



# 1 **Intercomparison of online and offline XRF spectrometers for determining the** 2 **PM<sub>10</sub> elemental composition of ambient aerosol**

3 Laura Cadeo<sup>1</sup>, Beatrice Biffi<sup>2</sup>, Benjamin Chazeau<sup>3,4</sup>, Cristina Colombi<sup>2</sup>, Rosario Cosenza<sup>2</sup>, Eleonora  
4 Cuccia<sup>2</sup>, Manousos-Ioannis Manousakas<sup>3</sup>, Kaspar R. Daellenbach<sup>3,\*</sup>, André S. H. Prévôt<sup>3</sup>, Roberta  
5 Vecchi<sup>1,\*\*</sup>

6 <sup>1</sup>Department of Physics, Università degli Studi di Milano, Milan, 20133, Italy

7 <sup>2</sup>ARPA Lombardia, Milan, 20124, Italy

8 <sup>3</sup>Laboratory of Atmospheric Chemistry, Paul Scherrer Institute, 5232 Villigen PSI, Switzerland

9 <sup>4</sup>Aix Marseille Univ, LCE, Marseille, France

10

11 *\*Corresponding authors:* kaspar.daellenbach@psi.ch, roberta.vecchi@unimi.it

12

13 **Keywords:** XRF intercomparison, on-line spectrometry, PM<sub>10</sub>

14

## 15 **Abstract**

16 Measuring the elemental composition of atmospheric particulate matter (PM) can provide useful  
17 information on the adverse effects of PM and help the identification of emission sources. Carrying  
18 out these measurements at a high time resolution (1-h or less) allows to describe the fast processes to  
19 which aerosol particles are subjected in the atmosphere, leading to a better characterisation of the  
20 emissions. X-ray fluorescence spectrometry (XRF) is one of the most widespread techniques used to  
21 determine the elemental composition of PM. In recent years, new systems known as online XRF  
22 spectrometers have been developed to provide real-time measurements of the PM elemental  
23 concentration at a high time resolution. Among these advanced instruments, the Xact® 625i Ambient  
24 Metals Monitor by Cooper Environmental (USA) performs in situ automated measurements with a  
25 user selected time resolution ranging from 15 to 240 min. In this study, an Xact® 625i monitor was  
26 deployed for nearly 6 months (July-December 2023) in Milan (Po Valley, Italy) at a monitoring  
27 station of the Lombardy Regional Agency for Environmental Protection (ARPA Lombardia). The  
28 instrument was configured to quantify 36 elements, ranging from Al to Bi, with 1-h time resolution  
29 in the PM<sub>10</sub> fraction. The objective of the study was to verify the correct functioning of the instrument  
30 and to evaluate the quality and robustness of the data produced. Xact® 625i data were aggregated to  
31 24-h daily means and then compared to 24-h PM<sub>10</sub> filter data retrieved by ARPA Lombardia in the  
32 same station and analyzed offline for the elemental concentration with a benchtop XRF spectrometer.  
33 The intercomparison focused on the 16 elements (Al, Si, S, Cl, K, Ca, Ti, Cr, Mn, Fe, Ni, Cu, Zn, Br,  
34 Sr, and Pb) whose concentrations were consistently above their minimum detection limits (MDL) for  
35 both online and offline techniques. Results of the intercomparison were satisfying showing that the  
36 Xact® 625i elemental concentrations were found to be highly correlated to the offline XRF analyses



37 ( $R^2$  ranging from 0.67 to 0.99) and slopes ranging from 0.79 to 1.3 (just a couple of elements showed  
38 slopes up to 1.70).

## 39 **1. Introduction**

40 Measurement and quantification of the chemical composition of atmospheric particulate matter (PM)  
41 are key aspects of air quality monitoring. It has long been known that PM is associated with adverse  
42 impacts, which are influenced by the chemical composition of the particles. At the global scale, PM  
43 affects cloud formation and Earth's radiative budget (Fuzzi et al., 2015); at the local scale, its  
44 harmfulness on human health is of particular concern (Brunekreef and Holgate, 2002; Kelly et al.,  
45 2012; Rohr and Wyzga, 2012; Daellenbach et al., 2020). Therefore, it is important to achieve a more  
46 detailed knowledge about which chemical components are responsible for these negative effects.  
47 Determining the composition of PM is also a fundamental step to perform source apportionment  
48 studies for the identification of the emission sources, which help the implementation of mitigation  
49 strategies (WHO, 2013).

50 Trace elements, in particular metals, although they generally do not contribute substantially to the  
51 mass of PM are of interest because they act as tracers for specific sources (Visser et al., 2015) and  
52 some of them are associated to adverse health effects even at ambient level concentrations (Chen and  
53 Lippmann, 2009). The quantification of these elements in PM samples can be obtained through  
54 various techniques, such as X-ray fluorescence spectrometry (XRF), particle-induced X-ray emission  
55 (PIXE), and inductively coupled plasma mass spectrometry (ICP-MS). These methods require the  
56 collection of aerosol particles on filters, followed by laboratory analysis. The XRF method is a  
57 widespread non-destructive technique, largely and successfully applied to aerosol speciation in the  
58 last decades (e.g. Marazzan et al., 2001). Samples do not require a pre-treatment procedure, in  
59 contrast to ICP-MS for which sample preparation is very laborious, time consuming and leads to the  
60 destruction of filters (Olesik, 1991). XRF spectrometers are compact devices typically composed of  
61 an X-ray source and a detector; recent technical advances in this field have resulted in the  
62 development of small and transportable commercial systems. In contrast, PIXE analysis requires  
63 access to accelerator facilities, which is limited due to a very high demand for availability; however,  
64 there are very effective set-up where a high number of samples can be robustly analyzed (e.g., at the  
65 INFN-LABEC facility in Florence, Italy, the typical irradiation time for each daily sample is 45-60 s  
66 depending on the mass loading but also 1h resolution samples can be analyzed in 1 min per spot; see  
67 e.g. Calzolari et al., 2010, 2015; Lucarelli et al., 2011)

68 PM samples are usually collected by air quality monitoring networks with a time resolution ranging  
69 from 24-h to one week, to ensure that enough PM mass is available for the analytical analysis, which



70 is carried out in a laboratory. The elemental composition of PM is then obtained with a considerable  
71 time delay and at low temporal resolution. In recent years, there has been a growing interest in  
72 developing instruments for high temporal resolution measurements. Sampling at a high time  
73 resolution (1-h or less) allows to capture fast processes which aerosol particles are subjected to in the  
74 atmosphere and to retrieve information about the typical hours of activity of a certain source, leading  
75 to a better characterization of PM emissions. However, due to the short integration times, high-time  
76 resolution measurements are often close to the minimum detection limits (MDL) of the analytical  
77 techniques (Malaguti et al., 2015).

