

1 **Carbon emission reduction requires attention to the contribution of natural gas use:**
2 **Combustion and leakage**

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13 Abstract: Natural gas will continue to replace coal in the process of global energy structure reform,
14 but its leakage potential can delay the realization of global carbon neutrality. To quantify its impact,
15 we established a carbon dioxide (CO₂) and methane (CH₄) emission flux detection platform on the
16 220-m platform of the Institute of Atmospheric Physics, Chinese Academy of Sciences, located in
17 northwestern Beijing. The observation results indicated that the daily mean CO₂ and CH₄ fluxes
18 were $12.21 \pm 1.75 \mu\text{mol} \cdot \text{m}^{-2} \cdot \text{s}^{-1}$ and $95.54 \pm 18.92 \text{ nmol} \cdot \text{m}^{-2} \cdot \text{s}^{-1}$, respectively. The daily variations
19 in the emissions of these two gases were highly consistent, and The fluxes were significantly
20 correlated with natural gas consumption, indicating that natural gas has become a common source
21 of CH₄ and CO₂, the combustion of which releases CO₂, while its leakage processes emit CH₄.

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22 Vehicle-based identification demonstrated that CH₄ can escape at the production, storage and use
23 stages of natural gas. Based on natural gas consumption data, the upper limit of the calculated natural
24 gas leakage rate in Beijing reached $1.12\% \pm 0.22\%$, indicating that the contribution of CH₄ to
25 climate change could reach 23 % of that of CO₂ on a 20-year scale. Natural gas leakage was
26 estimated to delay the time for China to achieve carbon neutrality by at least almost fourthree years.

27 **KEY WORDS:**

28 CO₂ flux, CH₄ flux, Eddy covariance, Natural gas leakage, Climate forcing, Carbon neutrality

29 **1. INTRODUCTION**

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30 In 2015, the 1.5 °C temperature control target was proposed in the Paris Agreement to reduce

31 the occurrence of extreme weather events(Seneviratne et al., 2018). To achieve this goal, it is
32 necessary to actively promote the low-carbon development transformation of the economic system,
33 especially energy transformation. In this process, natural gas plays an important role, and typical
34 countries have indicated a trend of coal reduction and gas increase during energy structure
35 adjustment over the past century. It is expected that global natural gas consumption will continue to
36 increase by 2035.

37 Natural gas is commonly referred to as a clean alternative to coal, but its main component is
38 methane, with a global warming potential (GWP) that is 29.8 times greater than that of carbon
39 dioxide at the hundred-year scale(Environmental-Protection-Agency, 2024). If 3.4 % of methane
40 leaks into the atmosphere before natural gas combustion, the advantages of natural gas over coal
41 will become negligible(Kemfert et al., 2022). Recent studies have suggested that the average loss
42 rate of natural gas in cities worldwide ranges from 3.3 % to 4.7 %(Sargent et al., 2021). According
43 to statistics from the International Energy Agency (www.iea.org) in 2020, methane leakage in the
44 global oil and gas industry reached 72 million tons and amounted to 6 billion tons of carbon dioxide
45 equivalent (CO₂e) within 20 years. Therefore, it is unclear whether natural gas can become a
46 bridging material for energy transformation.

47 One important prerequisite is to determine the contribution of natural gas leakage during coal-
48 to-gas conversion to urban methane(CH₄) emissions and its climate effects. At present, conventional
49 CH₄ monitoring methods include ground, aviation, and satellite monitoring methods. Ground
50 monitoring aims to detect the atmospheric CH₄ concentration through the installation of sensors and
51 monitoring stations at fixed locations or on vehicles(Wunch et al., 2016). Notably, monitoring
52 equipment is often installed near potential emission sources, with high detection accuracy but
53 generally a limited spatial range. The aviation monitoring method can be employed to identify large-
54 scale CH₄ emissions through measurement techniques such as drones or aircraft but cannot be used
55 to achieve long-term monitoring(Duren et al., 2019; Frankenberg et al., 2016; Sherwin et al., 2024).
56 Satellite methods can compensate for the shortcomings of the former two methods(Chen et al., 2022;
57 Cusworth et al., 2018; Shen et al., 2023), which exhibit interference from clouds and require
58 significant labor and financial investments.

59 The eddy covariance method, which is based on tall towers, enables long-term monitoring of
60 methane emissions, thus facilitating the identification of methane sources in specific areas. However,

61 it should be noted that this method has certain limitations during urban flux measurements at higher
62 altitudes, as larger air volumes in the measurement system may lead to a significant imbalance
63 between the observed vertical turbulence exchange and surface net flux compared with those at
64 typical measurement heights. However, this deficiency should be considered in conjunction with
65 the advantages of urban tower measurements because cities typically correspond to deeper rough
66 sublayers that can extend to 2–5 times the average building height(Barlow, 2014). Therefore,
67 increasing the measurement altitude can help characterize the turbulent exchange between this layer
68 and the inertial sublayer.

69 Developing countries are the main driving force behind the continuous growth in global energy
70 demand. As Beijing is the capital of the world's largest developing country and the first city within
71 China to complete the coal-to-gas conversion process, clarifying the natural gas leakage process in
72 Beijing can provide guidance for energy transformation in developing countries regionally and even
73 globally. In this study, three aspects related to natural gas were investigated as follows. First, the
74 fluxes of ~~methane~~CH4 and ~~carbon dioxide~~CO2 were observed simultaneously via the eddy covariance method, which was
75 used to investigate the impact of the coal-to-gas policy on CO₂ and CH₄ in Beijing, including the
76 magnitude of CO₂ emission and the common effects on the sources of both. Second, with navigation
77 experiments, the natural gas leakage process in Beijing has been confirmed, and the emission levels
78 of natural gas at different stages have been further roughly estimated, which provides certain
79 effective insights for the control of natural gas leakage in Beijing. Third, we discuss climate forcing
80 caused by natural gas leakage while considering the CO₂ flux as a basis, calculate the natural gas
81 leakage rate with statistical data, and estimate the impact of natural gas leakage on China's carbon
82 peak and carbon neutrality in conjunction with existing reports.

