

1 **Dissolved organic matter fosters core mercury-methylating**  
2 **microbiome for methylmercury production in paddy soils**

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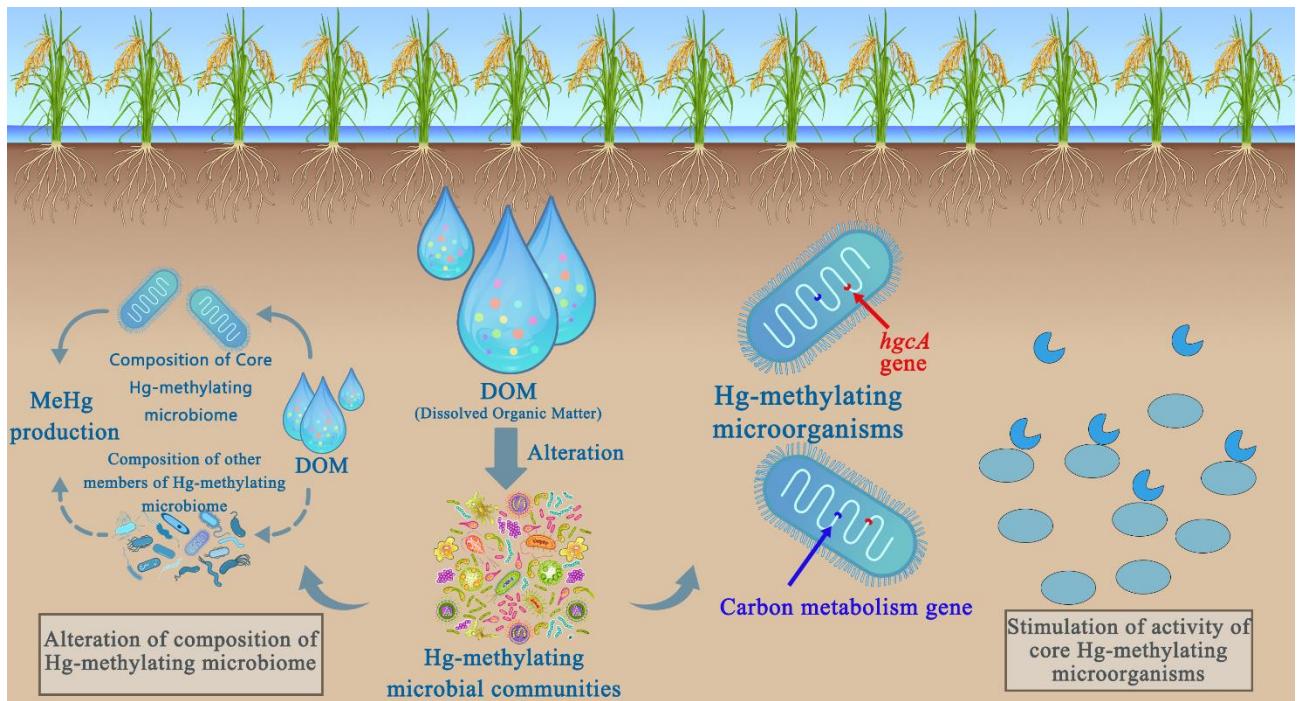
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13

14 **Abstract.** Methylmercury (MeHg), accumulated in rice grain, is highly toxic for human. Its production is largely driven  
15 by microbial methylation in paddy soils; however, dissolved organic matter (DOM) is a critical component for soil  
16 biogeochemistry process, yet its interactions with microorganisms involved in MeHg production, remains poorly  
17 understood. Here, we conducted *hgcA* gene sequencing and genome-resolved metagenomic analysis to identify core Hg-  
18 methylating microbiome and investigate the effect of DOM on core Hg-methylating microbiome in paddy soils across a  
19 Hg contamination gradient. In general, the Hg-methylating microbial communities varied largely with the degree of Hg  
20 contamination in soils. Surprisingly, a core Hg-methylating microbiome was identified that was exclusively associated  
21 with MeHg concentration. The partial Mantel test revealed strong linkages among core Hg-methylating microbiome  
22 composition, DOM and MeHg concentration. Structural equation model further indicated that core Hg-methylating  
23 microbiome composition significantly impacted soil MeHg concentration, contributing to 89% of the observed variation;  
24 while DOM play a crucial in determining core Hg-methylating microbiome composition, accounting for 65%. These  
25 results suggested that DOM regulates MeHg production by altering the composition of core Hg-methylating microbiome.  
26 The presence of various genes associated with carbon metabolism in the metagenome-assembled genome of core Hg-  
27 methylating microorganisms suggests that different DOMs stimulate the activity of core Hg-methylating microorganisms  
28 to methylate Hg, which was confirmed by pure incubation experiment with *Geobacter sulfurreducens* PCA (a core Hg-  
29 methylating microorganism) amended with natural DOM solution extracted from investigated soils. Overall, DOM  
30 simultaneously changes core Hg-methylating microbiome composition and functional activity and thus enhances MeHg  
31 production in paddy soils.

32 **Keywords.** Rice paddy; Mercury methylator; Methylmercury formation; Core microbiome

33 Graphical abstract



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35

36 **1 Introduction**

37 Mercury (Hg) is a toxic contaminant since it can be transformed into neurotoxic methylmercury (MeHg) and biomagnified  
38 in food chains (Driscoll et al., 2013). Human exposure to MeHg can cause neurocognitive deficits and cardiovascular  
39 effects (Oulhote et al., 2017; Roman et al., 2011). It is generally accepted that seafood consumption is the major route of  
40 exposure to MeHg in humans (Schartup et al., 2019). However, recent studies have demonstrated that rice consumption  
41 is another important route of human exposure to MeHg (Feng et al., 2008), as 3.5 billion individuals relying on rice as  
42 principal dietary component (Muthayya et al., 2014).

43 Compared to other environments such as wetlands and aquatic sediments, paddy fields present unique ecological  
44 conditions that make them significant hotspots for Hg methylation. The frequent flooding and draining cycles, high  
45 organic matter content, and dynamic redox conditions in paddy soils create an environment that supports high levels of  
46 microbial activity, particularly Hg-methylating microorganisms (Yin et al., 2013). These conditions not only enhance  
47 MeHg production but also increase the likelihood of MeHg entering the food web through rice consumption, posing  
48 significant health risks (Zhang et al., 2010). Understanding Hg methylation in paddy fields is therefore crucial, as rice is  
49 a critical exposure route for MeHg in humans.

