



# Formation and composition of organic aerosols from the uptake of glyoxal on natural mineral dust aerosols: a laboratory study

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## 22 Abstract

The uptake of glyoxal on realistic submicron mineral dust aerosol particles from a 23 natural soil (Gobi Desert) is investigated during experiments in a large simulation 24 25 chamber, under variable experimental conditions of relative humidity, irradiation, and ozone concentrations. The uptake of glyoxal on the dust particles starts as soon as the 26 glyoxal is injected in the chamber. At 80% RH, the measured uptake coefficient of 27 glyoxal on mineral dust is  $y = (9 \pm 5) \times 10^{-3}$ . The totality of the mass of reacting glyoxal 28 is transformed in organic matter on the surface of the dust particles. The uptake of 29 30 glyoxal is accompanied by the appearance marker peaks in the organic mass spectra and a persistent growth in the volume concentration of the dust particles. While the 31 mass of the organic matter on the dust rapidly reverts to values prior to uptake, the 32 organic composition of the dust is modified irreversibly. Glycolic and other organic 33 acids but also oligomers are detected on the dust. At 80% RH, compounds ranging 34 from C<sub>4</sub> to C<sub>10</sub> are observed as oligomerization products of glyoxal mono- and di-35 hydrate forms. The study suggests that dust aerosols could play a very substantial role 36 in the formation of organic aerosols at high relative humidity, but also that the reaction 37 38 could have potential important implications for the dust optical and hygroscopic properties, including their pH. 39





#### 40 **1 Introduction**

Mineral dust originates naturally from the wind erosion of arid or semi-arid soils, 41 resulting in the suspension of particles with diameters from fractions to hundreds of 42 43 microns, which can be transported over thousands of kilometres whilst in the atmosphere (Adebiyi et al., 2023; Mahowald et al., 2014). The total global mass of 44 mineral dust particles emitted annually in the atmosphere is of the order of 4600 Tg 45 yr<sup>1</sup>, accounting for approximately 40% of the total annual aerosol emissions (Knippertz 46 and Stuut, 2014; Kok et al., 2021). Major natural source areas of mineral dust are North 47 Africa (~50% of the global annual dust emissions), Asia (~40%), North America, and 48 the Southern Hemisphere (~10%; Kok et al., 2023). Anthropogenic emissions are 49 associated with soil erosion for agriculture, pasture, and deforestation (Tegen and 50 51 Fung, 1995; Webb and Pierre, 2018), but their contribution to the total annual dust 52 mass loading is uncertain, ranging from 5 to 60% (Chen et al., 2023). Mineral dust significantly impacts the Earth's energy balance by absorbing and scattering radiation 53 in the solar and terrestrial spectra (Di Biagio et al., 2019; Kok et al., 2023) and by 54 55 influencing the lifetime and optical properties of mixed-phase and ice clouds (e.g., Atkinson et al., 2013; Harrison et al., 2001; Steinke et al., 2016). Current estimates of 56 the effective radiative forcing (sum of direct and indirect) of natural mineral dust are in 57 the range of  $-0.07 \pm 0.18$  W m<sup>-2</sup> (Kok et al., 2023), owing to large uncertainties in the 58 59 atmospheric mass loading and properties of dust at emission and during transport (Castellanos et al., 2024; Li et al., 2021). 60

Gas-particle interactions along the dust lifecycle contribute to these uncertainties. 61 Numerous laboratory and field studies show that mineral dust is capable to adsorb 62 various reactive gaseous compounds, which may modify its chemical composition, and 63 in turn to alter optical properties, hygroscopicity and ice nucleation activity but also may 64 affect the oxidative capacity of the atmosphere (Bauer et al., 2007; Chirizzi, 2017; 65 Crowley et al., 2010; Joshi et al., 2017; Liu et al., 2013; Ooki and Uematsu, 2005; 66 Romanias et al., 2012; Seisel et al., 2004; Tang et al., 2017; Turpin and Huntzicker, 67 1995; Usher et al., 2003; Wagner et al., 2008). Dust aerosol may promote 68 photocatalytic reactions of inorganic gases such as SO<sub>2</sub> and NO<sub>2</sub>, initiating nucleation 69 events (Dupart et al., 2012; Nie et al., 2014). 70

The uptake of volatile organic compounds (VOCs) on mineral dust such as limonene,
toluene (Romanías et al., 2016), isoprene (Zeineddine et al., 2017), phenol





(Hettiarachchi and Grassian, 2024), and dicarboxylic acids (Ponczek et al., 2019), is
also documented. These reactions may alter the VOC budget in the atmosphere and
lead to the formation of secondary organic aerosols (Li et al., 2019; Tang et al., 2017;
Usher et al., 2003; Xu et al., 2023; Zeineddine et al., 2023), one of the key player of
atmospheric chemistry (Shrivastava et al., 2017).

Glyoxal (CHOCHO) is one of the most abundant VOCs in the troposphere (Lewis et 78 79 al., 2020). Atmospheric glyoxal is produced through the oxidation of aromatic compounds like benzene, toluene, and p-xylene (Volkamer et al., 2001) as well as by 80 81 the photochemical oxidation of isoprene (Chan et al., 2017). The global atmospheric concentrations have been evaluated in the range of 10 - 100 pptv by Fu et al. (2008). 82 However, case studies show sometimes higher concentrations. During a field study in 83 Shanghai in the summer of 2018, Guo et al. (2021) reported an average glyoxal 84 concentration of 164 ± 73 pptv, due to daytime photochemistry. Local concentrations 85 of up to 400 pptv have been documented in regions influenced by aromatic pollution 86 (Li et al., 2022). Satellite measurements of glyoxal show that the highest 87 concentrations in tropical and sub-tropical regions are found during warm, dry periods 88 influenced by biogenic emissions and vegetation fires, but also anthropogenic pollution 89 (Vrekoussis et al., 2009). Elevated glyoxal concentrations have been observed in aged 90 biomass burning plumes and tropical ocean regions, revealing model under-91 predictions in high-emission areas due to missing complex organic compound sources 92 (Kluge et al., 2023). Field measurements in the north east Atlantic Ocean reveal that 93 models generally underestimate glyoxal concentrations due to missing contributions 94 from acetaldehyde and other chemical precursors, and a potential glyoxal source from 95 96 the ocean surface organic microlayer, particularly significant at night (Walker et al., 2022). 97

98 Glyoxal is a very soluble molecule which readily oligomerises in water, leading to the formation of larger molecules (Kalberer et al., 2004; Shapiro et al., 2009). It also has 99 100 the ability to uptake onto aerosol particles, potentially serving as a significant source of organic aerosols (e.g., Liggio et al., 2005b, Carlton et al., 2007; Ervens and 101 Volkamer, 2010; Galloway et al., 2009; Knote et al., 2014). The uptake of glyoxal on 102 ammonium sulphate particles can lead to the formation of carbon-nitrogen compounds 103 (such as imidazole derivatives), oligomers, and organic acids (Galloway et al., 2009), 104 that has been observed to cause their browning (De Haan et al., 2020). The light-105





absorbing imidazole derivatives formed by glyoxal have been found to act as a photosensitizer, initiating radical chemistry under realistic irradiation conditions in the aerosol phase and initiating aerosol growth in the presence of limonene (Rossignol et al., 2014).

Shen et al. (2016) revealed that glyoxal can also uptake onto synthetic minerals proxies
of natural mineral dust, forming oligomers, organo-sulphates, formic acid, and glycolic
acid, henceforth suggesting a potential significant mechanism for organic aerosol
formation and modification of the optical and hygroscopic properties of mineral dust.

Following up from the pioneering study by Shen et al. (2016), in this paper we present the results of laboratory experiments using a large-scale simulation chamber to investigate the formation of OA from the uptake of glyoxal on realistic airborne mineral dust particles. Dust aerosols are generated from a natural parent soil from the Gobi Desert, one of the most important sources of tropospheric dust and representative of an area where this interaction could take place (Wang et al., 2015).

This paper has two major objectives. First, it provides experimental observations of the uptake of glyoxal on mineral dust aerosol, leading to the formation of organic aerosol mass upon interaction and measuring glyoxal uptake coefficient of mineral dust. Secondly, it presents the chemical composition of the mixed organic-dust aerosols, in terms of its oxidation state, molecular composition and the evolution of secondary organic aerosol content from glyoxal.

#### 126 **2 Experimental**

127 This study uses the CESAM atmospheric simulation chamber, a 4.2 m<sup>3</sup> cylindrical stainless-steel reactor initially described by Wang et al. (2011). CESAM is specifically 128 129 designed to study multiphase processes involving aerosol particles, gas-phase compounds and water, both in the vapour and liquid phases (Brégonzio-Rozier et al., 130 2016; Denjean et al., 2014; Giorio et al., 2017). CESAM is equipped with three 6.5 kW 131 high-pressure arc xenon lamps (model EX-170GM3-E, IREM SpA, Borgone, Italy) and 132 6 mm Pyrex plate filters to mimic the solar radiation. A 50 cm stainless-steel four-blade 133 fan located at the bottom of the chamber ensures a mixing time of about 1 minute for 134 the gas phase and the homogeneity of the internal composition. 135

The ageing experiments last up to five hours. Before each experiment, the chamber is evacuated down to  $10^{-4}$  mbar. The chamber is then filled with a mixture of 80% N<sub>2</sub>





138 (Messer, purity > 99.995%) and 20%  $O_2$  (Linde, 5.0) to an internal pressure exceeding by about 5 to 10 mbar the local atmospheric pressure, to prevent accidental 139 contamination during the experiments. For the experiments carried out in wet 140 conditions, the injection of water vapour precedes the injection of dust. The injection of 141 glyoxal (1 ppmv) was conducted at least after 30 minutes after the dust to ensure that 142 the dust particles are homogenously distributed. Irradiation is started within one hour 143 after the glyoxal uptake onto the particles. Ozone is used to verify the sensitivity of the 144 reactions to the presence of an oxidant. For those experiments, ozone is injected 145 before glyoxal. 146

## 147 2.1 Experimental protocols

Dust aerosols are generated and injected into the chamber according to the protocol 148 149 detailed in Battaglia et al. (2024). The natural soil sample used in this study is from the Gobi Desert (107.48°N; 36.49°E). Prior use, the soil is sieved at 1000 µm and dried at 150 100°C for less than an hour to remove adsorbed water and contamination from volatile 151 gases. A quantity ranging from 30 and 50 g is placed in a 1 L Büchner flask and shaken 152 at 100 Hz using a sieve shaker (Retsch® AS200) to simulate the saltation and 153 sandblasting mechanisms through which wind erosion generates airborne dust in the 154 real atmosphere (Di Biagio et al., 2017). An Aerodynamic Aerosol Classifier (AAC, 155 Cambustion®) is placed between the dust generator and the chamber to inject mono-156 modal dust centred between 300 and 400 nm in geometric diameter. 157

Glyoxal is prepared by heating a mixture of equal amounts of its trimer hydrate (Fluka® Analytical) and  $P_2O_5$  (Sigma – Aldrich ReagentPlus<sup>®</sup>, 99%) at 150°C (Horowitz et al., 2001). The trimer decomposition occurs inside a vial connected to a vacuum gas manifold. Glyoxal is collected as yellow crystals in a second vial immersed in an ethanol – liquid nitrogen cold trap at around -90°C and then vaporised in a 2.1 L glass bulb to a controlled pressure. This vial is connected to the simulation chamber to inject the glyoxal through a nitrogen flow.

Ozone is generated by a Corona discharge in pure O<sub>2</sub> using a commercial dielectric ozone generator (MBT 802N, Messtechnik GmbH, Stahnsdorf, Germany). Water vapour is generated by heating ultrapure water (Milli-Q IQ 7000, Merk<sup>™</sup>) inside a pressurised stainless-steel vessel, previously rinsed at least three times. The total organic carbon (TOC) content of the ultrapure water is monitored in each experiment





to evaluate the influence on the production of organic particles, which was found to be
minor (see Text S1 in the Supplementary Material). The relative humidity (RH) inside
the chamber is measured by a HMP234 Vaisala<sup>®</sup> humidity and temperature
transmitter.

#### 174 **2.2 Measurements and instrumentation**

#### 175 2.2.1 Gas-phase composition

CESAM is equipped with an in-situ long-path FTIR spectrometer (Bruker Tensor 37), 176 which enables the collection of spectra with a time resolution of 5 minutes. The spectral 177 range covered is 700 – 4000 cm<sup>-1</sup>, at a resolution of 0.5 cm<sup>-1</sup> and an optical path length 178 of 120 meters. To avoid major water interference, glyoxal is quantified by integrating 179 the peak corresponding to the stretching of the C-H bonds, in the 2720 to 2930 cm<sup>-1</sup> 180 181 interval. Additional species quantified by FTIR spectroscopy are formic acid (HCOOH), carbon monoxide (CO), and ozone ( $O_3$ ). HCOOH is quantified by integrating the 182 absorption band centred at 1105 cm<sup>-1</sup>, associated with the C-O bond vibration typical 183 of the carboxylic group of formic acid. CO is quantified by integrating the typical gas 184 phase stretching absorption centred at about 2143 cm<sup>-1</sup>, and  $O_3$  is quantified by 185 integrating the absorption band of its asymmetric stretching, centred at about 1043 186 cm<sup>-1</sup>. 187

A CAPS (Cavity Attenuated Phase Shift) NO<sub>2</sub> analyser (Model T500U, from Teledyne 188 API), with a concentration range of 10 - 1000 ppbv and an integration time of 30 s was 189 also connected to the chamber for glyoxal detection. VOCs are monitored by a PTR-190 ToF-MS (KORE Technology<sup>®</sup>, second generation) operated in H<sub>3</sub>O<sup>+</sup> ionization mode 191 at a time resolution of 1 minute. The reactor pressure and temperature are 1.35 mbar 192 and 60°C, respectively, leading to an E/N ratio of 131 Td, where E is the electric field 193 and N is the concentration of neutral particles. This ratio is used to determine the 194 effectiveness of the ion-molecule collisions and is given in the unit Townsend (Td). Ion 195 signals measured by PTR-MS are normalized by signals of reagent ions (i.e. H<sub>3</sub>O+ and 196  $H_3O+(H_2O)$ ) to account for variability in instrumental conditions, following equation 5.2. 197 Additional gas analysers are used to monitor NO<sub>x</sub> (APNA-370 Horiba®; measurement 198 range 1 - 1000 ppb; sampling flow 0.8 L min<sup>-1</sup>; response time 120 s) and CO/CO<sub>2</sub> 199 (APEE ProCeas<sup>®</sup>; CO<sub>2</sub> limit of detection 5 ppm; CO limit of detection 10 ppb; sampling 200 flow 0.2 L min<sup>-1</sup>; response time 45 s). 201





## 202 **2.2.2 Aerosol total number concentration and size distribution**

The aerosol total number concentration above 2.5 nm is measured by a Condensation 203 Particle Counter Condensation Particle Counter (TSI<sup>®</sup> model 3075, sampling time 1 s, 204 operated at 1.5 L min<sup>-1</sup>). The aerosol number size distribution is measured by a 205 combination of a Scanning Mobility Particle Sizer (SMPS) consisting of a Differential 206 Mobility Analyser (TSI<sup>®</sup>, model 3080) coupled with a Condensation Particle Counter 207 (TSI<sup>®</sup> model 3072, sampling time 180 s, operated at 0.3/3.0 L min<sup>-1</sup> aerosol flow/sheath 208 flow) measuring particles with mobility diameters between 19.5 and 881.7 nm (107 size 209 210 channels) and an Optical Particle Counter (sky-GRIMM<sup>®</sup> OPC model 1.109, sampling flow=1.2 L min<sup>-1</sup>, laser wavelength = 655 nm, sampling time 12 s) measuring particles 211 with optical mobility diameters between 0.265  $\mu$ m to 31  $\mu$ m (31 size channels). 212

The procedure for combining the aerosol size distributions measured by the SMPS and the sky-GRIMM<sup>®</sup> OPC is based on the method by Baldo et al. (2023), as described in detail in Battaglia et al. (2024). The number size distributions, expressed in d*N*/dlog*D* (cm<sup>-3</sup>), are used to evaluate the total particle surface S ( $\mu$ m<sup>2</sup> cm<sup>-3</sup>) and volume *V* ( $\mu$ m<sup>3</sup> cm<sup>-3</sup>) by assuming spherical particles as

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$$S = \int \pi D^2 \frac{\mathrm{d}N}{\mathrm{dlog}D} \,\mathrm{dlog}D \tag{1}$$

$$V = \int \frac{\pi}{6} D^3 \frac{\mathrm{d}N}{\mathrm{dlog}D} \,\mathrm{dlog}D \tag{2}$$

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## 222 2.2.3 Aerosol chemical composition

The aerosol chemical composition is measured by a combination of online and offline methods.

