

1 **Intercomparison of online and offline XRF spectrometers for determining the** 2 **PM₁₀ elemental composition of ambient aerosol**

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13 **Keywords: X-ray fluorescence analysis; XRF intercomparison, on-line spectrometry, PM₁₀**
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15 **Abstract**

16 Measuring the elemental composition of atmospheric particulate matter (PM) can provide useful
17 information on the adverse effects of PM and help the identification of emission sources. Carrying
18 out these measurements at a high time resolution (1-h or less) allows to describe the fast processes to
19 which aerosol particles are subjected in the atmosphere, leading to a better characterisation of the
20 emissions. Energy dispersive X-ray fluorescence spectrometry (ED-XRF) is one of the most
21 widespread techniques used to determine the elemental composition of PM. In recent years, new
22 systems known as online XRF spectrometers have been developed to provide real-time measurements
23 of the PM elemental concentration at a high time resolution. Among these advanced instruments, the
24 Xact® 625i Ambient Metals Monitor by Cooper Environmental (USA) performs in situ automated
25 measurements with a user selected time resolution ranging from 15 to 240 min. In this study, an
26 Xact® 625i monitor equipped with a PM₁₀ inlet was deployed for nearly 6 months (July-December
27 2023) in Milan (Po Valley, Italy) at a monitoring station of the Lombardy Regional Agency for
28 Environmental Protection (ARPA Lombardia). The instrument was configured to quantify 36
29 elements, ranging from Al to Bi, with 1-h time resolution in the PM₁₀ fraction. The objective of the
30 study was to verify the correct functioning of the instrument and to evaluate the quality and robustness
31 of the data produced. Xact® 625i data were aggregated to 24-h daily means and then compared to
32 24-h PM₁₀ filter data retrieved by ARPA Lombardia in the same station and analyzed offline for the
33 elemental concentration with a benchtop ED-XRF spectrometer. The intercomparison focused on the
34 16 elements (Al, Si, S, Cl, K, Ca, Ti, Cr, Mn, Fe, Ni, Cu, Zn, Br, Sr, and Pb) whose concentrations
35 were consistently above their minimum detection limits (MDL) for both online and offline
36 techniques. Results of the intercomparison were satisfying showing that the Xact® 625i elemental

37 concentrations were found to be highly correlated to the offline ED-XRF analyses (R^2 ranging from
38 0.67 to 0.99) and slopes ranging from 0.79 to 1.3 (just a couple of elements showed slopes up to 1.70).

39 **1. Introduction**

40 Measurement and quantification of the chemical composition of atmospheric particulate matter (PM)
41 are key aspects of air quality monitoring. It has long been known that PM is associated with adverse
42 impacts, which are influenced by the chemical composition of the particles. At the global scale, PM
43 affects cloud formation and Earth's radiative budget (Fuzzi et al., 2015); at the local scale, its
44 harmfulness on human health is of particular concern (Brunekreef and Holgate, 2002; Kelly et al.,
45 2012; Rohr and Wyzga, 2012; Daellenbach et al., 2020). Therefore, it is important to achieve a more
46 detailed knowledge about which chemical components are responsible for these negative effects.
47 Determining the composition of PM is also a fundamental step to perform source apportionment
48 studies for the identification of the emission sources, which help the implementation of mitigation
49 strategies (WHO, 2013).

50 Trace elements, in particular metals, although they generally do not contribute substantially to the
51 mass of PM are of interest because they act as tracers for specific sources (Visser et al., 2015) and
52 some of them are associated to adverse health effects even at ambient level concentrations (Chen and
53 Lippmann, 2009). The quantification of these elements in PM samples can be obtained through
54 various techniques (see e.g., Ogrizek et al., 2022), among the most widespread there are e.g., energy
55 dispersive X-ray fluorescence spectrometry (ED-XRF), particle-induced X-ray emission (PIXE), and
56 wet-chemistry inductively coupled plasma mass spectrometry (ICP-MS). All these methods require
57 the collection of aerosol particles on filters, followed by laboratory analysis. ED-XRF is a non-
58 destructive technique and does not require any sample pre-treatment (e.g., repeated analyses on the
59 same sample and quantification of different chemical components in the same sample are possible),
60 detects simultaneously multiple elements (20-30) with $Z > 10$ using an X-ray tube for irradiating the
61 samples, and it is typically operated using benchtop instruments. For decades until today, it has been
62 largely applied to aerosol analysis in research laboratories as well as in monitoring networks like e.g.,
63 the U.S. Environmental Protection Agency Chemical Speciation Network
64 (<https://www.epa.gov/amtic/chemical-speciation-network-csn-general-information>). One advantage
65 of ED-XRF is that it is quite stable and does not require frequent calibrations so that it is suitable for
66 automated spectrometer development.

67 PIXE analysis uses accelerated particles (often protons with energies of a few MeV) as irradiation
68 source and it has been traditionally used to assess the elemental composition in aerosol filter samples
69 (see e.g., Lucarelli et al., 2020; and therein cited literature). Although being more sensitive than ED-

70 XRF, PIXE has some features in common with ED-XRF such as the capability of providing
71 quantitative information for elements with $Z > 10$ (being both based on fluorescence X-rays detection),
72 the unnecessary sample pre-treatment and the non-destructive character. While the need of an
73 accelerator facility makes the beam-time availability for PIXE analysis a shortcoming, the existence
74 of very effective PIXE set-up where a high number of filter samples can be robustly and effectively
75 analyzed in short times helps a lot in large monitoring campaigns with many samples to be
76 characterized. As an example, at the INFN-LABEC facility in Florence, Italy, the typical irradiation
77 time for each daily aerosol sample is 45-60 s depending on the mass loading (vs. approximately 1-h
78 per sample with ED-XRF) and, more interestingly, also 1-h resolution samples can be analyzed in 1
79 min per spot (see e.g., Calzolari et al., 2010, 2015; Lucarelli et al., 2011).

80 ICP-MS is a very sensitive and fast analytical technique for detecting trace and ultra-trace elements
81 (> 50 elements simultaneously) in aerosol samples (see e.g., Duarte et al., 2021); it is ideal for heavy
82 metals accurate quantification which is performed on solubilized samples by strong acid digestion
83 thus requiring a time-consuming step, introducing a dependence on the extraction efficiency and
84 possible sample contaminations, and destroying the filter sample. In addition, ICP-MS instruments
85 need frequent calibrations and strict quality control checks to ensure stable and robust element
86 detection. As far as aerosol source tracers are concerned, a major drawback of ICP-MS is the poor
87 detection of elements like Si which is a key tracer for mineral dust particles (see e.g., Canepari et al.,
88 2009; Niu et al., 2010).

89 It is well-known that the ED-XRF technique is characterized by higher minimum detection limits
90 (MDL) compared to ICP-MS (up to two orders of magnitude; see e.g., Hyslop et al., 2024) and PIXE
91 (up to one order of magnitude; see e.g., Calzolari et al., 2008); this is a limiting factor when very low
92 aerosol loadings or trace/ultra-trace elements are of interest but e.g., for source apportionment
93 purposes it proved to be effective also when analyzing sub-daily samples or size-segregated samples
94 (see e.g., Bernardoni et al., 2011a; Bernardoni et al., 2011b). The filter type used for the aerosol
95 samples also play a role in the technique performance as reported by previous literature works (see
96 e.g., Calzolari et al., 2008; Ogrizek et al., 2022). As far as low Z elements are concerned, especially
97 for heavy loaded samples (Hyslop et al., 2024), a limitation of techniques based on the detection of
98 fluorescence X-rays is the matrix effect, whereby emitted X-rays are reabsorbed by other particles in
99 the sample matrix or are self-absorbed within single coarse particles (Hunter, and Rhodes, 1972; Van
100 Grieken and Markowicz, 1993) thus leading to an underestimation of the low- Z elemental
101 concentrations. However, these effects can be properly taken into account using correction factors
102 that can be either experimentally retrieved (see e.g., the use of PIGE-Particle Induced Gamma-ray

103 Emission analysis jointly with PIXE in Ariola et al., 2002; Calzolari et al., 2014) or by theoretical
104 calculations (see e.g., Hunter and Rhodes, 1972a, 1972b; Criss, 1976; Foster et al., 1996).

105 PM samples are usually collected by air quality monitoring networks with a time resolution ranging
106 from 24-h to one week, to ensure that enough PM mass is available for the analytical analysis, which
107 is carried out in a laboratory. The elemental composition of PM is then obtained with a considerable
108 time delay and at low temporal resolution. In recent years, there has been a growing interest in
109 developing instruments for high temporal resolution measurements. Sampling at a high time
110 resolution (1-h or less) allows to capture fast processes which aerosol particles are subjected to in the
111 atmosphere and to retrieve information about the typical hours of activity of a certain source, leading
112 to a better characterization of PM emissions. However, due to the short integration times, high-time
113 resolution measurements are often close to the MDL of the analytical techniques (Malaguti et al.,
114 2015).

115 Regarding the ED-XRF method, new systems have been developed which are able to sample PM
116 particles with a sub-hourly or hourly time resolution and to automatically measure their elemental
117 concentration, providing near-real time data access. These instruments are known as online XRF
118 spectrometers and can be employed for long monitoring periods (months, years) at a site with the
119 advantage of requiring limited maintenance. However, their high cost may prevent the simultaneous
120 use of multiple devices at different sites or the investigation of different size classes (Furger et al.,
121 2017). One of these advanced instruments is the Xact® 625i Ambient Metals Monitor by Cooper
122 Environmental (USA), which performs in situ automated measurements of the elemental
123 concentration of PM with a user selected time resolution ranging from 15 to 240 min. During
124 operation, remote access to the data is available, enabling continuous, near-real-time monitoring of
125 the instrument and ambient metal concentrations. Although the Xact® is currently one of the most
126 widely used online ED-XRF analyzers, it is worth noting that other instruments with similar working
127 principles are also available, such as the Horiba PX-375 ED-XRF monitor, whose setup and
128 performance are described in detail in Asano et al., (2017) and Trebs et al., (2024).