83

84 **2. METHODS**

85 **2.1 Instrument setup for eddy covariance measurement**

86 The measurements were conducted at a 325-m high meteorological tower in northwestern
87 Beijing, with a closed-path observation system installed on a platform at a height of 220 m, which
88 included a dual laser gas analyzer (QC-TILDAS-DUAL, Aerodyne Research Inc., USA), three-
89 dimensional ultrasonic anemometer (Gill Instruments, Ltd., Lymington, Hampshire, UK), vacuum
90 pump (XDS35i, BOC Edwards, UK), data collector (CR6, Campbell Scientific Inc., USA), and

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91 other accessories. In the dual laser gas analyzer, tunable infrared laser direct absorption spectroscopy
92 (TILDAS) technology is used to detect the most significant fingerprint transition frequencies of
93 molecules within the mid-infrared wavelength range. The analyzer has an optical path of up to 76
94 m and can measure H₂O, CO₂ and CH₄ simultaneously. Similar instruments have been applied to
95 observe outdoor ecosystems(Zöll et al., 2016). Under the action of a vacuum pump, the air sample
96 enters the instrument room at a flow rate of 2 lpm through a polytetrafluoroethylene (PTFE)
97 sampling tube with a length of 3 m and an inner diameter of 3 mm (Figure S1). Instrument
98 calibration includes zero-point and range calibration processes. High-purity nitrogen gas (>99.999%)
99 was used for zero-point calibration at 1-hour intervals. In this process, the corresponding solenoid
100 valve was opened, which was automatically controlled by TDLWintel software, and range
101 calibration was performed at the factory. In addition, before the experiment, we calibrated the gas
102 analyzer using CO₂ (401 ppm) and CH₄ (2190 ppb) standard gases. We found that the measured and
103 standard gas concentrations differed by less than 1%, indicating satisfactory instrument performance.
104 Therefore, we did not perform range calibration later. The instrument was placed in an insulated box
105 equipped with air conditioning to ensure normal operation of the laser. Both instruments were
106 operated at a sampling frequency of 10 Hz. The data collector and high-frequency instrument were
107 timed according to the network and global positioning system (GPS), respectively, to maintain
108 synchronization. To minimize the twisting effect of the flux tower on the incoming air, a three-
109 dimensional ultrasonic anemometer was installed at the end of a 1.5-m long support arm facing
110 southeast China in summer. This measurement lasted from June 11 to September 7, 2022, during
111 which the nitrogen cylinder was replaced, and the instrument was debugged on June 18 and 19.
112 From July 12 to 26, the experiment was stopped due to failure of the tower power supply.

113 2.2 Flux data processing

114 The flux data processing operation in this study is based on the eddy covariance technique via
115 EddyPro software (version 6.2.1, Li COR, Inc.; Lincoln, Nebraska, USA). An average flux
116 calculation period of 30 minutes was selected(Lee, 2004). Before calculating half-hourly fluxes,
117 ~~skipped the first 10 min to let the fluxes adapt to the new time window, and then take a moving window with a width equal to 1/6 of the averaging period (typically 5 minutes)~~
118 al., (1997): Take a moving window with a width equal to 1/6 of the averaging period (typically 5
119 minutes) and calculate the mean and standard deviation of the time series within the window. Define
120 outliers as any data points deviating from the mean by n times the standard deviation (initial n =

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121 3.5) Replace the identified outliers with linearly interpolated values from adjacent points.
122 Consecutive outliers ≤ 3 are treated as a single outlier; consecutive outliers ≥ 4 are considered local
123 trends and excluded from outlier classification. Iteratively increase n by 0.1 per cycle until no
124 outliers are detected or 20 iterations are reached. Advance the window by half its width (step size)
125 and repeat outlier detection/removal for the next window. Continue this process until all outliers are
126 processed within the averaging period. If outliers exceed 1 % of the total data points in any averaging
127 period, discard that entire period.

128 –The double rotation method proposed by Kaimal et al., (1994) was employed for tilt
129 correction. The delay time caused by the spatial separation of gas analyzers and three-dimensional
130 ultrasonic anemometers (as well as the injection pipeline of closed-path systems) was corrected via
131 the maximum covariance method(Fan et al., 2012). Webb, Pearman, and Leuning (WPL) correction
132 was not applied here(Webb et al., 2007) because the instrument room was in a state of constant
133 temperature and pressure that converted the real-time concentration into a dry volume mixing ratio,
134 and the longer pipeline of the closed-path system avoided the influence of temperature fluctuations.
135 The limitations of eddy covariance systems can lead to frequency loss in flux observations. Factors
136 such as a limited average period and linear detrending can cause low-frequency loss, whereas
137 instrument separation, path averaging, insufficient high-frequency responses, and pipeline
138 attenuation can cause high-frequency loss. The method proposed by Moncrieff et al., (1996) was
139 employed for frequency response correction. After the above correction of the flux data, in this paper,
140 the 0-1-2 quality labeling scheme proposed by Mauder and Foken(Mauder et al., 2004) was adopted
141 for data quality control purposes. Notably, a value of 0 represents data with the best quality, a value
142 of 1 represents data with good quality, and a value of 2 represents data with poor quality. In this
143 study, flux data marked as 2 were excluded from the subsequent analysis. In addition, the flux source
144 area was evaluated via the method of Kljun et al., (2004) (Text. S1), and the flux source area covered
145 most of the urban area of Beijing and reflected the average emission characteristics of urban Beijing
146 (Figure S2).at the regional scale.

147 2.3 Spectral analysis

148 High-frequency signal loss can occur in closed-path systems. To determine the response
149 capability of the closed-path system to high-frequency turbulence signals, we analyzed the observed
150 gas exchange signals through the turbulence power spectrum. The selected time ranges from 12:00

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151 to 16:00 every day during the observation period, with a total of 8 and a half hours of data. The data
152 were integrated and averaged, and the data curve was then compared with the ideal slope in the
153 inertia subarea (Figure S3). $Co(wT)$ followed the theoretical $f_n^{-4/3}$ (where f_n denotes the normalized
154 frequency) in the inertial subregion. In contrast, the slopes of $Co(wCO_2)$ and $Co(wCH_4)$ were
155 slightly greater than $-4/3$, indicating that there was high-frequency loss in the flux observations of
156 the closed-loop system(Kaimal et al., 1972). Through high-frequency correction, the calculation
157 results indicated that the CO_2 and CH_4 fluxes were 7.73 % and 6.85 % greater, respectively, than
158 those before correction.

