



20 **Abstract**

21 China has implemented some air pollution management measures in recent years, yet severe  
22 ozone pollution remains a significant issue. The Southeastern Coast of China (SECC) is often  
23 influenced by hot extremes and tropical cyclones (TCs), and the two can occur simultaneously (TC-  
24 HDs). The compound TC-HDs show a rising trend in the summers of 2014-2019, potentially  
25 affecting ozone pollution. Here, we found that surface ozone concentrations over SECC are more  
26 elevated during extreme hot days than the summer climatology. However, compared to extreme hot  
27 days alone (AHDs), the maximum 8-hour average ozone (MDA8 O<sub>3</sub>) concentration increases by an  
28 average of 6.8 μg/m<sup>3</sup> in the Pearl River Delta (PRD) and decreases by 13.2 μg/m<sup>3</sup> in the Yangtze  
29 River Delta (YRD) during the compound TC-HDs. The meteorological conditions during AHDs  
30 favor the chemical production of ozone over SECC, exhibiting increased temperature and solar  
31 radiation and decreased relative humidity. Relative to AHDs, strong northeasterly winds prevail in  
32 SECC during TC-HDs, suggesting the potential of ozone cross-regional transport between YRD and  
33 PRD. The process analysis in the chemical transport model (GEOS-Chem) suggests that relative to  
34 AHDs, the chemical production of ozone is enhanced in YRD during TC-HDs while horizontal  
35 transport alleviates ozone pollution in YRD but worsens it in PRD through cross-regional transport.  
36 The results highlight the significant effects of cross-regional transport in modulating ozone pollution  
37 in the two megacity clusters during hot extremes accompanied by TC activities, giving insight into  
38 future ozone control measures over SECC under global warming.

39

## 40 1. Introduction

41 Tropospheric ozone (O<sub>3</sub>) is the predominant air pollutant in addition to fine particulate matter  
42 (PM<sub>2.5</sub>) in China (Fu et al., 2019), and it poses significant risks to human health and the ecosystem  
43 (Feng et al., 2015; [Wang et al., 2017](#)). Long-term exposure to high concentrations of ozone could  
44 lead to lung tissue damage and chronic obstructive pulmonary disease, thereby increasing premature  
45 death (Turner et al., 2016; Liu et al., 2018). Chen et al. (2023) demonstrated that the deterioration  
46 of ozone air quality is considered to be the primary factor responsible for the 90% increase in  
47 premature respiratory mortalities in China from 2013 to 2019. Furthermore, ozone can negatively  
48 impact crop yield by inhibiting plant photosynthesis, accelerating crop senescence, and reducing  
49 both yield and quality (Ainsworth et al., 2012; Song and Hao, 2023), ultimately affecting ecosystem  
50 stability (Gu et al., 2023).

51 To deal with the severe air pollution issues, China has implemented several pollution  
52 prevention and control measures, which are prioritized to tackle the problem of particulate matter,  
53 including the “Air Pollution Prevention and Control Action Plan” and the “Three-Year Action Plan  
54 for Winning the Blue Sky Defense Battle” ([P. Wang et al., 2022](#)). These efforts are making  
55 significant progress in lowering PM<sub>2.5</sub> concentrations. [Z. Wang et al. \(2022\)](#) indicated that the annual  
56 concentration of PM<sub>2.5</sub> in China has decreased by 13.41 µg/m<sup>3</sup> from 2014 to 2020. However, it is  
57 worth noting that most cities in China still frequently experienced severe ozone episodes ([Li et al.,](#)  
58 2019). It is demonstrated that surface ozone concentrations in most regions of China have increased  
59 by approximately 20% during the period from 2013 to 2017 (Huang et al., 2018). Also, Wei et al.  
60 (2022) recently found that both surface ozone concentration and ozone pollution days (the maximum

61 8-hour average ozone (MDA8 O<sub>3</sub>) concentrations exceed 160 µg/m<sup>3</sup>) in China exhibited an  
62 increasing trend from 2013 to 2020. Therefore, ozone pollution in China has received widespread  
63 attention over the past few decades (Ashmore, 2005; Gong and Liao, 2019; N. Wang et al., 2022).

64 Changes in tropospheric ozone are closely related to their precursor gases, including nitrogen  
65 oxides (NO<sub>x</sub>) and volatile organic compounds (VOCs) (Fu and Liao, 2014; Jacob et al., 1999; [Li](#) et  
66 al., 2019). Zhang et al. (2023) revealed that an increase in biogenic volatile organic compounds  
67 (BVOCs) increased local ozone production by 23% per year in Hong Kong. [Li](#) et al. (2019) pointed  
68 out that anthropogenic NO<sub>x</sub> emissions in China decreased by 21% from 2013 to 2017, whereas  
69 VOCs emissions changed little. Decreasing NO<sub>x</sub> would increase ozone under the VOC-limited  
70 conditions prevailed in urban China but decrease ozone under rural NO<sub>x</sub>-limited conditions. In  
71 addition to precursor emissions, meteorological conditions can significantly modulate surface ozone  
72 levels (Yin et al., 2019; Zhou et al., 2022). For example, severe ozone pollution commonly occurs  
73 during summertime with high temperatures, low relative humidity (RH), and strong solar radiation  
74 (Dai et al., 2023; Yin et al., 2019). Under low RH conditions, trees close the stomata (pores for  
75 exchanging CO<sub>2</sub> and water vapor), inhibiting the dry deposition of ozone and leading to ozone  
76 accumulation (Kavassalis and Murphy, 2017). Besides, low humidity conditions inhibit ozone  
77 breakdown as the O(<sup>1</sup>D) combines with water molecules (H<sub>2</sub>O) to produce hydroxyl radicals that  
78 promote ozone decomposition in high-humidity environments (M. Li et al., 2021). Wind direction  
79 and wind speed play an important role in the transport and diffusion of ozone (Banta et al., 2011;  
80 Jammalamadaka and Lund, 2006). It is proved that the co-occurred heat waves and atmospheric  
81 stagnation days with low wind speeds favor ozone pollution in the U.S. through promoting the ozone

82 production (Zhang et al., 2018). Reduced clouds and strengthened solar radiation during the hot  
83 extremes favor photochemical ozone production in the troposphere and thus increase the surface  
84 ozone concentrations (P. Wang et al., 2022). Large-scale atmospheric circulation can modulate  
85 surface ozone through changing the meteorological conditions (Yang et al., 2022). For instance, the  
86 stronger western Pacific subtropical high (WPSH) leads to higher temperatures, stronger solar  
87 radiation, lower RH, and less precipitation, which favors the production and accumulation of surface  
88 ozone in northern China (Jiang et al., 2021).

89 Extreme weather events, especially heat waves and tropical cyclones (TCs), have a significant  
90 impact on ozone pollution in eastern China (Lin et al., 2019; T. Wang et al., 2017). High ozone  
91 concentration events are typically associated with high temperatures, which lead to increased  
92 emissions of BVOCs and enhance the chemical formation of ozone (Lu et al., 2019; P. Wang et al.,  
93 2022). [P. Wang et al. \(2022\) indicate that ozone pollution levels are significantly higher during](#)  
94 [extreme heat events compared to other days in the North China Plain.](#) TC activities can substantially  
95 affect tropospheric ozone by affecting the chemical production, transport, and accumulation  
96 processes over the coastal regions of China (K. Meng et al., 2022; Qu et al., 2021; Zhan et al., 2022).  
97 [Shu et al \(2016\) investigated the impacts of Typhoon Utor on an ozone pollution episode over the](#)  
98 [Yangtze River Delta \(YRD\) of China from August 7 to 12, 2013.](#) They found that the peripheral  
99 circulations of the approaching typhoon intensified downward airflow, leading to a short-term local  
100 weather pattern of high temperatures, low humidity, intense solar radiation, and light winds, which  
101 exacerbated ozone pollution. Following the passage of TCs, the lower tropospheric transport of  
102 ozone-rich air and strong photochemical reactions also contribute to amplifying ozone pollution

103 (Zhan et al., 2020). For ozone pollution in Pearl River Delta (PRD), influenced by the strong  
104 downdrafts related to the periphery of TCs, the PRD region typically is dominated by high pressure,  
105 low humidity, and strong solar radiation, leading to the accumulation of ozone (Wei et al., 2016;  
106 Ouyang et al., 2022). Moreover, TC activities are proven to enhance the chemical interactions  
107 between anthropogenic and biogenic emissions, resulting in extreme ozone pollution over both the  
108 YRD and PRD regions (N. Wang et al., 2022).