## 225 2.2.3.1 Time-of-flight Aerosol Chemical Speciation Monitor

A Time-of-flight Aerosol Chemical Speciation Monitor (ToF–ACSM; Aerodyne Research Inc., Billerica, Massachusetts) equipped with a standard vaporiser provides quantitative unitary mass resolution spectra of OA submicronic particles (40 nm–1 μm in vacuum aerodynamic diameter). Particles are sampled with a time resolution of 6 minutes and a flow of 0.85 L min<sup>-1</sup> through a Nafion membrane dryer (model PD-50T-12) installed upstream of the ToF–ACSM. The incoming aerosol is thermally vaporised





at  $\sim$ 600°C. The resulting gas is ionised by electron impact ionization (EI) and the fragments are classified by the time-of-flight mass analyser.

- Data processing (including mass calibration, peaks integration and air beam correction of ion intensities) is conducted with Tofware version 3\_2\_40209, the ACSM data analysis package for the software Igor Pro 7.08 (Wavemetrics, Inc., Portland, OR, USA). The organic mass concentration m<sub>org</sub> is obtained considering a unitary collection efficiency (CE = 1) and a relative ionization efficiency (REI) of 1.4 (Nault et al., 2023). This applies also for the organic linked to mineral dust.
- 240 Given that the glyoxal fragment  $CH_2O^+$  (m/z = 30) has an isobaric interference with the NO<sup>+</sup> fragment from nitrate, the contribution of glyoxal to the organic signal at 30 m/z is 241 estimated with a minor modification of the standard fragmentation table made following 242 the method proposed by Galloway et al. (2009) in their study of glyoxal uptake on 243 244 ammonium sulphate (AS) particles. The contribution to the total signal at 30 m/z from nitrate is imposed to be 1.7 times the intensity of the nitrate signal at 46 m/z, which 245 corresponds to the 30/46 signal ratio measured during nitrate calibration. The 246 contribution to 30 m/z of the organic is then the total signal minus the contribution of 247 248 the nitrate and the contribution of air. The elemental ratios of the organic fraction O/C and H/C are calculated from the measured f44 and f43 respectively, following the 249 parametrizations proposed by Aiken et al. (2008) and Ng et al. (2011), respectively. 250

#### 251 2.2.3.2. Filter sampling

252 Filter samples are collected using a custom-made stainless-steel holder (6 mm diameter to concentrate particles on a small surface) operated at 10 L min<sup>-1</sup> and 253 preceded by an active charcoal denuder to remove ozone and VOCs. The sampling 254 time ranges from 30 minutes to 3 hours. Particles are collected on PTFE filters (Zefluor, 255 47 mm diameter, 2 µm pore size, Pall Life Sciences), and guartz fibre filters 256 (Tissuquartz 2500 QATUP, 47 mm diameter, Pall Life Sciences). Before sampling, the 257 PTFE filters and filter holders are cleaned with dichloromethane (99.8 %, HPLC grade) 258 in an ultrasonic bath. Quartz filters are pyrolyzed at 550°C for approximately 8 hours. 259 260 After sampling, filters are folded and placed in an aluminium paper envelope previously pyrolyzed (same protocol as for filters), and stored in a refrigerator at -18°C. For each 261 experiment, one blank sample is collected by sampling for about 20 min from the 262 chamber only filled with N<sub>2</sub> and O<sub>2</sub>. Analytical blanks, corresponding to pyrolyzed filters 263 264 that had not undergone any sampling, were also collected.





#### 265 2.2.3.3. SFE/GC-MS organic aerosol analysis

Supercritical fluid extraction coupled with gas chromatography mass spectrometry (SFE/GC-MS) is used to analyse the molecular composition of the aerosol organic fraction. It was originally developed by Chiappini et al., 2006 and was slightly modified by a Teledyne ISCO model 260D pump for the extraction and a GC (Clarus 680 PerkinElmer) –MS (Clarus MS SQ8C Perkin Elmer) for the analysis.

The analytical protocol of the SFE/GC-MS analysis begins by placing the quartz filters 271 inside the extraction cell. Prior to the extraction, 5 µL of two different solutions are 272 273 deposited on the filters using a precision syringe (CR700-20 1-20ul (22s/2"/3), Hamilton, USA): (a) an internal standard solution composed by 20 µg mL<sup>-1</sup> of Tridecane 274 (99%, Sigma-Aldrich) and o-Toluic acid (Sigma Aldrich, purity >97 %) in 275 dichloromethane (99.8%, HPLC grade) and (b) a derivatizing agent solution composed 276 by N,O-bis(trimethylsilyl)trifluoroacetamide (BSTFA) and 1% of trimethylchlorosilane 277 278 as catalyst, provided by Sigma-Aldrich. The first step of the analysis is a static extraction, in which the cell is filled with supercritical CO<sub>2</sub> (LINDE, reference CO<sub>2</sub> High 279 Purity)), that interacts with the filter at 300 bar and 60°C, for 40 min. During this step, 280 281 the trimethylsilylation of hydroxy and carboxy functions by BSTFA also occurs (generating trimethylsilyl (TMS) derivatives). The supercritical fluid containing the 282 analytes is transferred to the GC injector through a deactivated silica transfer line. The 283 injector is cooled at -20°C using liquid nitrogen flowing around the injector for 15 284 minutes, where the compounds are retained, and the gaseous CO2 is removed. Once 285 the extraction step was completed, the chemical analysis was continued with the 286 injection of the condensed compounds by heat flash on the GC injector from -20°C to 287 280°C. The compounds are then eluted with helium flowing at 1 mL min<sup>-1</sup> (Linde) and 288 transferred to the GC (Clarus 680 PerkinElmer) for separation. The temperature 289 gradient of the GC column (Rxi®-5Sil MS column (30 m, 0,25mmi.d., film thickness: 290 0.25 µm, Restek) goes from 60°C to 280°C at a rate of 5°C min<sup>-1</sup> and held at 280°C for 291 292 10 min. Detection is achieved through electron impact (70 eV electron energy) ionisation followed by a quadrupole mass spectrometer (Clarus MS SQ8C Perkin 293 Elmer) analysis that produces mass spectra from m/z 50 to m/z 300. 294

The data analysis is conducted using the proprietary software (TurboMass Version 6.1.0.1965 PerkinElmer<sup>®</sup>). The analysis is limited to the chromatographic peaks which elutes before the internal standards (around 42 minutes), as for higher retention times





298 the signal to noise ratio is lower and not conclusive. The chromatograms of each filter sample are compared with those of the analytical and procedural blanks and a mass 299 spectrum is extracted from each chromatographic peak that is not present in the blank. 300 To account for the method variability in extraction efficiency concentrations are 301 corrected using the internal standard o-Toluic acid TMS derivatives. The structural 302 analysis of the molecule generating every chromatographic peak is then carried out 303 using two methods. Each mass spectrum is compared with the reference spectra of 304 the National Institute of Standards and Technology (NIST) Mass Spectral Library 305 (Version 2.2), which assigns a structure to each spectrum with a relative probability. 306 For spectra for which the automatic structural assignment fails (low assignment 307 probability), we searched for target mass fragments derived from molecules linked to 308 glyoxal reactivity (see Table S1 in the Supplementary Material), in particular the 73 m/z 309 fragment corresponding to a TMS derivatization [Si(CH<sub>3</sub>)<sub>3</sub>]<sup>+</sup>, the 147 m/z fragment 310 corresponding to two TMS derivatizations [(CH<sub>3</sub>)<sub>2</sub>Si=OSi(CH<sub>3</sub>)<sub>3</sub>]<sup>+</sup>, the 131 m/z fragment 311 (Glyoxylic acid TMS derivatized – CH<sub>3</sub>), and 205 m/z fragment (Glyoxal monohydrate 312 313 - CH<sub>3</sub>).

#### 314 2.2.3.4. Electrospray ionization (ESI) high-resolution mass spectrometry

Molecular analysis of the organic fraction collected on the quartz filters is performed 315 by electrospray ionization (ESI) high-resolution mass spectrometry (Kourtchev et al., 316 2015). A high-resolution (mass resolution=100000 at m/z 400) LTQ Orbitrap Velos 317 mass spectrometer (Thermo Fisher, Bremen, Germany) equipped with a TriVersa 318 Nanomate robotic nanoflow chip-based ESI source (Advion Biosciences, Ithaca NY, 319 USA) is used to obtain high resolution mass spectra of the methanol extracts following 320 an adaptation of the procedure described in Kourtchev et al. (2015). Filters are 321 extracted one time in 1 mL of methanol (Optima TM grade, Fisher Scientific) and two 322 323 times in 0.5 mL of methanol under ultrasonic agitation in slurry ice for 15 min. Extracts are combined and filtered sequentially through a 0.45 µm pore size and a 0.2 µm pore 324 325 size Teflon filter (ISODiscTM Supelco), which are then reduced by volume to approximately 50–200 μL under a gentle stream of nitrogen. The resulting sample is 326 injected by direct infusion. The negative ionization mass spectra are collected in three 327 replicates at ranges m/z 50-500 and m/z 150-1000 and processed using Xcalibur 2.1 328 software (Thermo Scientific). In the settings of the data processing, the following atoms 329 are included in the peak assignment: C (from 1 to 100 atoms in the possible assigned 330





molecular formula), H (1-200), O (0-50), N (0-5), S (0-2),  $^{13}$ C (0-1) and  $^{34}$ S (0-1). The allowed mass accuracy in the formula assignment is ± 4 ppm.

The peak assignment to a molecular formula is done according to Zielinski et al. (2018). 333 The protocol includes internal calibration, noise removal, blank subtraction, and 334 additional atomic constraints for formula filtering: elemental ratios were set as 0.3 335 ≤H/C≤ 2.5, O/C≤ 2, N/C≤ 0.5, S/C≤ 0.2, <sup>13</sup>C/<sup>12</sup>C≤ 0.011 and <sup>34</sup>S/<sup>32</sup>S≤ 0.045, and 336 nitrogen rule. In the case of multiple assignments for the same peak, the formula with 337 the lowest mass error was kept. This process allowed for the retrieval of parameters 338 339 describing carbon oxidation, such as O/C and H/C ratios. Consequently, each mass spectrum was analysed to construct a van Krevelen diagram, which is a graphical 340 representation illustrating the sample composition in terms of carbon, oxygen, and 341 hydrogen in the identified molecular formulas (Patriarca et al., 2018). Identified 342 molecular formulas are categorised into the following groups: CHO, CHON, CHOS, 343 CHNS, and CHONS. 344

A targeted search for molecular formulas resulting from glyoxal reactivity is also done. 345 In this search, we included formulas associated with glyoxal chemical transformations 346 347 such as hydration, oxidation, and oligomerization. Starting from the glyoxal formula, C<sub>2</sub>H<sub>2</sub>O<sub>2</sub>, formulas for mono- and dihydration products are C<sub>2</sub>H<sub>4</sub>O<sub>3</sub> and C<sub>2</sub>H<sub>6</sub>O<sub>4</sub>, 348 respectively. Oxidation products included formic acid ( $CH_2O_2$ ), glycolic acid ( $C_2H_4O_3$ ), 349 glyoxylic acid ( $C_2H_2O_3$ ), and oxalic acid ( $C_2H_2O_2$ ). Oligomers formed by the hydrolysis 350 of hydrated glyoxal formulas were also sought. In the process of hydrolysis-driven 351 oligomerization, each successive molecular addition results in the loss of a water 352 molecule. If the oligomer is a ring, an additional water molecule is lost due to the 353 condensation of the linear oligomer terminations. Denoting n as the number of 354 molecules of the monohydrated form  $(C_2H_4O_3)$  and m as the number of molecules of 355 the dihydrated glyoxal form ( $C_2H_6O_4$ ) participating in the formation of an oligomer, the 356 generated linear oligomers will have the formula C<sub>2n+2m</sub>H<sub>4n+6m-2(n+m-1)</sub>O<sub>3n+4m-(n+m-1)</sub>, 357 358 where the terms -2(n+m-1) and -(n+m-1) in the hydrogen and oxygen atom stoichiometry indicate water loss from the oligomerization process. For ring oligomers, 359 formulas characterised by the stoichiometry C<sub>2n+2m</sub>H<sub>4n+6m-2(n+m)</sub>O<sub>3n+4m-(n+m)</sub> are 360 searched. Similarly, formulas resulting from the condensation of hydrated forms with 361 362 the listed organic acids are calculated and researched.