159 **2.4 Mobile CH_4 and CO_2 observations**

160 Vehicle-based experiments were conducted in the urban area of Beijing in the winter of 2023
161 and the summer of 2024, and the specific deployment of the mobile observation station is shown in
162 Figure S1. Notably, the car was equipped with a CO_2/CH_4 spectrometer (Los Gatos Research, Inc.,
163 USA), a laptop for data viewing, and a mobile power supply (Figure S4). Zero-point calibration of
164 the instrument was performed once pure nitrogen was used before the mobile experiment began.
165 Standard gases of methane and carbon dioxide were introduced to calibrate the instrument
166 simultaneously, and we found that the concentration of the instrument matched well with the
167 standard gas. Since we focused more on the enhancement in concentration rather than itself, we did
168 not calibrate it again afterward. The sampling port was located approximately 20 cm from the roof,
169 and ambient air was collected through a PTFE tube with a length of 2 m and an inner diameter of 3
170 mm. Before the particulate matter entered the instrument, it was removed using a filter head. The
171 IMET sounding instrument (International Met Systems, USA) is installed on the roof, with a
172 sampling frequency that is consistent with that of the other instruments, i.e., 1 s, real-time
173 concentration information of different latitudes and longitudes is obtained at a resolution of seconds
174 through the corresponding time between the GPS and the instrument; for example, if the GPS
175 sampling time delay is 3 s, the latitude and longitude coordinates are reassigned to the CH_4 reading
176 observed three seconds prior. Our observation sites include petrochemical plants located in
177 southwestern Beijing, natural gas storage tanks and landfills in the northeastern part, and power
178 plants with the highest natural gas usage in the southeastern part.

179 **3. RESULTS**

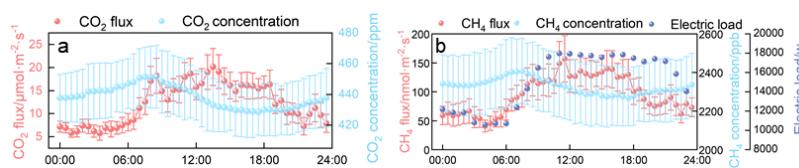
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180 **3.1 Diurnal variation in the flux**

181 A positive or negative flux reflects the vertical exchange direction of trace gases in the urban
182 canopy, which is positive upward and negative downward. (The uncertainty analysis is described in
183 the Text, S2 and Figure S5, respectively) Overall, both CO₂ and CH₄ fluxes are positive on a daily
184 scale, indicating that cities are the source of both gases. ~~The mean diurnal CO₂ flux is 12.21±1.75~~
185 ~~μmol·m⁻²·s⁻¹ (ranging from 6.05 to 19.66 μmol·m⁻²·s⁻¹ (Figure 1a), which is generally lower than~~
186 ~~that at 200 m in summer from 2013 to 2016 (mean 14.45 μmol·m⁻²·s⁻¹, ranging between 5 and 30~~
187 ~~μmol·m⁻²·s⁻¹)²² and at 140 m in summer from 2006 to 2009 (mean 16.19±4.12 μmol·m⁻²·s⁻¹,~~
188 ~~ranging from 8 to 20 μmol·m⁻²·s⁻¹)²². The diurnal CO₂ flux ranged from 6.05 to 19.66 μmol·m⁻²·s⁻¹ with an average of 12.21±1.75~~
189 ~~μmol·m⁻²·s⁻¹ (Figure 1a), which was generally lower than the summer observations by Cheng et al.,~~
190 ~~(2018) and Liu et al. (2012) at 200 m and 140 m at this tower, respectively (Table S1).~~ a smaller
191 deviation suggests that CO₂ may be dominated by a more stable source than before. We also obtained
192 observation results at 140 m in summer from 2009–2017 (Liu et al., 2020). The flux in 2022
193 significantly decreased compared with previous levels (Figure S6), which reflects the
194 transformation of Beijing's energy structure. The coal-to-gas policy implemented by Beijing these
195 years led to a gradual decrease in the proportion of coal in primary energy consumption, with a
196 steady increase in the proportion of natural gas in total consumption (Figure S7), the use of natural
197 gas results in much less coal ~~CO₂ carbon dioxide~~ than coal, generating the same amount of heat; moreover, Beijing
198 has increased the amount of electricity flow from other provinces in recent years (Figure S7), which
199 has further driven a decrease in the annual average concentration of PM_{2.5}, ~~which dropping to is expected to decrease to 30.5~~ $\mu\text{g}\cdot\text{m}^{-3}$
200 by 202~~42~~. In fact, previous studies have reported a high correlation between PM_{2.5} and CO₂ fluxes.
201 For example, Donato et al., (2019) found that the seasonal and daily variations in the particle
202 number flux in southern Italian suburbs are largely determined by both transportation activities and
203 household heating. Liu et al., (2020) confirmed that the CO₂ flux can explain 64 % of the interannual
204 variation in the PM_{2.5} concentration by fitting the correlation between the annual average PM_{2.5} and
205 CO₂ fluxes in Beijing from 2009 to 2017. Therefore, controlling CO₂ emissions can also greatly
206 control the concentration level of PM_{2.5}, thereby achieving the dual effects of mitigating climate
207 change and improving air quality. In terms of its diurnal variation, it did not follow a typical bimodal
208 pattern but rather remained high after reaching the first peak at 8:00, with a lower level at night,
209 reflecting high anthropogenic carbon emissions during the day, such as those resulting from

