrix of CSU codes and areal percentages for
225 each WaterSHED catchment and entered into a hierarchical cluster analysis using
226 Euclidean distances and Ward's approach (SPSS v28). The number of clusters selected
227 for further assessments was chosen manually to balance replication of catchments
228 within a reduced number of clusters. These landscape types thus formed the basis for
229 subsequent statistical assessments to understand how riverine transport of TSS and
230 DOC were modulated by hydrometeorological conditions across differing landscape
231 conditions across the basin.

232 *2.3.2 Hydrometeorology*

233 To contextualize recent conditions relative to those from the past, we compared
234 mean annual air temperatures and total annual precipitation from a meteorological
235 station during its climate normal period (1981–2010) with mean air temperature and



236 total precipitation from the 2019 to 2022 water years at the same station. Similarly, we
237 compared historic mean annual flow at a given long-term flow monitoring station (1981
238 to 2010) to mean annual flow from the water years 2019 to 2022 at the same station.
239 Total annual flow for each water year and each station was calculated by summing daily
240 flows. Total annual runoff yield (mm) was then calculated by dividing the annual flows
241 by catchment areas.

242 2.3.3 Concentrations, export, and catchment yields

243 We chose 13 of the 20 catchment stations to include in concentration, chemical
244 export, and catchment yield assessments since long-term flow data (and therefore
245 reliable rating curves) were not available yet at all catchments (Table 1). Eleven of
246 these 13 stations had paired long-term flow station records, while two new stations were
247 included in assessments (Rose, Tomahawk) since they had long-term flow records
248 available at upstream stations. TSS and DOC were selected as representative
249 parameters for this study because previous work from Alberta river monitoring programs
250 has shown notable concentration associations between TSS and several unfiltered
251 trace metals (e.g., Al, Co, Cr, Cu, Fe, Pb, V, Zn), while DOC concentrations showed
252 patterns similar to total and dissolved nitrogen and phosphorus, and dissolved trace
253 contaminants (e.g., As, Fe, Hg, V; Emmerton et al., 2023). Of note, DOC is the majority
254 of total organic carbon in these systems (Government of Alberta, 2024).

255 Quantification of export (i.e., concentration $\times$ flow) in rivers typically includes
256 assessments of parameter concentrations from grab samples collected relatively
257 infrequently and higher frequency river flow monitoring. Over a period of several years,
258 grab samples from a given station are collected across much of a river's hydrological
259 and biogeochemical conditions (Figure S1). Quantification of export (mass) from rivers
260 may be achieved using several different approaches. We used the R package *loadflex*
261 (Appling et al., 2015) to estimate annual export at catchment and NSR mainstem
262 monitoring stations using four different export models (i.e., interpolation, USGS Loadest,
263 log-log regression, Loadest-composite approach; see supporting information). Annual
264 tributary catchment yields (mass per area) were quantified by dividing estimated export
265 by gross catchment area for each station.



266 To quantify TSS and DOC export and storage (in ktonnes) of four NSR mainstem
267 reaches between water quality monitoring stations, we used median export of the four
268 models at all stations and a mass balance approach (after Yin et al., 2021):

$$\Delta S_{\text{NSR}} = I_{\text{MT}} + I_{\text{UT}} - (O_{\text{NSR(DS)}} - I_{\text{NSR(US)}}) + E \quad (1)$$

271 where I_{MT} denotes reach inputs from monitored tributary catchments; I_{UT} is unmonitored
272 tributary export to the reach using the product of catchment yields of other nearby
273 monitored tributaries and total unmonitored land area draining to the reach; $O_{\text{NSR(DS)}}$
274 represents the NSR mainstem station export at the downstream bound of the reach;
275 and $I_{\text{NSR(US)}}$ is the NSR mainstem station export at the upstream bound of the reach.
276 ΔS_{NSR} is assumed to be contributions to or subtractions from the NSR mainstem from
277 in-channel sediment deposition (–), in-channel/bed or bank erosion (+), or additions or
278 subtractions from riparian and floodplain regions ($\pm$). The residual (E) includes
279 groundwater or atmospheric contributions and losses of TSS or DOC, which are
280 assumed low relative to surface runoff contributions and not considered here.

282 2.3.4 Linear mixed models

283 To assess statistical differences between catchment yields of TSS and DOC across
284 different landscape types, between water years, and between four different export
285 models, we used linear mixed models (SPSS v28). The three fixed factors in the mixed
286 models were landscape type, water year, and export model while the random factor was
287 monitoring stations using autoregressive approach. Daily catchment yield data
288 (dependent variable) were log-transformed to meet normality standards in the mixed
289 model. Sidak multiple comparisons were used to assess statistical differences ($\alpha=0.05$)
290 between geometric means of TSS and DOC yields by landscape type, water year, and
291 export model. Models were run using yield data from all monitored catchments, but run
292 again excluding the regulated Brazeau River station to remove effects of the dam and
293 reservoir upstream of that station on sediment transport and DOC.