#### 363 2.2.3.5. X-ray photoelectron spectrometry (XPS)





364 X-ray photoelectron spectrometry (XPS) is used to quantify the elemental O/C ratio of the particle surface to a depth less than 10 nm. Measurements are performed with a 365 VG ES-CALAB 250 instrument using monochromatic AI K<sub> $\alpha$ </sub> radiation (1486.6 eV). The 366 O/C ratio is quantified by integrating the areas of  $O_{1s}$  and  $C_{1s}$  peaks. This last is 367 contributed by a number of functions, including -CO<sub>2</sub>, C-O, C-C/C-H as well as C-F 368 from the Teflon substrate. The contribution of the latter can be evaluated from the F1s 369 (approximately 690 eV) as described in Denjean et al., (2015). The contribution of SiO<sub>2</sub> 370 from the mineralogical composition of the dust to O1s was evaluated by integrating the 371 Si2p peak (107 eV) and applying the stoichiometric proportions between silicon and 372 oxygen in the composition of quartz (O/Si = 2). The O/C<sub>surf</sub> ratio is calculated as follows 373

374 375

$$O/C_{surf} = \frac{n[O_c]}{n[C]} = \frac{(n[O] - 2n[Si])}{n[C]}$$
(3)

376

Where  $n[O_c]$  is the signal coming from oxygen bonded to carbon atoms only, n[C] is the signal of  $C_{1s}$ , n[O] the area from  $O_{1s}$  and n[Si] is the area of the signal from  $Si_{2p}$ that is multiplied by 2 to take into account the silica stoichiometry. The XPS measurement on a filter collected during one ageing experiment are shown as an example in Figure S1 in the Supplementary Material.

## 382 2.3 Calculation of the glyoxal uptake coefficient

The uptake coefficient  $\gamma$  is defined as the probability of the gas to be taken up on the aerosol surface. It is a unit-less parameter expressed by the ratio between the number of molecules taken up on a surface and the total number of collisions of the gas on the surface as

- 387
- 388

 $\gamma = \frac{number \ of \ total \ molecules \ taken \ up}{total \ number \ of \ collisions} \tag{4}$ 

389

The uptake coefficient  $\gamma$  can be calculated in two ways. First of all, it can be estimated from the first-order heterogeneous loss rate of glyoxal ( $k_{het}$ , s<sup>-1</sup>) as

$$\gamma = \frac{k_{het}}{\omega}$$
(5)





394 where  $\omega$  is the rate of collisions (collision frequency) defined as 395 396  $\omega = \frac{cA_s}{4}$ (6) 397 where: 398 •  $c = 146 \times \sqrt{\frac{T}{MW}}$  is the mean molecular speed (m s<sup>-1</sup>), where T is the air temperature 399 (here 298 K) and MW the molecular weight of the compound of interest (in the case 400 of glyoxal MW = 58 g mol<sup>-1</sup>). 401 • A<sub>s</sub> is total aerosol surface concentration (m<sup>2</sup> m<sup>-3</sup>). 402 The total aerosol surface concentration (As) is calculated from the aerosol size 403 distribution recorded at the end of the dust injection. 404 The heterogeneous loss rate of glyoxal  $(k_{het})$  due to its uptake on dust particles can be 405 determined as the difference of the loss rate of glyoxal measured during the uptake 406 experiments  $(k_{obs})$  and the glyoxal loss rate on the chamber walls  $(k_{loss})$  as 407 408 409  $k_{het} = k_{obs} - k_{loss}$ (7)410 The glyoxal wall loss is represented by a partition equilibrium described by two first-411 order reactions: one for the adsorption of gas phase molecules onto the chamber walls, 412 and one for the reverse process. The rate constants for both processes have been 413 obtained experimentally through control experiments with only glyoxal in the chamber 414 in different relative humidity conditions. 415 If the uptake reaction is of the first rate,  $k_{het}$  is henceforth calculated as 416  $k_{het} = \frac{ln\left(\frac{[Gly]_0}{[Gly]_{obs}}\right) - ln\left(\frac{[Gly]_0}{[Gly]_{loss}}\right)}{t}$ (8) 417 418

where [Gly]<sub>0</sub> is the initial concentration of glyoxal, [Gly]<sub>obs</sub> represents the observed
evolution of glyoxal concentration in time, resulting from the sum of uptake and wall





loss, and [Gly]<sub>loss</sub> represents the estimated glyoxal concentration resulting from the wall
loss.

423 In addition,  $k_{\text{F-OM}}$ , the rate of formation of the particulate organic matter (POM) on pre-

424 existing particles following the uptake of glyoxal on the dust can be calculated as

425

$$k_{F-OM} = \frac{ln\left(\frac{[POM]_t}{[POM]_0}\right)}{t} \tag{9}$$

427

426

where [POM]<sub>0</sub> represents the initial POM concentration in the particle phase, and
[POM]<sub>t</sub> represents the concentration of the POM formed at a given time.

430 If the hypothesis that the POM formation is solely due to the uptake of glyoxal,  $\gamma$  can 431 be also evaluated as

.

 $\gamma = \frac{k_{F-POM}}{\omega} \tag{10}$ 

433

432

# 434 **3 Results and discussion**

# 435 3.1 Timeline of experiments

The ageing experiments of monodispersed mineral dust and glyoxal described in this paper are summarised in Table 1. All the aerosol data are corrected for dilution, wall loss, and particle loss through the tubing systems as detailed in Text S2 in the Supplementary Material. Gas phase concentrations are corrected for dilution only.

**Table 1.** Listing and initial conditions of the experiments considered in this study, including experiments with glyoxal only (experiment type GL), ammonium sulphate and glyoxal (AS + GL), dust only (D) and dust with glyoxal (D + GL). The glyoxal and ozone gas phase concentrations correspond to the maximum value measured by FTIR after the respective injections. V<sub>seed</sub> indicates the maximum volume concentrations of seed particles (either dust or ammonium sulphate) measured after the particle injection. The notation "dark/light" indicates experiments when filter samples were collected both in the dark and with irradiation.

Experiment type	Reagents	Date	Experiment number	RH, %	Light	[O₃], ppb <sub>v</sub>	Temp, K	[GL], ppbv	V <sub>seed</sub> , µm³ cm⁻³
Control	GL AS+GL	29/04/2021	G1	< 5	dark		292	1130	
		11/02/2022	G <sub>2</sub>	77	light	1440	291	627	
		21/02/2023	AS <sub>1</sub>	38	dark		298	527	50.1
		23/02/2023a	AS <sub>2</sub>	35	dark		298	516	48.3





		23/02/2023b	AS <sub>3</sub>	32	light		298	445	64.8
		07/09/2023	AS <sub>4</sub>	81	light		301	779	304.1
		08/09/2023	AS <sub>5</sub>	83	light		300	430	161.2
		31/01/2022	D1	< 5	<sup>dark</sup> /light		292		31.5
	D	03/02/2022	D <sub>2</sub>	75%	<sup>dark</sup> /light		293		55.4
		04/02/2022	D <sub>3</sub>	< 5	<sup>dark</sup> /light		293	690	35.6
		08/02/2023	D4	32	dark		294	940	21.5
		09/02/2023	D5	31	light		295	1050	52.7
	D+GL	10/02/2023	D <sub>6</sub>	35	dark		294	809	37.4
		13/02/2023	D7	34	light		296	850	51.3
		30/04/2021	D <sub>8</sub>	76	light		289	759	28.3
		03/05/2021	D9	79	light		290	607	38.7
Uptake		04/05/2021	D <sub>10</sub>	81	light		290	371	31.5
		05/05/2021	D11	78	dark		291	805	30.1
		06/05/2021	D <sub>12</sub>	82	dark		292	432	21.1
		08/02/2022	D <sub>13</sub>	81	<sup>dark</sup> /light	1270	293	555	64.0
		09/02/2022	D14	78	<sup>dark</sup> /light	1450	293	756	79.8
		10/02/2022	D <sub>15</sub>	75	dark/light		295	600	68.4
		14/02/2023	D <sub>16</sub>	83	dark		296	661	35.8
		15/02/2023	D <sub>17</sub>	75	light		298	444	41.0

448

Table 1 also lists the few control experiments using ammonium sulphate as seed particles, described in detail in Text S3 in the Supplementary Material. No POM formation is observed during control experiments with dust or glyoxal only, both dry and humid conditions and with and without irradiation.

The typical timelines of the particle concentrations (number and volume) and the nonrefractory composition measured in dry conditions and at 30% and 80% relative humidity are shown in Figure 1.









458 Figure 1. Timeline of ageing experiments of submicron dust with gas-phase glyoxal in dry conditions (top, experiment D<sub>3</sub>), 30% (middle, experiment D<sub>7</sub>), and 80% RH (bottom, experiment D<sub>15</sub>). Left (a): 459 460 aerosol total number (N) and volume (V) concentrations (blue and black lines, respectively) calculated 461 from the measured dust size distributions. Right (b): mass concentrations of ammonium, sulphate, 462 nitrate and organic (yellow, red, blue and green lines, respectively) measured by the ACSM. The yellow-463 highlighted portion of the graph indicates the interval where irradiation takes place, while the green 464 vertical dashed lines indicate the injection glyoxal in the chamber. The dust injection corresponds to the 465 time of the initial increase of the number and volume concentrations. Aerosol time series are corrected for dilution, wall loss and particle loss through the tubings. 466

467

Figure 1 shows that in dry conditions, there is no significant variation of either the aerosol number or the volume concentrations, nor the chemical composition (including organics) following the glyoxal injection.

471 At 30% RH, a small increase in the total volume concentration (approximately 5  $\mu m^3$ 

472 cm<sup>-3</sup>) is observed for about 30 minutes after the injection of glyoxal. This corresponds

to an increase of the POM of about 1  $\mu$ g m<sup>-3</sup>, approximately 20% more with respect to





the value measured before the uptake. On the other hand, the particle number concentration shows an apparent decrease at the beginning of the experiments, possibly because the particle loss correction model of Lai and Nazaroff (2000) does not fully apply to dust particles and humid conditions (see discussion in Battaglia et al. (2024). After that, and through the duration of the experiment, however, it remains constant, indicating that the increase in the particle volume occurs on the dust particles and not because of new particle formation.

At 80% RH, the increase in both the total volume and the POM concentrations is more 481 pronounced, approximately 10-15 µm<sup>3</sup> cm<sup>-3</sup> and 2 µg m<sup>-3</sup>, respectively. As for 30% RH, 482 both total particle volume and the POM concentrations return to values observed prior 483 the injection of glyoxal, within approximately 30 minutes from their maximum values, 484 likely due to evaporation. A similar behaviour is observed in the presence of ozone (not 485 shown). As for 30% RH, the particle number concentration slightly decreases in time 486 at the beginning of the experiment, but then remains constant, again excluding the 487 formation of new particles but rather confirming the formation of organic matter on pre-488 existing particles. This is also supported by the fact that the rate of increase of POM 489 and particle volume is the same (slope 3.2 10<sup>-4</sup> s<sup>-1</sup> and 3.2 10<sup>-4</sup> s<sup>-1</sup>, respectively for 490 POM and total volume), as shown by Figure S2 in the Supplementary Material. 491

On the other hand, the ratio between the observed increase of the POM (2  $\mu$ g m<sup>-3</sup>) and 492 that of the particle volume concentration (20  $\mu$ m<sup>3</sup> cm<sup>-3</sup>) corresponds to an estimated 493 mass density of the order of 0.1 g cm<sup>-3</sup>, that is, about 10 times lower than the value of 494 1 g cm<sup>-3</sup> expected for glyoxal. This would suggest that part of the organic matter formed 495 on dust is not detected by the ACSM, as will be further demonstrated in section 3.3. 496 On the other hand, Figure S2 shows that, after reaching its maximum value, the volume 497 concentration decreases at a lower rate than the POM (slope 3.9 10<sup>-5</sup> s<sup>-1</sup> and 6.1 10<sup>-5</sup> 498 s<sup>-1</sup>, respectively). This suggests that an additional process could contribute the particle 499 volume concentration partially compensating the loss of organic matter on the dust 500 501 particles.

This is confirmed by Figure 2 comparing the variation in time of the particle volume distributions, normalised to the total volume, at four different times of the experiments at 30% and the one at 80%: prior and when the glyoxal is injected, when the POM reaches its peak value, and at the end of the experiment.









507

509 Figure 2. Evolution of volume-size distributions for two glyoxal uptake experiments in different relative 510 humidity conditions. The images illustrate the progression of volume-size distributions recorded at four key moments during the experiments. The first distribution (orange) is recorded after the dust is injected 511 512 into the simulation chamber. The second distribution (vellow) is recorded at the moment of glyoxal injection. The third distribution (blue) corresponds to the peak uptake of glyoxal on the aerosol, and the 513 fourth (light blue) is recorded at the end of glyoxal uptake process. The left image depicts the evolution 514 for the experiment D7 conducted at 30% RH, while the image on the right shows the distributions for the 515 experiment D15 conducted at 80% RH. The results highlight that the distributions grow more significantly 516 517 at 80% RH, indicating a higher glyoxal uptake and organic formation at elevated humidity levels. 518

519 All distributions have a single mode. However, after the injection of glyoxal, the 520 geometric mean volume diameter, measured at the maximum POM concentration,





- 521 increases by up to 10% (from 310 to 340 nm) at 30% RH, and up to 20% (from 450 to 540 nm) at 80% RH. Interestingly, even at the end of the experiment, when the POM 522 concentration returns to its initial value, the increase in geometric mean diameter of 523 the aerosol is irreversible. This effect could be explained by the hypothesis that the 524 uptake of glyoxal enhances the dust hygroscopicity. After glyoxal uptake, the particle 525 becomes more hygroscopic and the difference in total volume between the beginning 526 and end of the experiment is due to water uptake which adds up to the formed organic 527 aerosol mass. 528
- 529 Finally, Figure 1 shows also that, while sulphate and ammonium are never detected, a background concentration of nitrate up to 1 µg m<sup>-3</sup> is measured by the ACSM as soon 530 as the dust particles are injected in the presence of water. We attribute it to the 531 heterogeneous interaction between NO<sub>2</sub> and the dust particles (Goodman et al., 1999) 532 533 as indeed, a background concentration of a few ppb of NO2 is present in the chamber as a result of the procedure used to reduce the TOC content in the injected water (see 534 Figure S3 in the Supplementary Material). However, since the contribution of nitrate 535 represents at maximum 1% of the injected dust mass and whether decreases or 536 remains constant throughout the experiment, its contribution to the particle growth and 537 538 overall ageing of the mineral dust should be negligible.
- Figure 3 shows the time series of the gas-phase compounds detected during the same
  experiment (D<sub>15</sub>) at 80% RH.







