

1 **Seasonal shifts in depth to water uptake by young thinned and overstocked lodgepole pine**
2 **(*Pinus contorta*) forests under drought conditions in the Okanagan Valley, British**
3 **Columbia, Canada**

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12 **Abstract:**

13 As drought and prolonged water stress become more prevalent in dry regions under climate
14 change, preserving water resources has become a focal point for maintaining forest health. Forest
15 regeneration after forest loss or disturbance can lead to over-stocked juvenile stands with high
16 water demands and low water-use efficiency. Forest thinning is a common practice with the goal
17 of improving tree health, carbon storage, and water use while decreasing stand demands in arid
18 and semi-arid regions. However, little is known about the impacts of stand density on seasonal
19 variation in depth to water uptake nor the magnitude of the effect of growing season drought
20 conditions on water availability. Existing reports are highly variable by climatic region, species,
21 and thinning intensity. In this study, stable isotope ratios of deuterium ($\delta^2\text{H}$) and oxygen ($\delta^{18}\text{O}$)
22 in water collected from soil varying depths and from branches of lodgepole pine (*Pinus contorta*)
23 under different degrees of thinning (control: 27,000 stems per ha; moderately thinned: 4,500
24 stems per ha; heavily thinned: 1,100 stems per ha) over the growing season were analyzed using
25 the MixSIAR Bayesian mixing model to calculate the relative contributions of different water
26 sources in the Okanagan Valley in the interior of British Columbia, Canada. We found that under
27 drought conditions the lodgepole pine trees shifted their depth to water uptake through the
28 growing season (June to October), to rely more heavily on older precipitation events that
29 percolated through the soil profile when shallow soil water became less accessible. Decreased
30 forest density subsequent to forest thinning did not cause a significant difference in isotopic
31 composition of branch water, but did cause changes in the timing and relative proportion of
32 water utilized from different depths. Thinned lodgepole pines stands were able to maintain water
33 uptake from 35 cm below the soil profile whereas the overstocked stands relied on a larger
34 proportion of deep soil water and groundwater towards the end of the growing season. Our
35 results support other findings by indicating that although lodgepole pines are drought tolerant
36 and have dimorphic root systems, they did not shift back from deep water sources to shallow soil
37 water when soil water availability increased following precipitation events at the end of the
38 growing season.

39 **Keywords:** *Pinus contorta*; stable water isotopes; forest thinning; water-use strategies;
40 preferential water uptake; dual-isotope analysis; Bayesian isotope mixing model; soil water
41 uptake; transpiration; the interior of British Columbia

42 1. Introduction

43 As forests recover after harvesting, carbon and water demands change, and future climate
44 projections of increased drought severity will further complicate biogeochemical cycling and
45 carbon-water trade-offs (Giles-Hansen et al., 2021; Wang et al., 2019). Overpopulated
46 regenerating stands can add further stress on ecosystems; for example, light competition in dense
47 juvenile stands increases stand water demands by driving vertical growth and canopy cover (Liu
48 et al., 2011). To mitigate this stress, management strategies such as systemic thinning of high-
49 density juvenile stands have been shown to promote forest regeneration while decreasing
50 competition and providing remaining vegetation with increased light availability, rooting space,
51 nutrient access, and space for horizontal branch growth (Giuggiola et al., 2016). Over a variety
52 of forest ecosystems, reductions in stand density increase light availability, tree water use, carbon
53 storage, and water-use efficiency, an indication of improved tree health, and to decrease stand
54 water use, reducing the intensity of water stress under drought conditions (Belmonte et al., 2022;
55 Fernandes et al., 2016; Giuggiola et al., 2016; Liu et al., 2011; Manrique-Alba et al., 2020;
56 Molina & del Campo, 2012; Park et al., 2018; Sohn et al., 2012, 2016; Wang et al., 2019).
57 Because the primary goal of forest thinning is to decrease stand water use and increase
58 productivity, papers reporting the effects of this management strategy often focus on changes in
59 carbon storage, tree growth, transpiration, and water-use efficiency (Giuggiola et al., 2016;
60 Manrique-Alba et al., 2020; Park et al., 2018; Sohn et al., 2016). However, few studies have
61 reported sources of water use for vegetation water uptake and shifts in depth to water uptake in
62 association with thinning treatments in overstocked naturally regenerating forests, particularly
63 under drought conditions.

64 Quantifying stand water use is imperative to predicting the future of water availability in our
65 ecosystems. However, various studies indicate that trees do not always use the most recent
66 precipitation, and that vegetation can utilize different sources of water at different soil depths
67 depending on availability or stress (Dawson & Pate, 1996; Grossiord et al., 2017; Wang et al.,
68 2017). Many studies also report the depth of water uptake of various species and the relationship
69 between co-existing species and shared water sources (Andrews et al., 2012; Brinkmann et al.,
70 2019; Grossiord et al., 2017; Langs et al., 2020; Liu et al., 2015; Maier et al., 2019; Meinzer et
71 al., 2007; Sánchez-Pérez et al., 2008; Szymczak et al., 2020; Wang et al., 2017; Warren et al.,
72 2005). In water-limited regions such as arid and semi-arid landscapes, some species have
73 adapted to derive water from various depths over time depending on seasonal water variability,
74 indicating higher ecological plasticity and drought tolerance (Langs et al., 2020; Wang et al.,
75 2017). Understanding where in the soil profile plants obtain water, over prolonged dry periods
76 and at different stand densities, is essential in assessing the impact of forest thinning and the
77 relative importance of different seasonal water sources during shifts in water availability in arid
78 regions and under future climate conditions (Evaristo et al., 2015; Prieto et al., 2012; Sohn et al.,
79 2016). The implications of depth to water uptake and seasonal changes in water utilization, in
80 conjunction with water-use efficiency, can emphasize the importance of the timing and volume
81 of precipitation events and primary contributors to vegetation water use.

82 Stable isotope ratios can be used as powerful natural tracers to identify distinct water sources
83 such as rainfall, snow, and groundwater (Brinkmann et al., 2018; Lin & Sternberg, 1993;
84 Sprenger et al., 2017; Stumpp et al., 2018). The isotopic signature of precipitation events is
85 altered by elevation, temperature, and evaporative fractionation creating distinctive layers within

86 the soil profile (Kleine et al., 2020; Sprenger et al., 2017; Stumpp et al., 2018). More
87 specifically, soil water reflects precipitation events as they infiltrate through the soil layer with
88 the influence of evaporative fractionation until mixing with older soil water and groundwater and
89 creating individualized water isotopic signatures throughout the soil profile (Andrews & Science,
90 2009; Brinkmann et al., 2018; Dawson & Pate, 1996; Sprenger et al., 2017; Stumpp et al., 2018).
91 The isotopic composition of plant water can correspond to the water uptake depth in the soil
92 profile (Brinkmann et al., 2019; Langs et al., 2020; Meinzer et al., 2007; Stumpp et al., 2018;
93 Wang et al., 2017). Due to these unique characteristics, stable water isotopes have been used by
94 researchers to assess sources of water used by plants and their possible shifts under altered
95 environmental conditions (Evaristo et al., 2015; Flanagan & Ehleringer, 1991; Meinzer et al.,
96 2001; Stumpp et al., 2018).

97 Lodgepole pine (*Pinus contorta* Douglas) is an early successional montane conifer with a deep
98 tap root, fine roots in shallow soil layers, and an adventitious rooting system which allow this
99 species to access water throughout the soil profile (Fahey & Knight, 1986; Halter & Chanway,
100 1993). Depending on the species, root structures have two main components; namely, lateral
101 roots to exploit soil near the surface, and, in species with dimorphic root systems, sinker roots or
102 a well-developed tap root to reach deeper soil water or groundwater when surface water is
103 limited. Species with dimorphic rooting systems can access water from different depths in the
104 soil profile depending on soil moisture content and water availability, making them more
105 resilient to water scarcity or prolonged drought conditions (Dawson & Pate, 1996; Meinzer et al.,
106 2013). Wang et al. (2019) studied the short-term effects of thinning overstocked juvenile (16-
107 year-old) lodgepole pine stands in the Upper Penticton Creek Watershed, British Columbia,
108 Canada, and found a significant positive relationship between growth and water use from
109 decreased stand density and that heavily thinned treatments showed the most drought resistance.
110 Andrews et al. (2012) compared water uptake strategies between Douglas-fir (*Pseudotsuga*
111 *menziesii* (Mirb.) Franco) and lodgepole pine in southern Alberta, and found that lodgepole pines
112 are able to minimize seasonal variations in stem water potential and that tap roots are deep
113 enough to access groundwater. These findings are consistent with other literature reporting that
114 decreased stem density can improve water-use efficiency and that conifer trees can access water
115 from different depths depending on moisture availability (Meinzer et al., 2007a; Warren et al.,
116 2005). The literature therefore indicates that lodgepole pines can access water from different soil
117 layers even under extreme or prolonged drought conditions, but little is known about the shifting
118 of water use under different stand densities as a result of thinning treatments and drought
119 conditions.