109 The southeastern coastal region of China (SECC), including the YRD and PRD regions,  
110 experiences both frequent TCs and heatwave events under global warming (W. Wang et al., 2016;  
111 Xiao et al., 2011). And there is a significant concurrent relationship between extreme heatwaves and  
112 TC activity that the peripheral circulations of TCs can promote extreme heatwaves (P. Wang et al.,  
113 2023; Zhang et al., 2024). Such compound hazards are more destructive than the individual extremes  
114 (Matthews et al., 2019). The SECC region has experienced a considerable rise in surface ozone  
115 concentrations during the period from 2013 to 2019 (X. Meng et al., 2022). Although the individual  
116 effects of heatwaves and TCs on ozone have been emphasized, the effects of the compound extremes  
117 of heat waves and TCs on ozone pollution have received limited attention.

118 In this study, we aim to investigate the impacts of the compound hazards of TCs and extreme  
119 hot days (TC-HDs) on ozone pollution in SECC (105-125° E, 10-35° N) based on the long-term  
120 (2014-2019) observational records of air temperatures and TCs, reanalysis datasets of ground-level  
121 ozone concentrations and meteorological parameters, and the GEOS-Chem model simulations. The  
122 mechanisms contributing to these impacts are also analyzed. The study is organized as follows:  
123 Section 2 presents the methods and data utilized, and Section 3 describes the spatial and temporal

124 variations of TC-HDs observed in SECC during 2014-2019. Section 4 describes the impacts of TC-  
125 HDs on ozone concentration in SECC, with a focus on YRD and PRD regions. In Section 5, we  
126 investigate the possible mechanisms for the anomalous ozone concentrations in SECC during TC-  
127 HDs as well as the differences between YRD and PRD by investigating the meteorological  
128 conditions and key chemical and physical process analysis with the GEOS-Chem model. Finally,  
129 Section 6 provides a summary and discussion.

## 130 **2. Data and methods**

### 131 **2.1 Observational datasets**

132 In this work, we focus on the ozone pollution over SECC where is vulnerable to both severe  
133 hot extremes and ozone pollution (Ma et al., 2019; P. Wang et al., 2022). A topographic map of the  
134 SECC including the YRD and PRD regions is shown in Figure 1. The daily maximum air  
135 temperatures ( $T_{\max}$ ) from observational sites in the domain for the period 2014-2019 are provided  
136 by the China Meteorological Administration (CMA). The best track dataset of TCs is also released  
137 by CMA (accessible at [https://tcdata.typhoon.org.cn/zjljsjj\\_sm.html](https://tcdata.typhoon.org.cn/zjljsjj_sm.html)). This dataset includes the time,  
138 location, and intensity of TCs, which records all TCs that have passed through the western North  
139 Pacific (WNP) since 1949 (Lu et al., 2021; Ying et al., 2014).

### 140 **2.2 Reanalysis datasets**

141 The ground-level ozone concentrations for the period 2014-2019 are obtained from the high-  
142 resolution and high-quality ground-level MDA8 ozone data (unit:  $\mu\text{g}/\text{m}^3$ ) for China (ChinaHighO<sub>3</sub>)  
143 at a resolution of  $0.1^\circ \times 0.1^\circ$ , which is one of the series of the high-resolution and high-quality  
144 ground-level air pollutants in China (Wei et al., 2022). The ChinaHighO<sub>3</sub> dataset provides reliable

145 estimates of MDA8 ozone, demonstrated by an average out-of-sample (out-of-station) coefficient  
146 of determination of 0.87 (0.80) and a root-mean-square error of 17.10 (21.10)  $\mu\text{g}/\text{m}^3$  across China  
147 (Wei et al., 2022).

148 Meteorological parameters including 2-meter air temperature (T2m), relative humidity (RH),  
149 surface solar radiation downwards (SSRD), geopotential height (HGT), eastward wind (uwnd),  
150 northward wind (vwnd), mean sea level pressure (MSLP), total cloud cover (TCC), vertical integral  
151 of eastward water vapour flux, vertical integral of northward water vapour flux and vertically  
152 integrated moisture divergence are from the fifth generation of the European Centre for Medium-  
153 Range Weather Forecasts (ECMWF) reanalysis data (ERA5), the latest global atmospheric  
154 reanalysis of ECMWF (Hersbach et al., 2020). The temporal resolution for T2m, SSRD, and TCC  
155 is 6 hours, while that for other meteorological parameters is 3 hours, which are all utilized to  
156 generate daily mean values.

### 157 **2.3 Identifications of extreme events**

158 Severe ozone pollution episodes generally occur in SECC during summertime, coinciding with  
159 frequent heatwaves and TCs (Ji et al., et al., 2024; Shu et al., 2016). In this work, we focus on the  
160 impacts of extreme high temperatures and TCs on surface ozone during summer only (June, July,  
161 and August) in SECC, as outlined in Figure 1. P. Wang et al. (2022) pointed out that most TC tracks  
162 over the western North Pacific passed through the SECC region during their lifetimes in the past  
163 several decades. The hot days (HDs) are defined as dates when the number of high-temperature sites  
164 ( $T_{\text{max}} \geq 35^\circ\text{C}$ ) exceeds more than 40% of the total number of observation stations within the SECC.  
165 We use the proportion of high-temperature sites exceeding 40% as a measure of HDs to ensure



166 adequate samples and extremely hot conditions over SECC during HDs. Note that the anomalies of  
167 surface ozone concentrations exhibit consistent spatial patterns during HDs identified by with a  
168 lower (30%) or higher (50%) criterion for the percentage of high-temperature sites (figures not  
169 shown). In this work, we classify all HDs into two categories: tropical cyclone-hot days (TC-HDs)  
170 and alone hot days (AHDs). TC-HDs are identified when hot days occur over the land regions within  
171 SECC, concurrent with the passages of tropical cyclones through the area; and AHDs indicate hot  
172 days that occur independently.

173 In this work, the least squares method is applied to fit the linear trend and the Student's t-test  
174 is used to test the significance of the trend ( $\alpha = 0.05$ ). A p-value  $< 0.05$  indicates the statistically  
175 significant trend (as shown in Figure 2). The Student's t-test is also used to evaluate the significance  
176 of the differences in ozone concentrations and meteorological variables between HD/TC-HDs and  
177 the long-term climatology.

#### 178 **2.4 GEOS-Chem model**

179 In this study, surface ozone concentrations during 2014-2019 are simulated by using the 3-D  
180 global chemical transport model (GEOS-Chem, version 13.4.1) with a horizontal resolution of  $2^\circ$   
181 latitude  $\times$   $2.5^\circ$  longitude and 47 vertical layers from the surface to 0.01 hPa. The GEOS-Chem has  
182 the fully coupled  $O_3$ - $NO_x$ -hydrocarbon-aerosol chemical mechanisms (Pye et al., 2009; Sherwen  
183 et al., 2016) used to simulate concentrations of gas-phase pollutants (such as ozone) and aerosols.  
184 It is driven by assimilated meteorological data of version 2 of Modern-Era Retrospective analysis  
185 for Research and Applications (MERRA2; Gelaro et al., 2017)

186 The anthropogenic emissions of ozone precursor gases, including CO, NO<sub>x</sub>, and non-methane

187 volatile organic compounds (NMVOCs) are provided by the Community Emissions Data System  
188 (CEDS) (Hoesly et al., 2018). Methane (CH<sub>4</sub>) concentrations are provided by the Global Monitoring  
189 Division (GMD) of the National Oceanic and Atmospheric Administration (NOAA). Biomass  
190 burning emissions are obtained from the Global Fire Emissions Database version 4 (GFEDv4) (Van  
191 Der Werf et al., 2017). The biogenic emissions are estimated with the Model of Emissions of Gases  
192 and Aerosols from Nature (MEGAN) version 2.1 (Guenther et al., 2012). Our previous studies show  
193 that GEOS-Chem model excels in accurately reproducing observed ozone concentrations and spatial  
194 distributions in China (Yang et al., 2022, 2014).