50 The accumulation of MeHg in rice is mostly attributed to microbial methylation of inorganic Hg in paddy soils (Meng  
51 et al., 2011). *In-situ* methylation and demethylation are deemed to be important processes controlling the net MeHg  
52 concentration in environments (Barkay and Gu, 2022; Helmrich et al., 2021; Li and Cai, 2012). Our recent study showed  
53 that Hg transformation processes, such as methylation, demethylation, oxidation, and reduction, occurred simultaneously  
54 in paddy soils, with Hg methylation being the most active (Liu et al., 2023). Therefore, paddy soil is a typical "hotspot"  
55 for Hg methylation, which is mainly a biotic process mediated by many abiotic factors, such as Hg bioavailability and  
56 redox conditions (Li and Cai, 2012). The diversity and activity of Hg-methylating microorganisms in paddy soils controls  
57 MeHg production (Gilmour et al., 2013; Liu et al., 2018). However, among the various Hg-methylating microorganisms  
58 currently known, the core microbiome controlling MeHg production and its interaction with environmental variables in  
59 paddy soils have yet to be identified.

60 Physicochemical factors in soils, such as organic matter, pH, salinity, redox potential, iron, and sulfur, have been shown  
61 to regulate the activity of Hg-methylating microorganisms and play an important role in controlling MeHg production in  
62 rice fields (Ullrich et al., 2001). Among the different variables, soil organic matter, which is ubiquitous in paddy soils (Li  
63 et al., 2018), play a vital role in Hg methylation (Yin et al., 2013). Dissolved organic matter (DOM), the most mobile  
64 organic matter fraction, increases MeHg production under sulfidic conditions (Graham et al., 2012). DOM increases  
65 microbial Hg bioavailability for methylation by stabilizing  $\beta$ -HgS(s) nanoparticles to prevent aggregation. In addition,  
66 Hg speciation in Hg-contaminated paddy soils was found to be predominantly regulated by organic matter (Liu et al.,  
67 2022), and the high bioavailability of DOM-bound Hg in rice paddies contributed to an increase in MeHg production (Liu  
68 et al., 2022). In contrast, other studies reported that DOM had a high affinity for Hg compounds (Skylberg et al., 2006),  
69 suppressing MeHg production due to strong Hg-DOM complexation (Schartup et al., 2015). As a result, the role of paddy  
70 soil DOM on Hg methylation remains elusive. Our recent study showed a significant and strong relationship between  
71 MeHg production and low-molecular-weight DOMs in paddy soils collected from major rice-producing areas across  
72 China (Abdelhafiz et al., 2023). Given paddy soil DOM's significant chemodiversity (Li et al., 2018), it is reasonable to  
73 hypothesize that the effect of DOM on MeHg production cannot be assessed solely based on Hg speciation and  
74 bioavailability, suggesting that other factors also play roles in MeHg production.

75 MeHg production is controlled by the synergy of Hg bioavailability and Hg-methylation capacity (Peterson et al., 2023),  
76 indicating that Hg-methylating microbial communities may also play an important role in DOM-regulated MeHg  
77 production. Concentration and composition of DOM have been shown to regulate MeHg production via alteration of the  
78 composition of the soil microbial community (Fagervold et al., 2014; Hu et al., 2021; Oloo et al., 2016). However, the  
79 core Hg-methylating microorganisms were not identified within these studies. Zhao et al. (2017) reported that two model  
80 Hg methylators exhibited an opposite response to DOM at the strain level. Therefore, we hypothesized that DOM fosters  
81 a core Hg-methylating microbiome that regulates MeHg production, since the core microbiome has a pivotal role in the  
82 functioning of ecosystems (Banerjee et al., 2018; Chen et al., 2019; Xun et al., 2021).

83 Thus, an attempt was made within this study to verify the crucial role of DOM in fostering the core Hg-methylating  
84 microbiome for MeHg production by (1) identifying the core Hg-methylating microbiome in paddy soils across a gradient  
85 of Hg contamination, (2) quantifying the relevance of DOM to core Hg-methylating microbiome and MeHg production  
86 in paddy soils compared with other soil physicochemical parameters, and (3) elucidating the mechanism of core Hg-  
87 methylating microorganisms in response to different DOMs. These results broaden our understanding of DOM as the  
88 prominent factor in altering Hg-methylating microbial communities and highlight the contribution of the core Hg-  
89 methylating microbiome to MeHg production in paddy soils.

## 90 **2 Materials and methods**

### 91 **2.1 Soil sampling and physico-chemical analysis**

92 Two field sampling campaigns were conducted in September 2020 and August 2022 in this study. Specifically, paddy  
93 fields from an abandoned Hg mining area (Sikeng, SK), an artisanal Hg smelting area (Gouxi, GX), and a regional  
94 background area (Huaxi, HX) in Guizhou Province, SW-China, were selected in September 2020 (Table S1, S1- S27). In  
95 each study area (SK, GX, and HX), nine sampling sites were randomly selected. Similarly, additional 19 sampling sites  
96 from the rice producing areas in 12 provinces of China were selected in August 2022 (Table S1, S28-S46). At each site,  
97 one rice paddy field was randomly selected. Paddy soil was taken from the root zone (10-20 cm deep) and comprised a  
98 composite of three subsamples from the same paddy field. A total of 46 soil samples were obtained in this study to  
99 represent different Hg contamination levels and bioavailability, net MeHg production, DOM concentration and  
100 composition, soil microbial community composition and structure, and other physicochemical characteristics. Soil  
101 samples were collected in the sterile PP bottles (Nalgene®, Thermo Fisher, USA) without any headspace, immediately  
102 shipped back to the laboratory on ice packs (~4°C) and divided into two subsamples before use. One subsample was  
103 stored at -20°C for microbial analysis, and the other was stored at 4°C for the analysis of soil physicochemical properties.  
104 Freeze-dried samples (-80 °C; Eyela FDU-2110, China) were screened to remove gravel and residue, then ground and  
105 evenly mixed using a mortar and pestle to pass through a 200-mesh sieve. The processed soil samples were analysed for  
106 pH, total carbon (TC), total nitrogen (TN), and various mercury species (water-soluble Hg, total Hg (THg), and MeHg),  
107 water-soluble sulfate (SO<sub>4</sub><sup>2-</sup>) and nitrate (NO<sub>3</sub><sup>-</sup>), DOM concentration (measured as water-soluble dissolved organic  
108 carbon), DOM composition (measured as optical properties of DOM) and low-molecular-weight organic acids. Fresh soil  
109 samples were also centrifuged to obtain pore water for the analysis of iron and sulfur (measured as Fe<sup>2+</sup> and S<sup>2-</sup> in soil  
110 pore water). Detailed measurement procedures are provided in Supplementary Text S1. It should be noted that Fe<sup>2+</sup> and  
111 S<sup>2-</sup>data were limited to soil samples obtained in August 2022.

112 **2.2 Soil DNA extraction**

113 We extracted DNA from 0.5 g of soil using the FastDNA Spin Kit for Soil (MP Biomedicals, France), following the  
114 manufacturer's instructions. The quality and concentration of the isolated DNA were assessed using spectrophotometry  
115 (Nanodrop ND1000, USA) and 1.0% agarose gel electrophoresis. The DNA was then stored at -80 °C for further analysis.