120 In this study, we build on the research from Wang et al. (2019) which looked at the effects of
121 thinning on water-use efficiency during a drought and non-drought year by analyzing the stable
122 isotope ratios ($\delta^2\text{H}$ and $\delta^{18}\text{O}$) of soil and xylem water to evaluate at what depths overstocked and
123 thinned stands access water over a growing season to further our understanding of the
124 ecosystem-level impacts of thinning as a management strategy. We hypothesized that lodgepole
125 pine primarily relies on spring snowmelt, but reductions in shallow source water during the
126 growing season would drive trees to utilize deeper sources of water as the season progressed. We
127 also hypothesized that decreased stand density (thinning) would increase shallow soil
128 evaporation due to decreased canopy cover, but also decrease competitive limitations in tree
129 rooting zones so that at lower densities trees could better maintain mid-level soil water uptake.

130 Through a detailed partitioning of tree water sources, we can better understand how lodgepole
 131 pine uses water, estimate proportional dependence of lodgepole pine on specific source waters,
 132 and determine if thinning affects tree water use and uptake strategies under drought conditions.

133

134 2. Methods

135 2.1. Study site

136 The study was conducted in the Upper Pentiction Creek experimental watershed (UPC) northeast
 137 of Pentiction in the interior of British Columbia, Canada (49°39'34" N, 119°24',34" W). The site
 138 elevation is approximately 1675 m a.s.l. with steep, rocky terrain and a southern aspect (Wang et
 139 al., 2019). The luvisolic soils were formed from granite; the texture is coarse sandy-loam and is
 140 well drained with a low water holding capacity (Hope, 2011; Winkler et al., 2021; Winkler &
 141 Moore, 2006). The biogeoclimatic region is the Engelmann Spruce-Subalpine Fir zone with cold,
 142 snowy conditions from November
 143 to early June and seasonal drought
 144 conditions during the summer
 145 months, June to October (Coupe et
 146 al., 1991; Wang et al., 2019). This
 147 research site was initially
 148 established as a paired watershed
 149 experiment in the early 1980s to
 150 quantify the impact of forest
 151 harvesting on water resources
 152 (Creed et al., 2014; Moore &
 153 Wondzell, 2005; Winkler et al.,
 154 2021).

155 The juvenile thinning experiment
 156 began in 2016 when 16-year-old,
 157 evenly aged, regenerating lodgepole
 158 pine stands were thinned to different
 159 densities than a control (Control - C:
 160 27,000 stem ha⁻¹, T1: 4,500 stems ha⁻¹, and T2: 1,100 stems ha⁻¹) where C represents the control
 161 stands, T1 represents the moderately thinned stands, and T2 represents the heavily thinned stands
 162 (Figure 1). The three treatments were repeated across three replicate blocks. Each block was 75
 163 m long and 25 m in width with three 20 m² plots and 5 m between treatment plots. After the
 164 initial thinning, all debris was left on site.

165

166 2.2. Climate and soil moisture monitoring

167 Climate stations (HOBO weather station, Onset Computer, Bourne MA, USA) were deployed
 168 across Block 1 treatments and have measured meteorological data since 2016 (ambient
 169 temperature, relative humidity (rH), wind speed, precipitation, and solar radiation) in 10-minute
 170 intervals. From these data, we calculated daily vapor pressure deficit (VPD) as well as daily and

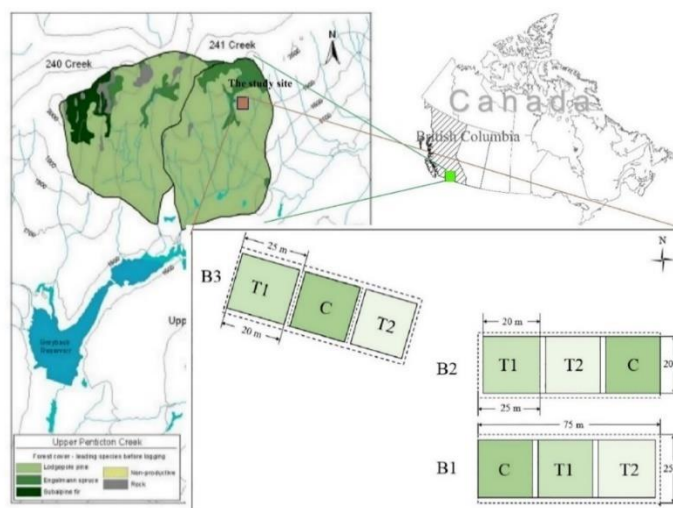


Figure 1 Watershed location and treatment plots of moderately thinned (T1), heavily thinned (T2), and the controlled (C) over-populated stands across the three replicate blocks (Wang et al., 2019)

171 monthly potential evapotranspiration (PET) (Flint & Childs, 1991; Russell, 1960; Streck, 2003).
 172 Recorded historical precipitation (1997-2008) was acquired from a long-term climate station in a
 173 lodgepole pine forest in the 241 experimental watershed (climate station P7) (Moore et al.,
 174 2021).

175 Rainfall and temperature data from Block 1 were related to historical data to calculate the
 176 monthly dryness (PET/P), standardized precipitation index (SPI), and standardized precipitation
 177 evapotranspiration index (SPEI) (Beguería et al., 2014; Stagge et al., 2014; Wu et al., 2005). In
 178 the middle of the growing season in 2021, four soil moisture probes (HOBO TEROS 11 Soil
 179 Moisture/Temp Probes) were deployed in each treatment in Block 1 to measure changes in soil
 180 moisture and temperature at 5 cm and 35 cm at 15-minute increments (n=12).

181 2.3 Sample collection

182 We sampled three trees per treatment across the three blocks and three in an adjacent mature plot
 183 south of the study site (n = 3333) four times over the 2021 growing season in approximately six-
 184 week intervals (June 11-12, July 21-22, September 10-11, and October 7-8) around noon to
 185 capture peak transpiration time (Table 1). We used a pole pruner to cut a mid-canopy branch in
 186 the live crown. We peeled the bark off branch segments with no needle coverage to remove outer
 187 bark and phloem, placed them into 10 mL glass tubes that were then with Parafilm® wrap,
 188 covered in aluminum foil, and set in a cooler until the end of the day when they were transferred
 189 to a freezer at -18°C. During the last two sampling periods, some trees had red needles, likely an
 190 indication of dryness or higher temperatures from an early growing season heat dome that began
 191 in June.

192 *Table 1 Overview of the branch, soil, and precipitation samples collected over the four sampling periods during the*
 193 *2021 growing season with an additional campaign to collect groundwater and stream water.*

Sampling Period		1	2	3	NA	4
Sampling Date		June 11-12	July 21-22	September 10-11	October 1	October 9
Sample Type	Branches	33	33	33	0	33
	5	9	9	9	0	9
	20	0	6	0	0	0
	35	9	9	9	0	9
	40	0	6	0	0	0
	60	0	6	0	0	0
	80	0	6	0	0	0
	100	0	6	0	0	0
	Rain	1	0	1	0	0
	Precipitation Snow	1	0	0	0	1
Stream	0	0	0	8	0	
Groundwater	0	0	0	6	0	

194

195 Soil samples were collected horizontally from 40 cm soil pits randomly dug within each
 196 treatment plot at 5 and 35 cm depths from the surface from June to October of 2021. Large rocks
 197 were removed from the profile. We conducted soil ribbon field tests to ensure that clay

198 composition was less than 10% (soil ribbons were less than 20 mm in length). Soils were taken
199 directly from the pit, then sealed in freezer seal bags and frozen until cryogenic distillation for
200 water extraction. In July, 1 m pits were dug. From the vertical pit, samples were collected in 20
201 cm increments to determine the depth of tree water access. After samples were collected, the
202 larger rocks and soils were used to fill the pits. We assumed that the isotopic signature of soil
203 water below 40 cm would be similar throughout the growing season and would be representative
204 of deep soil water. Soil samples were stored in a freezer at -18°C until cryogenically distilled.

