

1 Measurement of greenhouse gas fluxes in agricultural soils with a flexible, open-design  
2 automated system

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## Abstract

Over the last decades and due to the current climate change situation, the study of the impacts of human activities on climate has reached great importance, being agriculture one of the main sources of soil greenhouse gas. There are different techniques to quantify the soil gas fluxes, such as micrometeorological techniques or chamber techniques, being the last one capable to assess different treatment at the same site. Manual chambers are the most common one. However, manual chambers are characterised by low sampling frequency, typically one sample per day is considered a high sampling frequency. Therefore, a great deal of effort is required to monitor short-term emission events such as fertilisation or rewetting. For this reason, automated chamber systems are an opportunity to improve soil gas flux determination, but their distribution is still scarce due to the cost and challenging technical implementation. The objective of this study was to develop an automated chamber system for agricultural systems compare it with a manual chamber system.. Moreover, over a period of one month the soil gas fluxes were determined by both systems to compare their capabilities to capture the temporal variability of soil gas emissions. The automated system reported soil GHG fluxes up to 58 and 40% greater for CO<sub>2</sub> and N<sub>2</sub>O fluxes compared to the manual chamber system. Additionally, the higher sampling frequency of the automated chamber system allowed to capture the daily flux variations, resulting in a more accurate estimation of cumulative soil gas emissions. Furthermore, the assessment of various sampling intervals for single-day measurements indicated that soil gas sampling between 10:00 and 12:00 was the optimal time interval for obtaining representative daily emissions. The study emphasises the importance of chamber dimension and shape in the development of chamber systems, as well as sampling frequency and sampling hour, for manual chamber system.

## 1. Introduction

Agriculture and land-use changes are significant contributors to climate change, accounting for a 22% of total global emissions of greenhouse gases (GHG) (IPCC, 2023). Moreover, agricultural emissions are expected to increase along with food demand (Wiebe et al., 2019). Microbial activity is the primary driver of the production and emission of different soil GHG. Microbial processes are influenced by several abiotic factors such as soil water content, soil temperature or nutrient availability. The different farming practices – i.e. crop rotation, fertilization, irrigation – have a significant impact on these factors, and, therefore, they can have a great influence on soil GHG emissions (Oertel et al., 2016). By accurately measuring soil GHG emissions, it is possible to identify the major sources and understand the impact associated with various farming practices. This valuable information can be provided to policymakers and regulators to develop science-based policies and regulations that incentivize farmers to adopt more sustainable practices. Thus, measuring soil GHG emissions in agriculture is crucial to promote sustainable farming practices, that can mitigate climate change.

The use of manual chambers is one of the most widespread methods for studying soil GHG emissions at small spatial and temporal scales (Collier et al., 2014). Chambers are designed to establish an enclosed environment, facilitating the periodic collection of gases emitted from or consumed in the soil. Subsequently, the gas samples are subjected to laboratory analysis through gas chromatography (Harvey et al., 2020). These analyses determine the concentration of GHG within the chamber headspace and allow the calculation of emission rates based on the change in gas concentration over a given time span. This method is characterized by its simplicity and versatility as chambers are relatively simple to use and can be employed across diverse ecosystems and soil types (de Klein et al., 2020). Manual chambers are relatively simple to construct and can be tailored to fit specific research requirements. Besides, compared to alternative methods, they entail relatively low cost. However, they have as well some limitations. For instance, their

measurement frequency is restricted due to the time-intensive nature of manual sampling and subsequent analysis, making high-frequency sampling impractical. Usually, sampling frequency is not higher than one sampling per day, but it's well established that sampling frequency affects annual GHG estimations (Barton et al., 2015). For this reason, efforts are often concentrated on intense sampling frequencies during short periods (hours to days) when significant emissions peaks are expected, but later, during the rest of the campaign, samplings are carried out every 1 to 4 weeks (or even sometimes not considered). Another aspect to consider involves the notable soil disruption caused when samples need to be collected, such as after an irrigation event.

In contrast to manual chambers, the utilization of automated chambers coupled with an in-situ gas analyser allows sampling at a higher temporal frequency. Consequently, these automated systems more comprehensively capture temporal variations, enhancing insight into the dynamics of soil GHG emissions on a daily and seasonal basis (Grace et al., 2020). Automation also ensures capturing fluxes linked to unexpected events (such as rainstorms), obtaining data in areas of difficult access, and reducing the impact of soil disturbance on measurements. However, this method requires costly equipment and skilled operators and implies different infrastructure constraints, factors that result in lower spatial coverage compared to what can be achieved with manual systems. Moreover, these automated chamber systems are beginning to be manufactured and distributed by companies dedicated to the manufacture of gas analysers, with the limitation of being close systems to be modified. Based on that situation, over recent decades, several groups have crafted automated systems (Lognoul et al., 2017, Lawrence and Hall, 2020).

To date, the number of experiences using automated chambers coupled with in situ gas analysers is scarce and, as far as we have been able to find out, none of these previous studies used chamber systems consisting of a total of 12 individual chambers. The objective of this paper is to present an innovative non-commercial soil GHG measurement system based on automated chambers linked to an in situ photoacoustic multigas analyser and describe its operational details.

A, a comparison between this automated system and the manual static chamber methodology is also presented.

## 2. Materials and Methods

### 2.1. Automated system description

In this section, we present an automated chamber system tailored for monitoring soil gas emissions. By integrating openness, cost-effectiveness, and versatility, this system facilitates precise and dynamic measurements of soil GHG fluxes. Our design principles focused on building an adaptable configuration and real-time functionality, alluding to its potential importance in agricultural and environmental research. The system consists of three main parts: the chambers, the set of solenoid valves controlled by a computer (central control unit) and the multigas analyser (Figure 1).

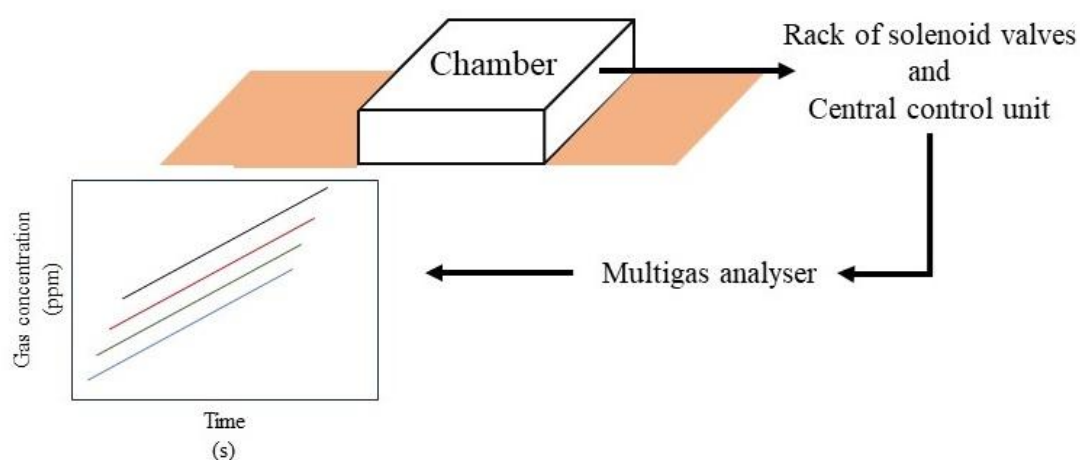


Figure 1. General scheme of the automated soil GHG measurement system.

