Vira Leng

lengvira@yahoo.com

Cambodian Conservation Agriculture Research for Development Center
Dept. of Agricultural Land Resources Management, GDA, MAFF
No. 54B / 49E, Street 395-656, Teuk Laak 3, Toul Kork, Phnom Penh, Cambodia
AIDA, University of Montpellier, CIRAD, Montpellier, France
Av. Agropolis, 34980 Montferrier-sur-Lez, France
+855 12 321 568

Dear Editors,

On behalf of all the co-authors, I sincerely appreciated the executive editor for giving us the opportunity to resubmit our manuscript entitled "Diachronic assessment of soil organic C and total N dynamics under long-term no-till cropping systems in the tropical upland of Cambodia" for consideration in the SOIL journal.

We appreciated the reviewers' time and effort in providing valuable comments and suggestions on our manuscript, and we have carefully considered and addressed all of them. In addition to the specific responses to the reviewers' comments, we have taken into account the editor's final comment throughout the manuscript to improve readability.

In response to the reviewers' comments, we made major revisions to improve the clearness, robustness, and overall quality of our manuscript. Below are the brief significant modifications we made:

- We modified the abstract, introduction, materials and methods, results, and conclusion sections in compliance with the detailed point-by-point comments and suggestions from the reviewers.
- We completely removed the information related to reference vegetation (RV) from the main text.
- We moved the tables 2 and 5 and figure 3 from the main text to the supplementary materials.
- We provided a more detailed description of the treatments; however, we decided not to change the title of the manuscript and the name of the treatments from "NT" to "Conservation Agriculture," as suggested by a reviewer because all treatments are under no-tillage, but not all of them can be considered as conservation agriculture, for instance no-tillage with monocropping. We therefore believe that "no-tillage cropping systems" is the best term to use in our study. We have however refined the treatment names to be clearer (CT into CTM, NT1 into NTM, NT2 into NTR1, and NT3 into NTR2).

Thank you very much for considering our revised manuscript for publication in the SOIL Journal. We look forward to your feedback, and if there are any additional comments and suggestions from the reviewers and the editors, we are pleased to carefully consider and address them for further improving our manuscript to meet the requirements for publication in the SOIL.

Sincerely,

Vira Leng

Phnom Penh, Cambodia, 21.08.2024

Tel: +855 12 321 568

Email: lengvira@yahoo.com

# **Diachronic** assessment

## of soil organic C and N dynamics under long-term no-till 2 cropping systems in the tropical upland of Cambodia 3 Vira Leng<sup>1,2\*</sup>, Rémi Cardinael<sup>3,4,5</sup>, Florent Tivet<sup>1,3,6</sup>, Vang Seng<sup>1</sup>, Phearum Mark<sup>1</sup>, 4 5 Pascal Lienhard<sup>3,7,8</sup>, Titouan Filloux<sup>3</sup>, Johan Six<sup>9</sup>, Lyda Hok<sup>10</sup>, Stéphane Boulakia<sup>3</sup>, Clever Briedis<sup>11</sup>, João Carlos de Moraes Sá<sup>12</sup>, Laurent Thuriès<sup>13,14</sup> 6 7 <sup>1</sup>Cambodian Conservation Agriculture Research for Development Centre, Department of Agricultural 8 Land Resources Management, General Directorate of Agriculture, MAFF, Phnom Penh, Cambodia <sup>2</sup>Doctoral school GAIA, SupAgro Montpellier, University of Montpellier, France 9 <sup>3</sup>AIDA, University of Montpellier, CIRAD, Montpellier, France 10 11 <sup>4</sup>CIRAD, UPR AIDA, Harare, Zimbabwe <sup>5</sup>Department of Plant Production Sciences and Technologies, University of Zimbabwe, Harare, 12 13 Zimbabwe <sup>6</sup>CIRAD, UPR AIDA, Phnom Penh, Cambodia. 14 <sup>7</sup>Northern Mountainous Agriculture - and Forestry Science Institute - Phu Ho commune - Phu Tho District 15 16 Phu Tho, Viet Nam.

- 17 <sup>8</sup>CIRAD, UPR AIDA, Hanoi, Viet Nam.
- 18 <sup>9</sup>Department of Environmental Systems Sciences, ETH Zurich, 8092, Switzerland
- 19 <sup>10</sup>Center of Excellence on Sustainable Agricultural Intensification and Nutrition, Royal University of
- 20 Agriculture, Phnom Penh, Cambodia
- 21 <sup>11</sup>Department of Agronomy, Federal University of Vicosa, Av. Peter Henry Rolfs s/n 36570-900,
- 22 Viçosa, MG, Brazil
- 23 <sup>12</sup>Department of Soil Science and Agricultural Engineering, State University of Ponta Grossa, Av.
- 24 Carlos Cavalcanti 4748, 84030-900, Ponta Grossa, PR, Brazil
- 25 <sup>13</sup>CIRAD, UPR Recyclage et Risque, F-34398 Montpellier, France.
- <sup>14</sup>Recyclage et Risque, University of Montpellier, CIRAD, Montpellier, France.
- 28 \*Corresponding author: Vira LENG (lengvira@yahoo.com)

27

## Abstract

30

31

32

33

34

35

36

37

38

39

40

41

42

43

44

45

46

47

48

49

50

No-till (NT) cropping systems have been proposed as a strategy to combat soil degradation by storing soil organic carbon (SOC) and total nitrogen (TN). We quantified the impacts of NT cropping systems on the changes in SOC and TN stocks and in particulate and mineral-associated organic matter fractions (POM and MAOM), to 100 cm depth, from three 13-year-old experiments in a tropical red Oxisol in Cambodia using diachronic and equivalent soil mass approaches. Established in 2009 and arranged in a randomized complete block design with triplicates, the experiments included maize (MaiEx)-, soybean (SoyEx)-, and cassava (CasEx)-based cropping systems. Each experiment comprised three treatments: (1) monocropping of main crops (maize, soybean, and cassaya) under conventional tillage (CTM); (2) monocropping of main crops under NT systems with the use of cover crops (NTM); and (3) bi-annual rotation of main crops under NT systems with the use of cover crops (NTR), both crops being presented every year and represented by NTR1 and NTR2. Soil samples were collected in 2021, 10 years after the last sampling. All the NT systems significantly (p < 0.05) increased SOC stock in the topsoil in SoyEx and MaiEx and down to 40 cm in CasEx. Considering the whole profile (0-100 cm), the SOC accumulation rates ranged from 0.86-1.47 and 0.70-1.07 Mg C ha<sup>-1</sup> yr<sup>-1</sup> in MaiEx and CasEx, respectively. Although SOC stock significantly increased in CTM in 0-20 cm in MaiEx and CasEx, it remained stable in 0–100 cm in all the experiments. In 0–5 cm, NTR systems significantly increased TN stock in all the experiments, while in NTM system it was only significant in MaiEx and SoyEx. In 0–100 cm, TN stock in all the experiments remained stable under NTR systems, whereas significant decrease was observed under NTM systems in SoyEx and CasEx. Although C-POM stock significantly increased under all NT systems limited to 0–10 cm in MaiEx and SoyEx, all the NT systems significantly increased

51 C-MAOM stock in the 0–10 cm layer in MaiEx and SoyEx and down to 40 cm in CasEx. All the NT
52 systems significantly increased N-POM stock in 0–10 cm in MaiEx and SoyEx, while a significant
53 decreased in N-MAOM stock was observed below 5 cm in CasEx and below 40 cm in MaiEx and SoyEx.
54 Our findings showed that long-term NT systems with crop species diversification accumulated SOC not
55 only on the surface but also in the whole profile by increasing both SOC in the POM and MAOM, even
56 in the cassava-based system. This study highlights the potential of NT systems for storing SOC over time,
57 but raises questions about soil N dynamics.

# 1 Introduction

Land and soil degradation is a global challenge with consequences not only for food and nutrition security but also for livelihoods, environmental pollution, climate change, water scarcity, and biodiversity. The main processes that cause soil degradation are water and wind erosion, chemical depletion, physical deterioration, decline in soil organic carbon (SOC), loss in biodiversity, acidification, and salinity (Lal, 2015a; Stavi and Lal, 2015; Dragović and Vulević, 2020; Barbier and Di Falco, 2021). It was estimated that about 562 million ha of land are degraded worldwide, with 5 to 10 million ha of land lost each year as a result of severe degradation (Stavi and Lal, 2015; Nkonya et al., 2016). The major factors contributing to soil degradation are deforestation and land clearance, the overuse of agrochemicals, and intensive agricultural management practices (Dragović and Vulević, 2020). Tropical soils have the highest risks of degradation due to the combination of high rainfall intensity and the ongoing intensification of agriculture to meet the food demand of a fast-growing population, which is also constrained by the limited availability

71 Di Falco, 2021). 72 Cambodia, located in the tropical region of Southeast Asia, is one of the highest land degradation hotspots 73 in the world, and about 55% of the country's population reside in these hotspot areas (UNCCD, 2018). In 74 the last two decades, human-induced activities including deforestation, land clearance for agriculture, 75 climate change, and intensive farming practices have further worsened Cambodia's already poor soil 76 fertility (UNCCD, 2018; Ken et al., 2020; ADB, 2021). Over the past two decades, 30%, or about 4.24 77 million ha, of forest areas were converted to croplands, putting pressure on natural resources, biodiversity, 78 and threatening the provision of several ecosystem services (World Bank Group, 2023). In the Northwest 79 rainfed uplands, like in other parts of the country, studies on soil erosion at field scale and modelling 80 reported that the annual soil loss rate in conventional plough-based tillage (CT) ranged from 0.33 to more than 80 Mg soil ha<sup>-1</sup> yr<sup>-1</sup>, depending on soil type and land slope (CARDI, 2017; Nut et al., 2021; Sourn et 81 82 al., 2022). The amplitude of soil erosion increased by 41% from an annual erosion rate of 2.92 Mg soil 83 ha<sup>-1</sup> yr<sup>-1</sup> in 1998 at the beginning of the forest conversion to agriculture with extensive, more diversified farming practices to 4.98 Mg soil ha<sup>-1</sup> yr<sup>-1</sup> in 2018 under CT maize- and cassava-based monocropping 84 85 systems (Nut et al., 2021; Sourn et al., 2022, 2023). It was estimated that approximately 3-4 mm of topsoil 86 is washed away annually for this Northwestern region of Cambodia (Nut et al., 2021; Sourn et al., 2023). 87 Erosion induces soil degradation and a loss of SOC for the eroded fields (Polyakov and Lal, 2004). It was 88 estimated that from 2000 to 2010, Cambodia lost approximately 1.98 million Mg C in the top 0-30 cm 89 depth as the consequence of forest conversion to other land uses (MAFF, 2018). Cambodian soils are 90 seriously threatened by intensive agricultural systems. The returns on taking actions against land

of land to be converted to cropland (Barbier and Hochard, 2018; Craswell and Lefroy, 2001; Barbier and

degradation through restoration and adoption of sustainable agricultural management practices are 92 estimated at 3 US dollars for every dollar invested in restoring degraded land in Cambodia, highlighting 93 the strong economic benefits (UNCCD, 2018). 94 SOC serves as the foundation of soil physical, chemical, and biological processes that sustain essential 95 ecosystem functions, and it is the reservoir of plant nutrients and energy for biota (Lal, 2015a). Therefore, 96 adopting sustainable management practices that lead to increase in SOC stock (Beillouin et al., 2023) is 97 part of the key strategies to reverse the soil degradation trends and to minimize the economics and 98 environmental impacts related to land degradation (Lal, 2015a; Obalum et al., 2017). Studies reported 99 that agricultural practices, in particularly those based on CT, weaken soil structure, accelerate soil erosion, 100 and deplete SOC stock (Tivet et al., 2013; Sá et al., 2014; Briedis et al., 2018; Oliveira et al., 2020; Tiecher 101 et al., 2020). By contrast, conservation agriculture (CA), defined by three key principles, (i) minimum or 102 no soil disturbance, (ii) permanent soil cover with mulch or cover crops, and (iii) crop diversification 103 through rotation or association, is a potential strategy to overcome soil degradation (Luo et al., 2010; Lal, 104 2015b; Powlson et al., 2016; Obalum et al., 2017). No-till cropping systems (NT), are part of the 105 conservation agriculture practice approach, and involve a range of practices with a reduction or an absence 106 in soil tillage and a high diversity of crop and cover crop species. The benefits of NT cropping systems 107 on soil health improvement have been reported worldwide. Diversified NT cropping systems enhance 108 both SOC stock (Hok et al., 2015) and pools (Sá et al., 2014; Briedis et al., 2018; Sithole et al., 2019; 109 Cooper et al., 2021; Rodrigues et al., 2022), especially through an increase in physical protection of 110 particulate organic C (Sithole et al., 2019) and mineral-associated organic C (Rodrigues et al., 2022) 111 inside soil aggregates. In addition, numerous studies have reported the co-benefits of NT cropping