78 Regarding the XRF method, new systems have been developed which are able to sample PM particles  
79 with a sub-hourly or hourly time resolution and to automatically measure their elemental  
80 concentration, providing near-real time data access. These instruments are known as online XRF  
81 spectrometers and can be employed for long monitoring periods (months, years) at a site with the  
82 advantage of requiring limited maintenance. However, their high cost may prevent the simultaneous  
83 use of multiple devices at different sites or the investigation of different size classes (Furger et al.,  
84 2017). One of these advanced instruments is the Xact® 625i Ambient Metals Monitor by Cooper  
85 Environmental (USA), which performs in situ automated measurements of the elemental  
86 concentration of PM with a user selected time resolution ranging from 15 to 240 min. During  
87 operation, remote access to the data is available, enabling continuous, near-real-time monitoring of  
88 the instrument and ambient metal concentrations. The instrument and its forerunner versions have  
89 been successfully employed in several field studies in the past years, which compared its online  
90 measurements to daily samples analyzed with more established laboratory techniques (Bhowmik et  
91 al., 2022; Tremper et al., 2018; Furger et al., 2017; Park et al., 2014; US-EPA, 2012). Among these  
92 studies, only in Park et al., (2014) the daily filters were analyzed with the ED-XRF (Energy-  
93 Dispersive X-Ray Fluorescence) technique; in all the other cases, the elemental concentration of daily  
94 samples was retrieved with ICP-MS and ICP-OES (inductively coupled plasma optical emission  
95 spectrometry). In the latest cases, the comparison was then influenced by the different choice of the  
96 analytical technique. Moreover, in most of these studies, the experimental campaigns were carried  
97 out only for a few weeks/months, leading to a very limited number of points available for the  
98 intercomparison. An evaluation of the performances of Xact® 625i (compared with the ICP-MS  
99 technique) during different seasons was realized only by Bhowmik et al., (2022) who conducted the  
100 field campaigns during summer (June-July) and winter (October-December) 2019 at two sites in  
101 Delhi.

102 In this study, an Xact® 625i monitor was deployed for nearly 6 months (July-December 2023) in  
103 Milan (Po Valley, Italy) at a monitoring station of the Lombardy Regional Agency for Environmental



104 Protection (ARPA Lombardia), where air quality measurements are performed continuously. Xact®  
105 625i hourly samples measured online with ED-XRF were compared to daily filters measured offline  
106 by ARPA Lombardia with a benchtop ED-XRF spectrometer in their laboratory. The goals of this  
107 paper are (1) to assess the on-line instrument performance in typical summer and winter elemental  
108 concentration ranges for PM<sub>10</sub> collected at an urban background site in the Po valley (Italy); (2) to  
109 evaluate the quality of the obtained data for the selected elements in relation to their MDLs; (3) to  
110 quantify the data robustness based on intercomparison between Xact® 625i and elemental  
111 concentrations provided by a benchtop XRF spectrometer.

## 112 **2. Materials and methods**

### 113 **2.1 Site characteristics**

114 The field campaign was performed at the permanent station Milano Pascal of the ARPA Lombardia  
115 Air Quality Network from 6 July until 12 December 2023. This is an urban background site located  
116 in the eastern side of Milan, in the University campus area called “Città Studi” (45.478°N, 9.236°E;  
117 122 m a.s.l); the station is placed in a public park about 130 m from road traffic. The metropolitan  
118 city of Milan is the second most densely populated area in Italy (ca. 2300 inhabitants km<sup>-2</sup>, almost  
119 doubled by daily commuters) and is located in the Po Valley, a well-known European pollution  
120 hotspot. The site is characterized by wintertime episodes of high pollutant concentrations, due to  
121 emissions from a variety of sources (e.g., residential heating, traffic, and industries) and prolonged  
122 atmospheric stability conditions related to the presence of the mountain chains of the Alps and the  
123 Apennines (Vecchi et al., 2007, 2009). Moreover, in Milan more than 80% of the days in a year are  
124 characterized by wind speed lower than 2 m s<sup>-1</sup> (Vecchi et al., 2019). The site is well documented  
125 with respect to gas-phase pollutants (e.g. NO<sub>x</sub>, SO<sub>2</sub>, O<sub>3</sub>), PM<sub>10</sub> and PM<sub>2.5</sub> chemical characterization,  
126 and source apportionment (e.g., Amato et al., 2016; Altuwayjiri et al., 2021).

### 127 **2.2 Xact® 625i**

128 The Xact® 625i Ambient Metals Monitor (Cooper Environmental Services (CES), Beaverton, OR,  
129 USA) is an online XRF spectrometer, designed for continuous measurements of the elemental  
130 composition of ambient aerosol. The device operates using a reel-to-reel filter tape sampling  
131 technique, followed by the analysis of metals in the resulting PM spot through energy-dispersive X-  
132 ray fluorescence (ED-XRF). Ambient air is drawn inside the instrument through a PM size-selective  
133 inlet which was PM<sub>10</sub> in this study, with a flow rate of 16.7 lpm and the PM is collected onto a Teflon  
134 filter tape. After each sampling interval is completed, the filter tape is automatically advanced to the  
135 XRF system, where the resulting PM deposit is irradiated with an X-ray tube (rhodium anode, max



136 voltage: 50 kV, current: 1 mA) with three excitation conditions (see Table S1 in the Supplementary  
137 Material) and the fluorescence X-rays are measured by a silicon drift detector (SDD). In the  
138 meantime, the next sample is collected on a clean spot of the filter tape and the process is repeated  
139 during each sampling interval, which was set at 60 min for this study. The XRF spectra thus produced  
140 are automatically analyzed by a proprietary software for spectral analysis and elemental  
141 quantification which is installed on the built-in computer. The software, through a linear least-squares  
142 deconvolution algorithm, fits each measured spectrum with a library of pure element reference spectra  
143 to obtain the concentration data for each calibrated element in  $\text{ng m}^{-3}$ . Data can be then downloaded  
144 and monitored remotely with an internet connection. Sampling and XRF analysis are performed  
145 continuously and simultaneously, except for the time required for tape advancement ( $\sim 20$  s). Quality  
146 assurance (QA) checks are performed every day at midnight for 30 min and consist of an energy  
147 calibration (using a rod coated with Cr and Nb) and an upscale measurement to monitor the stability  
148 of the instrument response (for Cr, Pb, and Cd). Therefore, the sample following midnight is collected  
149 with a sampling interval limited to 30 min (00:30-01:00 LT).

150 The instrument was located inside a temperature-controlled cabinet outside the ARPA Lombardia  
151 monitoring station. If any errors are detected during operation, the system halts sampling, ramps the  
152 X-rays down for safety, and displays the cause of the error. The instrument was configured to quantify  
153 36 elements: Al, Si, P, S, Cl, K, Ca, Ti, V, Cr, Mn, Fe, Co, Ni, Cu, Zn, Ga, Ge, As, Se, Br, Rb, Sr, Y,  
154 Zr, Cd, In, Sn, Sb, I, Ba, Hg, Tl, Pb, and Bi; in addition, Nb is also detected for daily QA checks.  
155 Before the beginning of the experimental campaign, each of these elements was calibrated with a  
156 reference standard sample. For each element,  $1\sigma$  interference-free MDLs ( $\text{MDL}_{1\sigma}$ ) for 1-h of  
157 sampling are reported in Table S2, provided for Xact® 625i following the approach reported in Currie  
158 (1977). In XRF analyses, MDLs are inversely proportional to the square root of the irradiation time,  
159 which in the case of Xact® 625i corresponds to the sampling interval. Therefore, lower MDLs are  
160 reached for longer sampling durations.