### 2.2. Soil chamber design

Soil chambers, ‘Queensland’ design, have been built following a model provided by the Terrestrial Bio-Geo-Chemistry Division (Institute of Meteorology and Climate Research, Atmos. Environ.al Research (IMK-IFU), Karlsruhe Institute of Technology (KIT)). Chambers consisted of an aluminum structure of 0.50 x 0.50 m length and width and 0.15 m height closed with

methacrylate panels and two lids 0.50 x 0.25 m width and length that are controlled by four pneumatic actuators, two per lid (Figure 2a). Besides, lids open at a 90° angle allowing rainfall or irrigation water supply to reach the soil surface of the area covered by the chambers. All methacrylate panels were coated with an aluminum bubble foil to keep the internal chamber temperature homogeneous during the enclosure time. Moreover, a rubber seal was fixed to the lids and the bottom part of each chamber to ensure a hermetic close and avoid gas leakage during the sampling process.

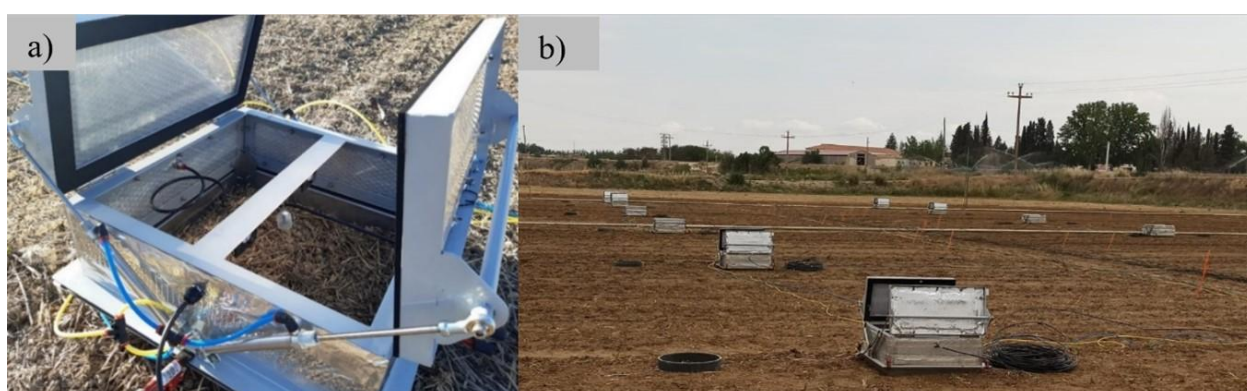


Figure. 2. (a) Open automated chamber deployed in the field trial ('Queensland' design). (b) Set of chambers deployed in the field trial. Dark rings next to chambers are the bases for manual chambers.

The gas sample line (polyethylene coated aluminum tube, Eaton Sinflex. 6/4mm external internal diameter, respectively) entered each chamber via one of the side panels, positioned approximately halfway up. In the central area of the chamber, the tube was bent facing downwards and the tip was protected by a small PVC funnel to prevent water condensation at the tube inlet. A vent (matching the material and diameter of the gas sampling line) was positioned on the opposite side panel to equalize pressure between the chamber's interior and exterior during flux measurements. Moreover, each chamber has two small fans (60x60x25 mm 12V; 4000 rpm. EVERCOOL EC6025L12EA) to promote air mixing inside the chamber.

Three chambers were equipped with a threaded cable gland on a lateral methacrylate panel for mounting a thermistor (107, Campbell Scientific Ltd., UK) to monitor internal chamber temperature. Chambers were attached by clamps to stainless steel bases (0.5 x 0,5 x 0.1 m) with sharp edges at the bottom that were inserted 0.10 m into the soil . Plants (crop and weeds) growing inside the chambers were cut since the crop during this experiment was maize (*Zea mays* L.) (Figure 2b). The cost of each chamber, including the solenoid valve and the sampling line is 600 €.

### ***2.3. Automated chamber operation***

The chambers opened and closed by means of pneumatic actuators. This setup comprised an air compressor delivering pressure to the pneumatic actuators. Inside a shed located next to the field trial, three solenoid valves installed in a panel, received air from the compressor (6 bar) and directed compressed air to the chambers. Routing of compressed air was facilitated by an external relay controller (8 relay board, 24V 6.5A, YWBL-WH) directly linked to the computer. In the configuration of this study, three sets of four chambers each opened and closed simultaneously. Similarly, each sampling line from each chamber was connected to a two-way solenoid valve that regulated the entry of the gas sample from each of the chambers to the photoacoustic multi-gas analyser (Gasera One, Gasera Ltd, Finland). The two-way solenoid valves were connected to a relay board (16 relay board, 24V 6.5A, YWBL-WH) that controlled which valve was activated (Figure 3).

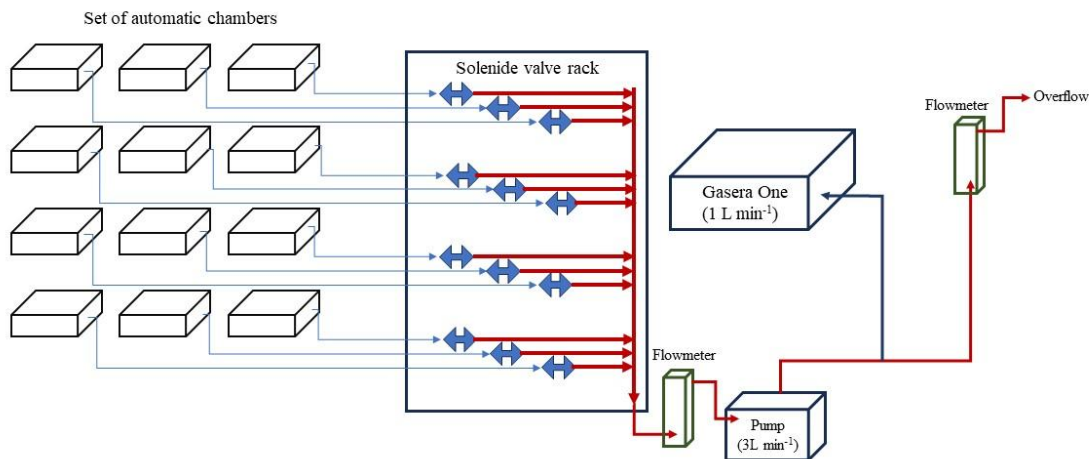


Figure. 3. Description of the automated chamber system.

To bring the gas from the chamber to the gas analyser, an external diaphragm pump (KNF NMP830KNDC 12V, KNF Neuberger, Inc, Freiburg im Breisgau, Germany) was coupled to the two-way solenoid valve bank. This pump continuously drew air from the activated sampling line, maintaining a flow rate of  $3\text{ L min}^{-1}$ . The gas analyser (Analysis cell volume  $30\text{ mL}$ ) drew sample gas from this primary line at a rate of  $1\text{ L min}^{-1}$  for a duration of six seconds every one and a half minutes (Figure 3c). Two flowmeters were attached to the main line. The initial one, positioned after the pump and preceding the gas analyser, regulated the gas flow delivered to the analyser. The second flowmeter ensured a continuous overflow greater than  $1\text{ L min}^{-1}$ , guaranteeing sufficient gas flow from the active sampling line to the gas analyser (Figure 3).