210 transportation and energy generation activities. The diurnal pattern of the CH₄ flux was similar to
211 the observation results of Giolo et al., (2012) and Helfer et al., (2016) (Figure 1b), reflecting an
212 ~~maximum daily CH₄ flux point (0.157 nmol m⁻² s⁻¹) at 11:30, and the CH₄ concentration (16,000 ppb)~~
213 ~~at 08:30, and then remained stable until after 10:30, when it began to rise rapidly again, reaching its~~
214 ~~daily peak of approximately 157.1 nmol m⁻² s⁻¹ around 11:30. After 17:30, it slowly declined. Its~~
215 ~~diurnal variation pattern showed some differences compared to CO₂ flux, which increased beginning~~
216 ~~at 03:30 to around 08:30 similar to CH₄ flux. However, the peak for CO₂ flux occurred around 13:30,~~
217 ~~then slowly decreased and decreased rapidly after 18:30. Assuming that the average CH₄ flux at~~
218 ~~midnight (00:00 to 06:00) can be employed as the baseline for nighttime emissions, it accounted for~~
219 ~~58% of the daily average flux. The CH₄ flux demonstrated a pronounced diurnal pattern, indicating~~
220 ~~a significant daily variation in the background source in the source area.~~



221 Figure 1 Daily variations in the CO₂ and CH₄ concentrations, fluxes, and electricity loads

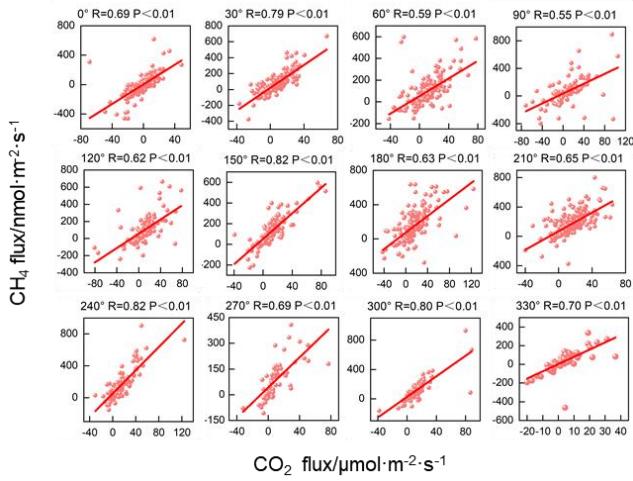
224 Figure 1 Daily variations in the CO₂ and CH₄ concentrations, fluxes, and electricity loads

225 3.2 Homology between CO₂ and CH₄

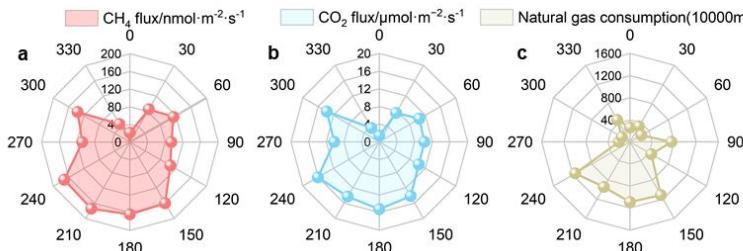
226 ~~The CO₂ fluxes exhibited similar diurnal patterns to the CH₄ fluxes, with the midnight mean~~
227 ~~accounting for 54% of the daily mean, which suggests that the fluxes may be driven by the same~~
228 ~~emission source. From the perspective of correlation statistics, the CO₂ and CH₄ fluxes showed a~~
229 ~~significant correlation along all directions (Figure 2), with correlation coefficients greater than that~~
230 ~~at the center of Loz, Poland (0.50) (Pawlak et al., 2016), but the low correlation between the CO₂~~
231 ~~and CH₄ fluxes and the temperature excludes the conclusion that biological sources dominate their~~
232 ~~emissions (Figure S8). Therefore, CO₂ and CH₄ share the same anthropogenic sources within the~~
233 ~~source area. This homology is also reflected in their spatial distributions, with high fluxes distributed~~
234 ~~mainly south of the tower, which is more densely populated and encompasses complex industrial~~
235 ~~structures, and much lower fluxes in the northern forest and park areas (Figure 3a, b). The~~
236 ~~correlation between the spatial distributions of the CO₂ and CH₄ fluxes reached 0.98, demonstrating~~

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237 the common impact of similar anthropogenic sources on their emissions. The linear fitting results at
238 150° and 240° indicated the highest correlation coefficient (0.82) along all directions (Figure 2),
239 further supporting this viewpoint.



241 Figure 2 Linear fitting results for the 30-minute CH₄ and CO₂ fluxes in the 12 directions



243 Figure 3 Mean CH₄ and CO₂ concentrations, fluxes and natural gas consumption in the 12
244 directions

245 4. DISCUSSION

246 4.1.3.3 Driver of the homology between CO₂ and CH₄

247 After the introduction of natural gas in 1985, the proportion of natural gas in the fossil fuel
248 industry of Beijing increased annually, especially when coal was replaced with natural gas and
249 electricity in 2014 and 2018, respectively, and natural gas became the most consumed fossil fuel
250 (Figure 4a). According to the 2022 Beijing Statistical Yearbook

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252 (<https://nj.tjj.beijing.gov.cn/nj/main/2023-tjnjkz/e/indexch.htm>), natural gas is used mainly for
253 thermal power generation and heating (accounting for 69%). Owing to the low proportion of heating
254 in summer, natural gas in Beijing is mostly used for thermal power generation in summer. Owing to
255 the difficulty in obtaining hourly electricity generation data, we obtained a daily variation curve of
256 the electricity load in Beijing based on the statistical data (power plants usually calculate the
257 required electricity generation based on the electricity load) (<https://www.gov.cn/xinwen/2019-12/30/5465088/files/e3682ce168c8427b886a43a790d66c2c.pdf>) (<https://www.gov.cn/zhengce/zhengceku/2020/12/03/5566580/files/ea9a93782e514543861bde434e86666.pdf>) (Figure 1b). The
258 daily variation in the electricity load is highly consistent with that in the CH₄ flux, with the
259 maximum CH₄ flux occurring at 11:00 pm during the peak electricity consumption period. After
260 16:00 pm, the electricity load and CH₄ flux decrease synchronously. Thus, the daily variation in the
261 CH₄ flux is driven by natural gas consumption. We gridded the natural gas consumption data (Figure
262 S9) and calculated the mean natural gas consumption along all directions within the flux source area
263 (Figure 3c). Notably, a high consistency between the spatial distributions of the CO₂ and CH₄ fluxes
264 and natural gas consumption was found, which reflects that after the adjustment of the energy
265 structure in Beijing, natural gas became the main source of CO₂ and CH₄. Considering the high
266 photosynthetic absorption of CO₂ by plants in summer, this conclusion also applies to the other
267 seasons, which supports the hypothesis that natural gas is the main source of winter CO₂ emissions
268 in Beijing, as determined based on the isotope tracing method(Wang et al., 2022; Wang et al., 2022).
269

