1 Supplement of

- 2 Dissolved organic matter fosters core
- 3 mercury-methylating microbiome for methylmercury
- 4 production in paddy soils
- 5 Qiang Pu et al.
- 6 Correspondence to: Bo Meng (mengbo@mail.gyig.ac.cn)

Supplementary Texts

- 8 Text S1. Measurement of soil physico-chemical properties, mercury and dissolved organic matter.
- 9 S1.1 Characterization of soil physico-chemical properties.
 - Soil pH was measured using a pH meter (PD-501, SANXIN, China) after extraction with ultrapure deionized water (soil: water = 1:2.5 w/v). Soil total carbon and total nitrogen were measured by an organic elemental analyzer (vario MACRO cube, Elementar, Germany). Water-soluble SO₄²⁻ was extracted with ultrapure deionized water (soil: water = 1:10 w/v) by a horizontal oscillator at 220 rpm for 16 h in the dark. The supernatant solution was obtained by centrifugation (2500 × g for 10 min) and filtration (0.45 μm, PES, Bizcomr, China). Water-soluble SO₄²⁻ was analyzed with a UV-Vis spectrophotometer (UV-5100B, METASH, China). Water-soluble NO₃⁻ was extracted with 2 M KCl, filtered using PES membranes (0.45 μm, Bizcomr, China), and measured by UV spectrophotometry (UV-1200, Macy Analysis Instrument Co. Ltd., China). The S²⁻ and Fe²⁺ in soil pore water were obtained by centrifugation (3500 × g for 15 min). The S²⁻ in soil pore water was measured by using methylene blue method (Cline, 1969), with a detection limit of 0.13 μM. The Fe²⁺ in soil pore water was measured by using ferrozine method (Viollier et al., 2000), with a detection of 10 μM.

22 S1.2 Analysis of mercury.

The water-soluble Hg concentration (representing Hg bioavailability) in paddy soils was determined according to Shi et al. with slight modifications (Shi et al., 2005). Briefly, ~0.5 g of soil was extracted in 8 mL of Milli-Q water with continuous agitation for 2 h. The suspension was centrifuged at 2850 g for 30 min and vacuum filtered through 0.45-µm mixed cellulose acetate filters (Whatman, USA). The amount of water-soluble Hg in solution was measured by cold vapor atomic fluorescence spectrometry (CVAFS, Brooks Rand Model III, Brooks Rand Laboratories) according to USEPA method 1631 (EPA, 2002). To determine total Hg (THg), ~0.2 g of soil was digested in 5 mL of freshly prepared aqua regia (HCl:HNO3 = 3:1 v/v) with 5 mL of Milli-Q water at 95 °C for 55 mins. The total Hg amount in the digest solution was measured by cold vapor atomic fluorescence spectrometry (CVAFS, Brooks Rand Model III, Brooks Rand Laboratories) according to USEPA method 1631 (EPA, 2002). Approximately 0.3-0.4 g of soil was extracted using CuSO4-methanol solvent for MeHg quantification via gas chromatography CVAFS

34 (GC-CVAFS, Brooks Rand Model III, Brooks Rand Laboratories) following the procedure of USEPA
35 method 1630 (EPA, 2001). To ensure the accuracy of THg and MeHg quantification in soils, method
36 blanks and standard reference materials, GSS-5 (THg: 290 ± 30 ng g⁻¹) and ERMCC580 (MeHg: 75.5 ±
37 3.7 ng g⁻¹), were analyzed. The THg and MeHg recoveries from GSS-5 and ERMCC580 were 113.0% ±
38 7.1% (n=6) and 103.6% ± 3.5% (n = 3), respectively. The relative standard deviation (RSD%) for THg

39 and MeHg analysis in triplicate was less than 5.2% and 3.9%, respectively.

S1.3 Analysis of dissolved organic matter concentration and composition.