116 **2.3 Amplicon sequencing and bioinformatic analysis**

117 Soil Hg-methylating microbial communities were characterized by Illumina MiSeq sequencing of the *hgcA* gene using  
118 the primer pair ORNL-HgcAB-uni-F (5'-AAYGTCTGGTGYGCNGCVGG-3') and the reverse primer ORNL-HgcAB-  
119 uni-32R (5'-CAGGCNCCGCAYTCSTRCA-3') (Gionfriddo et al., 2020). Amplicons were equimolarly mixed, and  
120 sequenced using the Illumina MiSeq instrument (Illumina Inc., San Diego) in 2×300 bp mode. Poor-quality reads, adapters  
121 and primers were trimmed with SICKLE and CUTADAPT (Joshi and Fass, 2011; Martin, 2011). USEARCH (version  
122 8.0) was used to truncate, derePLICATE, sort and remove singletons (Edgar, 2013). The set of sequences obtained was  
123 clustered at a 60% similarity cutoff with cd-hit-est (Fu et al., 2012). Using USEARCH (version 8.0), the sequences were  
124 then mapped to the resulting clusters' representative sequences to build a count table. The sequences were annotated with  
125 amino acid sequences from Hg-MATE-Db (V1.01142021) (Gionfriddo et al., 2021) by using a Hidden Markov Model  
126 (HMM) based on HMMER (Eddy, 2011). In addition, the abundance of the Hg-methylating gene *hgcA* (which encodes a  
127 corrinoid protein essential for methylating inorganic Hg) was quantified in an Applied Biosystem 7500. The quantification  
128 of the *hgcA* gene is provided in Text S2.

129 **2.4 Metagenomic sequencing and bioinformatic analysis**

130 DNA from nine randomly selected paddy fields at each site in September 2020 was equimolarly mixed to obtain >1 µg  
131 of DNA for shotgun metagenomic sequencing. For paddy soils collected in August 2022, three replicates of each sample  
132 were utilized to ensure sufficient quantity and quality of DNA for metagenomic sequencing. A total of 22 samples were  
133 analysed using an Illumina HiSeq 2500 system (Illumina Corp., USA).

134 The detection and taxonomic identification of the *hgcAB* gene (full operon responsible for Hg methylation pathway)  
135 was performed with marky-coco (Capo et al., 2023). The metagenomic sequences were trimmed to eliminate low-quality  
136 reads using fastp with the following parameters: -q 30 -l 25 --detect\_adapter\_for\_pe --trim\_poly\_g --trim\_poly\_x (Chen  
137 et al., 2018). These high-quality reads were then assembled into contigs using megahit 1.1.2 with default settings (Li et  
138 al., 2016). The annotation of the contigs for prokaryotic protein-coding gene prediction was conducted using prodigal  
139 2.6.3 (Hyatt et al., 2010). To search for *hgc* homologs, a profile of HMM derived from Hg-MATE.db.v1 was applied to  
140 amino acid FASTA file generated from each assembly with the function hmmssearch from HMMER 3.2.1 (Finn et al.,  
141 2011). To eliminate paralogs of *hgcA*, we removed the sequences without the conserved putative cap helix motif  
142 [N(V/I)WCA(A/G)GK] reported previously (Parks et al., 2013). We further filtered the sequences by retaining only  
143 sequences with more than four transmembrane domains as identified by TMHMM (v.2.0) (Krogh et al., 2001). Finally,  
144 the obtained contigs with *hgcA* homologs were classified taxonomically following a previously described method (Zhang  
145 et al., 2023). In addition, to estimate the relative abundance of the *hgcA* gene, metagenomic reads were mapped to  
146 representative genomes of the *hgcA* dataset using Bowtie2 (Capo et al., 2023). The relative abundances of each gene were  
147 calculated by normalizing the total length of successfully mapped reads by gene length and the total number of reads in  
148 the metagenome.

149 Contigs ≥ 1000 bp were used to carry out binning analysis with the MetaWRAP pipeline (v1.3.2) (Uritskiy et al., 2018).  
150 The quality of reconstructed metagenome-assembled genomes (MAGs) was assessed using CheckM (Parks et al., 2015).  
151 High-quality MAGs (completeness ≥ 90% and contamination ≤ 10%) were used to detect *hgcA* homologs, and taxonomy

152 of these retrieved MAGs was conducted using GTDB-tk (v2.1.0) with its reference database (version release\_207V2)  
153 (Parks et al., 2022). To explore what fractions of DOM can be metabolized by core Hg-methylating microorganisms, core  
154 Hg-methylating microbial-associated MAGs were mapped to the protein sequence of the Kyoto Encyclopedia of Genes  
155 and Genomes (KEGG) database using eggNOG mapper (Huerta-Cepas et al., 2017).

156 **2.5 Pure incubation of *Geobacter sulfurreducens* PCA with different DOMs**

157 To validate that different concentrations and molecular weights of DOM stimulate the activity of core Hg-methylating  
158 microorganisms, we incubated *Geobacter sulfurreducens* PCA (*G. sulfurreducens* PCA), identified as a core Hg-  
159 methylating microorganism in this study, with Hg<sup>2+</sup>, and a natural DOM solution extracted from NMS, MMS, or HMS  
160 soils. *Geobacter* was selected for these pure incubation experiments due to its dominant role in mercury methylation and  
161 its ability to isolate the effects of DOM on methylation rates without the interference of soil matrix complexity. More  
162 details on the descriptions for the pure incubation experiment can be found in Text S3.

163 **2.6 Statistical analysis**

164 Statistical analysis was conducted with SPSS 27 (SPSS, Chicago, IL), AMOS (SPSS, Chicago, IL), and R platform  
165 (version 3.6.1). All statistical tests were considered significant at  $p < 0.05$ . The Kruskal-Wallis test was used to compare  
166 microbial alpha diversity among all samples. Hg-methylating microbial communities across differentially-polluted soils  
167 were compared by analysing dissimilarity matrices using Bray-Curtis distance and visualized using principal coordinates  
168 analysis (PCoA) and Adonis with the "ade4" and "vegan" packages (Dray and Dufour, 2007; Oksanen et al., 2017). To  
169 determine the relationship between THg and MeHg, Spearman correlation was performed using "ggpubr" and visualized  
170 using "ggplot2" packages (Kassambara, 2018; Wickham, 2009). Variation partitioning analysis was performed using  
171 "vegan" package (Oksanen et al., 2017). The major predictors of Hg-methylating microbial communities and their  
172 significance were identified using random forest analysis with "randomForest", "rfPermute" and "A3" packages (Archer,  
173 2018; Fortmann-Roe, 2015; Liaw and Wiener, 2002). To investigate the co-occurrence patterns among microbial taxa  
174 related to MeHg production, co-occurrence networks were established in the R platform using "psych" package (Revelle,  
175 2023), and visualized in Gephi 0.9.2 (Bastian et al., 2009) based on strong (Spearman's  $r > 0.8$ ) and significant ( $p < 0.01$ )  
176 correlations (De Caceres and Legendre, 2009). The modules in Hg-methylating microbial network were identified using  
177 default parameters from Gephi. To explore the relationship between the modules and environmental parameters, we  
178 correlated dissimilarities of bacterial composition in core Hg-methylating microbiome with those of environmental factors  
179 as previously described (Sunagawa et al., 2015). The structural equation model (SEM) was conducted using AMOS 28  
180 to evaluate the impacts of DOM and core Hg-methylating microbiome on MeHg production. A *prior* model was  
181 established based on the known relationships among drivers impacting MeHg production (Fig. S1). We further calculated  
182 the contribution of ecological parameters, including DOM, to the core Hg-methylating microbiome, and the contribution  
183 of the core Hg-methylating microbiome to MeHg production, following the approach described by Tao et al. This  
184 calculation was performed by determining the proportion of the squared path coefficient of each parameter relative to the  
185 sum of the squared path coefficients of all parameters influencing the same target variable (Tao et al., 2015).