The solenoid valve banks, pneumatic system, chamber sampling lines, and gas analyser were all managed through a custom script created using R statistical software version 4.2.2 (R Core Team, 2022). This R script, governed by the time taken by the analyser to process the sample, can be easily modified by setting the total number of chambers or, if it is necessary to work by blocks, by setting the number of blocks and the number of chambers per block. One of the advantages of this system is the self-made multiplexer that allows to modify the number of



chambers easily compared to other multiplexers like Gasera Multipoint Sampler (Gasera Ltd, Finland) which has a close configuration of 8 or 12 channels. Moreover, the use of relay boards that could be configured by Arduino or easily integrated into the R script as the selected ones, as an alternative to control modules, for example, I-7060D (ICP DAS CO, LTD) that only have four channels per module, simplifies the configuration of the script, since just with one board it's possible to handle all the chambers. For this field experiment, the current setup consists of 3 blocks of four chambers each block. This configuration responds to the needs of the current experimental design, however, since it is an open system, the configuration is variable and can be individualised for each of the chambers.

#### ***2.4. Evaluation of the automated measurement system***

Over the last decade, the current research team members have successfully conducted several GHG flux studies using a manual closed chamber system (Álvaro-Fuentes et al., 2016, Franco-Luesma et al., 2019, 2020a, 2020b, 2022). Based on that, an evaluation experiment was carried out to compare the soil gas fluxes obtained via the newly developed automated chamber system against the conventional manual chamber system used regularly by the research group. This evaluation experiment was aimed to evaluate the impact of i) the chamber design and ii) the sampling frequency and time on the differences in soil GHG fluxes between a manual and an automated chamber measurement system.

Manual chambers consisted of a Polyvinyl Chloride (PVC) cylinder of 0.315 m diameter and 0.2 m height coated by white thermal paint to avoid internal air temperature increasing during the deployed time. Each chamber was placed same diameter PVC collar inserted 0.05 m into the soil. A rubber septum was affixed atop the chamber to enable gas sampling via a plastic syringe equipped with a needle. Gas samples from each chamber were transferred to a 12 mL pre-evacuated glass vial (Exetainer Labco®). The concentrations of CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O in the gas

samples were determined by gas chromatography Agilent 7890B (Agilent, Santa Clara, CA, United States) equipped with an autosampler (PAL3 autosampler, Zwingen, Switzerland). Soil gas fluxes were determined based on the increase of the gas concentration during the deployment period. Further details of the gas chromatography method and manual chamber design could be found in Franco-Luesma et al. (2022).

The evaluation experiment took place in a maize (*Zea mays* L.) field trial sown on 10/05/2023 under irrigation conditions. The soil is a *Typic Xerofluvent* (Soil Survey Staff, 2015) with a silty loam texture, characterized by a basic pH of 8, a calcium carbonate content ( $\text{CaCO}_3$ ) of 48%, a total organic carbon content of 0.6% and a bulk density of  $1.33 \text{ g cm}^{-3}$  in the first 0.25 m soil depth. The area is characterized by a Mediterranean semiarid climate with a mean annual air temperature of  $14.1^\circ\text{C}$ , mean annual precipitation of 298 mm and mean annual reference evapotranspiration (ET<sub>o</sub>) of 1,243 mm. The meteorological data were obtained from a meteorological station situated at 0.5 km from the experimental site.

The evaluation experiment had two different steps. The first step consisted of simultaneous gas sampling with both manual and automated chamber systems on four different dates (i.e. 19/06/2023, 20/06/2023, 21/06/2023 and 28/06/2023). On June 19<sup>th</sup> and 20<sup>th</sup>, chambers were sampled once during 08:00 to 09:30. On June 21<sup>st</sup> and June 28<sup>th</sup>, chambers were sampled four times between 08:00 to 14:00. Consequently, a total of ten samplings were performed, covering four different days and different hours of the day to capture the possible diurnal variation of soil gas emission.

In this short time experiment, two chambers of each block were selected to compared with the manual chambers. The sampling sequence for the automated system was programmed to sample each chamber every five minutes, with a total enclosure time of 28 minutes. However, due

to the sequence configuration, the computable time for determining the soil gas fluxes was 20 minutes as described in the sequence diagram (Figure S1). The two manual chambers of each block were closed at the same time as the automated chamber and gas sampling was done at time 0 (first automated chamber sampling), at time 10 minutes and at time 20 minutes (coinciding with the last automated chamber sampling).

The second step of the evaluation experiment consisted of assessing the impact of the sampling time (i.e. hour of the day) and sampling frequency (i.e. 16 daily measurements vs 1 daily measurement for the automated and the manual chamber system, respectively) on the estimation of the soil gas fluxes. For that propose, from 22 of May 2023 to 29 of June 2023, soil CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O fluxes were measured simultaneously by the manual and automated chamber systems in the same field experiment

During this period, the sampling frequency and configuration of the automated chamber system was the same as it was used during the step one of the evaluation experiments. The twelve chambers were grouped in three set of four chambers each, being sampling every five minutes for 28 minutes, resulting in a total of 5 sampling points per chamber (Figure S1). However, the procedure followed in the manual chamber system was different and it consisted of the collection of three gas samples at time 0, 20 and 40 minutes after closing the chamber. The sampling frequency followed a daily frequency over the first five days and, afterwards, weekly measurements till the end of the experiment. For both chamber systems, the measuring instrument (i.e. photoacoustic multi-gas analyser and gas chromatography for automated and manual chamber systems, respectively) were calibrated by using 4 different ultra-high purity gas standards (Carbueros Metálicos, Barcelona, Spain, standard 1, 400 ppm CO<sub>2</sub>, 1.5 ppm CH<sub>4</sub>, 0.3 ppmN<sub>2</sub>O, standard 2, 800 ppm CO<sub>2</sub>, 2 ppm CH<sub>4</sub>, 1 ppmN<sub>2</sub>O, standard 3, 1500 ppm CO<sub>2</sub>, 4 ppm CH<sub>4</sub>, 3 ppmN<sub>2</sub>O, standard 4, 3000 ppm CO<sub>2</sub>, 6 ppm CH<sub>4</sub>, 6 ppmN<sub>2</sub>O) in order to standardize the concentration values obtained.

## 2.5. Data analysis

Soil gas flux (mg of gas m<sup>2</sup> day<sup>-1</sup>) of CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O, i.e.,  $f_{CO_2}$ ,  $f_{CH_4}$  and  $f_{N_2O}$  was calculated using the following equation (Eq. 1)

$$f_{gas} = \frac{Fit * MW * p * h}{R * T} * fT * fU \quad (Eq. 1)$$

where *Fit* represents the linear increase of gas concentration in the chamber over the enclosure time, *MW* is the molar weight of the atom in the gas molecule (i.e. 12 g mol<sup>-1</sup> for CO<sub>2</sub>-C and CH<sub>4</sub>-C and 28 g mol<sup>-1</sup> for N<sub>2</sub>O-N), *p* is the atmospheric pressure in Pa, *h* is the chamber height in m, *R* is the ideal gas constant in J K<sup>-1</sup> mol<sup>-1</sup>, *T* is the chamber air temperature in K, *fT* is the correction factor of time units, 1440 minutes day<sup>-1</sup> and *fU* is the unit correction factor, 10<sup>3</sup>. Cumulative soil CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O emissions were calculated using the trapezoid rule (Levy et al., 2017). Comparison between systems was done by linear fitting considering only soil gas fluxes that presented a R<sup>2</sup> higher than 0.8. Moreover, comparison in cumulative emissions between chamber system over one month was evaluated by one-way ANOVA. All analyses were done using the R statistical software version 4.2.2 (R Core Team, 2022).

