



Implications of VOC Oxidation in Atmospheric Chemistry: Development of a Comprehensive AI Model for Predicting Reaction Rate Constants

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Abstract. Volatile Organic Compounds (VOCs) significantly influence global atmospheric chemistry through oxidative reactions with oxidants. These reactions produce key precursors to the formation of atmospheric fine particulate matter (PM_{2.5}) and ozone (O₃), which in turn play a crucial role in regulating O₃ pollution and reducing PM_{2.5} concentrations. With the increasing diversity of VOCs, the need for advanced modeling techniques to accurately estimate the atmospheric oxidation reaction rate constants (k_i , where $i \in \{\bullet\text{OH}, \bullet\text{Cl}, \text{NO}_3, \text{or O}_3\}$) has become more urgent. Here we introduce Vreact, a Siamese message passing neural networks (MPNN) architecture that jointly models VOC–oxidant reactivity. The model simultaneously predicts $\log_{10}k_i$ values and achieves a mean squared error (MSE) of 0.299 and a coefficient of determination (R^2) of 0.941 on the internal test set. This framework overcomes the single-oxidant constraint of traditional models, enabling unified and scalable prediction of VOC oxidation kinetics across multiple oxidants. An interactive web tool (<http://vreact.envwind.site:8001>) is provided to facilitate non-expert access to reactivity screening. Vreact offers valuable insights into the formation and evolution of atmospheric pollutants, and serves as a critical resource for developing effective control and emission strategies, ultimately supporting global efforts to mitigate air pollution and improve public health.

1 Introduction

The rapid advancement in data-driven methodologies has revolutionized various fields, such as protein structure prediction (Abramson et al., 2024), molecular generation (Zhang et al., 2023), organic reaction prediction (Burés and Larrosa, 2023), and bioinformatics (Theodoris et al., 2023). Environmental challenges, particularly those associated with atmospheric chemistry and climate change, have also benefited from these innovations. As pollutants evolve under both anthropogenic and natural influences, the understanding of their chemical and physical properties has become increasingly vital for addressing global air quality and climate issues. Volatile Organic Compounds (VOCs) are organic chemicals that readily



vaporize at ambient temperature, contributing significantly to the complexity of atmospheric processes. Sources of VOCs are both natural and anthropogenic, with human activities such as industrial production, petrochemical processing, and vehicle exhaust contributing to the emission of a variety of VOCs. Additionally, biosphere sources, such as plants and forests, release compounds like isoprene and monoterpenes, which further complicate atmospheric VOC dynamics (Qin et al., 2021; Sindelarova et al., 2014). These highly reactive VOCs drives critical atmospheric reactions, such as the formation of ozone and secondary organic aerosols (SOA), and significantly contribute to environmental pollution. For instance, VOCs interact with nitrogen oxides (NO_x) and radicals to form tropospheric O₃ and SOA (Finlayson-Pitts and Pitts, 1997; Hallquist et al., 2009; Han et al., 2018; Zhang et al., 2020; Ziemann and Atkinson, 2012). The role of VOCs in the formation of secondary pollutants such as PM_{2.5} (Huang et al., 2014; Zhao et al., 2015) and O₃ is a growing concern due to the adverse impacts on human health (Kamarrudin et al., 2013), including respiratory diseases, cardiovascular conditions, and overall mortality. The dynamic interactions between VOCs and atmospheric oxidants determine the persistence and transformation of these pollutants, which in turn influence their contribution to global haze, photochemical smog, and acid deposition.

VOCs undergo degradation and removal from the troposphere through diverse mechanisms driven by atmospheric oxidants. During the daytime, •OH radicals serve as the primarily oxidants, facilitating rapid VOC oxidation. At night, however, the concentration of •OH decreases sharply due to the lack of photochemical reactions, shifting the dominant oxidation pathways to NO₃ radicals and O₃. The reaction rates of VOCs with •OH are approximately 30 times faster than those with NO₃ radicals, with NO₃ radicals, significantly influencing the spatial and temporal variation of the atmosphere's self-cleaning capacity and the formation of organic aerosols (Palmer et al., 2022; Zha et al., 2023). For example, regions with high isoprene concentrations often reflect differences in its reaction products and rates with •OH and NO_x rather than solely high emissions (Wells et al., 2020). Additionally, the structural diversity of VOCs determines their reaction mechanisms, influencing reaction rates. Highly reactive compounds such as alkenes, multi-substituted aromatics, and phenols exhibit higher reaction rates, whereas alkanes, alkyl nitrates, and ketones demonstrate relatively low reactivity (Ziemann and Atkinson, 2012). These variations underscore the significance of atmospheric oxidation reaction rates as key indicators of the persistence of organic pollutants in the atmosphere. Accurate assessment of these rates is essential for understanding the fate of VOCs, elucidating SOA formation processes, and addressing global challenges related to PM_{2.5} and ozone development.

Given their importance, accurately predicting the atmospheric oxidation rates of VOCs is critical for understanding their persistence, transformation, and contribution to secondary pollutant formation. Traditionally, such predictions have relied on experimental kinetic modeling methods and quantitative structure-activity relationship (QSAR) approaches (Basant and Gupta, 2018; Liu et al., 2021). Experimental methods involve tracking reactant and product concentrations using techniques like chemical ionization mass spectrometry (CIMS), followed by kinetic fitting to determine Arrhenius parameters (Logan, 1982; Wells et al., 1996). However, these methods are time-consuming and cover only a narrow subset of atmospheric VOCs. QSAR models offer a scalable alternative by leveraging molecular descriptors and statistical learning. Notable examples include AOPWIN™ module integrated in US EPI Suite™ software, which applies Partial Least Squares (PLS) regression to 109 gas-phase reaction with hydroxyl radicals (Atkinson, 1986, 1987; Kwok and Atkinson, 1995), and later expansions using