### 161 **2.3 Daily PM<sub>10</sub> filter samples**

162 Daily PM<sub>10</sub> samples were collected on mixed cellulose ester membrane filters (47 mm diameter) with  
163 a SWAM Dual Channel Monitor (FAI Instruments, Rome, Italy) equipped with PM<sub>10</sub> and PM<sub>2.5</sub> inlets.  
164 The elemental composition of PM samples was determined offline by ED-XRF spectrometry in the  
165 laboratories of the Environmental Monitoring Sector of ARPA Lombardia. An Epsilon 4 spectrometer  
166 from Malvern Panalytical (Monza, Italy) was used for the ED-XRF analysis. Four different irradiation  
167 conditions, which are reported in Table S3, were chosen to optimize the measurement of 19 elements,  
168 i.e., Al, Si, P, S, Cl, K, Ca, Ti, V, Cr, Mn, Fe, Ni, Cu, Zn, Br, Rb, Sr, and Pb. For these measurements,



MDLs based on 24-h sampling time were evaluated as three times the square root of the counts in the background below the peak of the element divided by the corresponding sensitivity in the blank filter ( $MDL_{3\sigma}$ ) (Jenkins, 1981; Lindgren, 2006), are reported in Table S4.

## 2.4 Data coverage

The Xact® 625i measurements started on 6 July 2023 16:00 LT (local time) and ended on 12 December 2023 22:00 LT. The sampling interval of the instrument was set to 1-h. During the summer, in July and August, Xact® 625i suffered from high temperatures during the heatwaves, causing the X-ray tube to reach temperatures above 45° C. This led to automatic shutdowns of the measurements and to subsequent manual restarts, mostly performed remotely. The issue was mainly observed in the central hours of the day, from 13:00 to 16:00 LT. Nevertheless, it was still possible to attain a data coverage above 80% for Xact® 625i data in the central hours of the day during summertime. As a precautionary measure to avoid heat damage to the X-ray tube, Xact® 625i was switched off from 12 to 24 August. During those days, a power failure in the ARPA Lombardia cabin caused also the interruption of daily measurements. Another power failure occurred from 22 October to 8 November, leading to a long pause of hourly measurements. The X-ray tube of Xact® 625i started malfunctioning on 6 December. The issue could not be resolved and the X-ray tube had to be replaced, resulting in a premature end of the experimental campaign.

Overall, the Xact® 625i dataset consists of 2693 valid 1-h samples out of 3822 possible samples, attaining a coverage of 70%. For the daily filters, the dataset consists of 149 samples out of 157 possible samples, reaching a coverage of 95%. A timeline of the periods in which data are missing is reported for both hourly and daily measurements in Figure S1. A summary of the periods of interruption of the measurements lasting more than 12 hours is reported in Table S5. The number N of overlapping days with validated data is reported in Table 1 for each element considered for the intercomparison.

Xact® 625i data, which were originally reported in LT, were synchronized to daily samples and expressed in UTC+1 time zone.

## 2.5 Treatment of data below MDLs

Following the approach of Furger et al., (2017) and Tremper et al., (2018), for the intercomparison study here presented the  $MDL_{3\sigma}$  was considered also for Xact® 625i; indeed, the  $MDL_{3\sigma}$  assures a high statistical confidence (99.7%) and a better comparability with previous literature works. Hereafter,  $MDL_{3\sigma}$  will be referred simply to as MDL.

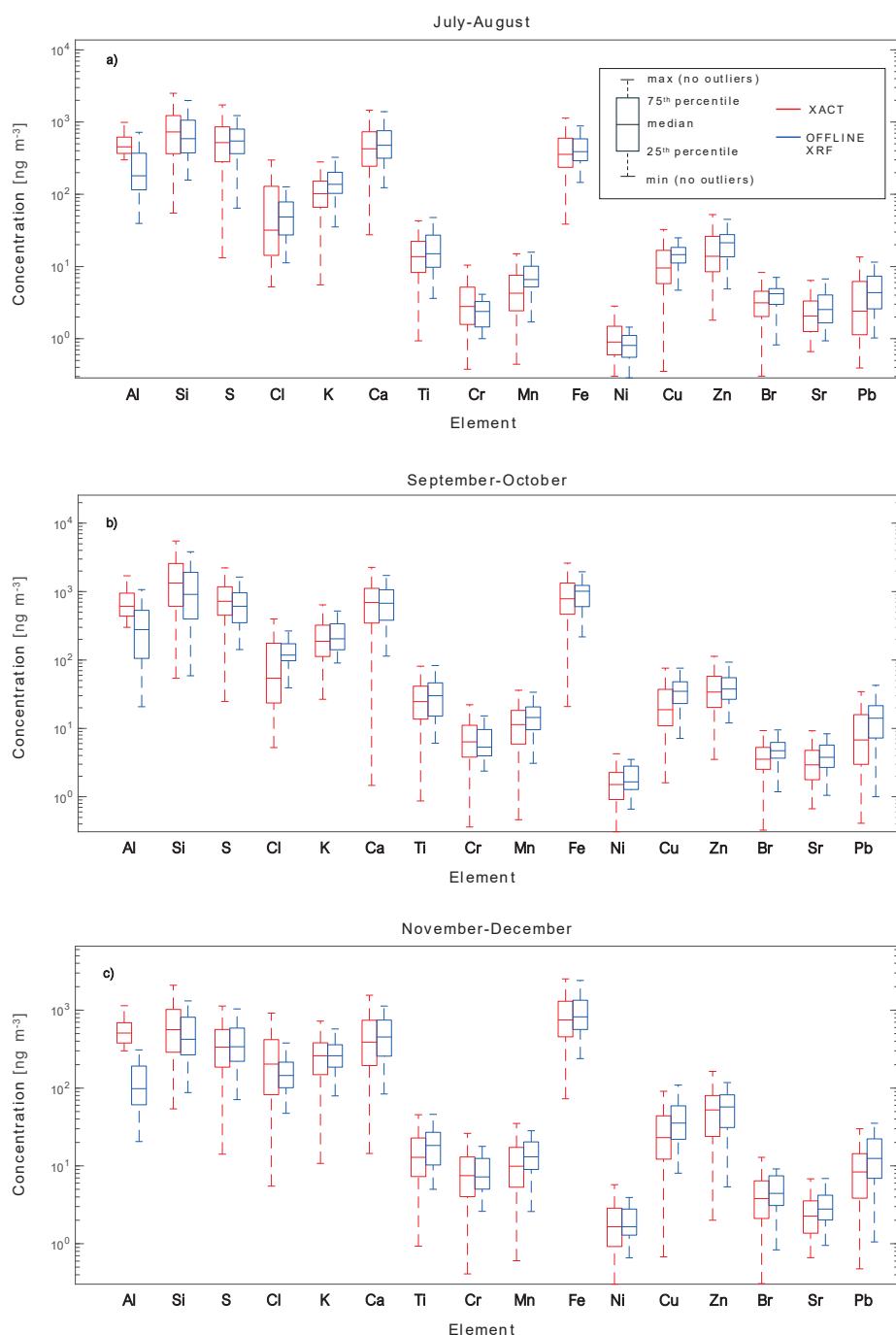


200 All the elements measured on the daily filters by offline XRF have less than 35% of their data points  
201 below their MDL. Among the elements detected by Xact® 625i, 13 of them (P, Co, Ga, Ge, Y, Cd,  
202 In, Sn, Sb, I, Hg, Tl, and Bi) have more than 90% of their data points below MDL; therefore, these  
203 elements were excluded from the intercomparison analysis. V and Rb have >70% of their data points  
204 below MDL leading to a less robust intercomparison with offline XRF (see Figure S2). In Table S6,  
205 the number of data points with concentrations above the MDL is reported for each element measured  
206 by Xact® 625i and by offline XRF.  
207 The intercomparison between Xact® 625i and daily PM<sub>10</sub> elemental concentrations was finally  
208 performed on 16 elements (Al, Si, S, Cl, K, Ca, Ti, Cr, Mn, Fe, Ni, Cu, Zn, Br, Sr, and Pb) which  
209 were measured by both techniques and were consistently above their MDLs.