Figure 3. Time series of the gas-phase composition observed during experiment D<sub>15</sub>: (a) concentrations 542 543 of carbon monoxide, glyoxal and formic acid measured by FTIR (for CO the measured of the online analyser are also shown); (b) various VOC ions (m/z 47, 59, 43, 46 and 61) measured by the PTR-MS. 544 545 Ion signals measured by PTR-MS are normalized by signals of reagent ions (i.e.  $H_3O^+$  and  $H_3O^+(H_2O)$ ) 546 and therefore expressed in normalized counts (ncps). The blue vertical dashed lines indicate the 547 injection of water in the chamber; the green vertical dashed lines indicate the injection of glyoxal in the 548 chamber, the yellow dashed lines indicate the beginning of irradiation and the grey dashed lines indicate 549 the end of irradiation.

550

The measured glyoxal concentration after the injection (Figure 3a) is lower than the 551 nominal concentration of 1 ppm and goes to zero within minutes due to the rapid 552 interactions with the walls of the chamber, water vapour, and the dust particles. Upon 553 irradiation, formic acid and carbon monoxide are formed, as expected by the photolysis 554 of glyoxal (De Haan et al., 2020). Fragments m/z = 46 and 47 are observed during 555 556 water injection and photolysis, and could originate from the deprotonated and protonated form of formic acid, respectively. This suggests that a minor fraction of the 557 formic acid could result from the desorption of compounds (including glyoxal) from the 558 chamber walls. Fragments m/z = 43 and m/z = 61, and occasionally m/z = 45 (not seen 559 560 during experiment D<sub>15</sub> and therefore not shown in Figure 3b), are observed at a normalised intensity two orders of magnitude lower than that of formic acid, but not 561 attributed. The quantification with both PTR-MS and FTIR in our experimental RH 562 conditions is complicated by the presence of water, which reduces the sensitivity of 563 PTR-MS and can interfere with the absorption of various organic compounds, making 564 their quantification less accurate. 565





#### 566 3.2 Glyoxal uptake coefficient

For experiment D<sub>15</sub> at 80% RH, Figure 4 shows the temporal evolution of the natural logarithm of the glyoxal concentration measured by the FTIR, compared to that measured during a typical blank experiment without dust particles. Figure 4 additionally shows the variation of the aerosol organic fraction measured by the ACSM during the same time period.



573

Figure 4. Measurement of glyoxal uptake coefficient on dust for the experiment D<sub>15</sub>, conducted at 80%
 RH. The figure compares the two methods for measuring the uptake coefficient. In the top image, results
 are shown for the method based on monitoring the decay of gas-phase glyoxal. The red and blue curves





577 represent the logarithm of the ratio between the calculated decay of gas-phase glyoxal in the absence 578 and presence of dust aerosols, respectively. The black lines represent the linear fit whose slope provides 579 the heterogeneous kinetic constants of the two processes. The image at the bottom displays the result 580 of the uptake coefficient measurement for the same experiment, obtained from the organic formation on 581 the aerosol monitored by ACSM. The blue time series shows the logarithm of the ratio between the 582 measured organic concentration divided by the initial organic on dust aerosol, while the black line is the 583 linear fit representing the kinetics of organic formation.

584

585 Within the first 10 minutes after the injection of glyoxal, the decrease of the natural 586 logarithm concentrations ratio with time in the presence of dust is linear (the rate is constant). After that, the loss slightly deviates from linearity. The difference from 587 588 linearity is more evident for the blank experiment, when it occurs earlier than when the dust is present. These observations indicate that, within the first 10 minutes, the uptake 589 590 of glyoxal on the dust particles can be considered to follow a first-order kinetic and its rate represents an initial uptake coefficient. In the following 20 minutes approximately, 591 592 the uptake slows down, possibly because all the sites available on the particle surface become occupied, but also because that desorption from the particle surface could 593 594 reinject glyoxal in the reactive mixture. On the particle phase, the natural logarithm of the organic concentration, normalised by its initial value increases rapidly and linearly, 595 almost on the same time scale of that of the loss of glyoxal, but then decreases to 596 return to its initial value within approximately one hour. These observations confirm 597 that the uptake of glyoxal results in a formation of OA on the dust particles, but that 598 this process is reversible. 599

- The uptake coefficients calculated as the linear fit of the glyoxal and particle organic concentration are presented in Table 2.
- 602

603 Table 2. Uptake coefficients for glyoxal on mineral dust and ammonium sulphate calculated from the 604 loss of gas-phase glyoxal and the rate of OA formation for the experiments conducted at 80% RH. The 605 initial glyoxal concentration is reported. The aerosol surface concentration (A<sub>s</sub>) corresponds to the value 606 preceding the glyoxal injection. Ozone concentration is the maximum concentration measured by FTIR 607 spectroscopy after the injection. For ammonium sulphate, only the γ values calculated from the loss of 608 gas phase glyoxal are presented, as the ACSM collection efficiency (CE) for ammonium sulphate varies 609 significantly during OA formation (Matthew et al., 2008).

Date	Experiment ID	RH%	[GL]₀, ppb <sub>v</sub>	Ozone (ppb)	As (m <sup>2</sup> m <sup>-3</sup> )	ω (s <sup>-1</sup> )	γgas	<b>Υ</b> ΟΑ glyxoal
30/04/2021	D <sub>8</sub>	76	759		4.8 10 <sup>-4</sup>	3.9 10 <sup>-2</sup>	6.0 10 <sup>-3</sup>	1.0 10 <sup>-3</sup>
03/05/2021	D9	79	607		6.1 10 <sup>-4</sup>	5.0 10 <sup>-2</sup>	1.5 10 <sup>-2</sup>	1.5 10 <sup>-2</sup>
04/05/2021	D <sub>10</sub>	81	371		5.1 10 <sup>-4</sup>	4.2 10 <sup>-2</sup>	1.7 10 <sup>-2</sup>	9.0 10 <sup>-3</sup>
05/05/2021	D11	78	805		4.6 10 <sup>-4</sup>	3.8 10 <sup>-2</sup>	8.0 10 <sup>-3</sup>	5.0 10 <sup>-3</sup>
06/05/2021	D <sub>12</sub>	82	432		3.5 10-4	2.9 10 <sup>-2</sup>	1.2 10 <sup>-2</sup>	2.3 10 <sup>-2</sup>
08/02/2022	D13	81	555	1270	7.1 10 <sup>-4</sup>	5.8 10 <sup>-2</sup>	4.0 10 <sup>-3</sup>	4.0 10 <sup>-3</sup>
09/02/2022	D <sub>14</sub>	78	756	1450	8.5 10 <sup>-4</sup>	7.0 10 <sup>-2</sup>	4.0 10 <sup>-3</sup>	5.0 10 <sup>-3</sup>
10/02/2022	D <sub>15</sub>	75	600		8.4 10 <sup>-4</sup>	6.9 10 <sup>-2</sup>	1.0 10 <sup>-2</sup>	5.0 10 <sup>-3</sup>





14/02/2023	D <sub>16</sub>	83 81	661	 6.0 10 <sup>-4</sup>	4.9 10 <sup>-2</sup>	4.0 10 <sup>-3</sup>	1.5 10 <sup>-2</sup>
08/09/2023	AS <sub>4</sub> AS <sub>5</sub>	83	430	 2.0 10 <sup>-3</sup>	1.7 10 <sup>-1</sup>	1.2 10 <sup>-3</sup>	
Average dust						9 (± 5) 10 <sup>-3</sup>	9 (± 7) 10 <sup>-3</sup>
Average AS						1.1 (± 0.2) 10 <sup>-4</sup>	

610

611 The average uptake coefficients for glyoxal on the Gobi mineral dust calculated at 80% RH from the gas-phase uptake and the particle formation are  $\gamma_{Gly-Dust-gas} = 9 \times 10^{-3}$ 612 (standard deviation  $\pm$  5) and y<sub>Gly-Dust-OA</sub> = 9 × 10<sup>-3</sup> (standard deviation  $\pm$  7), respectively. 613 The two average values agree. This suggests that every glyoxal molecule in the gas 614 phase is taken up by the airborne dust particles. This suggests also that the uptake 615 occurs on airborne particles only, as expected as the dust particles are selected in the 616 submicron range and that minimal deposition of dust particles is observed in the first 617 30 minutes after injection. The primary mechanism of particle loss during this period is 618 dilution, which does not interfere with uptake. The standard deviations of the mean 619 620 values are large, being attributed to the fact that the state of the chamber walls and the dust size distribution vary from one experiment to the other, and that the 621 622 aerosol/chamber walls surface ratio is very low  $(0.08-1.5 \ 10^{-3})$ . The presence of ozone appears uninfluential. 623

The results of the current study can be compared with the literature. Shen et al. (2016) 624 investigated the uptake of glyoxal on mineral proxies, i.e. SO<sub>2</sub> and CaCO<sub>3</sub> under 625 various levels of relative humidity. These authors determined the uptake coefficients 626 after a long exposition of the surface to glyoxal (steady state uptakes) and found that 627 628 the uptake coefficients are reduced with increasing the gas phase concentration of glyoxal. At 1 ppb concentration and a relative humidity of 60% the uptake coefficients 629 determined on suspended particles of calcite (CaCO<sub>3</sub>,  $\gamma$  = (1.4 ± 0.1) × 10<sup>-4</sup>) and 630 alumina (AIO<sub>3</sub>,  $\gamma = (5.5 \pm 0.1) \times 10^{-5}$ ). Our values are measured in a shorter time frame 631 and correspond to an initial uptake of glyoxal. They are approximately one order of 632 633 magnitude higher than those obtained by Shen et al. (2016). These authors scaled the uptake coefficient to the specific surface area of the dust, which henceforth could 634 635 correspond to a lower limit. On the contrary, in our case, we use a geometric surface area (assuming spherical particles) which could lead to an overestimation of the uptake 636 coefficient. 637





Zogka et al. (2024) used a Knudsen cell to evaluate the initial and steady state glyoxal uptake coefficient bulk soil samples of various origins. At low relative humidity, these authors found that for Gobi soil sieved to less than 63 μm in diameter the initial uptake coefficient using the geometric surface area was 0.18 (corresponding to an upper limit of the uptake) independent of glyoxal concentration. However, the steady state uptake coefficients determined after a long processing of surface were found to decrease with increasing glyoxal concentration, due to aging of the surface.

Various reasons can explain these apparently different results. First of all, in CESAM we measure the initial uptake coefficient at humid conditions, which is independent of concentration. On the other hand, Shen et al. (2016) measured steady-state uptake coefficient at lower glyoxal concentrations (< 1 ppb) than we did (>400 ppb). As shown by both Shen et al. (2016) and Zogka et al. (2024), the steady-state uptake coefficient decreases with the concentration of glyoxal, regardless of the relative humidity.

Secondly, the uptake coefficient is inversely proportional to the available particle surface, which in our case is smaller than Zogka et al. (2024) who used soils sieved to 63 µm. Shen et al (2016) used standard mineral particles of various sizes from 35 nm to 5 µm, while our particle size distribution peaked between 300 and 400 nm. Shen et al. (2016) used single minerals, while Zogka et al. (2024) and our study share the same soil sample from Gobi. While the uptake coefficient should depend on the dust mineralogy, this is difficult to ascertain in the present study.

Overall, although the experiments performed in literature with those of the current study were performed under different conditions, results indicate that natural Gobi dust is an effective sink of glyoxal, with initial uptake coefficient independent of glyoxal concentration, pointing to a first order removal process. However, the long-term ageing of particles leads to lower uptake coefficients that strongly depend on glyoxal concentration due to the depletion of surface sites.

#### 664 3.3. Organic composition

This section discusses the chemical composition of the organic matter on mineral dust following the interaction with glyoxal. Details on the organic composition of the native dust are provided in Text S4 in the supplementary material.





## 668 **3.3.1. Measurements by the ACSM**

- 669 Figure 5 shows the time evolution of the intensity of the organic fragments detected by
- the ACSM at 80% RH (experiment D<sub>15</sub>) before the injection of glyoxal (dust only),
- during the POM formation, after the maximum concentration, and at the end of the
- experiment, when it returned to its initial value.
- 673







675 Figure 5. ACSM organic mass spectra (intensities normalized to the total organic concentration) 676 recorded during the experiment D<sub>15</sub>: (a) before glyoxal uptake (dust organic fraction composition), (b) 677 during glyoxal uptake, (c) after reaching the maximum uptake on the particles and (d) 1h later under 678 irradiation. Panels on the left show the mass spectra ranging from 10 to 60 m/z, while the panels on the 679 right represent fragments from 60 to 200 m/z (their intensity is approximately one hundred times lower). 680 The inserts display the time series of organic concentrations measured by the ACSM. A black arrow 681 indicates the time corresponding to the mass spectrum shown. The yellow-highlighted shaded area 682 indicates the interval where irradiation takes place, while the green vertical dashed lines indicate the 683 moment of glyoxal injection in the chamber.





684

Fragments at 28 m/z and 44 m/z, typical of oxidised compounds, are ubiquitous at all 685 stages of the experiment. Their relative intensity follows the kinetic of the uptake, 686 decreasing during the formation of organic aerosols and reverting to their initial value 687 towards the end of the experiments. Fragment 69 m/z, attributed to C3HO2+ and 688  $C_4H_5O^+$ , has the highest intensity among ions above m/z 60 in all four spectra, 689 suggesting that it is not related to glyoxal reactivity. Galloway et al. (2009) observed 690 this fragment and identified it as a nitrogen-containing organic molecule with a formula 691 692  $C_3H_3NO^+$  from the reaction of glyoxal with ammonia during the uptake of glyoxal on ammonium sulphate. This fragment was also observed at significant intensities in 693 694 organic aerosols originating from phenolic derivatives, such as guaiacol and syringic acid, metabolites of plants (Sun et al., 2010), and could be explained as soil residues 695 from vegetation but also animals (Nieder et al., 2018). 696

The signal of fragments at 29 m/z (CHO<sup>+</sup>), 30 m/z (CH<sub>2</sub>O<sup>+</sup>), and 58 m/z (molecular 697 peak  $C_2H_2O_2^+$ ), characteristics of glyoxal, appear during the uptake (second panel from 698 699 the top in Figure 5) but their relative intensity decreases with time. Fragments m/z =700 105 and m/z = 131 observed during both the uptake and the photolysis are specific markers of the interaction between dust and glyoxal (Liggio et al., 2005). The former is 701 702 attributed to an oligomeric structure of glyoxal generated by the condensation of two 703 molecules of glyoxal hydrate (see also Table 3 in Liggio et al. (2005)). The latter is not identified, yet it consistently accompanies the glyoxal's primary ions. In photo-oxidation 704 experiments of glyoxal in the aqueous phase, Carlton et al. (2007) found that its 705 abundance increased proportionally with the increase in glyoxal concentration. In our 706 study, this fragment is detected in all conditions (with and without light or ozone). 707

Fragments at 119 and 120 m/z have been observed in organic aerosols derived from isoprene and attributed to an organic acid with the formula  $C_8H_8O_4$  (m/z 120) and its deprotonated form (Safi Shalamzari et al., 2013). In our experiments, they consistently increase after glyoxal uptake, and in particular upon irradiation. Although it is not straightforward to assign these fragments to a unique glyoxal derived formula, we hypothesise that they may arise from oxidised forms of oligomers.