a broader dataset (Öberg, 2005). Some models have also incorporated machine learning algorithms such as multiple linear regression (MLR) (Liu et al., 2020, 2022) for predicting reactions with NO_3 and $\bullet\text{OH}$ and artificial neural networks for predicting reactions with O_3 (Fatemi, 2006). Despite their utility, these models generally rely on predefined descriptors and are typically limited to reactions with a single type of oxidant. Recent advances in deep learning (DL), particularly graph
70 neural networks (GNN), have improved molecular representation by learning features directly from molecular graphs. This enables more flexible and accurate prediction of chemical properties without requiring predefined descriptors. GNNs have been successfully applied in atmospheric chemistry and other fields tasks, such as in predicting vapor pressures with GC^2NN (Krüger et al., 2025) and modeling reaction rate constants involving with $\bullet\text{OH}$ using GAT–GIN hybrid architectures (Huang et al., 2024). However, like traditional models, these GNN-based frameworks have been developed for single-molecule
75 systems and thus fall short in capturing the complexity of multi-molecule reactions in real environments. In contrast, the atmosphere involves competing and sequential reactions between VOCs and multiple oxidants— $\bullet\text{OH}$, NO_x , $\bullet\text{Cl}$, and O_3 —depending on time of day, region, and chemical conditions. This multiplicity underscores the urgent need for models that can simultaneously learn and predict VOC reactivity across multiple oxidants. To meet this need, message passing neural networks (MPNN) offer a powerful framework (Gilmer et al., 2017). MPNNs propagate information across molecular graphs,
80 capturing both atomic-level features and topological context. Extensions of MPNN, such as the communicative GraphRXN (Li et al., 2023) and directed MPNN Chemprop (Heid et al., 2024), have shown promise in learning reactivity across multiple reactants. Yet, their application has largely focused on synthesis or materials chemistry, not atmospheric multiphase oxidation.

This study addresses this gap by proposing Vreact, a novel Siamese MPNN architecture capable of jointly modelling
85 reactions between VOCs and four major atmospheric oxidants. Unlike previous models that treat each oxidant independently, Vreact processes VOC–oxidant pairs in a unified framework, it learns representations from the molecular graphs of VOCs and oxidants through the MPNN, and encodes their interactions *via* feature aggregation. This design enables the model to accept arbitrary VOC–oxidant combinations and simultaneously predict reaction rate constants k_i (where $i \in \{\bullet\text{OH}, \bullet\text{Cl}, \text{NO}_3, \text{O}_3\}$). Compared to traditional single-oxidant prediction models, Vreact shows significantly improved performance,
90 achieving higher accuracy and broader generalizability across multiple oxidants. Furthermore, the model’s interaction module captures atomic-level interaction patterns, providing mechanistic insights into VOC oxidation process *via* interpretable interaction weight matrices. Applying Vreact to 447 atmospheric VOCs not included in the training data revealed a wide distribution of oxidation reactivities and confirmed that alkenes and aromatics exhibit higher reactivity, acting as key precursors for ozone and SOA formation.



95 2 Methods and Data

2.1 Collection and Preprocessing of Reaction Rate Constant Dataset

The VOCs reaction rate constant dataset compiled by McGillen et al. is utilized in the study, which includes gas-phase reaction rate constants of natural atmospheric VOCs, halocarbons, and their degradation products with $\bullet\text{OH}$, $\bullet\text{Cl}$, NO_3 radicals, and O_3 , within a temperature range of 250-370K (McGillen et al., 2020). Under thermodynamic standard conditions at 298K, a total of 2802 gas-phase reaction rate constant data points were obtained, encompassing 1586 VOCs and 4 oxidants. This dataset includes k_i values for 1363 VOCs with $\bullet\text{OH}$, 735 VOCs with $\bullet\text{Cl}$, 393 VOCs with NO_3 radicals, and 311 VOCs with O_3 . Due to the wide range of reaction rate constants k_i in the dataset ($1.460 \times 10^{-21} \sim 7.550 \times 10^{-10} \text{ cm}^3/(\text{molecule} \cdot \text{s})$, S.D. $= \pm 1.040 \times 10^{-10}$), the data were log-transformed to $\log_{10} k_i$ to reduce skewness and mitigate the influence of outliers on the model. To ensure a balanced distribution of each type of oxidant in the training, validation, and internal test sets, the dataset was divided using stratified random sampling into training, validation and internal test sets in an 8:1:1 ratio (Table S1).

2.2 Construction and Training of the Vreact Model

All VOCs and oxidant molecules were converted into graphs $G(V, E)$ (Text S1). The generated molecular graph G includes ten types of atomic information for each non-hydrogen atom, such as element type, chirality, and atomic hybridization type, as well as four types of bond information, including bond type and conjugation (Table S2). A Siamese MPNN architecture-Vreact, was designed to simultaneously accept input features of VOCs and oxidant molecules (Fig. 1). The model takes the SMILES of VOCs and oxidants as input and primarily includes a VOC molecular graph representation layer and a MPNN layer, an oxidant molecular graph representation layer and MPNN layer, an interaction layer, and a prediction layer. The molecular graph $G(V, E)$ encoding layers of VOCs and oxidants containing node feature matrix X and edge feature matrix Y , which learn molecular properties through the MPNN layer (Gilmer et al., 2017). The MPNN forward propagation process consists of two phases: Message Passing Phase and Readout Phase and generates molecular feature tensors A for VOCs and B for oxidants. Subsequently, the interaction layer transforms the molecular features A of VOCs and B of oxidants into tensors A_I and B_I of the same shape and concatenates them into tensor Z . Reaction rate constants are determined not only by the molecular structure of the reactants but also by the interactions between the reactants. The interaction feature tensor I is dot-multiplied with A to obtain the oxidant-affected VOC feature tensor A' ; similarly, it is dot-multiplied with B to obtain the VOC-affected oxidant feature tensor B' . These operations embed the learned interaction features into the molecular structure features, providing a more comprehensive representation of the chemical reaction mechanisms between the two reactants. The prediction phase is composed of a pooling layer and three fully connected layers. The pooling layer uses the Set2Set method to achieve global average pooling, and the fully connected layers map the input features to the final predicted values ($\log_{10} k_i$). More details can be found in Text S2.



During model training, Adaptive Moment Estimation (Adam) (Kingma and Ba, 2017) was employed to address the fixed learning rates issue in traditional gradient descent methods. Adam adaptively adjusts the learning rate of each parameter using first-order moment estimates (mean of the gradients) and second-order moment estimates (exponentially moving average of the uncentered variance of the gradients), aiding in rapid model convergence. Bayesian optimization was utilized for hyperparameter tuning, which included the initial learning rate of the optimizer (lr), batch size, L2 regularization parameter (*weight decay*), dropout rate (p), and MPNN time steps (T) (Text S3). After identifying the optimal hyperparameter combination (Table S3) on the validation set, and the best model was saved. The predictive performance of the model was assessed using Mean Squared Error (MSE), Root Mean Square Error (RMSE), Mean Absolute Error (MAE), and coefficient of determination (R^2) (Text S4). For more information on the model implementation, please refer to Text S5.