### 3. Results and Discussion

#### 3.1. Automated system comparison

The comparison between the automated and manual measurement systems showed a linear response for the three gases compared. In the case of soil CO<sub>2</sub>, the automated system presented an average flux 58% greater compared to the manual system with a minimal flux difference of 425 mg CO<sub>2</sub>-C m<sup>2</sup> day<sup>1</sup> (Figure 4a). Data exhibited moderate dispersion ( $R^2=0.60$ ) revealing increased accuracy when manual fluxes were greater than 500 mg CO<sub>2</sub>-C m<sup>2</sup> day<sup>1</sup> (Figure 4a). Regarding CH<sub>4</sub> fluxes, the automated chamber system showed values greater than the fluxes obtained in the manual chamber system, showing a better fitting when fluxes were positive (Figure 4b). However, the lowest data dispersion between both measurement systems was obtained for soil N<sub>2</sub>O fluxes ( $R^2> 0.87$ ) but as observed for the other two gases, the automated chamber system reported fluxes values 40% greater than the manual chamber system (Figure 4c).

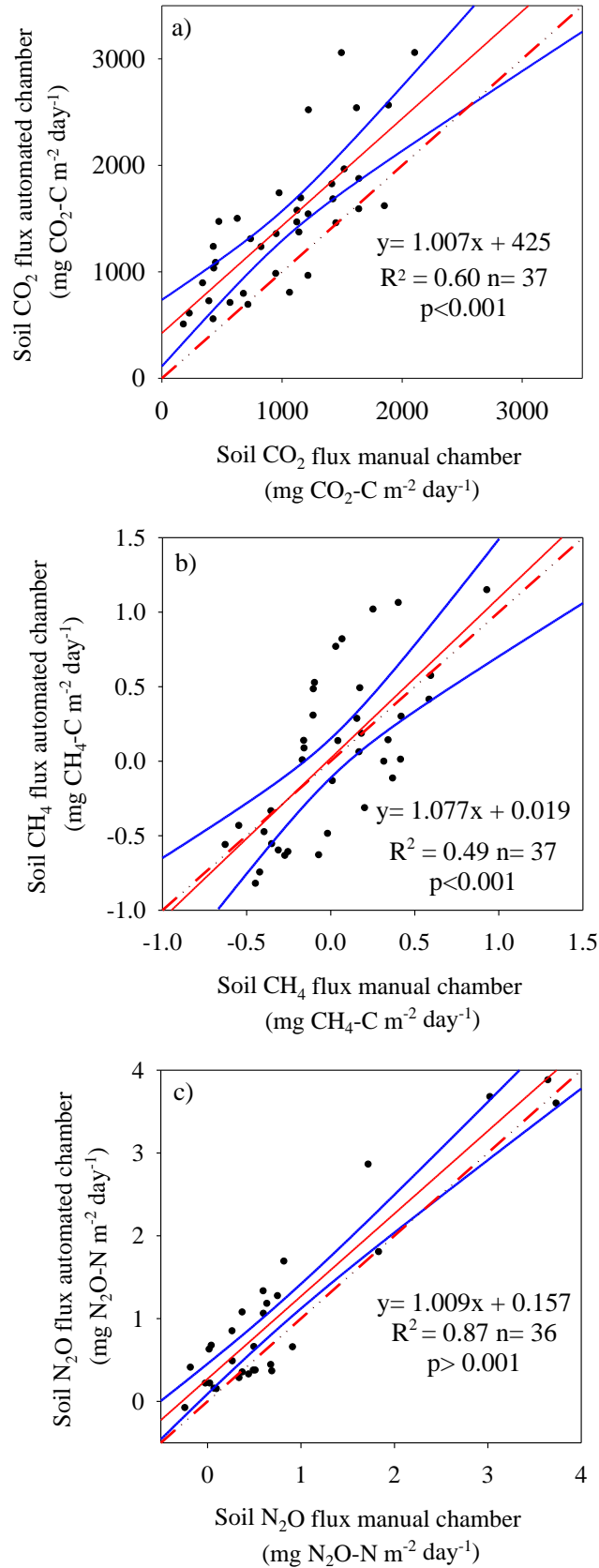


Figure 4. Comparison of soil gas flux between automated and manual chamber systems for carbon dioxide (CO<sub>2</sub>) fluxes (a), methane (CH<sub>4</sub>) fluxes (b) and nitrous oxide (N<sub>2</sub>O) fluxes (c). Blue solid lines represent 95% confidence intervals. Red dotted lines represent 1:1 line.

These differences between both measurement systems in flux magnitude and for the three studied gases may probably be a consequence of the different chamber shapes and dimensions. Hoffmann et al. (2018) found that the shape and dimension of the chamber have a significant effect on CO<sub>2</sub> fluxes, observing that small and cylindrical chambers tend to result in higher underestimation of CO<sub>2</sub> fluxes compared with large and squared chambers. In line with the previous authors, Pihlatie et al. (2013) also found a significant effect of the chamber shape and dimension on soil CH<sub>4</sub> flux determination. Similarly, Rochette and Eriksen-Hamel (2008) also concluded that chamber shape and dimensions are critical factors in the estimation of GHG fluxes.

All previous studies agreed that the area/perimeter ratio is a key factor in soil gas flux estimation and, hence, they recommended a ratio greater than 0.10 m (Clough et al., 2020). In our work, the two types of chambers compared presented different area/perimeter ratios with values of 0.125 and 0.089 m for the automated and the manual chamber systems, respectively. This difference in the area/perimeter ratio could explain the greater CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O fluxes measured by the automated chamber system compared with the manual system. Moreover, the use of fans to mix the internal air of the automated chambers might have also explained the higher fluxes measured in this system compared with the manual system. Air-mixing by fans is recommended to homogenize the internal air of the chamber, ensuring that the air sample aliquot is representative of the chamber headspace air (Clough et al., 2020).

In line with the previous explanation, the Minimum Detectable Flux (MDF) following the equation presented by Nickersen (2016) was calculated for methodologies. The MDF method not only considered the accuracy of the analyser but also considered the area and volume of the chamber and the enclosure time, factors that are different between both methodologies compared in this work. The MDFs for the automated chamber system were 1.209 mg CO<sub>2</sub>-C m<sup>-2</sup> day<sup>-1</sup>, 0.012 mg CH<sub>4</sub>-C m<sup>-2</sup> day<sup>-1</sup> and 0.059 mg N<sub>2</sub>O-N m<sup>-2</sup> day<sup>-1</sup>, while for the manual chamber system, MDFs values were 14.050 mg CO<sub>2</sub>-C m<sup>-2</sup> day<sup>-1</sup>, 0.143 mg CH<sub>4</sub>-C m<sup>-2</sup> day<sup>-1</sup> and 0.071 mg N<sub>2</sub>O-N m<sup>-2</sup> day<sup>-1</sup>.

<sup>1</sup>. MDF was greater for the automated chamber system for the three gases, considering a similar enclosure time of 20 minutes and an average air temperature during the experiment of 20° C. The differences in MDF found between both methodologies was another factor that explained the greatest fluxes values observed under the automated chamber system.

### ***3.2. Sampling time and frequency comparison***

The effect of sampling time and frequency on cumulative soil gas emissions was compared between the automated and the manual measurement systems. This analysis was performed during one month in which the automated chamber system ran continuously over the entire month, while in the manual chamber system sampling was only performed on nine different dates.

As expected, the automated chamber system was able to capture daily flux fluctuations, a fact that was not possible for the manual chamber system, because only one gas sampling was done for each of the selected dates (Figure 5). However, when fluxes temporal dynamics for each gas were evaluated, it had been observed differences for each gas.