### 210 **3. Results and discussion**

#### 211 **3.1 Data overview**

212 An overview of the data recorded during the experimental campaign is given in Fig. 1, focusing on  
213 the elements considered for the intercomparison. To account for seasonal differences in terms of  
214 meteorology and emissions, data were divided into 3 periods: July-August, September-October and  
215 November-December. The basic statistics of the dataset, including the mean, median, standard  
216 deviations, 25<sup>th</sup> and 75<sup>th</sup> percentiles are reported in Table S7.



217  
 218 **Figure 1:** Box plots of the concentrations for the elements considered for the intercomparison,  
 219 measured hourly online (in red) and daily offline (in blue) during the experimental campaign in (a)  
 220 July-August, (b) September-October, and (c) November-December. The bottom and the top of each



221 box are the 25<sup>th</sup> and 75<sup>th</sup> percentiles, respectively; the line in the middle of the box is the median; the  
222 bottom and top whiskers are the minimum and maximum value respectively.

223

### 224 **3.2 Intercomparison data analysis approach**

225 For the intercomparison between the two methods, Xact® 625i hourly data were averaged to 24-h to  
226 be comparable to the corresponding daily filter samples measured by offline XRF. Every day, during  
227 QA checks performed from 00:00 to 00:30 LT, Xact® 625i generates one sample with a 30-min time  
228 resolution so that this sample is added to the 23 hourly samples to calculate 24-h daily means. This  
229 procedure implicitly assumes that the half-hour sample collected during the first hour of sampling is  
230 representative of the entire hour. The hypothesis was tested conducting 23.5 h weighted means on a  
231 small number of samples, following the method of Furger et al., (2017). Tests showed a difference of  
232 less than 3% between the 23.5 h weighted mean and the 24-h mean, which was then chosen as  
233 calculation method. For this reason, Xact® 625i data were aggregated to 24-h daily means.

234 As previously stated, during the campaign summer days were affected by heat waves, which caused  
235 Xact® 625i to stop during the central hours of the day, leading to missing data. For this reason, the  
236 data coverage of Xact® 625i was evaluated for each day of the experimental campaign. In order to  
237 avoid misestimation of daily Xact® 625i concentrations, days with less than 18 hourly valid data  
238 (75% coverage) were excluded from the intercomparison. In addition, Xact® 625i daily means were  
239 not calculated when more than 6 hourly data were under the MDL for one day. In all comparisons,  
240 data under MDL were replaced by 0.5·MDL.

241 The comparisons between the daily PM<sub>10</sub> elemental concentrations retrieved by ARPA Lombardia  
242 through offline XRF and the daily means calculated from Xact® 625i hourly data were carried out  
243 using the Deming regression (Deming, 1943). This regression approach minimizes the sum of  
244 distances between the regression line and the X and Y variables, considering the experimental  
245 uncertainties in both variables. For the offline XRF measurements, the uncertainty included  
246 contributions of 5% from calibration standard uncertainty (U.S. EPA, 1999) and, for each spectrum,  
247 the contribution of counting statistics and fitting errors. For the Xact® 625i measurements, the  
248 uncertainty included contributions of 5% from calibration standard uncertainty (U.S. EPA, 1999),  
249 and an element-specific uncertainty derived from the spectral deconvolution calculated by the  
250 instrument software for each spectrum, which includes also the contribution of the flow and the  
251 sample deposit area. The mean relative uncertainties registered during the experimental campaign are  
252 reported for each element and for both online and offline methods in Table S8.



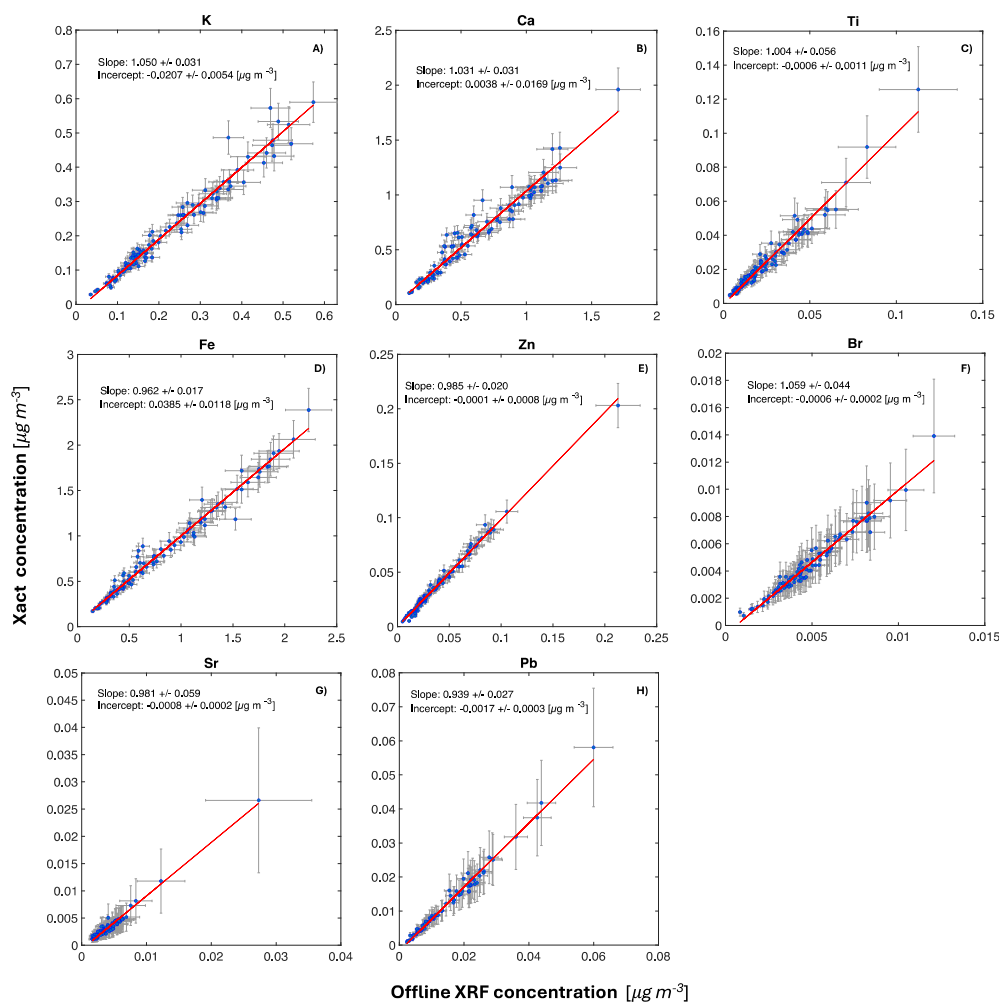
### 3.3 Intercomparison results

The results of the intercomparison between the PM<sub>10</sub> elemental concentrations retrieved offline and online are reported in Table 1. The Deming regression parameters are reported along with their uncertainties and the coefficient of determination of the linear regression; the number of data (N) considered for the comparison after data reduction is also reported.