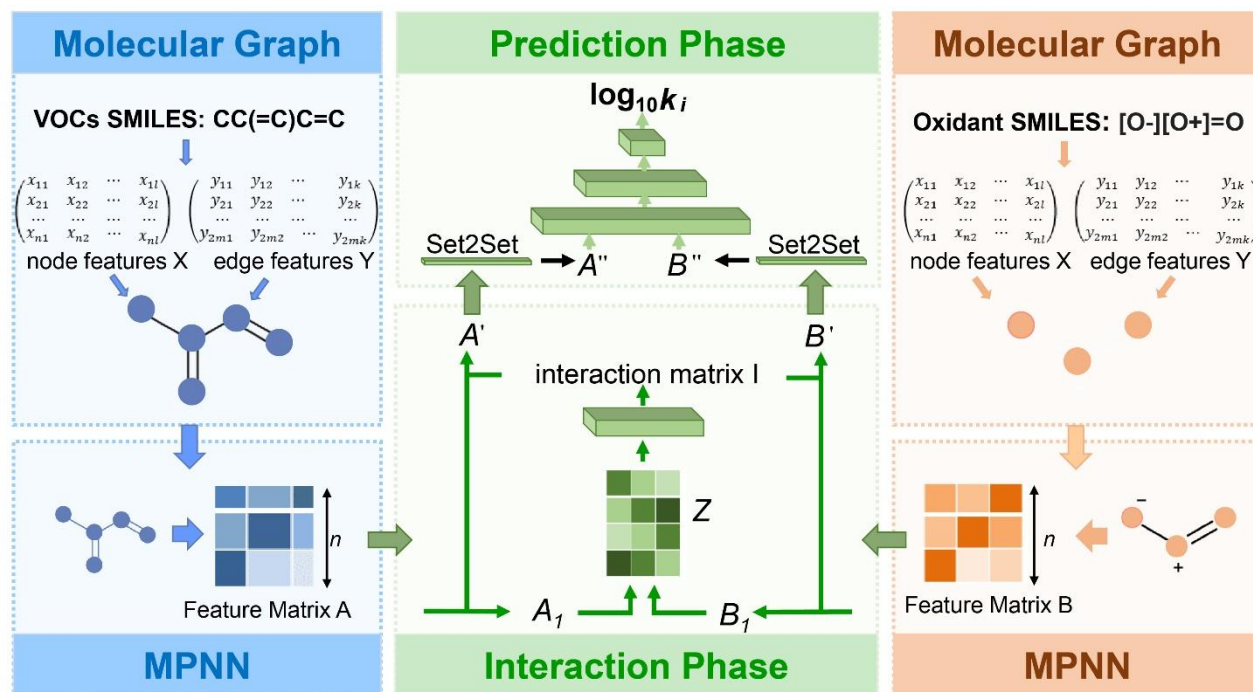


Figure 1. Schematic of the Vreact Architecture. SMILES of VOCs and oxidants are converted into molecular graphs, where nodes represent atoms and edges represent bonds. Atomic and bond features form matrices X and Y. Using a Siamese MPNN architecture, the Vreact model processes these features through separate MPNN layers for VOCs and oxidants. The final prediction layer outputs $\log_{10}k_i$, incorporating both molecular and interaction features.



3 Results and Discussion

3.1 Analysis of VOC and Oxidant Reaction Data Distribution and Characteristics

The categories and distribution characteristics of VOC and oxidant reaction data are first explored in the study, which includes $\log_{10}k_i$ data for 1586 VOCs with $\bullet\text{OH}$, $\bullet\text{Cl}$, NO_3 , and O_3 (Fig. 2A). The dataset contains the most data for $\bullet\text{OH}$, accounting for 48.64% of the total, as $\bullet\text{OH}$ plays a crucial role in the atmosphere, rapidly reacting with organic pollutants and dominating their removal process. The remaining data points are for $\bullet\text{Cl}$ (26.23%), NO_3 (14.03%), and O_3 (11.1%) in descending order of data quantity. O_3 is primarily produced through photochemical reactions involving NO_x and VOCs, while NO_3 , as the principal nighttime atmospheric oxidant, significantly contributes to the oxidation and removal of trace gases. The dataset encompasses VOCs with diverse chemical structures, including 22 functional groups such as double bonds, esters, benzene rings, and halogen atoms (F, S, Cl, Br, and I) (Fig. 2B). This extensive chemical structure space facilitates the model's ability to learn more structural features and enhances its generalization capability.

Moreover, although there is some overlap in the reactions of the four oxidants with VOCs, each oxidant also has specific VOC reactions (Fig. 2C). There are 747 VOCs with k_i data for only one oxidant and 839 VOCs with k_i data for multiple oxidants, of which 81 VOCs have data for all four oxidants. For example, isoprene can react with $\bullet\text{OH}$, NO_3 , and $\bullet\text{Cl}$ through hydrogen abstraction reactions, and undergo addition reactions with O_3 via its unsaturated double bonds. Furthermore, the four oxidants exhibit different $\log_{10}k_i$ value distribution with VOCs due to differences in chemical structures and reactivity (Fig. 2D). $\bullet\text{OH}$, due to its high oxidation potential, usually reacts quickly with VOCs via hydrogen abstraction, with $\log_{10}k_i$ concentrated in the range of -14.000 to -10.000. In contrast, O_3 typically undergoes slower addition reactions with unsaturated bonds in reactants (Ziemann and Atkinson, 2012), with $\log_{10}k_i$ ranging from -20.836 to -13.721. NO_3 can participate in both hydrogen abstraction and addition reactions, resulting in a wider range of $\log_{10}k_i$ values. The diverse reaction rates of these oxidants maintain the composition and oxidative state of aerosols in the atmosphere, but the uneven distribution of their values makes predicting k_i more challenging. Even for the same oxidant, VOCs with different structures exhibit varied reaction rates in gas-phase oxidation reactions. For example, NO_3 reacts very slowly with aromatic rings, with a k_i value of $3.900 \times 10^{-16} \text{ cm}^3/(\text{molecule} \cdot \text{s})$ for xylene. In contrast, NO_3 can rapidly abstract hydrogen from hydroxyl groups, with a k_i value of up to $1.72 \times 10^{-10} \text{ cm}^3/(\text{molecule} \cdot \text{s})$ for 3-methylcatechol.