The concentration of soil dissolved organic matter (DOM, reflected by water-soluble DOC) was extracted with Milli-Q water (soil: water = 1:10 w/v), filtered (0.45 µm Polypropylene Membrane Filters) and determined using a total organic carbon analyzer (Vario TOC cube, Elementar, Germany). The dissolved organic matter composition (reflected by optical properties of DOM) was characterized with UV-Vis absorption through Aqualog® absorption-fluorescence spectroscopy (Jobin Yvon, Horiba, Japan). UV-Vis absorption spectra for liquid samples were scanned from 230 nm to 800 nm (1 nm interval) (Liu et al., 2022). Internal filtering effects were minimized via pre-measurement dilution to a DOC < 10 mg/L (Jiang et al., 2018). S_R (spectral slope ratio over the ranges of 275-295 nm and 350-400 nm) of DOM was calculated for the optical property of DOM, and the detailed calculation and description can be found in previous works (Jiang et al., 2018; Zhang et al., 2023).

S1.4 Analysis of low-molecular-weight organic acids.

Soils from NMS, MMS and HMS were selected for analysis of low-molecular-weight organic acids. Soil (~10 g) was extracted with 20 mL milli-Q water for 12 h. The mixture was centrifuged at 15 000 × g for 15 min, and filtered through Whatman No. 42. Water-soluble low-molecular-weight organic acids were obtained by evaporating the solvent to dryness in a rotary evaporator at 40 °C and redissolving the residue in 1 mL of Mill-Q water. The water-soluble low-molecular-weight organic acids were identified and quantified by using reversed-phase high-performance liquid chromatography (HPLC, Shimadzu LC-20, Shimadzu, Osaka, Japan) with a diode array detector (van Hees et al., 1999).

Text S2. hgcA gene quantification.

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60 The abundance of the hgcA gene was quantified with primer set ORNL-Delta-HgcA-F: 61 GCCAACTACAAGMTGASCTWC and ORNL-Delta-HgcA-R: CCSGCNGCRCACCAGACRTT in 62 AB 7500 (Applied Biosystems, USA). The PCR setup was as follows: 10 µl of SYBR Premix Ex Taq 63 (TaKaRa Bio Inc., Japan), 0.5 μL (10 mM) of each primer, 2 μL of diluted DNA template (~20 ng), and 7 64 μL of sterilized DDW (double distilled water). The hgcA gene was quantified in triplicate under the 65 following thermal cycles: 3 min initial denaturation at 95°C, 40 cycles of 15 s at 95°C, 15 s at 50°C, and 66 15 s at 55°C, and 4 min at 72°C, followed by a plate read at 83 °C (Liu et al., 2018). Three no-template 67 controls were used to detect contamination during the amplification process.

Text S3. Bacterial culture and validation experiment.

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Geobacter sulfurreducens PCA was cultured in nutrient broth (NB) at 33°C (Hu et al., 2013). Cells were harvested at the middle log phase and washed three times with phosphate buffered saline (PBS, containing 0.137 M sodium chloride, 0.0027 M potassium chloride, 0.01 M sodium phosphate dibasic, and 0.0018 M potassium phosphate monobasic, pH 7.4) media before the validation experiment. The validation experiment was conducted in PBS medium amended with 5 mL of natural DOM solution extracted from paddy soils (i.e., NMS, MMS and HMS), 5 nM Hg(II), and a cell density of 2×108 cells mL⁻¹. The natural DOM extraction method referred to the procedure described in Text S1.3. All vials were immediately sealed with caps and kept in the dark on shaker. After incubation for 24 h at 33°C, triplicate sample vials were remove from the shaker and preserved at 4°C. An aliquot (10 mL) was filtered through 0.45 µm polyethersulfone (PES) membranes and analyzed for DOM concentration via total organic carbon analyzer (Vario TOC cube, Elementar, Germany). Another aliquot of the sample (3 mL) was acidified with trace metal grade HCl (0.2% (v/v)) and acetic acid (0.5% (v/v)), and analyzed for MeHg. The remaining aliquot was oxidized overnight in BrCl (1% (v/v)) and analyzed for total Hg. Total Hg and MeHg were measured by CVAFS (Brooks Rand Model III, Brooks Rand Laboratories) and cold vapor atomic fluorescence spectrometry (CVAFS, Brooks Rand Model III, Brooks Rand Laboratories), respectively (EPA, 2001; 2002). Treatment with HgCl2 and natural DOM solution was used to monitor abiotic Hg methylation. The

Treatment with HgCl₂ and natural DOM solution was used to monitor abiotic Hg methylation. The experimental procedures were consistent with the aforementioned experiment.