Upon irradiation, m/z = 147 is accompanied by m/z = 165. These fragments have been observed in organic aerosols produced in experiments of aqueous phase glyoxal oxidation via OH radicals (Lim et al., 2010). The fragment m/z 165 is attributed to the





- condensation of one molecule of glyoxal di-hydrate with one molecule of oxalic acid or two molecules of glyoxylic acid hydrate. The fragment m/z 147 could result from the dehydration of the aforementioned products. These fragments are present with similar intensity in the spectrum of native dust, and their increase in intensity under irradiated conditions might therefore be due to the reversibility of the interaction rather than an oxidation process of glyoxal.
- The occurrence of the fragment at 18 m/z is often resulting from the loss of a water molecule ( $H_2O$ ) from hydrated organic compounds. The slight decrease of the intensity of this fragment during glyoxal uptake could therefore be explained by the presence of the oligomerization of hydrated glyoxal molecules. This process leads to the loss of two hydroxide groups for each added molecule in favour of the formation of acetal or hemiacetal bonds in the structure of the resulting newly formed secondary organic aerosol.

## 730 3.3.2. Molecular analysis

The list and conditions of the samples analysed by SFE/GC-MS and ESI-Orbitrap are
reported in Table S2 in the Supplementary Material. The summary of the organic
molecules detected by those analysis is presented in Table 3 and discussed in the next
paragraphs.

735

736 **Table 3.** Summary of frequently observed compounds identified by SFE/GC-MS analysis and glyoxalrelated formulas observed with ESI-Orbitrap, along with the suggested structures under the different experimental conditions tested.

Molecular formula	Name	Tentative Structure	Technique	Experimental conditions
$C_2H_2O_3$	Glyoxylic acid	HO	ESI-Orbitrap	Dust+Gly, 80%, Dark, O <sub>3</sub>
C <sub>2</sub> H <sub>4</sub> O <sub>3</sub>	Glycolic acid	но он	ESI-Orbitrap SFE/GC-MS	Dust+Gly, 30%, Dark Dust+Gly, 30%, Light Dust+Gly, 80%, Light Dust+Gly, 80%, Light, O <sub>3</sub>
C₂H₂O₄	Oxalic acid	но он	ESI-Orbitrap	Dust+Gly, 80%, Light
C <sub>2</sub> H <sub>4</sub> O <sub>4</sub>	Glyoxylic acid monohydrate	но он	ESI-Orbitrap SFE/GC-MS	Dust+Gly, 80%, Light
C3H603	Lactic acid	HO	SFE/GC-MS	Dust, 80%, Light Dust+Gly, Dry, Dark Dust+Gly, 30%, Dark Dust+Gly, 30%, Light Dust+Gly, 80%, Light Dust+Gly, 80%, Dark, O <sub>3</sub> Dust+Gly, 80%, Light, O <sub>3</sub>





Levulinic acid	но	SFE/GC-MS	Dust, Dry, Dark Dust, 80%, Dark Dust, Dry, Light Dust+Gly, 30%, Dark Dust+Gly, 30%, Light Dust+Gly, 80%, Dark, O <sub>3</sub> Dust+Gly, 80%, Light, O <sub>3</sub>
Benzyl alcohol	04	SFE/GC-MS	Dust, 80%, Dark Dust+Gly, 30%, Dark Dust+Gly, Dry, Light Dust+Gly, 30%, Light Dust+Gly, 80%, Light
Cyclohexanone, 3,3,5-trimethyl	×,	SFE/GC-MS	Dust, 80%, Dark Dust+Gly, 30%, Dark Dust+Gly, 30%, Light Dust+Gly, 80%, Light, O <sub>3</sub>
Glycolic acid dimer	° Color	ESI-Orbitrap	Dust+Gly, 80%, Light
Glyoxal oligomer	о он	ESI-Orbitrap	Dust+Gly, 80%, Light
Glyoxal oligomer		ESI-Orbitrap	Dust+Gly, 80%, Light
Glyoxal oligomer		ESI-Orbitrap	Dust+Gly, 80%, Light
Glyoxal oligomer		ESI-Orbitrap	Dust+Gly, 80%, Dark
	Levulinic acid Benzyl alcohol Cyclohexanone, 3,3,5-trimethyl Glycolic acid dimer Glyoxal oligomer Glyoxal oligomer Glyoxal oligomer	Levulinic acid $\mu_{H} = \int_{0}^{0} \int_{0}^{0} \int_{0}^{0}$ Benzyl alcohol $\int_{0}^{0} \int_{0}^{0} \int_{0}^{0}$ Cyclohexanone, $\int_{0}^{0} \int_{0}^{0} \int_{0}^{0}$ Glycolic acid dimer $\int_{0}^{0} \int_{0}^{0} \int_{0}^{0}$ Glycolic acid dimer $\int_{0}^{0} \int_{0}^{0} \int_{0}^{0}$ Glyoxal oligomer $\int_{0}^{0} \int_{0}^{0} \int_{0}^{0} \int_{0}^{0}$ Glyoxal oligomer $\int_{0}^{0} \int_{0}^{0} \int_{0}^{0} \int_{0}^{0} \int_{0}^{0}$ Glyoxal oligomer $= \int_{0}^{0} \int_{0}$	Levulinic acid $in f + f + f$ SFE/GC-MSBenzyl alcohol $f + f + f$ SFE/GC-MSCyclohexanone, 3,3,5-trimethyl $j + f + f$ SFE/GC-MSGlycolic acid dimer $f + f + f + f$ ESI-OrbitrapGlyoxal oligomer $f + f + f + f + f + f + f + f + f + f +$

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# 740 3.3.2.1. SFE/GC-MS analysis

741 Figure 6 shows the comparison of four gas chromatograms obtained for experiments

 $D_2$ ,  $D_{13}$ ,  $D_{15}$  and  $D_{16}$  at 80% RH, with and without irradiation, with and without ozone

743 (Table 1).







744

745 Figure 6. SFE/GC-MS chromatograms recorded from filters collected during one dust control 746 experiment and four glyoxal uptake experiments under four different conditions. From up to bottom: the 747 first chromatogram is from Experiment D2, dust control experiment under 80% RH and irradiation. The 748 second is from Experiment D<sub>16</sub> with dust and glyoxal at 80% RH under dark conditions. The third is from Experiment D<sub>15</sub> with dust and glyoxal at 80% RH under irradiated conditions. The fourth and fifth 749 750 chromatograms are from Experiment D<sub>13</sub> with dust and glyoxal at 80% RH in the presence of ozone 751 under dark and irradiated conditions, respectively. The two peaks of higher intensity appearing after 41 752 minutes are from the internal standards added to the solution: tridecane around 41 minutes and ortho-753 toluic acid at about 41.8 minutes. The intensity is normalized to peak of the internal standard ortho-toluic 754 acid. The chromatograms start at 32 minutes as 15 minutes are required for the removal of CO<sub>2</sub> from 755 the extraction fluid, and approximately another 15 minutes represent the delay required for the solvent 756 to pass through the column and reach the electron ionization (EI) MS detection. For the second and third spectra, is noticeable that under irradiated conditions the number of peaks increases significantly 757 758 compared to dark conditions, likely due to enhanced chemical reactions driven by light. This effect 759 appears to be less pronounced in the presence of ozone.





761 The chromatograms of samples collected in the presence of glyoxal generally exhibit a higher number of peaks compared to samples of dust-only, indicating glyoxal 762 oxidation and production of organic aerosols. A higher number of peaks are detected 763 in samples after irradiation compared to dark conditions (see Figure 6). This suggests 764 that the exposure to light influences the chemical composition of the samples by 765 promoting pathways that alter the chromatogram profile, possibly through 766 photochemical reactions, which could lead to glyoxal oxidation products. However, in 767 the presence of ozone and glyoxal, the chromatogram profile recorded under dark 768 769 conditions is comparable to that recorded in the presence of light, suggesting that 770 ozone may play a significant role in the oxidation process, driving similar chemical reactions in both light and dark environments. This could imply that the oxidative 771 772 capacity of ozone is sufficient to promote glyoxal oxidation and organic formation independently of photolytic processes, resulting in comparable chromatogram profiles 773 regardless of the presence of light. 774

The compounds identified using SFE/GC-MS primarily consist of carboxylic acids and 775 relatively light-weight carbonyl compounds (<150 Da). Lactic and levulinic acids are 776 detected in 10 and 9 samples, respectively, regardless of the experimental conditions. 777 778 Compounds detected less frequently include benzyl alcohol, 3.3.5trimethylcyclohexanone, methylphosphonic acid (5 samples), decanal (4 samples). 779 and various organic acids, including heptanoic, propanedioic, and hydroacrylic acid (3 780 samples), as well as benzoic acid and 1-octanol (2 samples). 781

Glycolic acid and is only detected during experiments with glyoxal. Light and the 782 presence of ozone seem to favour its formation. Indeed, the fourth panel in Figure 6 783 suggests that ozone might substitute light in promoting the formation of glycolic acid 784 from glyoxal at high RH, suggesting an alternative oxidative pathway. Glycolic acid is 785 also detected at 30% RH (not showed), both with and without light, in agreement with 786 the experiments on dust by Shen et al. (2016), but differently than reported by Galloway 787 788 et al. (2009) on ammonium sulphate. Monohydrated glyoxylic acid is found in one sample at 80% RH under irradiated conditions, likely due to the known pattern of 789 oxidation of glyoxal and glycolic acid with OH radicals (Buxton et al., 1997). 790

Twelve mass spectra have a profile unrecognised by the NIST library. An example isshown in Figure 7.







794

**Figure 7.** SFE Electron ionization mass spectrum recorded from the filter collected during the experiment  $D_{13}$  (80% relative humidity in the presence of O<sub>3</sub>), where the characteristic peaks of the TMS functionalization of two hydroxyl groups (74 and 147 m/z) and one carboxylic group (117 m/z) are observed. The retention time of the peak corresponding to this mass spectrum was 39.58 min.

799

800 Notably, ten are collected under humid conditions (both 30% and 80% RH) and in the presence of light. Eleven are characterised by 73 m/z [Si(CH<sub>3</sub>)<sub>3</sub>]<sup>+</sup> for at least one 801 802 functionalization, 147 m/z [(CH<sub>3</sub>)<sub>2</sub>Si=OSi(CH<sub>3</sub>)<sub>3</sub>]<sup>+</sup> for at least two, and their multiples for a greater number of hydroxyl functionalities (-OH). In addition, the fragment at 117 m/z 803 804 [COO=Si(CH<sub>3</sub>)<sub>3</sub>]<sup>+</sup> is detected. These compounds are attributed to trimethylsilyl-multifunctionalised molecules from small oligomers of hydrated glyoxal (multiple 805 functionalities) that have undergone partial oxidation, as indicated by the presence of 806 carboxylic group peak in most of the spectra (117 m/z). 807

808 3.3.2.2. ESI-Orbitrap

On the 15 samples analysed by ESI-Orbitrap (Table S2 in the Supplementary Material), signals attributable to products of the oxidation, hydration, or oligomerization of glyoxal are found only for experiments at high relative humidity (80%), both in the dark but mostly in irradiated conditions. Figure 8 illustrates the mass spectrum and the assigned formula for experiment D<sub>10</sub>.







815

Figure 8. ESI Orbitrap MS from the filter D<sub>10</sub>: uptake of glyoxal on mineral dust at 80% RH and under
 irradiation. In the mass spectra, the peaks referring to formulas for which is possible to suggest a
 structure from glyoxal reactivity are labelled.

819

Molecules corresponding to oxidised compounds are predominantly observed on filters 820 821 collected under irradiated conditions: C<sub>2</sub>H<sub>4</sub>O<sub>3</sub> (also detected by SFE/GC-MS) attributed to glycolic acid; C<sub>2</sub>H<sub>2</sub>O<sub>4</sub>, attributed to oxalic acid, and C<sub>4</sub>H<sub>4</sub>O<sub>4</sub>, attributed to a dimer of 822 823 glycolic acid. Its monohydrated form, C<sub>2</sub>H<sub>4</sub>O<sub>4</sub> (also detected by SFE/GC-MS) is also observed in dark conditions. In the presence of ozone, glyoxylic acid is also observed 824 in dark conditions. Particle phase formic acid, observed on both dust and ammonium 825 sulphate by various authors (Galloway et al., 2009; Rubasinghege et al., 2013; Shen 826 et al., 2016), is not detected in our study neither by molecular analysis nor the ACSM. 827 Although the reasons remain unclear, Shen et al. (2016) suggested that, above 828 approximately 50% RH, adsorbed water could compete for surface reactive sites 829 resulting in suppressing the formation of organic acids onto the dust particles. In 830 contrast, the formation of glycolic and glycoxylic acid appears to be less affected by the 831 presence of adsorbed water, as they are detected solely in humid conditions. This is 832 possibly due to differences in their chemical pathways or their interactions with the dust 833 834 surface. These acids may form through mechanisms that are less competitive with water adsorption, or their precursors have a higher affinity for the reactive sites on dust 835





- particles (possibly due to the presence of two carbonyl regions) allowing their formation
- to proceed even in the presence of high humidity.
- Compounds from C<sub>4</sub> to C<sub>10</sub>, oligomerization products of the glyoxal mono and di 838 hydrate forms, are observed only at 80% RH. The following peaks are detected and 839 attributed: C<sub>4</sub>H<sub>6</sub>O<sub>6</sub> (1 monohydrated glyoxal + 1 dehydrated glyoxal forming a 5-atom 840 ring), C<sub>8</sub>H<sub>16</sub>O<sub>12</sub> (4 dehydrated glyoxal molecules forming an 8-membered ring), and 841 C10H12O11 (4 monohydrated glyoxal molecules + 1 dehydrated glyoxal molecule 842 forming a ring structure). The oligomer  $C_6H_6O_6$  (3 molecules of monohydrated glyoxal 843 844 forming a 6-membered ring) is detected only under dark conditions. C<sub>8</sub>H<sub>16</sub>O<sub>6</sub> corresponds to an oligomer previously observed by Shen et al. (2016) on mineral dust. 845 Based on the observations above, Figure 9 shows the suggested chemical 846 mechanism. 847



Figure 9. Proposed reaction scheme to explain the glyoxal-related molecular formulas detected through
 ESI-Orbitrap mass spectrometry and SFE/GC-MS.