205 Precipitation samples were collected cumulatively over individual field collection days where
206 precipitation was present (Table 1). Snow from a late spring event was collected on June 11th to
207 represent snow water isotopic composition during the sublimation and melt period of early 2021.
208 Another snow event was collected on October 11th during an active snowfall. A rain event was
209 collected on September 10th. Groundwater and stream samples were collected from the creek 241
210 watershed in early October 2021 at the beginning of the seasonal hydraulic recovery period
211 (Table 1). Groundwater was collected using a hand pump. Groundwater and stream samples
212 were collected at the end of the growing season as stream beds were dry and groundwater was
213 inaccessible during the dry period. Once the well had been pumped and cleared, 10 mL glass test
214 tubes were rinsed with ground water three times before being filled. Precipitation, groundwater
215 and stream samples were collected into 10 mL glass test tubes, sealed with Parafilm® and foil,
216 and stored in a fridge at 4°C .

217 2.4 Cryogenic extraction and isotopic analysis

218 Before extraction, branch samples remained sealed and were weighed in the glass test tubes used
219 for field collection. Branches remained in the test tubes until cryogenic distillation was complete
220 to ensure that any liquid water lost from the branch to the test tube was contained in the extract.
221 Soils samples were mixed in the Ziploc® bag, weighed, and transferred to a glass round bottom
222 flask. For stable isotope analysis, water was extracted from stem and soil samples using
223 cryogenic distillation (Orlowski et al., 2013; Percy et al., 2012). The test tube and branch
224 sample segment of the line was immersed in liquid nitrogen for 10 minutes until frozen
225 (Chillakuru, 2009). Soil sample size for extraction was roughly determined based on the
226 expected moisture of the frozen sample and soil moisture readings from continuous
227 measurements in the field. Soils were frozen for 45 minutes in a 500 mL round-bottom flask
228 using a dry-ice and 95% ethanol mixture before pumping out the air. Frozen samples were
229 pumped down to 60 mTorr, not disturbing the sample (Tsuruta et al., 2019). The vacuum-sealed
230 extraction unit was detached from the pump and transferred to a boiling water bath; the
231 extraction tube was submerged in liquid nitrogen. Branch samples were set to distill for 1 hour
232 and soil samples for 2 hours or until the tubing was clear to ensure all mobile and bound source
233 water was extracted (Orlowski et al., 2013; Tsuruta et al., 2019; Vargas et al., 2017; West et al.,
234 2006). As reviewed by Allen & Kirchner (2022), the cryogenic vacuum distillation of water from
235 plant tissues and soils can bias measurements of $\delta^2\text{H}$, the amount depending on species and soil
236 type. In contrast, bias in $\delta^{18}\text{O}$ values is close to zero (Allen & Kirchner, 2022). Reported biases
237 in $\delta^2\text{H}$ average about -6.1‰ for xylem water and -4‰ for water extracted from sandy soils, such
238 as the soils sampled here, which are of similar magnitude. Furthermore, all sources we identified
239 had a difference in $\delta^2\text{H}$ greater than 4‰ (with the minimum distance being 14‰ between
240 groundwater samples and deep soil water), minimizing any major effects on partitioning
241 calculations.

242 The volume of branch water extracted ranged from 1 to 3 mL depending on the size of the
 243 branch sample. Total extracted water varied dependent on the mass of the initial sample. The
 244 volume of soil water extract ranged from 1mL to 7 mL depending on the size of the sample
 245 prepared for extraction. Samples were also weighed after extraction and compared to oven dried
 246 samples to ensure distillation was complete. Water extracted from branch and soil samples
 247 accounted for $47.9 \pm 3.2\%$ and $9 \pm 6\%$ of mean sample weight \pm standard deviation.

248 All samples were pipetted and sealed into glass vials with screw tops and shipped to the
 249 University of California Davis Stable Isotope Facility (Davis, CA, USA) for ^{18}O and ^2H analysis
 250 using headspace gas equilibration on a GasBench-II interfaced to a Delta Plus XL isotope-ratio
 251 mass spectrometer (Thermo-Finnigan, Bremen, Germany) normalized to a range of secondary
 252 reference waters calibrated against three IAEA standard waters. Precision was less than or equal
 253 to 2.0‰ for $\delta^2\text{H}$ and 0.2‰ for $\delta^{18}\text{O}$. Results were returned in the "delta" notation expressing the
 254 isotopic composition of each sample as a ratio in parts per thousand, relative to VSMOW
 255 (Vienna-Standard Mean Ocean Water) where:

$$256 \quad \delta(\text{‰}) = \left(\frac{R_{\text{Sample}}}{R_{\text{Standard}}} - 1 \right)$$

257 Sample extract was situated in an isotope biplot and compared to the global meteoric water line
 258 (GMWL) along with a local meteoric water line for the Okanagan Valley (OMWL) ($\delta^2\text{H} = 6.6$
 259 ($\delta^{18}\text{O} - 22.7$) and local evaporative line (LEL) ($\delta^2\text{H} = 5 (\delta^{18}\text{O} - 48.4)$) calculated for the
 260 Okanagan Valley by Wassenaar et al. (2011). The LEL is a linear regression that indicates the
 261 departure of water sources from the OMWL to indicate the degree of evaporative processes
 262 fractionating the isotopic composition of water sources or variance in the isotopic composition of
 263 seasonal precipitation events.

264 One extreme outlier of B1C at the 20 cm depth was removed before analysis; the high $\delta^2\text{H}$ and
 265 $\delta^{18}\text{O}$ values were likely due to contamination or incomplete cryogenic distillation. To test the
 266 variance between thinning treatments, block replicates, dates collected, and soil depth, we first
 267 tested the normality of the subsets using the Shapiro-Wilk test and found that all subgroups were
 268 approximately normally distributed. Repeated measures ANOVAs were used to compare effects
 269 of date and treatment on $\delta^2\text{H}$ and $\delta^{18}\text{O}$ in branches, soils and groundwater to determine if
 270 changes in lodgepole pine uptake patterns occurred over time, if soil signatures varied between
 271 different depths (0-100 cm and groundwater) and densities, and if thinning juvenile stands
 272 changed seasonal shifts. All statistical analysis was conducted in R Studio (version 1.3.1073)
 273 using the appropriate tests to determine site distinctions and seasonal variability in depth to
 274 uptake (RStudio Team, 2020).

275 2.4 MixSIAR model scenarios

276 Process-based models (PBM) with a Bayesian approach include integrating other processes or
 277 existing information as priors allowing for a more informed approach than a simple linear model
 278 (Ogle et al., 2014). To accurately partition potential lodgepole pine water sources, we used the
 279 MixSIAR modeling package, a Bayesian mixing model (BMM) based on the Markov Chain
 280 Monte Carlo method (MCMC) (Langs et al., 2020; Stock, 2013/2022, p. 201; Stock et al., 2018;
 281 Wang et al., 2017; Wang et al., 2019). The MixSIAR modeling package was selected over
 282 previous iterations of the dual-isotope BMM (SIAR and Simmr) and other partitioning models

283 because of the accuracy in the analysis of covariates and the ability of the model to include
 284 source-specific uncertainties and
 285 discrimination factors (Stock et al., 2018;
 286 Wang et al., 2017). We partitioned
 287 potential water sources for five different
 288 scenarios using a combination of single
 289 and dual isotope approaches and different
 290 potential sources: scenario 1 – single
 291 isotope $\delta^{18}\text{O}$ two sources 5 cm and 35 cm
 292 depth; scenario 2 – single-isotope $\delta^2\text{H}$ two
 293 sources 5 cm and 35 cm depth; scenario 3
 294 – dual-isotope two sources 5 cm and 35
 295 cm depth; scenario 4 – dual isotope three
 296 sources 5 cm, 35 cm and 45-100 cm
 297 depth; scenario 5 – dual isotope three
 298 sources 5 cm, 35-100 cm and
 299 groundwater; and scenario 6 – dual
 300 isotope four sources 5 cm, 35 cm, 45-100
 301 cm and groundwater. In scenarios using
 302 deep soil water (35-100 cm depths), the
 303 isotopic composition was calculated as a
 304 weighted average between seasonally
 305 collected soil water from depth 35 and
 306 average soil water at depths collected in
 307 202 cm intervals during the early growing season ($n=38$ per season). There were no source
 308 concentration dependencies, and the discrimination was set to zero for both isotopes in the
 309 analysis. The run length of the Markov
 310 chain Monte Carlo (MCMC) was set to
 311 ‘normal’ (chain length = 100,000; burn
 312 =50,000; thin = 50; chains = 3). The
 313 Gelman-Rubin and Geweke diagnostic
 314 tests included in the model package were
 315 used to determine convergence (Gelman-
 316 Rubin score < 1.01). Scenarios that did
 317 not converge were run again with a longer
 318 runtime (chain length: 300,000; burn:
 319 200,000; thin: 100; chains = 3). No priors
 320 were used, so each water source was
 321 considered equally ($\alpha = 1$).