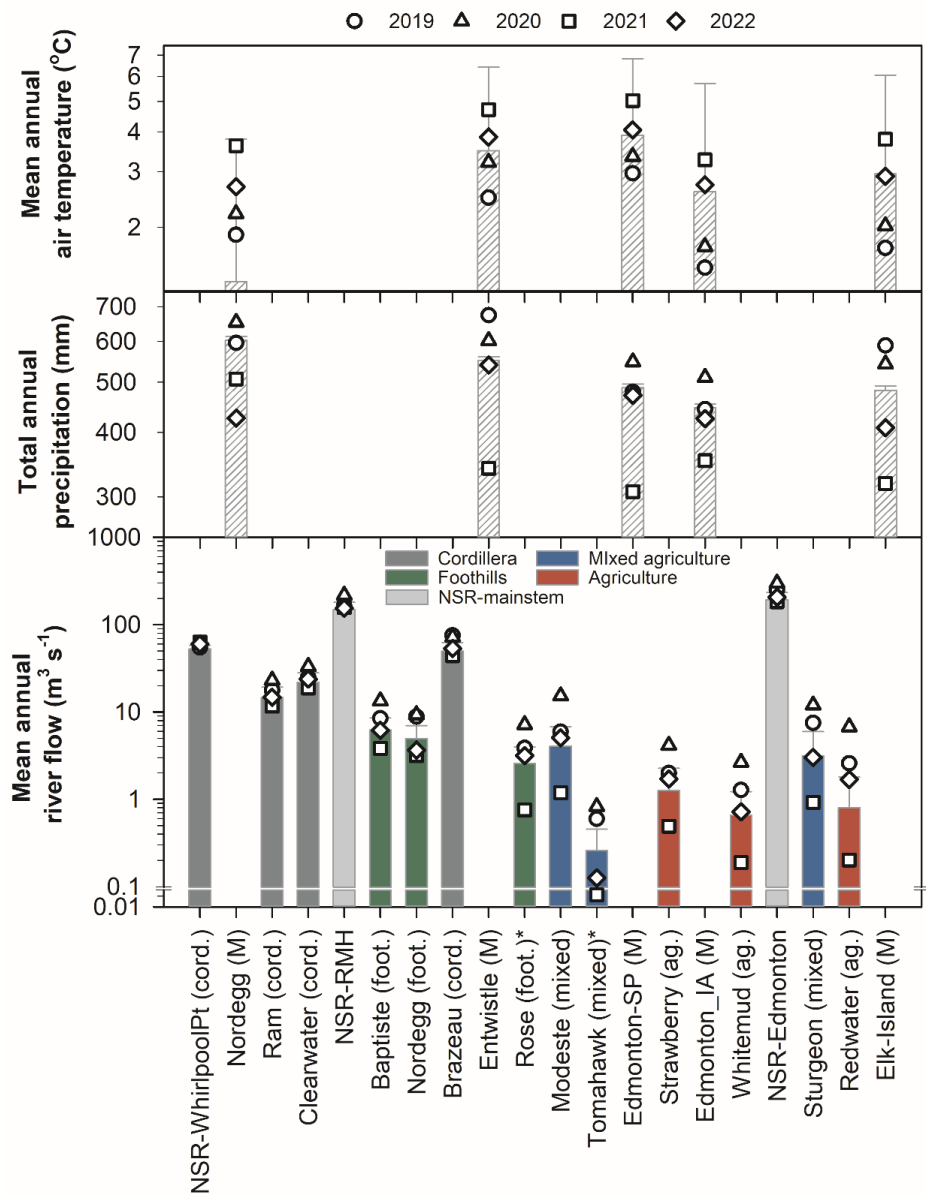
294 **3. Results**

295 **3.1 Catchment structure**

296 Using the CSU approach (Table S1a) and hierarchical clustering of all WaterSHED
297 catchments, we identified four, statistically distinct tributary landscape types across the
298 NSR basin (Figure 1; Table S1b). Cordillera catchments are dominated by steep
299 rock/rubble land cover, exposed bedrock, bare and forested slopes, and few wetlands
300 or anthropogenic land uses. Foothills catchments are typically forested with wetlands,
301 have coarse and fine surface sediments, and generally have rolling topography.
302 Forestry is the dominant land use, with less extensive footprints by oil and gas
303 development and linear features. Mixed agriculture catchments are dominated by low-
304 slope agricultural lands, with patches of mixedwood forest, grasslands, large lakes, and
305 energy production footprints. Wetland coverage is low and surficial geology is mixed
306 with coarse sediments dominating. Finally, agriculture catchments are heavily
307 agriculturalized landscapes with coarse tills, few wetlands, limited open water, and
308 abundant linear features relative to other classes due to proximities to urban areas and
309 rural road grids. The NSR at Whirlpool Point flow station was the only catchment to
310 drain a substantial glacier and these meltwaters were delivered downstream to other
311 NSR mainstem stations.

312 **3.2 Hydrometeorology**

313 Regardless of geography, monitoring stations across the NSR basin experienced
314 relatively cool air temperatures and the highest precipitation accumulations and
315 subsequent river flows during the water years 2019 and 2020 (Figure 2). One-year
316 standardized precipitation index mapping (Figure S2) suggested between normal and
317 one in three-year wet conditions persisted across most of the NSR basin in 2019 but the
318 Brazeau catchment experienced up to 1 in 25-year wet conditions. In 2020, the NSR
319 basin experienced widespread one in three-year wet conditions with patches of up to 1
320 in 12-year wet conditions. In contrast, 2021 was characterized by extreme heat and
321 drought highlighted by a widely-reported atmospheric blocking (heat dome) event over
322 much of western Canada (Zhang et al., 2023). Consequently, 1 in 25-year, up to 1 in
323 100-year dry conditions prevailed across the NSR basin, particularly in foothills and



324
325 **Figure 2** Canadian climate normals for mean annual air temperature (upper panel bars)
326 and total annual precipitation (centre panel bars) and mean annual river flow (lower
327 panel bars) from monitoring stations across the North Saskatchewan River (NSR) basin
328 (1981-2010; ± 1 standard error). Annual metrics from 2019 to 2022 water years are
329 overlain as symbols. All stations are ordered west to east through the basin and
330 mainstem stations include the prefix "NSR" in their labels and (M) labels denote
331 meteorological stations. *Station historic flow data from longer-term upstream stations
332 (Rose Creek at Alder Flats; Tomahawk Creek at Tomahawk).