| Element | Slope ± uncertainty<br>(online vs offline) | Intercept ± uncertainty [μg m <sup>-3</sup> ]<br>(online vs offline) | N   | R <sup>2</sup> |
|---------|--------------------------------------------|----------------------------------------------------------------------|-----|----------------|
| Al      | 1.29 ± 0.22                                | 0.1640 ± 0.0917                                                      | 42  | 0.83           |
| Si      | 1.69 ± 0.13                                | -0.2528 ± 0.1003                                                     | 97  | 0.94           |
| S       | 1.25 ± 0.02                                | -0.0469 ± 0.0070                                                     | 101 | 0.99           |
| Cl      | 1.69 ± 0.25                                | -0.0493 ± 0.0290                                                     | 75  | 0.67           |
| K       | 1.05 ± 0.03                                | -0.0207 ± 0.0054                                                     | 102 | 0.97           |
| Ca      | 1.03 ± 0.03                                | 0.0038 ± 0.0169                                                      | 102 | 0.95           |
| Ti      | 1.00 ± 0.06                                | -0.0006 ± 0.0011                                                     | 100 | 0.96           |
| Cr      | 1.30 ± 0.08                                | 0.0014 ± 0.0004                                                      | 77  | 0.86           |
| Mn      | 0.83 ± 0.02                                | -0.0002 ± 0.0003                                                     | 101 | 0.95           |
| Fe      | 0.96 ± 0.02                                | 0.0385 ± 0.0118                                                      | 102 | 0.98           |
| Ni      | 0.79 ± 0.06                                | 0.0004 ± 0.0001                                                      | 79  | 0.87           |
| Cu      | 0.85 ± 0.01                                | 0.0006 ± 0.0002                                                      | 102 | 0.99           |
| Zn      | 0.98 ± 0.02                                | -0.0001 ± 0.0008                                                     | 102 | 0.99           |
| Br      | 1.06 ± 0.04                                | -0.0006 ± 0.0002                                                     | 96  | 0.96           |
| Sr      | 0.98 ± 0.06                                | -0.0008 ± 0.0002                                                     | 74  | 0.97           |
| Pb      | 0.94 ± 0.03                                | -0.0017 ± 0.0003                                                     | 83  | 0.99           |

**Table 1:** Deming regression results and coefficient of determination for the comparison between Xact® 625i (Y) and offline XRF data (X). For each element, the number of points (N) available for the intercomparison is reported.

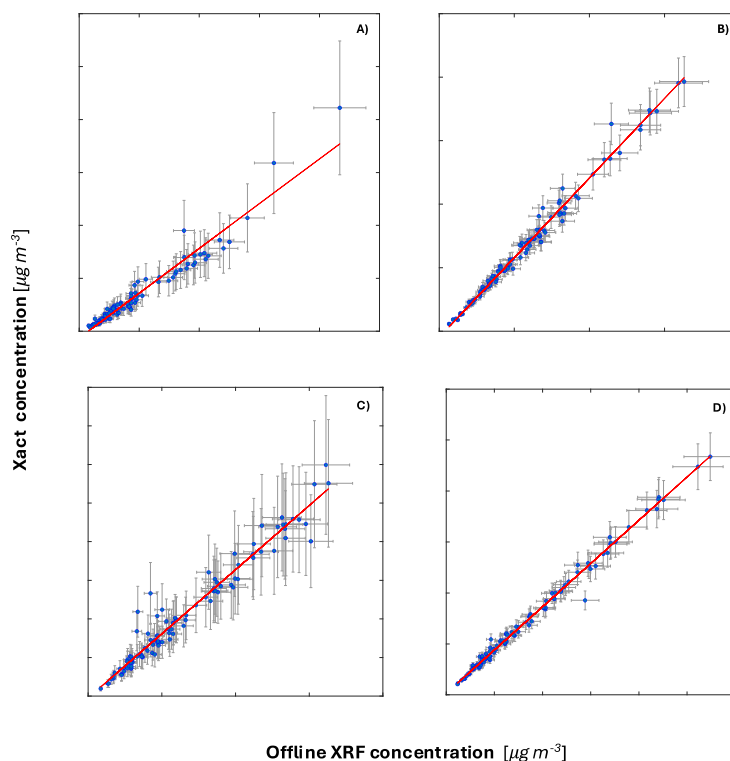
The scatterplots of the intercomparisons are presented in Figures 2-4. The time plots of the time series obtained by the two measurements methods are reported in Figures S3-S6. The 16 selected elements are compared by dividing them into three groups based on data characteristics.



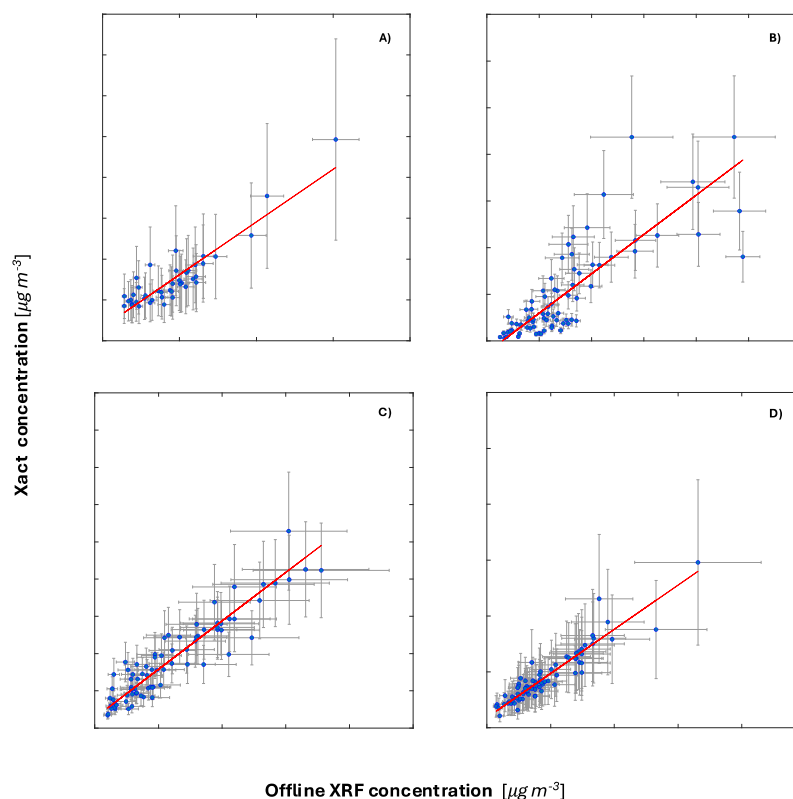
266

267 **Figure 2:** Scatterplots of the intercomparison between Xact® 625i data and offline XRF data for the  
 268 elements of Group A: K, Ca, Ti, Fe, Zn, Br, Sr, and Pb. The error bars represent the mean  
 269 experimental uncertainties reported in Table S8.

270



**Figure 3:** Scatterplots of the intercomparison between Xact® 625i data and offline XRF data for the elements of Group B: Si, S, Mn, and Cu. The error bars represent the mean experimental uncertainties reported in Table S8.



277

278 **Figure 4:** Scatterplots of the intercomparison between Xact® 625i data and offline XRF data for the  
279 elements of Group C: Al, Cl, Cr, and Ni. The error bars represent the mean experimental uncertainties  
280 reported in Table S8.