322

323 3. Results

324 3.1. Climate and soil water content

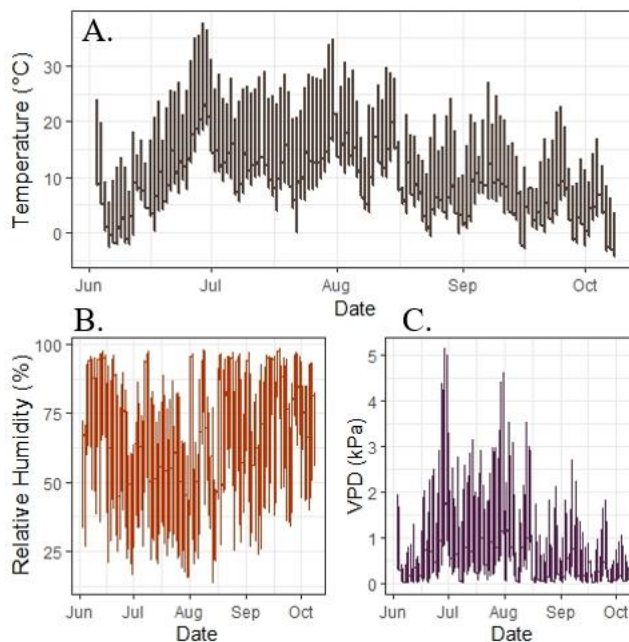


Figure 2 15-minute measurements of A. atmospheric temperature ($^{\circ}\text{C}$), B. Relative humidity (%), and C. vapor pressure deficit (VPD) (kPa).

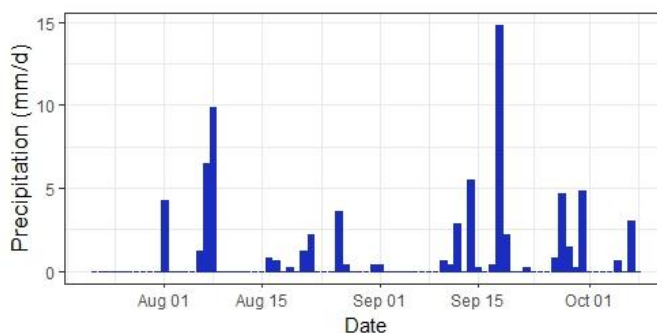
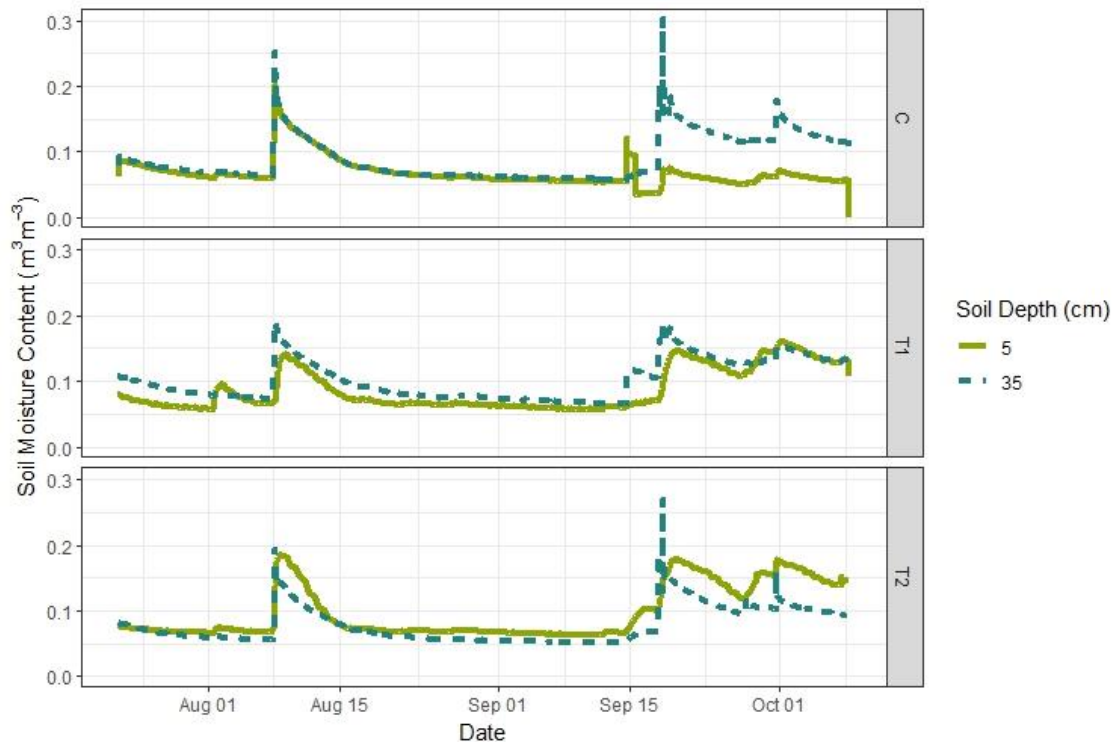


Figure 3 Rainfall (mm/d) from July 22 to October 8, 2021.

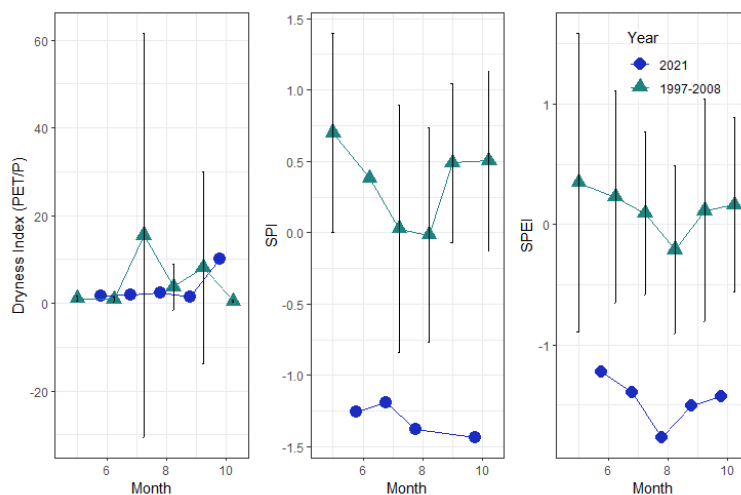
325 The ambient temperature peaked in the moderately thinned plot (T1) on June 29th with a
 326 maximum temperature of 36.3°C in an abnormally hot and dry summer (Figure 2). Relative
 327 humidity and VPD recorded in T1 showed the most variability and highest evaporative capacity
 328 during July. Atmospheric water vapor was higher in late September and October when
 329 precipitation was more frequent, and the watershed began to exhibit traits of hydrologic recovery
 330 (Figure 3). One indication of increased water availability was increased soil moisture at 5 cm and
 331 35 cm depths and more groundwater recharge in October (Figure 4). There was 17.5 mm of
 332 precipitation from September 16th to 18th that infiltrated to at least 35 cm below the soil surface
 333 along with subsequent rainfall events that likely infiltrated past the 35 cm sample depth changing
 334 the isotopic composition of deep soil water from what was measured during the deep pit
 335 sampling in July.



336

337 *Figure 4 Average in-situ continuous measurements (15-minute interval) of soil water content (m³/m³) from the*
 338 *control, moderately thinned, and heavily thinned stands in Block 1.*

339 Rainfall events recorded at a nearby long-term research station between June to October from
 340 1997-2008 represented
 341 approximately 30.1% of annual
 342 precipitation (Winkler et al.,
 343 2021). Over the 2021 study
 344 period, there was 147.8 mm of
 345 rainfall, while the mean summer
 346 rainfall from 1997 to 2008 was
 347 232.5 mm, and most of the
 348 rainfall occurred in the early
 349 growing season. SPI and SPEI
 350 were significantly lower in 2021
 351 than the mean historical range
 352 (Figure 5). Although there was
 353 precipitation and the beginning
 354 of hydraulic recovery in
 355 October, drought conditions
 356 persisted. Drought conditions of
 357 the study site reflected the
 358 drought conditions of the region
 359 as reported by Agriculture and Agri-Food Canada from June to August 2021 in moving from
 360 severe (level 2 drought) to exceptional (level 4) before recovering in September (Canada, 2014:
 361 [https://agriculture.canada.ca/en/agricultural-production/weather/canadian-drought-](https://agriculture.canada.ca/en/agricultural-production/weather/canadian-drought-monitor/drought-analysis)
 362 [monitor/drought-analysis](https://agriculture.canada.ca/en/agricultural-production/weather/canadian-drought-monitor/drought-analysis)).