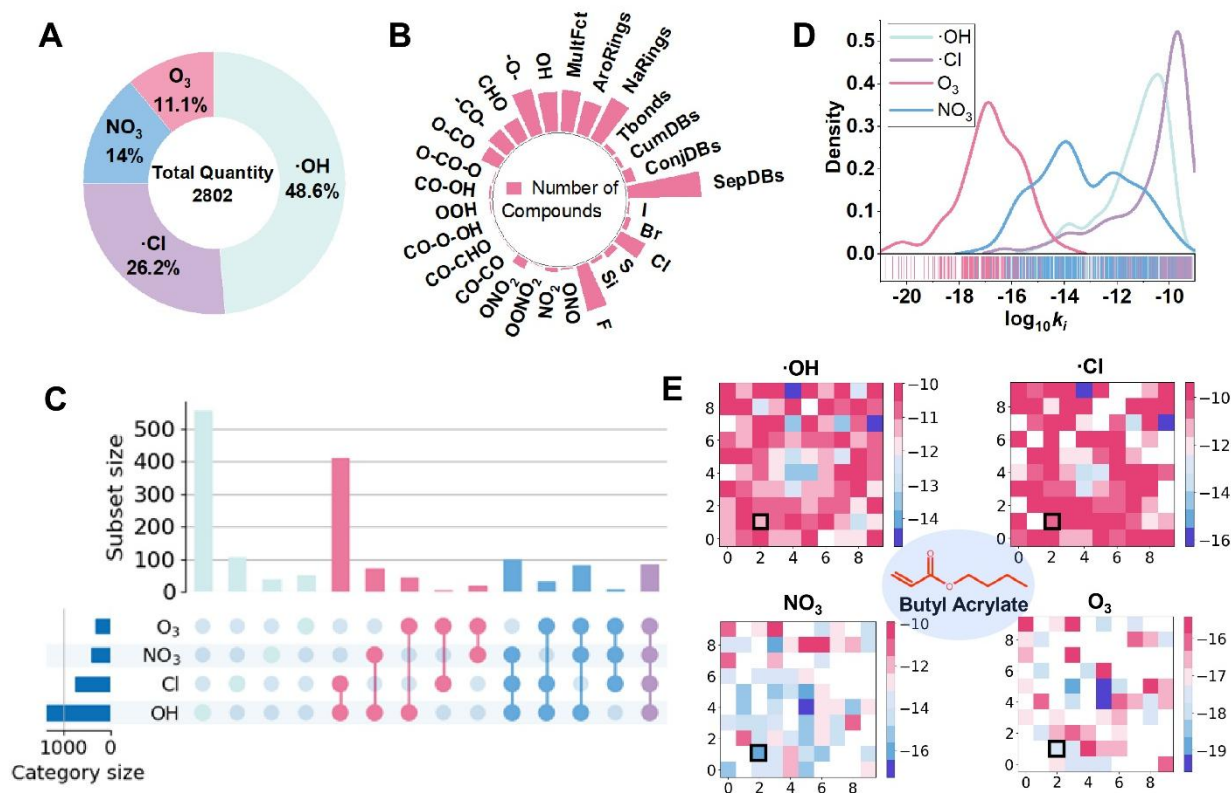


Figure 2. Visualization of VOCs Dataset. (A) Proportion of the four types of oxidants. (B) Number of VOCs containing each functional group. MultFct: multifunctional; AroRings: aromatic rings; NaRings: non-aromatic rings; Tbonds: triple bonds; CumDBs: cumulated double bonds; ConjDBs: conjugated double bonds; SepDBs: separated double bonds. (C) Number of VOCs that can undergo oxidation reactions with the four oxidants. (D) Distribution of $\log_{10}k_i$ values for the four oxidants. (E) Heatmap of reaction rate constants based on VOCs clustering, where each grid represents a cluster of structurally similar VOCs. The color gradient indicates the $\log_{10}k_i$ values, with red indicating higher $\log_{10}k_i$ values (faster reaction rates), blue indicating lower $\log_{10}k_i$ values (slower reaction rates), and white indicating the absence of $\log_{10}k_i$ data for that cluster. The cluster containing butyl acrylate are enclosed within the black box.

Furthermore, the same VOCs show different reaction rates with different oxidants. To explore the relationship between structural differences of VOCs and reaction rates, the study employed the Self-Organizing Map (SOM) algorithm (Kohonen, 2006) to visualize $\log_{10}k_i$ values. Based on the Morgan fingerprint similarity of VOCs, the VOCs were clustered into 100 groups, each containing VOCs with similar molecular structures. Each grid in Fig. 2E represents a cluster of VOCs, and the color gradient indicates the $\log_{10}k_i$ values of their reactions with the corresponding oxidants. By comparing the $\log_{10}k_i$ values of the same VOCs with four oxidants, the relationship between structural features and reaction rates for each oxidant can be evaluated. For example, butyl acrylate (CAS RN.141-32-2) reacts slowly with NO_3 radicals and O_3 , mainly due to the unsaturated addition reactions through the carbon-carbon double bond, where the ester group in the molecular structure produces an electron-withdrawing effect, reducing the electron density in the π bond and thus lowering the reaction rate (Gai et al., 2009; Wang et al., 2010). In contrast, it reacts faster with $\bullet\text{OH}$ and $\bullet\text{Cl}$ through hydrogen abstraction rather than



185 addition (Le Calvé et al., 1997; Ohta, 1984; Wang et al., 2018). This demonstrates that the dataset, which includes various oxidants and VOCs, exhibits diverse $\log_{10}k_i$ values. The overall $\log_{10}k_i$ values differ significantly between different oxidants. This diverse dataset enables the model to learn the reaction information between VOCs and different oxidants, thereby improving model performance and prediction accuracy.

3.2 Performance Evaluation of Vreact Model

190 The Siamese MPNN architecture of the Vreact captures both molecular features of VOCs and oxidants as well as their interaction dynamics simultaneously. During hyperparameter optimization, the set of hyperparameters that minimized MSE on the validation set was selected. After training for 46 epochs (Fig. S1), Vreact achieved robust predictive performance on the validation set, with R^2 of 0.961, MSE of 0.194 and MAE of 0.314 for $\log_{10}k_i$ (Fig. 3A). On the internal test set, the model achieved R^2 of 0.941, MSE of 0.299 and MAE of 0.322 for $\log_{10}k_i$ (Fig. 3A), indicating robust predictive capability and
195 excellent generalization ability for unseen VOC-oxidant combinations. The small MAE difference between the validation set and internal test sets, despite a larger difference in MSE, indicates that MSE is more sensitive to outliers or large errors, while MAE directly reflects the average absolute prediction error. Although the R^2 on the internal test set is slightly lower than on the validation set, this minor discrepancy does not affect the model's robust predictive ability. The result on the internal test set is available in Table S4.

