

1 **Influence of terrestrial and marine air-mass on the constituents and**
2 **intermixing of bioaerosols over coastal atmosphere**

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12 **Abstract:**

13 Coastal environments provide an ideal setting for investigating the intermixing
14 processes between terrestrial and marine aerosols. In this study, fine particulate matter
15 (PM_{2.5}) samples categorized into terrestrial, marine and mixed air masses were
16 collected from a coastal location in Northern China. The chemical and
17 biological constituents, including water-soluble ions (WSIIs), metallic elements, and
18 bacterial and fungal aerosols, were investigated from January to March 2018,
19 encompassing both the winter heating and spring dust seasons. Terrestrial air masses
20 constituted 59.94% of the total air masses throughout the sampling period, with a
21 significant increase during severe haze pollution (up to 90%). These air masses
22 exhibited higher concentration of PM_{2.5} (240 µg/m³) and carried more water-soluble
23 ions and metal elements. The terrestrial air mass also contained a larger number
24 of animal parasites or symbionts, as well as human pathogens from
25 anthropogenic emission, such as *Staphylococcus*, *Deinococcus*, *Sphingomonas*,

26 *Lactobacillus*, *Cladosporium* and *Malassezia*. Conversely, a significant quantity of
27 saprophytic bacteria such as hydrocarbon-degrading and gut bacteria from the genera
28 *Comamonas*, *Streptococcus*, *Novosphingobium*, *Aerococcus*, and the saprophytic
29 fungus *Aspergillus* were the most abundant species in the marine air mass samples. The
30 mixed air mass elucidates the intermixing process of terrestrial and marine sources, a
31 result of microorganisms originating from both anthropogenic and terrestrial emissions,
32 which including pathogenic microorganisms from hospitals and sewage treatment
33 plants, and a multitude of soil bacteria. A stronger correlation was noted between
34 microorganisms and continental elements in both terrestrial and mixed air mass samples,
35 specifically K^+ , Mg^{2+} , and Ca^{2+} derived from soil dust. Marine air masses exhibited a
36 significant correlation with sea salt ions, specifically Na^+ . In the mixed air mass sample,
37 a fusion of marine and terrestrial microorganisms is characterized by alterations in the
38 ratio of pathogenic to saprophytic microorganisms when compared to samples derived
39 from either terrestrial or marine sources. This study on the constituents and
40 amalgamation of bioaerosols over the coastal atmosphere encompassing distinct air
41 masses is crucial for understanding the transport, intermixing processes, and health
42 implications of terrestrial and marine air masses.

43 **Keywords:** $PM_{2.5}$, water-soluble ions, terrestrial air mass, marine air mass, bioaerosols

44 **1. Introduction**

45 Bioaerosols, encompassing bacteria, fungi, viruses, pollen, and cellular debris, etc.,
46 are vital aerosol particles in the atmosphere. Notably, bacterial aerosols which can be
47 either free-floating or attached to airborne particles, typically measure between 0.3 and
48 10 μm (Shen and Yao, 2023; Zhao et al., 2022). Such characteristics allow for extended
49 residence times and a heightened likelihood of long-distance transport from local to
50 distant regions within the atmospheric circulation (Galbán et al., 2021; Smets et al.,
51 2016). Fungal aerosols are prevalent in Earth's near-surface atmosphere, with their
52 origins being diversely found in water, soil, and plants (Fröhlich-Nowoisky et al., 2009).
53 Furthermore, the presence of pathogenic bacteria and fungi in the atmosphere may lead
54 to significant penetration efficiency within the human respiratory system (Fakunle et
55 al., 2021; Jiang et al., 2022).

56 Relevant studies have demonstrated that the bacterial and fungal concentration and
57 community structure are significantly influenced by emission sources, atmospheric
58 circulation, meteorological conditions, geographical and topographical factors (Liu et
59 al., 2021; Andrés et al., 2021; Li et al., 2019). The geographical and topographical
60 factors, such as terrestrial and marine environments, exhibit significant
61 differences in bioaerosol sources and pollution characteristics. Studies focusing on
62 continental environments, particularly in large inland cities like Beijing (Zhang et al.,
63 2019; Zhang et al., 2022), Xi'an (Wang et al., 2020), Shanghai (Yao et al., 2019), Jinan
64 (Wei et al., 2020) , and Nanjing (Hu et al., 2020) during severe air pollution, have
65 revealed that continental pollutant emissions significantly contribute to air pollution.
66 Furthermore, transboundary transportation of terrestrial air masses on hazy days plays
67 a crucial role in the progression of regional heavy pollution (Xie et al., 2020). The ocean
68 serves as both a potential source and sink for airborne microorganisms (Archer et al.,

69 2020; Mayol et al., 2017), exhibiting the intricate interplay between continental
70 communities and their transmission across oceanic regions (Cho and Hwang, 2011;
71 Deleon-Rodriguez et al., 2013). Bioaerosols from the oceans may be influenced by
72 long-distance transport from continental sources, such as plants and human pathogens
73 (Elbert et al., 2007; Sharoni et al., 2015). Studies have shown that the concentration and
74 diversity of bacterial and fungal aerosols from marine environment are typically lower
75 than those derived from continental sources (Cao et al., 2024; Xue et al., 2022; Shi et
76 al., 2022). Xu et al. (2019) undertook a thorough investigation of bacterial abundance
77 in Mt. Tai, China. Their findings indicate that variations in air mass from diverse sources
78 could potentially influence the chemical composition of PM_{2.5}, which in turn prompts
79 shifts in bacterial groups. Several studies have examined the impacts of terrestrial and
80 marine air masses on chemical constituents and microbial aerosols (Shi et al., 2022;
81 Aswini and Hegde, 2021; Lang-Yona et al., 2022). Generally, the chemical aerosols can
82 be affected by various sources of air masses, which may include local aerosols or
83 remotely transported aerosols. However, it remains unclear whether different air
84 trajectories contribute to the formation of bacterial communities within these particles.
85 There is a dearth of research focused on the influence of different air masses on the
86 chemical and biological composition in coastal cities.

87 Marine microbial aerosols can be released from ocean micro-surface or transported
88 from land (Prospero et al., 2005) and settle thousands of kilometers away from their
89 source of release (Mayol et al., 2014), exerting significant impacts on ecological and
90 climate systems. Gong et al (2020a) conducted an examination of the microbial
91 composition along Qingdao's coastlines, revealing a higher proportion of bacteria to
92 total microorganisms from continental sources compared to marine sources. Some
93 marine bacteria persist in aerosols after land transportation, Xu et al (2020b) studied the

94 diversity of bacterial populations in PM_{2.5} across urban and rural areas of Shanghai,
95 finding that airborne microbial communities over coastal cities are more influenced by
96 long-distance transport than those inland. Air samples taken at high altitude zones in
97 coastal Europe and Japan demonstrated the continental transportation of marine
98 microorganisms (Maki et al., 2014). The introduction of marine bacteria into
99 tropospheric free space may modify the airborne microbial composition in continental
100 regions (Cho and Hwang, 2011; Amato et al., 2007; Maki et al., 2014; Polymenakou et
101 al., 2008; Cáliz et al., 2018). Consequently, it is imperative to ascertain the impact of
102 terrestrial and marine air masses on the composition of atmospheric microorganisms,
103 with a particular focus on the influence of the mixing process between terrestrial and
104 marine air masses on bioaerosols. The coastal aerosols provide the ideal conditions for
105 understanding the mixing processes taking place between natural and anthropogenic air
106 masses from terrestrial and marine.

107 Weihai, a coastal city situated at the confluence of eastern Asia continent and the
108 Northwest Pacific Ocean, is prone to the impact of marine and terrestrial air masses.
109 Typically, Weihai maintains low pollutant emission with an annual average PM_{2.5}
110 concentration below 35 µg /m³. However, in winter and spring, regional air pollution
111 intensifies, leading to severe air quality issues. This is due to increased inter-regional
112 transportation, resulting in daily average PM_{2.5} concentration greater than 150 µg/m³
113 (Wei et al., 2020). In this study, we conducted an integrated atmospheric observation
114 experiment to examine the potential impact of terrestrial and marine air-mass on the
115 constituents and amalgamation of bioaerosols over the coastal atmosphere. This
116 approach allows for a comprehensive exploration of the effects of sea-land air mass
117 exchange on the spatial and temporal distribution of aerosols, as well as potential
118 intermixing processes in coastal regions.

