

Part 1: Zonal gradients in phosphorus and nitrogen acquisition and stress revealed by metaproteomes of *Prochlorococcus* and *Synechococcus*

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Research Article

Short summary

Ocean primary production helps regulate climate through carbon cycling. The magnitude of this process is governed by the availability of nutrients, such as nitrogen, phosphorus, iron and zinc. Here, we analyse zonal gradients in microbial nutrient acquisition strategies and potential nutrient limitation in the surface subtropical Atlantic Ocean to determine how changes in nutrient resources impact marine primary productivity. Nutrient and trace metal availability, biological activity, and protein biomarker abundance are determined to infer phosphorus, nitrogen and trace metal acquisition and metabolism in two dominant picocyanobacteria, *Prochlorococcus* and *Synechococcus*. We find phosphorus stress prevails for both *Prochlorococcus* and *Synechococcus* in the western Atlantic, but that *Prochlorococcus* becomes increasingly nitrogen, iron, zinc and cobalamin stressed in the east with coincidentally lower phosphorus biomarker proteins, indicating a switch in nutrient status across the transect. Our findings provide species and ecotype level insights into nutrient acquisition and metabolism in the ocean, combining biogeochemical and biological rate measurements with discovery and targeted proteomics to understand how microbial metabolism will respond to a changing climate.

Abstract

Ocean productivity is maintained by key nutrients, including nitrogen, phosphorus and trace metals. The magnitude and stoichiometry of nutrient fluxes to the ocean is changing. Here, we investigate how natural assemblages of marine microbes in the subtropical North Atlantic respond to variation in nutrient availability along a natural zonal gradient. We measure dissolved nutrient concentrations, biological rates, and characterize the microbial proteomes of the dominant picocyanobacteria, *Prochlorococcus* and *Synechococcus*. Moving west to east, dissolved organic phosphorus (DOP) and phosphate concentrations increased, and dissolved iron decreased. *Prochlorococcus* abundance increased eastwards, whereas *Synechococcus* abundance was highest in the west. Zonal distributions of protein biomarkers representing phosphorus (PstS, PhoA, PhoX), nitrogen (P-II, UrtA, AmtB) and trace metal metabolism (related to iron, zinc and cobalt) from metaproteomes, together with rates of alkaline phosphatase activity, indicate greater phosphorus stress the west than the east for both picocyanobacteria. In the east, elevated levels of protein biomarkers for nitrogen, iron, zinc and cobalamin in *Prochlorococcus* indicate a transition to nitrogen stress and greater influence of trace metal resources. Measured responses of *Prochlorococcus* ecotypes and *Synechococcus* clades to DOP, iron and zinc additions in incubation experiments further indicate divergent regulation of uptake and acquisition of phosphorus in these of picocyanobacteria across transect, albeit with caveat on potential for differences in regulation within a genus and

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between strains. Together our findings suggest a basin-scale transition from phosphorus stress in picocyanobacteria in the west to nitrogen stress in the east.

95 1. Introduction

Marine phytoplankton have an important role in biogeochemical cycles, supporting ecosystems and regulating climate. Global net primary productivity (NPP) is underpinned by availability of key nutrient resources, such as nitrogen (N), phosphorus (P), iron (Fe) and zinc (Zn) and others. In the subtropical open ocean, surface nutrient concentrations are chronically low and often limit NPP. Enhanced stratification, induced by ocean warming, alongside changes to natural and anthropogenic supply of fixed N (Chien et al., 2016; Kim et al., 2014; Wrightson and Tagliabue, 2020), P (Barkley et al., 2019) or Fe (Liu et al., 2022) to the global ocean are likely to perturb the magnitude and ratio at which nutrients are supplied to phytoplankton (Peñuelas et al., 2013), potentially expanding or intensifying nutrient limited ocean regions (Bopp et al., 2013; Chien et al., 2016; Lapointe et al., 2021). Detecting and understanding how nutrients regulate phytoplankton distribution, growth and activity is key to estimating the magnitude and direction of contemporary and future NPP, reducing uncertainty and assessing risks to ecosystem services (Tagliabue et al., 2021).

The nutrient that limits phytoplankton growth can be identified by adding single or multiple nutrients to seawater and measuring phytoplankton growth or other properties over time (Browning and Moore, 2023; Mahaffey et al., 2014; Mills et al., 2004; Moore et al., 2008). In addition, advances in 'omics' have enabled identification of protein biomarkers related to nutrient acquisition of stress in marine phytoplankton (Chappell et al., 2012; Hawco et al., 2020; Held et al., 2020, 2025; Rouco et al., 2018; Saito et al., 2014, 2015; Ustick et al., 2021). Identifying regions of N and/or Fe stress is straightforward (Browning and Moore, 2023) but detecting P stress can be more challenging. Systems-level interpretation of incubation results and biomarker abundances is needed to disentangle the effects of nutrient biogeochemistry and biological plasticity/activity. For instance, in low-phosphate regions dominated by the ecologically important picocyanobacteria, *Prochlorococcus*, phosphate addition experiments imply a lack of P stress, whereas genomic data identifies large areas of P stress for *Prochlorococcus* (Browning and Moore, 2023). This mismatch may be due to the flexibility in P acquisition strategies demonstrated by key marine phytoplankton (Duhamel et al., 2021; Martínez et al., 2012; Martiny et al., 2006, 2009; Moore et al., 2005; Ostrowski et al., 2010; Scanlan et al., 1993; Tetu et al., 2009). In addition, phosphate limited phytoplankton can deploy an array of strategies to acquire alternative sources of P from dissolved organic phosphorus (DOP) including esters (Sebastian and Ammerman, 2009; Tetu et al., 2009), polyphosphate (Moore et al., 2005), phosphite (Martínez et al., 2012) and phosphonate (Ilikchyan et al., 2010) or substituting P-rich lipids with P-free alternatives (Van Mooy et al., 2009). A hydrolytic metalloenzyme group, alkaline phosphatases, are responsible for cleaving P from esters (Hoppe, 2003). Enhanced activity of alkaline phosphatase (AP) has been used as an indicator of P limitation (Mahaffey et al., 2014; Su et al., 2023) although the substrate specificity (Srivastava et al., 2021), cellular localisation (Luo et al., 2009), AP allocation between ecotypes (Moore et al., 2005), uncertainty in the contribution of different phytoplankton groups to total enzyme activity (Held et al., 2025; companion study to this manuscript) and lack of knowledge on the efficiency of

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255 different AP enzymes raises [uncertainties](#). Collectively, the flexibility in P acquisition strategies, as well as the
perceived ability of *Prochlorococcus* to readily satisfy their P demands at ultra-low concentrations of phosphate
(Lomas et al., 2014) has led to the idea that *Prochlorococcus* evade nutrient stress, particularly by remodelling
their proteomes.

260 Comparing the physiological response of two ecologically important picocyanobacteria, *Prochlorococcus* and
Synechococcus, to P stress demonstrates the complexity of deciphering resource limitation in mixed populations,
between species, or even between strains of the same species. *Synechococcus* possess genes encoding a high
affinity periplasmic phosphate binding protein (pstS) and transport system (pstABC), as well as genes encoding
proteins essential for accessing organic P via alkaline phosphatase (phoA) and phosphonatase (phnC, D, E,
265 (Moore et al., 2005; Scanlan et al., 1993; Tetu et al., 2009). When phosphate is scarce, *Synechococcus* has been
shown to upregulate pstS, pstABC and phoA (Moore et al., 2005; Tetu et al., 2009), [the regulator gene ptrA](#)
(Ostrowski et al., 2010) [and the recent described high affinity AP gene psip1 \(in clade III only, Torcello-Requena](#)
et al., 2024), with a measurable increase in AP activity (Moore et al., 2005, [Torcello-Requena et al., 2024](#)),
implying that expression of these genes is indicative of P [stress](#) (Moore et al., 2005, [Torcello-Requena et al.](#)
270 [2024](#)). However, clade specific variations in response to phosphate limitation have been observed *in situ* (Sohm
et al., 2016, [Torcello-Requena et al. 2024](#)) and in culture (Moore et al., 2005). While *Prochlorococcus* also
possesses pstS and pstABC and has been shown to upregulate these genes alongside phoA under phosphate
deplete conditions (Martiny et al., 2006), strain specific variations in its ability to access organic P also exist. For
example, while the two most prevalent high light (HL) clades, MED4 (HLI) and MIT9312 (HLII) can grow
275 solely on phosphate, MED4 grows on a wider range of organic P compounds, [possess a high affinity AP \(psip1,](#)
[Torcello-Requena et al., 2024](#)) and dramatically increases AP activity when P starved compared to MIT9312
(Moore et al., 2005).

[In addition to species and clade specific responses across the microbial realm, AP enzymes are dependent](#)
[on a metal co-factor, with Zn and/or cobalt \(Co\) required for the protein PhoA](#) (Coleman, 1992) [and Fe and](#)
280 [calcium for the proteins PhoX and PhoD](#) (Rodriguez et al., 2014; Yong et al., 2014) [and Psip1](#) (Torcello-Requena
et al., 2024). [Although, 130 the active sites of PhoA and PhoX in marine microbes have yet to be biochemically](#)
[characterised, their metal requirements have been estimated assuming they are like the model organism,](#)
[Escherichia coli](#) and based on supporting evidence that the enzymes respond to the metals that they are expected
[to contain](#) (Cox and Saito, 2013; Mikhaylina et al., 2022; Ostrowski et al., 2010). [However, homology-based](#)
285 [annotation of enzymes is challenging and therefore the annotations herein should be considered putative. The](#)
[trace-metal content of these proteins](#) creates the potential for trace metals [to control P](#) acquisition via regulation
of AP activity [leading to](#) Fe-P or Zn-P co-limitation (Browning et al., 2017; Duhamel et al., 2021; Held et al.,
2025; Mahaffey et al., 2014). Observations of an accelerating stoichiometry of Co in the western North Atlantic
has led to hypotheses for the potential for Co use in oceanic alkaline phosphatases too (Held et al., 2025; Jakuba
290 et al., 2008; Saito et al., 2017). In culture studies, *Prochlorococcus* and *Synechococcus* have been shown to have
absolute requirements for Co but not Zn under replete P conditions (Hawco et al., 2020; Saito et al., 2002; Sunda
and Huntsman, 1995) but *Synechococcus* benefits from available Zn to produce AP under P scarcity (Cox and

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Saito, 2013). Thus, knowledge of the phytoplankton community structure, alongside their nutritional preferences and enzyme characteristics is key in deciphering nutrient limitation in the ocean.