- The primary oligomers detected are attributed to the condensation of hydrated glyoxal forms specifically the mono- and di-hydrated glyoxal forms, through hydrolysis. The
- 854 oligomer V is attributed to the condensation of two glycolic acid molecules.

## 855 3.4 Oxidation state and reversibility

The results on the oxidative properties of the aged dust are summarised in Figure 10 showing the van Krevelen diagram obtained from the ESI-Orbitrap analysis of a sample collected during experiment  $D_2$  (dust control experiments without glyoxal) and a sample collected during experiment  $D_{10}$  (dust and glyoxal), both in the dark and with irradiation and 80% RH.

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862

**Figure 10.** Van Krevelen diagrams recorded at 80% RH for: in the top line (a) experiments in the dark for mineral dust only (control experiment D<sub>2</sub>, left) and one ageing experiment of dust with glyoxal (experiment D10); bottom line (b) same with irradiation.

866

Under the different conditions presented, the samples exhibited varying levels of particulate organic matter and number of ESI Orbitrap peaks detected.





In the absence of glyoxal, only a few signals are detected. In dark conditions, the particulate organic matter of filter is 0.9  $\mu$ g with 102 peaks detected. In irradiated conditions, the particulate organic matter is 0.8  $\mu$ g with 86 peaks detected. Peaks are mostly in the range of O/C < 1 and 0.5 < H/C < 2.5, both in the dark and with irradiation. The distribution of values of both ratios is rather similar, while the appearance of molecules for families CHON and CHONS is observed when the lights are on.

The number of detected signals increases significantly in the presence of glyoxal, (right column of Figure 10), in particular in the presence of light. In dark conditions, the particulate organic matter on the filter is 1.7  $\mu$ g, yielding 398 signals detected. In irradiated conditions, the particulate organic matter is 0.6  $\mu$ g, resulting in 310 signals detected. The predominant family in this case is that of CHO molecules. The appearance of signals with O/C ratio higher than 1 is attributed to photo-oxidation.

For comparison, the O/C ratio of the bulk aerosol measured at 80% RH by the ACSM and XPS is shown in Figure 11 (results in dry and 30% RH conditions are shown in Figure S4 in the Supplementary Material).

884



**Figure 11.** Time series of O/C ratio measured with the ACSM during ageing and 80% RH (experiment D<sub>15</sub>). The black arrows show the duration of filter sampling and the corresponding O/C values obtained by XPS analysis. The yellow-highlighted portion of the graph indicates the interval where irradiation takes place, while the green vertical dashed lines indicate the moment of glyoxal injection in the chamber.





892 Figure 11 shows that the O/C ratio of the organic material in the native dust is around 1.5. During the uptake of glyoxal, the ratio decreases to 1 to finally revert to 1.5 within 893 approximately one hour. The agreement between the measurements of the ACSM and 894 895 the XPS analysis indicates that the ACSM probes the organic matter at the surface of the particles, which we expect to be involved in the reactivity towards glyoxal. The 896 comparison of the results in Figures 10 and 11 suggests that while glyoxal and high 897 volatility oxidation products evaporate, low volatility and heavy compounds such as 898 oligomers, remains on the dust and modifies in an irreversible way its surface 899 composition. This is in agreement with previous studies carried out on ammonium 900 sulphate seeds (Kroll et al., 2005; Galloway et al., 2009; De Haan et al., 2020; Hu et 901 902 al., 2022;).

## 903 4 Conclusive remarks

In this paper, we investigated the uptake of glyoxal by realistic submicron mineral dust aerosol particles from a natural soil (Gobi Desert), airborne in a large simulation chamber, and aged under variable experimental conditions of relative humidity, irradiation, and ozone concentrations. Results can be summarised and commented as follows:

The uptake of glyoxal on the dust particles occurs in humid conditions exceeding 909 • 30% RH. These observations agree with the results of Liggio et al. (2005a; 910 911 2005b) on the uptake of glyoxal on ammonium sulphate aerosols, observing the formation of organic matter only when RH exceeded 50%. Trainic et al. (2011) 912 found that the uptake of glyoxal on glycine and ammonium sulphate particles 913 occurred only when the relative humidity was above 35%. On the contrary, both 914 Shen et al (2016) and Zogka et al. (2024) demonstrated that the uptake can occur 915 in dry conditions too, which we did not observe. 916

• The uptake of glyoxal on the dust particles starts as soon as the glyoxal is injected in the chamber. In this study we used a single and instantaneous injection of glyoxal, and not a constant steady state flux. Furthermore, in humid conditions, upon injection, glyoxal is partitioned rapidly between the gas phase and the chamber walls. Both facts are actually an advantage to scale our results to ambient conditions. Indeed, Volkamer et al. (2005) estimated that the lifetime of glyoxal in the daytime is of around 1.3 hours. Alvarado et al. (2020) showed that





the long-range transport of glyoxal produced from a point source (Canadian wildfires) may be possible only by invoking the progressive oxidation of its longerlived precursors in the plume. In the scenario where dust aerosols interact with a glyoxal plume from a point source, one can expect an interaction time of a few minutes, compatible with that of this study.

At 80% RH, the measured uptake coefficient of glyoxal on mineral dust is  $\gamma = (9$ 929  $\pm$  5) × 10<sup>-3</sup>. Because the uptake follows a first order kinetic, the measured uptake 930 coefficient is independent on the glyoxal concentration and transferable to 931 atmospheric conditions. The uptake coefficient on dust is nearly two orders of 932 magnitude higher than for ammonium sulphate ( $y_{qas-AS} = 1.1 (\pm 0.2) \times 10^{-4}$ ) at the 933 same relative humidity (our study as well as Curry et al., 2018; De Haan et al., 934 2020; Galloway et al., 2009; Liggio et al., 2005b, a; Trainic et al., 2011) but lower 935 than  $\gamma_{gas-AS} = 2.9 \times 10^{-3}$  at a lower relative humidity (Liggio et al., 2005). 936

The difference could be due to the higher hygroscopicity of ammonium sulphate, 937 enhances water's competition with glyoxal for adsorption sites at 80% RH, when 938 indeed ammonium sulphate is deliquescent. This suggest nonetheless that dust 939 aerosols could play a very substantial role in the formation of organic aerosols at 940 high relative humidity compared to ammonium sulphate, which is often used as 941 an aerosol proxy. Additionally, we found that the uptake coefficient measured by 942 the loss of gas-phase glyoxal molecules agrees very well with the rate of 943 formation of secondary organic mass on the particles. This suggests that the 944 totality of the mass of reacting glyoxal is transformed in organic matter on the 945 surface of the dust particles. 946

The uptake of glyoxal and the formation of organic matter last approximately 20 947 minutes, after which evaporation occurs. However, we demonstrate the uptake 948 949 of glyoxal modifies irreversibly both the composition and the physical properties 950 of mineral dust. Oligomers and organic acids are detected on the dust even after the uptake has finished. Our findings support several key insights into the 951 irreversible uptake of glyoxal on mineral particles, as discussed by Shen et al. 952 (2016), that found that glyoxal oligomers exhibited a higher degree of 953 oligomerization (≥ 4) than previously reported in the aqueous phase and on acidic 954 seed particles (≤ 3; Liggio et al., 2005a; Nozière et al., 2009) and that adsorbed 955 water on particles favoured the formation of oligomers. This is also in agreement 956





with the field observations by consistent with the observations conducted byWang et al. (2015).

The presence of organic acids, such as glycolic and glycoxylic acids has 959 implications for aerosol pH, as they can influence the acidity of aerosols and their 960 ability to dissolve metals, potentially impacting atmospheric chemistry and the 961 reactivity of aerosol particles (Giorio et al., 2022). Changes in aerosol pH can, in 962 turn, affect the hygroscopic properties of aerosols, influencing their ability to 963 adsorb water and grow in size. Indeed, we observe that the volume of the dust 964 965 particle increases during the uptake and that this growth is persistent, henceforth becoming more efficient in interacting with visible light and form cloud droplets. 966 967 The newly formed organic matter from glyoxal on dust particles could also alter the aerosol's optical properties, affecting its ability to absorb solar radiation, as 968 969 recently observed in aqueous solution (De Haan et al., 2023) or on ammonium sulphate aerosols (De Haan et al. 2020; Trainic et al. 2011). For example, the 970 presence of hydrated glyoxal oligomeric structures has already been observed to 971 have UV radiation absorption properties (Kalberer et al., 2004; Shapiro et al., 972 2009). 973

In conclusion, our study reveals a significant quantitative transfer of gas-phase glyoxal 974 975 molecules to mineral dust aerosol surfaces, occurring within a timescale of a few minutes, underscoring the important role of dust-glyoxal interactions in the 976 977 atmosphere. Neglecting the uptake pathway on dust could result in an underestimation of glyoxal removal from the atmosphere, potentially leading to disparities between 978 979 model predictions and observed gaseous concentrations of glyoxal (Kluge et al., 2023; 980 Ling et al., 2020; Volkamer et al., 2007; Washenfelder et al., 2011). The results of this study could have important implications for the aerosol direct and indirect radiative 981 982 effect and aerosol pH. Further studies should investigate dust from different sources and mineralogy, poly-disperse size distribution including the coarse mode and lower 983 glyoxal concentration. 984

985

Data availability. The simulation chamber experiments that support the findings of this study are
 available through the Database of Atmospheric Simulation Chamber Studies (DASCS) of the
 EUROCHAMP Data Centre (https:// data.eurochamp.org/ data- access/ chamber- experiments/).

989 **Code availability**. The routine used for fitting the size distribution is available at 990 https://doi.org/10.5281/zenodo.8135133 (Baldo and Lu, 2023). Note that in this study we only





used the size distribution measured by the OPC instrument, which was fitted with a lognormalfunction.

Author contributions. PF, JFD and FB conceptualized the study. PF and FB led the paper 993 writing, with contributions from all the authors. JFD provided with expertise on multi-phase 994 995 chemistry. CB analysed the aerosol size distribution data. VM supervised the analysis of PTR-MS data. FB and CG performed the ESI-Orbitrap analysis of filter samples. FB and JFB 996 performed the analysis of ACSM observations. FB, TB and AG performed the SFE/CG-MS 997 analysis of filter samples. FB, GN and SC performed the thermo-optical analysis of filter 998 999 samples. FB, MC, AB, EP, VM, BPV and PF conducted the chamber experiments. MR 1000 provided with the soil sample and expertise on heterogeneous chemistry. PF provided with 1001 funding.

1002 **Competing interests**. The authors declare no competing interests.

1003 **Special issue statement**. This article is not part of a special issue. It is not associated with a 1004 conference.

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## 1015 **References**

1016	Adebiyi, A., Kok, J. F., Murray, B. J., Ryder, C. L., Stuut, JB. W., Kahn, R. A., Knippertz, P., Formenti,
1017	P., Mahowald, N. M., Pérez García-Pando, C., Klose, M., Ansmann, A., Samset, B. H., Ito, A.,
1018	Balkanski, Y., Di Biagio, C., Romanias, M. N., Huang, Y., and Meng, J.: A review of coarse mineral
1019	dust in the Earth system, Aeol. Res., 60, 100849, https://doi.org/10.1016/j.aeolia.2022.100849,
1020	2023.
1021	<ul> <li>Aiken, A. C., DeCarlo, P. F., Kroll, J. H., Worsnop, D. R., Huffman, J. A., Docherty, K. S., Ulbrich, I. M.,</li></ul>
1022	Mohr, C., Kimmel, J. R., Sueper, D., Sun, Y., Zhang, Q., Trimborn, A., Northway, M., Ziemann,
1023	P. J., Canagaratna, M. R., Onasch, T. B., Alfarra, M. R., Prevot, A. S. H., Dommen, J., Duplissy,
1024	J., Metzger, A., Baltensperger, U., and Jimenez, J. L.: O/C and OM/OC Ratios of Primary,
1025	Secondary, and Ambient Organic Aerosols with High-Resolution Time-of-Flight Aerosol Mass
1026	Spectrometry, Environ. Sci. Technol., 42, 4478–4485, https://doi.org/10.1021/es703009q, 2008.
1027	Baldo, C., Formenti, P., Di Biagio, C., Lu, G., Song, C., Cazaunau, M., Pangui, E., Doussin, JF.,
1028	Dagsson-Waldhauserova, P., Arnalds, O., Beddows, D., MacKenzie, A. R., and Shi, Z.: Complex
1029	refractive index and single scattering albedo of Icelandic dust in the shortwave part of the
1030	spectrum, Atmos. Chem. Phys., 23, 7975–8000, https://doi.org/10.5194/acp-23-7975-2023,
1031	2023.
1032 1033 1034	Baldo, C. and G. Lu. (2023). Research code supporting "Complex refractive index and single scattering albedo of Icelandic dust in the shortwave spectrum" (v1.0.0). Zenodo. https://doi.org/10.5281/zenodo.8135133
1035	Bauer, S. E., Mishchenko, M. I., Lacis, A. A., Zhang, S., Perlwitz, J., and Metzger, S. M.: Do sulfate and
1036	nitrate coatings on mineral dust have important effects on radiative properties and climate
1037	modeling?, J. Geophys. Res., 112, D06307, https://doi.org/10.1029/2005JD006977, 2007.
1038	Brégonzio-Rozier, L., Giorio, C., Siekmann, F., Pangui, E., Morales, S. B., Temime-Roussel, B., Gratien,
1039	A., Michoud, V., Cazaunau, M., DeWitt, H. L., Tapparo, A., Monod, A., and Doussin, JF.:
1040	Secondary organic aerosol formation from isoprene photooxidation during cloud condensation–
1041	evaporation cycles, Atmos. Chem. Phys., 16, 1747–1760, https://doi.org/10.5194/acp-16-1747-
1042	2016, 2016.
1043 1044	Buxton, G. V., Malone, T. N., and Arthur Salmon, G.: Oxidation of glyoxal initiated by OH in oxygenated aqueous solution, Faraday Trans., 93, 2889–2891, https://doi.org/10.1039/a701468f, 1997.
1045	Carlton, A. G., Turpin, B. J., Altieri, K. E., Seitzinger, S., Reff, A., Lim, HJ., and Ervens, B.: Atmospheric
1046	oxalic acid and OA production from glyoxal: Results of aqueous photooxidation experiments,
1047	Atmos. Environ., 41, 7588–7602, https://doi.org/10.1016/j.atmosenv.2007.05.035, 2007.
1048 1049 1050 1051	Castellanos, P., Colarco, P., Espinosa, W. R., Guzewich, S. D., Levy, R. C., Miller, R. L., Chin, M., Kahn, R. A., Kemppinen, O., Moosmüller, H., Nowottnick, E. P., Rocha-Lima, A., Smith, M. D., Yorks, J. E., and Yu, H.: Mineral dust optical properties for remote sensing and global modeling: A review, Remote Sensing of Environment, 303, 113982, https://doi.org/10.1016/j.rse.2023.113982, 2024.
1052	Chan Miller, C., Jacob, D. J., Marais, E. A., Yu, K., Travis, K. R., Kim, P. S., Fisher, J. A., Zhu, L., Wolfe,
1053	G. M., Hanisco, T. F., Keutsch, F. N., Kaiser, J., Min, KE., Brown, S. S., Washenfelder, R. A.,
1054	González Abad, G., and Chance, K.: Glyoxal yield from isoprene oxidation and relation to
1055	formaldehyde: chemical mechanism, constraints from SENEX aircraft observations, and
1056	interpretation of OMI satellite data, Atmos. Chem. Phys., 17, 8725–8738,
1057	https://doi.org/10.5194/acp-17-8725-2017, 2017.
1058	Chen, S., Chen, J., Zhang, Y., Lin, J., Bi, H., Song, H., Chen, Y., Lian, L., Liu, C., and Zhang, R.:
1059	Anthropogenic dust: sources, characteristics and emissions, Environ. Res. Lett., 18, 103002,
1060	https://doi.org/10.1088/1748-9326/acf479, 2023.
1061	Chiappini, L., Perraudin, E., Durand-Jolibois, R., and Doussin, J. F.: Development of a supercritical fluid
1062	extraction–gas chromatography–mass spectrometry method for the identification of highly polar
1063	compounds in secondary organic aerosols formed from biogenic hydrocarbons in smog chamber
1064	experiments, Anal. Bioanal. Chem., 386, 1749–1759, https://doi.org/10.1007/s00216-006-0744-
1065	3, 2006.