119 2. Materials and methods

120 2.1 Sample collection

121 The sampling site was situated at the national air sampling station (37.53°N,
122 122.06°E), 41 m above sea level, and approximately 1-2 km from the coast (Fig. S1).
123 The sampling platform was about 15 m above ground level, and devoid of any
124 significant obstructions. PM_{2.5} samples were gathered between January and March,
125 2018, during the winter heating and spring dust seasons in northern China. Two parallel
126 PM_{2.5} samplers (TH-150C-III, Wuhan Tianhong Instrument Co., Ltd., China) with a
127 particle size of $2.5 \pm 0.2 \mu\text{m}$ were utilized, equipped with Quartz membrane for the
128 collection of PM_{2.5}, inorganic ions, metal elements, and microorganisms at a flow rate
129 of 100 L min^{-1} . The Quartz membranes were cauterized in a muffle furnace at 450°C
130 for 6 hours to remove carbonaceous and contaminant materials. Intermittent sampling
131 was used and samples were collected twice a day (7:00 to 18:30 and 19:00 to 06:30 the
132 following day). During the sampling process, the pre-weighed filter membrane was
133 carefully positioned onto the filter mesh within a clean sampling folder using tweezers.
134 The membrane's rough side should be oriented towards the direction of air intake, and
135 it should be firmly pressed until there is no detectable air leakage. Prior to membrane
136 replacement, the membrane tray is wiped with alcohol cotton and dried. After sampling,
137 the membrane is removed using tweezers. Blank samples were procured by putting the
138 membrane into the sampler head without initiating the sampler. After sampling, the
139 blank samples were extracted simultaneously with the collected sample membranes. All
140 samples were then placed in a sterile filter cassette, encased in clean aluminum foil, and
141 stored at -80°C until further analysis. During the whole sampling period from January
142 1 to March 31, a total of 102 PM_{2.5} samples over 51 days were obtained, out of which

143 24 samples were chosen for subsequent analysis based on air mass categories.
144 Meteorological parameters, such as air temperature, relative humidity, wind direction,
145 and wind speed, were monitored in situ utilizing a PC-4 automatic weather station (PC-
146 4, JZYG, China). The hourly concentration of PM_{2.5}, PM₁₀, CO, SO₂, NO₂, and O₃ were
147 systematically retrieved from the National Ambient Air Quality Monitoring System
148 (<http://www.cnemc.cn/>).

149 The mass concentration of PM_{2.5}, water-soluble ions, and metal elements were
150 quantified after sampling. The membranes were meticulously weighed utilizing a
151 Mettler XP-6 balance with an accuracy of 10⁻⁶ g. Prior to the weighing, the membranes
152 were maintained in a controlled environment with consistent temperature and humidity
153 for a duration of 24 hours. Ion chromatography (ICS-2100, Chameleon 6. 8, AS-DV
154 autosampler Thermo Fisher) was employed to determine the concentration of water-
155 soluble ions such as Na⁺, K⁺, Ca²⁺, Mg²⁺, Cl⁻, NO₃⁻, SO₄²⁻, and NH₄⁺. These ions were
156 extracted by sonication with ionized water for 45 min and then separated by anion
157 (IonPacAS23) or cation column exchange (IonPacCS12A). Then, these ions were
158 detected using a conductivity detector with an anion separation column of IonPacAS23,
159 flow rate: 1.0 mL/min, an anion suppressor of AERS500, and conductivity detector. The
160 injection volume was 25 μL, and the cation separation column was IonPacCS12A, flow
161 rate:1.0 mL/min, cation suppressor was CERS500 (Zhang et al., 2022). Metallic
162 elements including Al, Fe, Ti, Mn, Co, Ni, Cu, Zn, Ga, Sr, Cd, Sn, Sb, Pb, V, Cr, and
163 As, were extracted using microwave digestion extraction (ETHOS ONE, Milestone),
164 with the concentrations determined by ICP-MS or ICP-OES (Thermo Fisher).

165 **2.2 Air mass clustering and classification**

166 The potential sources and transportation of air mass were examined using the

167 Meteoinfo backward trajectory model ([MeteoInfo 3.7.4 – Java,](http://www.meteothink.org/downloads/index.html)
168 <http://www.meteothink.org/downloads/index.html>) developed by the Chinese Academy
169 of Meteorological Sciences. Backward trajectories were simulated at one-hour intervals
170 and estimated over a 24-hour period. The meteorological data were sourced from
171 GDAS1 (<ftp://arlftp.arlhq.noaa.gov/pub/archives/gdas1/>). Backward trajectories of air
172 masses at an altitude of 500 meters were categorized and clustered, with daily plots
173 illustrating these trajectories. In this study, air mass categories were defined that if more
174 than 90% of the masses originated from the ocean, it was a typical sample affected by
175 marine air-mass; a typical terrestrial air-mass sample if more than 90% originated from
176 the continent; and a mixed air-mass sample if the proportions of terrestrial and marine
177 air masses were similar or accounted for more than 40% of the total air masses in one
178 day.

179 **2.3 DNA extraction and qPCR amplification**

180 Microbial genomic DNA were procured from filters utilizing the Fast-DNATM SPIN
181 kit for soil (MoBio Laboratories, Carlsbad, CA, USA). The concentration and quality
182 of the extracted DNA was measured via a Nanodrop spectrometer (Nanodrop 2000,
183 Thermo Scientific USA). The DNA concentration in the collected samples ranged from
184 10 to 20 ng/μL, satisfying the prerequisites for subsequent PCR amplification. The
185 quantitative polymerase chain reaction (qPCR) was employed to identify bacterial 16S
186 rRNA and fungal ITS gene copy numbers per cubic meter of air. The bacterial 16S V3-
187 V4 variable region was selected for PCR amplification using primer 338F (5'-
188 ATCTACGGGGGGCAGCAG-3') and 806R (5'GGACTACHVGGGTWTCTAAT-3')
189 (Masoud et al., 2011). The fungal ITS region were amplified using the primers ITS1
190 (5'-CTTGGTCATTTAGAGGAAGTAA-3') and ITS2 (5'-

191 GCTGCGTTCTTCATCGATGC-3') (Liu et al., 2021).

192 The PCR amplification conditions comprised an initial denaturation at 95°C
193 for 5 minutes, succeeded by 30 seconds at 95°C, 30 seconds at 50°C, and then 35 cycles
194 at 72°C for 40 seconds, followed by a final extension at 72°C for seven minutes to
195 ensure comprehensive amplification. Fluorescent signals were gathered during the
196 extension phase. For each sample, qPCR was conducted in triplicate, with ultrapure
197 water serving as a negative control to identify potential PCR contamination. Standard
198 curves were constructed using *E. coli* harboring the 16S rRNA gene and *Streptomyces*
199 plasmids containing the ITS gene. Additionally, gradient dilutions of these plasmids
200 were performed, ranging from 10²-10⁷ copies/μL. The copy number of 16S rRNA gene
201 in bacterial genome ranges from 1 to 21 with an average value of 5.5 according to the
202 rrnDB database (<https://rrndb.umms.med.umich.edu/>) (Stoddard et al., 2015). The
203 number of rRNA copies per genome ranges from 50 to 100 in filamentous fungi
204 (Rooney and Ward, 2005) and *Ascomycetes* typically have relatively smaller genome
205 sizes (Kullman et al., 2005). Consequently, the copy number of the fungal ITS gene in
206 their genome was set at 50 in this study. The FTC-3000 real-time quantitative
207 PCR system (Funglyn Biotech Inc., Scarborough, ON, Canada) was employed for
208 standard curve construction and data processing.

209 **2.4 16S rRNA and ITS gene sequencing and data processing**

210 The V3-V4 region of the bacterial 16S rRNA and the fungal ITS1 gene were targeted
211 for PCR amplification utilizing barcode-specific primers 338F-806R and ITS1F-ITS2,
212 respectively. To ensure optimal amplification efficiency and precision, a high-fidelity
213 enzyme (Phusion® High Fidelity PCR Master Mix from NewEngland Biolabs) in
214 conjunction with GC buffer was employed during PCR amplification.

215 This procedure entailed a pre-denaturation step at 98°C for one minute, followed by 30
216 cycles of 98°C for ten seconds; 58°C for bacteria (56°C for fungi) for 30
217 seconds; and 72°C for an additional 30 seconds, with a final extension at 72°C for five
218 minutes. Following amplification, samples were purified using the Agencourt Ampure
219 XP kit (Beckman Coulter, Brea, CA, USA). Subsequently, the purified samples were
220 combined to achieve equimolar concentrations and analyzed on the Illumina MiSeq
221 PE300 platform (Illumina, Inc. in San Diego, CA).