This study measures the biological response to nutrient transitions in the North Atlantic Gyre. Here, the Western basin is heavily influenced by Saharan aeolian dust (Jickells, 1999), while the eastern basin borders the upwelling system off northwest Africa (Menna et al., 2015). Both upwelling and dust deliver scarce resources to the region, creating strong gradients in nutrients and trace metals (Gross et al., 2015; Kunde et al., 2019; Reynolds et al., 2014; Sebastián et al., 2004) influencing productivity (Moore et al., 2008). DOP dynamics (Liang et al., 2022) and marine dinitrogen (N₂) fixation (Moore et al., 2009). Here, we exploit these strong natural gradients in nutrient and trace metal resources and biological activity to investigate nutrient acquisition strategies of natural assemblages of *Prochlorococcus* and *Synechococcus*.

Alongside measurements of biogeochemical states, specifically nutrients, dissolved iron, zinc, cobalt and DOP and biological rates, including AP activity and N₂ fixation, we investigated biological activity with non-targeted metaproteomics and quantitative targeted proteomics of the high affinity phosphate binding protein, PstS, and two alkaline phosphatases, PhoA and PhoX in *Prochlorococcus* and *Synechococcus* (Table 1). From the non-targeted metaproteomics analyses we specifically focus on proteins indicative of N acquisition (P-II, UrtA, AmtB) and proteins involved in iron (ferredoxin), zinc (zinc peptidase and transporter) and B₁₂ (cobalamin synthetase) metabolism (Table 1). This allowed us to firstly investigate the potential for *Prochlorococcus* and *Synechococcus* to be phosphorus-stressed in the subtropical Atlantic, challenging the view that avoidance of P limitation and hypothesised zonal gradients in proteins would reflect nutrient stress. We also assessed the potential for N, Fe and Zn to control the zonal distribution of *Prochlorococcus* and *Synechococcus*. Secondly, we assessed the potential for P acquisition to be regulated by the availability of DOP, Fe and Zn or Co. We hypothesised that the distribution of PhoA and PhoX would be reflected in rates of AP and alongside Fe and Zn, the limiting trace metal. We augmented *in-situ* sampling with nutrient bioassays, complimentary to those reported by Held et al., 2025 (companion manuscript), to further assess the potential for DOP substrate, alongside metals Fe and Zn to regulate AP activity and applied a quantitative proteomic approach targeting PstS, PhoA and PhoX only. Finally, we critically assessed our different approaches to delineate nutrient controls of the distribution and physiological strategies of *Prochlorococcus* and *Synechococcus*, highlighting the nuanced insights gained when bringing together biogeochemical measurements alongside ‘omics (Saito et al., 2024).

Table 1. Summary of the proteins targeted by metaproteome (all) and quantitative (*) protein analysis including their function and known characteristics.

Protein name or family	Function and reported characteristics
PstS*	Periplasmic phosphate-binding protein. Induced under P-limiting conditions
PhoA*	Alkaline phosphatase: cleaves phosphorus from organic compounds. Zinc metalloenzyme Induced under P-limiting conditions
PhoX*	Alkaline phosphatase: cleaves phosphorus from organic compounds. Iron metalloenzyme. Regulation unknown
P-II	Nitrogen regulatory protein. Indirectly controls the transcription of glutamine synthetase gene <i>glnA</i> .
AmtB	Ammonium transporter channel. Transmembrane
UrtA	An ABC-type, high-affinity urea permease. Substrate binding protein

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Ferredoxin	Iron metalloenzyme. Regulated by iron, more abundant under high iron conditions.
Zinc peptidase	Zinc metalloenzyme. Involved in proteolysis at the plasma membrane
Zinc transporter	Zinc metalloenzyme. ABC transporter, ATP-binding protein
Cobalamin synthetase	Cobalt metalloenzyme. Synthesis of cobalamin (vitamin B ₁₂)

355 2 Materials and methods

2.1 Sample collection from surface waters

360 [Samples were collected on a zonal transect between Guadeloupe and Tenerife at approx. ~22°N between 26th June and 12th August 2017](#) onboard the *RRS James Cook* (JC150, Fig. 1a). Sea surface temperature (SST) was measured via the underway seawater system using Seabird sensors. Using a trace-metal clean towed FISH and a Teflon diaphragm pump (Almatec A-15), seawater samples were collected every 2 h, at a resolution of ~ 25 km, from ~ 3 m below the surface (Fig. 1a), with seawater flow terminating into a class-100 clean air-laboratory.

2.2 Biogeochemical states and rates

Using unfiltered seawater samples from the towed FISH, concentrations of nitrate plus nitrite (Brewer and Riley, 1965), phosphate (Kirkwood 1989) and ammonium (Jones, 1991) were analysed onboard according to GO-SHIP nutrient protocols (Becker et al., 2020). Using filtered seawater from the towed FISH (Sartobran, Sartorius, 0.8/0.2 µm polyethersulfone membrane), concentrations of dissolved iron (Kunde et al., 2019) were measured onboard while concentrations of dissolved zinc (Nowicki et al., 1994) were determined at the University of Southampton. Concentrations of DOP were determined at the University of Liverpool using a modified version of (Lomas et al., 2010) as described by (Davis et al., 2019). Using unfiltered seawater from the towed FISH, rates of alkaline phosphatase were determined onboard every 4 h or ~ 50 km; Davis et al., 2019). *Prochlorococcus*, *Synechococcus* (or *Parasynechococcus*, Coutinho et al., 2016) and high and low nucleic acid bacteria (HNA and LNA, respectively) were enumerated every 2h at Plymouth Marine Laboratory using flow cytometry (Tarran et al., 2006). Surface ocean concentrations of chlorophyll *a* (on GF/F) were determined on every sample (Welschmeyer, 1994). Concentrations of dissolved cobalt were measured in separate samples collected from 40m from 4 stations only using high resolution inductively coupled plasma mass spectrometry (HR-ICP-MS), preceded by UV-digestion and off-line preconcentration into a chelating resin (WAKO) at the University of Southampton (Lough et al., 2019; Rapp et al., 2017).

380 2.3 Global metaproteomic analysis

At 7 stations, McLane pumps were deployed to 15 m (see Table S1 for deployment details). Data from station 1 was omitted from this study due to significant riverine influence (Kunde et al., 2019). Pumps were fitted with a trace metal clean mini-MULVS filter head. Between 17 and 359 L of seawater was filtered through a 51 µm (Nitex), 3 µm (Versapor) and 0.2 µm (Supor) filter stack. Filters were immediately frozen at -80°C, with subsequent transportation and storage at -80°C. Protein biomarker analysis was conducted on the 0.2 µm filter, representing the 0.2 to 3 µm particle fraction. Briefly, upon return to the laboratory, the total microbial protein was extracted using a detergent based method. The filter was unfolded and placed in an ethanol rinsed tube, then

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covered in 1 % SDS extraction buffer (1 % SDS, 0.1M Tris HCl pH 7.5, 10 mM EDTA), incubated at room temperature for 10 mins, then at 95 °C for 10 mins, and then shaken at room temperature for 1 h. The extract was
395 decanted and clarified by centrifugation before being concentrated by 5 kD membrane centrifugation to a small volume, washed in extraction buffer, and concentrated again. The total protein concentration was determined by BCA assay (kit) at this time. The proteins were precipitated in cold 50 % methanol 50 % acetone 0.5 mM HCl at 20 °C for one week, collected by centrifugation at 4 °C, and dried by vacuum. Purified protein pellets were resuspended in 1% SDS extraction buffer and redissolved for 1 h at room temperature. Total protein was again
400 quantified by BCA assay to assess recovery of the purification.

Extracted proteins were immobilized in a small volume polyacrylamide tube gel using a previously published method (Lu and Zhu, 2005; Saito et al., 2014). LC-MS/MS grade reagents were used and all tubes were ethanol rinsed. The gels were fixed in 50 % ethanol, 10 % acetic acid, then cut into 1mm cubes and washed in 50:50 acetonitrile: 25 mM ammonium bicarbonate for 1 h at room temperature, then washed again in the same
405 solution overnight. Next, the gels were dehydrated by acetonitrile treatment before protein reduction by 10 mM dithiothreitol treatment at 56 °C for 1 h with shaking. Gel pieces were rinsed in 50:50 acetonitrile: ammonium bicarbonate solution, then proteins were alkylated by treatment with 55 mM iodoacetamide at room temperature for 1 h with shaking. Gels were again dehydrated by acetonitrile treatment and dried by vacuum. Finally, proteins were digested by treatment with trypsin gold (Promega) prepared in 25 mM ammonium bicarbonate at the ratio
410 of 1:20 µg trypsin: µg total protein overnight at 37 °C with shaking. The next morning, any supernatant was decanted into a clean microfuge tube, and 50 µL protein extraction buffer (50 % acetonitrile, 5 % formic acid in water) was added to the gels, incubated for 20 mins, centrifuged and collected. The extraction was repeated and combined with the original supernatant. Peptides were concentrated to approximately 1 µg total protein per µL solution by vacuum at room temperature. 10 µL or 10 µg were injected per analysis.

Global metaproteome analysis, which is conducted with no prior determined targets, was performed in
415 Data-Dependent-Acquisition (DDA) mode using Reverse Phase Liquid Chromatography – active modulation – Reverse Phase Liquid Chromatography Mass Spectrometry (RPLC-am-RPLC-MS) (McIlvin and Saito, 2021). RPLC-am-RPLC-MS involves two orthogonal chromatography steps, which are performed in-line on a Thermo Dionex Ultimate 3000 LC system equipped with two pumps. The first separation was on a PLRP-S column
420 (200 µm × 150 mm, 3 µm bead size, 300 Å pore size, NanoLCMS Solutions) using an 8 h pH 10 gradient (10 mM ammonium formate and 10 mM ammonium formate in 90% acetonitrile), with trapping and elution every 30 mins onto the second column. The second separation occurred in 30 min intervals on a C18 column (100 m × 150 mm, 3 µm particle size, 120 Å pore size, C18 Reprosil-God, Maisch, packed in a New Objective PicoFrit column) using 0.1% formic acid and a 0.1% formic acid in 99.9% acetonitrile. The eluent was analyzed
425 on a Thermo Orbitrap Fusion mass spectrometer with a Thermo Flex ion source. MS1 scans were monitored between m/z 380 and 1,580, with an m/z 1.6 MS2 isolation window (CID mode), 50 ms maximum injection time and 5 s dynamic exclusion time.