129 The Xact® 625i and its forerunner versions have been successfully employed in several field studies
130 in the past years, which compared its online measurements to daily samples analyzed with more
131 established laboratory techniques (Bhowmik et al., 2022; Tremper et al., 2018; Furger et al., 2017;
132 Park et al., 2014; US-EPA, 2012). Among these studies, only in Park et al., (2014) the daily filters
133 were analyzed with the ED-XRF technique; in all the other cases, the elemental concentration of daily
134 samples was retrieved with ICP-MS and ICP-OES (inductively coupled plasma optical emission
135 spectrometry). In the latest cases, the comparison was then influenced by the different choice of the
136 analytical technique. Moreover, in most of these studies, the experimental campaigns were carried

out only for a few weeks/months, leading to a very limited number of points available for the intercomparison. An evaluation of the performances of Xact® 625i (compared with the ICP-MS technique) during different seasons was conducted only by Bhowmik et al., (2022) who conducted the field campaigns during summer (June-July) and winter (October-December) 2019 at two sites in Delhi.

In this study, an Xact® 625i monitor was deployed for nearly 6 months (July-December 2023) in Milan (Po Valley, Italy) at a monitoring station of the Lombardy Regional Agency for Environmental Protection (ARPA Lombardia), where air quality measurements are performed continuously. Xact® 625i hourly samples measured online with ED-XRF were compared to daily filters measured offline by ARPA Lombardia with a benchtop ED-XRF spectrometer in their laboratory. The goals of this paper are (1) to assess the on-line instrument performance in typical summer and winter elemental concentration ranges for PM₁₀ collected at an urban background site in the Po valley (Italy); (2) to evaluate the quality of the obtained data for the selected elements in relation to their MDLs; (3) to quantify the data robustness based on intercomparison between Xact® 625i and elemental concentrations provided by a benchtop ED-XRF spectrometer.

2. Materials and methods

2.1 Site characteristics

The field campaign was performed at the permanent station Milano Pascal of the ARPA Lombardia Air Quality Network from 6 July until 12 December 2023. This is an urban background site located in the eastern side of Milan, in the University campus area called “Città Studi” (45.478°N, 9.231°E; 122 m a.s.l); the station is placed in a public park about 130 m from road traffic. The metropolitan city of Milan is the second most densely populated area in Italy (ca. 2300 inhabitants km⁻², almost doubled by daily commuters) and is located in the Po Valley, a well-known European pollution hotspot. The site is characterized by wintertime episodes of high pollutant concentrations, due to emissions from a variety of sources (e.g., residential heating, traffic, and industries) and prolonged atmospheric stability conditions related to the presence of the mountain chains of the Alps and the Apennines (Vecchi et al., 2007, 2009). Moreover, in Milan more than 80% of the days in a year are characterized by wind speed lower than 2 m s⁻¹ (Vecchi et al., 2019). The site is well documented with respect to gas-phase pollutants (e.g. NO_x, SO₂, O₃), PM₁₀ and PM_{2.5} chemical characterization, and source apportionment (e.g., Amato et al., 2016; Altuwayjiri et al., 2021).

167 2.2 Xact® 625i

168 The Xact® 625i Ambient Metals Monitor (Cooper Environmental Services (CES), Beaverton, OR,
169 USA) is an online energy dispersive XRF spectrometer, designed for continuous measurements of
170 the elemental composition of ambient aerosol. The device operates using a reel-to-reel filter tape
171 sampling technique, followed by the analysis of metals in the resulting PM spot through energy
172 dispersive X-ray fluorescence (ED-XRF). Ambient air is drawn inside the instrument through a PM
173 size-selective inlet which was PM₁₀ in this study, with a flow rate of 16.7 lpm and the PM is collected
174 onto a Teflon filter tape. After each sampling interval is completed, the filter tape is automatically
175 advanced to the XRF system, where the resulting PM deposit is irradiated with an X-ray tube
176 (rhodium anode, max voltage: 50 kV, current: 1 mA) with three excitation conditions (see Table S1
177 in the Supplementary Material) and the fluorescence X-rays are measured by a silicon drift detector
178 (SDD). In the meantime, the next sample is collected on a clean spot of the filter tape and the process
179 is repeated during each sampling interval, which was set at 60 min for this study. The XRF spectra
180 thus produced are automatically analyzed by a proprietary software for spectral analysis and
181 elemental quantification which is installed on the built-in computer. The software, through a linear
182 least-squares deconvolution algorithm, fits each measured spectrum with a library of pure element
183 reference spectra to obtain the concentration data for each calibrated element in ng m⁻³. Data can be
184 then downloaded and monitored remotely with an internet connection. Sampling and XRF analysis
185 are performed continuously and simultaneously, except for the time required for tape advancement
186 (~ 20 s). Quality assurance (QA) checks are performed every day at midnight for 30 min and consist
187 of an energy calibration (using a rod coated with Cr and Nb) and an upscale measurement to monitor
188 the stability of the instrument response (for Cr, Pb, and Cd). Therefore, the sample following midnight
189 is collected with a sampling interval limited to 30 min (00:30-01:00 LT).

190 The instrument was located inside a temperature-controlled cabinet outside the ARPA Lombardia
191 monitoring station. If any errors are detected during operation, the system halts sampling, ramps the
192 X-rays down for safety, and displays the cause of the error. The instrument was configured to quantify
193 36 elements: Al, Si, P, S, Cl, K, Ca, Ti, V, Cr, Mn, Fe, Co, Ni, Cu, Zn, Ga, Ge, As, Se, Br, Rb, Sr, Y,
194 Zr, Cd, In, Sn, Sb, I, Ba, Hg, Tl, Pb, and Bi; in addition, Nb is also detected for daily QA checks.
195 Before the beginning of the experimental campaign, each of these elements was calibrated with a
196 reference standard sample. For each element, 1 σ interference-free MDLs (MDL_{1 σ}) for 1-h of
197 sampling are reported in Table S2, provided for Xact® 625i following the approach reported in Currie
198 (1977). In XRF analyses, MDLs are inversely proportional to the square root of the irradiation time,
199 which in the case of Xact® 625i corresponds to the sampling interval. Therefore, lower MDLs are
200 reached for longer sampling durations.

201 **2.3 Daily PM₁₀ filter samples**

202 Daily PM₁₀ samples were collected on mixed cellulose ester membrane filters (47 mm diameter) with
203 a SWAM Dual Channel Monitor (FAI Instruments, Rome, Italy) equipped with PM₁₀ and PM_{2.5} inlets.
204 The elemental composition of PM samples was determined offline by ED-XRF spectrometry in the
205 laboratories of the Environmental Monitoring Sector of ARPA Lombardia. An Epsilon 4 spectrometer
206 from Malvern Panalytical (Monza, Italy) was used for the ED-XRF analysis. Four different irradiation
207 conditions, which are reported in Table S3, were chosen to optimize the measurement of 19 elements,
208 i.e., Al, Si, P, S, Cl, K, Ca, Ti, V, Cr, Mn, Fe, Ni, Cu, Zn, Br, Rb, Sr, and Pb. For these measurements,
209 MDLs based on 24-h sampling time were evaluated as three times the square root of the counts in the
210 background below the peak of the element divided by the corresponding sensitivity in the blank filter
211 (MDL_{3σ}) (Jenkins, 1981; Lindgren, 2006) and are reported in Table S4.

212 **2.4 Data coverage**

213 The Xact® 625i measurements started on 6 July 2023 16:00 LT (local time) and ended on 12
214 December 2023 22:00 LT. The sampling interval of the instrument was set to 1-h. During the summer,
215 in July and August, Xact® 625i suffered from high temperatures during the heatwaves, causing the
216 X-ray tube to reach temperatures above 45° C. This led to automatic shutdowns of the measurements
217 and to subsequent manual restarts, mostly performed remotely. The issue was mainly observed in the
218 central hours of the day, from 13:00 to 16:00 LT. Nevertheless, it was still possible to attain a data
219 coverage above 80% for Xact® 625i data in the central hours of the day during summertime. As a
220 precautionary measure to avoid heat damage to the X-ray tube, Xact® 625i was switched off from 12
221 to 24 August. During those days, a power failure in the ARPA Lombardia cabin caused also the
222 interruption of daily measurements. Another power failure occurred from 22 October to 8 November,
223 leading to a long pause of hourly measurements. The X-ray tube of Xact® 625i started malfunctioning
224 on 6 December. The issue could not be resolved and the X-ray tube had to be replaced, resulting in a
225 premature end of the experimental campaign.

226 Overall, the Xact® 625i dataset consists of 2693 valid 1-h samples out of 3822 possible samples,
227 attaining a coverage of 70%. For the daily filters, the dataset consists of 149 samples out of 157
228 possible samples, reaching a coverage of 95%. A timeline of the periods in which data are missing is
229 reported for both hourly and daily measurements in Figure S1. A summary of the periods of
230 interruption of the measurements lasting more than 12 hours is reported in Table S5. The number N
231 of overlapping days with validated data is reported in Table 1 for each element considered for the
232 intercomparison.