200 To explore the predictive performance of the Vreact model for different types of oxidants, we evaluated the prediction performance for $\bullet\text{OH}$, $\bullet\text{Cl}$, O_3 , and NO_3 separately. The regression fit of predicted $\log_{10}k_i$ values versus experimental values for the four oxidants (Fig. 3B) shows that O_3 and NO_3 have higher dispersion compared to $\bullet\text{OH}$ and $\bullet\text{Cl}$. The R^2 values for the reactions of the four oxidants, in descending order, are $\bullet\text{OH} > \bullet\text{Cl} > \text{NO}_3 > \text{O}_3$, with $\bullet\text{OH}$ and $\bullet\text{Cl}$ having R^2 values of 0.942 and 0.916, respectively. The prediction performance for NO_3 radicals and O_3 is comparatively lower, with R^2 values
205 below 0.800. The $\bullet\text{OH}$ dataset is the most abundant and concentrated, whereas the $\log_{10}k_i$ values for NO_3 are highly dispersed, which may cause the model to have difficulty capturing all patterns and relationships for NO_3 , leading to prediction biases.

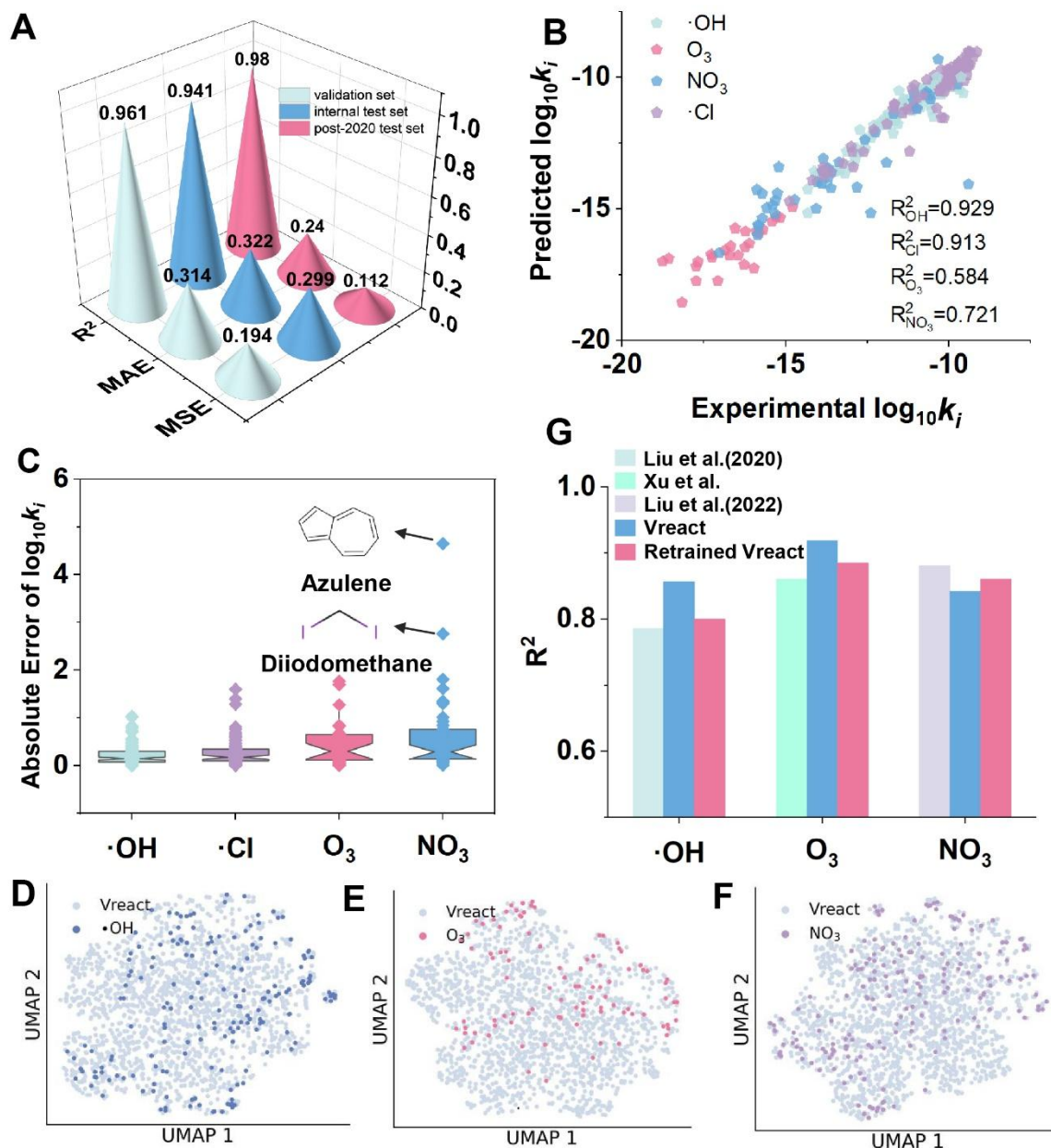


Figure 3. Evaluation and comparison of the predictive performance of the Vreact model. (A) MSE, MAE, R^2 of Vreact (trained on the McGillen et al. dataset) on the validation set, internal test set, and external post-2020 test set. (B) R^2 values for $\log_{10}k_i$ predictions of four oxidants' reactions in the internal test set. (C) Distribution of AE between predicted and experimental $\log_{10}k_i$ values for the four oxidants in the internal test set. (D-F) The chemical spatial distribution of VOCs in the $\cdot\text{OH}$, O_3 , and NO_3 datasets used in this study and prior literatures. (G) R^2 comparison among previously published single-oxidant models, the original Vreact (evaluated on literature test set), and Retrained Vreact (trained and tested using the same splits as the literature models) highlighting adaptability.



The Absolute Error (AE) between the predicted and experimental $\log_{10}k_i$ values for the four types of oxidants are presented in Fig. 3C. The median AE for $\bullet\text{OH}$ is 0.149, while O_3 and NO_3 exhibit median AEs of 0.301 and 0.287, respectively, which are slightly higher than that of $\bullet\text{OH}$. Overall, 84% of the AE values for O_3 and NO_3 are within 1. As depicted in the Fig. 3C, individual outliers in AE contribute to the increased RMSE and MAE for O_3 and NO_3 , and the consequent decrease in R^2 .