87 Supplementary Figures

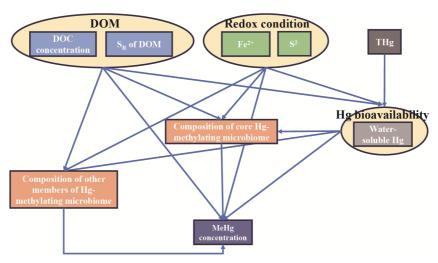


Fig. S1. A *priori* models for the structure equation models of variation in MeHg production based on the hypothesized causal relationships between multiple factors and MeHg production.

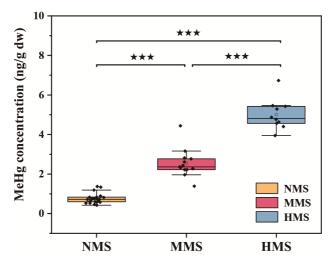


Fig. S2. Soil MeHg concentration in paddy soils. NMS, non-Hg polluted paddy soils (n = 23); MMS, moderate Hg-polluted paddy soils (n = 13); HMS, high Hg-polluted paddy soils (n = 10). " $\star\star$ " represents significant difference between different paddy soils (p < 0.001).

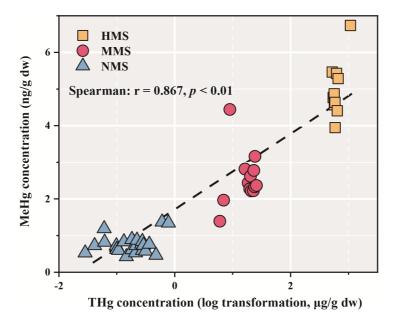


Fig. S3. Correlation between THg and MeHg concentration in paddy soils. NMS, non-Hg polluted paddy soils (n = 23); MMS, moderate Hg-polluted paddy soils (n = 13); HMS, high Hg-polluted paddy soils (n = 10).

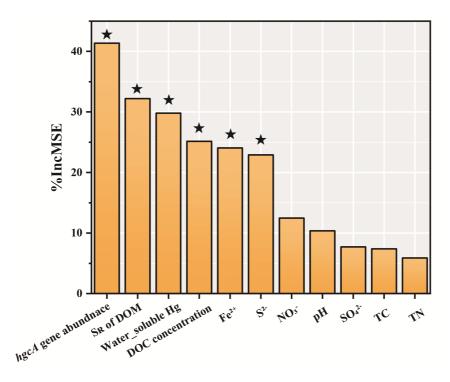


Fig. S4. Random forest modeling indicating the importance of different predictors for MeHg production. The number of trees used in model is 5000. " \star " represents a significant difference (p < 0.05).

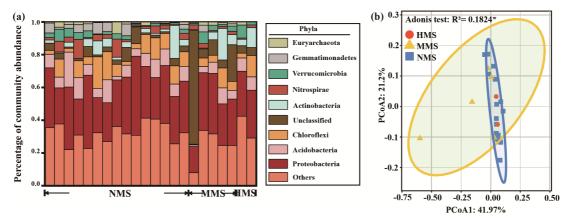


Fig. S5. (a) Taxonomic profiles of Hg-methylating microbial communities in different paddy soils. Phyla with low abundance phyla grouped together under "other phyla". **(b)** Principal coordinates analysis (PCoA) based on Bray-curtis distance showing the overall pattern of Hg-methylating microbial communities in paddy soils. NMS, non-Hg polluted paddy soils (n = 15); MMS, moderate Hg-polluted paddy soils (n = 5); HMS, high Hg-polluted paddy soils (n = 2).