Resulting spectra were searched in Proteome Discoverer 2.2 with SequestHT using a custom DNA sequence database consisting of over 30 genomes from cyanobacteria isolates and metagenomic data from the
430 Pacific and Atlantic oceans (including metagenomes from Metzyme and Geotraces cruise GA03). Annotations were derived using BLASTp against the NCBI non-redundant protein database. The corresponding protein

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435 FASTA file is available with the raw mass spectra files (see Supplement A). SequestHT parameters were set to
methionine (+16.0) and cysteine oxidation allowed. Protein identifications were made using Protein Prophet in
440 Scaffold (Proteome Software) at the 95 % peptide confidence level, resulting in <1 % protein and peptide FDRs.
Details of the peptides identified relative to protein name and organism can be found in Table S2 and the protein
report and analytical details can be found in Supplement B.

440 Global metaproteome protein abundances are reported in normalized spectral counts. The normalization
is performed by summing the total number of spectra in each sample, calculating the average number of spectra
across all the samples, and then multiplying each spectrum count by the average count over the sample's total
spectral count. This is done to control for small differences in the amount of sample injected into the mass
spectrometer.

445 **2.4 Quantitative proteomics analysis**

A small number of tryptic peptides were selected for absolute quantitative analysis in the samples from nutrient
addition experiments (see section 2.5 for details) and were analysed as described in detail by Held et al., 2025
(companion manuscript). The amino acid sequence for the protein biomarkers quantified in this study (PstS,
PhoA, PhoX) for *Prochlorococcus* and *Synechococcus* are summarised in Table S3 and peptide report and
450 analytical details are found in Supplement C.

2.5 Nutrient bioassay experiments

Trace-metal clean sampling and incubation protocols used to setup onboard bioassays are described in detail in
the Supplement D. Aliquots of Fe, Zn and Co solutions were added to unfiltered seawater to investigate metal
455 limitation of alkaline phosphatase and results are reported in Held et al., (submitted). Alongside these
experiments, we added DOP alone or with Fe and Zn to investigate the potential for organic P availability to
influence AP activity at stations 2 and 3 only, where concentrations of DOP were low (< 80 nM, Fig. 1a and e,
Table S4), and the results are reported here. Trace-metal clean 20L carboys were triple rinsed with unfiltered
seawater collected from 40m (to avoid contamination from the ship) via the FISH and filled and amended
460 accordingly (Table S4). At the start and end of 48 hours, we measured phytoplankton biomass (chlorophyll a,
abundance of *Prochlorococcus*, *Synechococcus*) and AP activity. After 48 hours, we collected samples to
quantify protein concentration (PstS, PhoA and PhoX) as described in section 2.4 (Table S3). Incubations were
conducted in triplicate. However, due to the biomass (therefore volume) required for protein analysis, we were
unable to collect samples from three incubation bottles for further analyses. Instead, all measurements were
465 collected from two incubation bottles, except aliquots for determination of AP, which was collected from three
incubation bottles. To compare the change in states or rates in treatments relative to the control, we considered a
significant change in a property to occur when the mean of the property in the amended incubation was 2-times
higher (or lower) than the mean control incubation. Incubations were conducted in a temperature controlled
container set to a temperature measured at 40m (between 25 and 27°C) and with 12:12h light:dark cycle
470 simulated by LED light panels (Part no: LED-PANEL-300-1200-DW and LED-PANEL-200-6-DW, Daylight
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3. Results and Discussion

3.1. Zonal trends in **nutrients, cell abundance and biological rates**:

Strong zonal gradients were evident in surface temperature and phosphorus concentrations. From west to east, SST decreased by ~ 3 °C (Fig. 1b), phosphate increased by ~ 15 nM (Fig. 1c) and DOP increased 3-fold (from ~ 50 nM to ~ 150 nM, Fig. 1d). By comparison, there were no clear zonal trends in fixed nitrogen, with concentrations of nitrate plus nitrite (N+N, herein nitrate) ranging from < 10 nM to ~ 40 nM (Fig. 1e) and ammonium, which ranged from 3 to 21 nM, being highest at stations 5 and 6 (Fig. 1f).

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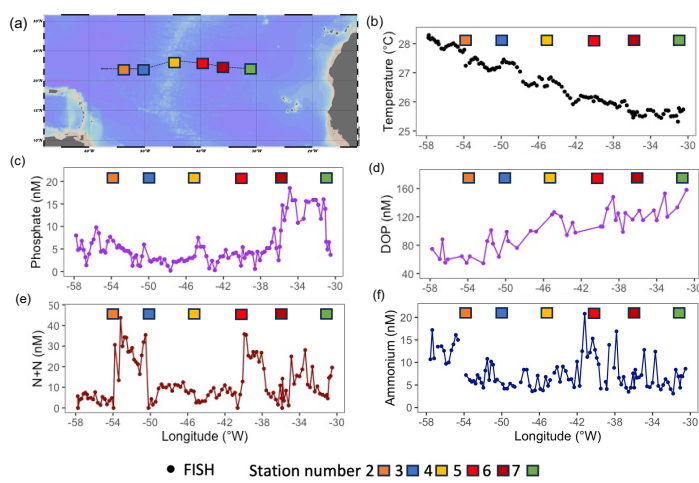


Figure 1. (a) Locations sampled during JC150 from the trace metal clean towed FISH (black circles) and stations (coloured squares) and surface ocean properties including (b) sea surface temperature (°C), (c) phosphate (nM), (d) dissolved organic phosphorus (DOP, nM), (e) nitrate+nitrite (N+N, nM), (f) ammonium (nM). Note that data from JC150 Station 1 (test station) has not been included in this manuscript due to the strong riverine influence (Kunde et al., 2019). Map produced using Ocean Data View (ODV).

There was a clear zonal trend in dissolved Fe concentrations, which decreased west to east by ~ 1.0 nM (Fig. 2a) owing to enhanced Saharan dust deposition in the western Atlantic Ocean (Kunde et al., 2019). In contrast, Zn concentrations were variable throughout the transect (ranging from 0.04 to 0.8 nM, Fig. 2b) and cobalt was constant (~ 11 pM to 14 pM, data not shown).

Microbial biomass, picocyanobacteria abundance and biological rates exhibited strong zonal gradients. From west to east, there were increases in chlorophyll *a* concentration (Fig. 3a) and *Prochlorococcus* cell abundance (Fig. 3b) whereas *Synechococcus* cell abundance decreased (Fig. 3c). HNA and LNA bacterial abundance (Fig.

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Deleted: SST decreased from ~ 28 °C in the west to ~ 25 °C in the east (Fig. 1b). In the upper 10 m, phosphate increased from ~ 5 nM to 20 nM from west to east (Fig. 1c), whereas nitrate plus nitrite (N+N, herein nitrate) ranged from < 10 nM to ~ 40 nM with no clear zonal trend (Fig. 1d). Ammonium concentrations ranged from 3 to 21 nM, with the highest concentrations observed between stations 5 and 6 (Fig. 1f). From west to east, DOP increased 3-fold (from ~ 50 nM to ~ 150 nM, Fig. 1e) alongside a 4-fold decrease in chlorophyll corrected AP activity (from > 2000 nmol P µg chl *a* d⁻¹ to < 500 nmol P µg chl *a* d⁻¹, Fig. 1g). Concurrent zonal gradients of phosphate, DOP and AP activity supports previous findings (Mahaffey et al., 2014) of an increase in AP activity as phosphate decreases (Fig. S1a), driving a decline in DOP (Fig. S1b).

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Deleted: Dissolved Fe concentrations decreased from ~ 1.4 nM in the west to ~ 0.4 nM in the east (Fig. 2a), with higher Fe in the west due to enhanced dust deposition from the Saharan dust source that is transported to the western Atlantic Ocean (Kunde et al., 2019). Zn was variable throughout the transect, ranging from 0.04 to 0.8 nM (Fig. 2b). Cobalt was measured at 40 m and at 4 stations only and ranged from 11 pM at stations 2 and 3 to 13-13.9 pM at stations 4 and 7 (data not shown).

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3d and e, respectively) also increased from east to west. There was also a 4-fold decrease in chlorophyll corrected APA (from $> 2000 \text{ nmol P } \mu\text{g chl } a \text{ d}^{-1}$ to $< 500 \text{ nmol P } \mu\text{g chl } a \text{ d}^{-1}$, Fig. 3f), likely in response to the observed gradient in P/DOP availability (Mahaffey et al., 2014, Fig. S1a, S1b).

In addition, the abundance of key diazotrophs *Trichodesmium* and UCYN-A increased from west to east (Cerdan-Garcia et al., 2022). Although rates of N_2 fixation in the east exceeded those in the west (3 to 10 nM d^{-1} and $< 3 \text{ nM d}^{-1}$, respectively), the highest rates were in the central transect between stations 4 and 5 (12 to 18 nM d^{-1} , Fig. 3g).

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Deleted: Chlorophyll *a* concentration increased from $\sim 0.05 \mu\text{g L}^{-1}$ to $\sim 0.15 \mu\text{g L}^{-1}$ from west to east (Fig. 3a). *Prochlorococcus* cell abundance increased from $\sim 5 \times 10^4 \text{ cells mL}^{-1}$ in the west to $2.5 \times 10^5 \text{ cells mL}^{-1}$ in the east (Fig. 3b), whereas *Synechococcus* cell abundance decreased from $\sim 8 \times 10^3 \text{ cells mL}^{-1}$ in the west to $1 \times 10^3 \text{ cells mL}^{-1}$ (Fig. 3c). Both HNA and LNA bacterial abundance increased ~ 2 -fold from west to east (Fig. 3d and e, respectively). Rates of N_2 fixation were highest between stations 4 and 5 (12 to 18 nM d^{-1}) and were elevated in the east (3 to 10 nM d^{-1}) compared to the west ($< 3 \text{ nM d}^{-1}$) (Fig. 1g). In addition, there was an increase in the abundance of key diazotrophs *Trichodesmium* and UCYN-A in the east relative to the west (Cerdan-Garcia et al., 2021).