354 *Figure 5 From left to right: dryness index (monthly PET using the*
 355 *Thornthwaite method divided by mean monthly precipitation), standard*
 356 *precipitation index (SPI) with a 3-month period, and standardized*
 357 *precipitation evapotranspiration index (SPEI) with a 3-month period.*

363

364 3.2. Water stable isotopes

365 The biplot of sample isotopic composition shows the distribution and effect of isotopic
 366 fractionation on source water isotope ratios of samples collected during the 2021 field season.
 367 Field collected samples were compared to the Okanagan Meteoric Water Line (OMWL)
 368 (Wassenaar et al., 2011). The slopes for branch and soil water were less steep than the OMWL,
 369 and the intercepts more negative, indicating that evaporative fractionation contributed to the
 370 isotopic composition of these pools at the UPC (Figure 6). Soil samples seemed to follow the
 371 LEL produced by Wassenaar et al. (2011) for the region indicating similar evaporative
 372 fractionation effects. Branch water more closely following the OMWL than soils, suggesting that
 373 most samples consisted of water that was accessed from deeper in the soil profile and had

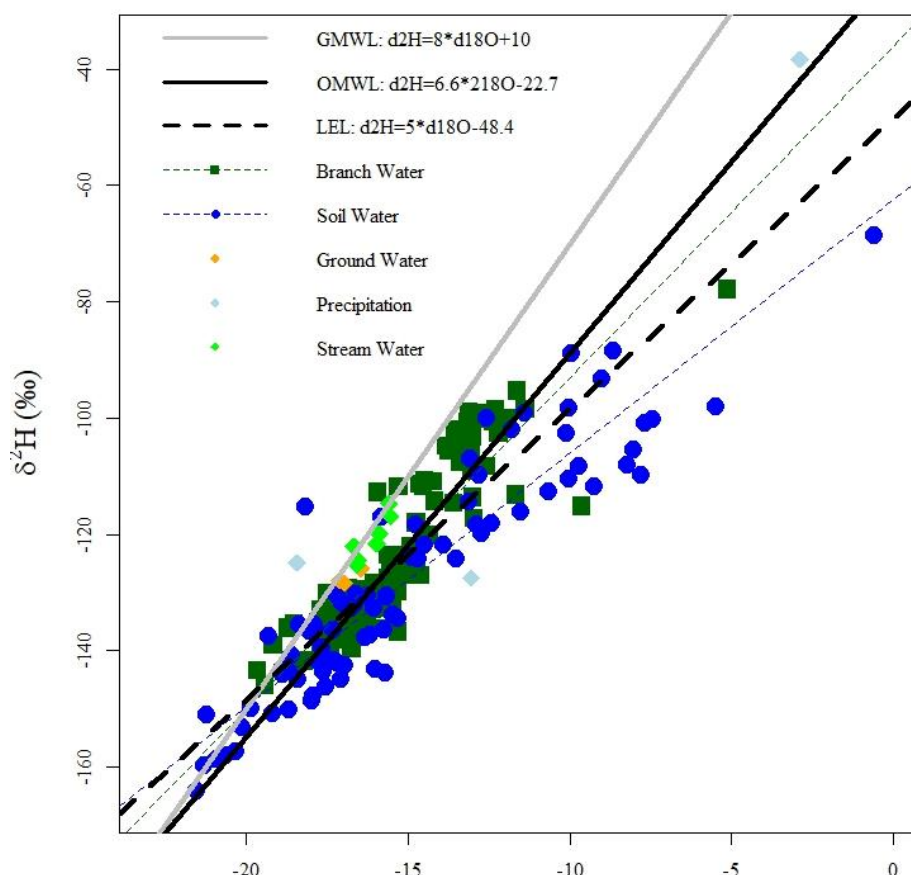


Figure 6 Biplot of $\delta^{18}\text{O}$ and $\delta^2\text{H}$ including all samples collected over the study period from branches, soils, streams, groundwater, and precipitation events outlined in Table 1 along with the global meteoric water line (GMWL), Okanagan Meteoric Water line (OMWL) and the Local Evaporative Line (LEL) developed by Wassenaar et al. (2011), and linear regressions for branch water and soil water.

374 infiltrated past the evaporative front. Precipitation samples collected during the field season fell
 375 along the OMWL (Wassenaar et al., 2011). The $\delta^2\text{H}$ and $\delta^{18}\text{O}$ of the June 11th rainfall event were
 376 -127.5‰ and -13.03‰ , respectively. The September rainfall event was much more enriched with
 377 a $\delta^2\text{H}$ of -38.4‰ and $\delta^{18}\text{O}$ of -2.89 (Figure 6). The snowfall collected on October 7th more
 378 closely resembled the lighter, colder, June precipitation event.

379 3.2.1. Soil moisture and seasonal water composition

380 Soil moisture probes and percent soil water
 381 content from samples collected for isotopic
 382 analysis were compared between treatments
 383 and deployment depths. Water content of soil
 384 samples was highest in June (21.5% at 5 cm
 385 and 21.6% at 35 cm) because of high snow
 386 melt and early spring precipitation, while
 387 soils were driest in September (6.32% at 5
 388 cm and 6.19% at 35 cm). Continuous soil
 389 moisture measurements showed that soil
 390 water began to increase in mid-September as
 391 precipitation became more frequent, daily
 392 solar radiation decreased, and water
 393 percolated into deeper soil layers. There were
 394 significant differences in the continuously
 395 measured soil moisture by depths,
 396 treatments, and month, respectively (5-35
 397 cm) (Depth: $F=3545.9$, $p<2e-16^{***}$)
 398 (Treatment: $F=1883.3$, $p<2e-16^{***}$) (Month:
 399 $F=3359.8$, $p<2e-16^{***}$) (Figure 7), but soil
 400 water content of samples for isotopic
 401 analysis only varied significantly by month
 402 (August – October) ($F=22$, $p<5.4e-9^{***}$).

403 Soil isotopic results were broken into two
 404 datasets to analyze the variation in isotopic
 405 composition over time and between

406 treatments, and then a profile of isotopic variance with depth was constructed. Soil water δ^2H
 407 and $\delta^{18}O$ varied significantly by depth (δ^2H : $p=2.57e-6^{***}$; $\delta^{18}O$: $p=2.45e-7^{***}$), being higher
 408 in the shallow soils than deeper in the profile (Figure 7.A. and 7.C.). δ^2H varied significantly
 409 across months ($p=2.72e-5^{**}$), but not between July and September and September and October.
 410 $\delta^{18}O$ also varied significantly across months ($p=1.5e-5^{**}$) except when directly comparing July
 411 to October and September to October. Despite treatment differences in soil moisture (Figure 4),
 412 there were no statistically significant treatment differences in the isotopic composition of soil
 413 water at either depth. In June, the mean soil water $\delta^{18}O$ at 5 cm was $-16.8\pm 2.57\%$ while the δ^2H
 414 was $-136.7\pm 13.6\%$; at 35 cm, the $\delta^{18}O$ was $-19.2\pm 1.52\%$ and δ^2H was $-149.2\pm 9.6\%$. Both $\delta^{18}O$
 415 and δ^2H increased more during the growing season at 5 cm than at 35 cm, and with more
 416 variability (Figure 7). In September, $\delta^{18}O$ and δ^2H at 5 cm were -8.75% and -106.23 and at 35
 417 cm were -14.71% and -127.64 respectively suggesting that soil isotopic composition nearer the
 418 soil surface follows trends in precipitation samples, being most enriched with O^{18} . By October,
 419 $\delta^{18}O$ and δ^2H at 5 cm reflected more recent precipitation events indicating that water availability
 420 in shallow soils began to increase.