195 Meteorological conditions can modulate surface ozone concentrations through affecting the  
196 physical and chemical processes including chemical production, horizontal advection, vertical  
197 advection, dry deposition and diffusion (Gong and Liao, 2019). In this work, we quantify the  
198 contributions of individual processes to ozone changes using the model outputs of GEOS-Chem  
199 (Gong and Liao, 2019; Ni et al., 2024; Shu et al., 2016). In the GEOS-Chem model, the change in  
200 ozone over time step are caused by chemical production, horizontal advection, vertical advection,  
201 dry deposition and diffusion, which can be used for separating out and quantifying the contributions  
202 of individual physical and chemical processes to the changes in the simulated ozone concentrations  
203 (Gong and Liao, 2019; Ni et al., 2024; Shu et al., 2016).

### 204 **3. Spatial and temporal variations of compound TC-HDs**

205 There are 63 days of TC-HDs over SECC during the summers of 2014-2019, accounting for  
206 around 70% of the total HDs (91 days), with 28 hot days (30%) occurring alone (AHDs) in the past  
207 6 years. The monthly variations of HDs, TCs, and TC-HDs over SECC from 2014 to 2019 are

208 displayed in Figure 2. All the weather extremes show notable intraseasonal variations with relatively  
209 higher occurrences in July and August but lower occurrences in June each year. Besides, both the  
210 occurrences of HDs and TCs demonstrate significant upward trends at the 90% confidence level,  
211 with a rate of 0.31 days per month and 0.33 days per month, respectively. TC-HDs also show an  
212 increasing trend with a rate of increase of 0.2 days/month, consistent with the variations of HDs and  
213 TCs in the past six years. Hence, it is essential to examine the effects of rising TC-HDs on ozone  
214 pollution in the SECC region.

215 Figure 3 demonstrates the spatial features of TC-HDs during 2014-2019. China Meteorological  
216 Administration classifies tropical cyclones into 6 classes based on their intensity, namely Tropical  
217 Depression (TD, 10.8-17.1 m/s), Tropical Storm (TS, 17.2-24.4 m/s), Strong Tropical Storm (STS,  
218 24.5-32.6 m/s), Typhoon (TY, 32.7-41.4 m/s), Strong Typhoon (STY, 41.5-50.9 m/s) and Super  
219 Typhoon (SSTY,  $\geq 51.0$  m/s). TCs are generally stronger before their landfalls, which can reach STY  
220 and even SSTY (Fig. 3a and 3b), consistent with previous findings (Han et al., 2022; Tuleya et al.,  
221 1984). TCs that affect the SECC primarily originate east of 135° E and move westward to the  
222 eastern coastal regions. Around 90% of TCs associated with TC-HDs make landfalls while the  
223 others weaken and dissipate over the ocean. Moreover, the air temperatures over land regions of  
224 SECC during the TC-HDs are generally high, where the average  $T_{max}$  of most sites are beyond 35°C  
225 and reach up to 38°C (Fig. 3a). Compared with the summer climatology from 2014 to 2019, the  
226 average  $T_{max}$  during TC-HDs increased approximately 4°C over most land regions of SECC (Fig.  
227 3b).

228 The temporal variations of the proportion of high-temperature sites ( $T_{max} \geq 35^\circ\text{C}$ ) and the

229 temperature anomalies relative to the summertime climatology averaged for the SECC land regions  
230 along with the movements of TCs are given in Figs. 3c and 3d. Specifically, in Figures 3c&3d, the  
231 colored dots along the movements of TC tracks represent the proportion of high-temperature sites  
232 and the average temperature anomalies relative to the summertime climatology in SECC,  
233 respectively. The two variables demonstrate similar patterns of variation. As TCs approach land  
234 regions, both the high-temperature sites and temperature anomalies are increasing gradually, with a  
235 particularly pronounced temperature increase associated with TCs moving towards the PRD regions.  
236 In contrast, there is typically a decline in both the average air temperatures and the proportion of  
237 high-temperature sites following the landfalls of TCs, which could be related to the strong wind and  
238 rainfall associated with the TCs (Gori et al., 2022; [Zhang et al. 2024](#)).

#### 239 **4. Influences of TC-HDs on surface ozone concentrations in SECC**

240 Extreme high temperatures and the accompanied meteorology such as high-pressure systems  
241 and strong radiation can affect local ozone pollution by enhancing chemical production and/or  
242 accumulation (Gong and Liao, 2019; P. Wang et al., 2022; X. Wang et al., 2009). It is also well  
243 demonstrated that extreme high temperatures and TCs can affect ozone pollution in coastal regions  
244 of China (Ding et al., 2023; Huang et al., 2011; N. Wang et al., 2022). In this work, we further focus  
245 on the impacts of TC-HDs on surface ozone concentrations over land regions of SECC. Figure 4  
246 presents the spatial distributions of ozone concentration anomalies during TC-HDs and AHDs  
247 relative to the summertime climatology from 2014-2019, as well as the differences between the two.  
248 Notable increases in surface ozone concentrations are observed over most land regions of SECC  
249 during the AHDs period, with anomalous MDA8 ozone reaching  $40 \mu\text{g}/\text{m}^3$  and  $20 \mu\text{g}/\text{m}^3$  over the

250 YRD (118-122° E, 30-33° N) and PRD (110-115.5° E, 21.5-24° N) regions, respectively (Fig. 4a),  
251 consistent with previous finding that hot extremes can worsen ozone pollution (P. Wang et al., 2022).  
252 During the compound TC-HDs, most land regions except Hainan Island and the northern part to 33°  
253 N of SECC experience a significant enhancement in surface ozone concentrations, with MDA8  
254 ozone anomalies above 20  $\mu\text{g}/\text{m}^3$  overall (Fig. 4b). Compared to AHDs, ozone concentrations in the  
255 easternmost coastal regions of China including YRD (outlined in Fig. 4c) are suppressed while  
256 ozone concentrations over the southernmost coastal regions of China including PRD (outlined in  
257 Fig. 4c) are promoted during the compound TC-HDs. Comparing to TCs occurring alone, surface  
258 ozone concentrations over most land regions of SECC increase during TC-HDs (Fig. S1), which  
259 should be attributed to the increases in air temperatures. In this work, we investigate the enhanced  
260 surface ozone concentration of PRD and the contrasting depression over YRD during the compound  
261 TC-HDs.

262 The temporal variations of anomalous MDA8 ozone in YRD and PRD regions relative to the  
263 summertime climatology with TC tracks during TC-HDs are shown in Figure 5. In Figures 5a&5b,  
264 the colored dots along the TCs track represents the anomalies of regional mean MDA8 ozone  
265 concentrations for YRD and PRD at that time compared with the summertime climatology for 2014-  
266 2019. Surface ozone concentrations over YRD are abnormally higher when TCs are positioned a  
267 long distance from land regions, whereas ozone concentrations fall dramatically when TCs approach  
268 and make landfall over land regions. And it should be noted that the surface ozone concentrations  
269 are elevated by the TCs moving westward YRD while most of the other TCs cause a decline in  
270 ozone concentrations with their approach to land regions, consistent with previous work (Shu et al.,

271 2016). Zhan et al. (2020) revealed that O<sub>3</sub> pollution episodes in YRD mainly occurred near the 24-  
272 hour warning line before a TC landed on the coastline. The region was impacted by an inland wind  
273 carrying substantial precursor substances from polluted areas, while approaching TCs led to  
274 increased precipitation and strong winds resulting in decreased ozone concentrations. For PRD,  
275 surface ozone concentrations increase noticeably when TCs approaching land regions. Particularly,  
276 TCs heading northeast to YRD favor the increases in ozone concentrations in the PRD whereas the  
277 other TC tracks tend to cause a reduction in surface ozone concentrations following landfalls (Fig.  
278 5b).

## 279 **5. Possible mechanisms underlying the impacts of TC-HDs on surface ozone**

### 280 **5.1 The dominant synoptic circulations**

281 The synoptic meteorological conditions during AHDs and TC-HDs are analyzed to disentangle  
282 the different responses of surface ozone to hot extremes superimposed by TCs. Specifically, we  
283 examine the composites of daily mean air temperature at 2m (T2m), relative humidity (RH), surface  
284 solar radiation downwards (SSRD), mean sea level surface (MSLP), geopotential height at 500 hPa  
285 (500hPa HGT) and 10-meter wind speeds, and total cloud cover (TCC) anomalies relative to  
286 summertime climatology from 2014 to 2019 during AHDs (Fig. 6) and TC-HDs (Fig. 7), as well as  
287 the differences between the two periods (Fig. 8). During the AHDs period, the land regions of SECC  
288 are covered by increased T2m, decreased RH, reduced TCC, and enhanced SSRD (Figs. 6a-d),  
289 which are favorable for the chemical formation of ozone therein (Yin et al., 2019), supporting the  
290 elevated surface ozone during AHDs as shown in Figure 4a. In the mid-upper troposphere, a band  
291 of positive HGT at 500hPa (H500) anomalies predominates over most land regions of China,

292 extending northeastward to Korea (Fig. 6e). This pattern features westerly wind anomalies in the  
293 northern flank and easterly wind anomalies in the southern flank, indicating a westward extension  
294 and strengthening of the western North Pacific subtropical high (WNPSH). These conditions favor  
295 the occurrences of hot extremes over southern China, as discussed by Luo & Lau (2017) and P.  
296 Wang et al. (2018). Besides, it's noticed that negative H500 anomalies and cyclonic circulation  
297 appear over the east of 125°E which are also seen at the surface (Fig. 6f) with anomalous  
298 southwesterly prevails over SECC land regions at the surface (Fig. 6f). Such a low-pressure system  
299 characterizes TC track which finally recurve northeastward (Fig. S2).