281

282 The first group, Group A (Figure 2), includes K, Ca, Ti, Fe, Zn, Br, Sr, and Pb. This group shows  
283 excellent correlation between the two measurements methods ( $R^2 > 0.95$ ) and is characterized by  
284 slopes compatible to unity within three times the uncertainty of the fitted slope ( $3\sigma$ ). For Ca, Ti, and  
285 Zn also the intercepts are compatible to 0 within  $3\sigma$ . Among this group, K, Ca, Ti, Fe, and Zn, are  
286 measured by Xact® 625i with relative uncertainties in the range 10-20% (see Table S8). Br, Sr, and  
287 Pb are instead measured by Xact® 625i with a higher uncertainty, on average 30-50% (see Table S8),  
288 and Sr and Pb hourly data are also more frequently under the MDL (20% of data).

289 The second group, Group B (Figure 3), consists of the elements Si, S, Mn, and Cu. This group is  
290 characterized by excellent correlation between the two measurements methods ( $R^2 > 0.95$ ) but, in  
291 contrast to Group A, the slopes of the regressions are not compatible to 1 within  $3\sigma$ . Si and S are  
292 among the lightest elements measured by Xact® 625i and, along with Al, it can be tricky to measure  
293 with ED-XRF because of absorption effects due to the presence of air in the irradiation chamber (e.g.



294 as typically occur in the XRF online measurements) and/or self-absorption inside the coarse particles  
295 themselves (Hunter, and Rhodes, 1972; Van Grieken and Markowicz, 1993); these effects can lead  
296 to an underestimation of low-Z element concentrations. Nevertheless, looking at the results for Al,  
297 Si, and S, absorption effects seem not to be the cause of the observed discrepancy as Xact® 625i data  
298 are typically higher than offline XRF analysis. Moreover, it should be noted that Si is detected by  
299 Xact® 625i with mean uncertainties of 30%, while S is detected with mean uncertainties of 10%. In  
300 the case of Mn and Cu, concentrations provided by Xact® 625i are constantly lower than daily offline  
301 measurements by approximately 15%.

302 A possible explanation for the observed discrepancies is related to the fact that, despite all samples  
303 are measured through ED-XRF technique, the spectra analysis for quantitative analysis is different  
304 and – more importantly - the two instruments are not calibrated with the same set of certified  
305 standards, which can lead to different quantification of concentrations. However, Xact® 625i data of  
306 the elements of this group can still be validated when compared to an offline measurement technique  
307 and used for high-time resolution elemental concentrations assessment, after harmonisation of the  
308 datasets.

309 The third group of elements, Group C (Figure 4), is composed of Al, Cl, Cr, and Ni. This group shows  
310 less comparability between the two methods, with  $R^2$  in the range 0.67-0.87. Cl, Cr, and Ni are  
311 frequently close or under the MDL for both experimental techniques and are characterized by mean  
312 relative uncertainties in the range of 30-50%. For these elements, the comparison could be improved  
313 by carrying out the Xact® 625i measurements on a 2 h time scale. Among the 16 elements evaluated  
314 for the intercomparison, Al is the one with the highest MDL for Xact® 625i hourly measurements  
315 and its hourly concentrations are under the MDL for nearly 35% of data points, while Al offline data  
316 are always above the MDL. Al is also measured by Xact® 625i with mean uncertainties of 50%. As  
317 can be seen also in Figure S6a, the Xact® 625i time series of Al is characterized by a constant upward  
318 shift in background concentrations, which is not observed for the other elements. The measurement  
319 of Al with Xact® 625i is complicated by the fact that the instrument uses an Al filter to carry out the  
320 analysis, as reported in Table S1. Another possible issue could be that the X-rays hit some internal  
321 parts of the instrument, causing a significant increase of the background. In the case of Cl, which  
322 shows quite scattered data, concentrations obtained by Xact® 625i are on average higher than the  
323 ones measured offline on daily samples. This could be explained by the volatility of Cl. Xact® 625i  
324 XRF measurements are performed immediately after the collection of the sample, while daily PM<sub>10</sub>  
325 filters are stored in the sampler at the monitoring station for up to 2 weeks before being taken to the  
326 laboratory for the offline XRF analysis.



327 The results of this study represent a significant step forward from Park et al., (2014), which is – as  
328 far as we know - the only previous study available in literature presenting a comparison between Xact  
329 hourly data and offline ED-XRF daily data. Park et al., (2014) realized an experimental campaign  
330 with a forerunner version of Xact (Xact 620) in Gwangju, South Korea. The campaign was carried  
331 out during February 2011 and lasted only 1 month, focusing on the PM<sub>2.5</sub> fraction. The daily filters  
332 were measured offline with an Epsilon 5 ED-XRF spectrometer (Malvern Panalytical). The study  
333 compared the online and offline concentrations of 12 elements (K, Ca, Ti, V, Mn, Fe, Ni, Cu, Zn, As,  
334 Ba, Pb), 9 of which were also analyzed in our study. For the 9 common elements (K, Ca, Ti, Mn, Fe,  
335 Ni, Cu, Zn, Pb), they observed a mean  $R^2$  of 0.89 and a slope of 1.31, with Xact measurements on  
336 average 30% higher than offline XRF. In our study, for these 9 elements, we found a much better  
337 correlation, with a mean  $R^2$  of 0.96 and slope of 0.94, which is closer to unity. Moreover, our study  
338 included also 7 elements (Al, Si, S, Cl, Cr, Br, Sr) which were not taken into account by Park et al.,  
339 (2014), and the measurement campaign lasted for a longer period (6 months), giving more robustness  
340 to the results.

341 Overall, considering all the 16 elements evaluated in this study, we found a good correlation (mean  
342  $R^2$  of 0.93) between the online and offline XRF, with a mean slope of 1.11. The results are also in  
343 agreement with Tremper et al. (2018), which compared Xact measurements to ICP-MS daily  
344 measurements in three sites in the United Kingdom. They observed a mean  $R^2$  of 0.93 and a slope of  
345 1.07 for the elements As, Ba, Ca, Cr, Cu, Fe, K, Mn, Ni, Pb, Se, Sr, Ti, V, and Zn. In the study by  
346 Furger et al. (2017), they found instead that the elemental measurements by an Xact® 625i were on  
347 average 28% higher than ICP-OES and ICP-MS measurements for S, K, Ca, Ti, Mn, Fe, Cu, Zn, Ba,  
348 and Pb.

#### 349 4 Conclusions

350 This study was realized to evaluate the performances of an Xact® 625i online XRF spectrometer. A  
351 six-month experimental campaign was carried out at the ARPA Lombardia monitoring station Milano  
352 Pascal (Milan, Italy) from July to December 2023. The instrument was configured to continuously  
353 measure 36 elements, ranging from Al to Bi, with 1-h time resolution. The measurement quality of  
354 Xact® 625i was tested by intercomparison with ED-XRF offline analyses on 24-h PM<sub>10</sub> samples with  
355 a well-established benchtop spectrometer. Xact® 625i hourly data were aggregated to 24-h means  
356 and compared to daily PM<sub>10</sub> data. The study focused on 16 elements which were measured by both  
357 techniques and were consistently above their MDLs (Al, Si, S, Cl, K, Ca, Ti, Cr, Mn, Fe, Ni, Cu, Zn,  
358 Br, Sr, and Pb).