233 Xact® 625i data, which were originally reported in LT, were synchronized to daily samples and
234 expressed in UTC+1 time zone.

235 **2.5 Treatment of data below MDLs**

236 Following the approach of Furger et al., (2017) and Tremper et al., (2018), for the intercomparison
237 study here presented the MDL_{3σ} was considered also for Xact® 625i; indeed, the MDL_{3σ} assures a
238 high statistical confidence (99.7%) and a better comparability with previous literature works.
239 Hereafter, MDL_{3σ} will be referred simply to as MDL.

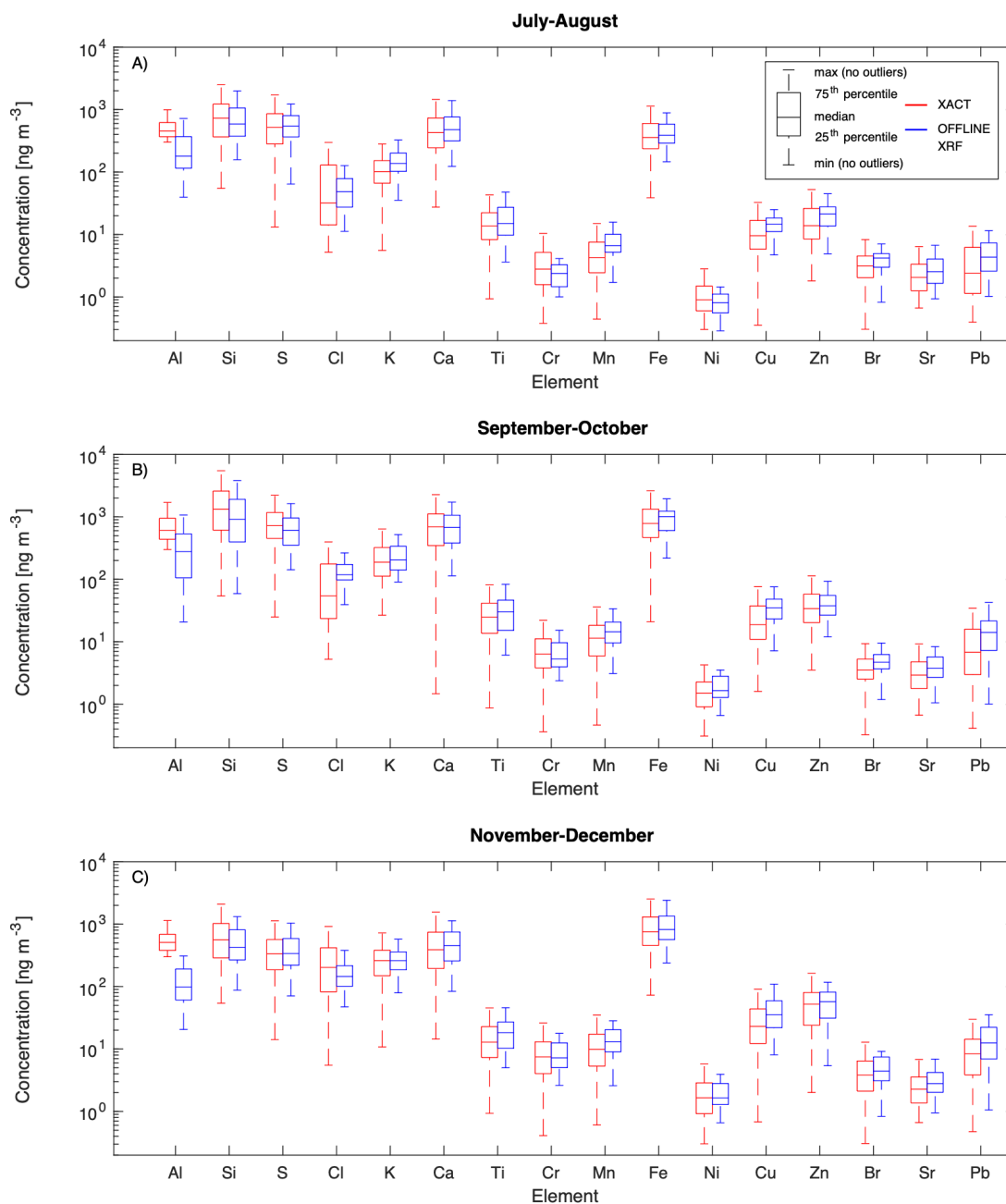
240 All the elements measured on the daily filters by offline ED-XRF have less than 35% of their data
241 points below their MDL. Among the elements detected by Xact® 625i, 13 of them (P, Co, Ga, Ge,
242 Y, Cd, In, Sn, Sb, I, Hg, Tl, and Bi) have more than 90% of their data points below MDL; therefore,
243 these elements were excluded from the intercomparison analysis. V and Rb have >70% of their data
244 points below MDL leading to a less robust intercomparison with offline ED-XRF (see Figure S2). In
245 Table S6, the number of data points with concentrations above the MDL is reported for each element
246 measured by Xact® 625i and by offline ED-XRF.

247 The intercomparison between Xact® 625i and daily PM₁₀ elemental concentrations was finally
248 performed on 16 elements (Al, Si, S, Cl, K, Ca, Ti, Cr, Mn, Fe, Ni, Cu, Zn, Br, Sr, and Pb) which
249 were measured by both techniques and were consistently above their MDLs.

250 **3. Results and discussion**

251 **3.1 Data overview**

252 An overview of the data recorded during the experimental campaign is given in Fig. 1, taking into
253 account all available valid data of the elements considered for the intercomparison. To account for
254 seasonal differences in terms of meteorology and emissions, data were divided into 3 periods: July-
255 August, September-October and November-December. The basic statistics of the dataset, including
256 the mean, median, standard deviations, 25th and 75th percentiles are reported in Table S7. As
257 previously mentioned, the Xact® 625i data coverage for July and August was impacted by the loss
258 of data mainly related to the time interval 13:00-16:00 LT when hot temperatures caused the X-ray
259 tube switch-off; therefore, the statistical robustness of the comparison is lower than in the other
260 represented periods.



261

262 **Figure 1:** Box plots of the concentrations for the elements considered for the intercomparison,
 263 measured hourly online (in red) and daily offline (in blue) during the experimental campaign in (a)
 264 July-August, (b) September-October, and (c) November-December. The bottom and the top of each
 265 box are the 25th and 75th percentiles, respectively; the line in the middle of the box is the median; the
 266 bottom and top whiskers are the minimum and maximum value respectively.

267

268 3.2 Intercomparison data analysis approach

269 For the intercomparison between the two methods, Xact® 625i hourly data were averaged to 24-h to
 270 be comparable to the corresponding daily filter samples measured by offline ED-XRF. Every day,
 271 during QA checks performed from 00:00 to 00:30 LT, Xact® 625i generates one sample with a 30-

min time resolution so that this sample is added to the 23 hourly samples to calculate 24-h daily means. This procedure implicitly assumes that the half-hour sample collected during the first hour of sampling is representative of the entire hour. The hypothesis was tested conducting 23.5 h weighted means on a small number of samples, following the method of Furger et al., (2017). Tests showed a difference of less than 3% between the 23.5 h weighted mean and the 24-h mean, which was then chosen as calculation method. For this reason, Xact® 625i data were aggregated to 24-h daily means. As previously stated, during the campaign summer days were affected by heat waves, which caused Xact® 625i to stop during the central hours of the day, leading to missing data. For this reason, the data coverage of Xact® 625i was evaluated for each day of the experimental campaign. In order to avoid misestimation of daily Xact® 625i concentrations, days with less than 18 hourly valid data (75% coverage) were excluded from the intercomparison. In addition, Xact® 625i daily means were not calculated when more than 6 hourly data were under the MDL for one day. In all comparisons, data under MDL were replaced by $0.5 \cdot \text{MDL}$.

The comparisons between the daily PM_{10} elemental concentrations retrieved by ARPA Lombardia through offline ED-XRF and the daily means calculated from Xact® 625i hourly data were carried out using the Deming regression (Deming, 1943). This regression approach minimizes the sum of distances between the regression line and the X and Y variables, considering the experimental uncertainties in both variables. For the offline ED-XRF measurements, the uncertainty included contributions of 5% from calibration standard uncertainty (U.S. EPA, 1999) and, for each spectrum, the contribution of counting statistics and fitting errors. For the Xact® 625i measurements, the uncertainty included contributions of 5% from calibration standard uncertainty (U.S. EPA, 1999), and an element-specific uncertainty derived from the spectral deconvolution calculated by the instrument software for each spectrum, which includes also the contribution of the flow and the sample deposit area. The mean relative uncertainties registered during the experimental campaign are reported for each element and for both online and offline methods in Table S8.

3.3 Intercomparison results

The results of the intercomparison between the PM_{10} elemental concentrations retrieved offline and online are reported in Table 1. The Deming regression parameters are reported along with their uncertainties and the coefficient of determination of the linear regression; the number of data (N) considered for the comparison after data reduction is also reported.