186 **3 Results**

187 **3.1 Mercury production in paddy soils**

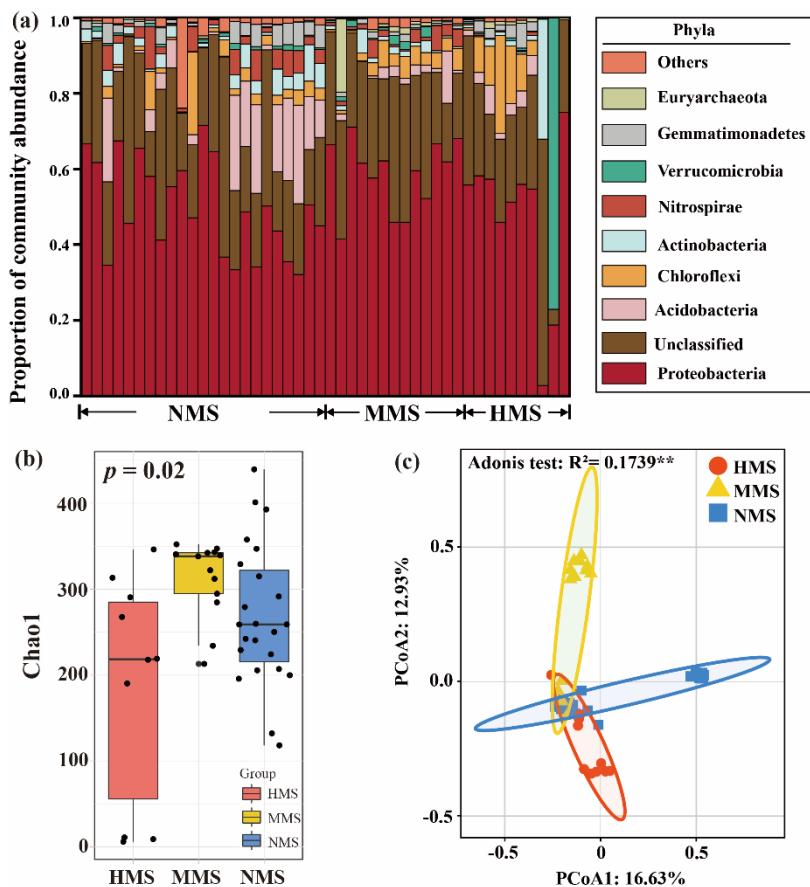
188 THg concentrations in paddy soils ranged from 0.03 to 1079.75 µg/g dw (Table S1). As reported in our previous study,  
189 dividing paddy soils by THg concentration rather than sampling sites facilitates a comprehensive investigation of the key

190 factors influencing Hg methylation (Abdelhafiz et al., 2023). Therefore, the paddy soils in this study were divided into  
191 three categories according to THg concentration: non-Hg contaminated soils (NMS, with average levels of  $0.24 \pm 0.18$   
192  $\mu\text{g/g dw}$ ,  $n=23$ ), moderate Hg-contaminated soils (MMS,  $18.28 \pm 6.77 \mu\text{g/g dw}$ ,  $n=13$ ), and high Hg-contaminated soils  
193 (HMS,  $637.79 \pm 160.93 \mu\text{g/g dw}$ ,  $n=10$ ). Furthermore, statistically significant differences in DOM concentrations  
194 (reflected by DOC concentration) and DOM composition (reflected by  $S_R$  of DOM) were found in NMS, MMS and HMS  
195 (Table S2). Specifically, DOC concentration varied significantly across the three soil types, with  $0.48 \pm 0.13$  in NMS,  
196  $0.40 \pm 0.07$  in MMS, and  $0.30 \pm 0.10$  in HMS. Similarly, the  $S_R$  of DOM differed markedly between NMS ( $1.40 \pm 0.76$ ),  
197 MMS ( $0.89 \pm 0.09$ ), and HMS ( $0.46 \pm 0.09$ ). However, no discernible differences in physicochemical properties (e.g., pH,  
198  $\text{S}^{2-}$ ,  $\text{SO}_4^{2-}$ ,  $\text{NO}_3^-$ , TN, TC,  $\text{Fe}^{2+}$ ) were observed in NMS, MMS and HMS (Table S3).

199 In this study, we found MeHg concentration in paddy soils in the order of HMS ( $5.01 \pm 0.77 \text{ ng/g dw}$ ,  $n=10$ ) > MMS  
200 ( $2.54 \pm 0.72 \text{ ng/g dw}$ ,  $n=13$ ) > NMS ( $0.76 \pm 0.25 \text{ ng/g dw}$ ,  $n=23$ ) (Fig. S2). Accordingly, a positive relationship was  
201 observed between total Hg and MeHg in different paddy soils (Fig. S3).