systems on soil health enhancement (Pheap et al., 2019; Koun et al., 2023), increased water infiltration. reduced soil erosion (TerAvest et al., 2015; Sithole et al., 2019), enhanced microbial activities (Hok et al., 2018) and abundance (Lienhard et al., 2013). Yet, there are still arguments about the benefits of NT cropping systems and associated factors that determine SOC accumulation. Particulate organic matter (POM) and mineral-associated organic matter (MAOM) are the two main fractions of the SOC pools. They differ in physical and chemical characteristics as well as their turnover rates. POM is more sensitive to soil tillage and land use than MAOM and total SOC (Blanco-Moure et al., 2013; Kan et al., 2021). Therefore, documentation of SOC fractions is desirable for a better understanding of SOC dynamics and stabilization processes (Lavallee et al., 2020). In a meta-analysis with the majority of the studies collecting samples between 0.15 and 0.3 m depth, Powlson et al., (2016) reported that SOC accumulation rate under CA systems ranged from 0.16 to 0.49 Mg C ha<sup>-1</sup> yr<sup>-1</sup> in tropical soils in the Indo-Gangetic Plains and from 0.28 to 0.96 Mg C ha<sup>-1</sup> yr<sup>-1</sup> in Sub-Saharan Africa. In a Ferralsol in Zimbabwe, Shumba et al., (2024) reported a SOC accumulation rate of 0.13 Mg C ha<sup>-1</sup> yr<sup>-1</sup> in the 0-5 cm layer only under CA, but not change under NT only. However, in meta-analyses, Angers and Eriksen-Hamel (2008) and Luo et al., (2010) found that conversion from CT to NT only changed the SOC distribution in the soil profile but did not significantly increase SOC stock in the whole profile. Boddey et al., (2010) and Xiao et al., (2020) reported that NT significantly increased SOC stock only at the soil surface but not in the deeper layers. It is therefore crucial to quantify SOC change in subsoil when assessing the impact of practices, especially NT systems. SOC storage is closely related to soil aggregate structure (Six et al., 2004; Liu et al., 2021). The complexity of cropping systems, characterized by crop species diversity through the use of cover crops,

112

113

114

115

116

117

118

119

120

121

122

123

124

125

126

127

128

129

130

131

crop rotation, and intercropping, was reported to enhance soil aggregation stability and the proportion of soil macroaggregates, along with an increase in SOC (Tiemann et al., 2015; Li et al., 2024). The diversity of crop species increased the quantity and chemical diversity of plant-derived litter inputs, which are the main sources of energy for soil microorganisms, and increased microbial activity and the abundance of fungal and bacterial communities (Tiemann et al., 2015; Zhang et al., 2023). The overall increase in fungal hyphae, plant roots, and aboveground biomass inputs under crop diversification are important organic binding agents that promote the formation of macroaggregates and facilitate the soil aggregation process (Tiemann et al., 2015). Furthermore, the increased amount and diversity of plant-derived C inputs in the forms of crop residues and root exudates provided a suitable microenvironment for soil microorganisms, which promoted microbial growth and turnover (Morugán-Coronado, 2022). The faster microbial growth and turnover rates increased the amount of microbial biomass and necromass, thus increasing SOC (Liang et al., 2011; Prommer et al., 2019). The amount, quality and frequency of the crop residues added to soil under a range of climate-driven decomposition rates, soil mineralogy and profile characteristics are important factors to consider to increase SOC stocks (Paustian et al., 1997; Six et al., 2002; Bayer et al., 2006; Ogle et al., 2012; Virto et al., 2012). It has been suggested that the amount of biomass-C inputs was the main factor explaining the variability in SOC storage between sites under NT (Virto et al., 2012). In a synthesis from tropical soils, Fujisaki et al., (2018) reported that the amount of biomass-C inputs was the main factor driving SOC stock change. In a meta-analysis in Sub-Saharan Africa, Corbeels et al., (2019) found that no-tillage alone does not lead to an increase in SOC stock, but CA systems combining the three principles could. It therefore seems that there is a hierarchy in CA principles to increase SOC stock, the most important one being the permanent soil cover, followed by a reduction in soil tillage and

133

134

135

136

137

138

139

140

141

142

143

144

145

146

147

148

149

150

151

152

155 where crop residue retention and cover crops were the most efficient CA practices to increase SOC 156 (Beillouin et al., 2023). 157 Two different soil sampling approaches are commonly used for assessing SOC stock change, the 158 diachronic and the synchronic approaches (Bernoux et al., 2005). The diachronic approach refers to 159 collecting samples on the same field plots over time. The synchronic approach, also known as the space-160 for-time method, refers to sample collection at the same time from different (often adjacent) field plots 161 under different land-use or management systems (Bernoux et al., 2005; Neto et al., 2010). Neto et al., 162 (2010) and Junior et al., (2013) revealed that the synchronic approach led to biased estimation of SOC 163 accumulation from long-term experiments in Brazil due to spatial heterogeneity and initial land use 164 history. They highlighted that diachronic soil sampling should be used for assessing soil SOC storage 165 rates due to changes in land-use or management patterns because it offers a more comprehensive view of 166 how SOC and N levels change under long-term tillage and cropping systems over time in which non-167 identical initial soil conditions cannot practically be excluded, making it more accurate and realistic for 168 the investigation of SOC and N dynamics, despite the fact that they are costly and require significant time 169 and resources (Bernoux et al., 2005; Neto et al., 2010; Junior et al., 2013). The synchronic approach, on 170 the other hand, is simpler, lower-cost, and less time-consuming, but comes with more uncertainty (Neto 171 et al., 2010; Junior et al., 2013). A change in soil bulk density is often observed when comparing NT 172 systems to CT, due to differences in tillage but also to root systems of cover crops. It is therefore required 173 to estimate SOC change using an equivalent soil mass approach instead of a fixed depth approach (Ellert 174 and Bettany, 1995).

improved rotations (Shumba et al., 2024). This has been confirmed in a recent second-order meta-analysis

175 NT cropping systems have been promoted to smallholders in various agroecosystems in Cambodia since 176 2009. The early effects of NT cropping systems on soil health, and SOC storage have been reported in 177 several studies (Hok et al., 2015, 2018, 2021; Pheap et al., 2019; Suong et al., 2019; Sar, 2021; Koun et 178 al., 2023), however, the information on the impact of long-term NT systems on the changes in SOC and 179 TN stocks remains scarce in the country as well as in Southeast Asia. There is a need to document the 180 long-term changes in SOC and TN stocks under NT cropping systems to fill in the knowledge gaps as 181 well as provide robust evidences to land use planners and policymakers. This could be profitable not only 182 for Cambodia but also for other countries in the region. 183 Therefore, this study aimed to quantify the impacts of CT and different NT cropping systems on the 184 changes in SOC and TN stocks and fractions over time (2011–2021) in Cambodia's tropical red Oxisol 185 using diachronic and equivalent soil mass (ESM) approaches. We hypothesized that implementation of 186 the three core technical principles of CA would significantly enhance the SOC stocks, both in the POM 187 and MAOM size fractions, including in the subsoils.

## 2 Materials and Methods

188

189

# 2.1 Study site description

The study was conducted at Bos Khnor Conservation Agriculture Research Station, the oldest CA research station in Southeast Asia, which belongs to the General Directorate of Agriculture (GDA), Department of Agricultural Land Resources Management (DALRM). It is located in Chamkar Leu district, Kampong Cham province (12°12'31.0"N 105°19'07.0"E, 118 m above sea level). The details of the study site were reported in Hok et al., (2015). Briefly, the site was the natural tropical rainforest, which was then converted to perennial cropland in 1937. The crops included cashew, coffee, mango, mulberry,

avocado, and rubber, which were planted soon after forest clearance. Because of the civil war (Khmer Rouge) between 1970 and 1982, the area was abandoned and taken over by several tree species, such as Tetrameles nudiflora R Br., Nauclea officinalis L., Cassia siamea (Lam.) H.S.Irwin & Barneby, and Leucaena leucocephala (Lam.) de Wit, which grew naturally. The farming was resumed, and cotton (Gossypium hirsutum L.) and banana (Musa acuminata spp.) were planted from 1982 to 2000. From 2000 to 2009, successive annual crops per year of cotton, followed by mung bean (Vigna radiata (L.) R. Wilczek), and sesame (Sesamum indicum L.), followed by soybean (Glycine max L.), were rotated under conventional plough-based management before the establishment of the three experiments. Mineral fertilizers such as NPK (15-15-15), ammonium phosphate (16-20-0), and potassium chloride (0-0-60) were applied to the crops without lime application. The soil of the study site is classified as a red Oxisol (USDA, 1999) or a Ferralsol in the world reference base (WRB) for soil resources (IUSS Working Group WRB, 2015), with 1% sand, 29% silt, and 69% clay in the top 0–20 cm and gradually increasing with soil depth to 78 % clay at 20–100 cm. The clay fraction is mainly made of kaolinite (Hok et al., 2015). The land on the site is flat, the land slope is < 1%. Prior to the establishment of the three experiments in 2009, the average SOC and TN stocks in the 0-20 cm layer were 33.6 Mg C ha<sup>-1</sup> and 3.33 Mg N ha<sup>-1</sup>. respectively. The research site's climate is defined as tropical monsoon, corresponding to the Am Köppen climate classification, with two main seasons: the wet season from May to October and the dry season from November to April. The mean annual temperature from 2009–2021 was 27.5°C, while the average annual minimum and maximum temperatures were 22°C and 35°C, respectively. The annual rainfall from 2009–2021 ranged between 1,650 and 2,000 mm.

196

197

198

199

200

201

202

203

204

205

206

207

208

209

210

211

212

213

214

## 2.2 Experimental design, treatment description, and crop management

216

217

218

219

220

221

222

223

224

225

226

227

228

229

230

231

232

233

234

235

236

The detailed history of the research site, experimental design, treatment description, and fertilizer application were reported in Hok et al., (2015) and Pheap et al., (2019). Our study covers three separate experiments, implemented in 2009, including (i) maize (Zea mays L.) (which was a former rice (Oryza sativa L.)-based trial from 2009 to 2019 and shifted to maize-based trial in 2020), (ii) soybean, and (iii) cassava (Manihot esculenta Crantz)-based cropping system trials, hereafter called MaiEx, SoyEx, and CasEx, respectively. These represent the most important annual upland crops in Cambodia as well as in some Southeast Asian countries. Each experiment is arranged in a randomized complete block design with three replicates. The elementary plot dimensions are 8 m x 37.5 m, equivalent to 300 m<sup>2</sup>. Each experiment consists of three treatments including: (1) monocropping under conventional tillage (CTM), in which the main crops, i.e., maize (Mz), sovbean (Sb), or cassava (Cs), were monocropped with land preparation done by disc ploughing (CTM-Mz, CTM-Sb, and CTM-Cs); (2) monocropping under NT systems with the use of cover crops (NTM), in which the main crops (maize, soybean, or cassava) were cropped in a one-year frequency pattern with no soil tillage and with addition of cover crops (NTM-Mz, NTM-Sb, and NTM-Cs); and (3) bi-annual rotation of the main crops under NT systems with the use of cover crops (NTR), where the main crops were presented every year in two separate elementary plots designated as NTR1 and NTR2. For treatment (3) of SoyEx and CasEx, represented by NTR1-Sb and NTR1-Cs, respectively, the main crops (i.e., soybean and cassava) were grown in a bi-annual rotation with maize, represented by NTR2-Sb and NTR2-Cs for SoyEx and CasEx, respectively. For the treatment (3) of Mai-Ex, the main crop (i.e., maize represented by NTR1-Mz) was grown in a bi-annual crop rotation with soybean, represented by NTR2-Mz (Table 1). Under all the NT systems, the species, sowing dates,

and methods of cover crop establishment varied depending on the design of treatments for each experiment, the types and cycles of the main crops, and the species and cycles of the cover crops (Table 1). For instance, stylo (Stylosanthes guianensis (Aublet) Sw.) and Brachiaria (Brachiaria ruziziensis R.Germ. & C.M.Evrard) were associated with rice and soybean, respectively, by manual broadcasting at the full flowering stage of rice before the end of September and at the first yellow leaves of soybean in the mid of October. Stylo was associated by line sowing with a NT planter at the same date of maize cultivation and 20 days after planting for cassava. In addition, if the development and/or density of the cover crop sown the previous year were considered insufficient, short cycle cover crop species, i.e., pearl millet (Pennisetum typhoides (L.) Morrone) or sorghum (Sorghum bicolor (L.) Moench), was sown alone or mixed with cowpea (Vigna unguiculata (L.) Walp. and sunnhemp (Crotalaria juncea L.) at the beginning of the rainy season (in the first week of May). Cover crops were then grown for 60–75 days to increase the biomass inputs prior to the cultivation of the main cycle of rice, soybean, or maize (Table 1). The establishment and harvest of the main crops varied depending on the species. For maize, upland rice, and soybean, with a life cycle of approximately 110–120 days, these crops were mainly seeded between the last week of June to mid-July and harvested between mid-October and mid-November, whereas cassava was planted in early May and harvested around 10 months old in the mid-February of the following year. For main crop residue management in MaiEx and SoyEx, all crop residues were retained in the soil in all the tillage systems. In CasEx, under CTM-Cs, all the cassava fallen leaves and branches were retained in the soil, while 100% of the cassava main stems and original cuttings were completely removed from the plot after harvest, representing standard farmers' practices. For all the NT-Cs systems, all the cassava fallen leaves and branches were returned to the soil, while 50% of the cassava main stems