222 Following sequencing, the barcode sequences were extracted and subsequently
223 stored in fastq format utilizing the QIIME toolkit (Caporaso et al., 2010). Sequences
224 shorter than 200 base pairs, with mass of less than 25, and containing ambiguous bases
225 underwent quality control screening using Trimmomatic (Bolger et al., 2014) and
226 Mothur (Schloss et al., 2009). The assembled sequences were then de-duplicated and
227 trimmed to equal length. De-chimerized sequences were eliminated (Edgar, 2013), and
228 sequences were clustered into operational taxonomic units (OTUs) using UPARSE at a
229 97% identity threshold. Subsequently, individual OTUs were removed. Taxonomic
230 assignments were determined using a Basic Local Comparison Search Tool (BLAST)
231 search against the bacterial SILVA taxonomy release 138.2 dataset and fungal UNITE
232 ITS reference v8.0 dataset. The original raw sequences were deposited in the Sequence
233 Read Archive under the accession number PRJNA1101427 and PRJNA1101176.

234 **2.5 Statistical analysis**

235 Bacterial community functional analysis was conducted using FAPROTAX, a
236 manually constructed database that maps prokaryotic taxa to metabolic or other
237 ecological functions, such as sulfur, nitrogen, hydrogen, and carbon cycling (Chen et
238 al., 2022). FUNGuild (Fungi Functional Guild) was used to predict the fungal

239 ecological function. This tool could classify and analyze fungal communities by the
240 microecological guild based on current published literature or data from authoritative
241 websites to classify fungi functionally (Nguyen et al., 2016). Three primary groups are
242 obtained based on the nutritional mode: Pathotroph, Symbiotroph, and Saprotroph.
243 Samples affected by different air masses were examined for intergroup species
244 variability, based on community abundance data. This was achieved using rigorous
245 statistical methods to identify species demonstrating differences in abundance within
246 the microbial communities of different groups, and hypothesis testing to evaluate the
247 significance of these observed differences. Statistical analysis, including Analysis of
248 variance (ANOVA) and Kruskal Wallis tests, were employed to discern bacteria and
249 fungi with varying abundances between samples and groups. A *p* value less than 0.05 or
250 0.01 was considered significant. ANOVA was used to analyze variation in a response
251 variable measured under conditions defined by discrete factors (Martin G. Larson,
252 2008). The Kruskal-Wallis test determines whether there is a statistically significant
253 difference between the medians of three or more independent groups (Kassambara,
254 2019). The Mantel analysis was utilized to reveal the correlation between microbial
255 community composition and various environmental factors. The Pearson's *r* coefficient
256 at $p < 0.05$ and $p < 0.01$ indicates the significant correlation.

257 **3. Results**

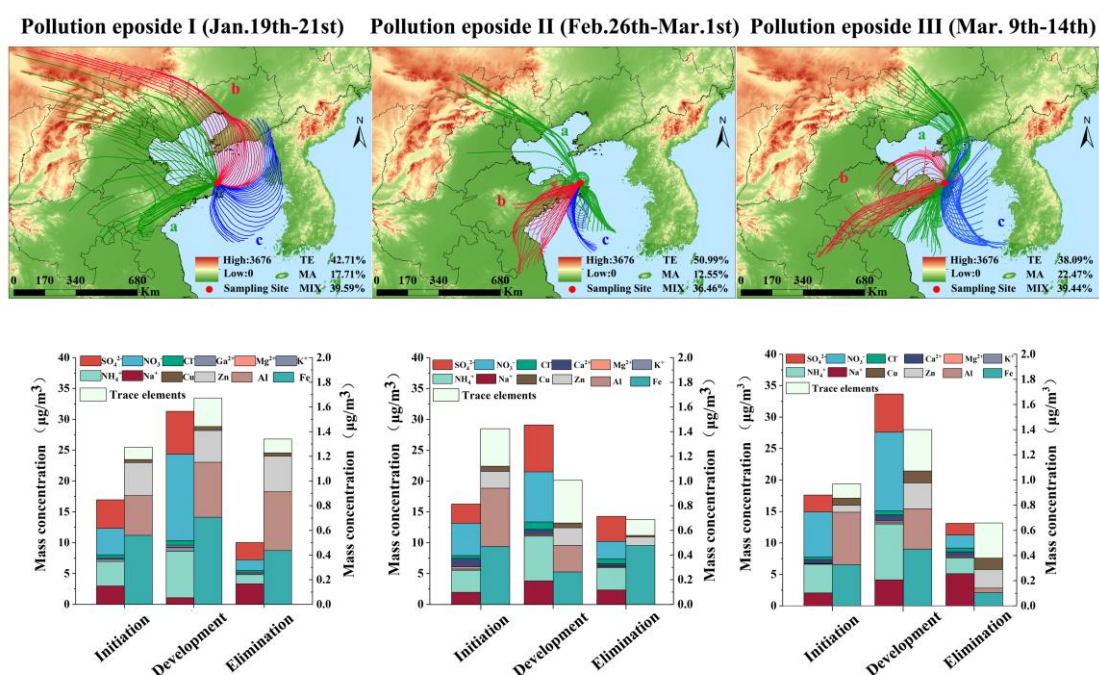
258 **3.1 Air masses categorization and typical pollution processes**

259 The Meteolnfo backward trajectory model was employed to simulate the trajectories
260 of air masses at an altitude of 500 m in the coastal city over 24-hour period,
261 identifying and classifying potential sources of air mass transport (Fig. S1). The
262 terrestrial air mass accounted for 59.94% of the total air masses at the sampling

263 site throughout the period, exhibiting an average $PM_{2.5}$ concentration of 36.15 ± 26.52
264 $\mu\text{g}/\text{m}^3$. Severe air pollution episodes occurred on January 20, with $PM_{2.5}$ was the
265 primary pollutant and the concentration reaching $240 \mu\text{g}/\text{m}^3$. During regional haze
266 pollution, the terrestrial air masses primarily influenced the coastal city were typically
267 originating from the North China Plain and the surrounding areas. The primary
268 contributors to emissions in the terrestrial air mass of this region are the dense
269 population, industrial and agricultural activities. The impact of the spring dust is mainly
270 reflected in the primary pollutant PM_{10} from March 30 to March 31. A significant
271 increase in PM_{10} concentration was observed on March 31, with an hourly maximum
272 value of $167 \mu\text{g}/\text{m}^3$. A relatively low $PM_{2.5}/PM_{10}$ ratio of 0.25 indicated pronounced
273 dust pollution. The typical marine air masses mainly from the eastern and southern areas
274 from the Yellow Sea and the western Pacific edge. Marine air masses from northern
275 areas partly come from the Bohai Sea, crossing inland areas and possibly making
276 contact with the land prior to reaching the study area (Fig. S1). A total of 14% of the
277 sampling days were influenced by marine air masses, resulting in an average $PM_{2.5}$
278 concentration of $23.99 \pm 11.00 \mu\text{g}/\text{m}^3$. Mixed air masses, characterized by simultaneous
279 influence from the northwestern winds of Inner Mongolia and the offshore air masses
280 of the Bohai Sea or Yellow Sea, accounted for 27% of all affected sampling days. These
281 mixed air masses yielded an average $PM_{2.5}$ concentration of $45.11 \pm 12.69 \mu\text{g}/\text{m}^3$. In
282 comparison, under the influence of mixed air masses, pollutant concentrations were
283 notably high. During the spring dust season, there is a notable increase in the proportion
284 of mixed air mass. This elevated concentration of particulate matter correlates with the
285 transmission of sand dust (Xie et al., 2020).

286 Three heavy pollution events were investigated to explore the alterations in air mass
287 during these episodes (Fig. 1, Fig. S2). Generally, the initiation and development stages

288 of pollution events were predominantly characterized by terrestrial and mixed air
 289 masses. Cold northwestern airmass from the continent, and marine air masses from east
 290 or south, were the primary contributors during pollution mitigation. For pollution
 291 episode I occurred from January 19 to January 22, the western terrestrial air mass was
 292 predominantly responsible for initiating the pollution, evidenced by a PM_{2.5}
 293 concentration of 51.35 µg/m³. As the pollution progressed, it transitioned into a mixed
 294 air mass with a PM_{2.5} concentration peaking at 240 µg /m³. This change was
 295 accompanied by significant increases in WSIs and elemental concentrations. Pollution
 296 elimination was initiated with high winds, snowfall influenced by the cold northwestern
 297 airmass from the continent and marine air masses from the northeast sea. The average
 298 PM_{2.5} concentration was reduced to 7.92 µg /m³, accompanied by a decrease in both
 299 WSIs and elemental concentrations.