300 During TC-HDs, most land regions of SECC are covered by increased T2m, decreased RH,  
301 reduced TCC, and enhanced SSRD (Figs. 7a-d), favoring the enhanced chemical formation of ozone  
302 therein, supporting the elevated surface ozone during TC-HDs as shown in Figure 4b. In the mid-  
303 upper troposphere, positive H500 anomalies cover nearly the whole land region of China with  
304 maximum dominating Korea, accompanied by anomalous anticyclonic circulation. Such changes  
305 characterize the strengthening and westward extension of WPSH (Fig. 7e), favoring more hot  
306 extremes. A dipole pattern of MSLP anomalies can be observed at the surface, with an abnormal  
307 low-pressure center over the South China Sea and another to its east (Fig. 7f). Besides, the land  
308 regions are influenced by weak winds at the surface (Fig. 7f). Particularly, for the decrease MDA8  
309 ozone, we can see that during TC-HDs, Hainan island is covered by decreased T2m, increased RH,  
310 reduced SSRD, and increased TCC (Figs. 7a-d), along with negative H500 and MSLP anomalies  
311 (Figs. 7e&f). Such meteorological conditions in Hainan islands may suppress the chemical  
312 production of ozone and the oceanic winds may clean the air (Fig. 7e). On the other hand, for the

313 north part of SECC, the circulation anomalies favor strengthened southeastern moisture flow and  
314 enhance the convergence of water vapor flux there (Fig. S3), which can lead to increased relative  
315 humidity (Fig.7b). The local higher temperature (Fig. 7a) and humid conditions may favor  
316 convection activities (P. Wang et al., 2019b), characterized increased cloud cover (Fig. 7d) and  
317 decreased surface solar radiation (Fig. 7c). These meteorological conditions can inhibit the local  
318 ozone production and cause a lower ozone concentration in north part of the SECC during TC-HDs.

319 Figure 8 demonstrates the differences in meteorological variables between the periods of TC-  
320 HDs and AHDs. Compared to AHDs, the YRD experiences an increase in T2m of approximately  
321 1°C, while the PRD experiences a decrease in T2m of around 0.5°C during TC-HDs. In the  
322 meanwhile, RH and TCC values are increased (decreased) in PRD (YRD) while SSRD values are  
323 decreased (increased) in PRD (YRD) (Figs. 8a-d). The large-scale circulations in the middle-upper  
324 troposphere show positive H500 anomalies over the north and northeast parts of China including  
325 YRD while southern China including PRD is dominated by negative H500 anomalies, supporting  
326 the increased T2m (decreased) over YRD (PRD). Moreover, a cyclonic circulation controls southern  
327 China with YRD influenced by strong southeasterly winds while PRD is influenced by inland  
328 northeasterly winds (Fig. 8e). The changes in MSLP show a similar pattern to H500 anomalies with  
329 positive anomalies over YRD but negative anomalies over PRD, and an anomalous low-pressure  
330 system is observed over South China Sea, accompanied by strong cyclonic circulation at the surface.  
331 It should be noted that the YRD is influenced by strong easterly winds from the ocean which may  
332 alleviate ozone pollution as shown in Figure 4c. In contrast, PRD is influenced by strong  
333 northeasterly winds from inland which is opposite to the climatological southwesterlies of the



334 summer monsoon circulation, which may favor the accumulation of ozone pollution therein, as  
335 shown in Fig. 4c

336 It should be noted that the meteorological conditions in YRD favor the chemical production of  
337 ozone yet reduced ozone concentrations are observed in YRD during TC-HDs relative to AHDs.  
338 The key physiochemical processes associated with the alleviated ozone pollution over YRD but  
339 worsened ozone pollution over PRD during TC-HDs relative to AHDs are explored with GEOS-  
340 Chem model simulations in the following part.

## 341 **5.2 Process analysis with GEOS-Chem model simulations**

342 In this section, process analysis is conducted in the GEOS-Chem model to quantify the  
343 contributions of each process, including net chemical production, horizontal advection, vertical  
344 advection, and mixing (diffusion and dry deposition) to the anomalous surface ozone concentrations  
345 over PRD and YRD during TC-HDs, respectively. The observed and simulated anomalous surface  
346 MDA8 ozone during TC-HDs relative to the summer climatology mean and AHDs are given in  
347 Figure S4 and the regional averages of surface MDA8 ozone concentrations over YRD and PRD  
348 are summarized in Table 1. GEOS-Chem simulation can reasonably capture the spatial pattern of  
349 anomalous surface MDA8 ozone concentrations during TC-HDs relative to the summer climatology  
350 mean (Figs. S4a and b), and relative to AHDs (Figs. S4c and d). Particularly, compared to AHDs,  
351 the simulated surface MDA8 ozone concentrations are inhibited over YRD but enhanced over PRD  
352 during TC-HDs. As listed in Table 1, compared to AHDs, during TC-HDs, the observed (simulated)  
353 average surface MDA8 ozone concentrations in the YRD decreased by  $13.21 \mu\text{g}/\text{m}^3$  ( $21.94 \mu\text{g}/\text{m}^3$ ),  
354 whereas it increased by  $6.8 \mu\text{g}/\text{m}^3$  ( $7 \mu\text{g}/\text{m}^3$ ) in the PRD. This suggests that the model simulation

355 can reasonably capture the opposite changes in surface MDA8 ozone concentrations over the YRD  
356 and PRD during extreme HDs superimposed by TCs.

357 Figure 9 illustrates the vertical profiles of simulated daily mean ozone concentrations for YRD  
358 and PRD from surface to 500 hPa for the summer climatology, TC-HDs and AHDs during 2014-  
359 2019, respectively. For YRD, surface ozone concentrations are apparently higher during AHDs than  
360 those during TC-HDs and summertime climatology from surface up to the 500 hPa level (Fig. 9a).  
361 This indicates that the enhancements in ozone concentrations during AHDs occur not only at the  
362 surface but also to the middle of the troposphere. In addition, compared with the summertime  
363 climatology, ozone concentrations in YRD during TC-HDs are slightly enhanced from surface to  
364 850 hPa as well as between 600 hPa and 500 hPa while suppressed between 850hPa and 600 hPa.  
365 For the PRD, ozone concentrations during TC-HDs are higher than the summertime mean and AHDs  
366 from the surface up to 500 hPa (Fig. 9b). The ozone concentrations during AHDs are comparable to  
367 (slightly stronger than) the summertime mean below (above) 850 hPa (Fig. 9b) based on the GEOS-  
368 Chem model simulation.