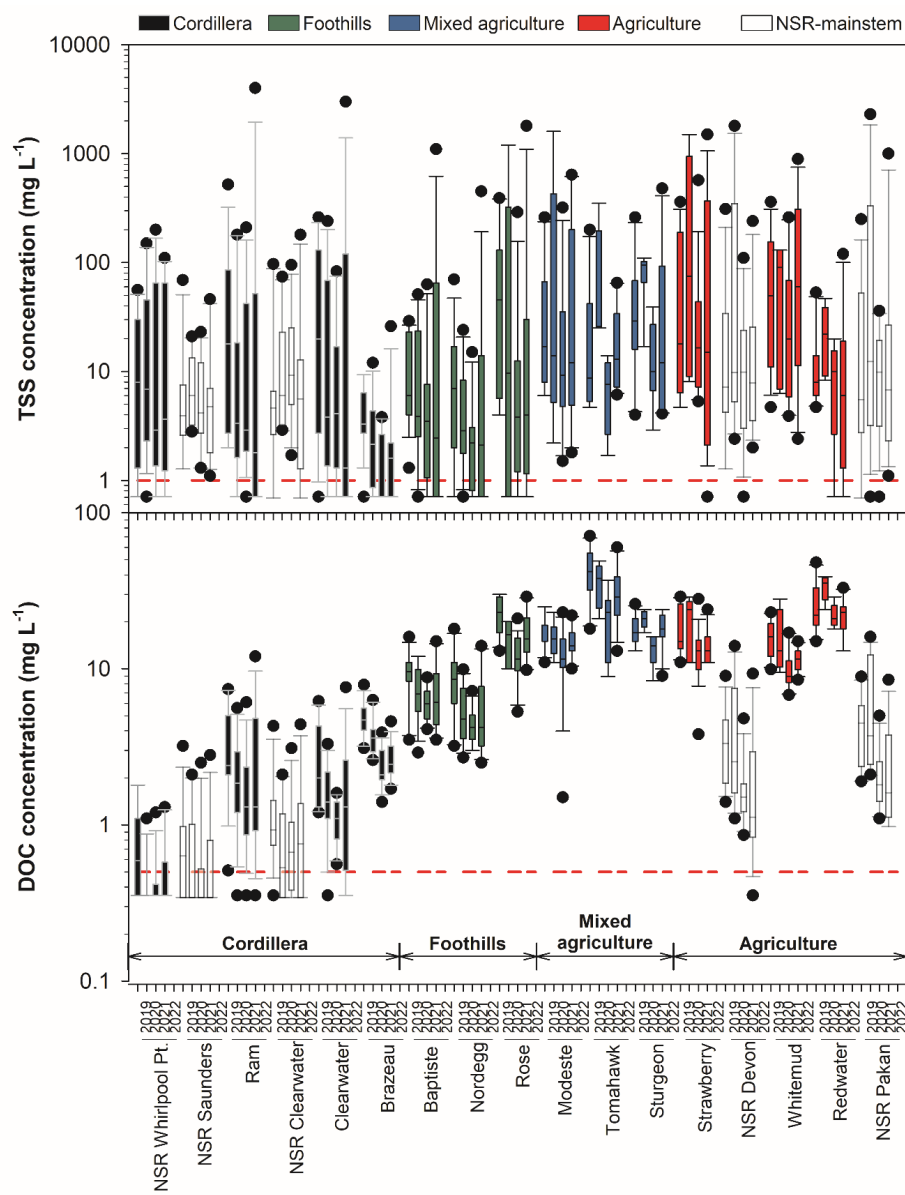
333 downstream plains regions. Meteorological conditions across most of the basin were
334 near-normal in 2022.

335 The cordillera NSR at Whirlpool Point and the regulated Brazeau River stations
336 were the largest tributary contributors of flow to the NSR of monitored stations, each
337 contributing ~24% of all flow at the NSR–Edmonton station between water years 2019
338 and 2022 (Figure 2). Other cordillera rivers including the Ram and Clearwater were the
339 next notable contributors of river water to the NSR, while downstream foothills and
340 plains rivers were less important for flow. Catchment water yields (runoff) also generally
341 followed this downstream progression (Table S2).

342 Between years (Figure 2; Table S2), the glacially-sourced NSR at Whirlpool Point
343 had the lowest variability in mean annual flow (50th–52nd percentile of historic record)
344 and was unique across the network as it experienced relatively high flows during much
345 of the hot and dry 2021 summer. The regulated Brazeau River produced stable but high
346 flows between 2019 and 2022 (57th–64th), particularly during a very wet 2019 in the
347 catchment. Other cordillera rivers showed lower relative flows and higher interannual
348 flow variabilities (44th–62nd). Both foothills (39th–69th) and plains rivers (38th–85th) were
349 even more variable in flow conditions between years. Of note, cordillera rivers Ram and
350 Clearwater showed an overall annual flow condition that was near normal in 2022
351 despite experiencing isolated storm events in early June 2022 that increased their flows
352 substantially for a short period. Flow at the NSR at Edmonton was mostly above its 115-
353 year upper quartile of flow in 2020 but mostly below its lower quartile in 2021. Other
354 years tended toward historical median flow. Total flow at the station ordered highest to
355 lowest was 9.4 km³ (2020), 7.9 km³ (2019), 6.5 km³ (2022), then 5.8 km³ (2021).

356 **3.3 TSS and DOC concentrations**

357 Downstream through the NSR basin, TSS concentrations in catchments showed a
358 step-change pattern (Figure 3). Relatively low, but variable concentrations occurred in
359 the cordillera (median of all samples $\pm$ standard error: 3.0 ± 25.6 mg L⁻¹) and foothills
360 (3.3 ± 11.1 mg L⁻¹) regions and higher concentrations in the agriculturalized plains
361 (mixed agriculture: 17.0 ± 24.0 mg L⁻¹; agriculture: 15.0 ± 20.8 mg L⁻¹). Concentrations
362 of TSS were often highest in the wetter 2019 and 2020 water years compared to the
363 drier 2021 and 2022 years, but differences occurred depending on the catchment. In



364
365 **Figure 3** Concentration boxplots of total suspended solids (TSS, upper panel) and
366 dissolved organic carbon (DOC, lower panel) by water year for all tributary catchment
367 (filled boxes) and North Saskatchewan River (NSR) mainstem (open boxes) monitoring
368 stations ordered west to east across the basin. Tributary catchments are classified by
369 landscape type (see Methods). Dashed red line indicates the analytical reporting limit
370 for each water quality parameter with the few samples below that line reported at half
371 the detection limit.



geographical agreement with the catchments, the NSR mainstem progressed from $5.3 \pm 2.7 \text{ mg L}^{-1}$ of TSS in the two upper basin stations to a more variable $8.5 \pm 32.5 \text{ mg L}^{-1}$ at the downstream plains stations. The flow regulated NSR mainstem stations reported more consistent concentrations between years compared to catchments.

Downstream concentration changes of DOC in catchments also showed a step-change, but between the cordillera catchments ($2.1 \pm 0.1 \text{ mg L}^{-1}$) and all other stations (range: $7.0 - 17.7 \text{ mg L}^{-1}$). NSR mainstem stations mirrored this pattern (upper stations: $0.9 \pm 0.1 \text{ mg L}^{-1}$; lower stations: $3.5 \pm 0.3 \text{ mg L}^{-1}$). Again, wetter years produced greater DOC concentrations compared to drier years. Variability in DOC concentrations between all sites and water years was lower compared to TSS (Figure 3).

With differing degrees of association, concentrations of TSS in most catchment and NSR mainstem stations coupled positively and statistically with flow and thus produced well-performing concentration–flow models (Table S3). An exception to this pattern occurred downstream of the dam-influenced NSR-Saunders station, whereas concentration–flow associations from the flow regulated Brazeau River were statistically stronger. DOC concentration–flow associations were also robust across the basin, though Redwater, Tomahawk, and NSR-Whirlpool Point catchments showed weaker associations compared to others.