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 304

Fig. 1 Transformation of terrestrial and marine air masses, and chemical composition in PM_{2.5} of three severe air pollution episodes. a, pollution initiation; b, pollution development; c, pollution elimination. TE, terrestrial air mass; MA, marine air mass; MIX, mixed air mass.

305 3.2 Water-soluble ions and metal elements over coastal atmosphere

306 During the sampling period, the concentration of water-soluble ions in PM_{2.5} was
307 22.92±12.19 µg/m³, ranging from 4.58 to 78.14 µg/m³. In marine air-mass samples, a
308 lower concentration of water-soluble ions was observed, with the concentration of
309 13.01±7.43 µg/m³, 27.94±13.61 µg/m³ and 30.38±11.38 µg/m³ in marine, terrestrial and
310 mixed air masses, respectively. Notably, NO₃⁻ had the highest proportion (26.94%,
311 6.4%~52.6%), followed by SO₄²⁻ (21.94%, 9.4%~33.4%) and NH₄⁺ (20.26%,
312 5.8%~35.6%). The SO₄²⁻, NO₃⁻ and NH₄⁺, constituted the majority of total ions in other
313 studies conducted in Beijing, Shanghai, and Guangzhou (Pathak et al., 2008; Hu et al.,
314 2014; Zhou et al., 2012). These inorganic secondary ions were significantly influenced
315 by both terrestrial and mixed air masses, with the latter exhibiting a more pronounced
316 effect (NO₃⁻, 10.65±3.26 µg/m³; NH₄⁺, 7.39±3.30 µg/m³; SO₄²⁻ 6.76±1.77 µg/m³) (Fig.
317 S3).

318 A high concentration of Na⁺ was observed, with a range of 3.15 ± 1.69 µg/m³, and
319 accounted for 14.47% of the total water-soluble ions. The concentration of Na⁺ and
320 Mg²⁺ exhibited analogous trends across terrestrial, marine, and mixed air mass samples.
321 The ratio of Mg²⁺/Na⁺, typical components of sea salt, was determined to be 0.11. This
322 value closely corresponded to the 0.12 found in seawater, indicating a potential origin
323 from marine environments (Sun et al., 2022). The concentration of K⁺ was 0.24±0.20
324 µg/m³ and 0.26±0.10 µg/m³ in the terrestrial and mixed air-mass samples, and was twice
325 as high as those in the marine air-mass samples (0.11±0.05 µg/m³), which suggested an
326 important contribution from anthropogenic emissions. The concentration of Cl⁻, and
327 Ca²⁺ was high in terrestrial and mixed air mass samples. Generally, these two ions
328 mostly come from anthropogenic or terrestrial environments, such as Cl⁻ mainly comes
329 from sea salt, coal and biomass combustion, and Ca²⁺ is a crustal element associated

330 with soil dust and sandstorms (Liang et al., 2021). Overall, from the composition and
331 concentration of water-soluble ions in PM_{2.5}, the coastal city was more affected by sea
332 salt, coal combustion, and dust events in early spring.

333 The elements concentration in different air-mass samples are depicted in Fig. S3. The
334 top ten elements were found to be Fe, Al, Zn, Cu, Sn, Pb, Mn, Ti, Ni, and V. These metal
335 elements were categorized into macro and trace elements. The macro metals,
336 specifically Iron (Fe), Aluminum (Al), and Zinc (Zn), constituted a significant
337 proportion of the total heavy metal elements, accounting for 34%, 25%, and 23%
338 respectively. In general, the concentration of both macro and trace metal elements in
339 marine air masses was found to be lower than that in terrestrial and mixed air masses.
340 The V/Ni ratio in marine air-mass samples was found to be 0.78, a value significantly
341 higher than that of both terrestrial and mixed air masses. The V/Ni ratio serves as an
342 indicator to assess the influence of ship emissions. A ratio exceeding 0.7 typically
343 indicates a significant impact from ship emission sources, and is commonly used as an
344 indicator in coastal cities (Zhang et al., 2014).

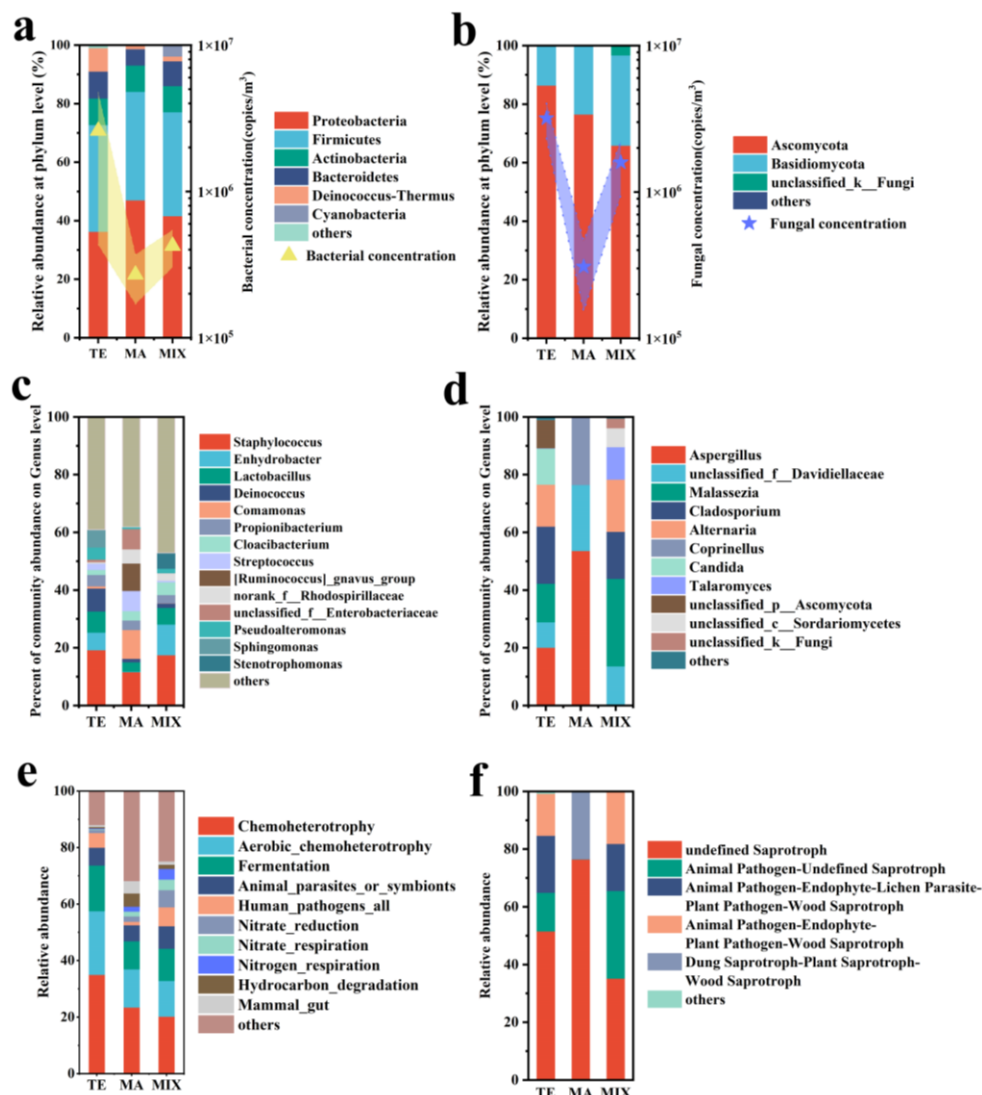
345 **3.3 Microbial community over coastal atmosphere**

346 Airborne bacterial and fungal concentration in PM_{2.5} were $3.29 \pm 2.38 \times 10^6$ copies/m³
347 ($1.99 \pm 1.44 \times 10^6$ bacteria m⁻³, $P=0.06$) and $1.69 \pm 0.55 \times 10^6$ copies m⁻³ ($3.39 \pm 1.10 \times 10^4$
348 fungal spores m⁻³, $P=0.04$). This value was comparable to the average abundance of
349 bacteria and fungi in the atmospheric boundary layer over land, which is approximately
350 1.9×10^4 bacteria m⁻³ and 2.4×10^4 fungal spores m⁻³ (Mayol et al., 2014). A high
351 concentration was observed in terrestrial air-mass samples, with the average value of
352 $2.59 \pm 2.16 \times 10^6$ copies m⁻³ ($P=0.12$) and $3.18 \pm 0.85 \times 10^6$ copies m⁻³ ($P=0.007$) for
353 bacteria and fungi, respectively. Microbial concentration in marine air-mass samples

354 was significantly lower, with the average concentrations of bacteria and fungi being
355 $2.70 \pm 1.00 \times 10^5$ copies m^{-3} ($P=0.04$) and $3.08 \pm 1.55 \times 10^5$ copies m^{-3} ($P=0.07$).