For example, the AE for the reaction of NO_3 with azulene (C_{10}H_8) is 4.653. Azulene, an aromatic hydrocarbon composed of a seven-membered ring fused to a five-membered ring, is an isomer of naphthalene (C_{10}H_8). NO_3 , as electrophilic reagents, tend to attack regions with higher electron density. Compared to naphthalene, the electron density distribution of azulene is uneven, with certain regions having high electron density that may facilitate effective interactions with NO_3 . Additionally, the structure of azulene may reduce steric hindrance, allowing NO_3 radicals easier access to reaction sites (Atkinson et al., 1992), resulting in a higher reaction rate constant and increasing the model's prediction difficulty. Similarly, the predicted $\log_{10}k_i$ value for the reaction of NO_3 with diiodomethane (CH_2I_2) is significantly lower than the true value (AE=2.763). This discrepancy may be attributed to the limited representation of iodine-containing VOCs in the dataset, with only iodomethane (CH_3I) and iodoethane ($\text{C}_2\text{H}_5\text{I}$) having k_i values in the training and validation sets. This limited data prevents the model from fully learning the reaction characteristics of iodine-containing compounds, resulting in a larger prediction error for diiodomethane with NO_3 radicals.

3.3 Comparison with Single-Oxidant Prediction Models

Most existing machine learning models for predicting VOC reaction rates constants are tailored for individual oxidants, limiting their applicability to complex atmospheric systems involving multiple oxidants. In contrast, the Siamese MPNN architecture of the Vreact enables simultaneous learning of molecular features and interaction patterns across different VOC–oxidant pairs within a unified framework. To benchmark Vreact against previously published single-oxidant QSAR/ML models, we selected three top-performing models developed under 298K conditions: Liu et al. (2020) for $\bullet\text{OH}$ (180 data points), Xu et al. (2013) for O_3 (95 data points), and Liu et al. (2022) for NO_3 radicals (189 data points). Prior to evaluation, we applied Uniform Manifold Approximation and Projection (UMAP) to reduce the dimensionality of the Morgan molecular fingerprints to visualize the chemical space of both the comparison datasets and the Vreact training set (Figs. 3D-F). The observed structural overlap confirms that Vreact's dataset spans a broad and diverse chemical space. Given that our study used different data than those reported in the literature, we employed two strategies for comparison. First, the pre-trained Vreact model (trained on the McGillen dataset) was directly applied to the test sets from the literature to evaluate extrapolation performance. Second, Vreact was retrained on each literature dataset using their original train/test splits, allowing a direct comparison with published models on literature test sets (Retrained Vreact).

As shown in Fig 3G, both the original Vreact model and its retrained version consistently outperformed the single-oxidant models from Liu et al. (2022) and Xu et al. (2013) on the $\bullet\text{OH}$ and O_3 literature test sets, achieving higher R^2 values and demonstrating superior regression fits between predicted and experimental values. These results highlight the capability of the Vreact architecture—whether trained on a broad multi-oxidant dataset or fine-tuned on smaller single-oxidant datasets—



to effectively learn structural features of VOCs and oxidants and capture complex molecular interactions through its Siamese
250 MPNN framework. For the NO_3 prediction task, the original Vreact model performed poorly, but the retrained Vreact model
showed an improvement in R^2 . This suggests that a unified dataset containing multiple oxidants may introduce additional
noise during training, affecting the model's ability to learn key interaction features between NO_3 and VOCs. Nonetheless, the
model's R^2 on the literature test set still reached 0.842, indicating only a slight loss in predictive accuracy, which is
acceptable.

255 3.4 Mechanism Insights Through Interaction Analysis

The interaction layer of the Vreact model can elucidate the atomic interaction mechanisms between VOCs and oxidants. The
interaction matrix, sized $n_1 \times n_2$, where n_1 represents the number of non-hydrogen atoms in the VOC molecule and n_2
represents the number of non-hydrogen atoms in the oxidant molecule. Mapping these interaction coefficients onto the
molecular structure highlights key atoms that determine the reaction rate.

260 To exemplify this mechanism, we analysed specific cases. 2-methyl-4-penten-2-ol is an unsaturated oxygenated volatile
organic compound (OVOC) that constitutes a significant proportion of the atmospheric VOCs, primarily sourced from
industrial solvents used in ink and jet ink manufacturing (Li et al., 2021). As shown in Fig. 4A, the interaction coefficient for
the distal unsaturated carbon atoms is the highest during the reaction with O_3 , indicating these are likely the reaction sites for
 O_3 attack. It is inferred that O_3 adds to the unsaturated carbon-carbon double bond through an addition reaction, forming
265 primary ozonides (POZs). These POZs are unstable intermediates that rapidly cleave to produce carbonyl compounds and
carbon-based radicals, which further rearrange to form secondary ozonides (SOZs). The SOZs and their reaction products are
precursors of SOA. Another example is γ -caprolactone (GCL), a five-membered ring ester used in perfumes, which rapidly
reacts and degrades with $\bullet\text{OH}$ upon entering the atmosphere. Interaction weight analysis shows that the carbon atom linked
to the ethyl group contributes most to GCL's oxidative degradation by $\bullet\text{OH}$ (Fig. 4B), suggesting that $\bullet\text{OH}$ initially attacks
270 this carbon atom, abstracting a H atom to form a carbon radical. Previous studies indicate that the reactivity of carbons
adjacent to the oxygen atom in lactones is particularly significant in reactions with $\bullet\text{OH}$, especially when alkyl substituents
are attached to this carbon, which enhances its reactivity (Barnes et al., 2014).