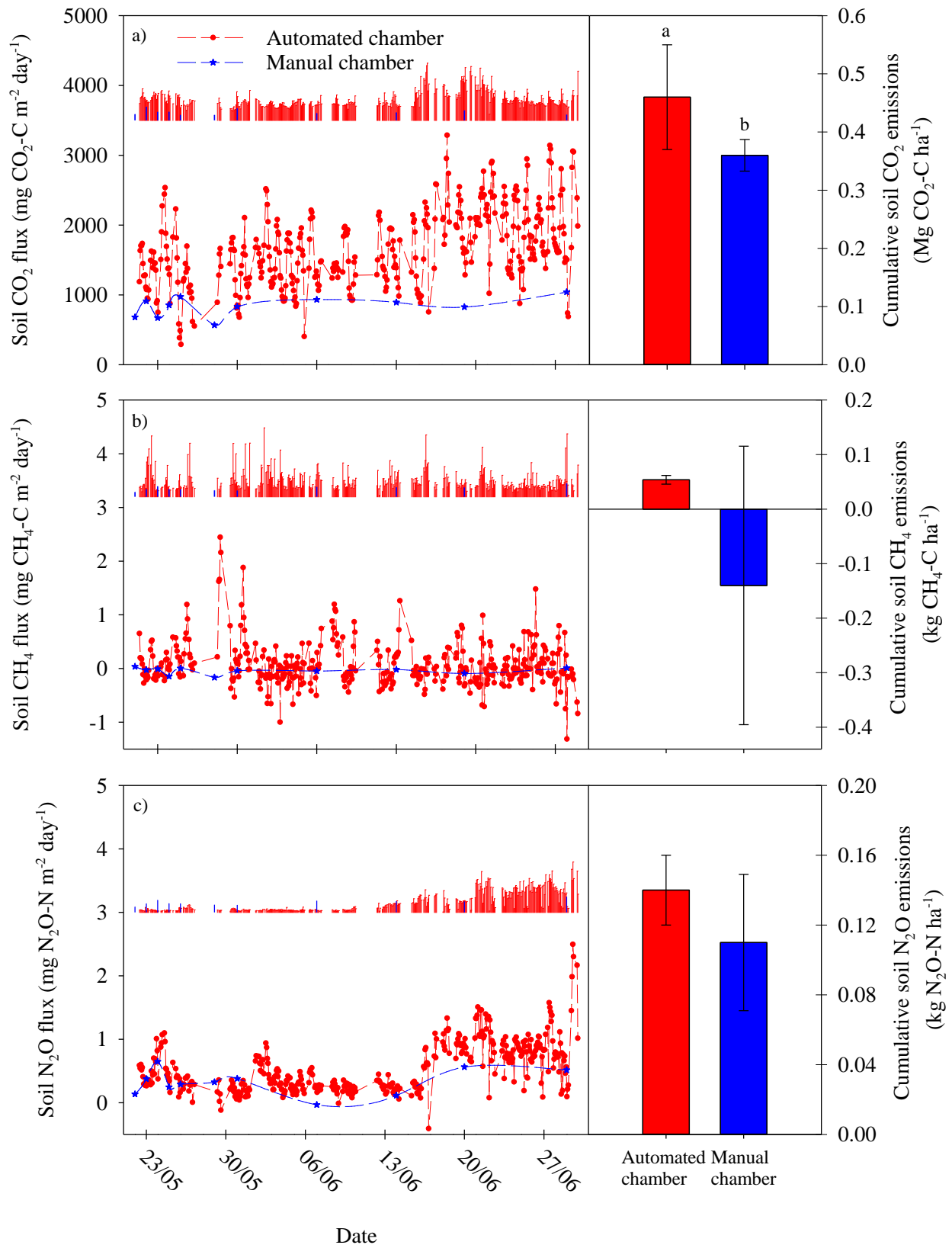


Figure 5. Comparison of soil gas flux and cumulative soil gas emissions between the automated (red line and bar) and the manual (blue line and bar) chamber system for carbon dioxide (CO<sub>2</sub>) fluxes (a), methane (CH<sub>4</sub>) fluxes (b) and nitrous oxide (N<sub>2</sub>O) fluxes (c). Vertical solid lines represent standard error SE (left panel) and cumulative soil emissions for the automated chamber systems (red bar) and manual chamber system (blue bar) (right panel).

Soil CO<sub>2</sub> and CH<sub>4</sub> fluxes determined by the manual chamber system showed similar behaviour, presenting a low variation in the fluxes magnitude over the evaluated period, being more pronounced for soil CH<sub>4</sub> fluxes (Figure 5a, 5b). For example, this was clearly observed in the CH<sub>4</sub> in which the automated system captured flux peaks greater than 2 mg CH<sub>4</sub>-C m<sup>-2</sup> day<sup>-1</sup> while the manual fluxes were close to 0 mg CH<sub>4</sub>-C m<sup>-2</sup> day<sup>-1</sup> over the entire measuring period (Figure 5b). Interestingly, the manual system was able to capture the temporal emission trend shown by the automated system for soil N<sub>2</sub>O fluxes, the gas that showed the greatest temporal variability over the period studied (Figure 5c).

The cumulative soil gas emissions of the three gases tended to be greater for the automated than the manual measurement system (Fig. 5). For example, cumulative soil CO<sub>2</sub> emissions presented significant differences between both sampling systems. The automated chamber system showed average values 16% more than the manual chamber system (Fig 5a). Indeed, this difference was even greater in CH<sub>4</sub> (more than 3-fold greater cumulative emissions in the automated than in the manual measurement system, Fig. 5b). Cumulative CH<sub>4</sub> emissions showed positive values for the automated chamber system while the average value for the manual chamber system was negative. However, the variability observed for the manual chamber system was 10 times greater rather than for the automated chamber system, a fact that resulted in the absence of significant differences between both sampling systems. Cumulative soil N<sub>2</sub>O emissions did not show significant differences between sampling systems despite that the average cumulative N<sub>2</sub>O emissions were 20% greater for the automated chamber system (Fig. 5c). As occurred with cumulative CH<sub>4</sub> emissions, the manual chamber system showed a greater variability than the automatic chamber system, reason that could explain the absence of significant differences between sampling systems.

Differences in the different cumulative emissions found between measurement systems might have been explained by the next three points: (i) construction differences, (ii) the sampling time in the manual system, and (iii) the height/enclosure time ratio (Clough et al., 2020). The automated chamber presented higher area/perimeter ratios and air-mixing by fans which could contribute to the greater fluxes found in this system compared with the manual system. Regarding the sampling time, this was especially critical for CO<sub>2</sub>. Manual sampling was performed at 08:00, resulting in an underestimation of the average daily emission (Pumpanen et al., 2003, Yu et al., 2013). In contrast, for N<sub>2</sub>O, underestimation was lower since 08:00 is considered a sampling time close to the optimal time for this gas (Wu et al., 2021). Finally, height/enclosure time ratio is also an important factor that affect the sensibility of the flux determination. As a recommendation, height/enclosure time ratio greater than 0.40 m hour<sup>-1</sup> is suggested to increase the minimum detectable flux and to reduce the impacts on air humidity, temperature and the gas diffusion process, variables that govern the soil gas fluxes between soil and atmosphere (Clough et al., 2020). In our study, the automated system resulted in height/enclosure ratios of 0.60 m hour<sup>-1</sup>, while in the manual system the ratios dropped to 0.30 m hour<sup>-1</sup>, explaining the lower cumulative emissions reported by the manual system.