3.4 TSS and DOC catchment yields

Catchment yields of TSS throughout the NSR basin were highly variable between water years (Table S4). All TSS yields ranged from <1 to $\sim 400 \text{ tonne km}^{-2} \text{ yr}^{-1}$ with highest observed yields from mixed agriculture, agriculture, and foothills catchments in 2020, and cordillera catchments in 2022. Punctuating this variability, TSS yields from the highest sediment yielding catchments in 2020 were approximately 18 to nearly 500 times larger than yields in the 2021 dry year. Only catchments that were regulated (i.e., Brazeau) or drained glacially influenced landscapes (NSR-Whirlpool Point) showed relatively consistent TSS yields between water years. DOC catchment yields were much smaller than TSS yields ranging from <1 to only $7 \text{ tonne km}^{-2} \text{ yr}^{-1}$ and yields were highest through the central portion of the NSR basin (Table S5). However, the variability of DOC yields between years and catchments was much lower than for TSS. For NSR



402 mainstem stations, both TSS and DOC yields were least variable between water years
403 in the headwaters and more variable in the plains regions.

404 Cordillera TSS catchment yields were statistically larger than those from mixed
405 agriculture and agriculture landscapes (Figure 4; Table S6). Foothills yields were not
406 different than those of other landscape types. These patterns were true for TSS
407 regardless of the differences in hydrometeorological conditions between water years or
408 if the regulated Brazeau catchment was included or not. Across all landscapes and
409 export model types, catchment yields of TSS were highest in the wet 2019 and 2020
410 years and both were statistically larger than yields in 2021 and 2022 (i.e.,
411 2019=2020>2022>2021). We observed increases in TSS yields that far outweighed
412 increases in flow between water years, particularly at certain foothills (e.g., Nordegg,
413 Rose), mixed agriculture (i.e., Modeste, Tomahawk), and agriculture (e.g., Strawberry,
414 Redwater) catchments (Table S7). For example, the Nordegg River in the foothills
415 reported TSS yields that were 235 times higher in 2020 compared to 2021, despite
416 runoff being only three times higher in 2020. Similarly, some cordillera rivers (i.e., Ram,
417 Clearwater) reported annual TSS yields 21 to 34 times higher in 2022 compared to
418 2021 with runoff only 1.3 times higher in 2022. TSS catchment yields were highly
419 variable (even magnitudes apart) between export models despite reporting the same
420 patterns between landscapes and water years (Figure 4), similar to other work (Kamrath
421 et al., 2023). Across the four different TSS export models, we found statistical
422 differences in yields between each model except for the Loadest and composite models
423 ($\alpha=0.05$; Table S6).

424 For DOC catchment yields, all years were statistically different following the
425 descending order 2019, 2020, 2022, and 2021 (Figure 4; Table S8). Though yields
426 showed the same patterns between landscape types as TSS, there were discrepancies
427 between years. DOC yields were not statistically different between landscape types in
428 the wetter 2019 and 2020 years, but cordillera DOC yields were higher in the drier 2021
429 and 2022 years compared to mixed agriculture and agriculture landscapes. Removal of
430 the Brazeau station did not affect these patterns. Most river stations across the NSR
431 basin reported interannual maximum to minimum DOC yield ratios of 1 to 94 (mean: 15)
432 compared to 1-57 (mean: 10) for runoff, which were much closer together relative to

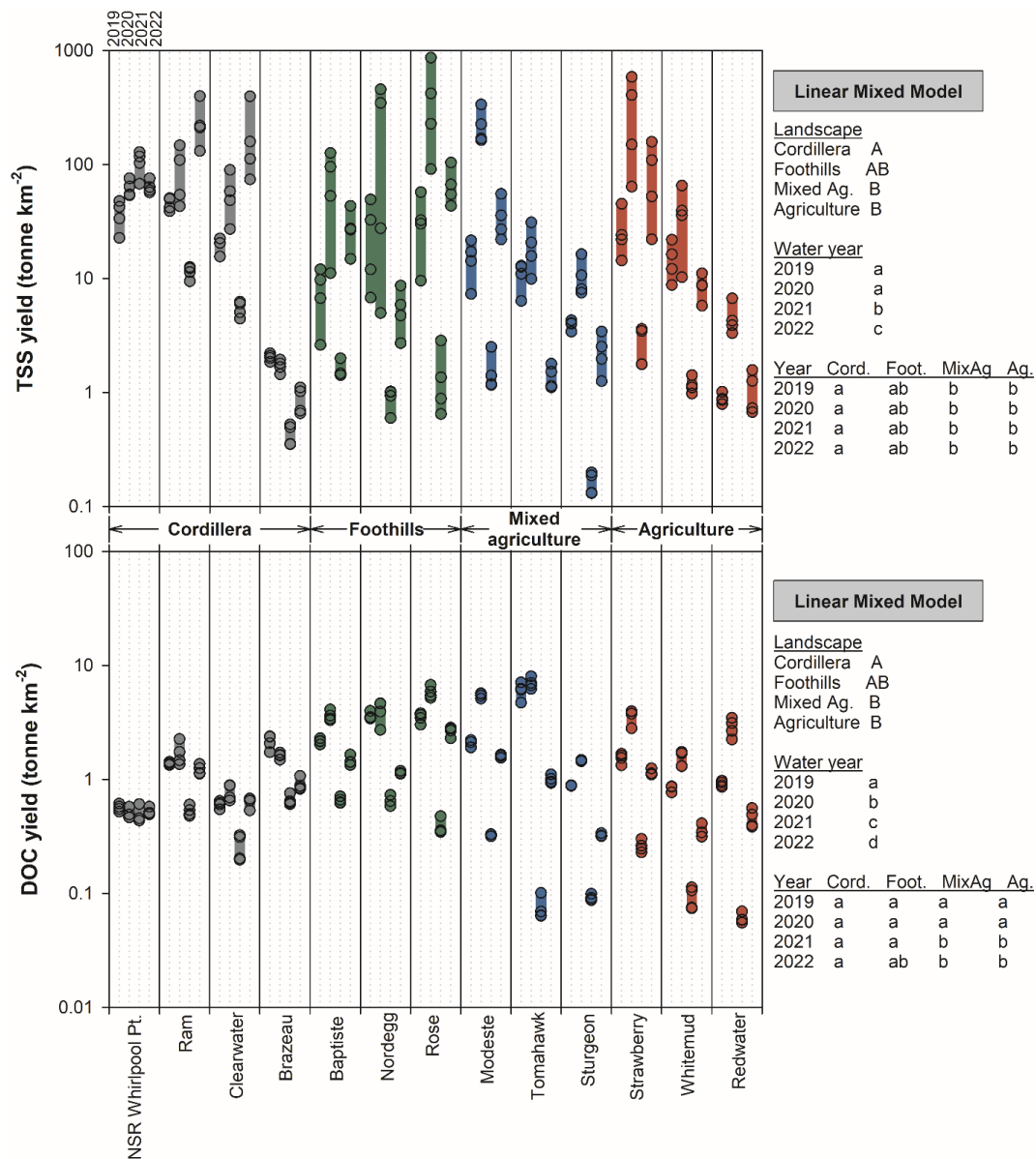


Figure 4 Total suspended solids (TSS; upper panel) and dissolved organic carbon (DOC; lower panel) catchment yields (filled circles) from water years 2019 to 2022 from tributary catchments of the North Saskatchewan River basin ordered upstream to downstream. Yield variability of each site reflects results from different *loadflex* export models (see Methods). Linear mixed model results assessing statistical differences (shown as different letters) between landscapes and water years together across all export models are shown to the right of each graph.