Xact® 625i was found to be a highly reliable instrument, suitable for measurements of elemental concentration of PM<sub>10</sub> in summer and winter conditions at 1-h time resolution. Xact® 625i elemental concentrations were found to be highly correlated to the offline XRF analyses of the daily samples ( $R^2$  in the range 0.67-0.99) albeit with slopes ranging from 0.79 to 1.70. Elements were divided into three groups according to their characteristics. The first group, Group A (K, Ca, Ti, Fe, Zn, Br, Sr, and Pb), shows excellent correlation between the two measurements methods ( $R^2 > 0.95$ ) and slopes compatible to 1 (range 0.94-1.06). Group B (Si, S, Mn, and Cu) is still characterized by excellent correlations between the two techniques, but the regression slopes are not compatible to 1. Xact® 625i performances are more critical for the elements of Group C (Al, Cl, Cr, Ni). These elements are frequently under the MDL for one or both experimental techniques and show the worst correlations between the two methods ( $R^2$  ranging from 0.67 to 0.87). The measurements of Al by Xact® 625i are also characterized by a constant value which adds up to concentrations, which is not observed for the other elements.

Future work should include an intercomparison between an Xact® 625i and an offline XRF spectrometer calibrated with the same certified standards, in order to avoid biases linked to the calibration of the instruments. Moreover, it would be interesting to assess the reliability of Xact® 625i high time resolution measurements by comparing it to other instruments/technique able to perform measurements of PM elemental concentration at high time resolution, like Horiba PX-375 automatic sampler (Asano et al., 2017).

## References

- Altuwayjiri, A., Soleimanian, E., Moroni, S., Palomba, P., Borgini, A., De Marco, C., Ruprecht, A. A., and Sioutas, C.: The impact of stay-home policies during Coronavirus-19 pandemic on the chemical and toxicological characteristics of ambient PM<sub>2.5</sub> in the metropolitan area of Milan, Italy, *Sci. Total Environ.*, 758, 143582–143582, <https://doi.org/10.1016/j.scitotenv.2020.143582>, 2021.
- Amato, F., Alastuey, A., Karanasiou, A., Lucarelli, F., Nava, S., Calzolari, G., Severi, M., Becagli, S., Gianelle, V. L., Colombi, C., Alves, C., Custódio, D., Nunes, T., Cerqueira, M., Pio, C., Eleftheriadis, K., Diapouli, E., Reche, C., Minguillón, M. C., Manousakas, M.-I., Maggos, T., Vratolis, S., Harrison, R. M., and Querol, X.: AIRUSE-LIFE+: a harmonized PM speciation and source apportionment in five southern European cities, *Atmos. Chem. Phys.*, 16, 3289–3309, <https://doi.org/10.5194/acp-16-3289-2016>, 2016.



- 389 Asano, H., Aoyama, T., Mizuno, Y., and Shiraishi, Y.: Highly Time-Resolved Atmospheric  
390 Observations Using a Continuous Fine Particulate Matter and Element Monitor, *Acs Earth and Space*  
391 *Chemistry*, 1, 580–590, [10.1021/acsearthspacechem.7b00090](https://doi.org/10.1021/acsearthspacechem.7b00090), 2017.
- 392 Bhowmik, H. S., Shukla, A., Lalchandani, V., Dave, J., Rastogi, N., Kumar, M., Singh, V., and  
393 Tripathi, S. N.: Inter-comparison of online and offline methods for measuring ambient heavy and  
394 trace elements and water-soluble inorganic ions ( $\text{NO}_3^-$ ,  $\text{SO}_4^{2-}$ ,  $\text{NH}_4^+$ , and  $\text{Cl}^-$ ) in  $\text{PM}_{2.5}$  over a heavily  
395 polluted megacity, Delhi, *Atmos. Meas. Tech.*, 15, 2667–2684, [https://doi.org/10.5194/amt-15-2667-](https://doi.org/10.5194/amt-15-2667-2022)  
396 2022, 2022.
- 397 Brunekreef, B. and Holgate, S. T.: Air pollution and health, *Lancet*, 360, 1233–1242,  
398 [https://doi.org/10.1016/S0140-6736\(02\)11274-8](https://doi.org/10.1016/S0140-6736(02)11274-8), 2002.
- 399 Calzolari, G., Chiari, M., Lucarelli, F., Nava, S., and Portarena, S.: Proton induced  $\gamma$ -ray emission  
400 yields for the analysis of light elements in aerosol samples in an external beam set-up, *Nucl. Instrum.*  
401 *Methods Phys. Res. B*, 268, 1540–1545, <https://doi.org/10.1016/j.nimb.2010.03.002>, 2010.
- 402 Calzolari, G., Lucarelli, F., Chiari, M., Nava, S., Giannoni, M., Carraresi, L., Prati, P., and Vecchi, R.:  
403 Improvements in PIXE analysis of hourly particulate matter samples, *Nucl. Instrum. Methods Phys.*  
404 *Res. B*, 363, 99–104. <https://doi.org/10.1016/j.nimb.2015.08.022>, 2015.
- 405 Chen, L. C. and Lippmann, M.: Effects of metals within ambient air particulate matter (PM) on human  
406 health, *Inhal. Toxicol.*, 21, 1–31, <https://doi.org/10.1080/08958370802105405>, 2009.
- 407 Currie, L.: Detection and quantification in X-ray fluorescence spectrometry. In: *X-Ray fluorescence*  
408 *analysis of environmental samples*, edited by: Dzubay, T. G., IX, Ann Arbor Science Publishers, 289–  
409 306, 1977.
- 410 Daellenbach, K. R., Uzu, G., Jiang, J., Cassagnes, L.-E., Leni, Z., Vlachou, A., Stefenelli, G.,  
411 Canonaco, F., Weber, S., Segers, A., Kuenen, J. J. P., Schaap, M., Favez, O., Albinet, A., Aksoyoglu,  
412 S., Dommen, J., Baltensperger, U., Geiser, M., El Haddad, I., Jaffrezo J.-L., and Prévôt, A. S. H.:  
413 Sources of particulate-matter air pollution and its oxidative potential in Europe, *Nature*, 587(7834),  
414 414–419, <https://doi.org/10.1038/s41586-020-2902-8>, 2020.
- 415 Deming, W. E.: *Statistical adjustment of data*, John Wiley & Sons; Chapman & Hall, New York:  
416 London, 261 pp., 1943.