Element	Slope $\pm$ uncertainty (online vs offline)	Intercept $\pm$ uncertainty [$\mu\text{g m}^{-3}$] (online vs offline)	N	R ²
Al	1.29 ± 0.22	0.1640 ± 0.0917	42	0.83
Si	1.69 ± 0.13	-0.2528 ± 0.1003	97	0.94
S	1.25 ± 0.02	-0.0469 ± 0.0070	101	0.99
Cl	1.69 ± 0.25	-0.0493 ± 0.0290	75	0.67
K	1.05 ± 0.03	-0.0207 ± 0.0054	102	0.97
Ca	1.03 ± 0.03	0.0038 ± 0.0169	102	0.95
Ti	1.00 ± 0.06	-0.0006 ± 0.0011	100	0.96
Cr	1.30 ± 0.08	0.0014 ± 0.0004	77	0.86
Mn	0.83 ± 0.02	-0.0002 ± 0.0003	101	0.95
Fe	0.96 ± 0.02	0.0385 ± 0.0118	102	0.98
Ni	0.79 ± 0.06	0.0004 ± 0.0001	79	0.87
Cu	0.85 ± 0.01	0.0006 ± 0.0002	102	0.99
Zn	0.98 ± 0.02	-0.0001 ± 0.0008	102	0.99
Br	1.06 ± 0.04	-0.0006 ± 0.0002	96	0.96
Sr	0.98 ± 0.06	-0.0008 ± 0.0002	74	0.97
Pb	0.94 ± 0.03	-0.0017 ± 0.0003	83	0.99

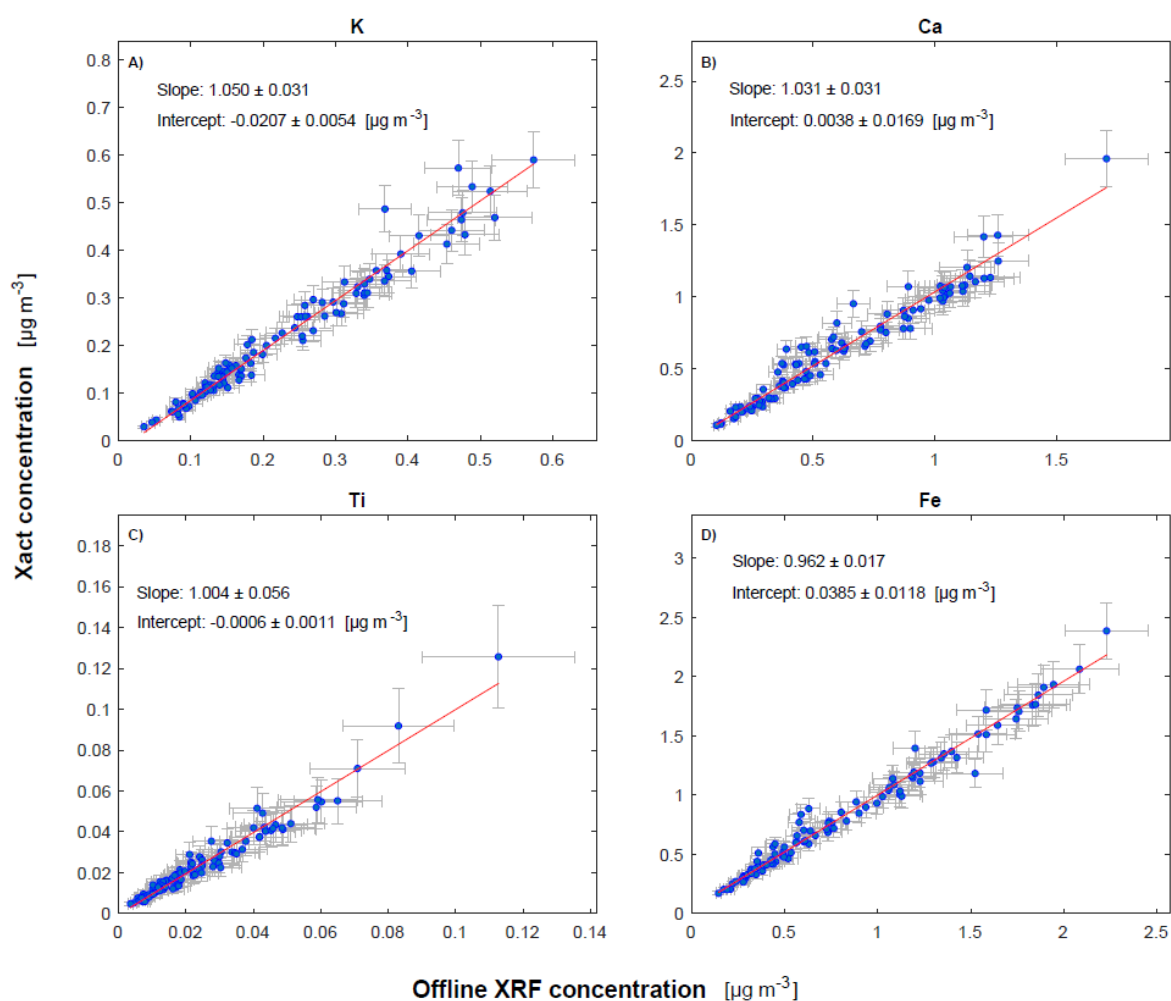
Table 1: Deming regression results and coefficient of determination for the comparison between Xact® 625i (Y) and offline ED-XRF data (X). For each element, the number of points (N) available for the intercomparison is reported.

The scatterplots of the intercomparisons are presented in Figures 2-5. The time plots of the time series obtained by the two measurements methods are reported in Figures S3-S6. The 16 selected elements are compared by dividing them into three groups based on data characteristics.

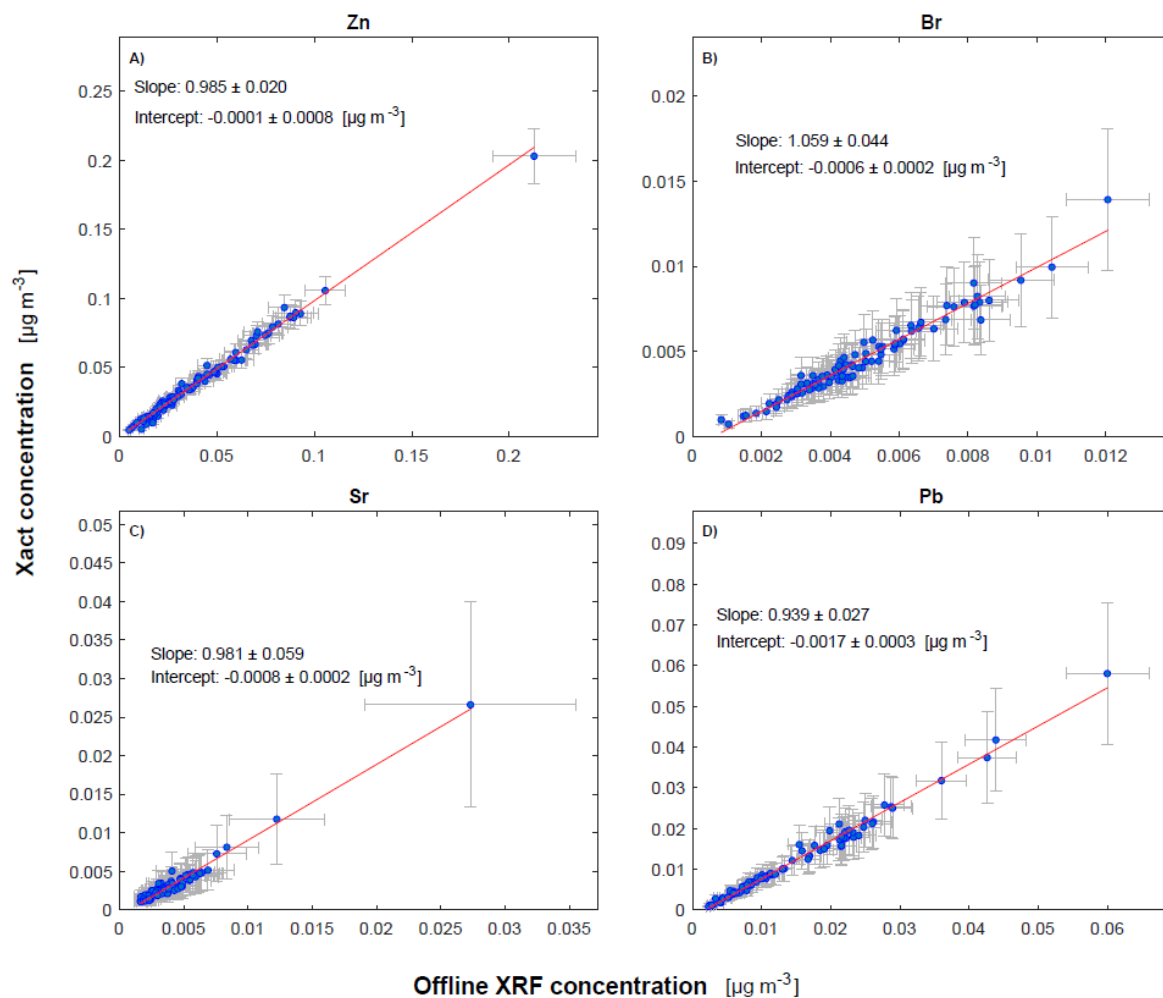
The first group, Group A (Figures 2-3), includes K, Ca, Ti, Fe, Zn, Br, Sr, and Pb. This group shows excellent correlation between the two measurements methods ($R^2 > 0.95$) and is characterized by slopes compatible to unity within three times the uncertainty of the fitted slope (3σ). For Ca, Ti, and Zn also the intercepts are compatible to 0 within 3σ . Among this group, K, Ca, Ti, Fe, and Zn, are measured by Xact® 625i with relative uncertainties in the range 10-20% (see Table S8). Br, Sr, and Pb are instead measured by Xact® 625i with a higher uncertainty, on average 30-50% (see Table S8), and Sr and Pb hourly data are also more frequently under the MDL (20% of data).