## 202 **3.2 Core mercury-methylating microbiome as predictors of MeHg production in paddy soils**

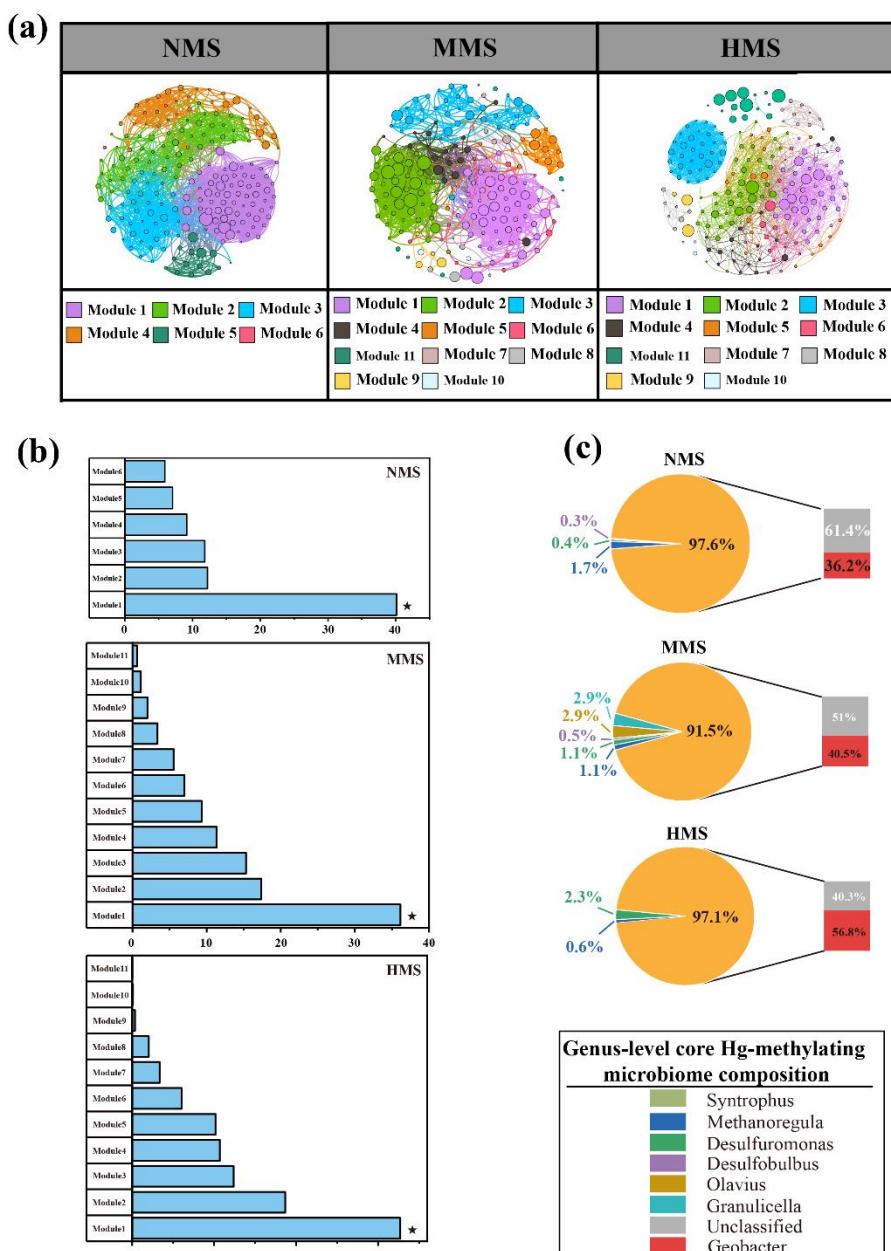
203 Random forest result revealed that *hgcA* gene abundance, DOM concentration, DOM composition, water-soluble Hg,  
204  $\text{Fe}^{2+}$ , and  $\text{S}^{2-}$  were significantly ( $p < 0.05$ ) associated with MeHg concentration (Fig. S4), with the *hgcA* gene as the  
205 strongest predictor. The *hgcA* gene-base taxonomic profiles of paddy soils reveal changes in Hg-methylating microbial  
206 community compositions across different levels of Hg pollution (Fig. 1a). Such observations were additionally supported  
207 by (1) the Chao1 index revealing the diversity of Hg-methylating microorganisms in the order of MMS ( $312.57 \pm 44.73$ ) >  
208 NMS ( $268.47 \pm 81.85$ ) > HMS ( $187.08 \pm 131.62$ ) ( $p < 0.05$ ; Fig. 1b) and (2) the divergent patterns of Hg-methylating  
209 microbial communities in paddy soils ( $p < 0.01$ ; Fig. 1c). The shotgun metagenomics results were consistent in detecting  
210 Hg-methylating microbial community composition and structure (Fig. S5). *Proteobacteria*, *Acidobacteria*, and  
211 *Chloroflexi* were the most abundant phyla in different paddy soils detected by both sequencing strategies. In summary,  
212 using both *hgcA* gene sequencing and metagenomic data, a significant difference in Hg-methylating microbial community  
213 structure and diversity was observed in paddy soils.



214

215 **Figure 1: Taxonomic profiles of Hg-methylating microbial communities in paddy soils based on amplicon sequencing.** (a) 216 Microbial community composition in differently polluted paddy soils. Phyla with low abundance grouped together under "other phyla". 217 (b) Microbial diversity (based on the Chao1 index) in differently polluted paddy soils. (c) Principal coordinates analysis (PCoA) based 218 on Bray-curtis distance showing the overall pattern of Hg-methylating microbial communities in differently polluted paddy soils. NMS, 219 non-Hg polluted paddy soils ( $n = 23$ ); MMS, moderate Hg-polluted paddy soils ( $n = 13$ ); HMS, high Hg-polluted paddy soils ( $n = 10$ ).

220 Network analysis captured six, eleven, and eleven modules (modularity index  $> 0.55$ ) in NMS, MMS, and HMS, 221 respectively (Fig. 2a, Table S4). Among all modules, Hg-methylating microorganisms in Module1 in NMS, MMS and 222 HMS were identified as core Hg-methylating microbiome based on their (1) higher connections to other modules and (2) 223 higher abundance in total Hg-methylating microbial community (Table S5). Importantly, the impact of various modules 224 in the microbial community on MeHg production was analyzed using random forest analysis. The results revealed that 225 the microbiome in Module 1 is a crucial bacterial group influencing soil methylmercury concentration (Fig. 2b). This 226 group is considered the core Hg-methylating microbiome in this study. Further analysis of the core Hg-methylating 227 microbiome composition revealed diverse core Hg-methylating microorganisms in paddy soils. Although most 228 microorganisms are not annotated, the three genera with the highest abundance in each soil type are as follows: In NMS, 229 *Geobacter* (36.2%), *Syntrophus* (1.7%), and *Desulfomonas* (0.4%) dominate; in MMS, *Geobacter* (40.5%), *Granulicella* 230 (2.9%), and *Olavius* (2.9%) are the most abundant; and in HMS, *Geobacter* (56.8%), *Methanoregula* (0.6%), and 231 *Granulicella* (2.3%) prevail (Fig. 2c). It is worth highlighting that, in this study, microorganisms belonging to *Geobacter* 232 were identified as the most significant core microorganisms for Hg methylation across all paddy soils.