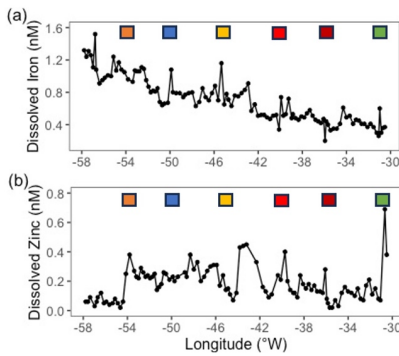
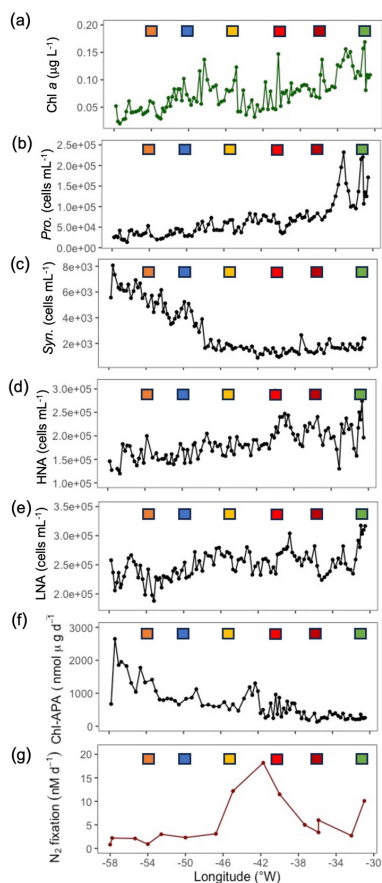


Figure 2. Zonal gradients in (a) dissolved iron concentrations (nM, from Kunde et al., 2019) and (b) dissolved zinc concentrations (nM). Samples captured from the towed FISH at $\sim 7\text{m}$. Coloured square represent stations sampled during JC150 (see Fig. 1 for station names).



565 Figure 3. Zonal gradients in (a) chlorophyll a concentrations ($\mu\text{g chl L}^{-1}$) and the abundance of (b)
Prochlorococcus (cells mL^{-1}), (c) *Synechococcus* (cells mL^{-1}), (d) high nucleic acid bacteria (HNA, cells mL^{-1})
 and (e) low nucleic acid bacteria (LNA, cells mL^{-1}). (f) chlorophyll a – corrected rates of alkaline phosphatase
 (nmol P mg chl d^{-1}) and (g) mean rates of dinitrogen (N_2) fixation (nM N d^{-1}) with error bars as standard
 deviation of triplicate incubations. Samples captured from the towed FISH at ~7m. Coloured squares represent
 570 stations sampled during JC150 (see Fig. 1 for station names).

575 These zonal gradients in hydrography, nutrients, biological rates and picocyanobacteria create two
 contrasting regions – one in the west (west of 46°W or west of station 4) and one in the east (east of 46°W or east
 of station 4). Thus, quantitative comparisons of key characteristics can be drawn between Station 2 at 54°W and
 Station 7 at 31°E (Fig. 1a, Table 2). Compared to the east, conditions in the west were characterized by notably

580 higher dissolved Fe and ammonium concentrations (3 to 4-fold higher), APA (4-fold higher) and *Synechococcus*
abundance (2-fold higher). In contrast, the east was characterized by relatively high phosphate, DOP, chlorophyll
585 *a*, *Prochlorococcus* abundance, rates of N₂ fixation, *Trichodesmium* and UCYN-A abundances (Table 2).

590 Based on these biogeochemical parameters, the phosphate-binding protein PstS, which is expressed
under P-limiting conditions, would be expected to be prevalent throughout the transect consistent with low
585 phosphate concentrations across the entire transect. Protein biomarkers would also be expected to indicate higher
alkaline phosphatase (AP) abundances in the west, corresponding with the observed trends in APA. In addition,
PhoX would be expected to be prevalent in the Fe-rich west, with greater prevalence of Fe-stress biomarkers in
the east.

Properties higher in the west (-fold)	Properties higher in the east (-fold)
Iron (3) [*]	Phosphate (4) [*]
Ammonium (4) [*]	DOP (3) [*]
APA (4) [*]	Chlorophyll (2) [*]
V _{max} /K _m (5) [†]	<i>Prochlorococcus</i> (6) [*]
<i>Synechococcus</i> (2) [*]	N ₂ fixation rates (3) ^{**}
	<i>Trichodesmium</i> (2) [†]
	UCYN-A (71) [†]
<i>Prochlorococcus</i> -Phosphate binding protein, PstS (2) [†]	<i>Prochlorococcus</i> - Nitrogen regulatory protein, PII (1.3) [†]
<i>Prochlorococcus</i> -alkaline phosphatase, PhoA (7) [†]	<i>Prochlorococcus</i> - Ammonium transporter, AmtB (1.7) [†]
<i>Synechococcus</i> -alkaline phosphatase, PhoA (29) [†]	<i>Prochlorococcus</i> -Urea permease, UrtA (1.6) [†]
SAR11-alkaline phosphatase, PhoA (24) [†]	<i>Prochlorococcus</i> -Ferredoxin (9) [†]
Total <i>Synechococcus</i> protein (1.3) [†]	<i>Prochlorococcus</i> -Zinc peptidase (1.3) [†]
	<i>Prochlorococcus</i> -Zinc transporter (4) [†]
	<i>Prochlorococcus</i> - Cobalamin synthetase (5) [†]
	SAR11 - alkaline phosphatase, PhoX (4) ^{††}
	Total <i>Prochlorococcus</i> protein (1.6) [†]

595 Table 2. Summary of states, rates and protein biomarkers that are higher in the west (left hand column) or east
(right hand column) of the transect. The numbers in brackets represent the approximate -fold difference between
west and east. Properties not reported (e.g. dissolved zinc, *Syn*-UrtA) displayed no clear difference between west
and east. We note if the differences in properties are statistically significant (*, $p < 0.05$) or not significant (**, $p > 0.05$).[†] indicates insufficient replication or measurements for statistical analysis.

600 3.2 Zonal gradients in phosphorus acquisition proteins

3.2.1. *Prochlorococcus*

605 Zonal gradients in *Prochlorococcus* P proteins generally indicate more severe P stress in the west.
Prochlorococcus (HLII) specific P proteins PstS and PhoA (*Pro*-PstS and *Pro*-PhoA, respectively) were almost
2-fold and 7-fold higher in the west relative to the east (Fig. 4a, Table 2), whereas there was no clear zonal trend
in PhoX (*Pro*-PhoX, Fig. 4a). Similar zonal trends for PstS (Fig. S2a), PhoA (Fig. S2b) and PhoX (Fig. S2c)
were observed irrelevant of the strain or ecotype of *Prochlorococcus*, thus reflecting true biological regulation
within the entire *Prochlorococcus* community, rather being contingent on variation in the abundance of one
clade/strain across the transect. Moreover, the increase in total *Prochlorococcus* protein (Fig. 4d) alongside
Prochlorococcus cell abundance (Fig. 3b) in the east suggests that trends in untargeted metaproteomics analysis

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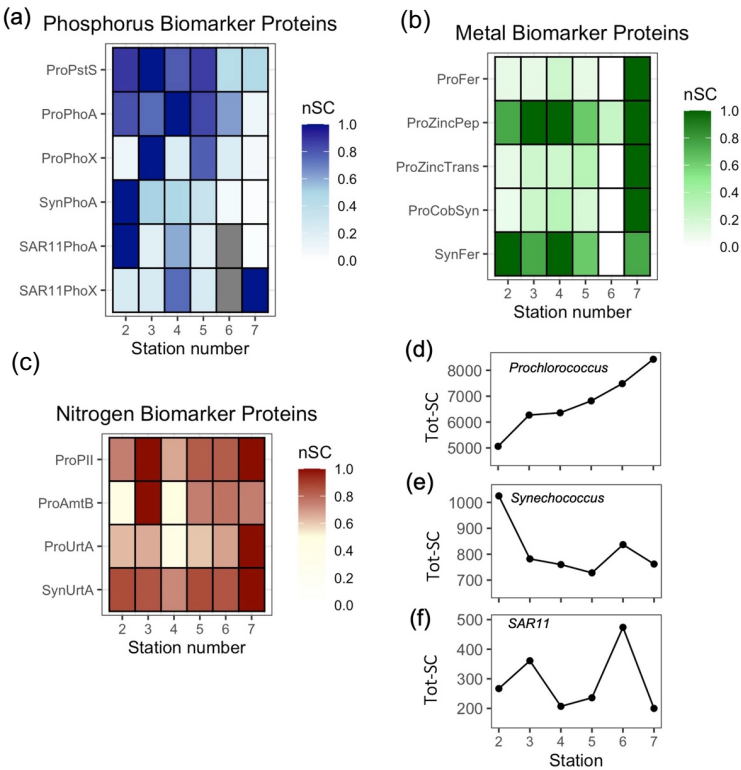
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are representative of microbial community structure. Thus, assuming all observed *Prochlorococcus* cells possess both genes, the higher *Pro*-PstS and *Pro*-PhoA in the west, where *Prochlorococcus* abundance was lower, reflects a physiological response to low phosphorus availability.

Deleted: Note that total *Prochlorococcus* protein (reported as total spectral counts, Fig. 4d) agreed with the zonal trend *Prochlorococcus* cell abundance (Fig. 3b) suggesting that untargeted metaproteomics analysis can capture trends in microbial community structure. Thus, higher *Pro*-PstS and *Pro*-PhoA in the west where there was lower *Prochlorococcus* reflects a physiological response to the nutrient environment rather than reflecting changes in biomass.

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Figure 4. Zonal gradients in the spectral counts (SC) of biomarker proteins in *Prochlorococcus* (Pro-), *Synechococcus* (Syn-) and SAR11 for (a) Phosphorus biomarker proteins; PstS, PhoA and PhoX, (b) Iron, zinc and cobalt biomarker proteins; Ferredoxin (Fd) and Zinc peptidase (ZincPep), Zinc transporters (ZincTrans) and Cobalamin Synthetase (CobW) (c) Nitrogen biomarker proteins: PII, AmtB and UrtA and (d) total protein for *Prochlorococcus*, (e) *Synechococcus* and (f) SAR11, presenting an independent measure of biomass. See Table 1 for details of the protein functions. nSC represents normalized spectral counts, which represents the spectral counts normalized to the maximum value of each protein across 6 stations. Tot-SC represents the sum of all normalized spectral counts for *Prochlorococcus*, *Synechococcus* or SAR11

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Correlations between *Prochlorococcus* abundance and other measured parameters also indicate a physiological response to nutrient availability. *Prochlorococcus* cell abundance was negatively correlated with *Pro*-PstS (Fig. 5a), *Pro*-PhoA (Fig. 5b) and APA (Fig. 5d). APA was also positively correlated with *Pro*-PstS and *Pro*-PhoA (Fig. 5f and g). Conversely, DOP concentration was positively correlated with *Prochlorococcus* cell abundance (Fig. 5c) but negatively correlated with *Pro*-PstS (Fig. 5e). Together these data suggest *Prochlorococcus* in the west were more P stressed than those in the east.