369 We further analyze the main physiochemical processes affecting ozone concentration,  
370 including net chemical production, vertical advection, horizontal advection and mixing (the sum of  
371 dry deposition and diffusion). Figures 10 and 11 show the vertical profiles of the anomaly of each  
372 process during TC-HDs and AHDs, relative to the summertime climatology over the YRD and PRD,  
373 as well as the differences between TC-HDs and AHDs. During AHDs, the increase in ozone  
374 concentrations in the YRD from the surface to 900 hPa is primarily contributed by chemical  
375 production, while transport processes tend to decrease ozone concentration. Between 900 hPa and

376 700 hPa, the mixing process contributes to the increases in ozone concentrations while the transport  
377 processes and chemical productions tend to decrease ozone concentrations. Between 700 hPa and  
378 500 hPa, horizontal transport hampers ozone increase, while chemical production and vertical  
379 transport contribute to ozone enhancement (Fig. 10a). During TC-HDs, the horizontal transport  
380 exhibits a suppressive effect on ozone concentrations in the YRD from the surface up to 500 hPa  
381 whereas the chemical production contributes to increases in ozone concentration. And two processes  
382 overtake the effects of vertical transport and mixing (Fig. 10b). In comparison with AHDs, during  
383 TC-HDs, ozone concentration increases due to chemical production exceed decreases caused by  
384 horizontal transport between the surface and 850 hPa. However, between 850 hPa and 700 hPa, the  
385 decrease in ozone concentration due to horizontal transport outweighs the increase from chemical  
386 production. Overall, compared to AHDs, the chemical production contributes to increased ozone  
387 concentrations in the YRD from the surface up to 500 hPa during TC-HDs, whereas the horizontal  
388 transport tends to lower ozone concentrations. The relatively small differences in physicochemical  
389 processes between AHDs and TC-HDs around 700 hPa in the vertical profile are due to the minor  
390 effects of these processes during both AHDs and TC-HDs (not shown). As listed in Table 2,  
391 compared to AHDs, ozone chemical production plays a predominant role in enhancing ozone  
392 pollution during TC-HDs in the YRD (1.24 Gg O<sub>3</sub>/day), whereas horizontal transport significantly  
393 depresses ozone concentration (−1.15 Gg O<sub>3</sub>/day). As stressed above, the strong and oceanic winds  
394 along YRD can help alleviate ozone pollution. The vertical transport and mixing processes play a  
395 less important role in affecting ozone concentrations in YRD during TC-HDs.

396 For PRD, the horizontal transport primarily contributes to decreased ozone concentrations in

397 the PRD between the surface and 500 hPa during AHDs. The chemical production, vertical transport  
398 and mixing processes exhibit promoting effects on ozone concentrations from surface to 700 hPa  
399 (Fig. 11a). During TC-HDs, horizontal transport predominantly contributes to increased ozone  
400 concentrations over PRD from surface to 900 hPa and between 850 hPa and 500 hPa and the  
401 chemical production increases ozone concentrations between 950 hPa and 800 hPa. In contrast, the  
402 mixing process and vertical transport play a dominant role in reducing ozone concentrations  
403 between the surface to 900 hPa and between 900 hPa to 500 hPa respectively (Fig. 11b). Compared  
404 to AHDs, during TC-HDs, transport processes predominantly contribute to the increases in ozone  
405 concentrations in the PRD from the surface up to 500 hPa. Additionally, between 900 hPa and 800  
406 hPa, chemical production plays a positive role in enhancing ozone concentrations. As listed in Table  
407 3, during TC-HDs compared to AHDs, the horizontal transport process in the PRD exhibits a  
408 significantly stronger enhancing effect on ozone concentrations from surface to 500 hPa (1.81 Gg  
409 O<sub>3</sub>/day) compared to the other three processes. Note that the summer climatology of ozone  
410 concentrations in YRD is significantly higher than that in PRD (Fig. S5) and the cross-section of  
411 wind anomalies demonstrates that strong winds from YRD to PRD exist during TC-HDs (Fig. S6),  
412 implying strong ozone mass transport from YRD to PRD. Thus, the cross-region transport may play  
413 a significant role in worsening ozone pollution over RPD during TC-HDs relative to AHDs.

414 In summary, favorable synoptic patterns during extreme hot days promote the chemical  
415 production of ozone, and thus exacerbate ozone pollution over both YRD and PRD. Compared to  
416 AHDs, the anomalously meteorological conditions during TC-HDs, enhance the chemical  
417 production of ozone in the YRD while the horizontal transport process mitigates ozone pollution

418 therein but worsens ozone pollution in PRD though cross-regional transport, finally resulting in an  
419 increase in ozone in PRD but a decrease in YRD.

## 420 **6. Conclusion and discussions**

421 China has implemented a series of emission reduction strategies to alleviate air pollution, and  
422 PM<sub>2.5</sub> concentrations have decreased significantly while ozone pollution remains a major concern.  
423 Ozone pollution is sensitive to meteorological conditions, especially the extreme weathers that  
424 frequently strike China under global warming. Eastern China is economically developed and  
425 densely populated, with serious ozone pollution. It is also vulnerable to both extreme hot weathers  
426 and TCs during the summer, [which can substantially affect surface ozone concentrations \(Lin et al.,](#)  
427 [2019; T. Wang et al., 2017\)](#). High ozone levels are [generally linked to high air temperatures, which](#)  
428 [increase BVOC emissions and enhance ozone formation \(e.g., Lu et al., 2019; P. Wang et al., 2022\)](#).  
429 [TC activities can modify surface ozone concentrations by affecting the transport and accumulation](#)  
430 [processes, exacerbating pollution in the YRD and PRD \(e.g., Shu et al., 2016; Zhan et al., 2020\)](#).  
431 [The SECC region experiences both frequent TCs and heatwave events under global warming \(W.](#)  
432 [Wang et al., 2016; Xiao et al., 2011\)](#) which has proven to have an intrinsic concurrent relationship  
433 [\(Matthews et al., 2019; P. Wang et al., 2023\)](#). However, [the impacts of the compound extremes of](#)  
434 [hot extremes and TCs on ozone pollution over SECC have received limited attention.](#)

435 In this work, we [systematically](#) investigate the impacts of extreme hot weathers on surface  
436 ozone for the summers of 2014-2019 over SECC coupled with (TC-HDs) and without TCs (AHDs).  
437 The associated synoptic conditions and physicochemical processes are assessed combined  
438 reanalysis dataset with the GEOS-Chem model simulations.

439 Results show that the surface ozone concentrations over most land regions within SECC are  
440 elevated during both TC-HDs and AHDs relative to the climatological mean, however, there are  
441 considerable differences between the [changes in ozone concentrations during TC-HDs and AHDs](#).  
442 Relative to AHDs, the surface ozone concentrations are noticeably decreased in the YRD region but  
443 increased in the PRD region during TC-HDs periods. The meteorological conditions suggest that  
444 though YRD is influenced by higher temperature, lower humidity, and stronger radiation during TC-  
445 HDs than AHDs, it is influenced by strong and sea breezes, which aid in ozone elimination. In  
446 contrast, the PRD is influenced by the strong northeasterly winds, opposite to the climatology, and  
447 may transport ozone pollution from polluted regions. Such a hypothesis is validated with the GEOS-  
448 Chem simulation. Compared with AHDs, among all the physicochemical processes, horizontal  
449 transport plays a crucial role in increasing (reducing) ozone levels over PRD (YRD). Compared to  
450 AHDs, the horizontal transport contributes to  $-1.15$  Gg O<sub>3</sub>/day and  $1.8$  Gg O<sub>3</sub>/day to the net  
451 changes in tropospheric ozone mass from surface to 500 hPa in YRD and PRD during TC-HDs. The  
452 findings will provide significant implications for the control measures of ozone pollution in PRD  
453 and YRD during hot extremes, to consider the significant impacts of TC activities.

454 [Note that throughout the lifetime of tropical cyclones, their positions exert varying influences](#)  
455 [on air temperatures and surface ozone concentrations over the land regions. Extremely high](#)  
456 [temperatures are typically observed before TCs make landfalls, driven by descending motions and](#)  
457 [intensified solar radiation \(Wang et al., 2023\). However, the high temperatures rapidly decline once](#)  
458 [TCs make landfalls, primarily due to the accompanying strong winds and precipitation \(Gori et al.,](#)  
459 [2022; Zhang et al., 2024\). Similarly, surface ozone concentrations over the YRD and PRD](#)

460 [regions tend to be elevated when TCs are at a distance from the land regions but they fall](#)  
461 [dramatically when TCs make landfalls \(Fig. 5\), consistent with previous findings \(e.g., Zhan et](#)  
462 [al., 2020\).](#)

463 Extreme hot weathers are projected to be more frequent and intensified in the future under the  
464 continued global warming (P. Wang et al., 2019a). Moreover, it is projected that TCs and heatwave  
465 compound events are projected to significantly increase in their intensity and frequency over the  
466 Northwest Pacific, causing potentially enlarged population exposures for the southeastern coast of  
467 China (Wu et al. 2022). Therefore, the potential impacts of extreme hot weather and the TC-HDs  
468 on surface ozone over China warrants future efforts. BVOCs are important precursors of ozone, and  
469 their emissions are greatly influenced by weather conditions. As revealed by N. Wang et al. (2022),  
470 the TCs over Northwest Pacific could intensify the chemical interactions between anthropogenic  
471 and biogenic emissions, resulting in extreme ozone pollution over YRD and PRD regions. The  
472 responses of biogenic emissions during hot extremes accompanied by TCs deserve further  
473 investigation.