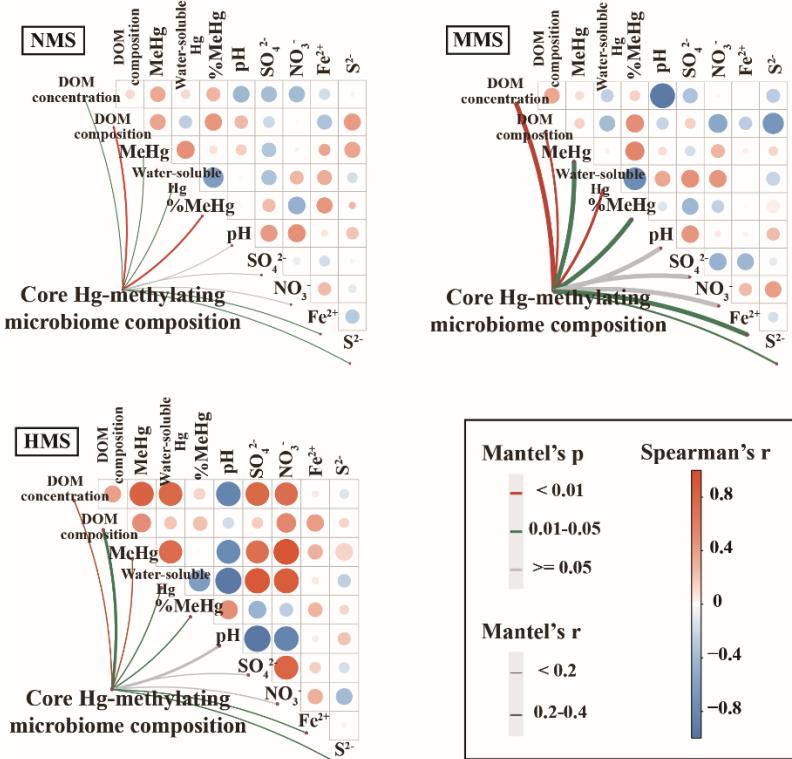
233

**Figure 2: Core Hg-methylating microbiome in paddy soils.** (a) Co-occurrence network of Hg-methylating microbial community in differently polluted paddy soils. Each node represents one OTU. The node size is proportional to the relative abundance of OTUs. (b) Predictors of the MeHg production in differently polluted paddy soils based on Random Forest analysis. Only predictors with significant effects are labeled asterisks. (c) Core Hg-methylating microbiome composition at genus level in differently polluted paddy soils. NMS, non-Hg polluted paddy soils (n = 23); MMS, moderate Hg-polluted paddy soils (n = 13); HMS, high Hg-polluted paddy soils (n = 10).

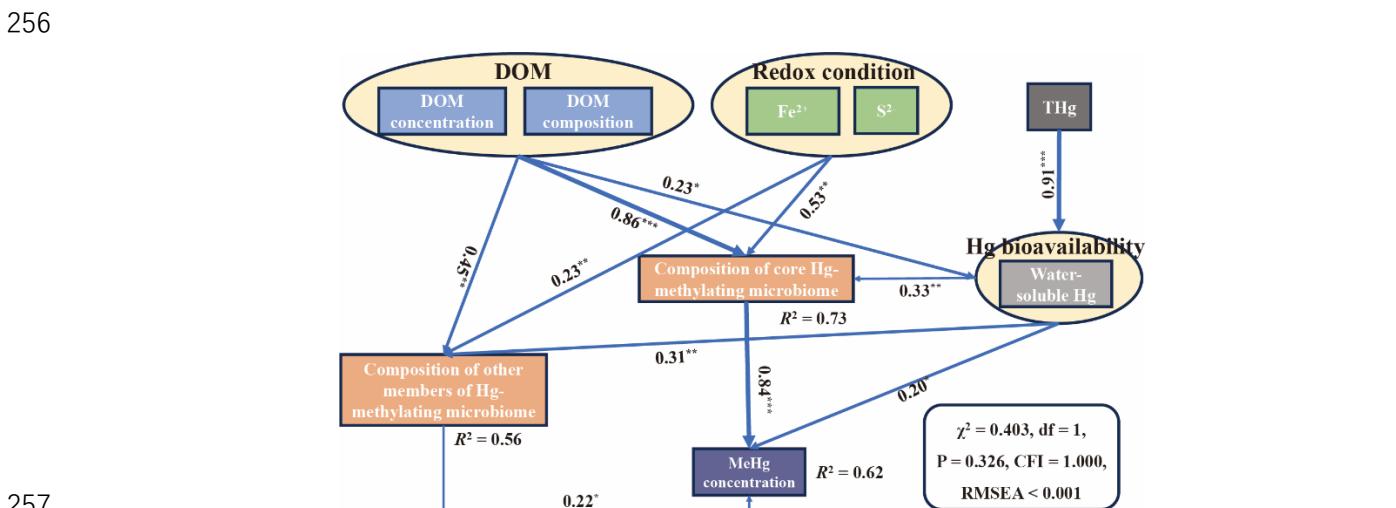
### 3.3 Dissolved organic matter as indicators of core mercury-methylating microbiome composition in paddy soils

Based on analysis of correlations, the results showed that there were significant correlations between core Hg-methylating microbiome composition, MeHg concentration, DOM concentration, DOM composition, water-soluble Hg, soil S<sup>2-</sup> and Fe<sup>2+</sup> (Fig. 3). Among all parameters, DOM is the most important factor influencing the composition of core Hg-methylating microbiome. This was supported by DOM explaining the most to core Hg-methylating microbiome composition (Fig. S6). Random forest analysis also showed that DOM concentration and composition were the most important predictors of the composition of core Hg-methylating microbiome (Fig. S7). Additionally, SEM result showed that the core Hg-methylating microbiome composition, which is closely linked to *hgcA* gene abundance, significantly regulated soil MeHg concentration ( $\lambda = 0.84, p < 0.001$ ) (Fig. 4). In comparison, the contributions of Hg bioavailability

249 and redox conditions to the core Hg-methylating microbiome composition are 10% and 25%, respectively, which are  
 250 much lower than that of DOM (65%) (Fig. 4).  
 251



252  
 253 **Figure 3. Pairwise comparisons of environmental factors and community taxonomic composition in core Hg-methylating**  
 254 **microbiome in differently polluted paddy soils.** NMS, non-Hg polluted paddy soils; MMS, moderate Hg-polluted paddy soils;  
 255 **HMS, high Hg-polluted paddy soils.**