- 417 Furger, M., Minguillón, M. C., Yadav, V., Slowik, J. G., Hüglin, C., Fröhlich, R., Petterson, K.,  
418 Baltensperger, U., and Prévôt, A. S. H.: Elemental composition of ambient aerosols measured with  
419 high temporal resolution using an online XRF spectrometer, *Atmos. Meas. Tech.*, 10, 2061–2076,  
420 <https://doi.org/10.5194/amt-10-2061-2017>, 2017.
- 421 Fuzzi, S., Baltensperger, U., Carslaw, K., Decesari, S., van Der Gon, H. D., Facchini, M. C., Fowler,  
422 D., Koren, I., Langford, B., Lohmann, U., Nemitz, E., Pandis, S., Riipinen, I., Rudich, Y., Schaap,  
423 M., Slowik, J. G., Spracklen, D. V., Vignati, E., Wild, M., Williams, M., and Gilardoni, S.: Particulate  
424 matter, air quality and climate: lessons learned and future needs, *Atmos. Chem. Phys.*, 15, 8217–  
425 8299, <https://doi.org/10.5194/acp-15-8217-2015>, 2015.
- 426 Hunter, C.B. and Rhodes, J.R.: Particle size effects in X-ray emission analysis: formulae for  
427 continuous size distributions. *X-Ray Spectr.*, 1, 107–111, 1972.
- 428 Jenkins, R., Gould, R. W., Gedcke, D. (editors); *Quantitative X-ray Spectrometry*, Marcel Dekker  
429 Inc., <https://doi.org/10.1201/9781482273380>, 1981.
- 430 Kelly, F. J., Fuller, G. W., Walton, H. A., and Fussell, J. C.: Monitoring air pollution: Use of early  
431 warning systems for public health, *Respirology*, 17, 7–19, [https://doi.org/10.1111/j.1440-](https://doi.org/10.1111/j.1440-1843.2011.02065.x)  
432 [1843.2011.02065.x](https://doi.org/10.1111/j.1440-1843.2011.02065.x), 2012.
- 433 Lindgren, E. S.: Energy Dispersive X-Ray Fluorescence Analysis. In *Encyclopedia of Analytical*  
434 *Chemistry* (editors R.A. Meyers and R.E. Grieken), <https://doi.org/10.1002/9780470027318.a6806>,  
435 2006.
- 436 Lucarelli, F., Nava, S., Calzolari, G., Chiari, M., Udisti, R., and Marino, F.: Is PIXE still a useful  
437 technique for the analysis of atmospheric aerosols? The LABEC experience. *X-Ray Spectrom.*, 40,  
438 162–167. <https://doi.org/10.1002/xrs.1312>, 2011.
- 439 Malaguti, A., Mircea, M., La Torretta, T. M. G., Telloli, C., Petralia, E., Stracquadanio, M., and  
440 Berico, M.: Comparison of online and offline methods for measuring fine secondary inorganic ions  
441 and carbonaceous aerosols in the central mediterranean area, *Aerosol Air Qual. Res.*, 15, 2641–2653,  
442 <https://doi.org/10.4209/aaqr.2015.04.0240>, 2015.
- 443 Marazzan, G. M., Vaccaro, S., Valli, G., and Vecchi, R.: Characterization of PM<sub>10</sub> and PM<sub>2.5</sub>  
444 particulate matter in the ambient air of Milan (Italy), *Atmos. Environ.*, 35, 4639–4650,  
445 [https://doi.org/10.1016/S1352-2310\(01\)00124-8](https://doi.org/10.1016/S1352-2310(01)00124-8), 2001.



- 446 Olesik, J. W.: Elemental analysis using ICP-OES and ICP/MS, *Anal. Chem.*, 63, 12A-21A, 1991.
- 447 Park, S. S., Cho, S. Y., Jo, M. R., Gong, B. J., Park, J. S., and Lee, S. J.: Field evaluation of a near-  
 448 real time elemental monitor and identification of element sources observed at an air monitoring  
 449 supersite in Korea, *Atmos. Pollut. Res.*, 5, 119–128, <https://doi.org/10.5094/APR.2014.015>, 2014.
- 450 Rohr, A. C. and Wyzga, R. E.: Attributing health effects to individual particulate matter constituents,  
 451 *Atmos. Environ.*, 62, 130-152, <https://doi.org/10.1016/j.atmosenv.2012.07.036>, 2012.
- 452 Tremper, A. H., Font, A., Priestman, M., Hamad, S. H., Chung, T.- C., Pribadi, A., Brown, R. J. C.,  
 453 Goddard, S. L., Grassineau, N., Petterson, K., Kelly, F. J., and Green, D. C.: Field and laboratory  
 454 evaluation of a high time resolution x-ray fluorescence instrument for determining the elemental  
 455 composition of ambient aerosols, *Atmos. Meas. Tech.*, 11, 3541–3557, [https://doi.org/10.5194/amt-](https://doi.org/10.5194/amt-11-3541-2018)  
 456 11-3541-2018, 2018.
- 457 U.S. EPA: Determination of metals in ambient particulate matter using X-Ray Fluorescence (XRF)  
 458 Spectroscopy, Agency, edited by: US-EPA (US Environmental Protection Agency), Cincinnati, OH  
 459 45268, USA, 1999.
- 460 U.S. EPA: Environmental Technology Verification Report – Cooper Environmental Services LLC†  
 461 Xact 625 Particulate Metals Monitor, EPA/600/R-12/680, [https://archive.epa.gov/nrmrl/archive-](https://archive.epa.gov/nrmrl/archive-etv/web/pdf/p100fk6b.pdf)  
 462 etv/web/pdf/p100fk6b.pdf, last access: September 2012.
- 463 Van Grieken R.E. and Marcowicz A.A. (editors): *Handbook of X-Ray spectrometry: methods and*  
 464 *techniques*, Marcel Dekker Inc., 1993. ISBN 0-8247-8483-9
- 465 Vecchi, R., Bernardoni, V., Fermo, P., Lucarelli, F., Mazzei, F., Nava, S., Prati, P., Piazzalunga, A.,  
 466 and Valli, G.: 4-hours resolution data to study PM<sub>10</sub> in a “hot spot” area in Europe, *Environ. Monit.*  
 467 *Assess.*, 154, 283–300. <https://doi.org/10.1007/s10661-008-0396-1>, 2009.
- 468 Vecchi, R., Marcazzan, G., and Valli, G.: A study on nighttime - daytime PM<sub>10</sub> concentration and  
 469 elemental composition in relation to atmospheric dispersion in the urban area of Milan (Italy), *Atmos.*  
 470 *Environ.*, 41, 2136–2144, <https://doi.org/10.1016/j.atmosenv.2006.10.069>, 2007.
- 471 Vecchi, R., Piziali, F. A., Valli, G., Favaron, M., and Bernardoni, V.: Radon-based estimates of  
 472 equivalent mixing layer heights: A long-term assessment, *Atmos. Environ.*, 197, 150–158.  
 473 <https://doi.org/10.1016/j.atmosenv.2018.10.020>, 2019.



474 Visser, S., Slowik, J. G., Furger, M., Zotter, P., Bukowiecki, N., Canonaco, F., Flechsig, U., Appel,  
475 K., Green, D. C., Tremper, A. H., Young, D. E., Williams, P. I., Allan, J. D., Coe, H., Williams, L.  
476 R., Mohr, C., Xu, L., Ng, N. L., Nemitz, E., Barlow, J. F., Halios, C. H., Fleming, Z. L., Baltensperger,  
477 U., and Prévôt, A. S. H.: Advanced source apportionment of size-resolved trace elements at multiple  
478 sites in London during winter, *Atmos. Chem. Phys.*, 15, 11291–11309, [https://doi.org/10.5194/acp-](https://doi.org/10.5194/acp-15-11291-2015)  
479 15-11291-2015, 2015.

480 WHO: Review of evidence on health aspects of air pollution – REVIHAAP Project, WHO,  
481 Copenhagen, technical report, 2013.

482

#### 483 **Author contribution**

484 **LC:** Data curation, Formal analysis, Visualization, Investigation, Writing – original draft preparation,  
485 Writing – review & editing; **BB:** Data curation, Investigation, Writing – review & editing; **BC:** Data  
486 curation, Writing – review & editing; **CC:** Resources, Project administration, Writing – review &  
487 editing; **RC:** Investigation; **EC:** Investigation, Data curation, Validation; Writing – review & editing;  
488 **MIM:** Data curation, Investigation, Writing – review & editing; **KRD:** Resources; Validation;  
489 Writing – review & editing; **ASHP:** Project administration, Resources, Writing – review & editing;  
490 **RV:** Supervision, Validation, Writing – review & editing.

491

#### 492 **Competing interests**

493 The authors declare that they have no conflict of interest.

494

#### 495 **Acknowledgements**

496 The Department of Physics of the University of Milan is acknowledged for the fellowship provided  
497 to Laura Cadeo.

498

499