270 ~~For the first time, we used mobile observations to determine the background concentrations of CO₂ and CH₄ in Beijing. The background concentrations were determined by calculating the mean concentrations of CO₂ and CH₄ in the air around large petrochemical plants, gas storage tanks, and power plants in Beijing. Given real-time variations in gas concentrations influenced by meteorological conditions and pollution transport, it was essential to determine background concentrations at each time point. The current mainstream approach for determining background values involves calculating the 5th or 10th percentile within a sliding window of 5 minutes (±2.5 min) or 10 minutes (±5 min) centered on the target timestamp (Pu et al., 2023; Well et al., 2018; Well et al., 2019). We compared and evaluated the results applying different combinations of time windows or percentile following the method of Schiferl et al., (2025). (Text S3). The 10-min time window with 5th percentile was used here to calculate the background value. The enhancement concentration can be defined as the difference between the observed value and the background value at the~~
271 mobile observations during winter and summer around large petrochemical plants, gas storage tanks,
272 and power plants in Beijing. Given real-time variations in gas concentrations influenced by
273 meteorological conditions and pollution transport, it was essential to determine background
274 concentrations at each time point. The current mainstream approach for determining background
275 values involves calculating the 5th or 10th percentile within a sliding window of 5 minutes (±2.5 min)
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280 can be defined as the difference between the observed value and the background value at the

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282 corresponding time. There was significant CH₄methane leakage around the gas storage tanks and power
283 plants in both winter and summer. Notably, the observed CH₄methane hotspots were located in the downzone
284 of potential leakage sources; therefore, we attribute the high CH₄methane concentration to the emissions of
285 these potential natural gas leakage sources. In winter, hotspots with concentrations higher than the
286 background value of 175963 ppb appeared around the gas storage tank (Figure 5a), corresponding to
287 an enhancement concentration of CH₄ (E-CH₄) and enhancement concentration of CO₂ (E-CO₂)
288 fingerprint line with a slope of 0.114 (Figure 6a). In addition, the enhancement concentration
289 fingerprint slopes of the other hotspot zones were 0.0625 and 0.075, respectively, indicating varying
290 degrees of leakage around the gas storage tank(Sun et al., 2019).The enhancement concentration
291 fingerprint in summer also revealed leakage related to gas storage equipment (Figure 5b), with a
292 slope of 0.043, analogous to that of 0.065 in winter. Similar to gas storage tanks, natural gas leakage
293 hotspots have been observed in various equipment in power plants. For example, fingerprints with
294 a slope of 0.005(Figure 6d)25 in summer reflected leakage related tocombustion gas storage devices or pipeline facilities
295 in power plants(Lamb et al., 1995), whereas fingerprints with a slope of 0.015, 0.02 or 0.0506 in summer reflected
296 leakage related tostorage combustion facilities (Figure 6c,d)(Hurry et al., 2016). We also discovered natural gas
297 leakage near the petrochemical plant (Figure 6e)25, the line with a slope of 0.02 was related to the gas
298 storage equipment, and the line with a slope of 0.005 was relevant to the natural gas combustion
299 ~~and the line with a slope of 0.005 was relevant to the natural gas combustion~~
300 also conducted mobile observations near a large landfill outside the Fifth Ring Road in Beijing,
301 which was a hotspot exhibiting a level exceeding the minimum concentration of 13785 ppb (Figure
302 5f). The concentration fingerprints were relatively disordered and significantly differed from those
303 of CH₄methane emissions dominated by natural gas (Figure 6f), indicating that waste disposal processes are
304 relatively complex and cannot be ignored in cities(Cusworth et al., 2024).

305 Converting observed concentration increments into emission rates is a simple means of
306 quantifying natural gas leakage, which is subject to atmospheric conditions and potential leak source
307 locations. Weller et al., (2018; 2019) developed a model based on the relationship between the
308 enhancement concentration and emission rate. The specific formula is shown in Text S4. The model
309 assumes that CH₄ enhancement is the best predictor of the leakage emission rate and that a greater
310 leakage emission rate corresponds to greater CH₄ enhancement. The method sets a minimum
311 threshold for the observed CH₄ concentration, which is 110% of the background value, to filter out

312 concentration changes caused by measurement. Moreover, when multiple detections are conducted
313 for the same leakage source, it is necessary to average the CH₄ enhancement values and then
314 substitute them into the above formula. We estimated the natural gas leakage emission rates from
315 different leakage sources with this method¹² and the confidence interval (CI) based on the Bootstrap
316 method was used to estimate the uncertainty of the leakage rate. The type of concentration
317 fingerprint can help define the types of leakage sources. The natural gas leakage rate from the gas
318 storage tank and power plant in winter were 7.4 ± 0.1 g/min and 0.6 ± 0.03 g/min, respectively, and
319 the natural gas leakage rate from the gas storage tank and power plant in summer were 1.2 ± 0.04
320 g/min and 2.1 ± 0.07 g/min, respectively. The natural gas leakage rate near the petrochemical plant
321 was 0.6 ± 0.04 g/min. The methane leakage rates from gas storage tanks and power plants during
322 winter observation were calculated as 1.02 – 4.10 g/min and 0.41 – 0.57 g/min, respectively, and 0.98
323 g/min and 0.52 – 1.45 g/min, respectively, during summer, which was lower than the results of Ars
324 et al., (2020) on the leakage rates of Toronto's natural gas distribution network (3.52–10.56 g/min),
325 but they noted that Well's method underestimated the leakage rate because it ignored smaller
326 concentration enhancements. A significant uncertainty in this method lies in the distance between
327 the leakage point and the vehicle; unfortunately, determining the distance between the two points in
328 practical operation is difficult, which may confound the estimation of methane leakage. Therefore,
329 sufficient mobile experiments should be conducted in subsequent work to accurately calculate
330 natural gas leakage in Beijing.