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Deleted: *Prochlorococcus* cell abundance (and in turn, total *Prochlorococcus* protein) was negatively correlated with *Pro*-PstS ($p=0.03$, Fig. 5a), *Pro*-PhoA ($p=0.035$, Fig. 5b) and APA ($p=0.001$, Fig. 5d) and positively correlated with DOP ($p=0.007$, Fig. 5c). *Pro*-PstS was negatively correlated with DOP ($p=0.03$, Fig. 5e) and both *Pro*-PstS and *Pro*-PhoA were positively correlated with AP activity ($p=0.013$ and $p=0.057$, respectively, Fig. 5f and g). In agreement with biogeochemical signatures for P stress, proteins, *Pro*-PstS and *Pro*-PhoA were more prevalent in the west (Fig. 4a, Table 2) and significantly positively correlated to AP activity (Fig. 5f and g). However, *Prochlorococcus* abundance was lower in the west compared to the east (Fig. 3b), and negatively related to overall AP activity, *Pro*-PstS and *Pro*-PhoA and positively related to DOP concentrations. Together these data imply increased P stress of *Prochlorococcus* in the west compared to the east. This was again evident in greatly increased P-stress biomarkers in the west, despite decreasing *Prochlorococcus* abundance, demonstrating the change in biomarkers is not simply due to changing biomass. ¶

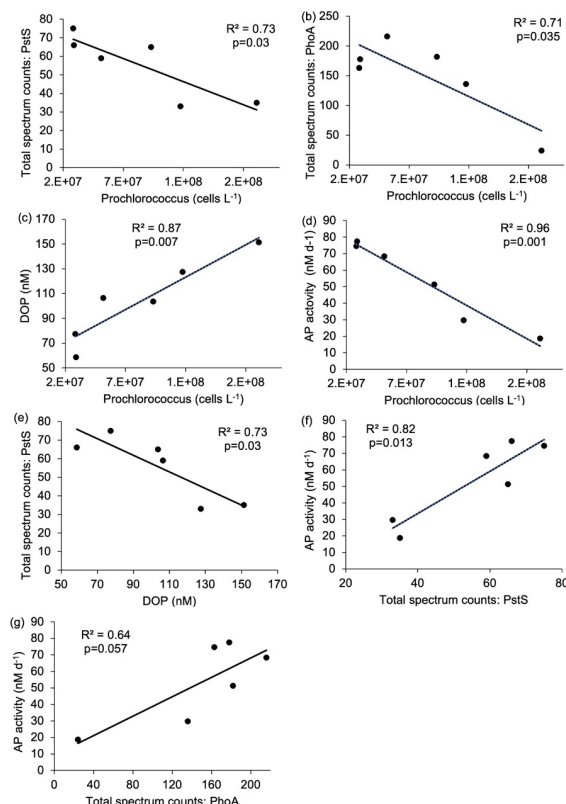


Figure 5. Relationship between (a) *Prochlorococcus* cell abundance (cells L⁻¹) and *Pro*-PstS (total spectrum counts), (b) *Prochlorococcus* cell abundance (cells L⁻¹) and *Pro*-PhoA (total spectrum counts), (c) *Prochlorococcus* cell abundance (cells L⁻¹) and dissolved organic phosphorus (DOP, nM), (d) *Prochlorococcus* cell abundance (cells L⁻¹) and rates of alkaline phosphatase (APA, nM d⁻¹), (e) DOP and *Pro*-PstS, (f) *Pro*-PstS and APA and (g) *Pro*-PhoA and APA. Results are linear regression as reported as R₂ value and p-value. Relationships shown in (a) to (f) are considered statistically significant as $p < 0.05$.

The bioassay experiments shed further light on the nutrient status of these communities. At Station 2, mean chlorophyll *a* increased (from 0.075 to 0.120 $\mu\text{g L}^{-1}$) after the addition of DOP alone, but with no increase in APA. Instead, DOP+Fe stimulated an increase in chlorophyll *a* (from 0.075 to 0.108 $\mu\text{g L}^{-1}$) alongside an increase in mean APA (3.03 to 9.70 nM d⁻¹, * denotes a 2-fold or greater increase relative to the control in Fig. 6a, Table S5). *Pro*-PhoX concentration more than doubled following DOP and, separately, DOP+Zn addition at Station 2 (Fig. 6a), however insufficient understanding of the controls on PhoX limits interpretation of this observation at this time. By comparison, no significant changes in chlorophyll *a* or APA were observed at Station 3 (Fig. 6b, Table S5).

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In the bioassays, DOP addition resulted in a decrease in the concentration of *Pro*-PstS and *Pro*-PhoA after the (Fig. 6a and b), implying P acquisition proteins are repressed in the presence of elevated DOP. These observations corroborate in-situ observations as *Pro*-PstS and *Pro*-PhoA both decreased to the east (Fig. 4a) where DOP and phosphate were elevated in surface waters (Fig. 1c and e). We interpret this DOP effect to be the result of DOP conversion to phosphate by alkaline phosphatase, and negative regulation of the Pho operon that controls both PstS and PhoA rather than DOP directly interacting with the regulatory system (Martiny et al., 2006). Alternatively, there may be another system that is directly regulated by DOP availability. For example, PtrA is an alternative phosphate-sensitive regulator identified in some *Synechococcus* and *Prochlorococcus* strains and that may be responsive to organic P (Ostrowski et al., 2010). However, flow cytometry-derived *Prochlorococcus* abundance declined in all experiments (Fig. 6a and b, Table S5), a common outcome for marine oligotrophs in bottle incubation experiments, and it is unclear whether the observed decline in *Pro*-PstS and *Pro*-PhoA in the bioassays was due to a physiological response to elevated DOP or a decline in *Prochlorococcus* biomass, or a combination of the two.

That said, knowledge of the dominant *Prochlorococcus* clades in the Atlantic Ocean (Johnson et al., 2006) alongside selection of protein markers to target specific clades allows us to interpret ecotype-level responses in experiments and in the biogeochemical transect (Saito et al., 2015). For example, *Prochlorococcus* HLII, the dominant clade in the oligotrophic subtropical ocean, can use ATP but no other organic P sources and minimally increases APA in response to P starvation as it lacks regulatory genes that respond to P-limitation (e.g. *ptrA*, (Moore et al., 2005). By contrast, HL1 (MED4), which possesses both regulatory genes involved in phosphorus metabolism, *phoBR* and *ptrA* (Martiny et al., 2006), can grow on a variety of organic P substrates and substantially increases AP activity when grown on organic P relative to phosphate (Moore et al., 2005), suggesting HL1 can upregulate AP in response to external organic P levels (Moore et al., 2005). In the global metaproteomes, a west to east increase in *Prochlorococcus* ecotypes HL1 (Fig. S3a) and HLII (Fig. S3b) was detected, which accompanied the increases in cell abundance and total *Prochlorococcus* protein. HL1 (MED4) is more prevalent in the eastern Atlantic (Zinser et al., 2007) and in this study, we observed an increase in the contribution of HL1 to total ecotype from 6 to 8% (Fig. S3c). The eastward increase in HL1 abundance alongside its increased plasticity to grow on a variety of organic P substrates may explain why *Prochlorococcus* abundance is higher where DOP is elevated in the eastern Atlantic.

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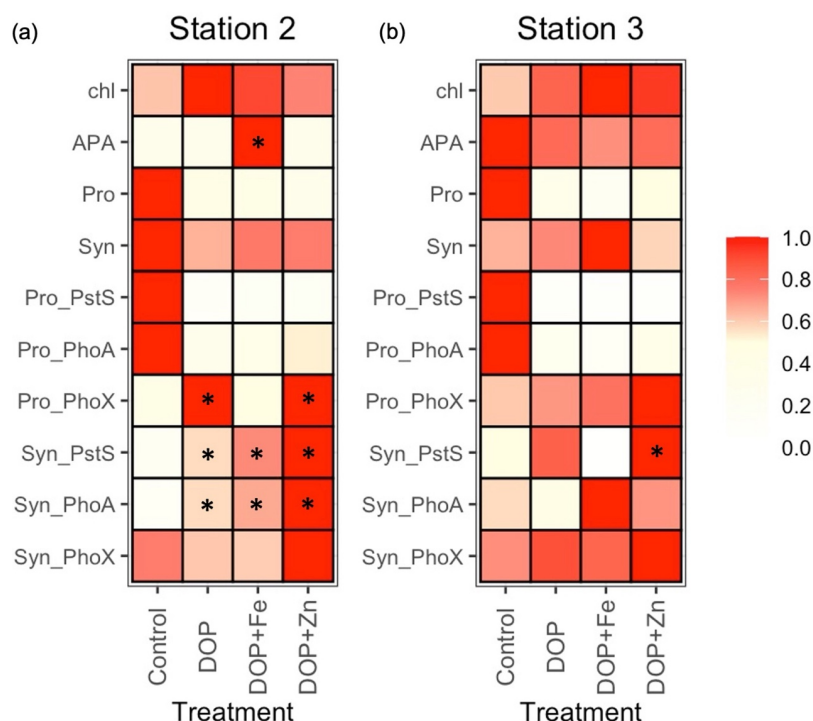


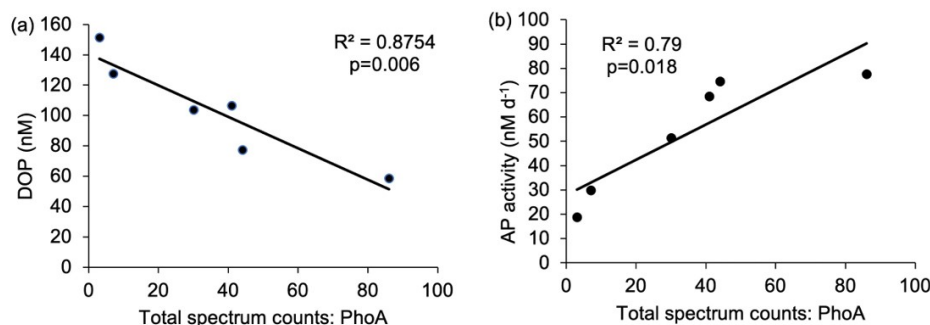
Figure 6. Fractional (scale 0 to 1) change in states, rates and individual proteins in *Prochlorococcus* (*Pro*-) and *Synechococcus* (*Syn*-) after the addition of dissolved organic phosphorus (DOP), DOP and iron (DOP+Fe) and DOP and zinc (DOP+Zn) at Station 2 (a) and Station 3 (b) for chlorophyll *a* (chl), rates of alkaline phosphatase activity (APA), *Prochlorococcus* (*Pro*), *Synechococcus* (*Syn*) and protein biomarkers PstS, PhoA and PhoX. Coloured squares represent the mean of duplicate or triplicate samples and are normalised as the fraction of the maximum of that property in each experiment. See Table S4 for a description of the experiments and Table S5 for raw data for all properties. * denotes a 2-fold or more change in the mean property relative to the control.