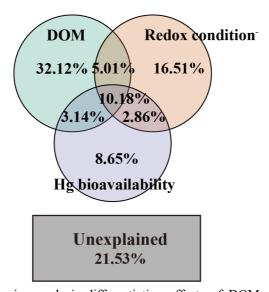


Fig. S6. Variation partitioning analysis differentiating effects of DOM, redox conditions, and Hg bioavailability on core Hg-methylating microbiome composition. DOM is reflected by DOM concentration and composition, which are measured as water-soluble DOC concentration and S_R (spectral slope ratio of $S_{275-295}:S_{350-400}$) values of DOM. Redox conditions are reflected by soil Fe^{2+} and S^{2-} , which are measured as concentrations of Fe^{2+} and S^{2-} in soil pore water. Hg bioavailability is reflected by water-soluble Hg. It should be noted that Fe^{2+} and S^{2-} data were limited to the soil samples obtained in August 2022.

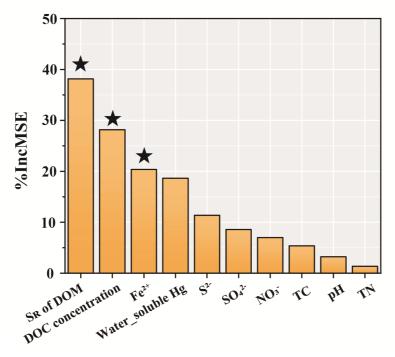


Fig. S7. Random forest modeling indicating the importance of different predictors for core Hg-methylating microbiome composition. The number of trees used in model is 5000. " \star " represents a significant difference (p < 0.05).

121 Supplementary Tables

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Table S1. Detailed information for paddy soils collected from 12 provinces across China.