237

238

239

240

241

242

243

244

245

246

247

248

249

250

251

252

253

254

255

256

and 100% of the original cuttings were retained in the soil and then crimped to speed up the decomposition process and facilitate field operation implementations in the following cropping season. The residues of all the cover crops were left as mulch under all the NT systems in all the experiments. There were a few adjustments on the cropping systems over the experimental period from 2009 to 2021, especially with the use of cover crops, crop varieties, and mineral fertilizer application types and rates. Details of main crops and cover crop successions and the cumulative amount of the aboveground biomass-C and N inputs from the crop residues are presented in Table 1. The C inputs were estimated from the dry aboveground biomass inputs recorded prior to the termination of the cover crops and at grains/tuber harvest for the main crops. The C and N inputs from the root systems were not recorded nor even estimated based on literature. In the case of missing data of aboveground biomass, the amount of biomass was estimated using the average of recorded data over time as reference in the case of cover crops and grain and/or tuber-aboveground biomass ratio for the main crops (e.g. rice, maize, soybean and cassava). In addition, the cumulative and annual N inputs were estimated from the amount of the cumulative and annual C inputs, respectively, by applying available C/N ratio values of each plant species that were obtained from the C and N concentration analysis by dry combustion. For land preparation, the CTM plots were ploughed twice to 15-20 cm depth using a 7-disc plough after early rains at the beginning of the wet season and then before the main crop cultivation. If sufficient early rain falls were received at the beginning of the wet season (in the 3<sup>rd</sup> week of March), sesame and mung bean were sown manually under CTM treatment in SoyEx and MaiEx, respectively, as early-cycle cash crops (April to June) prior to the main crops, i.e., soybean or maize (from July to November). If that was

258

259

260

261

262

263

264

265

266

267

268

269

270

271

272

273

274

275

276

277

278

not the case, the CTM plots remained fallow with the growth of natural grasses and broad leaves until the

main cycle crops. These cropping systems represent the standard farmers' practices. Under the NT systems (NTM, NTR1 and NTR2), a long cycle cover crop i.e., stylo was used as a cover crop and grown in association with the main crops. This cover crop was sown in the middle of the inter-row at 0, 15, 35 days after the sowing of the main crops, i.e., maize, cassava, rice, respectively, and by seed broadcasting at the first yellow leaves of soybean, approximately 4 weeks before harvest. In addition, if the development and/or density of the cover crop sown the previous year was considered insufficient, pearl millet or sorghum was sown alone for the treatments planted with soybean or mixed with sunnhemp and cowpea for the treatments planted with maize at the beginning of the rainy season as short-cycle cover crops. The cover crops were then grown for 60–75 days prior to the main cycle of rice, maize, or soybean. The main crops (rice, maize and soybean), both under CTM and all the NT systems, as well as the cover crops (at the beginning of the rainy season) were sown by a NT planter (Fitarelli pulled by power tiller, Vence Tudo, or Seamato lifted or pulled by tractor). From 2009 to 2020, cassava was planted along the furrows drawn by chiselling at 0.8 m spacing to approximately 20 cm depth, and then it was planted by a NT cassava planter (Planticenter) in 2021. Under the NT systems, the cover crops were terminated by crimping followed by the application of a mix of non-selective herbicides, i.e., glyphosate [N (phosphonomethyl) glycine] and 2,4-D [2,4-dichloro-phenoxyacetic acid], at a rate of 960 and 720 g active ingredient (a.i) ha<sup>-1</sup>, respectively. Since 2009, soil amendment was done with thermophosphate (16% P<sub>2</sub>O<sub>5</sub>, 31% CaO and 16% MgO) at the end of dry season (early April), and then basal fertilizers and top dressings on the main crops were applied with different rates of N, P, K depending on the types and phenological stage of each main crops using diammonium phosphate (18% N, 46% P<sub>2</sub>O<sub>5</sub>), ammonium phosphate (16% N, 20% P<sub>2</sub>O<sub>5</sub>), potassium

279

280

281

282

283

284

285

286

287

288

289

290

291

292

293

294

295

296

297

298

300 chloride (60%  $K_2O$ ), and urea (46% N). The application of the fertilizer inputs to each main crop are detailed in Table S1.

**Table 1.** Experiments, cropping systems and crop sequences, and associated cumulative and annual aboveground C and organic N inputs from crop residues during the experimental period (2009-2021).

Experiments and	Crop sequences from 2009 to 2021 <sup>b</sup>	C input (M	Ig ha <sup>-1</sup> )	N input (Mg ha <sup>-1</sup> )		
cropping systems <sup>a</sup>		Cumulative	Annual	Cumulative	Annual	
MaiEx						
CTM-Mz	$R - \underline{Mu}/R - \underline{Mu}/R - \underline{Mu}/R - \underline{Mu}/R - R - R - R - R - R - R - Mz - Mz$	28.60	2.20	0.64	0.05	
NTM-Mz	$Mi/\mathbf{R} - Mi/\mathbf{R} - Mi/\mathbf{R} - St/\mathbf{R} - St/\mathbf$	67.70	5.21	1.50	0.12	
NTR1-Mz	$Mi/\mathbf{R} - Mi + Su + St/\mathbf{Mz} - Mi + Su + St/\mathbf{R} - St/\mathbf{Mz} + St - St/\mathbf{R} + St - St/\mathbf{Mz} + St - St/\mathbf{Rz} + St - St/\mathbf{Mz} + St - St/\mathbf{Mz} + St - St/\mathbf{Mz} + St - St/\mathbf{So} + Su + \mathbf{R} - Mi + Su + Co/\mathbf{Mz} - So + Su + Co/\mathbf{Sb}$	73.08	5.62	1.62	0.12	
NTR2-Mz	$Mi/\mathbf{Mz} - Mi+P+St/\mathbf{R} - Mi+Su+St/\mathbf{Mz}+St - St/\mathbf{R}+St - St/\mathbf{Mz}+St - St/\mathbf{Rz}+St - St/\mathbf{Rz}+St$	70.12	5.39	1.56	0.12	
SoyEx						
CTM-Sb	$Sb-Sb-\underline{Se}/Sb-Sb-\underline{Se}/Sb-Sb-Sb-Sb-Sb-Sb-Sb-Sb-Sb-Sb$	23.18	1.78	0.52	0.04	
NTM-Sb	$Mi/\mathbf{Sb}+Br-Mi/\mathbf{Sb}+Br-Mi/\mathbf{Sb}+Br-Mi/\mathbf{Sb}+St-St/\mathbf{Sb}+St-St/\mathbf{Sb}+St-St/\mathbf{Sb}+St-St/\mathbf{Sb}+St-St/\mathbf{Sb}+St-St/\mathbf{Sb}+St-St/\mathbf{Sb}+St-St/\mathbf{Sb}+St-St/\mathbf{Sb}-S$	65.09	5.01	1.45	0.11	
NTR1-Sb	$Mi/\mathbf{Sb}+St-Mi+St/\mathbf{Mz}+Br-Mi/\mathbf{Sb}+St-Mi+Su/\mathbf{Mz}-So/\mathbf{Sb}+St-So+Su/\mathbf{Mz}-So+St/\mathbf{Sb}-So+Su/\mathbf{Mz}+St-St/\mathbf{Rb}-\mathbf{Rb}+So/\mathbf{Mz}-So+Su/\mathbf{Sb}-Mi+Su+Co/\mathbf{Mz}-So+Su/\mathbf{Sb}$	71.13	5.47	1.58	0.12	
NTR2-Sb	$Mi/\mathbf{Mz}+Br-Mi/\mathbf{Sb}+St-Mi+Su/\mathbf{Mz}+St-St/\mathbf{Sb}+St-So+Su/\mathbf{Mz}-So+Su/\mathbf{Sb}-So+Su/\mathbf{Mz}-So+Su/\mathbf{Sb}-So+Su/\mathbf{Mz}-So+Su/\mathbf{Sb}-So+Su/\mathbf{Mz}-So+Su+Co/\mathbf{Sb}-Mi+Su+Co/\mathbf{Mz}$	78.94	6.07	1.75	0.13	
CasEx						
CTM-Cs	Cs-Cs-Cs-Cs-Cs-Cs-Cs-Cs-Cs-Cs-Cs	17.64	1.36	0.39	0.03	
NTM-Cs	$\mathbf{Cs} + St - St/\mathbf{Cs} + St - St/\mathbf{Cs} + St - St/\mathbf{Cs} + St - \mathbf{Cs} + \mathbf$	46.92	3.61	1.04	0.08	
NTR1-Cs	$\mathbf{C}\mathbf{s}+St-Mi+\mathbf{M}\mathbf{z}+St-St/\mathbf{C}\mathbf{s}+St-Mi+Su/\mathbf{M}\mathbf{z}+St-St/\mathbf{C}\mathbf{s}+St-St/\mathbf{M}\mathbf{z}+St-St/\mathbf{C}\mathbf{s}-St/\mathbf{M}\mathbf{z}+St-St/\mathbf{C}\mathbf{s}-So+Su/\mathbf{M}\mathbf{z}-So+Su/\mathbf{C}\mathbf{s}-Mi+Su+Co/\mathbf{M}\mathbf{z}-\mathbf{C}\mathbf{s}$	64.25	4.94	1.43	0.11	
NTR2-Cs	$Mi/\mathbf{Mz}+St-St/\mathbf{Cs}+St-Mi+Su/\mathbf{Mz}+St-St/\mathbf{Cs}+St-Mi+Su/\mathbf{Mz}+St-St/\mathbf{Cs}+St-St/\mathbf{Mz}+St-St/\mathbf{Cs}-So+Su/\mathbf{Mz}-So+Su/\mathbf{Cs}-So+Su/\mathbf{Mz}-Cs-Mi+Su+Co/\mathbf{Mz}$	67.10	5.16	1.49	0.11	

<sup>a</sup>MaiEx: maize-based cropping system trial; SoyEx: soybean-based cropping system trial; and CasEx: cassava-based cropping system trial; CTM: monocropping under conventional tillage; NTM: monocropping under NT systems with notill mulch-based cropping systems associated with different crop sequences, and NTR1 and NTR2 refer to bi-annual crop rotational systems under NT systems with no-till mulch-based cropping systems associated with different crop sequences.

<sup>b</sup>Br: brachiaria (*Brachiaria ruziziensis* R.Germ. & C.M.Evrard); Co: cowpea (*Vigna unguiculata* (L.) Walp.); Cs: cassava (*Manihot esculenta* Crantz); Mi: millet (*Pennisetum glaucum* (L.) R.Br.); Mu: Mungbean (*Vigna radiata* (L.) R. Wilczek) Mz: maize (*Zea mays* L.); P: pigeon pea (*Cajanus cajan* L.); R: rice (*Oryza sativa* L.); Rb: ricebean (*Vigna umbellata* (Thunb.) Ohwi & H. Ohashi); Sb: soybean (*Glycine max* L.); Se: sesame (*Sesame indicum* L.); So: Sorghum (*Sorghum bicolor* (L.) Moench); St: stylo (*Stylosanthes quianensis* (Aubl.) Sw.); Su: sunnhemp (*Crotalaria juncea* L.). The letters in bold, underlined, and italicized indicate the main crops, cash crops, and cover crops, respectively. "–" indicates the period between the year; "/" indicates relay cropping with varying planting dates; and "+" indicates crops planted in association (same or staggered sowing dates). The C inputs were estimated from the amount of aboveground biomass of each crop; the belowground biomass was not included.