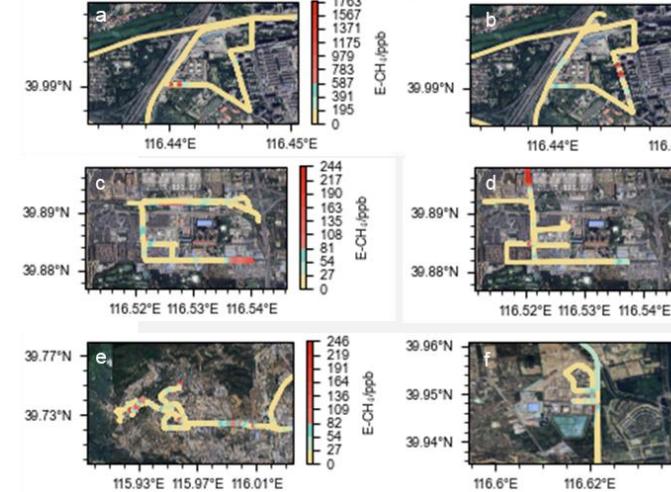
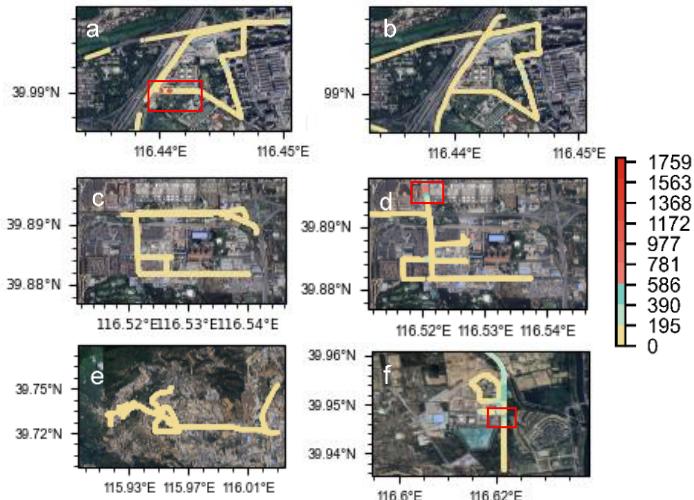


Figure 5 CH₄ enhancement concentration distribution map based on vehicle observations (a, c show storage tanks and thermal power plants in winter; b, d show storage tanks and thermal power plants in summer; e shows petrochemical plants; f shows waste disposal station; and the red box represents high leakage value)

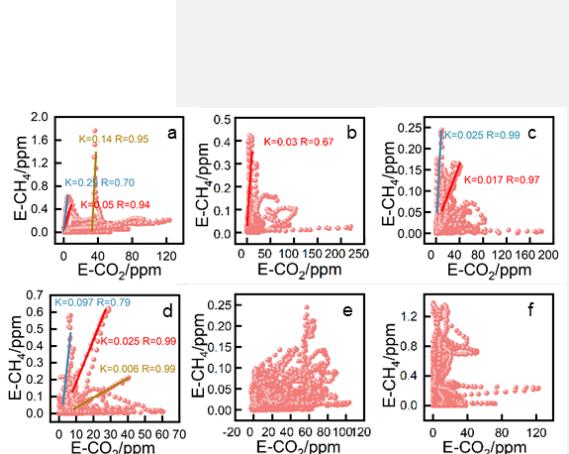
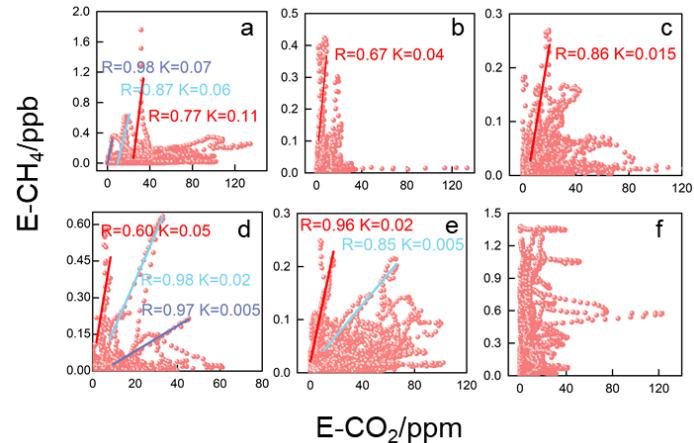


Figure 6 Fitting of the CO₂ and CH₄ concentration enhancement values (a, c show the fitting results for the gas storage tanks and power plants in winter; b, d show the fitting results for the gas storage tanks and power plants in summer; e shows the petrochemical plants; and f shows the waste disposal stations. Different fitting lines represent various leakage sources.)

342
343

4.23.4 Climatic effects of natural gas (NG) losses and their impact on carbon neutrality

344 Based on the natural gas consumption and flux data for the flux source area, the estimated
345 upper limit of the natural gas methane leakage rate in Beijing reached $1.12\% \pm 0.22\%$ (Text. S5),
346 and the lower limit of natural gas leakage in Beijing was estimated to be 0.82 % considering the
347 emissions from biogenic sources (Text. S6). If the CH₄ fluxes were attributable solely to pipeline
348 leakage processes, the CH₄ fluxes should remain relatively stable throughout the day without
349 significant diurnal variations, given the constant pressure in urban pipeline pressures. However in
350 our observations, the CH₄ fluxes exhibited pronounced diurnal patterns and their spatial distribution
351 positively correlated with natural gas consumption. This indicates that CH₄ emissions in Beijing
352 originate predominantly from consumption-oriented leakage processes. Consequently, as natural
353 gas consumption surges during winter heating periods, CH₄ emissions from these processes (e.g.,
354 fugitive emissions from electrical devices) also increase. As a result, the ratio of emissions to
355 consumption (leakage rate) remains relatively stable. Thus, the CH₄ leakage rate measured in
356 summer is representative of year-round leakage rate of natural gas, which is lower than the value of
357 2.07% calculated based on the purchase and sales statistics and the statistical mean value of 1.1
358 1.65% reported by the American Petroleum Institute (<https://www.api.org/>), we assume that the
359 leakage rate does not have significant seasonal variability because of the positive correlation
360 between methane flux and natural gas consumption.