474

#### 475 **Data availability**

476 The ground-level ozone concentrations are obtained from the high-resolution and high-quality  
477 ground-level daily maximum 8-hour average ozone (MDA8 O<sub>3</sub>) data for China (ChinaHighO<sub>3</sub>,  
478 <https://doi.org/10.5281/zenodo.4400043>). Daily maximum air temperature is provided by the  
479 National Meteorological Information Center of the China Meteorological Administration (CMA,  
480 <http://data.cma.cn/en/>). The best track dataset of tropical cyclones is also released by CMA (<https://>

481 tcddata.typhoon.org.cn/zjljsjj\_sm.html). Meteorological conditions data are derived from the fifth  
482 generation of ECMWF reanalysis data (ERA5, <https://cds.climate.copernicus.eu/>). The GEOS-  
483 Chem model is available at <http://acmg.seas.harvard.edu/geos/>.

#### 484 **Author contributions**

485 C. Qi performed the analyses and wrote the initial draft. P. Wang and Y. Yang conceived and  
486 supervised the study. H. Li performed the GEOS-Chem simulations. P. Wang reviewed and edited  
487 the initial draft. All the authors discussed the results and contributed to the final manuscript.

#### 488 **Competing interests**

489 The authors declare that they have no competing interest.

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735

736 **Table 1.** Observed and simulated averaged surface MDA8 ozone ( $\mu\text{g}/\text{m}^3$ ) in the Yangtze River Delta  
 737 (YRD, 118-122° E, 30-33° N) and the Pearl River Delta (PRD, 110-115.5° E, 21.5-24° N) during  
 738 TC-HDs and AHDs, as well as their differences (TC-HDs minus AHDs).

	YRD		PRD	
	Observed	Simulated	Observed	Simulated
TC-HDs	134.71	177.17	97.93	123.91
AHDs	147.92	199.11	91.13	116.91
Differences	-13.21	-21.94	6.8	7

739  
 740 **Table 2.** Contributions of difference processes for the anomalous ozone mass ( $\text{Gg O}_3/\text{day}$ ) from  
 741 surface to 500 hPa for TC-HDs and AHDs relative to the summer climatology and their difference  
 742 (TC-HDs minus AHDs) for YRD.

	Net chemical production	Vertical advection	Horizontal advection	Mixing
TC-HDs	1.70	0.03	-2.18	0.004
AHDs	0.46	-0.04	-1.03	-0.12
Differences	1.24	0.07	-1.15	0.12

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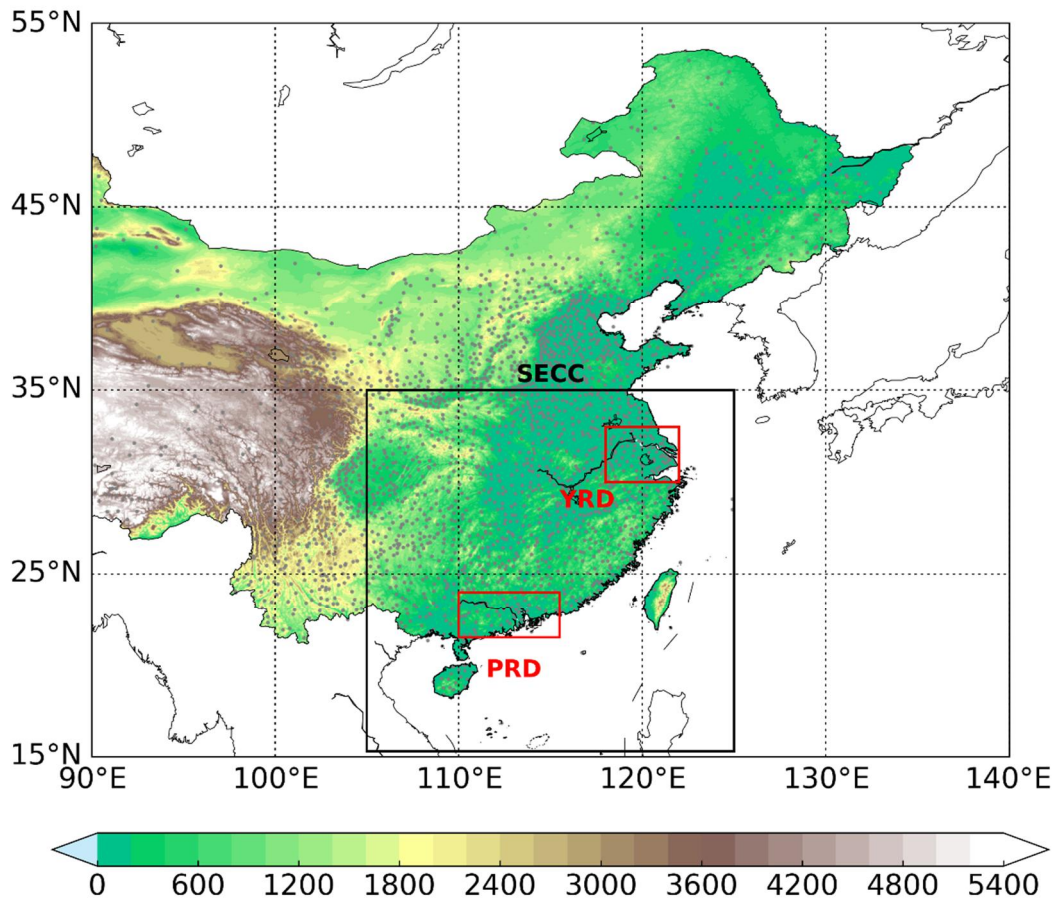
745

746 **Table 3.** Contributions of difference processes for the anomalous ozone mass (Gg O<sub>3</sub>/day) from  
 747 surface to 500 hPa for TC-HDs and AHDs relative to the summer climatology and their difference  
 748 (TC-HDs minus AHDs) for PRD.

	Net chemical production	Vertical advection	Horizontal advection	Mixing
TC-HDs	0.21	-0.13	0.97	-0.002
AHDs	0.06	0.09	-0.84	-0.05
Differences	0.15	-0.22	1.81	0.05

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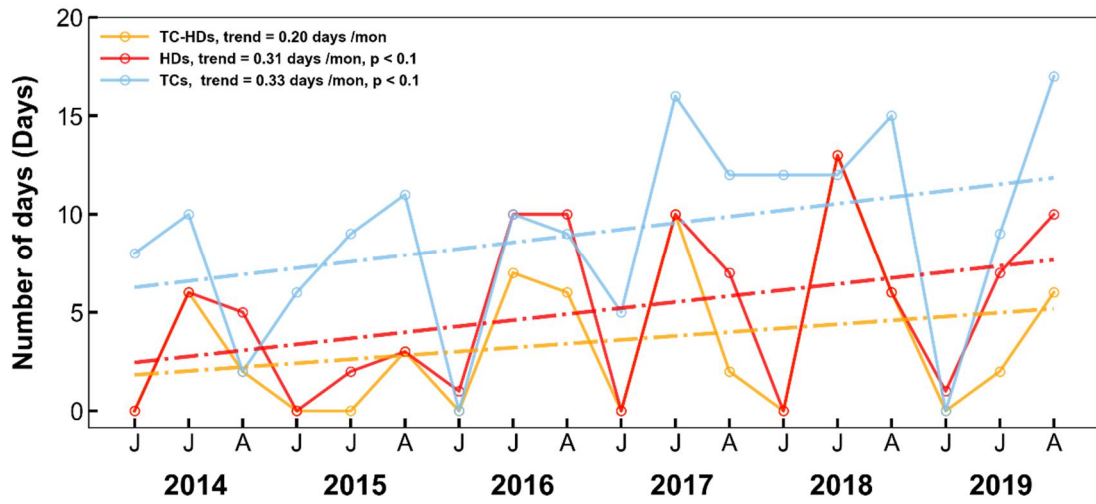


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752 **Figure 1.** Topographic map of China with SECC (black box) and megacity clusters YRD and PRD  
 753 outlined (red boxes). Gray points represent the temperature observation stations.