# 2.3 Soil sampling and processing

302

The study was a diachronic analysis from 2011 to 2021. In 2009, prior to the establishment of the 303 304 experiments, soil and bulk density (pb) samples were collected as the pre-experiment (PE) from three 305 randomly selected sampling points per replicate of each experimental location at four depths: 0-5, 5-10, 306 10–20, and 20–30 cm. The individual soil samples from the same depth and replicate were composited, 307 resulting in three composites per depth and per experiment. The composite samples were oven dried at 308 40°C and sieved through a 2mm mesh for chemical property analysis. Bulk density samples were 309 collected using core samplers of 5 cm in diameter and 5 cm in height, oven dried at 105°C for 48 h. The 310 SOC and TN stocks of PE in 2009 in the top 0–20 cm were 33.3, 35.0, and 32.4 Mg C ha<sup>-1</sup> and 3.34, 3.41, and 3.26 Mg N ha<sup>-1</sup> in MaiEx, SoyEx, and CasEx, respectively (Fig. S1). 311 312 In November 2011, soil sampling was conducted to assess the early effects of tillage systems on soil 313 organic C and N concentrations and stocks. The details of the sampling are described in Hok et al., (2015). 314 Briefly, two pits (1m x 1m) were opened per elementary plot for soil and ob sample collection. Individual samples for chemical analysis were collected from two undisturbed sides of each pit at 0-5, 5-10, 10-20, 315 20-40, 40-60, 60-80, and 80-100 cm. The individual samples from the same depth and the same pit were 316 317 composited, and as a result, two composite samples per layer were collected per elementary plot. The 318 composites were oven-dried at 40°C before being softly disrupted, sieved through a 2-mm sieve, and 319 homogenized. Soil bulk density samples were taken from the same two undisturbed sides of each pit at 320 the same soil depths as for SOC analysis using core samplers of 5 cm in diameter and 5 cm in height. The 321 soil cores were oven-dried at 105°C for 48 h.

In December 2021, we re-sampled the soil to assess the changes in ρb, SOC and N concentrations and stocks ten years after the study conducted by Hok et al., (2015). The samples were collected at the same 7 layers. From each treatment and replicate, we collected four individual samples by an automatic soil column cylinder auger (a gasoline-powered percussion hammer Cobra TT with inner diameter of 85 mm, Eijkelkamp, the Netherlands) in a diagonal "X" shape from four points within each plot, avoiding overlapping on the pits opened in 2011. In addition, we dug a 1 m x 1 m pit in the middle of each plot for sample collection; three individual soil samples and three ρb cores were collected from three undisturbed sides of the pit at each depth. Soil samples were air-dried at room temperature, gently broken down, and sieved through a 2 mm mesh sieve. Finally, the seven individual samples from the same layer were mixed and homogenized to make a composite sample per elementary plot. The samples of ρb were oven-dried at 105°C for 48 h.

# 2.4 Soil organic C and total N analyses

The concentrations of SOC and TN of the soil samples collected in 2009 and 2011 were determined by dry combustion using an elemental CN analyzer (TruSpec CN, LECO, St. Joseph, USA). The details of the analysis were described in Hok et al., (2015). Sub-samples of the composite soils (n = 3 per layer) collected in 2021 were finely ground (<150 µm) before analysis for total C and N by dry combustion using the LECO<sup>®</sup> CHN628 analyzer at the Sustainable Agroecosystems Lab, ETH Zurich University, Switzerland.

# 2.5 Soil organic C and total N stocks calculation

340

- 341 In this Oxisol, there was no coarse fraction (i.e. gravels) left after sieving at 2 mm. Therefore, the bulk 342 density of soil equals the bulk density of fine earth. To avoid inaccurate stock calculation due to 343 differences in bulk density between treatments when using the fixed depth method, the equivalent soil 344 mass (ESM) approach was applied to compute SOC and TN stocks (Ellert and Bettany, 1995; Von Haden 345 et al., 2020; Fowler et al., 2023). Since the pb of the treatments differed between the two sampling years 346 (2011 and 2021) at each sampling depth (Table S2), we defined the reference soil mass as the lowest soil 347 mass observed at each sampling depth, regardless of sampling years, cropping systems or land use. For this reference, soil mass layers (480, 518, 1061, 1873, 1766, 1809, and 1779 Mg ha<sup>-1</sup>) corresponded to 348 349 the depth layers (0-5, 5-10, 10-20, 20-40, 40-60, 60-80, and 80-100 cm, respectively). We applied 350 these reference soil masses to compute the SOC and TN stocks in 2021 and recalculated the stocks of the 351 PE and the treatments of the three experiments collected in 2009 and 2011.
- To correct for differences in pb, SOC and TN stocks were computed according to Eq. 1 and 2.

353 
$$M_{(Soilmin,i)} = \rho b_{(i)} \times T_{(i)} \times 1000$$
 (eq. 1)

354 
$$SOC \ or \ TN \ stock = \sum (i=1)^n n \left[ \left( M_{(soilmin,i)} \times \ conc._{(i)} \right) + \left( \left( M_{(soil,i)} - M_{(soilmin,i)} \right) \times \ conc._{(i-1)} \right) \right] \times discorrection$$

Where:  $M_{(soilmin,i)}$  is the minimal soil mass per unit area in the ith layer (Mg ha<sup>-1</sup>) recorded over the treatments and used as a reference.  $\rho b_{(i)}$  is the bulk density of the ith layer (g cm<sup>-3</sup>).  $T_{(i)}$  is the thickness of the ith layer (m).  $conc._{(i)}$  is the concentration of SOC in ith layer.  $conc._{(i-1)}$  is the concentration of

- SOC in i 1th layer.  $M_{(soil.i)}$  is the designated soil mass of each layer (i.e., the maximum soil mass). The
- numbers 1000 and 0.001 are unit conversion coefficients.
- We defined delta stock ( $\Delta$ ) of SOC and TN, as the stock change within the same treatment and depth
- between 2021 and 2011 sampling years (diachronic) and calculated it as follows:
- 363  $\Delta SOC$  or TN stock. diach = SOC or TN stock<sub>treat.(i)2021</sub> SOC or TN stock<sub>treat.(i)2011</sub> (eq. 3)
- 364 Where: i represents the treatments.
- 365 To compare the synchronic and diachronic approaches for SOC stock change, the stock change estimated
- 366 by the synchronic approach was computed as follows using the CTM treatment as the control treatment:
- 367  $\Delta SOCstock.synch = SOCstock_{NT(i)2021} SOCstock_{CTM2021}$  (eq. 4)
- 368 Where: NT(i) represents NTM, NTR1, and NTR2 treatments.
- 369 The SOC and TN stock change (accumulation or loss) rates (Mg C or N ha<sup>-1</sup> yr<sup>-1</sup>) of each treatment were
- 370 calculated by dividing  $\Delta$  SOC or TN stock by the number of years between the 1<sup>st</sup> and 2<sup>nd</sup> samplings (10
- 371 years):

372 SOC or TN stock change 
$$rate_{treatment(i)} = \frac{\Delta SOC \text{ or TN } stock_{treatment(i)}}{10}$$
 (eq. 5)

# 373 **2.6 Particle-size fractionation of soil organic matter**

- 374 The soil organic C was physically fractionated using a sub-sample of the composite soil for all the
- 375 treatments and seven depths. The particle-size fractionation was implemented in accordance with the
- 376 procedure described in Hok et al., (2015). Briefly, 40 g of soil samples were dispersed in a solution of
- 377 1.25 g of sodium hexametaphosphate and 100 mL of deionized water and stored at 10 °C for 16 hours.
- 378 The sample was then horizontally shaken at 100 rpm for 8 hours with three 10-mm-diameter agate balls.

The soil mixture was wet-sieved with deionized water through a 53-µm sieve to get the proportion of particulate organic matter (POM) sized between 53µm and 2,000µm. The <53-µm fraction was flocculated with 2-g CaCl<sub>2</sub> in a 1-L glass cylinder and left overnight for sedimentation. The supernatant was syphoned after full sedimentation. This <53-µm fraction is made up of mineral-associated organic matter (MAOM). The two fractions were oven-dried at 40°C until reaching constant weight and finely ground for determining SOC and TN concentrations by dry combustion using the LECO® CHN628 analyzer at the Sustainable Agroecosystems Lab, ETH Zurich University, Switzerland.

# 2.7 Statistical analysis

Statistical analysis was conducted using R software, version 4.3.1 (Core Team, 2023). Linear mixed models (*ImerTest* package) were fitted on all data: sampling years, soil depths, and treatments were defined as the fixed factors, while the replicates were defined as the random factors. Prior to the analysis, the normality of each variable was checked by Shapiro's test. Then we applied Levene's test to check the homoscedasticity of the data. A diachronic approach was used to assess statistical significance between the two sampling years (i.e., 2021 vs. 2011) of the same treatment at the same soil depth by analysis of variance with Fisher tests (degrees of freedom calculated by Satterthwaite method), computing of estimated marginal means (EMMs) and p-value adjustment using the Tukey method. The same approach was used to compare SOC stocks between treatments in the same sampling year, calculated at equivalent soil mass and using the synchronic approach. In addition, we applied the same statistical procedures to assess the statistical significance of cumulative SOC and TN stocks.

## 398 **3 Results**

- 399 The effects of cropping systems on the concentrations, stocks of SOC and TN as well as their fractions in
- 400 the physical size classes between 2011 and 2021 varied among the three experiments and across the soil
- 401 profile.

403

# 402 3.1 Impact of cropping systems on SOC concentration and stock

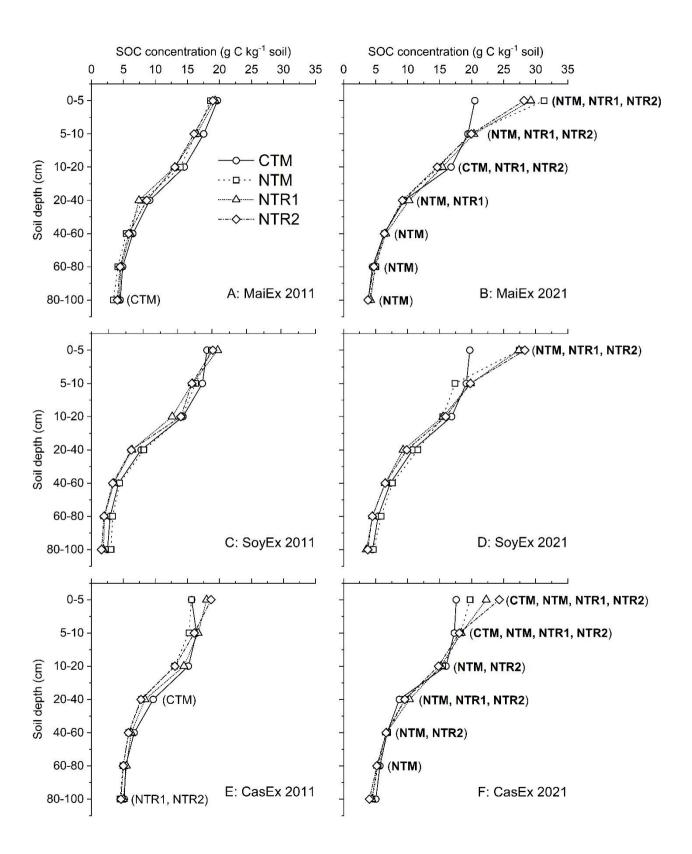
## 3.1.1 SOC concentration

- 404 The SOC concentration of all the treatments in 2011 and 2021 was highest in the topsoil (0–5 cm) and
- decreased with soil depth in all the experiments (Fig.1 with Table S3 as duplication).
- 406 Over a 10-year period of historical cropping sequences from 2011 to 2021, all the NT cropping systems
- 407 had significant effects (p < 0.05) in the increase of SOC concentration. On the other hand, the SOC
- 408 concentration under CTM remained stable with exception of a few significant increases detected in the
- 409 tilled layers in MaiEx and CasEx (Fig. 1).
- 410 In 2021, the monocropping of main crops under NT systems in MaiEx (NTM-Mz) and CasEx (NTM-Cs)
- 411 exhibited a similar trend in increasing SOC concentration significantly (p < 0.05) across the soil profile
- 412 compared to 2011 (Figs. 1B and 1F). The SOC concentration under NTM-Mz increased by 68%, 21%,
- 413 16%, 17%, 23%, and 16% at 0–5, 5–10, 20–40, 40–60, 60–80, and 80–100 cm, respectively (Fig. 1B).
- The significant increase in SOC concentration under NTM-Cs was detected from 0 to 80 cm with a gain
- 415 of 26%, 20%, 19%, 22%, 18%, and 10% in the 0-5, 5-10, 10-20, 20-40, 40-60, and 60-80 cm,
- 416 respectively (Fig. 1F).