442 TSS (1-437; mean 95; Table S7). DOC yield estimates were much less variable
443 between export models compared to TSS results (Figure 4). DOC export models
444 reported that yields computed by composite and regression approaches, and
445 interpolation and Loadest approaches, were statistically similar (Table S8).

446 **3.5 TSS and DOC export**

447 Export of TSS through the catchment and NSR mainstem network displayed sharp
448 interannual contrasts (Figure 5; Table S9). Across all years, catchment (and
449 unmonitored land) export of TSS was highest in the near-normal 2022 (median of
450 models $\pm$ SE: $2,616 \pm 330$ ktonne), followed by the wetter years 2020 ($2,411 \pm 246$
451 ktonne) and 2019 (622 ± 23 ktonnes), then the drought 2021 year (302 ± 25 ktonne).
452 Though most tributaries showed their highest exports in the wet 2020 year, extremely
453 high sediment export from the cordillera Ram and Clearwater rivers during substantial,
454 short-term rainfall events in June dominated tributary sediment export in 2022.

455 Storage of sediment (Figure 5) in the upstream NSR-Saunders reach (189 – 864
456 ktonnes yr^{-1} , 2019 to 2022) was closely aligned with incoming delivery of sediment from
457 catchments (214 – 892 ktonnes yr^{-1}). This pattern was true even during substantial
458 glacier melt-related transport of sediments during the 2021 heat dome event. The
459 behaviour of Abraham reservoir upstream of the Bighorn Dam as a major sediment sink
460 is consistent with this overall relationship between inputs and storage.

461 The NSR-upstream Clearwater reach showed variable inputs of sediments from
462 catchments each year (30 – 630 ktonne yr^{-1}), with highest export in 2022. From 2019 to
463 2020, there was storage of sediment in the river (30 – 44 ktonne yr^{-1}) before a modest
464 release of sediment in 2021 (16 ktonne yr^{-1}). In 2022, sediment contributed by the Ram
465 River and unmonitored regions after June storms (630 ktonne yr^{-1}) was mostly stored in
466 this reach of the NSR (472 ktonne yr^{-1}).

467 Through the lower-slope NSR-Devon reach in 2019 to 2021, addition of sediment to
468 the river from stores (48 – 1,994 ktonne yr^{-1}) outpaced catchment sediment additions by
469 1.2 to 2.6 times, meaning a majority of suspended sediment in this reach of the NSR
470 was sourced from areas within and adjacent to the river itself. The high-flow year of
471 2020 resulted in major additions of sediment to the river from in-river sources. In 2022,



Figure 5 Total suspended solids (TSS) export arrow diagrams of the North Saskatchewan River (NSR) basin for the water years (November to October) of 2019 to 2022. Tributary and storage exchanges are represented by horizontal arrows and NSR mainstem stations and reaches are shown as vertical arrows. Alternating shading delineates NSR mainstem reaches. Relative station spatial separations not to scale.

contributions of sediment from Clearwater R. and other tributaries were stored in the reach (487 ktonne yr⁻¹), in contrast to previous years.

Finally, through the NSR-Pakan reach, the banks and channels of the NSR itself contributed 205 and 834 ktonne yr⁻¹ of sediment in 2020 and 2022, respectively, while 2019 reported 53 ktonne yr⁻¹ of sediment storage through the reach while 2021 had storage of 70 ktonne yr⁻¹. Across all NSR mainstem stations, sediment storage occurred



484 mostly in the upper two station reaches (7/8 occurrences) compared to sediment losses
485 in the two downstream reaches (5/8 occurrences). Overall, catchment export relative to
486 the furthest downstream mainstem station (NSR at Pakan) was 50-66% in the wet 2019
487 and 2020 years up to 262% (2022) and 347% (2021) during the normal to drier years.

488 In contrast to the highly variable and active storage terms year to year for TSS,
489 export of DOC downstream through the NSR basin showed more consistent patterns
490 (Figure 6; Table S10). In upper reaches, export of DOC from tributaries was relatively
491 low but consistent between years delivering 1.4 to 6.4 ktonne yr⁻¹ following the
492 interannual pattern 2020>2019>2022>2021. Consistent storage of DOC in the dam-
493 influenced NSR-Saunders reach each water year mirrored TSS patterns. DOC was
494 stored through the NSR-upstream Clearwater reach but storage was nearer to net zero
495 compared to the NSR-Saunders reach. Through the NSR-Devon reach, DOC was
496 stored except during the wet 2020 season when DOC was added to the river. The NSR-
497 Pakan reach reported near zero storage in 2019 before consistent DOC addition to the
498 river (1.1 – 2.8 ktonne yr⁻¹) from 2020 to 2022. Overall, cycling of DOC between the
499 NSR mainstem and its stores within and adjacent to the river was less consequential
500 than for TSS, resulting in relatively consistent tributary contributions of DOC relative to
501 NSR at Pakan station of between 98 and 128%.