Moreover, when the daily emission pattern of the automated chamber was, it was observed that soil CO<sub>2</sub> fluxes presented the maximum fluxes rate between 10:00 and 16:00, a daily pattern that respond to the changes in air and soil temperature, similar to the results reported by Pumpanen et al. (2003) and Yu et al. (2013). Differences between maximum and minimum daily fluxes (Figure 6a). Soil CH<sub>4</sub> fluxes did not show a clear daily pattern, excepted for some dates for (May 26<sup>th</sup> and 31<sup>st</sup>). Differences between the maximum and the minimum CH<sub>4</sub> fluxes were lower since soil CH<sub>4</sub> fluxes only ranged between -0.5 to 0.5 mg CH<sub>4</sub>-C m<sup>-2</sup> day<sup>-1</sup> for most of the nine selected

dates, expected for May 26<sup>th</sup> and 31<sup>st</sup> when soil CH<sub>4</sub> fluxes above 1 mg CH<sub>4</sub>-C m<sup>-2</sup> day<sup>-1</sup> were observed at midday (Figure 6b).

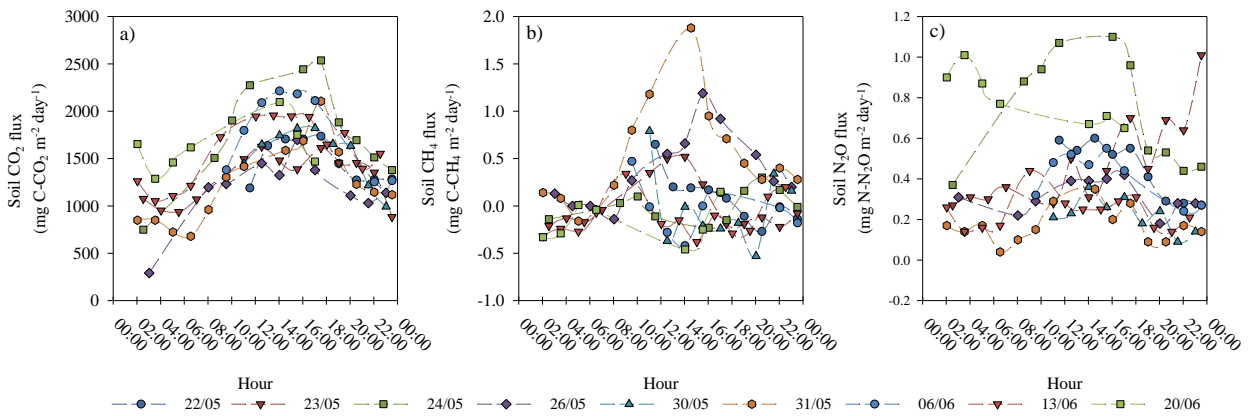


Figure 6. Daily soil gas flux determined by the chamber system for carbon dioxide (CO<sub>2</sub>), methane (CH<sub>4</sub>) fluxes (b) and nitrous oxide(N<sub>2</sub>O) fluxes (c). Vertical solid bars represent the two-hour intervals 08:00 to 10:00 (red), 10:00 to 12:00 (purple) and 12:00 to 14:00 (green), selected for evaluating the daily bias between sampling hours.

Soil N<sub>2</sub>O fluxes also presented a daily emission pattern characterized by reaching the maximum soil N<sub>2</sub>O from 10:00 to 18:00 and the minimum during nighttime but not being as clear as emission pattern observed for soil CO<sub>2</sub> fluxes (Figure 6c). This daily emission pattern was also observed by Wu et al. (2021) in a metanalysis which evaluated global daily N<sub>2</sub>O emission patterns. A possible explanation to the daily pattern observed in all three gases would be the temperature dependence of the biological process that governs the production and emission of soil GHG (Lloyd and Taylor, 1994, Smith and Dobbie, 2001, Davidson and Janssens, 2006.). This dependence would explain the higher emissions observed during daytime compared to nighttime (Fig. 6c).

To evaluate the effect of the selected sampling time for single day measurements could have over the soil fluxes estimation, three two intervals, 08:00-10:00, 10:00-12:00 and 12:00-14:00, were selected. For cumulative soil CO<sub>2</sub> emissions, the selected interval had a large impact on the estimation of the emissions, reporting cumulative emissions 23% greater for the 12:00-14:00 intervals against doing the gas sampling between 08:00 and 10:00 (Table 1). Cumulative

soil CH<sub>4</sub> emissions presented the highest values for the 08:00-10:00 time interval, however differences in absolute values were too low, indicating that sampling hour could have a lower impact on emissions estimation compared to CO<sub>2</sub> emissions for example (Table 1). Finally, time interval also has an impact on cumulative soil N<sub>2</sub>O emissions, observing the greatest emissions for 10:00-12:00 time interval. The large difference what found between 08:00-10:00 and 10:00-12:00 time interval, with a difference of 7.45% in cumulative emissions (Table 1).

Table 1. Cumulative soil emissions (carbon dioxide, CO<sub>2</sub>, methane, CH<sub>4</sub>, nitrous oxide, N<sub>2</sub>O) at three different time intervals (08:00-10:00, 10:00-12:00, 12:00-14:00).

Cumulative emissions	Time intervals			Bias between intervals (%)		
	08:00 - 10:00 (1)	10:00 - 12:00 (2)	12:00 - 14:00 (3)	1 vs 2	1 vs 3	2 vs 3
CO <sub>2</sub> (Mg CO <sub>2</sub> -C ha <sup>-1</sup> )	0.40	0.48	0.53	16.12	23.60	8.92
CH <sub>4</sub> (g CH <sub>4</sub> -C ha <sup>-1</sup> )	0.23	0.21	0.20	-8.81	-13.40	-4.22
N <sub>2</sub> O (g N <sub>2</sub> O-N ha <sup>-1</sup> )	0.91	0.99	0.95	7.45	3.79	-3.95

Based on the daily emissions pattern observed couple with the evaluation of the effect of different time interval, sampling time (i.e. hour) can have an impact on soil gases flux estimation, especially when only one sampling is done per day, as occurred with manual chamber systems. For soil CO<sub>2</sub> emissions, carried out the manual sampling at 08:00 supposed an underestimation of 43% respect to the mean daily flux estimated over 24 hours estimated with automated chamber system for the nine dates evaluated. Average soil CO<sub>2</sub> fluxes determined with the manual chamber system over the nine dates was 836 mg CO<sub>2</sub>-C m<sup>-2</sup> day<sup>-1</sup>, while the 24 hours CO<sub>2</sub> flux for the same nine date measuring with the automated chambers system was 1469 mg CO<sub>2</sub>-C m<sup>-2</sup> day<sup>-1</sup>. In contrast, sampling hour had a minimum impact on soil CH<sub>4</sub> fluxes, obtaining the similar average flux in both systems, 0.066 and 0.068 836 mg CH<sub>4</sub>-C m<sup>-2</sup> day<sup>-1</sup> for the manual and the automated chamber system, respectively.

Regarding N<sub>2</sub>O emissions, 08:00 resulted in an adequate sampling hour to obtain a representative daily emission. Average soil N<sub>2</sub>O flux of the nine manual sampling was 0.38 mg N<sub>2</sub>O-N m<sup>-2</sup> day<sup>-1</sup>, while the daily average for the same nine dates estimated with the automated chamber system was 0.41 mg N<sub>2</sub>O-N m<sup>-2</sup> day<sup>-1</sup>, resulting the fluxes determined with the manual chambers in an underestimation of 7% compared to the daily average soil N<sub>2</sub>O flux determined with the automated chambers. Perhaps, considering the evaluation of the three different time intervals, carrying out the soil gas sampling between 10:00 to 12:00 could be reduce this underestimation.