356 Airborne microbial community structure exhibited significant variation influenced
357 by marine and terrestrial air masses (Fig. 2). Predominantly, *Proteobacteria* (40.06%),
358 *Firmicutes* (36.30%), *Actinobacteria* (8.97%), *Bacteroidetes* (8.29%), and
359 *Deinococcus-Thermus* (4.59%) were identified as the most abundant bacteria. Notably,
360 *Actinobacteria* and *Deinococcus-Thermus* were found in high concentration in samples
361 from terrestrial air masses. In particular, *Deinococcus-Thermus* demonstrated a relative
362 abundance of 7.9% in terrestrial air mass samples, significantly surpassing that of
363 marine (1.2%) and mixed air mass samples (1.5%). *Cyanobacteria* exhibit a higher
364 concentration in both terrestrial and mixed air-masses samples. *Cyanobacteria* are
365 prevalent in freshwater, soil, surface crusts in deserts (Cordeiro et al., 2020; Temraleeva
366 et al., 2016; Curren and Leong, 2020). In comparison to terrestrial environments, the
367 prevalence of *Cyanobacteria* is notably diminished in colder marine environments
368 (Koh et al., 2012). Atmospheric *Cyanobacteria* in the coastal city may originate from
369 the terrestrial environment, subsequently transported via terrestrial and mixed air
370 masses. *Proteobacteria* was the most prevalent taxon in marine air-mass samples. The
371 predominant metabolic activity in deep-sea sediments is attributed to this group, with
372 major taxa being found in marine sediments (Huang et al., 2021). The predominant
373 bacterial genera included *Staphylococcus* (21.02%), *Enhydrobacter* (6.43%),
374 *Lactobacillus* (6.03%), and *Deinococcus* (4.56%). These bacteria included a series of
375 opportunistic pathogens and were found abundant in terrestrial and mixed air masses
376 samples. Relative abundance of bacteria such as *Comamonas*, *Streptococcus*,
377 *Ruminococcus*, and *Enterobacteriaceae* were higher in marine air-mass samples. These
378 predominantly consisted of saprophytic bacteria and intestinal microorganisms.

379 The dominant fungal phyla were *Ascomycota* (77.29%) and *Basidiomycota* (21.58%),
 380 which were similar to the previous studies (Du et al., 2018; Liu et al., 2019; Zeng et al.,
 381 2019). Fungal community influenced by terrestrial and mixed air masses were quite
 382 similar, with relatively higher abundances of opportunistic pathogens such as
 383 *Malassezia*, *Alternaria*, *Cladosporium*. In contrast the saprophytic *Aspergillus*,
 384 *Davidiellaceae* and *Coprinellus* were abundant in marine air-mass samples.



385

386 **Fig. 2. Bacterial and fungal community composition and function influenced by different air**
 387 **masses. The community concentration, main phylum (a), (b), genus (c), (d), and community**
 388 **function (e), (f) are indicated.**

389 The FAPROTAX analysis revealed that the primary bacterial ecological functions
 390 were chemoheterotrophy, aerobic chemoheterotrophy, fermentation, and human

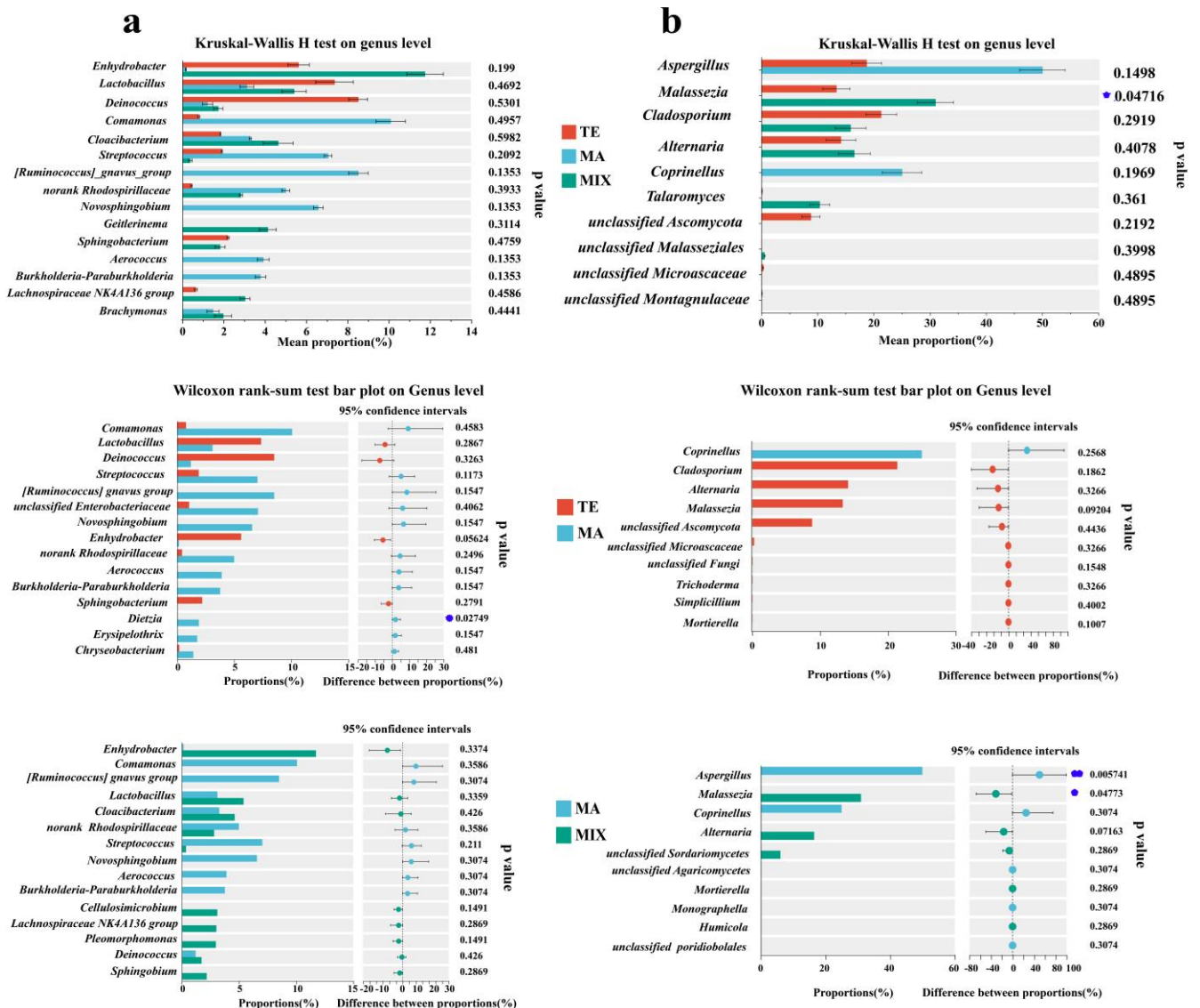
391 pathogens. These accounted for 74%, 47%, and 44% respectively in terrestrial, marine,
392 and mixed air-mass samples. Notably, human pathogens and animal parasites or
393 symbionts were more prevalent in terrestrial and mixed air-mass samples than in marine
394 air-mass samples. Marine air-mass samples were enriched with mammal gut bacteria,
395 as well as hydrocarbon and aromatic compound degradation bacteria. The fungal
396 community function in terrestrial and mixed air-mass samples were similar, with
397 undefined Saprotroph, Animal Pathogen-Undefined Saprotroph, Animal Pathogen-
398 Endophyte-Lichen Parasite, Animal Pathogen-Endophyte-Plant Pathogen were the
399 main functions, which totaled 99.27%, 99.98%, 99.27% in the terrestrial, marine, and
400 mixed air-mass samples. In particular, the observed prevalence of Saprotroph fungi was
401 higher in samples from marine air masses, such as those containing *Aspergillus*. Notably,
402 the fungi associated with terrestrial air masses predominantly carried animal pathogens
403 and exhibited greater pathogenicity, including species such as *Malassezia* and
404 *Alternaria*.

405 **4. Discussion**

406 **4.1 Community disparities influenced by terrestrial and marine air masses**

407 Overall, bacterial and fungal community concentration influenced by terrestrial and
408 mixed air masses were higher than those in marine air mass samples. Bioaerosols
409 originating from anthropogenic environments, such as sewage treatment plants and
410 hospitals, may harbor more resistant bacteria. Microorganisms from arid areas like
411 deserts possess a higher number of species adapted to harsh conditions, including
412 resistance to temperature fluctuations, dryness, and UV radiation. These include
413 bacteria such as *Bacillus*, *Streptococcus*, and *Deinococcus* (Maki et al., 2010; Park et
414 al., 2018; Qi et al., 2021). These microorganisms can adhere to particulate matter during

415 long-distance transport of air masses after being released from their terrestrial habitats.