417 When compared to 2011, the bi-annual rotation of main crops under NT systems in MaiEx (NTR1-Mz and NTR2-Mz) and CasEx (NTR1-Cs and NTR2-Cs) significantly (p < 0.05) increased SOC 418 419 concentration from the surface down to subsoil depth in 2021 (Figs. 1B and 1F). On average NTR-Mz 420 (average of NTR1-Mz and NTR2-Mz), significantly increased SOC concentration by 50%, 24%, and 15% 421 at 0-5, 5-10, and 10-20 cm depth, respectively. Significant increase was still observed under NTR1-Mz 422 at 20-40 cm depth (Fig. 1B). In 2021, among the two treatments of NTR-Cs crop rotation systems (NTR1-423 Cs and NTR2-Cs), NTR2-Cs significantly increased SOC concentration from the top 0 to 60 cm by 30%, 424 12%, 13%, 23%, and 15% in the 0-5, 5-10, 10-20, 20-40, and 40-60 cm, respectively, while a significant 425 decrease of -13% was recorded in the 80–100 cm depth (Figs. 1E and 1F). Under NTR1-Cs, significant 426 increases in SOC concentration were observed up to 40 cm depth with a gain of 24%, 10%, and 23% in 427 0-5, 5-10, and 20-40 cm, respectively, with a significant decrease by -12% in 80-100 cm depth (Figs. 428 1E and 1F). 429 Unlike MaiEx and CasEx, the significant increase (p < 0.05) in SOC concentration under all the NT 430 cropping systems in SoyEx (NTM-Sb, NTR1-Sb, and NTR2-Sb) in 2021 was only observed in the top 0– 431 5 cm with a similar increase amount of ~7.5 g C kg<sup>-1</sup> soil (Fig 1D). 432 Over a decade of monocropping of main crops under conventional tillage in all the experiments (CTM-433 Mz, CTM-Sb, and CTM-Cs), the SOC concentration remained stable overall, with the exception of a few 434 significant increases detected in the tilled layers in MaiEx (CTM-Mz at 10–20 cm) and CasEx (CTM-Cs

435

at 0–5 and 5–10 cm) (Figs. 1B, 1D, and 1F).



**Figure 1.** SOC concentration distribution across the soil profile (0-100 cm) of the treatments under different experiments in 2011 and 2021. CTM: monocropping under conventional tillage; NTM: monocropping under NT systems with no-till mulch-based cropping systems associated with different crop sequences, and NTR1 and NTR2 refer to bi-annual crop rotational systems under NT systems with no-till mulch-based cropping systems associated with different crop sequences as described in Table 1. A: MaiEx 2011 - SOC concentration of the treatments in maize-based trial measured in 2011; B: MaiEx 2021 - SOC concentration of the treatments in soybean-based trial measured in 2021; C: SoyEx 2011 - SOC concentration of the treatments in soybean-based trial measured in 2021; E: CasEx 2011 - SOC concentration of the treatments in cassava-based trial measured in 2021; and F: CasEx 2021 - SOC concentration of the treatments in cassava-based trial measured in 2021; Treatment(s) in bold within the brackets indicate the gain and significant (p < 0.05) difference in concentrations between 2011 and 2021 in the same treatment at the same soil depth.

## **3.1.2 SOC stock**

436

437 From 2011 to 2021, there were significant (p < 0.05) increases in SOC stock, which varied depending on 438 tillage, cropping systems, and the experiments (Table 2 and Table S5). In 2021, in the case of MaiEx, the SOC stock under NTR-Mz crop rotation systems (average of NTR1-Mz and NTR2-Mz) significantly (p 439 440 < 0.05) increased by 4.6, 2.6, and 2.2 Mg C ha<sup>-1</sup> at 0–5, 5–10, and 10–20 cm depth, respectively. NTR1-Mz showed a significant increase in SOC stock at a deeper profile at 20–40 cm, with a gain of 4.5 Mg C 441 ha<sup>-1</sup> (Table 2 and Table S5). In the case of CasEx, the soils under NTR-Cs crop rotation systems (average 442 443 of NTR1-Cs and NTR2-Cs) in 2021 significantly (p < 0.05) increased SOC stock by an average of 2.4, 1.1, 1.4, and 2.9 Mg C ha<sup>-1</sup> in 0–5, 5–10, 10–20, and 20–40 cm, respectively, but significantly decreased 444 by an average of -0.9 Mg C ha<sup>-1</sup> in 80–100 cm (Table 2 and Table S5). For SoyEx, NTR-Sb crop rotation 445 446 systems (average of NTR1-Sb and NTR2-Sb) significantly accumulated SOC stock by an average of 3.55 and 1.75 Mg C ha<sup>-1</sup> in 0–5 and 5–10 cm, respectively, along with a positive trend from 10 to 80 cm depth 447 448 (Table 2 and Table S5).

Unlike SOC concentration, the significant effect of increasing SOC stock of the monocropping of main 449 crops under NT systems varied across the three experiments (Table 2). NTM-Cs showed the significant 450 (p < 0.05) accumulation of SOC stock by 2.0, 1.6, 2.5, and 2.2 Mg C ha<sup>-1</sup> in 0–5, 5–10, 10–20, and 20– 451 452 40 cm, respectively (Table 2 and Table S5). NTM-Mz significantly increased SOC stock in the surface layers (0-10 cm) by 6.1 and 2.2 Mg C ha<sup>-1</sup> in the 0-5 and 5-10 cm, respectively, while the NTM-Sb 453 significantly increased the stock by 3.6 Mg C ha<sup>-1</sup> in the 0–5 cm (Table 2 and Table S5). 454 455 In the case of monocropping of main crops under conventional tillage, despite there were a few significant increases in SOC stock were detected in the till layers in CTM-Mz (MaiEx) by 0.9 and 2.3 Mg C ha<sup>-1</sup> in 456 457 the 5–10 and 10–20 cm, respectively, the significant decline in SOC stock was observed below 60 cm with the decreasing by approximately -1.2 Mg C ha<sup>-1</sup> at the 60–80 and 80–100 cm depth (Table 2). For 458 459 CasEx and SoyEx, despite the significant increase in SOC stock was observed in CTM-Cs from 0 to 20 460 cm, the accumulation rate was 2 times lower than those NT systems, while no significant changes were recorded under CTM-Sb (Table 2). 461 Over the 10-year period from 2011 to 2021, considering a 100 cm layer as a single stratum, all the NT 462 463 cropping systems significantly increased SOC stock, with accumulation rates ranging from 0.86 to 1.47 and 0.70 to 1.07 Mg C ha<sup>-1</sup> yr<sup>-1</sup> Mg C ha<sup>-1</sup> yr<sup>-1</sup> for CA-Mz and CA-Cs, respectively (Table 2). Although 464 non-significant difference detected, all the NT-Sb systems increased SOC stock with annual accumulation 465 rates ranging from 0.65 to 1.00 Mg C ha<sup>-1</sup> yr<sup>-1</sup> (Table 2). Despite there were a few significant increases in 466 467 SOC stock were observed under CTM, the whole profile SOC stock in all the CTM (CTM-Mz, CTM-Sb, 468 and CTM-Cs) remained stable in 2021 (Table 2).

**Table 2.** Mean SOC stocks in 2021 and SOC stock change rate between 2021 and 2011.

Experiments <sup>a</sup>		Cropping systems <sup>b</sup>									
		CTM		NTM	NTR1	NTR2	CTM	NTM	NTR1	NTR2	
		SOC stock in 2	202	21 (Mg C ha <sup>-1</sup> )		SOC stock change rate 2021-11 (Mg C ha <sup>-1</sup> yr <sup>-1</sup> )					
MaiEx	0-5	9.83 (±0.19)		15.01 (±0.38) A	14.03 (±0.32) A	13.52 (±0.54) A	0.04 (±0.03)	<b>0.61</b> (±0.06)	<b>0.48</b> (±0.02)	<b>0.44</b> (±0.02)	
	5-10	10.16 (±0.18)	4	10.91 (±0.11) A	11.16 (±0.06) A	10.89 (±0.46) A	<b>0.09</b> (±0.03)	<b>0.22</b> (±0.01)	<b>0.23</b> (±0.00)	<b>0.22</b> (±0.06)	
	10-20	18.28 (±0.26)	4	16.31 (±0.57)	17.08 (±0.07) A	16.32 (±0.69) A	<b>0.23</b> (±0.02)	$0.11~(\pm 0.06)$	<b>0.25</b> (±0.04)	<b>0.19</b> (±0.09)	
	20-40	19.90 (±0.72)		18.61 (±0.43)	20.28 (±0.70) A	18.64 (±0.58)	0.11 (±0.06)	$0.16~(\pm 0.07)$	<b>0.45</b> (±0.08)	0.07 (±0.08)	
	40-60	12.58 (±0.37)		12.18 (±0.25)	12.75 (±0.42)	12.31 (±0.19)	-0.04 (±0.03)	$0.08~(\pm 0.08)$	$0.10~(\pm 0.03)$	0.02 (±0.05)	
	60-80	9.01 (±0.61) I	3	9.58 (±0.31)	9.64 (±0.48)	9.33 (±0.26)	<b>-0.12</b> (±0.02)	$0.07~(\pm 0.08)$	-0.02 (±0.01)	-0.02 (±0.01	
	80-100	7.17 (±0.66) I	3	7.51 (±0.54)	7.97 (±0.38)	$7.26~(\pm 0.05)$	<b>-0.13</b> (±0.05)	$0.05~(\pm 0.08)$	-0.02 (±0.02)	-0.06 (±0.05	
	0-100	86.92 (±2.18)		90.13 (±1.82) A	92.90 (±2.30) A	88.27 (±2.11) A	$0.17~(\pm 0.15)$	<b>1.30</b> (±0.38)	<b>1.47</b> (±0.11)	<b>0.86</b> (±0.20)	
SoyEx	0-5	9.45 (±.0.05)		13.17 (±0.19) A	13.14 (±0.30) A	13.58 (±0.31) A	0.02 (±0.03)	<b>0.36</b> (±0.02)	<b>0.32</b> (±0.02)	<b>0.39</b> (±0.05)	
	5-10	9.95 (±0.10)		9.74 (±1.45)	11.05 (±0.12) A	10.80 (±0.21) A	0.03 (±0.03)	0.04 (±0.12)	<b>0.18</b> (±0.02)	<b>0.17</b> (±0.02)	
	10-20	18.35 (±0.12)		16.79 (±1.84)	17.27 (±0.35)	17.43 (±0.43)	0.11 (±0.05)	-0.02 (±0.16)	0.17 (±0.04)	0.08 (±0.07)	
	20-40	22.34 (±0.50)		22.66 (±1.18)	19.26 (±0.43)	19.69 (±0.75)	0.19 (±0.18)	0.17 (±0.17)	0.19 (±0.07)	0.18 (±0.13)	
	40-60	15.20 (±0.61)		15.86 (±0.55)	13.00 (±0.65)	13.08 (±0.43)	0.14 (±0.18)	0.10 (±0.10)	0.13 (±0.03)	0.17 (±0.06)	
	60-80	11.16 (±0.52)		11.95 (±0.62)	9.48 (±0.32)	9.32 (±0.44)	0.04 (±0.14)	0.06 (±0.04)	0.01 (±0.04)	0.04 (±0.06)	
	80-100	8.76 (±0.57)		9.37 (±0.82)	7.03 (±0.04)	7.25 (±0.37)	-0.04 (±0.11)	-0.05 (±0.01)	-0.08 (±0.05)	-0.02 (±0.02	
	0-100	95.22 (±2.06)		99.53 (±4.26)	90.23 (±1.52)	91.15 (±1.87)	$0.48~(\pm 0.70)$	$0.65~(\pm 0.58)$	0.92 (±0.13)	1.00 (±0.29)	
CasEx	0-5	8.44 (±0.23)	A	9.48 (±0.09) A	10.68 (±0.00) A	11.67 (±0.34) A	<b>0.09</b> (±0.01)	<b>0.20</b> (±0.03)	<b>0.21</b> (±0.01)	<b>0.27</b> (±0.06)	
	5-10	8.97 (±0.25)	A	9.52 (±0.17) A	9.82 (±0.04) A	9.76 (±0.22) A	<b>0.05</b> (±0.01)	<b>0.16</b> (±0.02)	<b>0.10</b> (±0.00)	<b>0.12</b> (±0.03)	
	10-20	17.15 (±0.30)	A	16.66 (±0.49) A	16.63 (±0.15) A	16.23 (±0.24) A	<b>0.09</b> (±0.03)	<b>0.25</b> (±0.03)	<b>0.09</b> (±0.01)	<b>0.18</b> (±0.04)	
	20-40	18.18 (±0.94)		18.84 (±0.74) A	20.77 (±0.22) A	19.27 (±1.02) A	-0.09 (±0.11)	<b>0.22</b> (±0.06)	<b>0.29</b> (±0.06)	<b>0.29</b> (±0.01)	
	40-60	13.12 (±0.44)		13.42 (±0.36)	13.95 (±0.22)	12.99 (±0.70)	-0.05 (±0.11)	0.14 (±0.05)	0.12 (±0.03)	0.13 (±0.02)	
	60-80	11.11 (±0.08)		10.78 (±0.47)	10.61 (±0.49)	10.18 (±0.59)	$0.02~(\pm 0.08)$	0.09 (±0.04)	-0.01 (±0.01)	0.04 (±0.01)	
	80-100	9.47 (±0.09)		8.71 (±0.33)	8.62 (±0.57) B	7.91 (±0.46) B	$0.02~(\pm 0.04)$	0.01 (±0.01)	<b>-0.10</b> (±0.01)	<b>-0.08</b> (±0.05	
	0-100	86.44 (±2.12)		87.42 (±2.52) A	91.07 (±0.92) A	88.01 (±3.03) A	0.14 (±0.35)	<b>1.07</b> (±0.18)	<b>0.70</b> (±0.11)	<b>0.95</b> (±0.07)	

<sup>&</sup>lt;sup>a</sup> MaiEx: Maize-based experiment; SoyEx: Soybean-based experiment; and CasEx: Cassava-based experiment.