#### **4. Conclusion**

The presented system features an open design, cost-effective components, and adaptable configuration, offering benefits in flexibility, compatibility, and affordability, which in the end resulted in a more precise monitoring of the time flux variability. Moreover, it has been highlighted that the shape, dimension, and configuration of the chamber system are critical factors that must be considered in the design of the chambers, being critical in setting area/perimeter and height/enclosure time ratios greater than 0.10m and 0.40m h<sup>-1</sup>, respectively. Likewise, in case there is not option to implement an automated system, the sampling time of the manual measurement system is critical resulting in significant over or underestimation. Our results showed that 08:00 ~~06:00GMT~~ was an optimal sampling time for soil N<sub>2</sub>O emissions but resulted in an underestimation of soil CO<sub>2</sub> and CH<sub>4</sub> emissions, being more adequate performing the gas sampling between 10:00 to 12:00 to reduce emissions overestimation compared to the mean daily emissions, especially for CO<sub>2</sub>. Therefore, based on the results presented in this work, automated chamber systems are a powerful tool for quantifying GHG fluxes from the soil, allowing to capture the large temporal variability that characterizes GHG emissions. However, when the implementation of automated measurement system was not possible, sampling time for manual chamber must be

442 considered to avoid underestimation of soil fluxes, being the time interval between 10:00 to 12:00  
443 the most appropriate time interval.

## 444 **Data availability**

445 All raw data can be provided by the corresponding authors upon request.

446

## 447 **Code and data availability**

448 The R code and notebooks of the automatic chamber system and the data used to compare and  
449 validate the function of the automatic chamber system can be found at  
450 <https://zenodo.org/records/15189414>. (Franco-Luesma et al., 2025).

451

## 452 **Author contributions**

453 SF-L: Conceptualization, formal analysis, software, data acquisition, writing—original draft, data  
454 curation, methodology, formal analysis, investigation. MA-A: Conceptualization, methodology,  
455 writing—review; editing. BW: Writing—review. BL: Conceptualization ,methodology, software,  
456 writing—review, funding acquisition. JA-F: Conceptualization ,methodology, writing—review;  
457 editing, supervision, project administration, resources, funding acquisition

458

## 459 **Competing interests**

460 The authors declare that they have no conflict of interest.

461

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## References

- Álvaro-Fuentes, J., Arrúe, J. L., Cantero-Martínez, C., Isla, R., Plaza-Bonilla, D., & Quílez, D. (2016). Fertilization Scenarios in Sprinkler-Irrigated Corn under Mediterranean Conditions: Effects on Greenhouse Gas Emissions. *Soil Sci. Soc. Am. J.*, 80(3), 662-671. <https://doi.org/10.2136/sssaj2015.04.0156>
- Barton, L., Wolf, B., Rowlings, D., Scheer, C., Kiese, R., Grace, P., Stefanova, K., Butterbach-Bahl, K. Sampling frequency affects estimates of annual nitrous oxide fluxes. *Sci Rep* 2015, 5 (1), 15912. <https://doi.org/10.1038/srep15912>
- Clough, T. J., Rochette, P., Thomas, S. M., Pihlatie, M., Christiansen, J. R., & Thorman, R. E. (2020). Global Research Alliance N<sub>2</sub>O chamber methodology guidelines: Design considerations. *J. Environ. Qual.*, 49(5), 1081-1091. <https://doi.org/10.1002/jeq2.20117>
- Collier, S. M., Ruark, M. D., Oates, L. G., Jokela, W. E., & Dell, C. J. (2014). Measurement of greenhouse gas flux from agricultural soils using static chambers. *JoVE*, (90), e52110. <https://doi.org/10.3791/52110>
- Davidson, E. A., & Janssens, I. A. (2006). Temperature sensitivity of soil carbon decomposition and feedbacks to climate change. *Nature*, 440(7081), 165-173. <https://doi.org/10.1038/nature04514>
- de Klein, C. A. M., Harvey, M. J., Clough, T. J., Petersen, S. O., Chadwick, D. R., & Venterea, R. T. (2020). Global research alliance nitrous oxide chamber methodology guidelines: Introduction, with health and safety considerations. *J. Environ. Qual.*, 49(5), 1073-1080. <https://doi.org/10.1002/jeq2.20131>
- Ding, W., Cai, Z., & Tsuruta, H. (2004). Diel variation in methane emissions from the stands of *Carex lasiocarpa* and *Deyeuxia angustifolia* in a cool temperate freshwater marsh. *Atmos. Environ.*, 38(2), 181-188. <https://doi.org/10.1016/j.atmosenv.2003.09.066>

- Ferrara, R. M., Carozzi, M., Decuq, C., Loubet, B., Finco, A., Marzuoli, R., Gerosa, G., Di Tommasi, P., Magliulo, V. & Rana, G. (2021). Ammonia, nitrous oxide, carbon dioxide, and water vapor fluxes after green manuring of faba bean under Mediterranean climate. *Agric. Ecosyst. Environ.*, 315, 107439. <https://doi.org/10.1016/j.agee.2021.107439>
- Forte, A., Fiorentino, N., Fagnano, M., & Fierro, A. (2017). Mitigation impact of minimum tillage on CO<sub>2</sub> and N<sub>2</sub>O emissions from a Mediterranean maize cropped soil under low-water input management. *Soil Tillage Res.*, 166, 167-178. <https://doi.org/10.1016/j.still.2016.09.014>
- Francis Clar, J. T., & Anex, R. P. (2020). Flux intensity and diurnal variability of soil N<sub>2</sub>O emissions in a highly fertilized cropping system. *Soil Sci. Soc. Am. J.*, 84(6), 1983-1994. <https://doi.org/10.1002/saj2.20132>
- Franco-Luesma, S., Álvaro-Fuentes, J., Plaza-Bonilla, D., Arrúe, J. L., Cantero-Martínez, C., & Caverro, J. (2019). Influence of irrigation time and frequency on greenhouse gas emissions in a solid-set sprinkler-irrigated maize under Mediterranean conditions. *Agric. Water Manage.*, 221, 303-311. <https://doi.org/10.1016/j.agwat.2019.03.042>
- Franco-Luesma, S., Caverro, J., Plaza-Bonilla, D., Cantero-Martínez, C., Tortosa, G., Bedmar, E. J., & Álvaro-Fuentes, J. (2020). Irrigation and tillage effects on soil nitrous oxide emissions in maize monoculture. *Agron. J.*, 112(1), 56-71. <https://doi.org/10.1002/agj2.20057>
- Franco-Luesma, S., Caverro, J., Plaza-Bonilla, D., Cantero-Martínez, C., Arrúe, J. L., & Álvaro-Fuentes, J. (2020). Tillage and irrigation system effects on soil carbon dioxide (CO<sub>2</sub>) and methane (CH<sub>4</sub>) emissions in a maize monoculture under Mediterranean conditions. *Soil Tillage Res.*, 196, 104488. <https://doi.org/10.1016/j.still.2019.104488>
- Franco-Luesma, S., Lafuente, V., Alonso-Ayuso, M., Bielsa, A., Kouchami-Sardoo, I., Arrúe, J. L., & Álvaro-Fuentes, J. (2022). Maize diversification and nitrogen fertilization effects

517 on soil nitrous oxide emissions in irrigated Mediterranean conditions. *Front. Environ. Sci.*, 10,  
518 914851. <https://doi.org/10.3389/fenvs.2022.914851>