417 **Fig. 3 Bacterial (a) and fungal (b) community disparities influenced by terrestrial (n=10),**
 418 **marine (n=6), and mixed air mass (n=8).**
 419

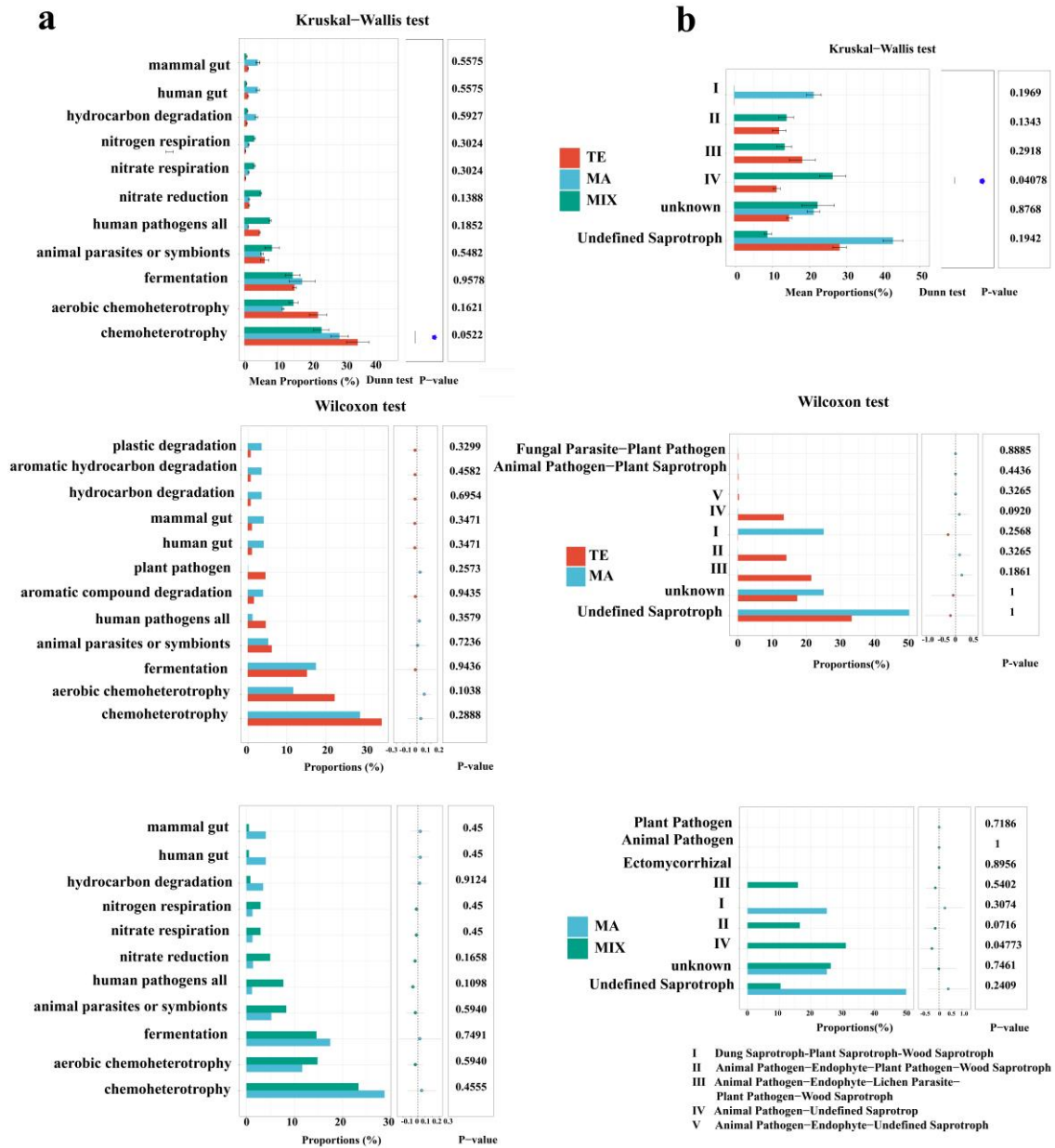
420 Community disparities influenced by terrestrial, marine, and mixed air masses was
 421 shown in Table S2, S3, Fig. 3 and Fig. 4. The Principal Coordinates Analysis (PCoA)
 422 revealed distinct clusters corresponding to terrestrial, marine, and mixed air mass
 423 samples (Fig. S4). In terms of bacterial community variations, numerous dominant
 424 microorganisms exhibited obvious differences in abundance between marine and
 425 terrestrial air mass samples, although these differences were not statistically significant
 426 ($p > 0.05$). For instance, the bacterial enrichments in terrestrial and mixed air-masses

427 included *Staphylococcus*, *Deinococcus*, *Lactobacillus*, and *Sphingomonas*.
428 *Staphylococcus* is a pathogenic bacterium widely found in human skin and intestines
429 that can cause various infections (Cheung et al., 2021). As a radiation-resistant
430 bacterium, *Deinococcus-Thermus* is capable of withstanding harsh environmental
431 conditions (Callegan et al., 2008; Rainey et al., 2007). *Lactobacillus*, a genus of Gram-
432 positive bacteria, has also been identified as abundant in atmospheric dust (Federici et
433 al., 2018; Xu et al., 2017). *Sphingomonas*, typically found in water and soil, which is
434 widely known as the metabolic mechanism ability to endure poor nutrients and
435 degradation of refractory compounds, such as polycyclic aromatic hydrocarbons (PAHs)
436 (Hu et al., 2007; Sun et al., 2018). The majority of these bacteria possess the ability to
437 produce spores, which are capable of withstanding harsh conditions such as low
438 temperatures, aridity, and radiation during long-distance transmission, thereby ensuring
439 their survival throughout this process (Griffin et al., 2003). *Comamonas* was identified
440 as the dominant bacterium in the coastal atmosphere of Weihai, particularly prevalent
441 in marine air-mass samples. This bacterium is commonly associated with
442 environmental bioremediation and is predominant in oligotrophic environments (Yan et
443 al., 2012; Zhang et al., 2024). *Streptococcus* are mostly found in the oral and
444 gastrointestinal tracts and also play an important role in human infections (Brouwer et
445 al., 2023). *Ruminococci*, and *Enterobacteriaceae* are typical gut microorganisms
446 present in humans and animals. In the coastal atmosphere, the identified sequences of
447 these gut bacteria might be associated with intestinal microorganisms in marine fish
448 and mammal animals.

449 For fungal community, *Aspergillus* demonstrated a significant differentiation
450 between marine and mixed air mass samples ($P=0.005$). The highest proportion was
451 noted in samples from marine air masses, at 53.7%. In contrast, the values were 20.1%

452 and 0.3% in terrestrial and mixed air masses respectively. Previous studies have
453 confirmed that *Aspergillus* is widely distributed in nature and unpolluted environments
454 (Kendrick, 1995). This fungus is predominantly found in offshore regions, such as
455 Qingdao, China (Li et al., 2011). Furthermore, the saprophytic *Aspergillus* was also
456 prevalent in clean samples during periods of haze pollution and was frequently detected
457 on non-haze days (Yan et al., 2016). *Malassezia* was higher in terrestrial and mixed air-
458 mass samples ($P=0.047$), which has been found to be widespread in a variety of animals.
459 As a parasitic fungus, *Malassezia* causes the majority of skin diseases, such as dandruff
460 and seborrheic dermatitis caused by *Malassezia sphericalis* (Deangelis et al., 2007).
461 Furthermore, *Cladosporium* exhibits higher abundance in terrestrial and mixed air-mass
462 samples compared to marine air masses, although this difference lacks statistical
463 significance ($P>0.05$). This species is ubiquitous worldwide, commonly found in a wide
464 variety of plants, known to be a common endophyte as well as a foliar fungus (El-Morsy,
465 2000; Islam and Hasin, 2000). It is frequently isolated from soils, foodstuffs and organic
466 matters (Bensch et al., 2012). Moreover, *Cladosporium* have been commonly observed
467 in terrestrial atmospheric environments, demonstrating the potential origins from
468 continental environments (Frączek et al., 2017; Han et al., 2019).