502 **4. Discussion**

503 ***4.1 Sediment mobility driven by site-specific precipitation and erosion thresholds***

504 Numerous studies have demonstrated that different catchment characteristics can
505 influence downstream water quality (Allan, 2004; Lintern et al., 2018b; Shi et al., 2024).
506 These landscape to water quality connections require a source of landscape material in
507 the catchment, precipitation and runoff, entrainment or dissolution of material, and
508 downstream delivery. Typically, we expect steep cordillera regions to have colluvial
509 sediment stores of large grain sizes that produce fewer suspended sediments in rivers
510 (Sinha and Friend, 1994; Townsend-Small et al., 2008) compared to thicker,
511 unvegetated soils of erosion-prone agricultural plains and urban environments (Giri,
512 2021). Forested landscapes often have moderate soil stores resistant to erosion due to
513 vegetation cover (Veldkamp et al., 2020). This pattern is reflected in the NSR basin by

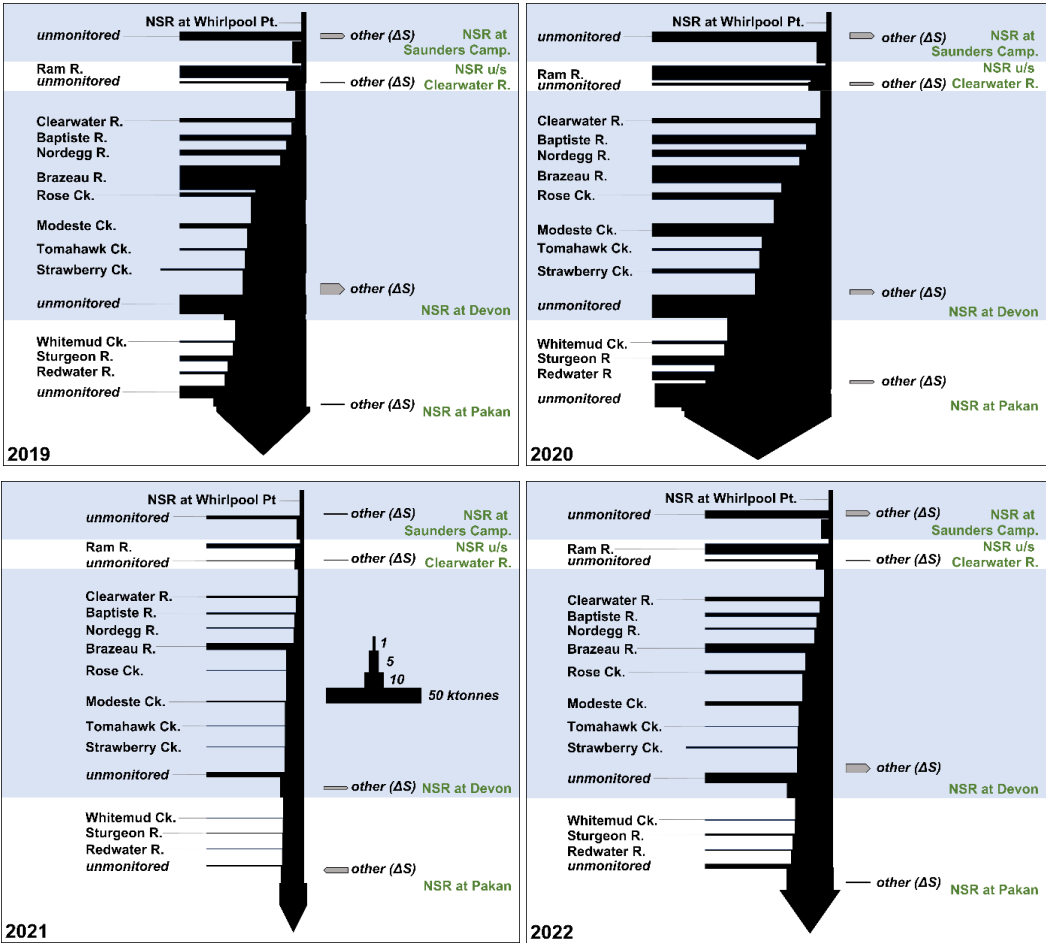


Figure 6 Dissolved organic carbon (DOC) export arrow diagrams of the North Saskatchewan River (NSR) basin for the water years (November to October) of 2019 to 2022. Tributary and storage exchanges are represented by horizontal arrows and NSR mainstem stations and reaches are shown as vertical arrows. Alternating shading delineates NSR mainstem reaches. Relative station spatial separations not to scale.

differences observed between lower concentration cordillera regions and those downstream (Figure 3). However, sediment yields were highest from the wetter cordillera areas, pointing to precipitation and runoff controlling sediment mobility across the basin (i.e., transport limited; Su et al., 2023). This finding is supported by strong and positive TSS concentration-flow (C-Q) relationships at most stations and suggests there



525 is ample sediment supply across the basin but hydrological energy is required to
526 mobilize it (Creed et al., 2015), as observed elsewhere in Alberta (Loiselle et al., 2020).

527 Statistical differences in TSS yields and variability in sediment export between water
528 years (Figures 4, 5) reflected the variability in precipitation and runoff conditions
529 observed across the NSR basin during the study. Similarly, TSS yield patterns across
530 landscape types were the same across all water years, also suggesting
531 hydrometeorological conditions were the dominant control on sediment yields. During
532 the open water period in central Alberta (May-October), rainfall frequency and intensity
533 is related to hourly-scale, patchy convective rainfall events and daily-scale low pressure
534 systems from the west and south (Vickers et al., 2001). These drivers contribute to a
535 mosaic of moisture conditions across the basin (Figure S2). These episodic processes
536 lead to shorter-term hydrologic connections across landscapes and subsequent erosion
537 that may mobilize a substantial portion of a river's annual export of material. In between
538 events, or during dry years of this continental basin, evapotranspiration is strong,
539 leading to desiccation of landscapes, prolonged lack of hydrologic connectivity, and
540 higher groundwater contributions to rivers relative to surface flow (Tetzlaff et al., 2024).
541 This juxtaposition of conditions at least partially drives the variability of sediment yields
542 and export between years and between rivers across the basin. However, rainfall and
543 runoff accumulation does not fully explain TSS yield variability in space and time.

544 Though transport-limitation and hydrometeorology drive sediment yields in both
545 catchments and the NSR proper, we observed non-linear (i.e., threshold) response of
546 sediment yields in selected tributaries relative to changes in flow, which is commonly
547 observed in rivers (Vercruysse et al., 2017). This is consistent with rapid connection,
548 disconnection, and reconnection events that occur between runoff and sediment stores
549 in catchments (Fryirs, 2013). The striking sediment yield increases in 2020 in selected
550 foothills and plains catchments and in 2022 in selected cordillera catchments
551 demonstrates the variability (i.e., site specificity) of this threshold behavior. Threshold
552 response also helps explain wide variability in model results (Figure 4) as extreme high
553 events perturb model fits at the high end of C-Q associations (Asselman, 2000).
554 Further, our nested monitoring program allowed us to observe how alluvial sediment
555 stores in major rivers interact with flow. Downstream through the NSR mainstem



556 reaches, we observed mostly loss of sediment stores due to erosion within the NSR
557 (2019-2021) before a reversal to sediment storage (2022; Figure 5). In 2020, a flow
558 threshold was crossed where bank and bed erosion were more intense in the NSR-
559 Devon reach compared to other reaches and years. These results demonstrate how
560 mainstem flows begin to erode and entrain in-channel and marginal sediment stores
561 and mobilize sediment downstream as an erratic “conveyor belt” (Fryirs, 2013). This
562 behavior is particularly robust in headwater basins that transition from mountainous,
563 highland-type landscapes to low-relief plains regions. This type of dynamic sediment
564 transport is also typical across deglaciaded landscapes like the NSR basin.