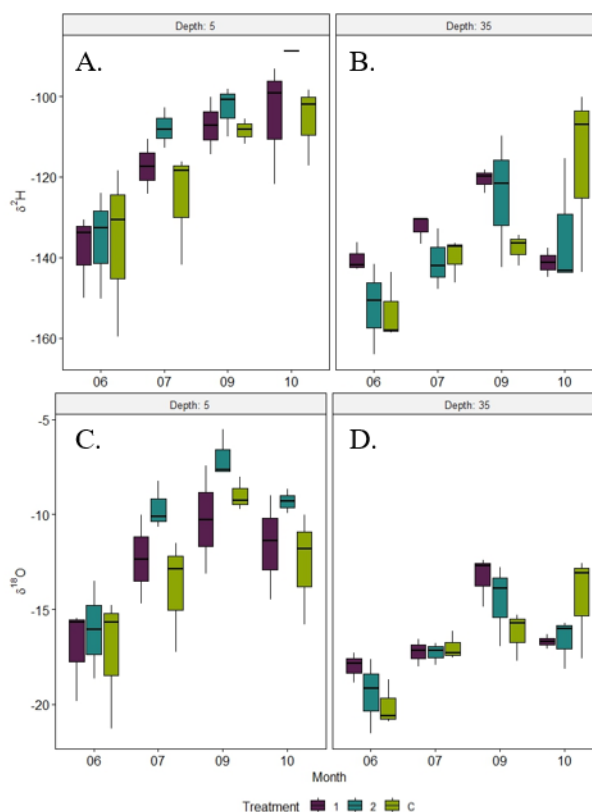


Figure 7 Soil water δ^2H (top) and $\delta^{18}O$ (bottom) at 5 (left) and 35 cm (right) depths collected repeatedly over the growing season from each treatment and block.

421 From the isotopic soil
 422 profile, there were three
 423 significant groupings of
 424 isotopic composition
 425 ($p < 0.05$): shallow soil
 426 water (5-20 cm), deep soil
 427 water (35-100 cm), and
 428 groundwater. Mean
 429 groundwater collected at
 430 the end of the growing
 431 season most closely
 432 resembled spring and fall
 433 snowfall events. The mean
 434 $\delta^{18}\text{O}$ of groundwater was -
 435 $16.82 \pm 0.34\text{‰}$, which
 436 resembles that in the soil
 437 profile, but mean $\delta^2\text{H}$ was
 438 slightly higher than soil
 439 water ($n=4$). This isotope
 440 fractionation may be due to

441 interactions with bound soil water and soils as the water infiltrates through the vadose zone, but
 442 the spread of values as potential uptake sources was greater than any predicted bias from
 443 cryogenic vacuum extraction therefor groundwater was included in the model as a isotopically
 444 distinct potential source for lodgepole pine water use (Allen & Kirchner, 2022; Vargas et al.,
 445 2017).

446 The more negative values for both $\delta^{18}\text{O}$ and $\delta^2\text{H}$ with soil depth indicate that snow melt is the
 447 main source of water to the deep unsaturated zone and that enriched summer precipitation is not
 448 infiltrating deeper soil layers (Figure 8).

449

450 3.2.2. Isotopic variability in branch xylem water

451 Branch xylem for each treatment across the three blocks and the adjacent mature stand were
 452 compared for each sampling period. All treatments closely resembled the mature stand in both
 453 $\delta^{18}\text{O}$ and $\delta^2\text{H}$. There were no statistically significant differences in both $\delta^{18}\text{O}$ and $\delta^2\text{H}$ of xylem
 454 water across thinning treatments; there was, however, significant variation over time ($\delta^{18}\text{O}$:
 455 $F=24.8^*$; $\delta^2\text{H}$: $F=146.6^*$). More specifically, $\delta^{18}\text{O}$ and $\delta^2\text{H}$ of xylem water varied by month for
 456 all months collected except for between June and September and July and September (Figure 9).
 457 Because the isotopic composition of xylem water showed significant change over the growing
 458 season but did not follow the same seasonal trends as soil water, the trees were likely changing
 459 their primary water source within the soil profile.

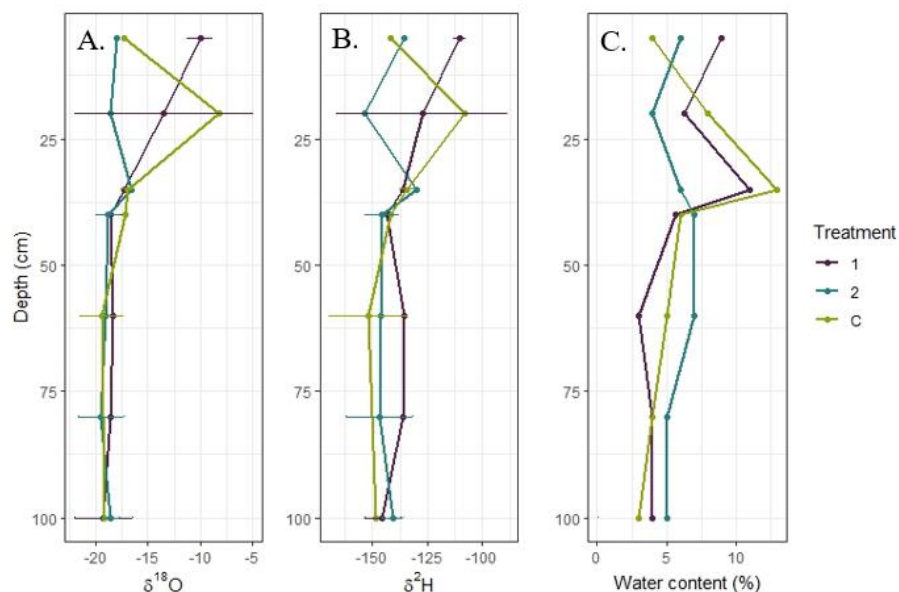


Figure 8 Vertical isotopic profiles and soil water content from treatments in Block 2 and samples collected in mid-July.

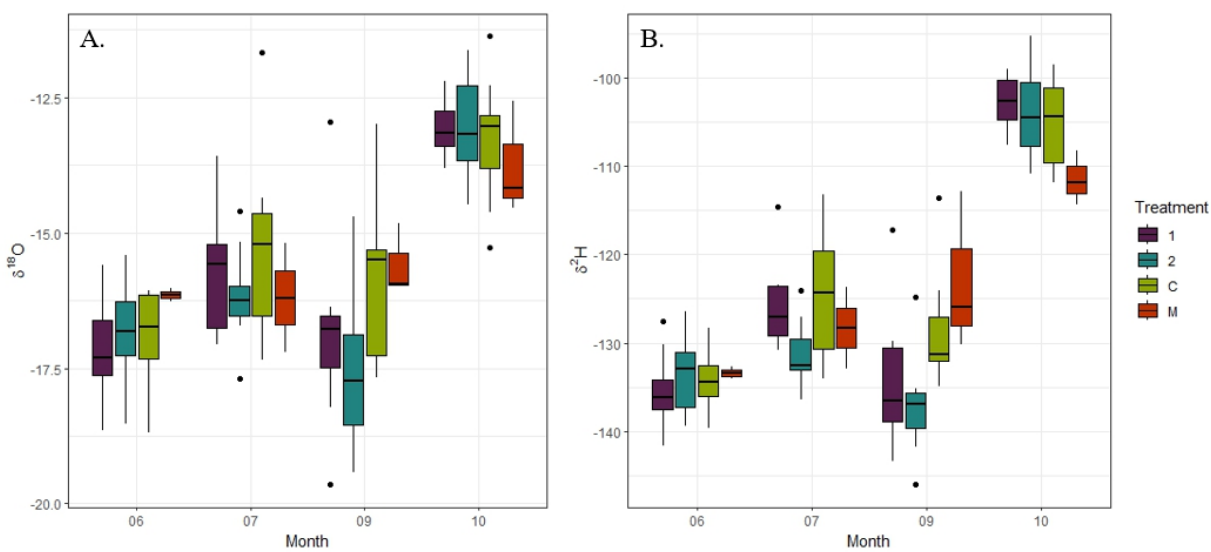


Figure 9 Branch A. $\delta^{18}\text{O}$ and B. $\delta^2\text{H}$ by month and treatments (control (C), lightly thinned (T1), heavily thinned (T2), and mature (M)).