469 The significant variance in the ecological function of microbial community
470 was predominantly evident in the Animal Pathogen-Undefined Saprotrophs ($P=0.04$).
471 These pathogenic fungi were more prevalent in terrestrial and mixed air-mass samples
472 than in marine air-mass samples. Marine air-mass samples were enriched with mammal
473 gut bacteria, hydrocarbon and aromatic compound degradation bacteria, and undefined
474 Saprotroph fungi.



475

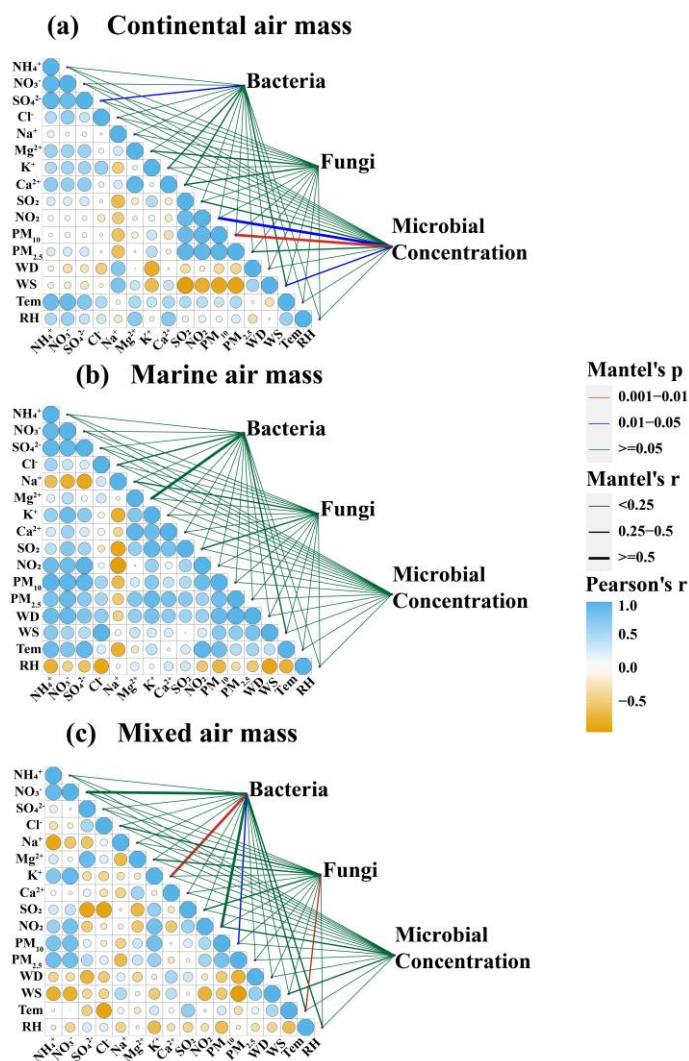
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477

Fig. 4. Bacterial (a) and fungal (b) community function disparities influenced by terrestrial (n=10), marine (n=6), and mixed mass (n=8).

478 4.2 Implications of environmental factors on coastal airborne microbes

479 To elucidate the contributions of marine and terrestrial air masses to microbial
480 aerosols in the coastal city, a Mantel correlation analysis was conducted between
481 microbial aerosols and various environmental factors (Fig.5, Fig.S4). The bacterial
482 community exhibited a significant positive correlation with SO_4^{2-} due to the influence
483 of terrestrial and mixed air masses ($P < 0.05$). Both bacterial and fungal communities
484 demonstrated positive correlations with continental ions such as K^+ , Mg^{2+} , and Ca^{2+}
485 (Shi et al., 2022). Additionally, a positive correlation was observed between bacterial
486 and fungal concentrations and NO_2 ($P < 0.05$), as well as a significant positive
487 correlation with PM_{10} ($P < 0.01$). Air masses transported over long distances from the
488 continent appear to harbor diverse and abundant microbial populations (Kakikawa et
489 al., 2009; Deleon-Rodriguez et al., 2013). For instance, the long-distance transportation
490 of dust particles from northwestern winds in the Inner Mongolia region may have
491 altered the community structure and abundance (Qi et al., 2021; Xia et al., 2024). Dust-
492 borne bacteria, such as *Staphylococcus*, *Delftia*, *Pseudoalteromonas* and *Deinococcus*,
493 are likely introduced into the atmosphere during Asian Dust events. Most of these
494 bacteria accompany the transportation of dust particles to the downwind coastal city.
495 Similarly, microbial communities showed high positive correlation with ions from
496 continental sources, such as K^+ , Mg^{2+} , and Ca^{2+} in terrestrial and mixed air mass
497 samples. K^+ is the signature ion of biomass combustion (Mason et al., 2016; Yu et al.,
498 2018). Mg^{2+} and Ca^{2+} are mostly derived from crustal elements (Zhang et al., 2012).
499 The dominant bacteria within the microbial community exhibited a significant positive
500 correlation with these three ions simultaneously, suggesting a similar sources. They
501 generally originate from terrestrial sources, including anthropogenic biomass
502 combustion and soil dust.



503

504 **Fig.5 Mantel analysis reveal the correlation between microbial community and various**
 505 **environmental factors under different air masses, terrestrial air masses (a) marine air**
 506 **masses (b) and mixed air masses (c). The Pearson's r coefficient indicates the significant**
 507 **correlation at $p < 0.05$ and $p < 0.01$.**

508

509 The concentrations of particulate matter and water-soluble ions in samples affected
 510 by the marine air mass are notably low, indicating that the majority of chemical
 511 elements did not significantly influence the composition and abundance of the
 512 microbial community. However, a significant positive correlation was observed
 513 between Na^+ and certain dominant species such as *Talaromyces*, *Monographella*, and
 514 *Phoma* ($P < 0.05$). Similarly, Mg^{2+} was found to be significantly positively correlated
 515 with *Malassezia* in mixed air masses ($P < 0.01$) (Fig. S5). Besides anthropogenic sources,

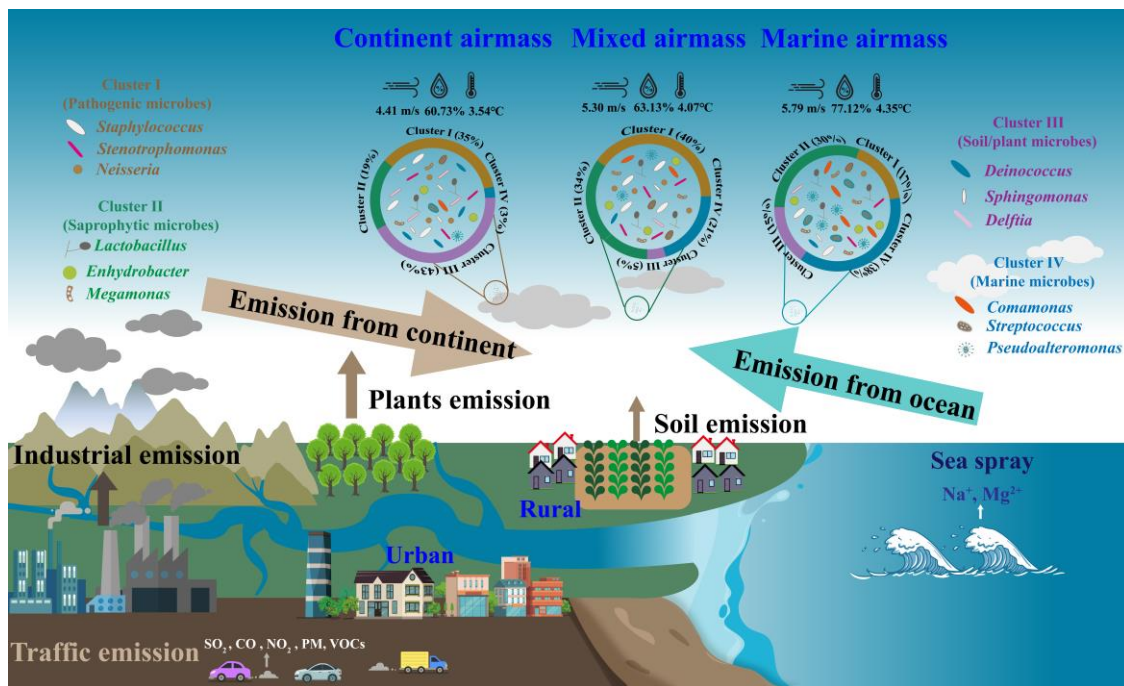
516 such as industrial emissions, the origins of Mg^{2+} and Na^+ in coastal regions should also
517 consider the impact of sea salt (Sun et al., 2022). Samples collected under the influence
518 of marine air masses had a positive correlation between Mg^{2+} and Na^+ ($r^2 = 0.67$),
519 indicating that they have a similar origin of sea salt. Furthermore, the mean Mg^{2+}/Na^+
520 ratio was found to be 0.11, which is proximate to the seawater value of 0.12 (Seinfeld
521 and Pandis, 1997).