<sup>&</sup>lt;sup>b</sup> CTM: monocropping under conventional tillage; NTM: monocropping under NT systems with no-till mulch-based cropping systems associated with different crop sequences, and NTR1 and NTR2 refer to bi-annual crop rotational systems under NT systems with no-till mulch-based cropping systems associated with different crop sequences as described in Table 1. Different uppercase letters accompanying the values indicate significant difference within the same treatment between 2011 and 2021 (diachronic) at the same soil depth at p < 0.05 (Tukey's test). Values of SOC stock change rates in bold indicate changes for a given treatment and depth between 2021 and 2011 are significantly different from 0 at p < 0.05 (Tukey's test).

0.05 (Tukey's test). Positive values of SOC stock change rate indicate a SOC accumulation; negative values indicate a SOC loss. Values in the parentheses indicate standard errors (n=3).

## 3.2 Impact of cropping systems on TN concentration and stock

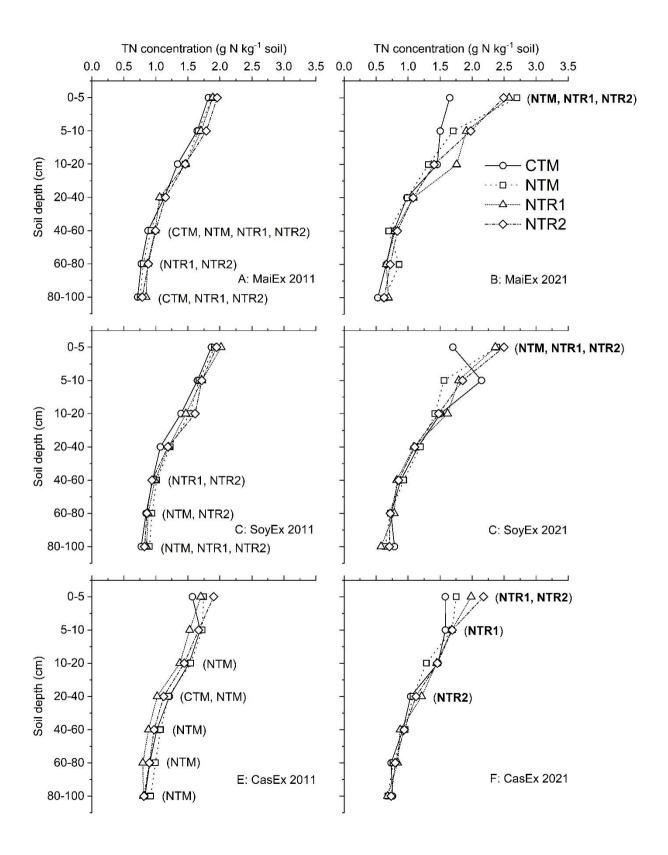
## 3.2.1 Total N concentration

469

470

471 Over 10 years of cultivation from 2011 to 2021, surprisingly, the response of soil TN concentration to 472 tillage and cropping systems differed from SOC (Fig. 2, Table S4 as duplication). The positive (p < 0.05) 473 effect on TN concentration was mainly observed on the surface layer under NT systems. However, the 474 significant (p < 0.05) decrease in TN concentration varied across tillage, cropping systems and 475 experiments observed below 20 cm (Fig. 2, Table S4). In 2021, NTR systems (NTR1 and NTR2) 476 significantly (p < 0.05) increased TN concentration in the top 5 cm of MaiEx (NTR-Mz) and SoyEx 477 (NTR-Sb) by 32% and 23%, respectively (Figs. B2 and 2D), but decreased TN concentration significantly 478 (p < 0.05) below 60 to 100 cm by -18 to -21% and -10 to -25% in MaiEx and SoyEX, respectively (Figs. 479 2A and 2B). Under CasEx in 2021, the soil TN concentration significantly (p < 0.05) increased by 16% 480 in the top 0-5 cm under the NTR system (average of NTR1-Cs and NTR2-Cs), while overall TN 481 concentration remained stable below 5 cm, except for significant increases under NTR1-Cs by 10% and 482 19% in the 5–10 cm and 20–40 cm, respectively (Fig. 2F). 483 From 2011 to 2021, in the case of monocropping under NT systems in MaiEx and SoyEx, the significant 484 (p < 0.05) increase in TN concentration by 44% and 25% under NTM-Mz and NTM-Sb, respectively 485 (Figs 2B and 2C). However, the TN concentration was significantly (p < 0.05) decreased by -24% in the 486 40-60 cm under NTM-Mz and from -23 to -29% in the 60-100 cm depth under NTM-Sb (Figs. 2A and 487 2B). After 10-years of cassava monocropping under NT system (NTM-Cs), TN concentration did not 488 change in the top 0–10 cm, but the concentration significantly (p < 0.05) decreased below 10–100 cm

- 489 depth from -10 to -25% in 2021 (Fig. 2E).
- 490 In contrast to NT systems, after 10-years of conventional tillage-based monocropping of soybean (CTM-
- 491 Sb) and cassava (CTM-Cs), the soil TN concentration in 2021 remained constant across the whole profile,
- 492 except a significant (p < 0.05) decrease of -14% in the 20–40 cm layer detected under CTM-Cs (Fig. 2D
- and 2E), while in cassava monocropping under CT (CTM-Mz), soil TN concentration remained stable
- from 0 to 40 cm, then significantly (p < 0.05) decreased by -12% and -26% in the 40–60 and 80–100 cm
- depths, respectively (Fig. 2A and 2B).



**Figure 2.** TN concentration distribution across the soil profile (0-100 cm) of the treatments under different experiments in 2011 and 2021. CTM: monocropping under conventional tillage; NTM: monocropping under NT systems with no- till mulch-based cropping systems associated with different crop sequences, and NTR1 and NTR2 refer to bi-annual crop rotational systems under NT systems with no-till mulch-based cropping systems associated with different crop sequences as described in Table 1. A: MaiEx 2011 - TN concentration of the treatments in maize-based trial measured in 2011; B: MaiEx 2021 - TN concentration of the treatments in soybean-based trial measured in 2021; C: SoyEx 2011 - TN concentration of the treatments in soybean-based trial measured in 2021; E: CasEx 2021 - TN concentration of the treatments in cassava-based trial measured in 2011; and F: CasEx 2021 - TN concentration of the treatments in cassava-based trial measured in 2021; Treatment(s) in bold within the brackets indicate the gain and significant (p < 0.05) difference in concentrations between 2011 and 2021 in the same treatment at the same soil depth.

## 3.2.2 Total N stock

Over the past decade, cultivating the main crops (maize, soybean, and cassava) under NTR systems (NTR1 and NTR2) significantly (p < 0.05) increased TN stock in the soil surface in all the experiments in 2021. However, the response of TN stock below the surface layers to the NTR systems differed between the three experiments (Table 3, Table S5). In the case of MaiEx (NTR1-Mz and NTR2-Mz), TN stock increased by 0.3 Mg N ha<sup>-1</sup> in the 0–5 cm; the stock remained stable in the 5–40 cm, but then significantly (p < 0.05) decreased in the 40–100 cm depths between -0.25 and -0.40 Mg N ha<sup>-1</sup> (Table 3). Similar to MaiEx, the soil TN stock in the soils under NTR systems of SoyEx (NTR1-Sb and NTR2-Sb) significantly increased by 0.25 Mg N ha<sup>-1</sup> in the 0–5 cm layer, remained constant in the 5–60 cm, and then significantly decreased in the 60–100 cm with a rough amount of -0.02 Mg N ha<sup>-1</sup> (Table 3). In contrast to MaiEx and SoyEx, among the two NTR-Cs crop rotation system, NTR2-Cs significantly (p < 0.05) increased TN stock by 0.10 Mg N ha<sup>-1</sup> in the surface 0–5 cm, whereas the significant increase in TN stock was detected in the 0–5, 5–10 and 20–40 cm by 0.10, 0.10 and 0.03 Mg N ha<sup>-1</sup>, respectively, under NTR1-Cs. Unlike,

- 509 MaiEx and SoyEx, there was no significant decrease in TN stock in the subsoil layers under the NTR-Cs
- 510 (Table 3).
- When compared to 2011, TN stock in 2021 of MaiEx and SoyEx under the NTM system significantly (p
- 512 <0.05) increased in the topsoil (0-5 cm) by 0.40 and 0.20 Mg N ha<sup>-1</sup> under NTM-Mz and NTM-Sb,
- respectively. However, significant (p < 0.05) decreases in TN stock were detected in the subsoils under
- 514 NTM-Mz by -0.40 Mg N ha<sup>-1</sup> in the 40–60 cm and by -0.30 to -0.40 Mg N ha<sup>-1</sup> in the 60–100 cm under
- 515 NTM-Sb (Table 3). In the case of CasEx, TN stock in the NTM-Cs soil remained constant in the top 0–
- 516 10 cm, then significantly decreased from -0.3 to -0.5 Mg N ha<sup>-1</sup> in the 10–100 cm (Table 3).
- 517 For the CTM of all the experiments, from 2011 to 2021, TN stock in the topsoil layers remained stable,
- 518 whereas losses were observed in the layers below 20 cm. CTM-Sb significantly (p < 0.05) increased TN
- stock by 0.20 Mg N ha<sup>-1</sup> in the 10–20 cm, then remained constant below 20 cm with a significant ( $p < 10^{-2}$
- 520 0.05) reduction by -0.20 Mg N ha<sup>-1</sup> detected in the 60–80 cm (Table 3). In CTM-Mz, TN stock did not
- 521 change in the 0-40 cm but significantly declined between -0.20 to -0.30 Mg N ha<sup>-1</sup> from 40-100 cm
- 522 (Table 3). In CTM-Cs soil, TN stock did not change in the top 0–20 cm but significantly decreased from
- 523 -0.20 to -0.30 Mg N ha<sup>-1</sup> in the 20–100 cm (Table 3).
- Measured in the whole profile (0–100 cm), over the past decade, the TN stock under NTR systems of all
- 525 the experiments remained stable (Table 3). Monocropping of soybean and cassava under NT systems
- 526 (NTM-Sb and NTM-Cs) caused a significant (p < 0.05) reduction of TN stock at the annual depletion rate
- of -0.10 and -0.17 Mg N ha<sup>-1</sup> yr<sup>-1</sup>, respectively, while nearly a decade of upland rice monocropping then
- 528 recent shift to maize under NTM system (NTM-Mz) did not change TN stock (Table 3). In the case of
- monocropping of main crops under CT, the TN stock under CTM-Cs significantly (p < 0.05) decreased

- at the rate of -0.11 Mg N ha<sup>-1</sup>yr<sup>-1</sup>. The TN stock of soil under CTM-Sb remained stable, while the CTM-
- Mz showed the depletion trend at the rate of -0.11 Mg N ha<sup>-1</sup> yr<sup>-1</sup> although non-significant (Table 3).