Sample	Province	Site	Longitude	Latitude	Total Hg (μg/g)	Sampling time	Category
S1	Guizhou	HX-1	106°31′34″	26°25′15″	0.22	Sep. 2020	NMS
S2	Guizhou	HX-2	106°31′20″	26°25′18″	0.27	Sep. 2020	NMS
S3	Guizhou	HX-3	106°31′21″	26°25′15″	0.29	Sep. 2020	NMS
S4	Guizhou	HX-4	106°31′28″	26°25′16″	0.31	Sep. 2020	NMS
S5	Guizhou	HX-5	106°31′19"	26°25′17"	0.28	Sep. 2020	NMS
S6	Guizhou	HX-6	106°31′26″	26°25′21″	0.28	Sep. 2020	NMS
S7	Guizhou	HX-7	106°31′31″	26°25′20″	0.36	Sep. 2020	NMS
S8	Guizhou	HX-8	106°31′28″	26°25′14"	0.18	Sep. 2020	NMS
S9	Guizhou	HX-9	106°31′15″	26°25′19″	0.21	Sep. 2020	NMS
S10	Guizhou	GX-1	109°09′25″	27°33′23″	19.74	Sep. 2020	MMS
S11	Guizhou	GX-2	109°11′22″	27°33′37″	24.3	Sep. 2020	MMS
S12	Guizhou	GX-3	109°10′09″	27°33′36″	20.24	Sep. 2020	MMS
S13	Guizhou	GX-4	109°12′42″	27°33′53″	23.94	Sep. 2020	MMS
S14	Guizhou	GX-5	109°11′02″	27°33′28″	22.79	Sep. 2020	MMS
S15	Guizhou	GX-6	109°13′38″	27°33′56″	25.67	Sep. 2020	MMS
S16	Guizhou	GX-7	109°10′35″	27°33′33″	23.25	Sep. 2020	MMS
S17	Guizhou	GX-8	109°09′55″	27°33′43″	20.86	Sep. 2020	MMS
S18	Guizhou	GX-9	109°09′12″	27°33′31″	18.56	Sep. 2020	MMS
S19	Guizhou	SK-1	109°12′34″	27°30′41″	639.87	Sep. 2020	HMS
S20	Guizhou	SK-2	109°12′48″	27°31′05″	586.56	Sep. 2020	HMS
S21	Guizhou	SK-3	109°12′24″	27°30′36″	524.16	Sep. 2020	HMS
S22	Guizhou	SK-4	109°12′27″	27°30′25″	543.04	Sep. 2020	HMS
S23	Guizhou	SK-5	109°12′35″	27°30′39″	583.72	Sep. 2020	HMS
S24	Guizhou	SK-6	109°12′18″	27°30′50″	567.14	Sep. 2020	HMS
S25	Guizhou	SK-7	109°12′30″	27°30′52″	621.2	Sep. 2020	HMS
S26	Guizhou	SK-8	109°12′38″	27°30′55″	570.8	Sep. 2020	HMS
S27	Guizhou	SK-9	109°12′49″	27°31′02″	661.62	Sep. 2020	HMS
S28	Jilin	5N-3	125°57′0″	43°42′54″	0.03	Aut. 2022	NMS
S29	Jilin	5N-8	125°44′7″	44°6′26″	0.06	Aut. 2022	NMS
S30	Liaoning	5M-5	123°6′45″	41°49′46″	0.04	Aut. 2022	NMS
S31	Hubei	3I-1	113°10′48″	29°31′48″	0.06	Aut. 2022	NMS
S32	Guangxi	1A-28	108°45′58″	21°48′23″	0.1	Aut. 2022	NMS
S33	Hunan	3I-11	112°24′0″	28°34′12″	0.1	Aut. 2022	NMS
S34	Guangdong	1B-14	110°43′25″	21°51′23″	0.11	Aut. 2022	NMS
S35	Guangxi	1A-22	110°14′59″	22°21′20″	0.13	Aut. 2022	NMS
S36	Sichuan	3K-3	104°20′24″	30°49′12″	0.15	Aut. 2022	NMS
S37	Guizhou	4G-15	106°20′28″	26°25′59″	0.18	Aut. 2022	NMS
S38	Jiangsu	2E-11	119°9′29″	33°28′45″	0.21	Aut. 2022	NMS
S39	Hunan	3I-18	109°38′24″	28°37′12″	0.48	Aut. 2022	NMS
S40	Guangxi	1A-1	109°45′33″	23°11′26″	0.61	Aut. 2022	NMS
S41	Zhejiang	2D-5	119°41′31″	30°19′6″	0.78	Aut. 2022	NMS
S42	Liaoning	5M-1	123°7′9″	41°19′15″	6.01	Aut. 2022	MMS
S43	Shaannxi	6Q-23	109°27′55″	33°5′1″	7	Aut. 2022	MMS
S44	Guizhou	4G-8	109°24′38″	27°39′59″	8.95	Aut. 2022	MMS
S45	Guizhou	4G-7	109°15′13″	27°31′58″	16.34	Aut. 2022	MMS
S46	Chongqin	3J-1	108°55′12″	28°37′48″	1079.75	Aut. 2022	HMS

Paddy soils were divided into three categories according to mercury concentration: NMS, non-Hg polluted soils;

MMS, moderate Hg-polluted soils; HMS, high Hg-polluted soils. Aut., August; Sep., September.

Table S2. Characterization of dissolved organic matter (DOM) in paddy soils.