3.2.2. *Synechococcus*

PhoA in *Synechococcus* (*Syn*-PhoA) was 29-fold higher in the west than the east (Fig. 4a, Table 2) and significantly negatively correlated with DOP (Fig. 7a) and positively correlated with APA (Fig. 7b). Unlike *Prochlorococcus*, there was no correlation between cell abundance and proteins, DOP or AP (Fig. S4). The concentration of other *Synechococcus* P-related proteins (*Syn*-PstS and *Syn*-PhoX), were not detected in the sampled metaproteome but might have been present at concentrations below detection limits. In the west, *Synechococcus* abundance (Fig. 3c), APA (Fig. 1f), *Syn*-PhoA (Fig. 4a) and total *Synechococcus* protein count (Fig. 4e) were higher than in the east.

Deleted: (Zinser et al., 2007)(Johnson et al., 2006; Zinser et al., 2007) (Johnson et al., 2006; Zwirgmaier et al., 2007, 2008). This DOP effect was likely the result of DOP conversion to phosphate and negative regulation of the Pho operon that controls both PstS and PhoA rather than DOP directly interacting with the regulatory system. *Pro*-PhoX increased more than 2-fold after the addition of DOP alone and DOP+Zn at station 2 (Fig. 6a) but the lack of knowledge of controls on PhoX mean it is premature to interpret this observation. One caveat is that the abundance of *Prochlorococcus*, gleaned from flow cytometry, declined in all experiments and also eastward along the zonal transect (Fig. 6a and b, Table S5), meaning it is unclear if the decline in protein biomarkers, *Pro*-PstS and *Pro*-PhoA, in experiments and surface waters, was a physiological response to elevated DOP, or due to a decline in *Prochlorococcus* biomass. However, knowledge of the dominant *Prochlorococcus* clades in the Atlantic Ocean (Johnson et al., 2006) alongside selection of protein markers to target specific clades means that our ability to interpret at the ecotype level resolution both in experiments, as well as over large spatial transects is rapidly improving (Saito et al., 2015), highlighting the benefits of 'omics applications to study microbial biogeography.

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825 Figure 7. Relationship between (a) *Synechococcus* PhoA (total spectral counts) and concentrations of DOP (nM) and (b) *Synechococcus* PhoA (total spectral counts) and alkaline phosphatase activity (AP, nM d⁻¹). The R² value and p-values are reported. p<0.05 indicates that the relationship is statistically significant.

830 In the bioassays at Station 2, the addition of DOP, DOP+Fe and DOP+Zn resulted in declines in *Synechococcus* abundance by 23 to 35% relative to the control (Fig. 6a, Table S5). However, associated biomarker proteins increased. Mean concentrations of *Syn*-PstS increased by 2.7-, 3.5- and 4.7-fold after the addition of DOP, DOP+Fe and DOP+Zn, respectively, relative to the control after 48-h bioassays (Fig. 6a, Table S5). Similarly, the mean concentration of *Syn*-PhoA increased by 3.6-, 4.3- and 6.4-fold after the addition of DOP, DOP+Fe and DOP+Zn, respectively (Fig. 6a, Table S5). It is unclear why production of both *Syn*-PstS and *Syn*-PhoA was more stimulated after the addition of DOP+Fe than DOP at Station 2, assuming PhoA contains Zn or Co, and not Fe as metal co-factors (Coleman, 1992). However, replication was low (n=2) and variability between replicates was high, limiting a statistically robust interpretation.

840 At Station 3, *Synechococcus* abundance increased by 12% and 53% following DOP and DOP+Fe addition, respectively but decreased after DOP+Zn addition (Fig. 6b, Table S5). The change in protein concentration after nutrient additions was less pronounced at Station 3 than Station 2 (Table S5). DOP additions induced a 1.8-fold increase in *Syn*-PstS and a 30% decrease in *Syn*-PhoA. Addition of DOP+Fe induced a 90% decrease in *Syn*-PstS and a 1.7-fold increase in *Syn*-PhoA while addition of DOP+Zn induced a 2.1-fold increase in *Syn*-PstS and a 1.2-fold increase in *Syn*-PhoA (Fig. 6b, Table S5). There was no consistent change in *Syn*-PhoX after the addition of DOP, DOP+Zn or DOP+Fe, with *Syn*-PhoX increasing or decreasing by 20 to 40% at both stations (Fig. 6b, Table S5).

850 In-situ measurements and bioassays converge to imply that *Synechococcus* is reliant upon organic P accessed via APA. The zonal trends and bioassay results agree with culture experiments demonstrating that *Syn*-PstS and *Syn*-PhoA are produced in the presence of DOP and Zn to increase P acquisition when phosphate is low (Cox and Saito 2013). Higher APA and prevalence of *Syn*-PhoA in the low DOP and phosphate west implies that *Synechococcus* was P stressed. Enzyme kinetic bioassays indicate higher AP enzyme efficiency in the west (Fig.

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8a), ~~with enzyme efficiency~~ positively correlated with *Syn*-PhoA ($p=0.017$, Fig. 8b). ~~Thus, *Syn*-PhoA potentially governs this trend of enzyme efficiency, suggesting DOP hydrolysis was more efficient in the west than the east.~~

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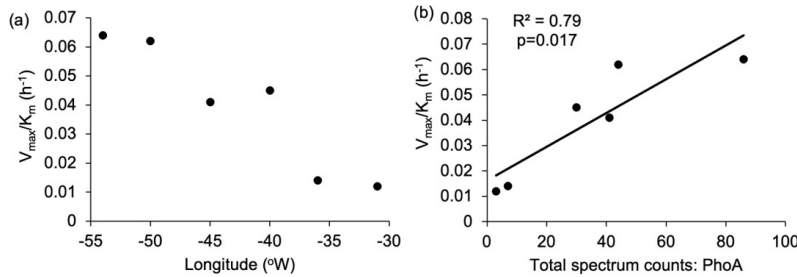


Figure 8. Enzyme efficiency for alkaline phosphatase was calculated as the ratio between V_{max} and K_m (h^{-1}); (a) zonal gradient in enzyme efficiency, indicating higher enzyme efficiency in the western compared to the eastern subtropical Atlantic and (b) positive significant ($p=0.017$) relationship between *Syn*-PhoA and enzyme efficiency.

A west-east gradient ~~was also observed for~~ PhoA in SAR11 (Fig. 4a), ~~which is~~ an abundant aerobic chemoheterotrophic alphaproteobacterial ~~contributing~~ to LNA bacterial counts (Fig. 3e). ~~The abundance of both HNA and LNA (Fig. 3d and e, respectively) and total SAR11 protein (Fig. 4f) increased from west to east.~~ However, SAR11-PhoA decreased 24-fold ~~and SAR11-PhoX increased 4-fold (Fig. 4a, Table 2) despite dissolved Fe concentrations being higher in the west relative to the east. The mechanism for this discrepancy is unclear.~~ However, PhoA is efficient at hydrolysing DOP under low P conditions and culture studies show that organic P is an important source of P for SAR11, representing up to 70% of its cellular P requirement when phosphate is non-limiting (Grant et al., 2019). Thus, we speculate that SAR11 might strategically use PhoA in the west with the zonal patterns in SAR11-PhoA and PhoX likely reflecting the preferential acquisition of DOP over phosphate. This further supports the premise that PhoA is an indicator of DOP acquisition across marine microbial taxa (Steck et al., 2025; Ustick et al., 2021).

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The decline in *Synechococcus* abundance and *Syn*-PhoA, despite the increase in DOP in the east suggests other factors, such as resource availability or competition with other microorganisms including *Prochlorococcus*, ~~inhibited growth of *Synechococcus* in the east. Zinc concentrations decreased eastwards from ~ 0.35 nM to 0.15 nM. Briefly, bioassays conducted during the same expedition indicated that Zn addition stimulated a 2- to 4-fold fold increase in *Syn*-PstS and *Syn*-PhoA after a 48-h incubation period (Held et al., 2025; companion manuscript), corroborating a role for zinc in PhoA metabolism. Cobalt also stimulated increases in *Syn*-PstS and *Syn*-PhoA, representing new field evidence for Co influencing AP (Held et al., 2025; companion manuscript). Co may effectively substitute Zn at the active site of PhoA within marine cyanobacteria, consistent with trends observed in accelerating Co stoichiometry and APase abundances in the North Atlantic Ocean (Saito et al., 2017).~~

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~~Deleted:~~ Results from bioassays would suggest that *Syn*-PhoA should increase to the east as DOP concentration increase. Instead, we observed an eastward decline in *Synechococcus* abundance and *Syn*-PhoA alongside the increase in DOP, which suggests that other factor(s) affected the P acquisition strategy and growth of *Synechococcus*. Averaged over the upper 40m, the mean Zn concentration declined from ~ 0.35 nM in the west to 0.15 nM in the east and may have limited *Syn*-PhoA activity. The dependence of PhoA on Zn in *Synechococcus* has been observed in culture with the production of PstS and PhoA only occurring at low phosphate as well as in the presence of Zn (Cox and Saito 2013). Held et al., (submitted) observed more than a 2- to 4-fold fold increase in *Syn*-PstS and *Syn*-PhoA after the addition of Zn or Co relative to the control, with the addition of Fe also stimulating an increase in *Syn*-PstS. This finding represents new field evidence for Co influencing AP (Held et al., (submitted), implying that Co may effectively substitute Zn at the active site of PhoA within marine cyanobacteria, consistent with trends observed in accelerating Co stoichiometry and APase abundances in the North Atlantic Ocean (Saito et al., 2017). Competition for P, trace metals and other resources with other plankton, including *Prochlorococcus* and N_2 fixers, may have stunted growth of *Synechococcus* in the east relative to the western subtropical Atlantic. ¶