1066	Chirizzi, D., Cesari, D., Guascito, M. R., Dinoi, A., Giotta, L., Donateo, A., and Contini, D., Influence of
1067	Saharan dust outbreaks and carbon content on oxidative potential of water-soluble fractions of
1068	PM2. 5 and PM10. Atmos. Environ., 163, 1-8, 2017
1069	Crowley, J. N., Ammann, M., Cox, R. A., Hynes, R. G., Jenkin, M. E., Mellouki, A., Rossi, M. J., Troe,
1070	J., and Wallington, T. J.: Evaluated kinetic and photochemical data for atmospheric chemistry:
1071	Volume V – heterogeneous reactions on solid substrates, Atmos. Chem. Phys., 10, 9059–9223,
1072	https://doi.org/10.5194/acp-10-9059-2010, 2010.
1073	Curry, L. A., Tsui, W. G., and McNeill, V. F.: Technical note: Updated parameterization of the reactive
1074	uptake of glyoxal and methylglyoxal by atmospheric aerosols and cloud droplets, Atmos. Chem.
1075	Phys., 18, 9823–9830, https://doi.org/10.5194/acp-18-9823-2018, 2018.
1076	De Haan, D. O., Hawkins, L. N., Jansen, K., Welsh, H. G., Pednekar, R., De Loera, A., Jimenez, N. G.,
1077	Tolbert, M. A., Cazaunau, M., Gratien, A., Bergé, A., Pangui, E., Formenti, P., and Doussin, J
1078	F.: Glyoxal's impact on dry ammonium salts: fast and reversible surface aerosol browning, Atmos.
1079	Chem. Phys., 20, 9581–9590, https://doi.org/10.5194/acp-20-9581-2020, 2020.
1080	De Haan, D. O., Hawkins, L. N., Wickremasinghe, P. D., Andretta, A. D., Dignum, J. R., De Haan, A. C.,
1081	Welsh, H. G., Pennington, E. A., Cui, T., Surratt, J. D., Cazaunau, M., Pangui, E., and Doussin,
1082	JF.: Brown Carbon from Photo-Oxidation of Glyoxal and SO <sub>2</sub> in Aqueous Aerosol, ACS Earth
1083	Space Chem., 7, 1131–1140, https://doi.org/10.1021/acsearthspacechem.3c00035, 2023.
1084	Denjean, C., Formenti, P., Picquet-Varrault, B., Katrib, Y., Pangui, E., Zapf, P., and Doussin, J. F.: A
1085	new experimental approach to study the hygroscopic and optical properties of aerosols:
1086	application to ammonium sulfate particles, Atmos. Meas. Tech., 7, 183–197,
1087	https://doi.org/10.5194/amt-7-183-2014, 2014.
1088	Di Biagio, C., Formenti, P., Balkanski, Y., Caponi, L., Cazaunau, M., Pangui, E., Journet, E., Nowak, S.,
1089	Caquineau, S., Andreae, M. O., Kandler, K., Saeed, T., Piketh, S., Seibert, D., Williams, E., and
1090	Doussin, JF.: Global scale variability of the mineral dust long-wave refractive index: a new
1091	dataset of in situ measurements for climate modeling and remote sensing, Atmos. Chem. Phys.,
1092	17, 1901–1929, https://doi.org/10.5194/acp-17-1901-2017, 2017.
1093	Di Biagio, C., Formenti, P., Balkanski, Y., Caponi, L., Cazaunau, M., Pangui, E., Journet, E., Nowak, S.,
1094	Andreae, M. O., Kandler, K., Saeed, T., Piketh, S., Seibert, D., Williams, E., and Doussin, JF.:
1095	Complex refractive indices and single-scattering albedo of global dust aerosols in the shortwave
1096	spectrum and relationship to size and iron content, Atmos. Chem. Phys., 19, 15503–15531,
1097	https://doi.org/10.5194/acp-19-15503-2019, 2019.
1098	Dupart, Y., King, S. M., Nekat, B., Nowak, A., Wiedensohler, A., Herrmann, H., David, G., Thomas, B.,
1099	Miffre, A., Rairoux, P., D'Anna, B., and George, C.: Mineral dust photochemistry induces
1100	nucleation events in the presence of SO <sub>2</sub> , Proc. Natl. Acad. Sci. U.S.A., 109, 20842–20847,
1101	https://doi.org/10.1073/pnas.1212297109, 2012.
1102	Ervens, B. and Volkamer, R.: Glyoxal processing by aerosol multiphase chemistry: towards a kinetic
1103	modeling framework of secondary organic aerosol formation in aqueous particles, Atmos. Chem.
1104	Phys., 10, 8219–8244, https://doi.org/10.5194/acp-10-8219-2010, 2010.
1105	Fu, TM., Jacob, D. J., Wittrock, F., Burrows, J. P., Vrekoussis, M., and Henze, D. K.: Global budgets
1106	of atmospheric glyoxal and methylglyoxal, and implications for formation of secondary organic
1107	aerosols, J. Geophys. Res., 113, D15303, https://doi.org/10.1029/2007JD009505, 2008.
1108	Galloway, M. M., Chhabra, P. S., Chan, A. W. H., Surratt, J. D., Flagan, R. C., Seinfeld, J. H., and
1109	Keutsch, F. N.: Glyoxal uptake on ammonium sulphate seed aerosol: reaction products and
1110	reversibility of uptake under dark and irradiated conditions, Atmos. Chem. Phys., 2009.
1111	Giorio, C., Monod, A., Brégonzio-Rozier, L., DeWitt, H. L., Cazaunau, M., Temime-Roussel, B., Gratien,
1112	A., Michoud, V., Pangui, E., Ravier, S., Zielinski, A. T., Tapparo, A., Vermeylen, R., Claeys, M.,
1113	Voisin, D., Kalberer, M., and Doussin, JF.: Cloud Processing of Secondary Organic Aerosol from
1114	Isoprene and Methacrolein Photooxidation, J. Phys. Chem. A, 121, 7641–7654,
1115	https://doi.org/10.1021/acs.jpca.7b05933, 2017.





- Giorio, C., D'Aronco, S., Di Marco, V., Badocco, D., Battaglia, F., Soldà, L., Pastore, P., and Tapparo,
   A.: Emerging investigator series: aqueous-phase processing of atmospheric aerosol influences
   dissolution kinetics of metal ions in an urban background site in the Po Valley, Environ. Sci.:
   Processes Impacts, 24, 884–897, https://doi.org/10.1039/D2EM00023G, 2022.
- Goodman, A. L., Underwood, G. M., and Grassian, V. H.: Heterogeneous Reaction of NO2: Characterization of Gas-Phase and Adsorbed Products from the Reaction, 2NO2(g) + H2O(a) → HONO(g) + HNO3(a) on Hydrated Silica Particles, J. Phys. Chem. A, 103, 7217–7223, https://doi.org/10.1021/jp9910688, 1999.
- 1124Guo, Y., Wang, S., Zhu, J., Zhang, R., Gao, S., Saiz-Lopez, A., and Zhou, B.: Atmospheric1125formaldehyde, glyoxal and their relations to ozone pollution under low- and high-NOx regimes in1126summertime1127Shanghai,1127https://doi.org/10.1016/j.atmosres.2021.105635, 2021.
- Hettiarachchi, E. and Grassian, V. H.: Heterogeneous Reactions of Phenol on Different Components of Mineral Dust Aerosol: Formation of Oxidized Organic and Nitro-Phenolic Compounds, ACS EST
   Air, 1, 259–272, https://doi.org/10.1021/acsestair.3c00042, 2024.
- Horowitz, A., Meller, R., & Moortgat, G. K. (2001). The UV–VIS absorption cross sections of the α dicarbonyl compounds: pyruvic acid, biacetyl and glyoxal. J. Photochem. Photobiol., A, 146(1-2),
   19-27.
- Hu, J., Chen, Z., Qin, X., and Dong, P.: Reversible and irreversible gas-particle partitioning of dicarbonyl
   compounds observed in the real atmosphere, Atmos. Chem. Phys., 22, 6971–6987,
   https://doi.org/10.5194/acp-22-6971-2022, 2022.
- Joshi, N., Romanias, M. N., Riffault, V., and Thevenet, F.: Investigating water adsorption onto natural mineral dust particles: Linking DRIFTS experiments and BET theory, Aeolian Res, 27, 35–45, https://doi.org/10.1016/j.aeolia.2017.06.001, 2017.
- Lai, A. C., Nazaroff, W. W., Modeling indoor particle deposition from turbulent flow onto smooth surfaces.
   J. Aerosol Sci., 31, 463-476, 2000.
- Kalberer, M., Paulsen, D., Sax, M., Steinbacher, M., Dommen, J., Prevot, A. S. H., Fisseha, R.,
  Weingartner, E., Frankevich, V., Zenobi, R., and Baltensperger, U.: Identification of Polymers as
  Major Components of Atmospheric Organic Aerosols, Science, 303, 1659–1662,
  https://doi.org/10.1126/science.1092185, 2004.
- Kluge, F., Hüneke, T., Lerot, C., Rosanka, S., Rotermund, M. K., Taraborrelli, D., Weyland, B., and
  Pfeilsticker, K.: Airborne glyoxal measurements in the marine and continental atmosphere:
  comparison with TROPOMI observations and EMAC simulations, Atmos. Chem. Phys., 23, 1369–
  1401, https://doi.org/10.5194/acp-23-1369-2023, 2023.
- Knippertz, P. and Stuut, J.-B. W. (Eds.): Mineral Dust: A Key Player in the Earth System, Springer
   Netherlands, Dordrecht, https://doi.org/10.1007/978-94-017-8978-3, 2014.
- Knote, C., Hodzic, A., Jimenez, J. L., Volkamer, R., Orlando, J. J., Baidar, S., Brioude, J., Fast, J.,
  Gentner, D. R., Goldstein, A. H., Hayes, P. L., Knighton, W. B., Oetjen, H., Setyan, A., Stark, H.,
  Thalman, R., Tyndall, G., Washenfelder, R., Waxman, E., and Zhang, Q.: Simulation of semiexplicit mechanisms of OA formation from glyoxal in aerosol in a 3-D model, Atmos. Chem. Phys.,
  14, 6213–6239, https://doi.org/10.5194/acp-14-6213-2014, 2014.
- Kok, J. F., Adebiyi, A. A., Albani, S., Balkanski, Y., Checa-Garcia, R., Chin, M., Colarco, P. R., Hamilton,
  D. S., Huang, Y., Ito, A., Klose, M., Li, L., Mahowald, N. M., Miller, R. L., Obiso, V., Pérez GarcíaPando, C., Rocha-Lima, A., and Wan, J. S.: Contribution of the world's main dust source regions
  to the global cycle of desert dust, Aerosols/Atmospheric Modelling/Troposphere/Physics (physical
  properties and processes), https://doi.org/10.5194/acp-2021-4, 2021.
- Kok, J. F., Storelvmo, T., Karydis, V. A., Adebiyi, A. A., Mahowald, N. M., Evan, A. T., He, C., and Leung,
   D. M.: Mineral dust aerosol impacts on global climate and climate change, Nat Rev Earth Environ,
   4, 71–86, https://doi.org/10.1038/s43017-022-00379-5, 2023.
- Kourtchev, I., Doussin, J.-F., Giorio, C., Mahon, B., Wilson, E. M., Maurin, N., Pangui, E., Venables, D.
   S., Wenger, J. C., and Kalberer, M.: Molecular composition of fresh and aged secondary organic