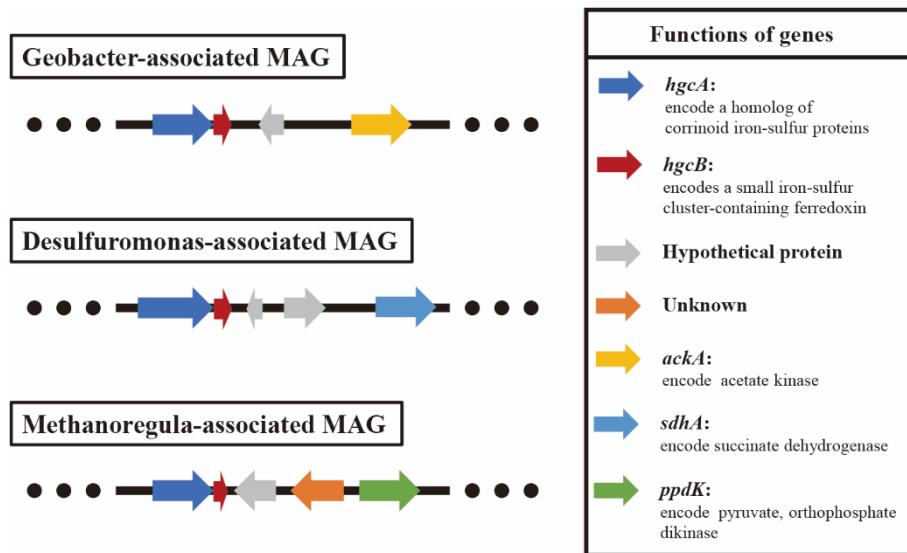


256  
 257 **Figure 4. Structural equation models showing the effects of DOM, redox conditions, and Hg bioavailability on MeHg**  
 258 **production.** NMDS1 values of the NMDS analysis were used for the representation of DOM and Redox condition in the SEMs.  
 259 Numbers adjacent to arrows are standardized path coefficients, and numbers in brackets denote p values. 'Statistically nonsignificant'  
 260 results are not shown in the figure. R<sup>2</sup> denotes the proportion of variance explained.  
 261

### 262 3.4 Dissolved organic matter stimulates activity of core mercury-methylating microorganism enhancing 263 methylmercury production in paddy soils

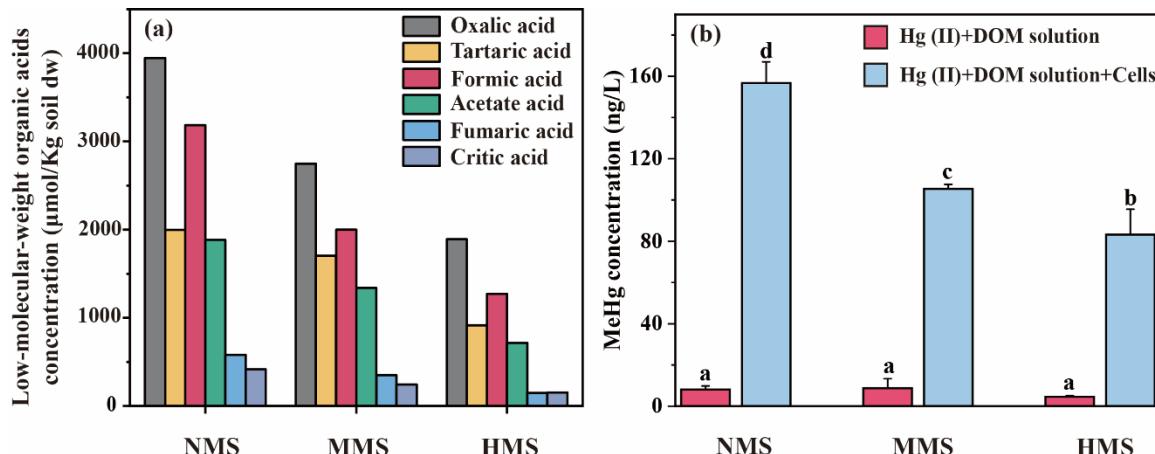
264 The results of metagenomic-binning revealed that three core Hg-methylating microbial-associated metagenome-  
 265 assembled genomes (MAGs, completeness  $\geq 90\%$  and contamination  $\leq 10\%$ ) carried different carbon utilization genes  
 266 (*ackA*, *sdhA*, or *ppdK* gene) (Fig. 5), which are responsible for acetate kinase, succinate dehydrogenase, pyruvate and

267 orthophosphate dikinase. These results indicated that the low-molecular-weight DOMs in soil selectively stimulate the  
 268 activity of core Hg-methylating microorganism that preferentially utilize them for metabolism, leading to the increase of  
 269 MeHg concentration.



270  
 271 **Figure 5. Analysis of the genetic context of *hgcA* gene and genes involved in carbon metabolism in core Hg-methylating**  
 272 **microbial-associated MAGs.** The extents and directions of genes are shown by arrows labeled with gene names.

273 To validate this hypothesis, *Geobacter sulfurreducens* PCA, core Hg-methylating microorganism identified in this  
 274 study, was incubated with HgCl<sub>2</sub> and various DOM solutions extracted from investigated paddy soils. The results showed  
 275 distinct patterns in MeHg production (Fig. 6), confirming that different concentration of low-molecular-weight DOMs  
 276 significantly regulates MeHg production by influencing the activity of core Hg-methylating microorganisms.



277  
 278 **Figure 6. Effect of natural DOM solution extracted from paddy soils on MeHg production by core Hg methylator (*Geobacter***  
 279 ***sulfurreducens* PCA).** (a) The concentration of low-molecular-weight organic acids in paddy soils from non-Hg polluted soils (NMS),  
 280 moderate Hg-polluted soils (MMS) and high Hg-polluted soils (HMS). (b) MeHg concentration by *G. sulfurreducens* PCA. Data (n =  
 281 3) are presented as mean value  $\pm$  SD, with error bars representing standard deviations. Significant differences among different  
 282 treatments were tested with Tukey's honest significance test; different lowercase letters in each bar indicate significant differences  
 283 among treatments ( $p < 0.05$ ).

#### 284 4 Discussion

285 Our study found that MeHg concentration was strongly linked to *hgcA* gene abundance even compared to abiotic factors,  
 286 which suggested that MeHg production is a microbially-mediated process (Parks et al., 2013; Podar et al., 2015). Our  
 287 study further revealed that although there are significant differences in the Hg-methylating microbial communities in  
 288 different polluted paddy soils, they all have a core Hg-methylating microbiome, which plays a more important role than

289 other Hg methylators in regulating MeHg production. As illustrated by a previous study, the major module (also known  
290 as the core microbiome) in microbial community network contributes to the stability of soil microbiome, enhancing its  
291 resistance to climate changes and nutrient fertilization (Jiao et al., 2022). These findings establish the presence of a major  
292 module contributing exclusively to Hg methylation in paddy soils, although there are many more Hg-methylating  
293 microorganisms present. In fact, microorganisms containing the *hgcA* gene are able to methylate Hg, but this does not  
294 mean that they are automatically active in Hg methylation.