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			DOM	DOM composition ^a				
Sample	Category	Site	concentration (g kg ⁻¹)	SUVA ₂₅₄	S_R	BIX	FI	HIX
S1	NMS	HX-1	0.36	1.21	1.9	0.63	1.48	0.8
S2	NMS	HX-2	0.36	0.65	0.54	0.69	1.57	0.7
S3	NMS	HX-3	0.37	0.85	1	0.66	1.57	0.73
S4	NMS	HX-4	0.43	1.27	1.28	0.6	1.48	0.78
S5	NMS	HX-5	0.38	0.6	1.2	0.7	1.57	0.71
S6	NMS	HX-6	0.38	1.5	0.99	0.61	1.47	0.83
S7	NMS	HX-7	0.43	0.81	1.2	0.64	1.53	0.76
S8	NMS	HX-8	0.36	0.6	0.51	0.7	1.57	0.71
S9	NMS	HX-9	0.34	1.32	1.7	0.66	1.53	0.76
S10	MMS	GX-1	0.36	1.12	0.77	0.51	1.55	0.99
S11	MMS	GX-2	0.36	0.97	0.82	0.52	1.54	0.98
S12	MMS	GX-3	0.36	0.09	0.79	0.5	1.37	0.88
S13	MMS	GX-4	0.39	1.14	0.93	0.52	1.4	0.98
S14	MMS	GX-5	0.38	1.03	0.92	0.5	1.42	1
S15	MMS	GX-6	0.38	1.01	0.97	0.6	1.59	0.93
S16	MMS	GX-7	0.37	0.63	0.84	0.58	1.5	0.85
S17	MMS	GX-8	0.37	1.08	0.82	0.54	1.42	0.87
S18	MMS	GX-9	0.38	1.14	0.89	0.58	1.5	0.77
S19	HMS	SK-1	0.21	1.04	0.42	0.51	1.17	0.75
S20	HMS	SK-2	0.2	1.08	0.39	0.52	1.19	0.72
S21	HMS	SK-3	0.32	1.32	0.57	0.44	1.08	0.77
S22	HMS	SK-4	0.19	1.53	0.39	0.62	1.24	0.56
S23	HMS	SK-5	0.33	1.15	0.6	0.45	1.2	0.78
S24	HMS	SK-6	0.31	1.48	0.35	0.45	1.12	0.76
S25	HMS	SK-7	0.28	1.59	0.56	0.43	1.06	0.77
S26	HMS	SK-8	0.32	1.16	0.35	0.53	1.2	0.74
S27	HMS	SK-9	0.27	1.15	0.33	0.35	1.2	0.78
S28	NMS	5N-3	0.61	0.16	1.2	0.76	1.63	0.78
S29	NMS	5N-8	0.56	0.10	4.1	0.76	1.57	1
S30	NMS	5M-5	0.54	0.33	1.21	0.70	1.69	0.91
S31	NMS	3I-1	0.5	0.33	1.13	0.63	1.63	0.84
S31	NMS	1A-28	0.64	1.11	0.86	0.63	1.03	0.84
S32	NMS	3I-11	0.51	0.23	1.21	0.57	1.47	0.93
S34	NMS	1B-14	0.55	0.23	1.72	0.63	1.04	0.9
S35	NMS	1B-14 1A-22	0.51	0.18	2.87	0.37	1.77	0.80
S35 S36	NMS	3K-3	0.58	0.22	1.34	0.83	1.50	0.79
S30 S37	NMS		0.38	0.74	1.34	0.54	1.56	0.89
S37 S38	NMS NMS	4G-15 2E-11	0.5	0.38	1.47	0.56	1.36	0.93
S39	NMS	3I-18	0.45	0.29	0.91	0.55	1.64	0.9
S40	NMS	1A-1	0.89	0.17	1.13	0.73	1.75	0.87
S41	NMS	2D-5	0.54	1.53	1.32	0.73	1.97	0.34
S42	MMS	5M-1	0.39	0.32	0.84	0.58	1.44	0.98
S43	MMS	6Q-23	0.42	0.39	1.06	0.54	1.57	0.92
S44	MMS	4G-8	0.4	0.41	0.99	0.68	1.64	0.79
S45	MMS	4G-7	0.61	0.64	0.93	0.76	1.63	0.95
S46	HMS	3J-1	0.53	1.75	0.52	0.55	1.14	0.62

Paddy soils were divided into three categories according to mercury concentration: NMS, non-Hg polluted soils;

biological index (BIX), humification index (HIX) and fluorescence index (FI) are the fluorescence compounds and calculated indices from EEM fluorescence spectra of DOC.