1167 1168	aerosol from a mixture of biogenic volatile compounds: a high-resolution mass spectrometry study, Atmos. Chem. Phys., 15, 5683–5695, https://doi.org/10.5194/acp-15-5683-2015, 2015.
1169	Kroll, J. H., Ng, N. L., Murphy, S. M., Varutbangkul, V., Flagan, R. C., and Seinfeld, J. H.: Chamber
1170	studies of secondary organic aerosol growth by reactive uptake of simple carbonyl compounds,
1171	J. Geophys. Res., 110, 2005JD006004, https://doi.org/10.1029/2005JD006004, 2005.
1172	Lewis, A. C., Hopkins, J. R., Carslaw, D. C., Hamilton, J. F., Nelson, B. S., Stewart, G., Dernie, J.,
1173	Passant, N., and Murrells, T.: An increasing role for solvent emissions and implications for future
1174	measurements of volatile organic compounds, Phil. Trans. R. Soc. A., 378, 20190328,
1175	https://doi.org/10.1098/rsta.2019.0328, 2020.
1176	Li, L., Mahowald, N. M., Miller, R. L., Pérez García-Pando, C., Klose, M., Hamilton, D. S., Gonçalves
1177	Ageitos, M., Ginoux, P., Balkanski, Y., Green, R. O., Kalashnikova, O., Kok, J. F., Obiso, V.,
1178	Paynter, D., and Thompson, D. R.: Quantifying the range of the dust direct radiative effect due to
1179	source mineralogy uncertainty, Atmos. Chem. Phys., 21, 3973–4005, https://doi.org/10.5194/acp-
1180	21-3973-2021, 2021.
1181	Li, Q., Gong, D., Wang, H., Wang, Y., Han, S., Wu, G., Deng, S., Yu, P., Wang, W., and Wang, B.: Rapid
1182	increase in atmospheric glyoxal and methylglyoxal concentrations in Lhasa, Tibetan Plateau:
1183	Potential sources and implications, Sci. Total Environ., 824, 153782,
1184	https://doi.org/10.1016/j.scitotenv.2022.153782, 2022.
1185	Liggio, J., Li, SM., and McLaren, R.: Heterogeneous Reactions of Glyoxal on Particulate Matter:
1186	Identification of Acetals and Sulfate Esters, Environ. Sci. Technol., 39, 1532–1541,
1187	https://doi.org/10.1021/es048375y, 2005a.
1188 1189	Liggio, J., Li, S., and McLaren, R.: Reactive uptake of glyoxal by particulate matter, J. Geophys. Res., 110, 2004JD005113, https://doi.org/10.1029/2004JD005113, 2005b.
1190	Lim, Y. B., Tan, Y., Perri, M. J., Seitzinger, S. P., and Turpin, B. J.: Aqueous chemistry and its role in
1191	secondary organic aerosol (OA) formation, Atmos. Chem. Phys., 10, 10521–10539,
1192	https://doi.org/10.5194/acp-10-10521-2010, 2010.
1193	Ling, Z., Xie, Q., Shao, M., Wang, Z., Wang, T., Guo, H., and Wang, X.: Formation and sink of glyoxal
1194	and methylglyoxal in a polluted subtropical environment: observation-based photochemical
1195	analysis and impact evaluation, Atmos. Chem. Phys., 20, 11451–11467,
1196	https://doi.org/10.5194/acp-20-11451-2020, 2020.
1197	Liu, C., Chu, B., Liu, Y., Ma, Q., Ma, J., He, H., Li, J., and Hao, J.: Effect of mineral dust on secondary
1198	organic aerosol yield and aerosol size in α-pinene/NOx photo-oxidation, Atmos. Environ., 77,
1199	781–789, https://doi.org/10.1016/j.atmosenv.2013.05.064, 2013.
1200	Mahowald, N., Albani, S., Kok, J. F., Engelstaeder, S., Scanza, R., Ward, D. S., and Flanner, M. G.: The
1201	size distribution of desert dust aerosols and its impact on the Earth system, Aeolian Research,
1202	15, 53–71, https://doi.org/10.1016/j.aeolia.2013.09.002, 2014.
1203	Nault, B. A., Croteau, P., Jayne, J., Williams, A., Williams, L., Worsnop, D., Katz, E. F., DeCarlo, P. F.,
1204	and Canagaratna, M.: Laboratory evaluation of organic aerosol relative ionization efficiencies in
1205	the aerodyne aerosol mass spectrometer and aerosol chemical speciation monitor, Aerosol
1206	Science and Technology, 57, 981–997, https://doi.org/10.1080/02786826.2023.2223249, 2023.
1207	Ng, N. L., Canagaratna, M. R., Jimenez, J. L., Chhabra, P. S., Seinfeld, J. H., and Worsnop, D. R.:
1208	Changes in organic aerosol composition with aging inferred from aerosol mass spectra, Atmos.
1209	Chem. Phys., 11, 6465–6474, https://doi.org/10.5194/acp-11-6465-2011, 2011.
1210 1211 1212	Nie, W., Ding, A., Wang, T., Kerminen, VM., George, C., Xue, L., Wang, W., Zhang, Q., Petäjä, T., Qi, X., Gao, X., Wang, X., Yang, X., Fu, C., and Kulmala, M.: Polluted dust promotes new particle formation and growth, Sci Rep, 4, 6634, https://doi.org/10.1038/srep06634, 2014.
1213	Nieder, R., Benbi, D. K., and Reichl, F. X.: Soil-Borne Particles and Their Impact on Environment and
1214	Human Health, in: Soil Components and Human Health, Springer Netherlands, Dordrecht, 99–
1215	177, https://doi.org/10.1007/978-94-024-1222-2_3, 2018.





1216 Nozière, B., Dziedzic, P., and Córdova, A.: Products and Kinetics of the Liquid-Phase Reaction of 1217 Glyoxal Catalyzed by Ammonium lons (NH4+), J. Phys. Chem. A, 113, 231-237, 1218 https://doi.org/10.1021/jp8078293, 2009. 1219 Ooki, A. and Uematsu, M.: Chemical interactions between mineral dust particles and acid gases during 2004JD004737, events, 1220 Asian dust Res.. 110. .1 Geophys. https://doi.org/10.1029/2004JD004737, 2005. 1221 1222 Patriarca, C., Bergquist, J., Sjöberg, P. J. R., Tranvik, L., and Hawkes, J. A.: Online HPLC-ESI-HRMS 1223 Method for the Analysis and Comparison of Different Dissolved Organic Matter Samples, Environ. 1224 Sci. Technol., 52, 2091–2099, https://doi.org/10.1021/acs.est.7b04508, 2018. 1225 Ponczek, M., Hayeck, N., Emmelin, C., and George, C.: Heterogeneous photochemistry of dicarboxylic 1226 acids on mineral dust, Atmos. Environ., 212, 262-271, 1227 https://doi.org/10.1016/j.atmosenv.2019.05.032, 2019. Romanias, M. N., El Zein, A., and Bedjanian, Y.: Heterogeneous Interaction of H<sub>2</sub>O<sub>2</sub> with TiO<sub>2</sub> Surface under Dark and UV Light Irradiation Conditions, J. Phys. Chem. A, 116, 8191–8200, 1228 1229 1230 https://doi.org/10.1021/jp305366v, 2012. 1231 Rossignol, S., Aregahegn, K. Z., Tinel, L., Fine, L., Nozière, B., and George, C.: Glyoxal Induced 1232 Atmospheric Photosensitized Chemistry Leading to Organic Aerosol Growth, Environ. Sci. Technol., 48, 3218-3227, https://doi.org/10.1021/es405581g, 2014. 1233 1234 Rubasinghege, G., Ogden, S., Baltrusaitis, J., and Grassian, V. H.: Heterogeneous Uptake and 1235 Adsorption of Gas-Phase Formic Acid on Oxide and Clay Particle Surfaces: The Roles of Surface 1236 Hydroxyl Groups and Adsorbed Water in Formic Acid Adsorption and the Impact of Formic Acid 1237 Adsorption on Water Uptake, J. Phys. Chem. Α, 117, 11316-11327, 1238 https://doi.org/10.1021/jp408169w, 2013. 1239 Safi Shalamzari, M., Ryabtsova, O., Kahnt, A., Vermeylen, R., Hérent, M., Quetin-Leclercq, J., Van Der 1240 Veken, P., Maenhaut, W., and Claeys, M.: Mass spectrometric characterization of organosulfates 1241 related to secondary organic aerosol from isoprene, Rapid Comm Mass Spectrometry, 27, 784-794, https://doi.org/10.1002/rcm.6511, 2013. 1242 1243 Seisel, S., Lian, Y., Keil, T., Trukhin, M. E., and Zellner, R.: Kinetics of the interaction of water vapour 1244 with mineral dust and soot surfaces at T = 298 K, Phys. Chem. Chem. Phys., 6, 1926–1932, 1245 https://doi.org/10.1039/B314568A, 2004. 1246 Shapiro, E. L., Szprengiel, J., Sareen, N., Jen, C. N., Giordano, M. R., and McNeill, V. F.: Light-absorbing 1247 secondary organic material formed by glyoxal in aqueous aerosol mimics, Atmos. Chem. Phys., 1248 9, 2289-2300, https://doi.org/10.5194/acp-9-2289-2009, 2009. 1249 Shen, X., Wu, H., Zhao, Y., Huang, D., Huang, L., and Chen, Z.: Heterogeneous reactions of glyoxal on 1250 mineral particles: A new avenue for oligomers and organosulfate formation, Atmos. Environ., 131, 1251 133–140, https://doi.org/10.1016/j.atmosenv.2016.01.048, 2016. 1252 Shrivastava, M., Cappa, C. D., Fan, J., Goldstein, A. H., Guenther, A. B., Jimenez, J. L., Kuang, C., 1253 Laskin, A., Martin, S. T., Ng, N. L., Petaja, T., Pierce, J. R., Rasch, P. J., Roldin, P., Seinfeld, J. 1254 H., Shilling, J., Smith, J. N., Thornton, J. A., Volkamer, R., Wang, J., Worsnop, D. R., Zaveri, R. 1255 A., Zelenyuk, A., and Zhang, Q.: Recent advances in understanding secondary organic aerosol: 1256 for 55, 509-559, Implications global climate forcing, Rev. Geophys., https://doi.org/10.1002/2016RG000540, 2017. 1257 1258 Sun, Y. L., Zhang, Q., Anastasio, C., and Sun, J.: Insights into secondary organic aerosol formed via 1259 aqueous-phase reactions of phenolic compounds based on high resolution mass spectrometry. 1260 Atmos. Chem. Phys., 10, 4809–4822, https://doi.org/10.5194/acp-10-4809-2010, 2010. Tang, M., Huang, X., Lu, K., Ge, M., Li, Y., Cheng, P., Zhu, T., Ding, A., Zhang, Y., Gligorovski, S., 1261 1262 Song, W., Ding, X., Bi, X., and Wang, X.: Heterogeneous reactions of mineral dust aerosol: 1263 implications for tropospheric oxidation capacity, Atmos. Chem. Phys., 17, 11727-11777, 1264 https://doi.org/10.5194/acp-17-11727-2017, 2017. 1265 Tegen, I. and Fung, I.: Contribution to the atmospheric mineral aerosol load from land surface 1266 modification, J. Geophys. Res., 100, 18707, https://doi.org/10.1029/95JD02051, 1995.





- Turpin, B. J. and Huntzicker, J. J.: Identification of secondary organic aerosol episodes and quantitation of primary and secondary organic aerosol concentrations during SCAQS, Atmos. Environ., 29, 3527–3544, https://doi.org/10.1016/1352-2310(94)00276-Q, 1995.
- 1273 Usher, C. R., Michel, A. E., and Grassian, V. H.: Reactions on Mineral Dust, Chem. Rev., 103, 4883– 1274 4940, https://doi.org/10.1021/cr020657y, 2003.
- Volkamer, R., Platt, U., and Wirtz, K.: Primary and Secondary Glyoxal Formation from Aromatics:
   Experimental Evidence for the Bicycloalkyl-Radical Pathway from Benzene, Toluene, and p Xylene, J. Phys. Chem. A, 105, 7865–7874, https://doi.org/10.1021/jp010152w, 2001.
- Volkamer, R., San Martini, F., Molina, L. T., Salcedo, D., Jimenez, J. L., and Molina, M. J.: A missing sink for gas-phase glyoxal in Mexico City: Formation of secondary organic aerosol, Geophysical Research Letters, 34, 2007GL030752, https://doi.org/10.1029/2007GL030752, 2007.
- 1281 Vrekoussis, M., Wittrock, F., Richter, A., Burrows, J. P., and Cho, C.: Temporal and spatial variability of 1282 glyoxal as observed from space, Atmos. Chem. Phys., 2009.
- Wagner, C., Hanisch, F., Holmes, N., de Coninck, H., Schuster, G., and Crowley, J. N.: The interaction
   of N<sub>2</sub>O<sub>5</sub> with mineral dust: aerosol flow tube and Knudsen reactor studies, Atmos. Chem. Phys.,
   2008.
- Walker, H., Stone, D., Ingham, T., Hackenberg, S., Cryer, D., Punjabi, S., Read, K., Lee, J., Whalley,
  L., Spracklen, D. V., Carpenter, L. J., Arnold, S. R., and Heard, D. E.: Observations and modelling
  of glyoxal in the tropical Atlantic marine boundary layer, Atmos. Chem. Phys., 22, 5535–5557,
  https://doi.org/10.5194/acp-22-5535-2022, 2022.
- Wang, G., Cheng, C., Meng, J., Huang, Y., Li, J., and Ren, Y.: Field observation on secondary organic aerosols during Asian dust storm periods: Formation mechanism of oxalic acid and related compounds on dust surface, Atmos. Environ., 113, 169–176, https://doi.org/10.1016/j.atmosenv.2015.05.013, 2015.
- Washenfelder, R. A., Young, C. J., Brown, S. S., Angevine, W. M., Atlas, E. L., Blake, D. R., Bon, D. M.,
  Cubison, M. J., De Gouw, J. A., Dusanter, S., Flynn, J., Gilman, J. B., Graus, M., Griffith, S.,
  Grossberg, N., Hayes, P. L., Jimenez, J. L., Kuster, W. C., Lefer, B. L., Pollack, I. B., Ryerson, T.
  B., Stark, H., Stevens, P. S., and Trainer, M. K.: The glyoxal budget and its contribution to organic
  aerosol for Los Angeles, California, during CalNex 2010: Glyoxal Budget for Los Angeles, J.
  Geophys. Res., 116, https://doi.org/10.1029/2011JD016314, 2011.
- Webb, N. P. and Pierre, C.: Quantifying Anthropogenic Dust Emissions, Earth's Future, 6, 286–295,
   https://doi.org/10.1002/2017EF000766, 2018.
- Zeineddine, M. N., Romanias, M. N., Gaudion, V., Riffault, V., and Thévenet, F.: Heterogeneous
  Interaction of Isoprene with Natural Gobi Dust, ACS Earth Space Chem., 1, 236–243, https://doi.org/10.1021/acsearthspacechem.7b00050, 2017.
- Zielinski, A. T., Kourtchev, I., Bortolini, C., Fuller, S. J., Giorio, C., Popoola, O. A. M., Bogialli, S.,
  Tapparo, A., Jones, R. L., and Kalberer, M.: A new processing scheme for ultra-high resolution
  direct infusion mass spectrometry data, Atmos. Environ., 178, 129–139,
  https://doi.org/10.1016/j.atmosenv.2018.01.034, 2018.
- 1309 Zogka, A. G., Lostier, A., Papadimitriou, V. C., Thevenet, F., Formenti, P., Rossi, M. J., Chen, H., and 1310 Romanias, M. N.: Unraveling the Uptake of Glyoxal on a Diversity of Natural Dusts and 1311 Surrogates: Linking Dust Composition to Glyoxal Uptake and Estimation of Atmospheric 1312 Lifetimes, ACS Earth Space Chem., 8. 1165-1178, https://doi.org/10.1021/acsearthspacechem.3c00359, 2024. 1313