519 Franco-Luesma, S., Alonso-Ayuso, M., Wolf, B., Latorre, B., & Álvaro-Fuentes, J. (2025).  
520 AutoCam-GHG. In SOIL. Zenodo. <https://doi.org/10.5281/zenodo.15189414>

521 Grace, P. R., van der Weerden, T. J., Rowlings, D. W., Scheer, C., Brunk, C., Kiese, R.,  
522 Butterbach-Bahl, K., Rees, R. M., Robertson, G. P. & Skiba, U.M. (2020). Global Research  
523 Alliance N<sub>2</sub>O chamber methodology guidelines: Considerations for automated flux measurement.  
524 *J. Environ. Qual.*, 49(5), 1126-1140. <https://doi.org/10.1002/jeq2.20124>

525 Harvey, M. J., Sperlich, P., Clough, T. J., Kelliher, F. M., McGeough, K. L., Martin, R. J.,  
526 & Moss, R. (2020). Global Research Alliance N<sub>2</sub>O chamber methodology guidelines:  
527 Recommendations for air sample collection, storage and analysis. *J. Environ. Qual.*, 49(5), 1110-  
528 1125. <https://doi.org/10.1002/jeq2.20129>

529 Hoffmann, M., Pehle, N., Huth, V., Jurisch, N., Sommer, M. and Augustin, J. (2018). A  
530 simple method to assess the impact of sealing, headspace mixing and pressure vent on airtightness  
531 of manually closed chambers. *J. Plant Nutr. Soil Sci.*, 181, 36-40.  
532 <https://doi.org/10.1002/jpln.201600299>

533 Intergovernmental Panel on Climate Change, IPCC (2014). Climate Change 2014:  
534 Synthesis Report. Contribution of Working Groups I, II and III to the Fifth Assessment Report of  
535 the Intergovernmental Panel on Climate Change (IPCC).

536 Isla, R., Castillo, M. G., Medina, E. T., Latorre, B., de Viteri, D. Q. S., & Caverio, J. (2022).  
537 Greenhouse gas emissions associated to sprinkler-irrigated alfalfa under semi-arid Mediterranean  
538 conditions. *Span. J. Agric. Res.*, 20(3), 4. <https://doi.org/10.5424/sjar/2022203-18416>

539 Lawrence, N. C., & Hall, S. J. (2020). Capturing temporal heterogeneity in soil nitrous  
540 oxide fluxes with a robust and low-cost automated chamber apparatus. *Atmos. Meas. Tech.*, 13(7),  
541 4065-4078. <https://doi.org/10.5194/amt-13-4065-2020>

542 Lloyd, J. and Taylor J, A. (1994). On the temperature dependence of soil respiration.  
543 *Functional Ecology*, 8, 315–323. <https://doi.org/10.2307/2389824>

544 Lognoul, M., Theodorakopoulos, N., Hiel, M. P., Regaert, D., Broux, F., Heinesch, B., ...  
545 & Aubinet, M. (2017). Impact of tillage on greenhouse gas emissions by an agricultural crop and  
546 dynamics of N<sub>2</sub>O fluxes: Insights from automated closed chamber measurements. *Soil Tillage*  
547 *Res.*, 167, 80-89. <https://doi.org/10.1016/j.still.2016.11.008>

548 Nickerson, N. (2016). Evaluating gas emission measurements using Minimum Detectable  
549 Flux (MDF). Eosense Inc., Dartmouth, Nova Scotia, Canada.

550 Oertel, C., Matschullat, J., Zurba, K., Zimmermann, F., & Erasmi, S. (2016). Greenhouse  
551 gas emissions from soils—A review. *Chem Erde-Geochem.*, 76(3), 327-352.  
552 <https://doi.org/10.1016/j.chemer.2016.04.002>

553 Parkin, T. B. (2008). Effect of sampling frequency on estimates of cumulative nitrous oxide  
554 emissions. *J. Environ. Qual.*, 37(4), 1390–1395. <https://doi.org/10.2134/jeq2007.0333>

555 Pihlatie, M. K., J. R. Christiansen, H. Aaltonen, J. F. J. Korhonen, A. Nordbo, T. Rasilo,  
556 G. Benanti, M. Giebels, M. Helmy, J. Sheehy, S. Jones, R. Juszczak, R. Klefoth, R. Lobo-do-Vale,  
557 A. P. Rosa, P. Schreiber, D. Serça, S. Vicca, B. Wolf, and J. Pumpanen. (2013). Comparison of  
558 static chambers to measure CH<sub>4</sub> emissions from soils. *Agric. For. Meteorol.*, 171, 124-136.  
559 <https://doi.org/10.1016/j.agrformet.2012.11.008>

Pumpanen, J., Ilvesniemi, H., Perämäki, M., & Hari, P. (2003). Seasonal patterns of soil CO<sub>2</sub> efflux and soil air CO<sub>2</sub> concentration in a Scots pine forest: comparison of two chamber techniques. *Glob. Change Biol.*, 9(3), 371-382. <https://doi.org/10.1046/j.1365-2486.2003.00588.x>

R Core Team (2022). R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria. Version 4.2.2. <https://www.R-project.org/>

Reeves, S., Wang, W., Salter, B., & Halpin, N. (2016). Quantifying nitrous oxide emissions from sugarcane cropping systems: Optimum sampling time and frequency. *Atmos. Environ.*, 136, 123-133. <https://doi.org/10.1016/j.atmosenv.2016.04.008>

Rochette, P., & Eriksen-Hamel, N. S. (2008). Chamber Measurements of Soil Nitrous Oxide Flux: Are Absolute Values Reliable? *Soil Sci. Soc. Am. J.*, 72(2), 331. doi:10.2136/sssaj2007.0215

Savage, K., Phillips, R., & Davidson, E. (2014). High temporal frequency measurements of greenhouse gas emissions from soils. *Biogeosciences*, 11(10), 2709-2720. <https://doi.org/10.5194/bg-11-2709-2014>

Smith, K. A., & Dobbie, K. E. (2001). The impact of sampling frequency and sampling times on chamber-based measurements of N<sub>2</sub>O emissions from fertilized soils. *Glob. Change Biol.*, 7(8), 933-945. <https://doi.org/10.1046/j.1354-1013.2001.00450.x>

Yu, L., Wang, H., Wang, G., Song, W., Huang, Y., Li, S. G., Liang, N., Tang, Y. & He, J. S. (2013). A comparison of methane emission measurements using eddy covariance and manual and automated chamber-based techniques in Tibetan Plateau alpine wetland. *Environ. Pollut.*, 181, 81-90. <https://doi.org/10.1016/j.envpol.2013.06.018>

Wang, Z.-P., & Han, X.-G. (2005). Diurnal variation in methane emissions in relation to plants and environmental variables in the Inner Mongolia marshes. *Atmos. Environ.*, 39(34), 6295–6305. <https://doi.org/10.1016/j.atmosenv.2005.07.01>

Wiebe, K., Robinson, S., & Cattaneo, A. (2019). Climate change, agriculture and food security: impacts and the potential for adaptation and mitigation. *Sustainable Food and Agriculture*, 55-74. <https://doi.org/10.1016/B978-0-12-812134-4.00004-2>

Wu, Y. F., Whitaker, J., Toet, S., Bradley, A., Davies, C. A., & McNamara, N. P. (2021). Diurnal variability in soil nitrous oxide emissions is a widespread phenomenon. *Glob. Change Biol.*, 27(20), 4950-4966. <https://doi.org/10.1111/gcb.15791>