MMS, moderate Hg-polluted soils; HMS, high Hg-polluted soils.

¹²⁸ a SUVA₂₅₄ (specific UV absorbance at a wavelength of 254 nm) and S_R (spectral slope ratio of S₂₇₅₋₂₉₅: S₃₅₀₋₄₀₀) are properties from UV-Vis absorption spectra of DOM.

Table S3. Physicochemical properties in paddy soils

				SO ₄ ²⁻	NO_3^-	TN	TC	S ²⁻	Fe ²⁺
Sample	Category	Site	pН	(mg kg ⁻¹)	(mg kg ⁻¹)	(%)	(%)	(μ M)	(μ M)
S1	NMS	HX-1	7.52	347.87	17.67	0.4	5.49	No data	No data
S2	NMS	HX-2	7.53	369.9	19.52	0.28	3.15	No data	No data
S3	NMS	HX-3	7.51	354.61	18.38	0.41	5.47	No data	No data
S4	NMS	HX-4	7.5	369.68	20.26	0.41	5.5	No data	No data
S5	NMS	HX-5	7.52	356.8	18.73	0.21	7.82	No data	No data
S6	NMS	HX-6	7.54	351.22	18.24	0.29	3.13	No data	No data
S7	NMS	HX-7	7.51	358.16	18.76	0.21	7.79	No data	No data
S8	NMS	HX-8	7.53	343.73	18.26	0.29	3.15	No data	No data
S9	NMS	HX-9	7.51	348.69	18.21	0.21	7.77	No data	No data
S10	MMS	GX-1	7.52	348.01	17.62	0.42	5.72	No data	No data
S11	MMS	GX-2	7.52	349.67	17.55	0.21	7.99	No data	No data
S12	MMS	GX-3	7.5	337.66	17.57	0.42	5.74	No data	No data
S13	MMS	GX-4	7.5	371.17	17.78	0.28	3.21	No data	No data
S14	MMS	GX-5	7.49	335.06	17.63	0.43	5.71	No data	No data
S15	MMS	GX-6	7.5	381.51	17.79	0.27	3.19	No data	No data
S16	MMS	GX-7	7.5	333.01	17.6	0.47	6.08	No data	No data
S17	MMS	GX-8	7.51	377.65	17.55	0.21	7.88	No data	No data
S18	MMS	GX-9	7.5	339.97	17.5	0.21	7.93	No data	No data
S19	HMS	SK-1	7.51	254.52	15.65	0.48	6.12	No data	No data
S20	HMS	SK-2	7.48	255.55	14.58	0.21	7.84	No data	No data
S21	HMS	SK-3	7.47	266.23	16.1	0.46	6.1	No data	No data
S22	HMS	SK-4	7.45	271.33	15.52	0.32	4.71	No data	No data
S23	HMS	SK-5	7.51	251.63	15.56	0.28	7.92	No data	No data
S24	HMS	SK-6	7.45	256.33	14.92	0.29	3.14	No data	No data
S25	HMS	SK-7	7.48	246.24	16.06	0.46	6.07	No data	No data
S26	HMS	SK-8	7.45	241.53	14.5	0.22	7.89	No data	No data
S27	HMS	SK-9	7.47	219.44	16.4	0.21	3.18	No data	No data
S28	NMS	5N-3	7.36	393.09	15.91	0.25	2.71	0.87	16752.81
S29	NMS	5N-8	7.68	291.38	18.36	0.18	6.73	0.28	20058.81
S30	NMS	5M-5	6.84	421.42	7.98	0.18	6.7	0.38	34030.78
S31	NMS	3I-1	7.11	296.93	17.07	0.24	2.76	0.92	8490.57
S32	NMS	1A-28	6.95	278.44	17.39	0.36	4.94	1.31	19670.1
S33	NMS	3I-11	6.84	293.2	18.57	0.24	2.71	2.21	3476.21
S34	NMS	1B-14	6.88	286.05	17.68	0.18	6.82	1.38	10140.85
S35	NMS	1A-22	7.48	282.71	17.03	0.18	6.87	1.13	3147.96
S36	NMS	3K-3	7.01	206.27	14.03	0.41	5.26	2.06	1869.54
S37	NMS	4G-15	7.06	268.33	16.82	0.18	6.68	0.87	77.94
S38	NMS	2E-11	6.54	290.39	16.03	0.23	2.74	0.72	1567.21
S39	NMS	3I-18	7.01	293.8	18.55	0.23	4.91	0.46	3912.43
S40	NMS	1A-1	7.12	273.16	15.44	0.18	6.78	0.55	44307.05
S41	NMS	2D-5	7.03	185.56	18.12	0.16	5.23	0.48	64217.64
S42	MMS	5M-1	6.88	250.39	17.8	0.36	4.92	0.46	2849.95
S43	MMS	6Q-23	7.05	394.55	8.76	0.25	2.69	0.55	3031.35
S44	MMS	4G-8	7.52	292.47	18.19	0.25	4.7	0.17	7086.89
S45	MMS	4G-7	5.57	289.84	18.25	0.34	4.72	0.17	17456.81
S46	HMS	3J-1	6.18	436.82	18.86	0.35	4.73	0.75	4433.16
510	111/10	JJ-1	0.10	150.02	10.00	0.55	1113	0.13	1133.10