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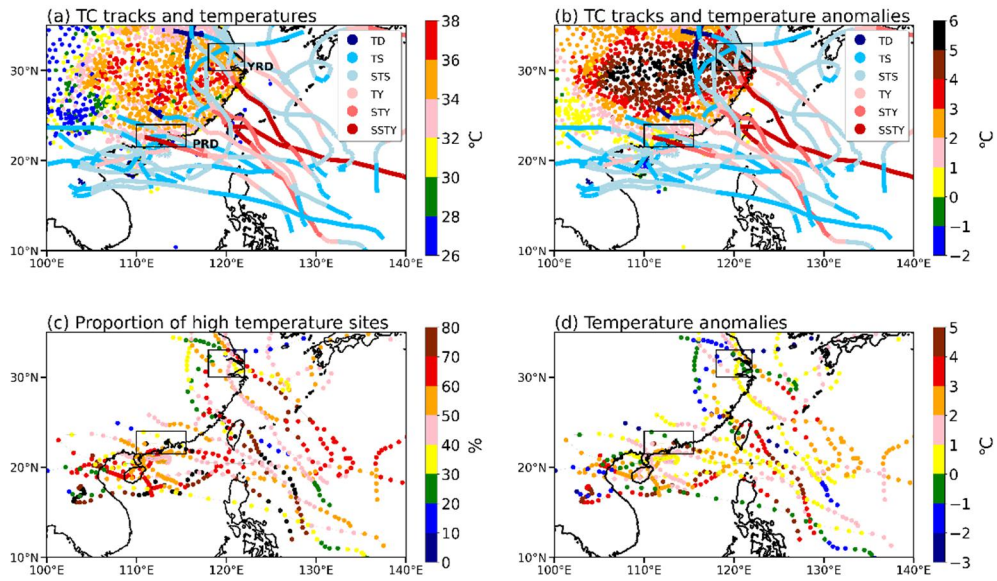
756

757 **Figure 2.** Monthly variations of the number of days of TC-HDs, total HDs, and total TCs during the

758 summer season (June, July and August) from 2014-2019.

759





760

761 **Figure 3.** (a) The distribution of average Tmax (dots) and TC tracks during TC-HDs. (b) The Area

762 average of Tmax anomalies during TC-HDs period relative to the summer climatology (June to

763 August of 2014-2019), along with the TC tracks categorized by different intensities. (c) The

764 proportion of high-temperature sites ( $T_{max} \geq 35^{\circ}\text{C}$ ) over land region of SECC along with the

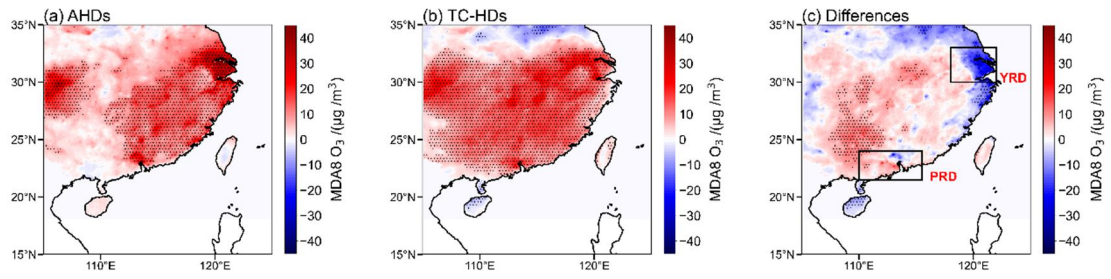
765 movements of the TCs. The proportion of high-temperature sites refers to the percentage of high-

766 temperature sites within all stations in the SECC region. (d) The average of Tmax anomalies for all

767 observational sites within SECC relative to the summer climatology, along with the movements of

768 TCs. YRD and PRD regions are outlined in black boxes in each panel.

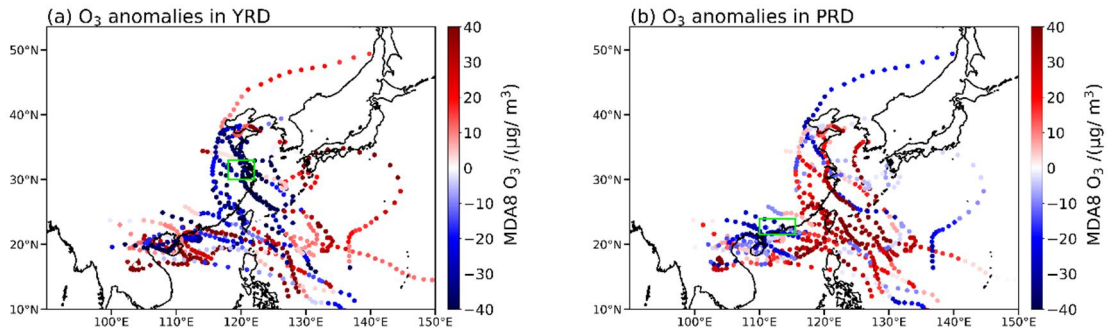
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770

771 **Figure 4.** The spatial distribution of surface MDA8 ozone concentration anomalies during the  
 772 periods of (a) AHDs, (b) TC-HDs relative to the summertime climatology, as well as (c) their  
 773 differences (TC-HDs minus AHDs). Stippling regions indicate ozone anomalies that are  
 774 significantly difference from zero at the 95% confidence level. YRD and PRD regions are outlined  
 775 in black boxes in panel c.

776



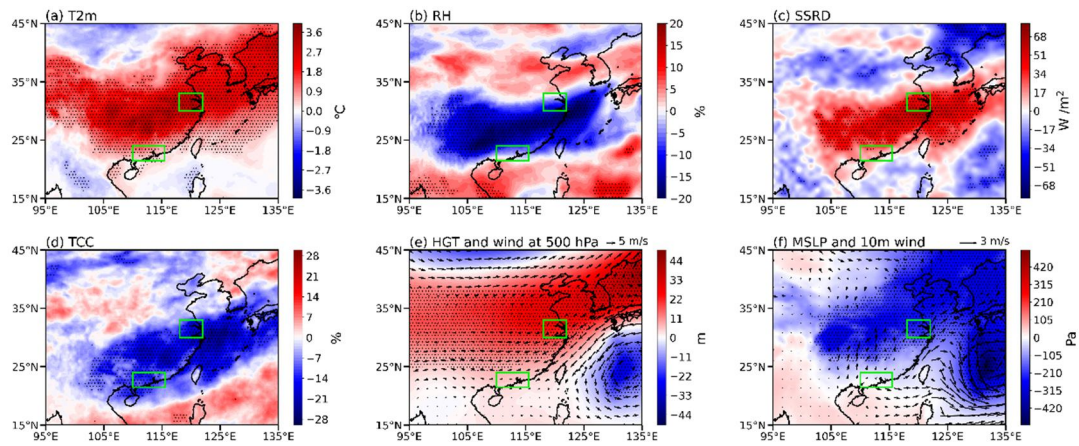
777

778 **Figure 5.** The average anomalies of surface MDA8 ozone concentrations over land regions of SECC

779 along with the movements of TCs associated with TC-HDs for (a) YRD and (b) PRD regions. YRD

780 (a) and PRD (b) regions are outlined in green boxes in panel (a) or (b).

781



782

783 **Figure 6.** The spatial distribution for the composites anomalies of (a) air temperature at 2m (T2m),

784 (b) relative humidity (RH), (c) surface solar radiation downwards (SSRD), (d) total cloud cover

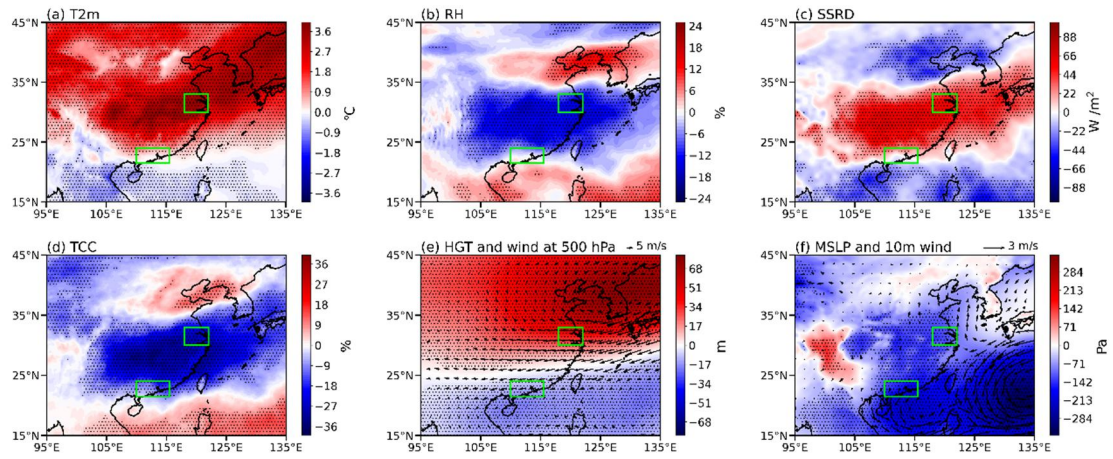
785 (TCC), (e) geopotential height (HGT) and winds at 500hPa, and (f) mean surface level pressure

786 (MSLP) and 10-meter winds during AHDs relative to the summer climatology. Stippling indicates

787 statistically significant anomalies above 95% confidence level. YRD and PRD regions are outlined

788 in green boxes in each panel.