3.3. Non-targeted metaproteomic indicators of nutrient status in picocyanobacteria

Higher PstS and PhoA in the west compared to the east, alongside the positive relationship between *Pro*-PstS, *Pro*-PhoA and *Syn*-PhoA with AP activity and negative relationship with DOP corroborate that these protein biomarkers are P-stress biomarkers in both *Prochlorococcus* (Martiny et al., 2006; Moore et al., 2005; Reistetter et al., 2013) and *Synechococcus* (Scanlan et al., 1993; Tetu et al., 2009). However, DOP addition stimulated distinct biomarker responses in *Prochlorococcus* and *Synechococcus* (Fig. 6, Table S5). For *Prochlorococcus*, DOP addition reduced *Pro*-PstS and *Pro*-PhoA but increased PhoX after 48 h relative to the control (Fig. 6, Table S5). In contrast, for *Synechococcus*, DOP addition increased *Syn*-PstS and *Syn*-PhoA, with no change in PhoX (Fig. 6, Table S5). We intuit that the protein biomarkers changed due to a physiological response rather than change in cell abundance because the per cell protein content (Fig. S5) showed the same pattern (with the caveat that the protein is clade specific yet likely targeted a major ecotype, whereas cell abundance represents all cells). DOP addition stimulated a decrease in *Pro*-PstS and *Pro*-PhoA per cell and increase *Pro*-PhoA per cell relative to the control (Fig. S5a), whereas DOP addition stimulated an increase in *Syn*-PstS, *Syn*-PhoA and *Syn*-PhoX per cell (Fig. S5b).

Either the protein regulatory pathway differs between *Prochlorococcus* and *Synechococcus*, and/or the strain specific differences in quantified proteins is complicating our interpretation response of proteins across different strains. Here we describe evidence for the former hypothesis. In *Prochlorococcus*, the Pho regulon controls P-acquisition genes such as pstS (phosphate transporter) and phoA and includes the two-component regulatory genes, phoB and phoR (Martiny et al., 2006). The *Pro*-phoX gene is controlled by the pho regulon as it sits within a genomic island with other P stress responsive (Kathuria and Martiny, 2011). In contrast, *Synechococcus* (WH8102) has a two-tiered phosphate response system, where the PhoBR regulator controls pstS using a Pho box (Cox and Saito, 2013; Tetu et al., 2009) and second regulator, PtrA, controls one of the phoA phosphatase copies, Zn transport, and various other cellular processes (Ostrowski et al., 2010). The gene neighbourhood containing phoA (SYNW2391) in *Synechococcus* is also located near efflux transporter and close to the ferric uptake regulator, Fur. To our knowledge, regulation of PhoX and its interaction with PhoA regulation in the marine picocyanobacteria is not well understood, but analysis of the gene neighbourhood in the model organism *Prochlorococcus* sp. NATL1A reveals that phoX is not within the phoA neighbourhood and is in the vicinity of a putative manganese transporter. For *Synechococcus* (WH8102), the position of phoX (SYN1799) is like *Prochlorococcus* and is located directly next to the futAB iron ABC transport system, consistent with the iron requirement of this enzyme. The separation of phoA and phoX within the genome in both *Prochlorococcus* and *Synechococcus* (at least in the representative strains described above) implies their regulation may be distinct in the different organisms. Consistent with prior observations (Browning et al., 2017; Mahaffey et al., 2014; Rouco et al., 2018), PhoA and PhoX may be regulated by both metals and phosphorus availability but the specific regulatory system in picocyanobacteria is complex and still unknown.

The alternative interpretation is that differences in strain specificity among the identified proteins explains the differences in the *Prochlorococcus* vs *Synechococcus* response patterns. Proteins concentrations are measured by

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detecting peptides, and biological specificity (e.g. to strains or species) is determined by comparing the amino acid sequence of the peptide to isolate genomes or annotated MAGs. A least-common-ancestor (LCA) analysis can then be performed to assess the level of biological specificity that is represented by that peptide (Saunders et al., 2023) (Table S3, Supplement C). The targeted peptides were selected based on their abundance in a preliminary metaproteomics analysis, suggesting that these were the most abundant proteins within the microbial community (see Held et al., 2025 for a more complete description of how targeted peptides were selected for this study region). Peptide sequences for up to 5 strains of *Prochlorococcus* were targeted with a focus on the HLII clade, particularly strain MIT9314, allowing comparison between proteins of the same strain. In contrast, protein sequences for *Synechococcus* were compared across clades because the peptide sequence for PhoA and PhoX targeted WH8102 (clade III) but the peptide sequence for PstS targeted RCC307 (clade X, Table S3). While co-occurring clades III and X are geographically positively correlated in warm oligotrophic waters (Sohm et al., 2008), RCC307 possesses a different putative alkaline phosphatase gene compared to WH8102 (likely PhoA, see Tetu et al., 2009)). This mismatch in targeted strains and clades means that interpretation of the response of *Synechococcus* (and perhaps *Prochlorococcus*) to nutrient addition needs to be treated with some caution until the physiology and regulatory pathways of protein production are better understood. However, based on current knowledge of phosphate acquisition genes in marine *Synechococcus* and *Prochlorococcus*, we do expect that the targeted proteins/strains are major players in our study region.

3.4. Influence of trace metals on alkaline phosphatase and associated protein biomarkers

Unlike other cyanobacteria, where the trace metal availability aligns well with the biogeography of PhoA and PhoX (e.g. *Trichodesmium*, Rouco et al., 2018), there were no consistent trends between iron or zinc concentrations and proteins PhoX and PhoA respectively in this study. The distribution of *Pro*-PhoX from metaproteomes (Fig. 4a) did not reflect iron availability (Fig. 2a). Despite elevated iron in the west, *Pro*-PhoA concentrations were 2.7 to 4.7-fold higher than *Pro*-PhoX (reported fmol L⁻¹, Table 3), with *Pro*-PhoX being greater than *Pro*-PhoA in the east where Fe was lowest (Fig. 2a). PhoX was not detected for *Synechococcus* in metaproteome analysis but was detected quantitatively at the start of bioassay experiments (Table 3). *Syn*-PhoX concentrations also did not reflect iron availability (Table 3) and there was no consistent trend in the ratio between *Syn*-PhoA and *Syn*-PhoX (Table 3). Zonal trends in quantitative versus metaproteome-derived *Syn*-PhoA were different and likely driven by differences in depth horizons sampled (40m for experiments, 15 meters for metaproteome analysis) as well as the different *Synechococcus* populations captured using quantitative peptide analysis (see Table S3, clade III and X) compared to metaproteomes.

Table 3: Concentration of proteins (fmol L⁻¹) PhoX and PhoA for *Prochlorococcus* and *Synechococcus* at the start of the nutrient bioassay experiments at stations 2, 3, 4 and 7 (see Fig. 1a for locations) to illustrate how the relative concentration and ratio of PhoA to PhoX differ between *Prochlorococcus* and *Synechococcus* and across the zonal transect.

Protein biomarker	<i>Prochlorococcus</i>			<i>Synechococcus</i>		
	PhoX	PhoA	PhoA/PhoX	PhoX	PhoA	PhoA/PhoX
Protein conc. (fmol L ⁻¹)						
Station 2 control	17 ± 2	45 ± 14	2.7	21 ± 6	7 ± 1	0.3

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- Deleted:** Although strong zonal trends in dissolved Fe were observed (Fig. 2a), the distribution of *Pro*-PhoX gleaned from the metaproteome analysis (Fig. 4a) did not reflect iron availability. Instead, protein concentrations quantified in the control of the nutrient bioassays reveal that for *Prochlorococcus*, PhoA concentrations were 2.7 to 4.7-fold higher than PhoX (reported as the number of enzymes per litre seawater basis, Table 3) at stations, 2, 3 and 5, where Fe was elevated, but *Pro*-PhoX was greater than *Pro*-PhoA at station 7 where Fe was lowest (Fig. 2a). ¶
- Deleted:** Between stations 2 and 4, *Syn*-PhoX decreased from 21 ± 6 fmol L⁻¹ to 9 ± 5 fmol L⁻¹ in line with a decline in dissolved Fe but increased again to 16 ± 2 fmol L⁻¹ at station 7 where Fe was lowest (Fig. 2a). However, the
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Station 3 control	16 ± 11	48 ± 10	3.0	15 ± 2	22 ± 12	1.8
Station 4 control	7 ± 0.1	31 ± 13	4.7	9 ± 5	8 ± 3	0.8
Station 7 control	8 ± 2	3 ± 1	0.4	16 ± 2	28 ± 13	2.4

The systems biology of the PhoX enzyme is poorly understood compared to that of PstS and PhoA, where the latter is known to be regulated by phosphate and zinc (Cox and Saito, 2013; Martiny et al., 2006; Ostrowski et al., 2010; Tetu et al., 2009). Despite the lack of correlation between trace metal availability, PhoA and PhoX in the surface ocean in this study, results from bioassays conducted during the same expedition (Held et al., 2025; companion manuscript) lends support to the potential for a direct metal control on APA (Browning et al., 2017; Jakuba et al., 2008; Mahaffey et al., 2014; Saito et al., 2017). In the west, Zn addition stimulated a 6-fold increase in *Syn-PhoA* relative to the control. Cobalt addition simulated a 7-fold increase in *Syn-PhoA* and 8-fold increase in *Pro-PhoX*. Finally, iron addition stimulated a 2-fold increase in *Pro-PhoX* in the iron-deplete eastern Atlantic (Held et al., submitted).