361 Our measured leakage rate was lower than the value of 2.07 % calculated based on the purchase
362 and sales statistics and the statistical mean value of 1.1 %–1.65 % reported by the American
363 Petroleum Institute (<https://www.api.org/>). The natural gas leakage rate in Beijing is relatively low,
364 as noted in existing reports. Nevertheless, the contributions of CH₄ to climate warming are 8.37 %
365 and 23.17 % of those of CO₂ at the 100- and 20-year scales, respectively, according to the
366 determined CO₂ and CH₄ fluxes and the GWP of methane. With the arrival of the winter heating
367 season, climate forcing will further increase on a yearly scale. Assuming that the natural gas
368 consumption in Beijing during the heating season is 5 times greater than that during the other
369 seasons (according to Beijing Gas in 2019), that oil consumption does not significantly fluctuate
370 throughout the year and that both the CO₂ and CH₄ fluxes are positively correlated with fossil fuel
371 consumption and natural gas leakage, the climate forcing effect of natural gas leakage in 2022 was

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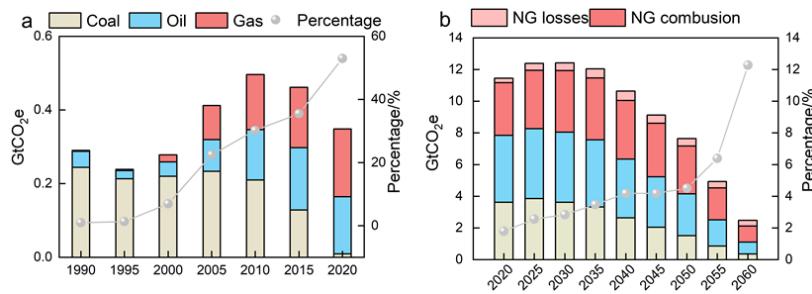
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11.47 % on a 100-year scale and could reach as high as 31.56 % on a 20-year scale. However, when the same amount of heat is generated, the use of natural gas could yield CO₂ emission reductions of 50 % relative to coal and of only approximately 30 % relative to oil. Therefore, the reduction in greenhouse gas emissions resulting from natural gas combustion compared with that resulting from the combustion of other fossil fuels may be offset by the climate forcing effect of [CH₄methane](#) leakage in the short term, making it difficult for natural gas to become a transitional energy source for energy transition.

379 To assess the impact of natural gas leakage on carbon peak and carbon neutrality based on our
380 quantified leakage rate, scaling the Beijing-derived leakage rate to a national level is needed.
381 However, due to the absence of leakage rate data from other cities, we can provide only a rough
382 estimate based on available data as follows: according to the 14th Five-Year Plan for National Urban
383 Infrastructure Development (<https://www.gov.cn/zhengce/zhengceku/2022-07/31/5703690/files/d4ebd608827e41138701d06fe6133cdb.pdf>), cities in China are divided into
384 three categories—major cities (natural gas penetration rate $\geq 85\%$), medium cities (natural gas
385 penetration rate $\geq 75\%$), and small cities (natural gas penetration rate $\geq 60\%$). The China Gas
386 Development Report 2023 further supplements pipeline coverage
387 progress(<https://www.emerinfo.cn/download/zgtrqfzbg2003001.pdf>), indicating that large cities
388 and developed regions (e.g., Beijing, the Yangtze River Delta, the Pearl River Delta) accounted for
389 approximately 30%–40% of the national pipeline length in 2022, here set at 35%. Small/medium
390 cities constituted 60%–70% of the total pipeline length, here set at 65%. A study based on Bayesian
391 network modeling revealed that leakage probabilities in small/medium cities are 1.8 times higher
392 than those in major cities (95% CI: 1.6–2.0)(Gao et al., 2024). Consequently, the national leakage
393 rate was calculated as $1.7\% (95\% \text{ CI: } 1.57\% - 1.85\%) = 0.35 \times 1.12\% + 0.65 \times 1.12\% \times 1.8 (95\% \text{ CI: } 1.6 - 2.0)$.
394
395

402 the China Energy Outlook 2060 (SINOPEC 2021)(Economics-and-Development-Research-
 403 Institute 2021) ~~and it is predicted that the total CO₂ emissions will peak in 2030. According to the~~
 404 ~~China Energy Outlook 2060 (SINOPEC 2021), the total CO₂ emissions will peak in 2030. According to the~~
 405 ~~the CO₂e in China will still peak by 2030. However, the CO₂e resulting from natural gas leakage~~
 406 ~~will reach 0.37 Gt (95 % CI: 0.34 Gt–0.40 Gt) in 2060, compared to 0.26 Gt previously. This~~
 407 ~~accounts for approximately 16.6 % (95 % CI: 15.4 %–17.9 %) of the total CO₂ emissions (excluding~~
 408 ~~natural gas leakage) and 35.9 % (95 % CI: 33.2 %–38.8 %) of the total CO₂ emissions from natural~~
 409 ~~gas combustion, which is comparable to the CO₂ emissions from coal combustion (0.35 Gt). Since~~
 410 ~~natural carbon sinks do not show significant short-term fluctuations, the future increase in carbon~~
 411 ~~sinks will mainly rely on carbon capture and storage (CCS) technology. Given the current estimated~~
 412 ~~CO₂ capture rate of CCS technology (0.1 Gt/year, as estimated by the China Energy Outlook 2060~~
 413 ~~(SINOPEC 2021)), the achievement of carbon neutrality in China will likely be delayed by nearly~~
 414 ~~three to four years. Therefore, when determining future natural gas consumption levels, it is~~
 415 ~~necessary to both consider the leakage effects of natural gas and utilize carbon modeling.~~