460 3.3. Partitioning xylem source water and seasonal fluxes using MixSIAR

461 With a “normal” runtime (chain length: 100,000; burn: 50,000; thin: 50; chains: 3), scenarios 1, 2
 462 and 6 approached the Gelman-Rubin diagnostic, which indicates convergence when the variable
 463 is less than 1.05 (Table S2). Scenarios 4 and 6 were rerun with the run time set to “long” (chain
 464 length: 300,000; burn: 200,000; thin: 100; chains: 3). The Gelman-Rubin diagnostic variable for
 465 scenario 4 was 120, and for scenario 6 was 17, meaning scenario 6 was closer to convergence
 466 (>1.05). Results of scenario 6 indicate that, in June, trees in each treatment acquired the most
 467 water from the 5 cm depth (C: 76%; T1: 77%; T2: 79%) (Figure 10). In July, shallow soil water
 468 was still the primary source for T1 and T2 at 47% and 61%, but C had 55% water from 45-100
 469 cm deep and only 33% from 5 cm below the surface. By September, all treatments acquired less
 470 than 15% of tree water from shallow soil. Lodgepole pine water use in treatments 1 and 2 was
 471 composed of approximately 48% and 54% from around 35 cm, while 72% of water in control
 472 stand trees was from 35-100 cm. By October, although SPEI results indicate more moisture and
 473 less evaporative demand, scenario six indicated that all three treatments had most water uptake
 474 from below 45 cm in the soil profile (Figure 10). Results of the MixSIAR model support findings
 475 of branch water stable isotope trends over the growing season where the branch water started
 476 with mean $\delta^{18}\text{O}$ and $\delta^2\text{H}$ values of $-16.9 \pm 0.89\text{‰}$ and $-134.37 \pm 3.8\text{‰}$ in June, becoming slightly
 477 more enriched in July. There was a shift to a source with a higher concentration of lighter
 478 isotopes in September. Branch water was most enriched with heavy isotopes in October, like
 479 shallow soil water, with mean $\delta^{18}\text{O}$ and $\delta^2\text{H}$ of $-12.9 \pm 1.76\text{‰}$ and $-103.8 \pm 7.0\text{‰}$, respectively.
 480 However, the MixSIAR model does not account for potential changes in the isotopic
 481 composition of water from precipitation events from mid-September to mid-October.

482

483 The branch water in October was more enriched in heavy oxygen isotopes for each treatment
 484 than soil water at a depth of 35 cm and was more isotopically similar to soil water at 5 cm.
 485 Deuterium also followed a similar trend.

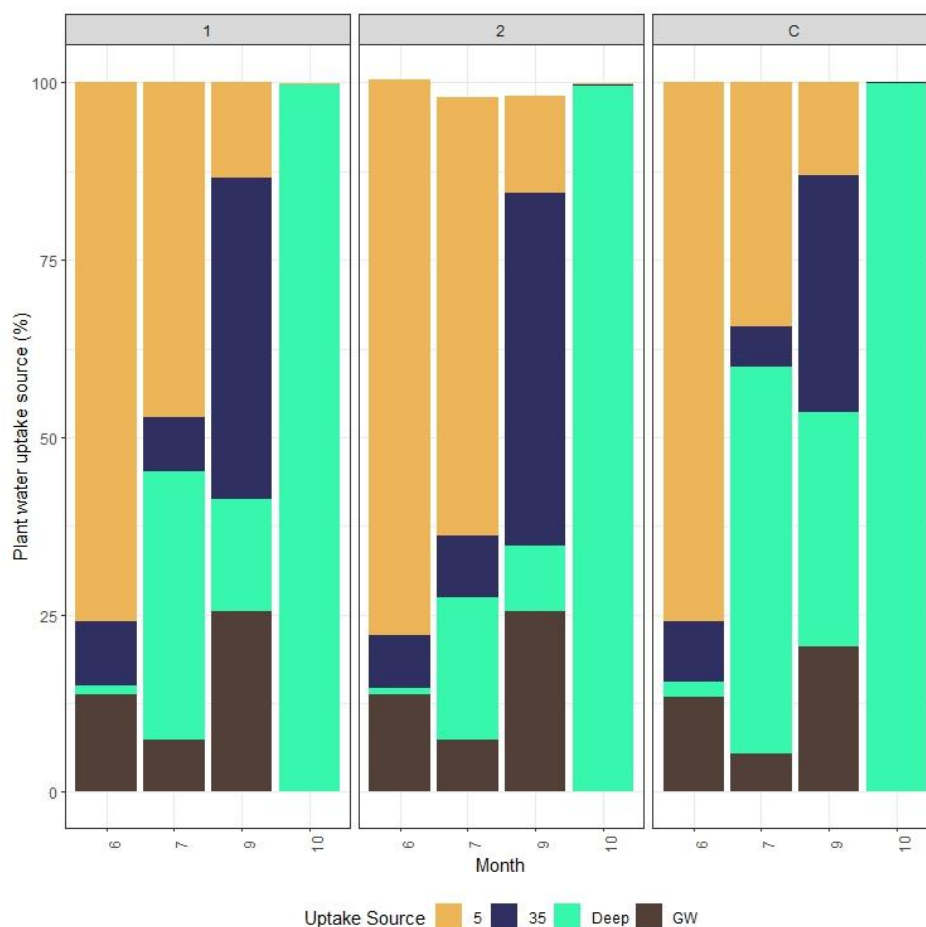


Figure 10 Partitioned relative contribution of different sources of water in the soil profile by the MixSIAR model of scenario 6 with long runtime.

486

487 4. Discussion

488 4.1. Seasonal variability in soil water

489 Deep soil water showed mixed gradient of older, more depleted, water molecules deeper in the
 490 profile indicating that deep soil water mainly originates from spring snowmelt during the
 491 summer months. Low intensity and less frequent summer precipitation events are evaporated out
 492 of the shallow soil layers and do not infiltrate past the evaporative front to recharge the
 493 unsaturated zone or groundwater. Although there was not a statistically significant difference in
 494 the depth to water uptake by thinning treatments, there was increased evaporative enrichment, or
 495 a higher concentration of oxygen-18, in the shallow soils of the heavily thinned stand (Figure
 496 4.C.). The muted enrichment of $\delta^{18}\text{O}$ around 35 cm depth in the soil indicates a mixing of the
 497 left-over summer precipitation with older and lighter water. Our results do not indicate that

498 differences in soil exposure canopy coverage were effective enough to significantly affect the
499 isotopic composition of soil water below 5 cm in depth.

500 4.2. Seasonal lodgepole pine water use

501 Literature utilizing stable water isotopic analysis to determine plant preferential water uptake in
502 arid regions indicates that vegetation can utilize precipitation despite the temporal origin
503 (Andrews et al., 2012; Brinkmann et al., 2019; Ehleringer et al., 1991). Seasonal water
504 availability depends on precipitation, soil water holding capacity and drainage, and evaporative
505 loss (Gibson & Edwards, 2002; Kleine et al., 2020; Stumpp et al., 2018). Based on the seasonal
506 shift in the isotopic composition, soil water at a depth of 5 cm was more enriched with heavier
507 isotopes over the growing season than at 35 cm due to more evaporative isotopic fractionation
508 near the soil surface and a lack of rainfall intense enough to drive precipitation deeper into the
509 soil profile before September 16, 2021 (Figure 3). The effect of evaporative enrichment of the
510 near surface soil water was most obvious in July and September in the heavily thinned stand
511 (T2). However, variability in branch isotopic composition did not follow the same trends. Our
512 results indicate that lodgepole pines access water from multiple depths in the soil profile.
513 Regardless of depth and forest density, spring snowmelt is the main source for lodgepole pines as
514 it infiltrates through the vadose zone.

515 The MixSIAR isotopic partitioning model results from each of the six scenarios indicated a
516 seasonal shift in the depth to water uptake of lodgepole pine, regardless of changes in stem
517 density, over the growing season. At the beginning of the growing season, when snow meltwater
518 is more available at shallow depths and beginning to infiltrate through the soils, lodgepole pines
519 obtain most of their water from snow melt in shallow soils with small contributions from other
520 potential sources (< 25% of June water uptake in all treatments). Then, in July, the trees in the
521 control treatment were using less shallow soil water (34.3% of plant water uptake from 5 cm
522 below the soil profile) whereas the moderately thinned and heavily thinned plots maintained a
523 greater proportion of shallow water uptake (47.1% and 61.5% respectively). The mean $\delta^{18}\text{O}$ and
524 $\delta^2\text{H}$ of branch water from each treatment in September had a higher concentration of lighter
525 stable water isotopes than in July and a larger proportion of tree water was from 35-100 cm deep
526 in the soil profile as shallow soils were dry from a lack of rainfall and surface soil evaporation.
527 By September, the control stand was more dependent on deeper soil water and groundwater with
528 only 33.4% of plant water uptake originating from 35 cm in the soil profile, whereas both
529 thinning treatments maintained more than 45% of water uptake from 35 cm in the soil profile. In
530 October, all treatments were completely dependent on deep soil water, but it is likely that the
531 isotopic profile of deep soil water sampled in July skewed the results. It is plausible that the trees
532 began to rely on shallow soil water towards the end of the growing season when soil water
533 content increased. Further research is needed with more intensive sampling of deep soil water
534 during the hydrological recharge period at the end of the growing season and beginning of
535 senescence.