Proteins relating to iron, zinc and B₁₂ metabolism in *Prochlorococcus* increased in the east alongside the increase in *Prochlorococcus* cell abundance. Ferredoxin and zinc transporters increased eastward by 3 to 9-fold and protein annotated as CobW, a member of the COG0523 family implicated in metal chaperone functions (Edmonds et al., 2021) and Co chaperone for B₁₂ synthesis (Young et al., 2021) also increased 3 to 10-fold (Fig. 4c). The eastward increase in three independent proteins (up to 10-fold) was greater than the increase in total protein for *Prochlorococcus* (~1.6-fold) implying a regulated molecular increase in response to resource limitation or competition, rather than reflecting a change in biomass only. However, zinc protein annotations in *Prochlorococcus* are putative and alignment-based transporter annotations are unable to discern cognate metal use. In addition, the role of zinc in *Prochlorococcus* physiology is uncertain. *Prochlorococcus* does not have an obligate Zn requirement when phosphate is available (Saito et al., 2002), and Zn is highly toxic to a Pacific Ocean strain of *Prochlorococcus* 29/08/2025 10:57:00. In addition, while CobW is an abundant protein among the ~20 genes involved in cobalamin biosynthesis, there are currently no known biomarkers for cobalt or zinc metabolism in *Prochlorococcus*, with studies producing negative results (Hawco et al., 2020). For *Synechococcus*, there were no clear trends in ferredoxin (Fig. 4c) and flavodoxin was infrequently detected. These findings highlight the difficulty of predict the direct metal requirement alongside metals controlling multiple APs *in-situ*, and makes a strong case for continued biochemical characterization of cyanobacterial trace metal physiology and enzymes.

3.5. Influence of nitrogen acquisition on the biogeography of *Prochlorococcus* and *Synechococcus* in the subtropical North Atlantic

We used nontargeted metaproteomics to interpret the role of nitrogen acquisition on the biogeography of *Prochlorococcus* and *Synechococcus*. Upwelling in the eastern Atlantic (Menna et al., 2015) alongside nitrogen fixation (Fig. 3g, Cerdan-Garcia et al., 2022) and dust deposition (Kunde et al., 2019) deliver fixed nitrogen including, nitrate, ammonium and urea to the surface subtropical Atlantic (dust ref here). Nitrate and ammonium concentrations were low across the transect (Fig. 1e and f), with maximum ammonium concentration coinciding

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There was no consistent or significant increased phytoplankton growth, increase in *Prochlorococcus* and *Synechococcus*, or AP activity after the addition of Fe, Zn or Co alone (Held et al., submitted). However, at station 2 in the west, *Syn-PhoA* increased 6 and 7-fold after the addition of Zn (38 ± 0.56 fmol L⁻¹) and Co (47 ± 6.8 fmol L⁻¹) relative to the control (6.7 ± 1.5 fmol L⁻¹), respectively. At station 4 over the mid-Atlantic ridge, *Pro-PhoX* increased 8-fold upon addition of Co relative to the control. At station 7 in the east, *Pro-PhoX* also increased over 2-fold upon Fe addition (to 18 ± 2.6 fmol L⁻¹) relative to the control (8.2 ± 2.4 fmol L⁻¹, Held et al., submitted). Results from these bioassays illustrate the potential for a direct metal control on AP, lending support to the hypothesis for local, albeit patchy metal-phosphorus co-limitation in the subtropical North Atlantic (Jakuba et al., 2008, Mahaffey et al., 2014, Saito et al., 2017, Browning et al., 2017).

Alongside metal control of AP, there were gradients in proteins relating to iron, zinc and B₁₂ metabolism in *Prochlorococcus*. Ferredoxin increased from west (1 to 2 spectral counts) to the east (9 spectral counts) alongside zinc transporters (1 to 3 spectral counts in the west, 9 in the east) although there was no clear zonal trend in zinc peptidases (Fig. 4c, Table 2). We consider these zinc protein annotations in *Prochlorococcus* as putative because this transporter annotated in the *Prochlorococcus* genome has not been validated for Zn transport, and alignment-based transporter annotations tend to have difficulty in discerning cognate metal use. Moreover, under phosphate-replete conditions, *Prochlorococcus* does not have an obligate Zn requirement (Saito et al., 2002), and Zn is highly toxic to a Pacific Ocean strain of *Prochlorococcus* (Hawco et al., 2018), contributing to the uncertainty of the role of zinc in *Prochlorococcus* ([22])

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with the highest N₂ fixation rates (Fig. 3g). *Prochlorococcus* (and *Synechococcus*) are prime beneficiaries of N exuded from N₂ fixers (Caffin et al., 2018). *Prochlorococcus*-AmtB and UrtA was lowest where N₂ fixation and ammonium were highest (Figure 4b), reflecting alleviation of N stress as *Prochlorococcus* benefited from N exudates, as observed in the North Pacific gyre (Saito et al., 2014, 2015). Otherwise, eastward increases in protein biomarkers in *Prochlorococcus*, specifically P-II, ammonium transporter AmtB and urea transporter UrtA (30-70%, Fig. 4b and Table 2) indicate increasing N stress towards the eastern Atlantic for *Prochlorococcus*.

Synechococcus appeared to be N stressed throughout the transect as UrtA spectral counts were constant and more than 5 times higher than for *Prochlorococcus* (Fig. 4b). This is likely due to its larger cell size and less efficient surface-area to volume ratio for nutrient acquisition (Chisholm 1992). P-II and AmtB (or NtcA) was not detected in the metaproteome of *Synechococcus*, perhaps because *Synechococcus* was 5 to 10 times less abundant in the metaproteomes compared to *Prochlorococcus*. The dominance of proteins for ammonium and urea acquisition of *Synechococcus* and *Prochlorococcus* are consistent with the premise that while marine *Synechococcus* and some *Prochlorococcus* strains have the genetic makeup to assimilate nitrate (Berube et al., 2015; Domínguez-Martín et al., 2022; Martiny et al., 2009), it accounts for < 5% of their total N demand, and instead ammonium and urea are the dominant N sources (Berthelot et al., 2019; Casey et al., 2016; Painter et al., 2008)

4.0. Conclusions

This study exploited natural gradients in nutrient resources created by upwelling in the east and dust deposition in the west. Combining biogeochemical states, enzyme rate measurements, and 'omics approaches, in the spirit of the developing 'BioGeoSCAPES program (Saito et al., 2024), we studied the nutrient acquisition strategies for *Prochlorococcus* and *Synechococcus* in-situ and using nutrient bioassays. Using protein biomarkers alongside biogeochemical signatures for nutrient stress, we concluded that *Prochlorococcus* and *Synechococcus* were P-stressed in the western Atlantic and *Prochlorococcus* was N-stressed in the eastern Atlantic, with *Synechococcus* showing signs of N-stress throughout the transect. Our findings are generally consistent with prior metagenomic observations on basin scale contrasts in N and P stress for *Prochlorococcus* in the Atlantic Ocean (at medium level, Ustick et al., 2021). There was evidence for trace metal control on alkaline phosphatase but the response of protein biomarkers to the addition of organic P, Zn and Fe differed between *Prochlorococcus* and *Synechococcus* (also see Held et al., 2025, companion manuscript), highlighting that the functions and systems biology of alkaline phosphatase regulation differs across the organisms and for different environmental stimuli. This indicates that ongoing laboratory characterization of protein biomarkers and cyanobacterial physiology is needed defining the regulation and function not only at the species level, but also across strains within species.

Under future climate scenarios, stratification, aerosol dynamics, N₂ fixation and the bioavailability of organic P are predicted to change (e.g. (Buchanan et al., 2021; Chien et al., 2016; White et al., 2012; Wrightson and Tagliabue, 2020), all with the potential to perturb the availability of already scarce nutrient resources in the oligotrophic gyres. To identify and quantify the future trajectory of *Prochlorococcus* and *Synechococcus* under future ocean scenarios, a holistic view that considers the species and strain specific strategies used to access

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Moved up [3]: For *Synechococcus*, the urea transporter UrtA spectral counts were more than 5 times higher than for *Prochlorococcus* and were constant across the transect, implying *Synechococcus* was N stressed throughout the transect (Fig. 4b), likely due to its larger cell size and less efficient surface-area to volume ratio for nutrient acquisition (Chisholm 1992). We did not detect P-II and AmtB (or NtcA) in the metaproteome of *Synechococcus*, perhaps because *Synechococcus* was 5 to 10 times less abundant in the metaproteomes compared to *Prochlorococcus*.[†]

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resources, alongside representation of large scale forcings are required. We have shown here that there is utility in combining biochemical assays with untargeted and targeted omics approaches to reveal these patterns, generate hypotheses that can be tested in controlled laboratory experiments, and improve predictions of marine microbiology and biogeochemistry in a changing ocean.

Competing interests: The authors declare no competing interests

Data availability: All new data are provided in the Supplement or are available from the British Oceanographic Data Centre (BODC) with the following DOIs: Size-fractionated iron measurements ((Berube et al., 2015; Domínguez-Martín et al., 2022; Martiny et al., 2009)), inorganic nutrients, alkaline phosphatase, DOP, chlorophyll, flow cytometry: <https://doi.org/10.5285/284a411e-2639-93de-e063-7086abc0e9d8>, Experiment D (<https://doi.org/10.5285/1e9c4caa-b936-fc7c-e063-7086abc06ff6>). The mass spectrometry proteomics data have been deposited to the ProteomeXchange Consortium via the PRIDE partner repository with the dataset identifier PXD054252 and 10.6019/PXD054252

Supplement. Supplementary information is provided as individual files and 1 zip file. There are 5 supplements including Supplement A (Fasta file), Supplement B (protein file), Supplement C (peptide file) and Supplement D (trace metal clean protocols for nutrient bioassays). In the zip file, there are 5 supplementary tables provided as spreadsheets (Table S1 to S5) and 5 supplementary figures (Fig. S1 to S5).

Author contributions: CM, MCL and AT acquired the funding from NERC. CM and MCL led the research cruise. CD conducted AP measurements. KK conducted Fe measurements. NW conducted zinc measurements. MSC conducted nutrient measurements. LW conducted N₂ fixation measurements. CM, MCL, CD, KK, LW and NW conducted the large volume incubation experiments. KK and NH conducted the quantitative proteomics analysis and MM analysed samples using mass spectrometry at WHOI. NH and MM conducted the global metaproteome analyses. CM and NH wrote the manuscript with significant contributions from MS, ML, CD, KK and AT.

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I'm a little worried about this statement because it implies that there is a lot of P stress and that it is difficult to identify it, but the alternative is that there actually isn't a lot of P stress and that is why it is rarely found. We do point out some mechanisms that would make it more difficult to see P limitation (like biological plasticity) but those same mechanisms also exist for N and Fe.		
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