**Table 3**. Mean TN stock in 2021 and TN stock change rate between 2021 and 2011.

Experime nts <sup>a</sup>	Approxi mate soil depth (cm)	Cropping systems <sup>b</sup>							
		CTM	NTM	NTR1	NTR2	CTM	NTM	NTR1	NTR2
		N stock in 2021 (Mg N ha <sup>-1</sup> )				N stock change rate 2021-11 (Mg N ha <sup>-1</sup> yr <sup>-1</sup> )			
MaiEx	0-5	0.79 (±0.04)	1.30 (±0.02) A	1.24 (±0.03) A	1.20 (±0.04) A	-0.01 (±0.01)	<b>0.04</b> (±0.01)	<b>0.03</b> (±0.00)	<b>0.03</b> (±0.00)
	5-10	0.79 (±0.05)	0.95 (±0.01) A	1.04 (±0.02)	1.06 (±0.12)	-0.01 (±0.01)	$0.01 (\pm 0.00)$	0.01 (±0.00)	0.01 (±0.01)
	10-20	1.55 (±0.01)	1.44 (±0.03) B	1.87 (±0.23)	1.58 (±0.12)	0.01 (±0.01)	-0.01 (±0.01)	0.03 (±0.02)	0.00 (±0.02)
	20-40	1.98 (±0.10)	1.91 (±0.06)	2.18 (±0.11)	2.11 (±0.05)	-0.02 (±0.02)	-0.02 (±0.00)	0.01 (±0.02)	-0.02 (±0.01)
	40-60	1.46 (±0.03) B	1.33 (±0.11) B	1.48 (±0.12) B	1.56 (±0.05) B	<b>-0.02</b> (±0.01)	<b>-0.04</b> (±0.00)	<b>-0.03</b> (±0.01)	<b>-0.03</b> (±0.01)
	60-80	1.24 (±0.09) B	1.49 (±0.13)	1.25 (±0.04) B	1.36 (±0.02) B	<b>-0.02</b> (±0.01)	-0.01 (±0.01)	<b>-0.04</b> (±0.01)	<b>-0.04</b> (±0.00)
	80-100	1.00 (±0.08) B	1.22 (±0.02)	1.21 (±0.03) B	1.17 (±0.03) B	<b>-0.03</b> (±0.01)	-0.02 (±0.01)	<b>-0.03</b> (±0.01)	<b>-0.04</b> (±0.00)
	0-100	8.82 (±0.46)	9.63 (±0.49)	10.27 (±0.17)	10.03 (±0.30)	-0.11 (±0.02)	-0.06 (±0.06)	-0.03 (±0.06)	-0.09 (±0.03)
SoyEx	0-5	0.82 (±0.03)	1.15 (±0.01) A	1.13 (±0.03) A	1.20 (±0.06) A	-0.01 (±0.00)	<b>0.02</b> (±0.00)	<b>0.02</b> (±0.00)	<b>0.03</b> (±0.01)
	5-10	1.07 (±0.22)	0.87 (±0.11)	0.99 (±0.02)	1.00 (±0.03)	0.02 (±0.02)	0.00 (±0.01)	0.01 (±0.00)	0.01 (±0.00)
	10-20	1.75 (±0.12) A	1.53 (±0.08)	1.74 (±0.09)	1.62 (±0.02)	<b>0.02</b> (±0.01)	-0.01 (±0.00)	0.01 (±0.01)	-0.01 (±0.01)
	20-40	2.28 (±0.04)	2.29 (±0.12)	2.20 (±0.06)	2.14 (±0.04)	0.02 (±0.01)	-0.01 (±0.01)	-0.01 (±0.02)	-0.02 (±0.02)
	40-60	1.74 (±0.04)	1.80 (±0.05) A	1.60 (±0.09)	1.62 (±0.03)	0.00 (±0.01)	-0.02 (±0.01)	-0.03 (±0.02)	-0.02 (±0.01)
	60-80	1.44 (±0.05) B	1.47 (±0.08) B	1.44 (±0.03) B	1.39 (±0.02) B	<b>-0.02</b> (±0.00)	<b>-0.03</b> (±0.01)	<b>-0.02</b> (±0.01)	<b>-0.02</b> (±0.01)
	80-100	1.36 (±0.04)	1.21 (±0.17) B	1.18 (±0.04) B	1.26 (±0.01) B	-0.01 (±0.00)	<b>-0.04</b> (±0.02)	<b>-0.04</b> (±0.01)	<b>-0.02</b> (±0.00)
	0-100	10.45 (±0.39)	10.31 (±0.29) B	10.29 (±0.13)	10.23 (±0.12)	0.02 (±0.04)	<b>-0.10</b> (±0.03)	-0.06 (±0.03)	-0.06 (±0.03)
CasEx	0-5	0.76 (±0.01)	0.84 (±0.01)	0.95 (±0.01) A	1.05 (±0.04) A	0.00 (±0.00)	0.00 (±0.00)	<b>0.01</b> (±0.00)	<b>0.01</b> (±0.01)
	5-10	0.82 (±0.02)	0.88 (±0.01)	0.89 (±0.02) A	0.91 (±0.01)	$0.00 (\pm 0.00)$	$0.00 (\pm 0.00)$	<b>0.01</b> (±0.00)	0.00 (±0.00)
	10-20	1.57 (±0.03)	1.40 (±0.07) B	1.57 (±0.02) A	1.58 (±0.01) A	-0.01 (±0.01)	<b>-0.03</b> (±0.01)	0.01 (±0.00)	0.00 (±0.00)
	20-40	2.06 (±0.07) B	2.04 (±0.10) B	2.34 (±0.02) A	2.19 (±0.10)	<b>-0.03</b> (±0.01)	<b>-0.03</b> (±0.01)	<b>0.03</b> (±0.00)	0.00 (±0.01)
	40-60	1.69 (±0.05) B	1.74 (±0.05) B	1.75 (±0.03)	1.74 (±0.06)	<b>-0.02</b> (±0.01)	<b>-0.03</b> (±0.00)	0.01 (±0.00)	-0.01 (±0.01)
	60-80	1.48 (±0.10) B	1.51 (±0.11) B	1.54 (±0.02)	1.52 (±0.02)	<b>-0.02</b> (±0.02)	<b>-0.04</b> (±0.01)	0.00 (±0.00)	-0.02 (±0.01)
	80-100	1.32 (±0.09)	1.29 (±0.11) B	1.34 (±0.05)	1.35 (±0.03)	-0.02 (±0.01)	<b>-0.05</b> (±0.01)	-0.01 (±0.01)	-0.02 (±0.01)
	0-100	9.70 (±0.21) B	9.70 (±0.39) B	10.40 (±0.04)	10.34 (±0.25)	<b>-0.11</b> (±0.06)	<b>-0.17</b> (±0.07)	0.06 (±0.02)	-0.03 (±0.01)

<sup>a</sup>MaiEx: Maize-based experiment; SoyEx: Soybean-based experiment; and CasEx: Cassava-based experiment.

<sup>b</sup>CTM: monocropping under conventional tillage; NTM: monocropping under NT systems with no-till mulch-based cropping systems associated with different crop sequences, and NTR1 and NTR2 refer to bi-annual crop rotational systems under NT systems with no-till mulch-based cropping systems associated with cropping systems as described in Table 1. Different uppercase letters accompanying the values indicate significant difference within the same treatment between 2011 and 2021 (diachronic) at the same soil depth at p < 0.05 (Tukey's test). Values of TN stock change rates in bold indicate changes for a given treatment and depth between 2021 and 2011 are significantly different from 0 at p < 0.05

(Tukey's test). Positive values of TN stock change rate indicate a N accumulation; negative values indicate a N loss. Values in the parentheses indicate standard errors (n=3).

### 3.3 Impact of cropping systems on C and N stocks in size fractions

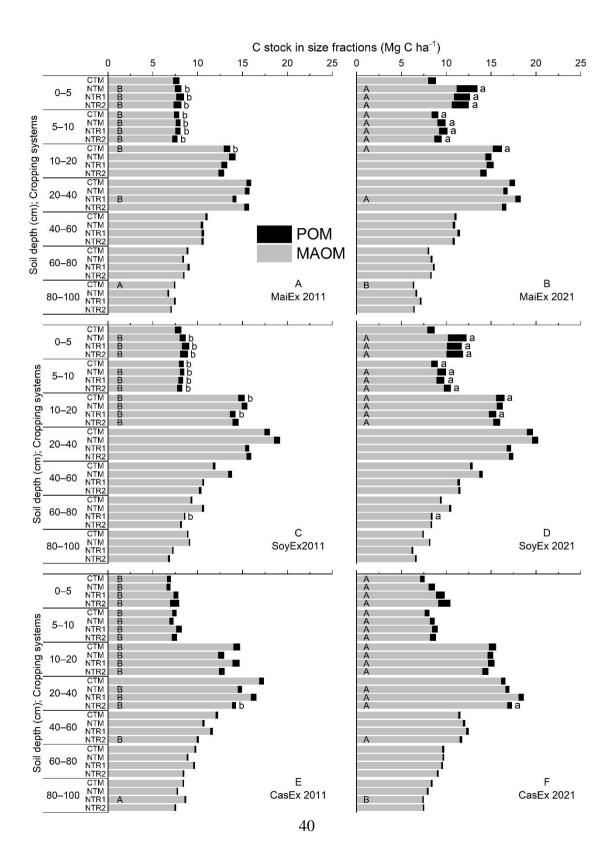
532

551

533 3.3.1 C stock in size fractions 534 In this diachronic study, over the 10-year period, the stocks of C-POM and C-MAOM were significantly 535 (p < 0.05) influenced by all the treatments. However, the effects varied across cropping systems and the 536 experiments (Fig. 4, Table S6 and S7). 537 The data showed that C-POM stock in 2021 significantly (p < 0.05) increased in the surface layers (0-10)538 cm) under all the NT systems in MaiEx and SoyEx, but it was not the case in CasEx (Figs. 4B, 4D, and 539 4F). The annual accumulation rates of C-POM stock in MaiEx and SoyEx were similar, with a range of approximately 0.15 and 0.04 Mg C ha<sup>-1</sup> yr<sup>-1</sup> under NTM system and 0.10 and 0.03 Mg C ha<sup>-1</sup> yr<sup>-1</sup> under 540 541 NTR systems (average of NTR1 and NTR2) in the 0-5 and 5-10 cm, respectively. This suggested the 542 consequence of the annual biomass inputs that were left on the soil surface under all the NT systems over 543 the experimental period (Table 1). Although the significant increase in C-POM stock was also detected 544 under CTM in the tilled layers (5–20 cm) in MaiEx and SoyEx, at the annual accumulation rates of only 0.02 Mg C ha<sup>-1</sup> yr<sup>-1</sup> across the two soil depths (5–10 and 10–20 cm), which is relatively low when 545 546 compared with NT systems (Figs. 4B and 4D). 547 In a similar trend to C-POM, C-MAOM stock increased significantly (p < 0.05) in the top soil depths 548 under all the NT systems in MaiEx and SoyEx in 2021. The annual accumulation rates were similar between NTM-Mz and NTR-Mz, with a rate of 0.33 and 0.15 Mg C ha<sup>-1</sup> yr<sup>-1</sup> in the 0–5 and 5–10 cm, 549 550 respectively (Fig. 4B). In SoyEx, all the NT systems exhibited the trend of C-MAOM stock accumulation

in the deeper layers (to 20 cm) than MaiEx, with approximate annual accumulation rates of 0.20, 0.15,

and 0.10 Mg C ha<sup>-1</sup> yr<sup>-1</sup> in the 0–5, 5–10, and 10–20 cm, respectively (Fig. 4D). In CasEx, despite the fact 552 553 that the C-POM stock remained constant over the past decade, the C-MAOC stock significantly (p < 0.05)554 increased down to 40 cm by all the NT systems in 2021, with similar accumulation rates from 0.09 to 555  $0.26 \text{ Mg C h}^{-1} \text{ yr}^{-1}$  in the 0-40 cm depths (Fig 4F). 556 Under CTM in 2021, an increase in C-MAOM stock was observed in the tilled layers across all experiments (Figs. 4B, 4D, and 4F). Specifically, in the MaiEx experiment, significant differences (p < 557 558 0.05) of C-MAOM stock between 2011 and 2021 were found in the 5–10 cm and 10–20 cm layers, with annual accumulation rates of 0.10 and 0.23 Mg C ha<sup>-1</sup> yr<sup>-1</sup>, respectively (Fig. 4B). In the case of SoyEx, 559 560 a significant increase in C-MAOM stock was only detected in the 10-20 cm layer, with an annual 561 accumulation rate of 0.11 Mg C ha<sup>-1</sup> yr<sup>-1</sup> (Fig. 4D). Meanwhile, in the CasEx experiment, the C-MAOM stock showed a significant annual increase at a rate of 0.05 Mg C ha<sup>-1</sup> yr<sup>-1</sup> across the topsoil 0–20 cm 562 563 (Fig. 4F).