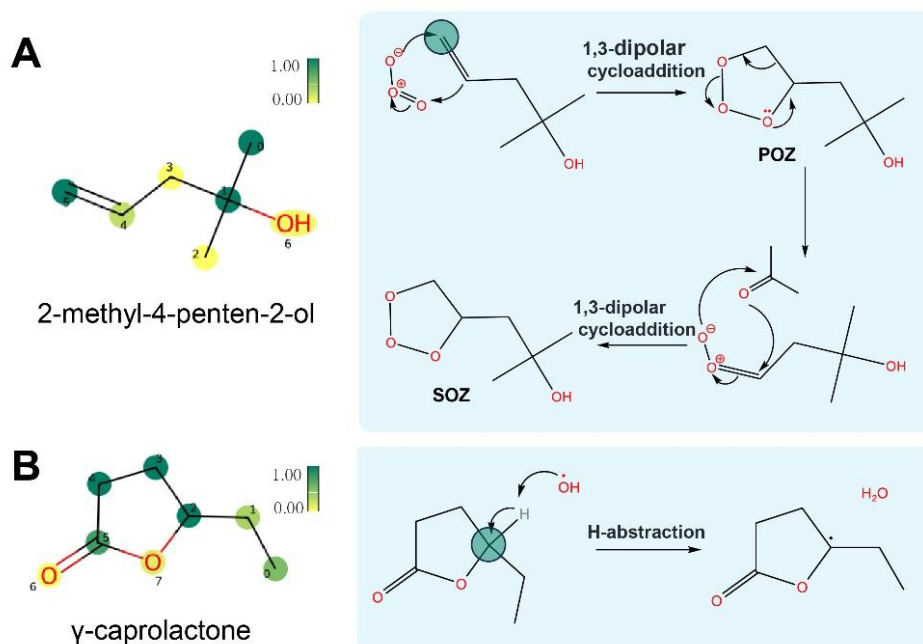


Figure 4. Visualization of atomic weights in VOC molecules. (A) Reaction process of 2-methyl-4-penten-2-ol with O₃. (B) Reaction process of γ -caprolactone with •OH. The darker the highlighted color of the atom, the stronger its interaction in the gas-phase oxidation reaction.

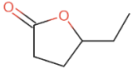
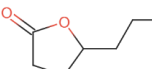
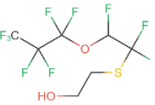
3.5 Evaluating Extrapolation Ability and Prioritizing VOCs for Environmental Impact

To further validate the extrapolation capability and generalization performance of the Vreact model, developed using a dataset compiled up to the year 2020 (Baptista et al., 2021; Joudan et al., 2022; Li et al., 2021), additional k_i data from experimentally measured VOCs and oxidants published after 2020 were collected as an external test set (post-2020 test set) (Table 1). The prediction results showed that the AE between the experimental $\log_{10}k_i$ and the predicted values was within 1, with the reaction rate constant prediction for γ -heptalactone and •OH exhibiting the smallest prediction error. The AE for γ -heptalactone with •OH was only 0.005, and the overall MAE was 0.240, with an MSE of 0.112 and an R² of 0.98 (Fig. 3A shown in red). The results indicate that the Vreact can accurately predict the atmospheric oxidation reaction rate constants of unknown VOCs, demonstrating its potential application in addressing complex atmospheric chemistry issues involving the interactions between VOCs and oxidants.

Table 1. The prediction results on the post-2020 test set.

VOC name	Chemical structure	Oxidant	Experimental $\log_{10}k_i$	Predicted $\log_{10}k_i$	AE	Ref.
2-methyl-4-penten-2-ol		O ₃	-17.370	-16.712	0.658	(Li et al., 2021)



γ -caprolactone		•OH	-11.194	-11.209	0.015	(Baptista et al., 2021)
		•Cl	-9.886	-10.149	0.263	(Baptista et al., 2021)
γ -heptalactone		•OH	-11.056	-11.051	0.005	(Baptista et al., 2021)
		•Cl	-9.770	-9.943	0.173	(Baptista et al., 2021)
FESOH		•OH	-11.377	-11.876	0.499	(Joudan et al., 2022)
		•Cl	-10.824	-10.759	0.065	(Joudan et al., 2022)

FESOH: 2- (1,1,2-trifluoro-2-heptafluoropropoxy-ethylsulfanyl)-ethanol; AE: absolute error

290

Despite the identification of hundreds of VOC species, the environmental behavior of most VOCs in the atmosphere and their potential contributions to particulate matter formation and ozone increase remain largely unclear. To address this gap, we employed the Vreact model to evaluate the atmospheric oxidation reaction rate constants of a broad spectrum of VOCs. Molecular structures for 447 VOCs with unknown atmospheric oxidation k_i values were collected from previous research, which evaluated more than 500 Chinese domestic source profiles, including literature and field measurements (Sha et al., 2021) (Table S5). After excluding VOCs already included in the Vreact dataset, 296, 339, 416, and 369 data points for •OH, •Cl, O₃, and NO₃ were retained, respectively. The prediction results indicated that, although the oxidation reaction rates of VOCs in the atmosphere vary (Fig. 5A), the differences in $\log_{10}k_i$ values are primarily influenced by the type of oxidant, with smaller variations in $\log_{10}k_i$ values observed for different VOCs reacting with the same oxidant. Among these, reactions with •OH and •Cl were the fastest, consistent with the results from the McGillen dataset analysis used in the modeling (Fig. 2D). Additionally, the changes in the proportion of VOC types within different reaction rate intervals (Fig. 5B) demonstrated that the composition of VOC types varied with reaction rates. Halocarbons exhibited relatively slower reaction rates, while alkenes and aromatics reacted relatively quickly, and oxygenated compounds showed a more uniform rate distribution. Consequently, areas with high emissions of alkenes and aromatics will produce more reaction products per unit time, providing precursors for O₃ and SOA formation (Gao et al., 2021).