536 Local monitoring close to the study site indicated that the depth to groundwater stayed at least
537 6.5 m below the surface from August through the end of the study period. The continued use of
538 deep soil water even during rewetting in late September and October suggests that the drought
539 conditions suppressed top soil water uptake, but that deeper soil was sufficiently saturated to

540 sustain root water uptake and tree function enough to limit groundwater uptake to less than 30%
541 for all treatments until the beginning of fall precipitation events recharging the saturated zone.

542 Our results indicate that lodgepole pine, like other pine species in arid regions, is flexible in its
543 ability to access deep soil water and can change its depth to water uptake over time depending on
544 water availability (Brinkmann et al., 2018; Grossiord et al., 2017; Kerhoulas et al., 2013; Kleine
545 et al., 2020; Moreno-Gutiérrez et al., 2011; Simonin et al., 2006; Sohn et al., 2014; Wang et al.,
546 2021). Our results of seasonal changes in depth to water uptake by lodgepole pine support the
547 findings of Andrews et al. (2012) on changes in lodgepole pine depth to water uptake in Alberta.
548 Tree species native to arid regions exhibit a variety of adaptations to long-term drought stress
549 and decreased water availability in the soil profile such as deep tap roots, access to the water
550 table, utilizing bound and mobile soil water, fine root mortality, and hydraulic redistribution in
551 ecosystems with low water holding capacity (Amin et al., 2020; Brinkmann et al., 2018;
552 Grossiord et al., 2017; Kerhoulas et al., 2013; Kleine et al., 2020; Langs et al., 2020; Meinzer et
553 al., 2007b; Prieto et al., 2012; Sohn et al., 2016; J. Wang et al., 2017, p. 201).

554 The literature is inconsistent across different biogeoclimatic regions and species with regards to
555 the effects of thinning on stand dynamics that influence inter-tree competition for water
556 resources or changes in depth to water uptake. (Kerhoulas et al., 2013; Moreno-Gutiérrez et al.,
557 2011; Sohn et al., 2016; Wang et al., 2021). We found no significant impact of forest thinning on
558 depth to water uptake. However, our observation of seasonal shifts in depth to water uptake
559 support results of a study on the impacts of thinning intensity on 60-year-old *Pinus halepensis*
560 Mill. in a semi-arid region of Spain which concluded that forest thinning reduced competition for
561 water resources but did not alter water uptake patterns (Moreno-Gutiérrez et al., 2011). Another
562 study on the impact of thinning *Pinus ponderosa* Dougl. on depth to water uptake concluded that
563 water was consistently more isotopically enriched in low-density stands potentially due to
564 prolonged evaporative fractionation in the soil profile, or that understory vegetation utilized
565 more shallow water sources (Kerhoulas et al., 2013). The impact of forest thinning on stand and
566 understory water use is highly variable and dependent on understory growth, canopy structure,
567 water availability, when forest thinning is implemented, and the time since stem removal
568 (Kerhoulas et al., 2013; Moreno-Gutiérrez et al., 2011; Sohn et al., 2016). More research is
569 needed to discern if lodgepole pine relies more on mobile or bound soil water, the extent of
570 lodgepole pine rooting zones, what biogeochemical factors cause seasonal shifts in water uptake,
571 and if severe seasonal drought has a lasting effect on water uptake strategies during hydrologic
572 recovery (Simonin et al., 2007; Vargas et al., 2017).

573 4.3. Impacts of the drought and implications for future climate conditions

574 The 2021 growing season was an abnormally hot and dry period for the interior of British
575 Columbia with severe to exceptional drought conditions. Wang et al. (2019) found that thinning
576 improved water-use efficiency, drought tolerance, and drought recovery by decreasing stand
577 density and improving carbon storage. Our results support the finding that lodgepole pine trees
578 can adjust to prolonged water scarcity, and over-populated stands may be more resilient than the
579 literature has initially indicated. In fact, drought conditions over the study period likely
580 intensified the change in xylem water isotopic composition over the growing season. However,
581 the scope of this study did not include pre-drought seasonal water use patterns nor the impact of
582 forest density on depth to water uptake during drought recovery. Because lodgepole pine depth

583 to water uptake changes during prolonged dry growing season conditions, the trees are more
584 reliant on winter snowpack and spring infiltration to recharge deeper source water below the
585 evaporative front. One experiment on juniper (*Juniperus monosperma* (Engelm.) Sarg.) and
586 piñon pine (*Pinus edulis* Engelm.) investigated the simultaneous stress of increased heat and
587 decreased precipitation on depth to water uptake and found that extreme temperatures and
588 decreased precipitation lead to less reversible embolism and more root death in surface soil
589 levels preventing trees from accessing shallow water sources if precipitation becomes more
590 available late in the growing season (Grossiord et al., 2017). It is becoming more imperative to
591 understand the climatic drivers of lodgepole pine water use and access as mean annual
592 temperatures continue to rise, the seasonal frequency and intensity of precipitation change, and
593 drought conditions become more severe. This study indicates that severe seasonal dryness pushes
594 lodgepole pines to rely more on snowmelt while losing function in shallow roots. Our results are
595 inconclusive in determining the depth to water uptake in September and October because of
596 limited deep soil water measurements. However, increased annual temperatures and more
597 variable precipitation patterns as a part of climate change projections are predicted to drive
598 decreases in winter snowpack and could drive lodgepole pine stands, regardless of stem density,
599 to rely on groundwater influencing water availability and depth to groundwater. These
600 projections could lead to prolonged inter-annual water scarcity along with seasonal water
601 scarcity during the late growing season.

602

603 5.1 Conclusions

604 Lodgepole pine, across all treatments, was able to shift access from shallow soil water at the
605 beginning of the growing season to deeper soil water as drought conditions intensified. The
606 quick-draining and sun-exposed soils of the UPC do not retain small summer precipitation
607 events, and these patterns are intensified in the shallow soil layer of the heavily thinned stand
608 because decreased canopy cover can be directly related to increased soil evaporation. As a result,
609 due to changes in water availability, lodgepole pines shift to a more readily available source in
610 the soil profile (Aranda et al., 2012; Prieto et al., 2012). Our findings support the literature that
611 lodgepole pines are a drought-tolerant species with dimorphic rooting systems allowing them to
612 access water from varying depths in the soil depending on water availability (Andrews et al.,
613 2012; Liu et al., 2011). Despite the ecological plasticity under extreme heat and low summer
614 precipitation conditions, there was no statistically significant variance in depth to water use
615 between the over-populated plots and thinned ones. Both thinned and unthinned lodgepole pine
616 stands were able to access shallow soil water during the early months (June and July), then
617 switched to deeper soil water and a larger proportion of groundwater during September.
618 Although there was not a statistically significant difference in isotopic composition of branch
619 water for the different treatments, our results indicate that decreased stem density may lead to the
620 prolonged use of soil water 35 cm below the surface during prolonged dry periods which would
621 decrease the dependency of lodgepole pine on deep soil water or ground water.

622 Future climate projections indicate hotter growing seasons and less precipitation (Allen et al.,
623 2010). Further investigation is needed to discern how lodgepole pines, under different stand
624 densities, use water during prolonged drought and drought recovery periods (Grossiord et al.,
625 2017; Navarro-Cerrillo et al., 2019; Simonin et al., 2007; Sohn et al., 2016). From our findings,

626 stand density did not prevent lodgepole pines from accessing soil water from various depths, but
627 decreased stem density did result in lodgepole pines using soil water higher in the soil profile for
628 longer under extremely dry conditions. Lodgepole pines indicate a strong level of drought
629 tolerance and ability to access water under extreme heat conditions. If summer precipitation
630 decreases, lodgepole pine in the interior of British Columbia can access deeper soil water from
631 spring snowmelt. However, if snowpack and spring snowmelt begin to decrease, lodgepole pine
632 may need to acclimate to these hydrological shifts.

633

634 *Code and Data Availability:*

635 The codes of the data analysis and plotting are available at <https://github.com/emory->
636 [ce/LodgepolePineWaterUseStrategies2021](https://github.com/emory-ce/LodgepolePineWaterUseStrategies2021) and are available upon request (ece58@nau.edu)

637

638 *Author Contributions:*

639 EE conceived the idea as a part of their Master's research with AW, and performed the
640 extractions with RG. Analysis was primarily conducted by EE with guidance from AW and RG.
641 All authors contributed to the manuscript.

642

643 *Competing Interests:*

644 None of the authors have competing interests.

645

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651

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