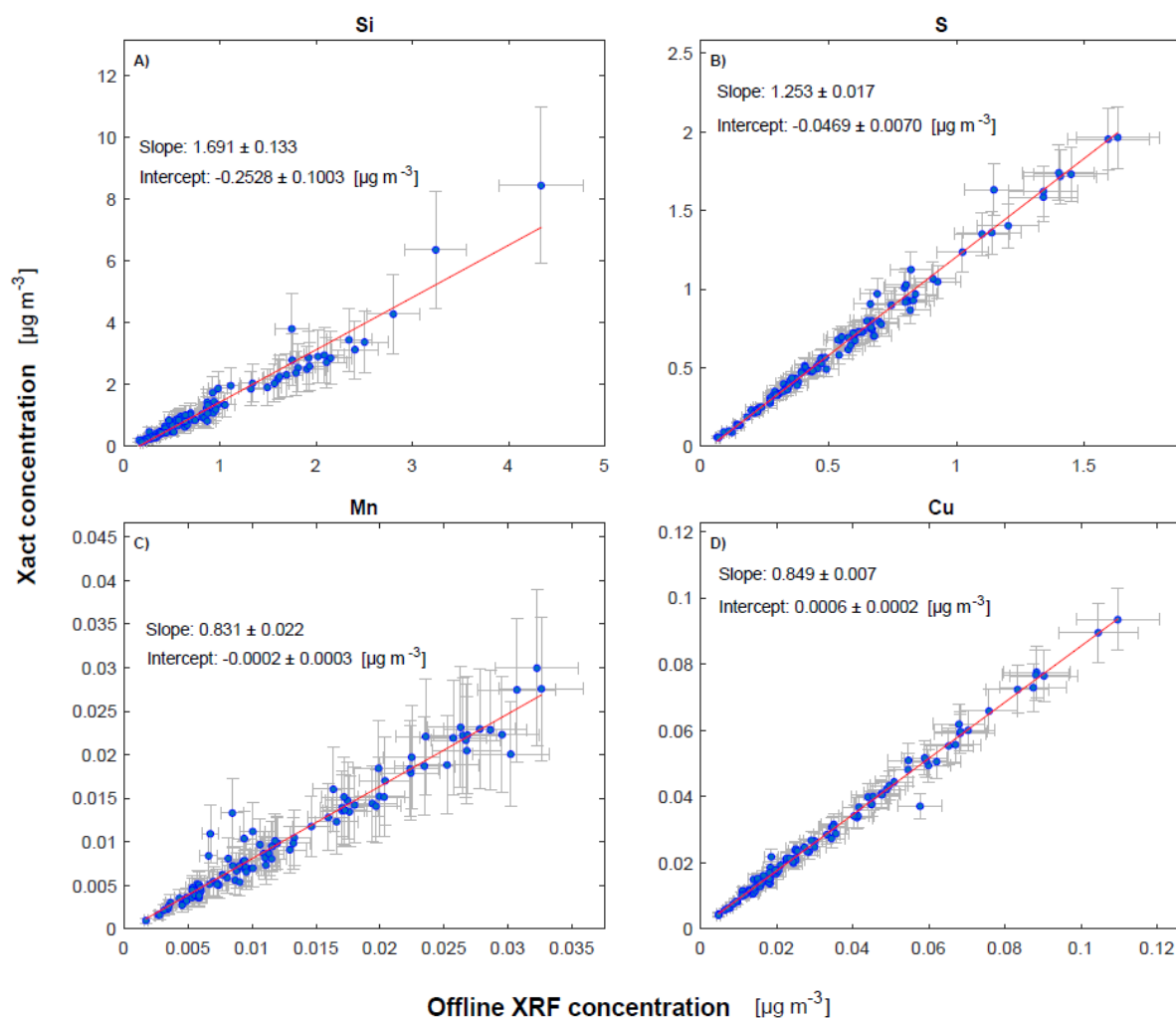
The second group, Group B (Figure 4), consists of the elements Si, S, Mn, and Cu. This group is characterized by excellent correlation between the two measurements methods ($R^2 > 0.95$) but, in contrast to Group A, the slopes of the regressions are not compatible to 1 within 3σ . Si and S are among the lightest elements measured by Xact® 625i and, along with Al, it can be tricky to measure with ED-XRF because of absorption effects due to the presence of air in the irradiation chamber (e.g. as typically occur in the XRF online measurements) and/or self-absorption inside the coarse particles

325 themselves (Hunter, and Rhodes, 1972; Van Grieken and Markowicz, 1993); these effects can lead
 326 to an underestimation of low-Z element concentrations. Nevertheless, looking at the results for Al,
 327 Si, and S, absorption effects seem not to be the cause of the observed discrepancy as Xact® 625i data
 328 are typically higher than offline ED-XRF analysis. Moreover, it should be noted that Si is detected
 329 by Xact® 625i with mean uncertainties of 30%, while S is detected with mean uncertainties of 10%.
 330 In the case of Mn and Cu, concentrations provided by Xact® 625i are constantly lower than daily
 331 offline measurements by approximately 15%.



332
 333 **Figure 2:** Scatterplots of the intercomparison between Xact® 625i data and offline ED-XRF data for
 334 the elements K, Ca, Ti, and Fe of Group A. The error bars represent the mean experimental
 335 uncertainties reported in Table S8.





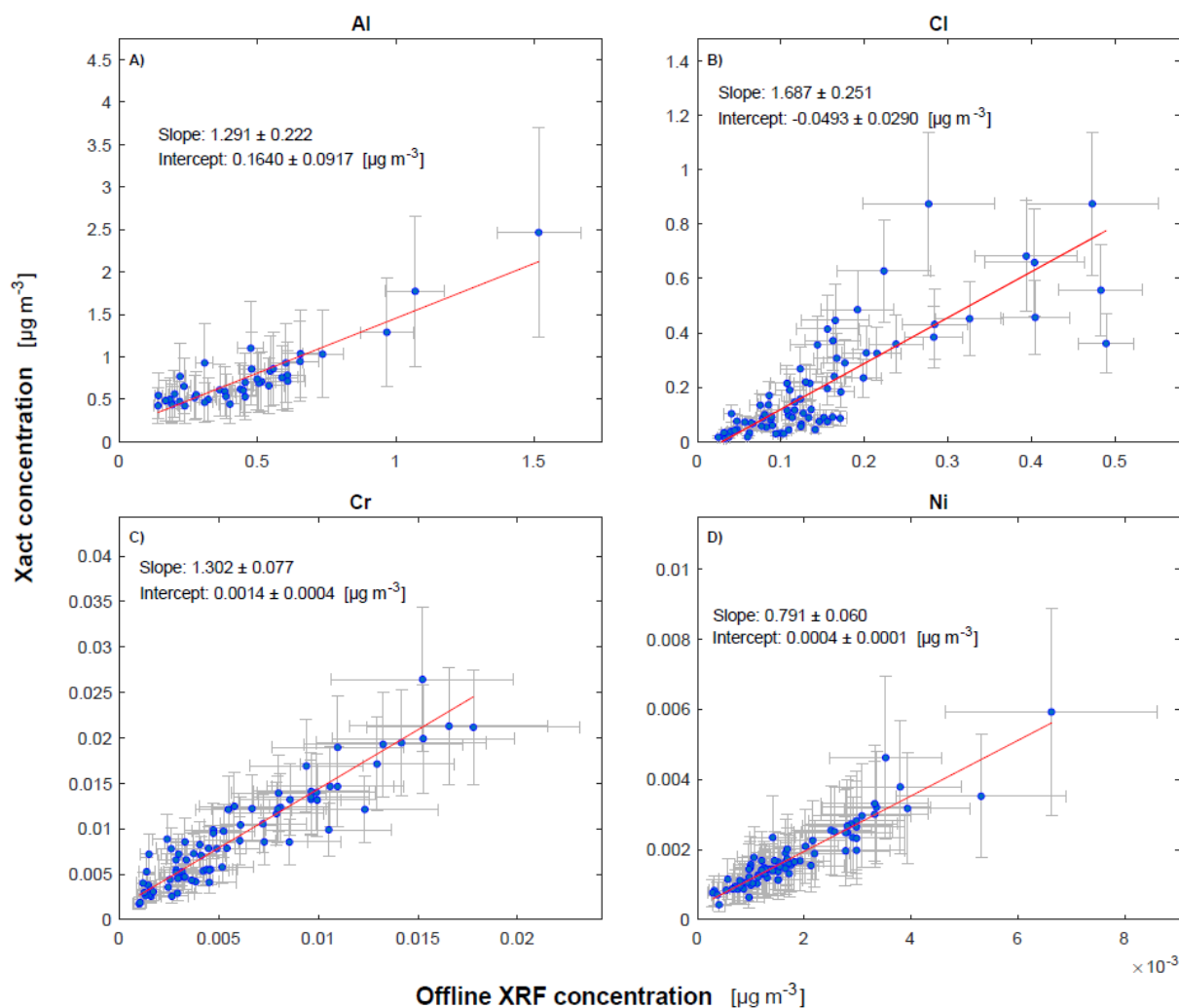
350
351

352 **Figure 4:** Scatterplots of the intercomparison between Xact® 625i data and offline ED-XRF data for
353 the elements of Group B: Si, S, Mn, and Cu. The error bars represent the mean experimental
354 uncertainties reported in Table S8.