295 The SEM analysis result indicated that although redox conditions and Hg bioavailability significantly affected the  
296 composition of core Hg-methylating microbiome, their contribution to the composition of core Hg-methylating  
297 microbiome was less and weaker than that of DOM. The explanation for this phenomenon may be that (1) the soil  
298 collected in the paddy field during the flooding period is in an anaerobic state, so the selection of redox conditions on  
299 core mercury-methylating microorganisms is weakened; (2) Hg is a toxic element to microorganisms and is usually not  
300 involved in microbial metabolism (Wang et al., 2020). Environmental Hg may induce the persistence of some  
301 microorganisms. Therefore, long-term Hg contamination often only elevates the abundance of specific microbial taxa  
302 capable of Hg tolerance (Frossard et al., 2018); (3) DOM, an important carbon source and nutrient in nature, is involved  
303 in microbial respiration and metabolism (Kujawinski, 2011). Consequently, the concentration and composition of DOM  
304 contributed significantly to core Hg-methylating microbiome. These results highlight the dominant role of DOM in  
305 shaping core Hg-methylating communities, as compared to redox conditions and Hg bioavailability.

306 Our study found that *Geobacter*, *Desulfuromonas*, and *Methanoregular*, *Syntrophus*, *Granulicella*, and *Olavius* are  
307 core Hg-methylating microorganisms in paddy soils. Previous studies confirmed that *Geobacter*, *Desulfuromonas*, and  
308 *Syntrophus* have the capability for Hg methylation (Bravo et al., 2018; Gilmour et al., 2013; Liu et al., 2018; Zhong et al.,  
309 2024). In addition, *Methanoregular* spp., as methanogenic archaea, show potential for Hg methylation (Jones et al., 2019).  
310 *Granulicella* affects the decomposition of complex organic materials (Pankratov and Dedysh, 2010), while *Olavius* plays  
311 a role in sulfur and nitrogen cycling (Blazejak et al., 2005). These roles suggest that both microorganisms could also be  
312 important potential Hg methylators. Although many core Hg-methylating microorganisms have not been annotated, our  
313 study emphasizes that the annotated Hg-methylating microorganisms play a much greater role in Hg methylation in paddy  
314 soils than previously thought.

315 Our study identified various DOMs components, including oxalic acid, tartaric acid, formic acid, acetate acid, fumaric  
316 acid, and citric acid, in paddy soils. These low-molecular-weight organic acids, particularly abundant in NMS soils, serve  
317 as key carbon sources for Hg-methylating microorganisms and stimulate the growth and activity of the core Hg-  
318 methylating microbiome. Pure incubation of *Geobacter sulfurreducens* PCA (core Hg-methylating microorganism  
319 identified in our paddy soils) further confirmed that different concentration of low-molecular-weight DOM solution  
320 extracted from natural paddy soils obtained from NMS, MMS and HMS had significant effects on MeHg concentration.  
321 These findings demonstrate that DOM composition strongly influences microbial Hg methylation by stimulating key  
322 metabolic pathways. For instance, *Geobacter sulfurreducens* and *Desulfovibrio desulfuricans* use acetate and fumarate in  
323 the TCA cycle, supporting anaerobic respiration and electron transport that enhance Hg methylation (Hu et al., 2013; Liu  
324 et al., 2018). Similarly, methanogenic archaea such as *Methanoregula* and *Methanosarcina* utilize formate and acetate  
325 through methanogenesis, further contributing to Hg methylation (Sakai et al., 2010; Schöne et al., 2022). Although  
326 metabolomic data were not included in this study, future research incorporating such analyses could provide valuable  
327 insights into how specific DOM components influence microbial metabolism and Hg methylation, revealing key  
328 metabolites and pathways such as acetate fermentation, methanogenesis, and electron transfer processes. This highlights  
329 how specific DOM components shape the core Hg-methylating microbiome and influence its role in MeHg production.

330 In contrast to low-molecular-weight organic acids, other DOM components such as aromatic compounds and humic  
331 substances may have limited influence on microbial Hg methylation due to their complex structures and reduced  
332 bioavailability. While aromatic compounds and humic substances were not directly analyzed in this study, their complex  
333 structures likely reduce Hg bioavailability or slow microbial degradation, resulting in weaker effects on Hg methylation  
334 compared to low-molecular-weight organic acids. Future research could integrate direct Hg speciation measurements with  
335 detailed DOM compositional analyses to better understand how specific DOM components and Hg species interact to  
336 influence microbial Hg methylation.

337 DOM's influence on microbial Hg methylation has been observed in other ecosystems, such as wetlands and sediments,  
338 where DOM shapes microbial community structures to promote methylmercury (MeHg) production. For instance, in  
339 wetlands, DOM-bound Hg has been found to change the community assembly of mercury for methylating microbes  
340 (Fagervold et al., 2014). This highlights the broader ecological significance of DOM's role in promoting Hg methylation  
341 and suggest that DOM-driven microbiome modulation is a critical process across diverse environments. Moreover, the  
342 knowledge gained in this study highlights how variation in DOM quality due to human activities and climate change (e.g.,  
343 changes in molecular weight, aromaticity, and bioactivity) could significantly alter MeHg production in different  
344 environmental compartments (Xenopoulos et al., 2021). For instance, long-term processes may scatter stable DOM, such  
345 as black carbon, globally through biomass combustion (Qi et al., 2020), while simpler and more reactive DOM may  
346 dominate in aquatic ecosystems (Xenopoulos et al., 2021). These changes could either enhance or diminish Hg ecotoxicity,  
347 depending on the specific conditions. Therefore, future in-depth studies coupling DOM quality, Hg speciation, and  
348 microbial Hg methylation are essential to deliver more accurate assessments of Hg's environmental and health impacts,  
349 particularly in the context of the Minamata Convention.

## 350 5 Conclusions

351 This study provides novel evidence that DOM significantly influences MeHg production by altering the composition and  
352 stimulating the activity of the core Hg-methylating microbiome. While DOM regulates the composition of other members  
353 of the Hg-methylating microbiome, its impact on MeHg production is primarily mediated through the core Hg-  
354 methylating microbiome. Using metagenomic binning and pure incubation experiments, we demonstrated that low-  
355 molecular-weight DOM directly promotes MeHg production by enhancing the metabolic activity of core Hg-methylating  
356 microorganisms. These findings underscore the central role of the core Hg-methylating microbiome in Hg cycling and  
357 highlight DOM as a critical driver of microbial Hg methylation. As human activities and climate change continue to alter  
358 DOM composition and concentration, their influence on Hg methylation dynamics warrants further investigation to better  
359 predict and mitigate Hg-related environmental and health risks.

360 *Data Availability.* The raw reads of *hgcA* gene amplicon sequencing have been deposited in the NCBI SRA under  
361 accession number PRJNA847325 and PRJNA972506. Shotgun metagenomic sequencing have been deposited in the

362 NCBI SRA under accession number PRJNA848068 and PRJNA972502. Other datasets generated during the current study  
363 are available from the corresponding author upon reasonable request.

364 *Author Contributions.* The study was designed by QP, BM, and XBF. QP, JL and YRL conducted the sampling, performed  
365 the DNA extraction and the bioinformatic analyses. JHH, KZ and MA performed the geochemical analyses. The  
366 manuscript was written by QP and BM, with assistance and input from co-authors.

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