522 Influenced by mixed air masses, temperature has a greater impact on fungal
523 community, which was positively correlated with *Malasseziales* and *Davidiellaceae*. In
524 marine air mass samples, a positive correlation between air temperature and certain
525 microorganisms (*Aerococcus*, *Cloacibacterium*, *Sphingobium*, *Enhydrobacterium*,
526 *Davidiellaceae*, *Malasseziales*) also indicated that the increase in air temperature in
527 spring favors the survival of airborne microbes (Jones and Harrison, 2004).

528 **4.3 Intermixing of bioaerosols from terrestrial and marine air-mass**

529 Coastal atmospheres exhibit a complex amalgamation of terrestrial and marine
530 aerosols, with the characteristics significantly influenced by the origins of the air
531 masses. Consequently, these coastal aerosols offer ideal conditions for examining the
532 mixing processes between natural and anthropogenic emission from terrestrial and
533 marine sources, respectively. This work shows that enormous levels of haze aerosols
534 from the terrestrial and mixed air mass can be rapidly transported into the coastal city
535 during regional haze or dust pollution. During the long-distance transportation of air
536 masses, a comprehensive mixture of soil-derived, biogenic, and anthropogenic
537 microorganisms from terrestrial air masses, and aquatic, saprophytic, gut
538 microorganisms from marine environments are fully mixed (Fig. 6). This amalgamation
539 is particularly evident in the samples of mixed air masses. Within these, a significant

540 quantity of chemical components, bacteria and fungi, as well as opportunistic pathogens,
541 were transported into the coastal city from the continent. Microbial communities were
542 strongly correlated with haze aerosols, e.g., WSIs in PM_{2.5} from terrestrial and mixed
543 air mass. The primary influence on terrestrial air mass was anthropogenic emissions,
544 with coal combustion for winter heating and biomass burning being the predominant
545 pollution sources. Moreover, dust events in spring carried higher concentration of
546 particulate matter. These air pollutants can act as initial source of bioaerosols such as
547 bacteria and fungi, thereby providing a site for attachment reproduction (Jiang et al.,
548 2022). Additionally, water-soluble ions in PM, primarily secondary inorganic ions, such
549 as sulfate, nitrate, and ammonium ions, can supply essential nutrients for microbial
550 growth (Fan et al., 2019). This elucidates the elevated concentration of particulate
551 matter and microorganisms in terrestrial air mass during heavy pollution.
552 Simultaneously, it is pertinent to highlight that during such pollution events, the
553 terrestrial air mass exacerbates the pollution process, leading to a marked increase in
554 the proportion of pathogenic microorganisms.



555

556 **Fig. 6 Implications of intermixing of bioaerosols from terrestrial and marine air-mass on the**
557 **composition of airborne microbial communities.**

558

559 In the coastal city of Weihai, marine air masses predominantly originate from the
560 expansive western Pacific Ocean, located to the eastern and southern areas. These air
561 masses are typically characterized by their cleanliness and contain a significant quantity
562 of saprophytic and intestinal microorganisms that infiltrate the atmosphere via ocean
563 droplets (Mayol et al., 2014). The presence of microorganisms in marine air masses
564 exhibited a significant correlation with sea salt ions, specifically Na^+ and Mg^{2+} . These
565 ions, when introduced into the atmosphere, form aerosolized particulate matter. The
566 introduction of marine bacteria into tropospheric free space may alter the airborne
567 microbial composition in continental regions (Cho and Hwang, 2011; Amato et al.,
568 2007; Maki et al., 2014; Polymenakou et al., 2008; Cáliz et al., 2018), thereby
569 significantly impacting bioaerosols over coastal atmospheres.

570 Notably, marine air masses from the northerly areas, cross China's inland Bohai Sea
571 and then pass through urban and rural areas with high human activity intensity. These
572 areas have more anthropogenic microorganisms, including pathogenic microorganisms

573 from hospitals and sewage treatment plants. In this study, samples within this category
574 are defined as those influenced by mixed air mass. This also signifies the amalgamation
575 of marine and terrestrial microorganisms. Consequently, within the mixed air mass
576 samples, the prevalence of pathogenic and saprophytic microorganisms is comparable.
577 This similarity underscores the complex impact of both terrestrial and marine sources.
578 Populations in coastal cities may also be susceptible to exposure to these bioaerosols
579 and pathogens, which are transported over long distances during regional air pollution.

580 **5. Conclusion**

581 This study evaluated the potential influence and the mixing effects between natural
582 and anthropogenic aerosols respectively from terrestrial and marine air mass on
583 biological constituents of PM_{2.5} over coastal atmosphere. The concentration of water-
584 soluble ions, metal elements and bioaerosols were higher in terrestrial and mixed air-
585 mass samples than marine air-mass samples. The V/Ni ration greater than 0.7 in marine
586 air-mass samples indicated significant influence from marine ship emissions. Bacterial
587 and fungal community influenced by terrestrial and mixed air masses were enriched in
588 pathogens, including the *Staphylococcus*, *Malassezia* and *Alternaria*. In contrast,
589 marine air mass carried more gut and saprophytic microbes, such as *Ruminococcus*,
590 *Enterobacteriaceae*, *Aspergillus*, and *Davidiellaceae*. Potential implications of
591 environmental factors on airborne microbes based on Mantel correlation analysis
592 suggested that the bacterial community was mainly significantly correlated with SO₄²⁻,
593 K⁺, Mg²⁺, and Ca²⁺ and PM₁₀ influenced by terrestrial and mixed air masses from
594 continental sources. In contrast, influenced by marine air masses, bacterial and fungal
595 community were strongly correlated with sea salt ions, such as Na⁺. The predominant
596 bacteria and fungi in marine and mixed air-mass samples exhibited a positive

597 correlation with air temperature in spring, suggesting elevated temperature is conducive
598 to the survival and reproduction of bioaerosols. Our research offers a novel perspective
599 on the amalgamation of terrestrial and marine aerosols in the coastal atmosphere. In the
600 mixed air-mass samples, a complex amalgamation of soil-derived and anthropogenic
601 microorganisms from terrestrial air masses, along with aquatic, saprophytic, and gut
602 microorganisms from marine environments, are fully intermixed during the long-
603 distance transportation of air masses. This process significantly influences bioaerosols
604 in coastal atmospheres. This understanding will enhance our comprehension of the
605 environmental and climatic impacts on microbial aerosols within the marine boundary
606 layer. Future research endeavors will concentrate on the local emission and long-
607 distance transport, as well as the contributions of both terrestrial and marine air masses
608 on bioaerosol components in coastal cities.

609 **CRedit authorship contribution statement**

610 All authors contributed to the manuscript and have given approval of the final version.
611 Min Wei designed the study. Qun He performed the data analysis and wrote the original
612 manuscript. Min Wei assisted with the sampling, Qun He, Zhaowen Wang and Rongbao
613 Duan conducted the experiments and performed the statistical analyses. Houfeng Liu,
614 Min Wei, and Pengju Xu contributed to the interpretation of results. Min Wei, Caihong
615 Xu and Jianmin Chen revised the manuscript.

616 **Data availability**

617 The hourly concentration of PM_{2.5}, PM₁₀, CO, SO₂, NO₂, and O₃ were systematically
618 retrieved from the National Ambient Air Quality Monitoring System
619 (<http://www.cnemc.cn/>); the historical air quality data was sourced from the nearest

620 environmental monitoring center (<https://www.aqistudy.cn/historydata/>); MeteInfo
621 backward trajectory model (MeteoInfo 3.7.4 - Java, <http://www.meteothink.org>
622 /downloads/index.html); meteorological data were obtained from GDAS1
623 (<ftp://arlftp.arlhq.noaa.gov/pub/archives/gdas1/>), and meteorological data as well as
624 bacterial and fungal concentrations are available from the authors on request
625 (minwei@sdnu.edu.cn). The raw sequences of bacterial 16S rRNA and fungal ITS gene
626 are available in the Sequence Read Archive (SRA) under accession number
627 PRJNA1101427
628 (https://www.ncbi.nlm.nih.gov/Traces/study/?acc=PRJNA1101427&o=acc_s%3Aa)
629 and PRJNA1101176
630 (<https://dataview.ncbi.nlm.nih.gov/object/PRJNA1101176?reviewer=mu6a68nm61vn>
631 h0g1lahff3qvma).

632 **Declaration of competing interest**

633 The authors declare that they have no conflict of interests.

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