355

356 A possible explanation for the observed discrepancies is related to the fact that, despite all samples
357 are measured through ED-XRF technique, the spectra analysis for quantitative analysis is different
358 and – more importantly - the two instruments are not calibrated with the same set of certified
359 standards, which can lead to different quantification of concentrations. However, Xact® 625i data of
360 the elements of this group can still be validated when compared to an offline measurement technique
361 and used for high-time resolution elemental concentrations assessment, after harmonisation of the
362 datasets.

363



364
365

366 **Figure 5:** Scatterplots of the intercomparison between Xact® 625i data and offline ED-XRF data for
367 the elements of Group C: Al, Cl, Cr, and Ni. The error bars represent the mean experimental
368 uncertainties reported in Table S8.

369

370 The third group of elements, Group C (Figure 5), is composed of Al, Cl, Cr, and Ni. This group shows
371 less comparability between the two methods, with R^2 in the range 0.67-0.87. Cl, Cr, and Ni are
372 frequently close or under the MDL for both experimental techniques and are characterized by mean
373 relative uncertainties in the range of 30-50%. For these elements, the comparison could be improved
374 by carrying out the Xact® 625i measurements on a 2 h time scale. Among the 16 elements evaluated
375 for the intercomparison, Al is the one with the highest MDL for Xact® 625i hourly measurements
376 and its hourly concentrations are under the MDL for nearly 35% of data points, while Al offline data
377 are always above the MDL. Al is also measured by Xact® 625i with mean uncertainties of 50%. As
378 can be seen also in Figure S6a, the Xact® 625i time series of Al is characterized by a constant upward
379 shift in background concentrations, which is not observed for the other elements. The measurement

of Al with Xact® 625i is complicated by the fact that the instrument uses an Al filter to carry out the analysis, as reported in Table S1; another possible issue could be that the X-rays hit some internal parts of the instrument, causing a significant increase of the background. Al concentrations cannot thus be corrected in a reliable way and further improvements in the instrument should be considered to enhance Al detection. In the case of Cl, which shows quite scattered data, concentrations obtained by Xact® 625i are on average higher than the ones measured offline on daily samples. This could be explained by the volatility of Cl. Xact® 625i ED-XRF measurements are performed immediately after the collection of the sample, while daily PM₁₀ filters are stored in the sampler at the monitoring station for up to 2 weeks before being taken to the laboratory for the offline ED-XRF analysis.

The results of this study represent a significant step forward from Park et al., (2014), which is – as far as we know - the only previous study available in literature presenting a comparison between Xact® hourly data and offline ED-XRF daily data. Park et al., (2014) conducted an experimental campaign with a forerunner version of Xact® (Xact® 620) in Gwangju, South Korea. The campaign was carried out during February 2011 and lasted only 1 month, focusing on the PM_{2.5} fraction. The Xact® 620 model, was the first commercially available near real time ambient metals monitor; it was able to detect elements starting from K and had higher detection limits (details can be found in Park et al., 2014), required more manual intervention for calibration and quality assurance processes and had a more basic interface with limited remote access capabilities. The daily filters were measured offline with an Epsilon 5 ED-XRF spectrometer (Malvern Panalytical). The study compared the online and offline concentrations of 12 elements (K, Ca, Ti, V, Mn, Fe, Ni, Cu, Zn, As, Ba, Pb), 9 of which were also analyzed in our study. For the 9 common elements (K, Ca, Ti, Mn, Fe, Ni, Cu, Zn, Pb), they observed a mean R^2 of 0.89 and a slope of 1.31, with Xact® measurements on average 30% higher than offline ED-XRF. In our study, for these 9 elements, we found a much better correlation, with a mean R^2 of 0.96 and slope of 0.94, which is closer to unity. Moreover, our study included also 7 elements (Al, Si, S, Cl, Cr, Br, Sr) which were not taken into account by Park et al., (2014), and the measurement campaign lasted for a longer period (6 months), giving more robustness to the results. Overall, considering all the 16 elements evaluated in this study, we found a good correlation (mean R^2 of 0.93) between the online and offline ED-XRF, with a mean slope of 1.11. The results are also in agreement with Tremper et al. (2018), which compared Xact® measurements to ICP-MS daily measurements in three sites in the United Kingdom. They observed a mean R^2 of 0.93 and a slope of 1.07 for the elements As, Ba, Ca, Cr, Cu, Fe, K, Mn, Ni, Pb, Se, Sr, Ti, V, and Zn. In the study by Furger et al. (2017), they found instead that the elemental measurements by an Xact® 625i were on average 28% higher than ICP-OES and ICP-MS measurements for S, K, Ca, Ti, Mn, Fe, Cu, Zn, Ba, and Pb.

414 A summary of previous literature studies and their characteristics is reported in Tables S9 – S10. In
415 these studies, several reasons for the differences observed between the Xact® data and the offline
416 techniques are described and some of them are shortly reported below. In general, as specified in
417 Tremper et al., (2018), the measured elements are chosen to represent a range of source categories
418 (i.e. regulatory, traffic, industry), plus the internal standard (Pd for Xact® 625 and Nb for Xact®
419 625i). The number of elements that are actually quantified and thus included in the intercomparison
420 results for each study depends on the ambient air concentrations, and thus the site, and MDL of the
421 techniques.

422 In the US-EPA (2012) work, intercomparison results are available only for 6 elements, as the others
423 were under the MDL of ICP-MS analysis and/or Xact® measurements; weak regression parameters
424 for Cu are explained by concentrations frequently close to MDL of both techniques. In Furger et al.,
425 (2017), Xact® and ICP-MS data showed high linearity and little scatter in the regressions for the
426 elements S, K, Ca, Ti, Mn, Fe, Cu, Zn, Ba, and Pb; the relative mean difference of 28% they have
427 found was attributed to many possible causes such as differences in the inlets used for the Xact® and
428 the high-volume samplers for ICP-MS filter samples, a slightly different location of the samplers,
429 possible calibration issues with the Xact® for S, values next to MDLs for one or both techniques,
430 XRF particle-size-dependent self-absorption effects for the lighter elements, and line interferences or
431 contaminations during the ICP-MS digestion and analysis procedures. Tremper et al. (2018) and
432 Bhowmik et al. (2022) both mentioned similar reasons for the differences observed between Xact®
433 and ICP-MS filter data; in addition, blank filters were found to be variable, the standards used for
434 Xact® calibration had a much higher concentration than ambient air and the calibration matrix
435 differed from the sample matrix.

436 **4 Conclusions**

437 This study was realized to evaluate the performances of an Xact® 625i online energy dispersive XRF
438 spectrometer. Although X-ray fluorescence is notably less sensitive than other analytical techniques
439 like ICP-MS, it is robust and stable so that online spectrometers can be deployed also in monitoring
440 networks due to easy use and little maintenance. Online spectrometers are still quite expensive and
441 only a reduced number of elements are detectable compared to e.g., ICP-MS but for source
442 apportionment studies the availability of high resolution elemental composition is currently key to
443 refined modelling applications. Indeed, the possibility of joining high-time resolution, which provide
444 details on temporal patterns, and low-time resolution elemental data, which allow the detection of
445 elemental tracers for specific sources, has been already proved to be effective for source
446 apportionment studies (see e.g., Crespi et al., 2016; Forello et al., 2019; Mooibroek et al, 2022).

447 A six-month experimental campaign was carried out at the ARPA Lombardia monitoring station
448 Milano Pascal (Milan, Italy) from July to December 2023. The instrument was configured to
449 continuously measure 36 elements, ranging from Al to Bi, with 1-h time resolution. The measurement
450 quality of Xact® 625i was tested by intercomparison with ED-XRF offline analyses on 24-h PM₁₀
451 samples with a well-established benchtop spectrometer. Xact® 625i hourly data were aggregated to
452 24-h means and compared to daily PM₁₀ data. The study focused on 16 elements which were measured
453 by both techniques and were consistently above their MDLs (Al, Si, S, Cl, K, Ca, Ti, Cr, Mn, Fe, Ni,
454 Cu, Zn, Br, Sr, and Pb).

455 Xact® 625i was found to be a highly reliable instrument, suitable for measurements of elemental
456 concentration of PM₁₀ in summer and winter conditions at 1-h time resolution. Xact® 625i elemental
457 concentrations were found to be highly correlated to the offline ED-XRF analyses of the daily samples
458 (R^2 in the range 0.67-0.99) albeit with slopes ranging from 0.79 to 1.70. Elements were divided into
459 three groups according to their characteristics. The first group, Group A (K, Ca, Ti, Fe, Zn, Br, Sr,
460 and Pb), shows excellent correlation between the two measurements methods ($R^2 > 0.95$) and slopes
461 compatible to 1 (range 0.94-1.06). Group B (Si, S, Mn, and Cu) is still characterized by excellent
462 correlations between the two techniques, but the regression slopes are not compatible to 1. Xact®
463 625i performances are more critical for the elements of Group C (Al, Cl, Cr, Ni). These elements are
464 frequently under the MDL for one or both experimental techniques and show the worst correlations
465 between the two methods (R^2 ranging from 0.67 to 0.87). An issue of the Xact® 625i instrument is
466 related to the quantification of Al, which is problematic so that the Al concentrations are basically
467 not reliable.

468 Future work should include an intercomparison between an Xact® 625i and an offline ED-XRF
469 spectrometer calibrated with the same certified standards, in order to avoid biases linked to the
470 calibration of the instruments. Moreover, it would be interesting to assess the reliability of Xact®
471 625i high time resolution measurements by comparing it to other instruments/technique able to
472 perform measurements of PM elemental concentration at high time resolution, like Horiba PX-375
473 ED-XRF automatic sampler (Asano et al., 2017; Trebs et al., 2024).