565 Accumulated glacial tills are often unstable and erodible and serve as a persistent
566 source of sediment to rivers, elevating sediment yields in an effect described as
567 paraglacial sedimentation (Ballantyne, 2002; Church and Ryder, 1972). Downstream
568 delivery of glacial derived sediment reduces sharply over time as glacial ice retreats and
569 easily mobilized sediment is either exhausted or becomes inaccessible to the fluvial
570 system. Ultimately sediment yields decline, marking the end of the paraglacial period.
571 However, fluvial reworking can create secondary stores of glaciogenic sediments
572 downstream and when coupled with extra-channel stores, can become ongoing sources
573 of sediment for much longer periods (Ballantyne, 2002; Orwin and Smart, 2004). This
574 sediment may then mobilize under certain conditions when kinetic energy thresholds
575 are crossed, typically during isolated, intense, short-term hydrological events. These
576 types of threshold conditions are now appearing in other northern landscapes that are
577 experiencing enhanced terrestrial-aquatic change in particulate and dissolved material
578 export due climatic shifts (Beel et al., 2020).

579 ***4.2 Organic matter mobility closely reflects broad hydrologic connectivity***

580 Yields and export of DOC across the NSR basin generally showed similar patterns
581 to TSS but with some notable differences. Riverine DOC is sourced from upstream
582 hydrological flushing of organic soil horizons, wetlands, and riparian areas (Neary et al.,
583 2009). From a DOC source perspective, weaker cordilleran C-Q associations and sharp
584 increasing concentrations out of the mountains, compared to TSS, suggest weaker
585 transport-limitation toward neutral or source-limitation behavior (Creed et al., 2015; Su
586 et al., 2023). Though sediment is sourced in cordillera areas either through glacial



587 inputs, colluvial processes, or secondary scouring of channels and beds, Canada's
588 Rocky Mountains have typically sporadic and thin organic soil and wetland cover in
589 valleys (Hoffmann et al., 2014). The cordillera NSR-Whirlpool station, for example,
590 showed a nearly flat C-Q association, meaning increasing flow in rivers in this relatively
591 wet region did not result in more incoming DOC from upstream catchments. Other
592 cordillera rivers had stronger C-Q associations but did not produce high concentrations
593 of DOC, likely owing to thin soils and marginal wetland coverage in downstream
594 portions of their catchments. Foothills regions have some wetland coverage and are
595 underlain by more developed luvisolic soils with shallow organic layers and thus have
596 greater organic matter stores compared to Rocky Mountain catchments. Further
597 downstream, deeper chernozemic soils of the agriculturalized plains have very large
598 stores of organic matter. However, similar to TSS, DOC yields were statistically highest
599 from wetter cordillera catchments and lowest in those of drier agricultural landscapes
600 despite clear differences in organic matter stores. Again, these results point to
601 hydrometeorology playing the dominant role in DOC mobility downstream, but with a
602 notable difference.

603 DOC yields across the basin, relative to TSS, reported less variability between
604 years for a given river, between rivers of a specific landscape type, and between rivers
605 across landscape types (Figure 4). DOC yields between landscapes were also different
606 for wetter (2019 and 2020) versus drier (2021 and 2022) conditions. Maximum to
607 minimum DOC yield ratios were similar to patterns reported for runoff (Table S7). These
608 results point to straightforward hydrologic connectivity driving DOC yields across the
609 basin compared to the erosive threshold behavior observed for TSS yields that
610 contributed to high variability within landscape types. Rain and runoff increase
611 hydrologic contact with shallow organic soils and wetlands, through runoff and interflow,
612 resulting in more DOC delivery to rivers (Neary et al., 2009; Ward et al., 2017). The
613 pulse-shunt process concept (Raymond et al., 2016) similarly outlines how
614 hydrologically active years in catchments flush organic matter from soils and transport it
615 downstream quickly and efficiently to higher order rivers. This process can result in
616 hydrologic connection to isolated sources of organic material across the landscape
617 before disconnection during drier years. This process also requires contact and gravity



618 flushing, rather than extra energy required to entrain particulate material. This concept
619 also demonstrates why model performance was better for DOC than TSS as flushing
620 produces less variability than erosion, particularly at the high end of C-Q associations.

621 Through the NSR mainstem, there was little evidence in most years that riparian
622 areas, or other sources of DOC within and adjacent to the mainstem, influenced
623 downstream DOC mobility. Rather, the NSR was more of a conduit of DOC sourced
624 from tributary catchments. Only in the years when wet conditions were widespread
625 (2020) did we observe likely riparian additions of organic matter from regions adjacent
626 to the NSR mainstem, though these inputs were still inconsequential.

627 Ultimately, in support of our hypothesis, interannual differences in
628 hydrometeorological conditions controlled both TSS and DOC mobility through the NSR
629 basin, but via differing processes (i.e., threshold erosion versus simple flushing).
630 However, we did not expect that inter-catchment variability of TSS and DOC yields
631 within a given landscape type would blur differences between obviously different
632 landscape types as determined statistically using high-resolution geospatial datasets.
633 This, therefore, points to the highly variable nature of dry, continental climates such as
634 those in Alberta, the ongoing legacy of past glaciation on sediment availability, and may
635 partially explain why landscape conditions here show a less measurable role in
636 mobilization of river material, compared to other locations (Lintern et al., 2018a).