789

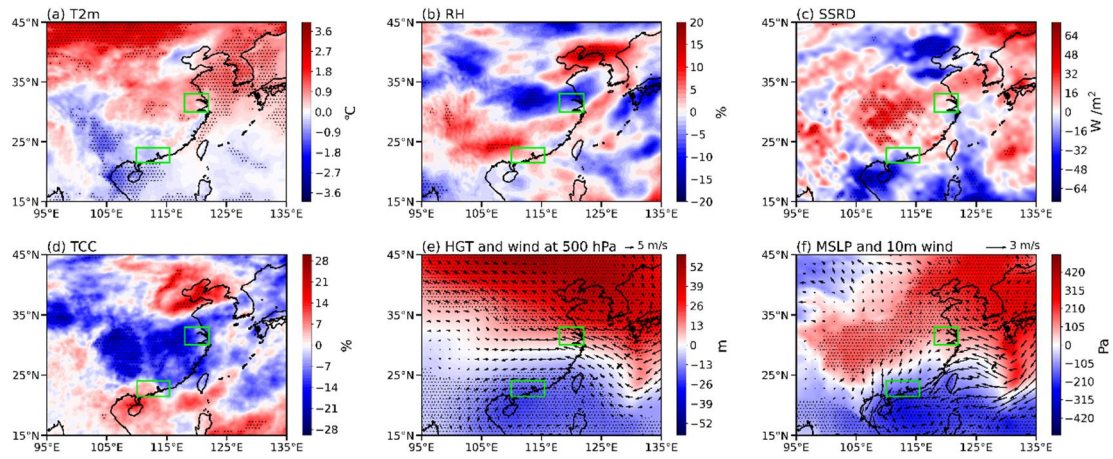


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**Figure 7.** Same as in Figure 6, but for the anomalous meteorological conditions during TC-HDs.

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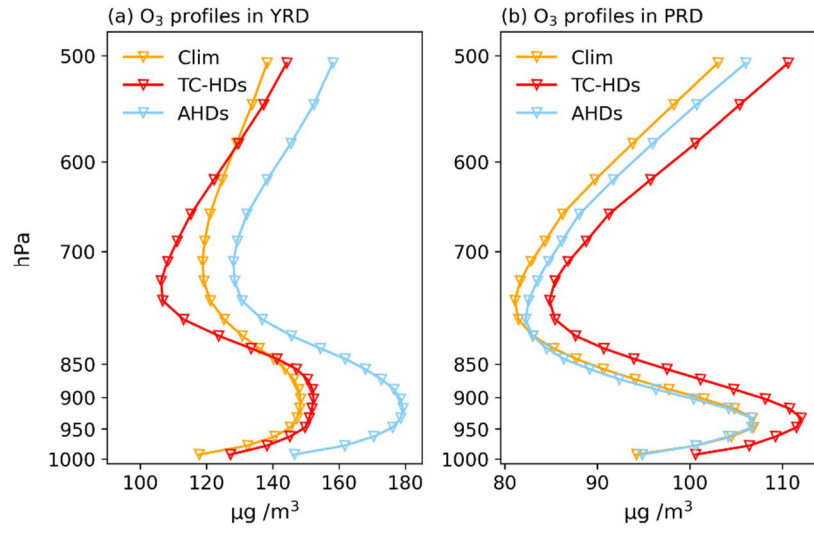


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794 **Figure 8.** Same as in Figure 6, but for the differences between TC-HDs and AHDs (TC-HDs minus

795 AHDs).

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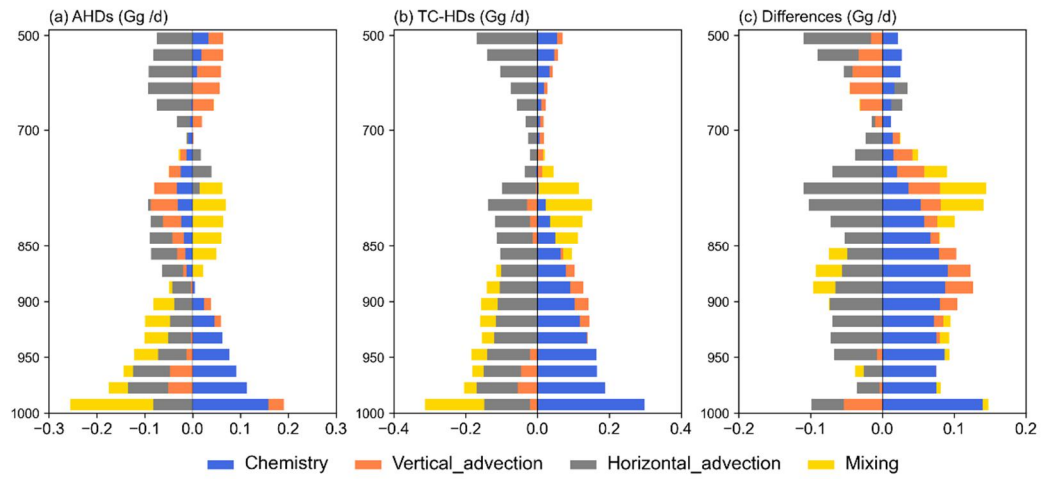
797

798 **Figure 9.** Vertical profiles of simulated daily ozone concentrations ( $\mu\text{g}/\text{m}^3$ ) averaged over land

799 regions of SECC for TC-HDs, AHDs and for the summertime climatology (Clim) during 2014-2019

800 for (a) YRD and (b) PRD.

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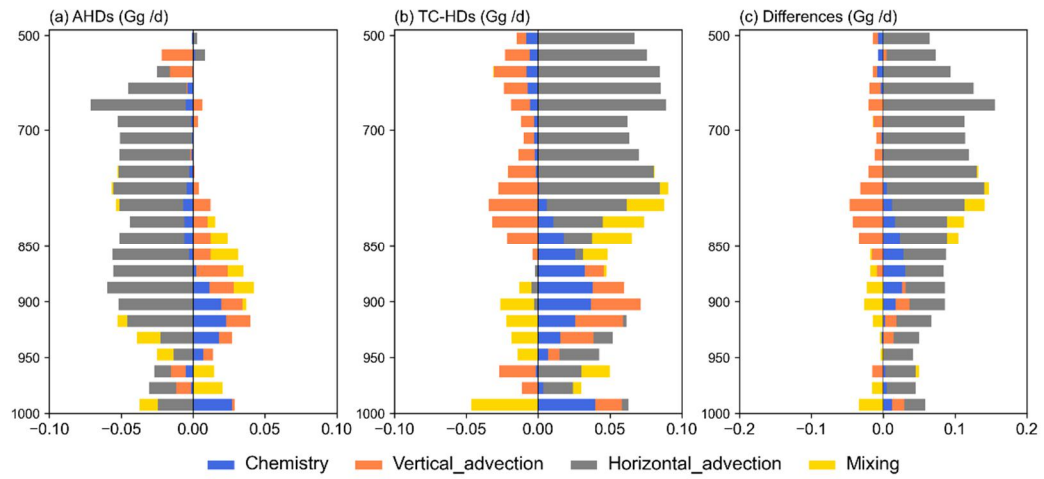
803 **Figure 10.** Vertical profiles of net changes in ozone mass (Gg O<sub>3</sub>/day) anomalies averaged for YRD

804 for each process during (a) AHDs, (b) TC-HDs relative to summertime climatology, and their

805 differences (TC-HDs minus AHDs).

806





807

808 **Figure 11.** Same as in Figure 10, but for the vertical profiles of net changes in ozone mass (Gg

809 O<sub>3</sub>/day) anomalies averaged for PRD for each process.