474 **References**

475 Altuwayjiri, A., Soleimanian, E., Moroni, S., Palomba, P., Borgini, A., De Marco, C., Ruprecht, A.
476 A., and Sioutas, C.: The impact of stay-home policies during Coronavirus-19 pandemic on the
477 chemical and toxicological characteristics of ambient PM_{2.5} in the metropolitan area of Milan, Italy,
478 Sci. Total Environ., 758, 143582–143582, <https://doi.org/10.1016/j.scitotenv.2020.143582>, 2021.

479 Amato, F., Alastuey, A., Karanasiou, A., Lucarelli, F., Nava, S., Calzolari, G., Severi, M., Becagli, S.,
 480 Gianelle, V. L., Colombi, C., Alves, C., Custódio, D., Nunes, T., Cerqueira, M., Pio, C., Eleftheriadis,
 481 K., Diapouli, E., Reche, C., Minguillón, M. C., Manousakas, M.-I., Maggos, T., Vratolis, S., Harrison,
 482 R. M., and Querol, X.: AIRUSE-LIFE+: a harmonized PM speciation and source apportionment in
 483 five southern European cities, *Atmos. Chem. Phys.*, 16, 3289–3309, [https://doi.org/10.5194/acp-16-](https://doi.org/10.5194/acp-16-3289-2016)
 484 3289-2016, 2016.

485 Ariola, V., Campajola, L., D'Alessandro, A., Del Carmine, P., Gagliardi, F., Lucarelli, F., Mandò,
 486 P.A., Marcazzan, G., Moro, R., Nava, S., Prati, P., Valli, G., Vecchi, R., Zucchiatti, A.: Aerosol
 487 characterisation in Italian towns by IBA techniques, *Nucl. Instr. Meth. Phys. Res. B*, 190 (1–4), 471-
 488 476, 2002.

489 Asano, H., Aoyama, T., Mizuno, Y., and Shiraishi, Y.: Highly Time-Resolved Atmospheric
 490 Observations Using a Continuous Fine Particulate Matter and Element Monitor, *Acs Earth and Space*
 491 *Chemistry*, 1, 580-590, [10.1021/acsearthspacechem.7b00090](https://doi.org/10.1021/acsearthspacechem.7b00090), 2017.

492 Bernardoni, V., Vecchi, R., Valli, G., Piazzalunga, A., Fermo, P.: PM₁₀ source apportionment in
 493 Milan (Italy) using time-resolved data, *Sci. Total Environ.*, 409, 4788-4795,
 494 <https://doi.org/10.1016/j.scitotenv.2011.07.048>, 2011a.

495 Bernardoni, V., Cuccia, E., Calzolari, G., Chiari, M., Lucarelli, F., Massabò, D., Nava, S., Prati, P.,
 496 Valli, G., Vecchi, R.: ED-XRF set-up for size-segregated aerosol samples analysis. *X-Ray Spectr.*,
 497 40, 79-87. doi:10.1002/xrs.1299, 2011b.

498 Bhowmik, H. S., Shukla, A., Lalchandani, V., Dave, J., Rastogi, N., Kumar, M., Singh, V., and
 499 Tripathi, S. N.: Inter-comparison of online and offline methods for measuring ambient heavy and
 500 trace elements and water-soluble inorganic ions (NO₃⁻, SO₄²⁻, NH₄⁺, and Cl⁻) in PM_{2.5} over a heavily
 501 polluted megacity, Delhi, *Atmos. Meas. Tech.*, 15, 2667–2684, [https://doi.org/10.5194/amt-15-2667-](https://doi.org/10.5194/amt-15-2667-2022)
 502 2022, 2022.

503 Brunekreef, B. and Holgate, S. T.: Air pollution and health, *Lancet*, 360, 1233–1242,
 504 [https://doi.org/10.1016/S0140-6736\(02\)11274-8](https://doi.org/10.1016/S0140-6736(02)11274-8), 2002.

505 Calzolari, G., Chiari, M., Lucarelli, F., Mazzei, F., Nava, S., Prati, P., Valli, G., and Vecchi, R.: PIXE
 506 and XRF analysis of particulate matter samples: an inter-laboratory comparison, *Nucl. Instrum.*
 507 *Methods Phys. Res. B*, 266(10), 2401-2404, <https://doi.org/10.1016/j.nimb.2008.03.056>, 2008.

508 Calzolari, G., Chiari, M., Lucarelli, F., Nava, S., and Portarena, S.: Proton induced γ -ray emission
 509 yields for the analysis of light elements in aerosol samples in an external beam set-up, Nucl. Instrum.
 510 Methods Phys. Res. B, 268, 1540–1545, <https://doi.org/10.1016/j.nimb.2010.03.002>, 2010.

511 Calzolari, G., Chiari, M., Lucarelli, F., Nava, S., Taccetti, F., Becagli, S., Frosini, D., Traversi, R.,
 512 Udisti, R.: PIXE–PIGE analysis of size-segregated aerosol samples from remote areas, Nucl. Instr.
 513 Meth. Phys. Res., 318, 125-129, <https://doi.org/10.1016/j.nimb.2013.05.097>, 2014.

514 Calzolari, G., Lucarelli, F., Chiari, M., Nava, S., Giannoni, M., Carraresi, L., Prati, P., and Vecchi, R.:
 515 Improvements in PIXE analysis of hourly particulate matter samples, Nucl. Instrum. Methods Phys.
 516 Res. B., 363, 99–104. <https://doi.org/10.1016/j.nimb.2015.08.022>, 2015.

517 Canepari, S., Perrino, C., Astolfi, M.L., Catrambone, M., Perret, D.: Determination of soluble ions and
 518 elements in ambient air suspended particulate matter: Inter-technique comparison of XRF, IC and
 519 ICP for sample-by-sample quality control. Talanta, 15, 77(5), 1821-9. doi:
 520 10.1016/j.talanta.2008.10.029, 2009.

521 Chen, L.C. and Lippmann, M.: Effects of metals within ambient air particulate matter (PM) on human
 522 health, Inhal. Toxicol., 21, 1–31, <https://doi.org/10.1080/08958370802105405>, 2009.

523 Crespi, A., Bernardoni, V., Calzolari, G., Lucarelli, F., Nava, S., Valli, G., Vecchi, R.: Implementing
 524 constrained multi-time approach with bootstrap analysis in ME-2: An application to PM_{2.5} data from
 525 Florence (Italy). Sci. Total Environ., 541, 502-511, doi: 10.1016/j.scitotenv.2015.08.159, 2016.

526 Criss, J.W.: Particle size and composition effects in X-ray fluorescence analysis of pollution samples.
 527 Analytical Chemistry, 48(1), 179-186, 1976.

528 Currie, L.: Detection and quantification in X-ray fluorescence spectrometry. In: X-Ray fluorescence
 529 analysis of environmental samples, edited by: Dzubay, T. G., IX, Ann Arbor Science Publishers, 289–
 530 306, 1977.

531 Daellenbach, K. R., Uzu, G., Jiang, J., Cassagnes, L.-E., Leni, Z., Vlachou, A., Stefenelli, G.,
 532 Canonaco, F., Weber, S., Segers, A., Kuenen, J. J. P., Schaap, M., Favez, O., Albinet, A., Aksoyoglu,
 533 S., Dommen, J., Baltensperger, U., Geiser, M., El Haddad, I., Jaffrezo J.-L., and Prévôt, A. S. H.:
 534 Sources of particulate-matter air pollution and its oxidative potential in Europe, Nature, 587(7834),
 535 414–419, <https://doi.org/10.1038/s41586-020-2902-8>, 2020.

Deming, W. E.: Statistical adjustment of data, John Wiley & Sons; Chapman & Hall, New York: London, 261 pp., 1943.

Duarte, R.M.B.O., Gomes, J.F.P., Querol, X., Cattaneo, A., Bergmans, B., Saraga, D., Villanueva, F.: Advanced instrumental approaches for chemical characterization of indoor particulate matter. *Applied Spectr. Rev.*, 57(8), 705–745, <https://doi.org/10.1080/05704928.2021.2018596>, 2021.

Forello, A.C., Bernardoni, V., Calzolari, G., Lucarelli, F., Massabò, D., Nava, S., Pileci, R.E., Prati, P., Valentini, S., Valli, G., Vecchi, R.: Exploiting multi-wavelength aerosol absorption coefficients in a multi-time source apportionment study to retrieve source-dependent absorption parameters. *Atmos. Chem. Phys.*, 19, 11235–11252, doi.org/10.5194/acp-2019-123, 2019.

Foster, R.D., Read, M.L., Usher, J.M.: Particle size and depth effect errors in the XRF determination of elements in aerosol collected onto filters. Health and Safety Laboratory, IS/96/03, 1996.

Furger, M., Minguillón, M. C., Yadav, V., Slowik, J. G., Hüglin, C., Fröhlich, R., Petterson, K., Baltensperger, U., and Prévôt, A. S. H.: Elemental composition of ambient aerosols measured with high temporal resolution using an online XRF spectrometer, *Atmos. Meas. Tech.*, 10, 2061–2076, <https://doi.org/10.5194/amt-10-2061-2017>, 2017.

Fuzzi, S., Baltensperger, U., Carslaw, K., Decesari, S., van Der Gon, H. D., Facchini, M. C., Fowler, D., Koren, I., Langford, B., Lohmann, U., Nemitz, E., Pandis, S., Riipinen, I., Rudich, Y., Schaap, M., Slowik, J. G., Spracklen, D. V., Vignati, E., Wild, M., Williams, M., and Gilardoni, S.: Particulate matter, air quality and climate: lessons learned and future needs, *Atmos. Chem. Phys.*, 15, 8217–8299, <https://doi.org/10.5194/acp-15-8217-2015>, 2015.