637 ***4.3 Implications for water users within a changing river***

638 The eastern slopes of the Canadian Rocky Mountains are a critical “water tower” for
639 approximately 5 million people and Canada’s centre of agricultural activity. This region
640 is also experiencing substantial changes in climate and river flow related to changes in
641 atmospheric circulation, glacial loss, seasonal shifts, and other hydrometeorological
642 factors (Anderson and Radić, 2020; Moore et al., 2009; Newton et al., 2021; Rood et al.,
643 2008). Climate modelling of conditions across the NSR basin for 2061–80 using data
644 from Canada’s Changing Climate Report predicts mostly unidirectional increases in
645 precipitation and temperature across most seasons (Figure S3; Bush and Lemmen,
646 2019). For example, across the basin, precipitation is expected to increase
647 approximately 5–25% relative to reference conditions (1986–2005) during winter and
648 open-water periods. Over the same period, seasonal air temperatures are expected to



649 increase 3.5 to 5.5 °C throughout the basin, potentially inducing greater humidity and
650 more intense convective and cyclonic rainfalls during the open water season. This
651 heating may also cause more intense droughts when atmospheric cycles like El Niño
652 induce drier conditions across western Canada (Asong et al., 2018).

653 Assuming similar river and catchment processes in the future, our results suggest
654 that the NSR basin may experience increased mobilization of TSS and DOC in
655 tributaries and the NSR mainstem in response to a wetting climate. A warming climate
656 may also increase glacial runoff and associated glacial sediment mobilization
657 (Staniszewska et al., 2021). For example, the relatively hot conditions that the NSR-
658 Whirlpool catchment endured in 2021 resulted in relatively greater storage of water and
659 sediment in the Bighorn Dam. Given the controlling factors of rainfall location, intensity,
660 and specific catchment conditions on TSS mobility, increasing precipitation and river
661 flow may create more turbid river systems whose conditions and behaviour are difficult
662 to predict spatially and temporally. DOC mobilization would be expected to be more
663 predictable based on open water season rainfall accumulations. Regardless,
664 mobilization of more turbid and organic-rich waters may create issues for aquatic
665 ecosystem health and human water users. Prolonged turbidity is a known stressor on
666 ecosystem health due to sediment burial and scouring processes that impact
667 periphyton, benthic invertebrates, and fish health and reproduction (Wood and
668 Armitage, 1997). Sharp changes in DOC concentrations and character can upset heat
669 and carbon cycling of local ecosystems (Spencer et al., 2012). From a human use
670 perspective, turbidity and organic matter can provide substantial challenges for the cost-
671 effective production of treated drinking water (Price and Heberling, 2018). Though
672 turbidity is effectively reduced in conventional treatment methods, higher turbidity
673 requires more alum additions (and therefore higher cost) to settle particles (Farhaoui
674 and Derraz, 2016). Similarly, high DOC in raw river water can require additions of
675 powdered activated carbon to avoid development of potentially hazardous disinfection
676 byproducts in treated water (Delpla et al., 2009); again at higher costs.

677 In terms of managing human-related pollution effluents into rivers, many
678 jurisdictions use a total maximum daily load (TMDL) approach to ensure effluent
679 additions do not stress aquatic ecosystems. TMDLs integrate data from point and non-



point sources of pollution in a river system to calculate how much of a concerning substance can be delivered to river systems but still meet water quality guidelines (USEPA, 2024). Turbidity and DOC can be important hosts binding metals and other contaminants of concern, including mercury, within their matrices (Delpla et al., 2009). Increases in natural background loading of these materials could have implications for the quality, release frequency, and volume of industrial and municipal effluents delivered to the NSR in the future. Similarly, climate change is expected to intensify wildfire activity across northern landscapes, potentially delivering greater amounts and different forms of material downstream with potential implications for TMDLs, and water users in general (Goode et al., 2012; Hanes et al., 2019).

Considering the potential for greater delivery of TSS and DOC downstream through the NSR and associated challenges for water users, it is important to contextualize these results from a source water protection perspective. In the NSR, source water protection narratives (EPCOR, 2020) focus on the cordillera and foothills regions of the basin, where most of the water in the NSR at Edmonton is sourced (Government of Canada, 2024b; Newton and Taube, 2023) and where future land use changes are anticipated to occur. Currently, foothills and cordillera regions are less impacted by human activities relative to downstream agriculturalized plains regions (Alberta Biodiversity Monitoring Institute, 2016) and contribute water low in sediment and organic matter to the NSR. Climate projections across the NSR basin expect weaker precipitation accumulation increases in the mountains and foothills relative to the agriculturalized plains (Bush and Lemmen, 2019) and long-term loss of glacial meltwater inputs to the NSR (Anderson and Radić, 2020). Coupled with expected future land use pressures in these regions means that it will be critical for land use decisions in the water-producing regions of the NSR be made in an informed manner, guided by data, and maintains ecosystem services (Palmer et al., 2009).

5. Conclusion

Export of material from major rivers is a key environmental metric to identify and quantify landscape response(s) to changes in climate and anthropogenic activities. However, proper use of this metric requires integrated monitoring programs that assess landscape heterogeneity across a river basin over long periods of time. This approach is



711 required to characterize complex relationships between landscapes, inter- and intra-
712 annual hydrometeorological variability, and how rivers mobilize and store material.

713 Using our multi-annual, representative monitoring design for the NSR basin, we
714 showed that catchment yields of TSS and DOC have significant variability not only
715 between catchments of different structural/landscape conditions, but also between
716 those that are similar. This pattern was consistent whether occurring within pristine or
717 disturbed landscapes. Significant variability in interannual yields also occurred and was
718 complicated by the influence of in-channel material stores and erosion threshold
719 behaviour in specific catchments.

720 The NSR in western Canada supplies water to over 1.5 million Canadians and is
721 experiencing change due to continued loss of glacial meltwater inputs, greater heating
722 and rainfall accumulations, and increasing human populations. Our study suggests that
723 NSR waters are expected to become more turbid and enriched in dissolved organic
724 material under this climate regime and will likely cause challenges for water users (e.g.,
725 drinking water utilities), river management strategies, and aquatic biota. However,
726 broader application of monitoring approaches that capture landscape heterogeneity,
727 replicate monitoring of diverse catchment conditions, and include entire river basins in
728 an integrated fashion will provide improved data and understanding of river systems and
729 ultimately better decisions when managing water resources.

730

731 Data availability statement:

732 All data presented in this manuscript are publicly available for download from: 1. Government of Alberta
733 water quality portal (<https://environment.extranet.gov.ab.ca/apps/WaterQuality/dataportal/>) or by request
734 (swq.requests@gov.ab.ca); 2. Water Survey of Canada hydrometric data
735 ([https://www.canada.ca/en/environment-climate-change/services/water-](https://www.canada.ca/en/environment-climate-change/services/water-overview/quantity/monitoring/survey.html)
736 [overview/quantity/monitoring/survey.html](https://www.canada.ca/en/environment-climate-change/services/water-overview/quantity/monitoring/survey.html)); and 3. Alberta River Basins portal (<https://rivers.alberta.ca/>).
737

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