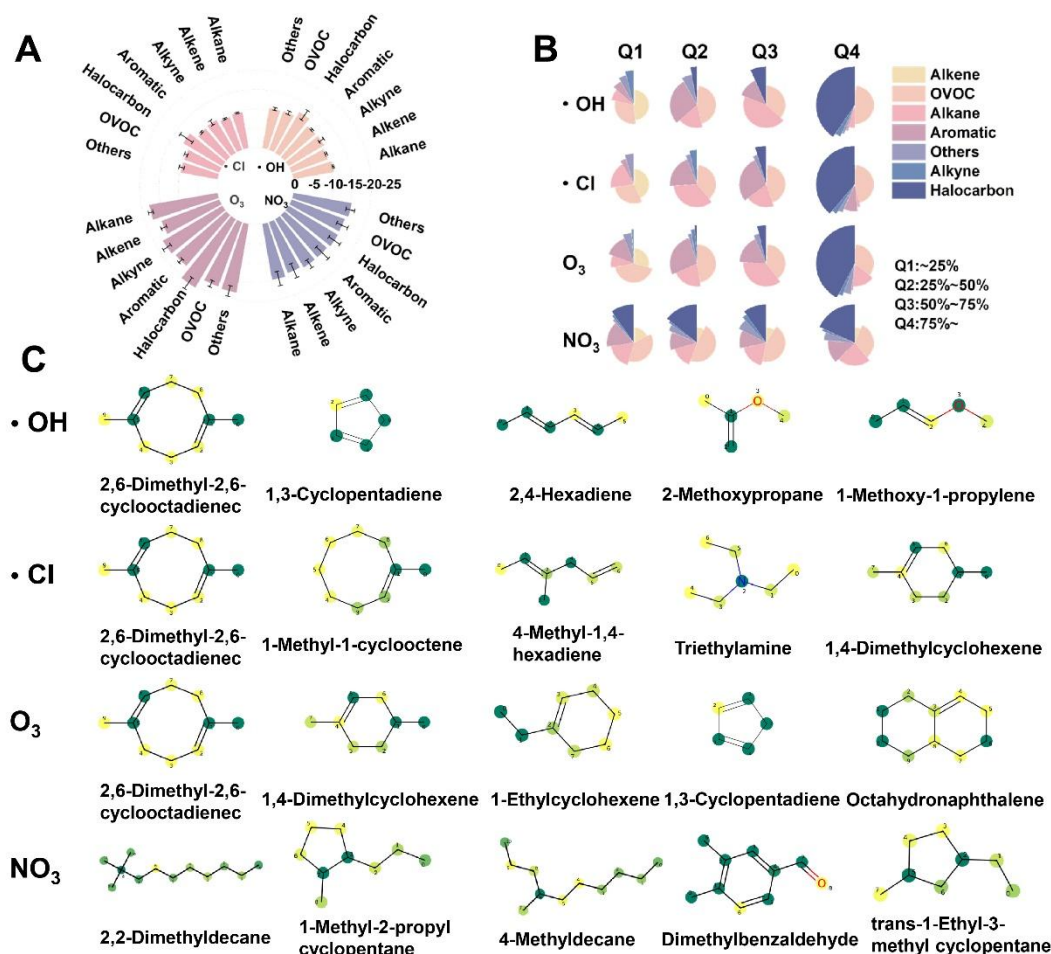
305

The top five VOCs with the fastest reaction rates with •OH, •Cl, O₃, and NO₃ were further examined in the study (Fig. 5C). Among these, 2,6-Dimethyl-2,6-cyclooctadiene (CAS RN: 3760-14-3) is a volatile compound with an irritating odor, exhibiting the fastest reaction rates with •OH, •Cl, and O₃. Additionally, 1,3-cyclopentadiene (CAS RN: 542-92-7) and 1,4-Dimethylcyclohexene (CAS RN: 70688-47-0) also showed high reaction rates with O₃, •Cl, and •OH, likely due to the presence of double bonds and cyclic structures in these molecules. The carbon atoms in the double bonds and those connected to methyl groups generally have high reactivity. Therefore, it could be inferred that these VOCs, or VOCs with

310



similar structures, may significantly contribute to the formation of fine particulate matter and the increase in ozone in the atmosphere.



315 **Figure 5. Predicted reaction rate constants for VOCs atmospheric oxidation reactions. (A) Predicted mean $\log_{10}k_i$ values for different types of VOCs. (B) Distribution of VOC types ranked by predicted reaction rates, divided into quartiles: the fastest 25% (Q1), 25%-50% (Q2), 50%-75% (Q3), and the slowest 25% (Q4). (C) Molecular structures of VOCs with the fastest reaction rates with the four oxidants.**

4 Concluding

320 Understanding the oxidation rates of VOCs is crucial for evaluating their impact on atmospheric chemistry and air pollution. In this study, Vreact, a deep learning-based model, is introduced to predict VOC oxidation reaction rate constants with multiple oxidants simultaneously. The model demonstrates high predictive accuracy while offering mechanistic insights into VOC oxidation by analyzing atomic-level interaction patterns. By incorporating a broad range of VOC structures and oxidant interactions, Vreact enhances its generalizability, allowing for large-scale screening of previously uncharacterized



325 VOC oxidation rates. Additionally, an interactive web-based tool (<http://vreact.envwind.site:8001>) is provided (Fig. S2),
which facilitates VOC oxidation rate predictions for non-experts. This tool significantly improves accessibility, enabling
researchers, policymakers, and environmental agencies to assess VOC reactivity and prioritize mitigation efforts effectively.
Given the increasing complexity of atmospheric pollution and its global impacts on human health and climate, these
advancements in predictive modeling offer valuable resources for addressing worldwide air quality challenges.

330 However, Vreact has several limitations. The model's performance depends on the availability and quality of experimental
kinetic data. The training dataset primarily relies on measured reaction rate constants, which may not cover all VOCs,
especially those with complex structures or containing halogen and sulfur groups. As a result, VOC classes that are
underrepresented in the dataset, such as iodine-containing compounds, may exhibit prediction errors, highlighting the need
for further data collection. Additionally, while Vreact captures essential molecular interactions, biases may arise from the

335 existing reaction rate constants datasets, especially when reaction conditions or mechanisms differ from those used in
training. Extrapolating reaction rate constants for uncharacterized VOCs presents another challenge. While the model shows
strong generalization capabilities, its accuracy may decrease for highly reactive or structurally unique compounds. To
improve predictions, future work could integrate high-throughput quantum chemical calculations and automated
experimental validation. Optimizing inference speed and integrating Vreact with atmospheric chemistry models could

340 enhance its applicability in real-time air quality simulations.

Data and Code Availability

The code and datasets used and/or analyzed during the current study are available at <https://github.com/Luo-Jiaqi/Vreact> and
supplemental information.

Supplementary Material

345 Detailed information about the learning curve of the Vreact training process (Figure S1); User interface of the web platform
for predicting VOC reaction rate constants using the Vreact model (Figure S2); Graph representation of molecular structures
(Text S1); MPNN message passing and readout phases for molecular graphs (Text S2); Regularization and early stopping
techniques in the Vreact model training (Text S3); Model performance evaluation metrics (Text S4); Implementation of the
Vreact model (Text S5); Distribution of VOCs reactions with atmospheric oxidants across datasets (Table S1); Atomic
350 features and bond features used in molecular graph representation (Table S2); Hyperparameter search space and optimal
settings for the Vreact model (Table S3); Experimental and predicted $\log_{10}k_i$ values for VOCs on the internal test dataset
(Table S4); 447 real-world atmospheric VOCs (Table S5).



Author Contributions

Methodology, Investigation, Formal analysis, Data curation, Visualization, Writing–original draft, X.Z. and J.Q.L.;
355 Resources, Conceptualization, Software, Writing–review & editing, Supervision, Funding acquisition, J.J.F and X.L.;
Software, Validation, Writing–review & editing, W.X.P and Q.X.; Software, Funding acquisition, Writing–review & editing,
A.Q.Z; Resources, Supervision, G.B.J.

Competing Interests

The authors declare no competing interests.

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