Paddy soils were divided into three categories according to mercury concentration: NMS, non-Hg polluted soils;

MMS, moderate Hg-polluted soils; HMS, high Hg-polluted soils. No data indicate that the concentrations of S²⁻ and

Fe²⁺ were unavailable in soil samples from September 2020 (S1-S26).

Table S4. Key characteristics of co-occurrence networks in paddy soils.

Category	Connected nodes	Edges	Module	Average degree	Network diameter	Modularity index
NMS	199	3062	6	15.31	9	0.554
MMS	193	1655	11	8.275	7	0.583
HMS	189	1714	11	8.57	7	0.591

NMS, non-Hg polluted paddy soils (n = 23); MMS, moderate Hg-polluted paddy soils (n = 13); HMS, high Hg-polluted paddy soils (n = 10).

Table S5. Identification of major module (also known as core microbiome) in different paddy soils.

ID	Module	The number of connections to other modules	Relative abundance (%)	Correlation with MeHg concentration	Correlation with %MeHg
NMS	module1	146	34	0.418*	0.787***
	module2	13	23.5	0.164	-0.374
	module3	28	19	0.206	-0.31
	module4	3	16.5	0.189	-0.256
	module5	17	6.5	0.116	-0.186
	module6	0	0.5	0.047	-0.05
MMS	module1	66	27.5	0.503*	0.863**
	module2	84	20.5	0.035	-0.165
	module3	8	17.5	-0.103	0.066
	module4	59	9	0.068	0.051
	module5	5	8	0.206	-0.204
	module6	5	5.5	-0.049	0.106
	module7	5	3	-0.085	0.04
	module8	3	2.5	0.039	-0.065
	module9	0	2	0.101	-0.169
	module10	1	1	-0.008	0.016
	module11	0	3.5	0.026	-0.064
HMS	module1	161	20	0.410*	0.872**
	module2	120	18.5	-0.351	0.531
	module3	0	16.5	0.318	-0.518
	module4	25	9.5	-0.113	0.219
	module5	43	9	-0.259	0.654*
	module6	29	7.5	-0.302	0.518
	module7	0	5	-0.101	0.215
	module8	0	3.5	0.002	-0.008
	module9	0	1.5	-0.029	0.126
	module10	0	1.5	-0.106	0.204
	module11	0	7.5	0.284	-0.579

NMS, non-Hg polluted paddy soils (n = 23); MMS, moderate Hg-polluted paddy soils (n = 13); HMS, high Hg-polluted paddy soils (n = 10).

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