Hyslop, N. P., Rosales, C. M. F., Weber, F. X., Dombek, T. L., Levine, K. E., McWilliams, A. C., and Spada, N. J.: A comparison of XRF and ICP-MS for PM_{2.5} elemental analysis in the chemical speciation network, *EM Magazine*, 2024(06), Article 15, 2024.

Hunter, C.B. and Rhodes, J.R.: Particle size effects in X-ray emission analysis: formulae for continuous size distributions. *X-Ray Spectr.*, 1, 107-111, 1972a.

Hunter, C.B. and Rhodes, J.R.: Particle size effects in X-ray emission analysis: Simplified formulae for certain practical cases. *X-Ray Spectr.*, 1, 113-117, 1972b.

Jenkins, R., Gould, R. W., Gedcke, D. (editors); Quantitative X-ray Spectrometry, Marcel Dekker Inc., <https://doi.org/10.1201/9781482273380>, 1981.

565 Kelly, F. J., Fuller, G. W., Walton, H. A., and Fussel, J. C.: Monitoring air pollution: Use of early
 566 warning systems for public health, *Respirology*, 17, 7-19, [https://doi.org/10.1111/j.1440-](https://doi.org/10.1111/j.1440-1843.2011.02065.x)
 567 [1843.2011.02065.x](https://doi.org/10.1111/j.1440-1843.2011.02065.x), 2012.

568 Lindgren, E. S.: Energy Dispersive X-Ray Fluorescence Analysis. In *Encyclopedia of Analytical*
 569 *Chemistry* (editors R.A. Meyers and R.E. Grieken), <https://doi.org/10.1002/9780470027318.a6806>,
 570 2006.

571 Lucarelli, F., Nava, S., Calzolari, G., Chiari, M., Udisti, R., and Marino, F.: Is PIXE still a useful
 572 technique for the analysis of atmospheric aerosols? The LABEC experience. *X-Ray Spectrom.*, 40,
 573 162–167. <https://doi.org/10.1002/xrs.1312>, 2011.

574 Lucarelli, F.: How a small accelerator can be useful for interdisciplinary applications: the study of air
 575 pollution. *Eur. Phys. J. Plus*, 135, 538. <https://doi.org/10.1140/epjp/s13360-020-00516-3>, 2020.

576 Malaguti, A., Mircea, M., La Torretta, T. M. G., Telloli, C., Petralia, E., Stracquadanio, M., and
 577 Berico, M.: Comparison of online and offline methods for measuring fine secondary inorganic ions
 578 and carbonaceous aerosols in the central mediterranean area, *Aerosol Air Qual. Res.*, 15, 2641–2653,
 579 <https://doi.org/10.4209/aaqr.2015.04.0240>, 2015.

580 Mooibroek, D., Sofowote, U.M., Hopke, P.K.: Source apportionment of ambient PM₁₀ collected at
 581 three sites in an urban-industrial area with multi-time resolution factor analyses, *Sci.Total Environ*,
 582 850, 157981, <https://doi.org/10.1016/j.scitotenv.2022.157981>, 2022.

583 Niu, J., Rasmussen, P. E., Wheeler, A., Williams, R., and Chénier, M.: Evaluation of airborne
 584 particulate matter and metals data in personal, indoor and outdoor environments using ED-XRF and
 585 ICP-MS and co-located duplicate samples, *Atmos. Environ.*, 44(2), 235-245,
 586 <https://doi.org/10.1016/j.atmosenv.2009.10.009>, 2010.

587 Ogrizek, M., Kroflič, A., Šala, M.: Critical review on the development of analytical techniques for
 588 the elemental analysis of airborne particulate matter, *Trends Environ. Anal. Chem.*, 33, e00155,
 589 <https://doi.org/10.1016/j.teac.2022.e00155>, 2022.

590 Olesik, J. W.: Elemental analysis using ICP-OES and ICP-MS, *Anal. Chem.*, 63, 12A-21A, 1991.

591 Park, S. S., Cho, S. Y., Jo, M. R., Gong, B. J., Park, J. S., and Lee, S. J.: Field evaluation of a near-
 592 real time elemental monitor and identification of element sources observed at an air monitoring
 593 supersite in Korea, *Atmos. Pollut. Res.*, 5, 119–128, <https://doi.org/10.5094/APR.2014.015>, 2014.

594 Rohr, A. C. and Wyzga, R. E.: Attributing health effects to individual particulate matter constituents,
 595 Atmos. Environ., 62, 130-152, <https://doi.org/10.1016/j.atmosenv.2012.07.036>, 2012.

596 Trebs, I., Lett, C., Krein, A., Matsumoto Kawaguchi, E., and Junk, J.: Performance evaluation of an
 597 online monitor based on X-ray fluorescence for detecting elemental concentrations in ambient
 598 particulate matter, Atmos. Meas. Tech., 17, 6791–6805, <https://doi.org/10.5194/amt-17-6791-2024>,
 599 2024.

600 Tremper, A. H., Font, A., Priestman, M., Hamad, S. H., Chung, T.- C., Pribadi, A., Brown, R. J. C.,
 601 Goddard, S. L., Grassineau, N., Petterson, K., Kelly, F. J., and Green, D. C.: Field and laboratory
 602 evaluation of a high time resolution x-ray fluorescence instrument for determining the elemental
 603 composition of ambient aerosols, Atmos. Meas. Tech., 11, 3541–3557, [https://doi.org/10.5194/amt-](https://doi.org/10.5194/amt-11-3541-2018)
 604 11-3541-2018, 2018.

605 U.S. EPA: Determination of metals in ambient particulate matter using X-Ray Fluorescence (XRF)
 606 Spectroscopy, Agency, edited by: US-EPA (US Environmental Protection Agency), Cincinnati, OH
 607 45268, USA, 1999.

608 U.S. EPA: Environmental Technology Verification Report – Cooper Environmental Services LLC†
 609 Xact 625 Particulate Metals Monitor, EPA/600/R-12/680, [https://archive.epa.gov/nrmrl/archive-](https://archive.epa.gov/nrmrl/archive-etv/web/pdf/p100fk6b.pdf)
 610 etv/web/pdf/p100fk6b.pdf, last access: September 2012.

611 Van Grieken R.E. and Marcowicz A.A. (editors): Handbook of X-Ray spectrometry: methods and
 612 techniques, Marcel Dekker Inc., 1993. ISBN 0-8247-8483-9

613 Vecchi, R., Bernardoni, V., Fermo, P., Lucarelli, F., Mazzei, F., Nava, S., Prati, P., Piazzalunga, A.,
 614 and Valli, G.: 4-hours resolution data to study PM₁₀ in a “hot spot” area in Europe, Environ. Monit.
 615 Assess., 154, 283–300. <https://doi.org/10.1007/s10661-008-0396-1>, 2009.

616 Vecchi, R., Marcazzan, G., and Valli, G.: A study on nighttime - daytime PM₁₀ concentration and
 617 elemental composition in relation to atmospheric dispersion in the urban area of Milan (Italy), Atmos.
 618 Environ., 41, 2136–2144, <https://doi.org/10.1016/j.atmosenv.2006.10.069>, 2007.

619 Vecchi, R., Piziali, F. A., Valli, G., Favaron, M., and Bernardoni, V.: Radon-based estimates of
 620 equivalent mixing layer heights: A long-term assessment, Atmos. Environ., 197, 150–158.
 621 <https://doi.org/10.1016/j.atmosenv.2018.10.020>, 2019.

622 Visser, S., Slowik, J. G., Furger, M., Zotter, P., Bukowiecki, N., Canonaco, F., Flechsig, U., Appel,
623 K., Green, D. C., Tremper, A. H., Young, D. E., Williams, P. I., Allan, J. D., Coe, H., Williams, L.
624 R., Mohr, C., Xu, L., Ng, N. L., Nemitz, E., Barlow, J. F., Halios, C. H., Fleming, Z. L., Baltensperger,
625 U., and Prévôt, A. S. H.: Advanced source apportionment of size-resolved trace elements at multiple
626 sites in London during winter, *Atmos. Chem. Phys.*, 15, 11291–11309, [https://doi.org/10.5194/acp-](https://doi.org/10.5194/acp-15-11291-2015)
627 15-11291-2015, 2015.

628 WHO: Review of evidence on health aspects of air pollution – REVIHAAP Project, WHO,
629 Copenhagen, technical report, 2013.

630

631 **Author contribution**

632 **LC**: Data curation, Formal analysis, Visualization, Investigation, Writing – original draft preparation,
633 Writing – review & editing; **BB**: Data curation, Investigation, Writing – review & editing; **BC**: Data
634 curation, Writing – review & editing; **CC**: Resources, Project administration, Writing – review &
635 editing; **RC**: Investigation; **EC**: Investigation, Data curation, Validation; Writing – review & editing;
636 **MIM**: Data curation, Investigation, Writing – review & editing; **KRD**: Resources; Validation;
637 Writing – review & editing; **ASHP**: Project administration, Resources, Writing – review & editing;
638 **RV**: Supervision, Validation, Writing – review & editing.

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640 **Competing interests**

641 The authors declare that they have no conflict of interest.

642

643 **Acknowledgements**

644 The Department of Physics of the University of Milan is acknowledged for the fellowship provided
645 to Laura Cadeo.

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