416
 417 Figure 4 Terminal consumption of coal, oil, and natural gas and their proportions from 1990 to
 418 2020(a) Since diesel-powered trucks are allowed only at night on the Fifth Ring Road and
 419 kerosene, which is used mainly in aviation and is not included in the flux source area, oil mainly
 420 comprises gasoline in this case), CO₂ equivalent from coal, oil and natural gas (losses and
 421 combustion) in the future scenario (estimated by China Energy Outlook 2060 released by
 422 SINOPEC in 2021), and CO₂ equivalent of natural gas leakage as a proportion of natural gas (NG)
 423 combustion emissions(b)

424 Notably, the Beijing-Tianjin-Hebei region has experienced the most severe air pollution in China.
 425 To ensure people's health, Beijing's coal-to-gas policy has been implemented most thoroughly with
 426 a well-established layout and control measures for natural gas leakage across China, so it is
 427 reasonable to apply the leakage rate from Beijing to all of China and attempt to estimate the impact

428 ~~the gas supply chain, including pipeline, storage, and distribution, and the natural gas~~

429 **4.33.5 Policy implications**

430 Our observations revealed a strong correlation between CH_4 methane emissions and natural gas
431 ~~and the gas supply chain, including pipeline, storage, and distribution, and the natural gas~~
432 terminal consumption process ~~may~~ drive natural gas leakage in Beijing. Liu et al., (2023) established a
433 bottom-up emission inventory and reported that the terminal use process in Beijing accounts for 80%
434 of the total methane emissions in the entire natural gas supply chain. Therefore, the Chinese
435 government may need to expand the detection of pipeline leakage to the entire natural gas industry
436 chain.

437 Notably, existing grid-based inventory products also exhibit significant uncertainty in terms of
438 methane sources. The extracted inventory originates from the Emissions Database for Global
439 Atmospheric Research (EDGAR) (<https://edgar.jrc.ec.europa.eu/EDGARv8.0>). Although the mean
440 methane flux ($126.34 \text{ nmol} \cdot \text{m}^{-2} \cdot \text{s}^{-1}$) within the source area is close to our results, the terminal use
441 process accounts for only approximately 13 % of the annual methane emissions, suggesting that
442 many potential urban methane sources could have been missed, which should be considered in
443 inventory refinement in the future.

444 In addition, minimizing the methane leakage rate could ensure the early realization of carbon
445 neutrality in China. Although methane emission control has been included in the agenda for the first
446 time in the Methane Emission Control Action Plan promulgated in 2023, which clearly highlights
447 the need to promote the application of leak detection and repair technology and to enhance the
448 comprehensive recovery and utilization of methane, methane leakage standards have not been
449 updated. Previous methane leakage standards focused only on controlling the amount of methane
450 leakage from a safe perspective, thereby ignoring the climate effects of natural gas leakage. China
451 must urgently develop a strict and detailed set of natural gas leakage standards.

452 **5. SUMMARY AND CONCLUSIONS**

453 This study utilized the eddy covariance method to measure CO_2 and CH_4 fluxes at 220-m height
454 in urban Beijing, providing critical insights into surface-atmosphere exchanges of greenhouse gases
455 in the region. First, urban areas unequivocally act as net sources of both CO_2 and CH_4 . The daily
456 mean fluxes were $12.21 \pm 1.75 \text{ } \mu\text{mol} \cdot \text{m}^{-2} \cdot \text{s}^{-1}$ for CO_2 and $95.54 \pm 18.92 \text{ nmol} \cdot \text{m}^{-2} \cdot \text{s}^{-1}$ for CH_4 , with
457 daytime emissions significantly exceeding nighttime levels, highlighting the importance of

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458 anthropogenic influences.

459 Although diurnal variation patterns differed slightly between CO₂ and CH₄ fluxes, their strong
460 correlation indicates shared dominant sources. Spatial distribution analysis revealed high
461 consistency between both fluxes and natural gas consumption patterns, confirming natural gas as a
462 common source. With Beijing's energy restructuring, natural gas has become the dominated
463 terminal energy consumption. Its combustion releases substantial CO₂, while leakage processes emit
464 CH₄, as validated by mobile observations detecting CH₄ fugitive emissions during production,
465 storage and use stages. Although biogenic sources could contribute to CH₄ emissions, they account
466 for at most 27 % of total CH₄ fluxes in the source area, ruling out the view that biological sources
467 dominate both emissions. Attributing all CH₄ emissions to natural gas usage, the upper leakage rate
468 of natural gas in Beijing was calculated as 1.12 %± 0.22 %.

469 The CH₄ emissions from natural gas will exacerbate climate warming. Calculated flux results
470 showed that the contribution of CH₄ to climate warming on a century and 20-year scale can reach
471 as high as 8.37 % and 23.17 % of CO₂, respectively. On the basis of predicted energy report and
472 calculated leakage rate, it is predicted that natural gas leakage will delay China's realization of
473 carbon neutrality, which necessitates urgent attention to mitigate associated climate effects.

474 **SUPPORTING INFORMATION**

475 Details about the Beijing Meteorological Tower, eddy observation system and navigation
476 observation station, daily summer variation in CO₂ flux from 2009 to 2017, total consumption,
477 electricity inflow and the proportion of natural gas in total energy consumption from 2013-2022,
478 spatial distribution of CO₂ and CH₄ fluxes with wind speed and direction, grid distribution of natural
479 gas consumption in Beijing, calculation methods of the flux source area and natural gas leakage rate,
480 uncertainty analysis of flux calculation, estimation of non-natural gas sources

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487 **DATA AVAILABILITY**

488 All the data generated or analyzed in this study are included in the published article and are available
489 from the authors upon reasonable request.

490 **COMPETING INTERESTS**

491 The authors declare that they have no competing interests.

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