**Figure 4.** Carbon stock in mineral-associated and particulate organic matter (MAOM and POM) fractions across the whole profile (01-100 cm) in 2011 and 2021 under different treatments and experiments. CTM: monocropping under conventional tillage; NTM: monocropping under NT systems with no-till mulch-based cropping systems associated with different crop sequences, and NTR1 and NTR2 refer to bi-annual crop rotational systems under NT systems with no-till mulch-based cropping systems associated with cropping systems as described in Table 1. The uppercase letters on the bars and lowercase letters in front of the bars indicate a significant difference (Tukey's test; p < 0.05) in C stock in MAOM and POM, respectively, between 2011 and 2021 for the same treatment and soil depth.

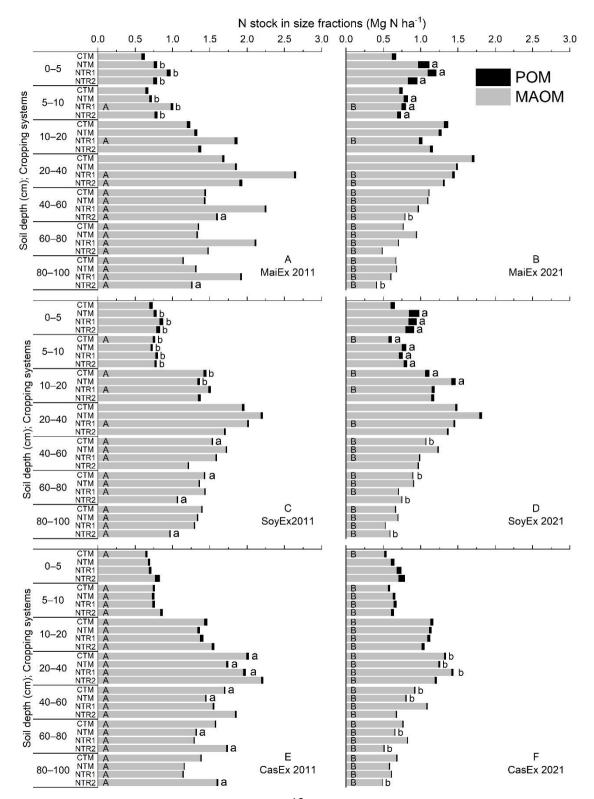
#### 3.3.2 N stock in size fractions

564

565 Over the past decade (2011–2021), cropping systems had varying effects on the stocks of N-POM and N-566 MAOM across soil depths and the experiments (Fig. 5, Tables S8 and S9). In 2021, N-POM stock 567 increased significantly (p < 0.05) in the topsoil (0-10 cm) under all the NT systems in MaiEx and SoyEx, with similar amounts of 0.10 and 0.01 Mg N ha<sup>-1</sup> in the 0–5 and 5–10 cm, respectively (Figs. 5B and 5D). 568 569 Below 10 cm, N-POM stock remained constant under all the NT systems in both experiments, except for 570 the depletion trend found under a NTR system, in particular under NTR2-Mz below 40 cm and NTR2-Sb 571 below 60 cm (Figs. 5B and 5D). In contrast to MaiEx and SoyEx, in CasEx, none of NT systems increased 572 N-POM stock in the top soils, but NTM-Cs and NTR-Cs systems significantly (p < 0.05) depleted it below 20 cm (Fig 5F). 573 574 In 2021, monocropping of soybean under conventional tillage (CTM-Sb) significantly accumulated N-POM stock in the tilled layers (5-10 and 10-20 cm) with an amount of 0.01 Mg N ha<sup>-1</sup> across the two 575 576 layers, but the significant depletion (p < 0.05) at a similar amount was observed below 40 cm (Fig. 5D). 577 Monocropping of upland rice over a decade and recent shift to maize under conventional tillage (CTM-578 Mz) did not change the N-POM stock across the soil profile (Fig. 5B), whereas the N-POM stock under CTM-Cs soil remained stable in the top 20 cm, but significantly declined by -0.01 Mg N ha<sup>-1</sup> from 20 to 579

580 60 cm (Fig. 5E and 5F).

Surprisingly, from 2011 to 2021, none of the tillage or cropping systems increased N-MAOM stock, but 581 582 decreased it with varying soil depths and experiments (Fig. 5). In MaiEx and SoyEx, N-MAOM stock 583 remained unchanging under NTM system (i.e., NTM-Mz and NTM-Sb) from 0 to 40 cm but declined significantly (p < 0.05) below 40 cm with the rate ranging from -0.036 to -0.063 Mg N ha<sup>-1</sup> yr<sup>-1</sup> (Figs. 5A 584 585 and 5C). Under NTR-Mz and NTR-Sb, the significant decrease of N-MAOM stock was detected below 586 5 cm to subsoil layers, but this was inconsistent between the two NTR systems (NTR1 and NTR2) and soil depths, with depletion rates ranging from -0.023 Mg N ha<sup>-1</sup> yr<sup>-1</sup> in the near soil surface to -0.140 Mg 587 N ha<sup>-1</sup> yr<sup>-1</sup> in the bottom of the soil profile (Figs. 5A and 5C). In CasEx, the N-MAOM stock in the surface 588 589 layer (0-5 cm) did not change under all the NT systems (NTM-Cs, NTR1-Cs, and NTR2-Cs), but decreased significantly (p < 0.05) below 5 cm with the annual depletion rates ranging from -0.009 to Mg 590 N ha<sup>-1</sup> yr<sup>-1</sup> in the 5 cm to -0.111 Mg N ha<sup>-1</sup> yr<sup>-1</sup> in the subsoil profile (Fig. 5E and 5F). 591 In 2021, the N-MAOM stock of the CTM-Mz soil remained steady at the 0-40 cm, whereas depletion 592 593 was detected from 40–100 cm at rates ranging from -0.032 to -0.058 Mg N ha-1 yr-1 (Figs. 5A and 5B). 594 CTM-Sb did not preserve N-MAOM stock even in tilled layers over the past ten years, but depleted it significantly below 5 cm to subsoil depths at rates of -0.016 to -0.073 Mg N ha<sup>-1</sup> yr<sup>-1</sup> (Figs. 5C and 5D), 595 while a significant decrease in N-MAOM stock was observed throughout the soil profile (0–100 cm) with 596 a depletion of -0.013 to -0.081 Mg N ha<sup>-1</sup> yr<sup>-1</sup> (Figs. 5E and 5F). 597



**Figure 5.** TN stock in mineral-associated and particulate organic matter (MAOM and POM) fractions across the whole profile (01-100 cm) in 2011 and 2021 under different treatments and experiments. CTM: monocropping under conventional tillage; NTM: monocropping under NT systems with no- till mulch-based cropping systems associated with different crop sequences, and NTR1 and NTR2 refer to bi-annual crop rotational systems under NT systems with no-till mulch-based cropping systems associated with cropping systems as described in Table 1. The uppercase letters on the bars and lowercase letters in front of the bars indicate a significant difference (Tukey's test; p < 0.05) in TN stock in MAOM and POM, respectively, between 2011 and 2021 for the same treatment and soil depth.

#### 4 Discussion

598

599

600

601

602

603

604

605

606

607

608

609

610

611

612

613

# 4.1 Change in SOC stock

Despite the contrasted effects among the NT systems and the experiments, our study showed that adopting NT systems with the use of cover crops in the long-term significantly increased SOC stock (Table 3). Several studies reported that long-term NT adoption accumulated SOC stock only on the surface soils, but the stock did not differ from CT when considering the whole soil profile (Blanco-Canqui and Lal, 2008; Luo et al., 2010; Blanco-Canqui et al., 2011; Du et al., 2017; Xiao et al., 2020). For example, a recent meta-analysis from 86 studies covering a range of crop productions across the world, Xiao et al., (2020) found that NT systems significantly accumulated the SOC stock only in the top 0–5 cm, and no significant change was found below 5 cm. Across climatic conditions, soil types, and various cropping systems in China, based on 95 comparisons between NT and CT, adopting NT led to increase in SOC stock by 3.8% in the upper 20 cm layer but to a decrease in SOC in the 30–40 cm layer (Du et al., 2017). Similarly, from a systematic review of global data of 69 paired-experiments, Luo et al., (2010) reported that long-term NT adoption only significantly affected SOC stock in the top 0–10 cm but not down to 40 cm depth. The authors also reported that increasing crop species diversity resulted in a lower SOC accumulation in the surface and a greater SOC loss in deeper layers.

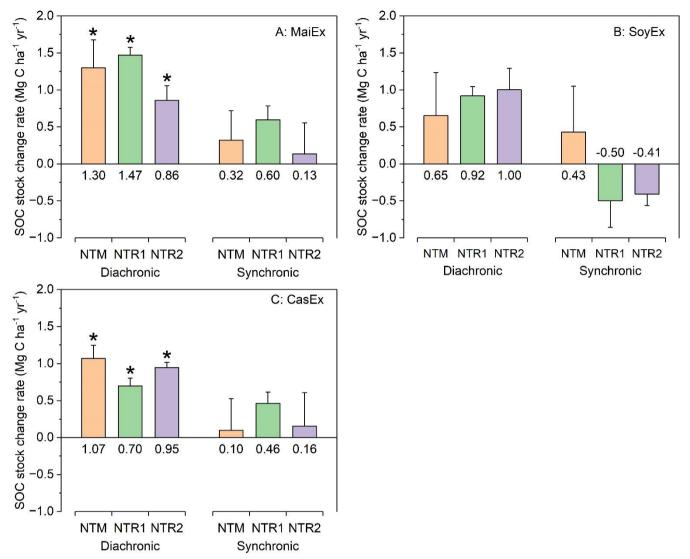
614 SOC stock changes reported under NT systems may differ according to climate, soil type and cropping 615 systems (Paustian et al., 1997; Six et al., 2002; Bayer et al., 2006; Ogle et al., 2012; Virto et al., 2012). 616 Soils in the tropical climate require diversified and large amounts of C inputs for NT viability due to fast 617 residue decomposition and difficulty in maintaining soil cover (Séguy et al., 2006; Castro et al., 2015). 618 In the same experiments as in our study, Hok et al. (2015) reported that NT systems with diverse crop 619 species significantly accumulated SOC at the surface 0-5 cm after 4 years of NT adoption. Our study 620 revealed that NT systems significantly increased SOC stock, although there was variability among the 621 NT systems and across the three experiments in the accumulation rates in the subsoil layers (Table 2). 622 Considering the cumulative SOC stock, all the NT systems significantly (p < 0.05) increased cumulative 623 SOC stock across the whole soil profile in MaiEx and CasEx. In SoyEx, significant increase in cumulative 624 SOC stock was limited to the top 0-20 cm under NTM-Sb, whereas NTR-Sb had significantly 625 accumulating SOC stock from 0 to 80 cm depths (Table S10). 626 Consistent with our findings, with the intensive NT systems and high C inputs retained to the soils, other 627 studies reported that long-term NT with the use of cover crops increased SOC stock beyond the surface 628 and the whole soil profile (Diekow et al., 2005; Boddey et al., 2010; Olson et al., 2014). From three long-629 term experiments (15–26 years) on Ferralsols in South Brazil, no-tillage with intensive cropping systems of maize and soybean production increased SOC with annual accumulation rates between 0.04 and 0.88 630 Mg ha<sup>-1</sup> in 0-30 cm, and from 0.48 to 1.53 Mg ha<sup>-1</sup> yr<sup>-1</sup> 0-100 cm (Boddey et al., 2010). After 12 years 631 632 of NT adoption with the use of cover crops for soybean and maize rotation in a humid continental sloping 633 land in Illinois, USA, SOC stock recovered from its initial SOC loss under CT before the experiment

implementation, with accumulation rates of 0.42, 0.78, and 1.21 Mg C ha<sup>-1</sup> yr<sup>-1</sup> at 0–15, 15–75, and 0–75 634 635 cm, respectively (Olson et al., 2014). 636 SOC storage and stabilization could be explained by several processes: (i) continuous supplies of large 637 quantities and diverse qualities of plant biomass-C inputs to the soil (Sá et al., 2014); (ii) the 638 transformation of this biomass-C by microbial communities into various organic C forms (Frasier et al., 639 2016; Schmidt et al., 2019); (iii) the stabilization of newly derived C by physical protection, binding with 640 organo-mineral particles, and biochemically stabilization through the formation of recalcitrant soil 641 organic matter (Six et al., 2002); and (iv) distribution of SOC over the soil profile through biological 642 processes, from root systems (Lorenz and Lal, 2005) and soil fauna (Lavelle et al., 2016). 643 In NT systems, multiple crop species were sown in the same unit area of land through the rotation of cash 644 crops and the use of cover crops by intercropping or during the fallow period, producing a large quantity 645 and diverse quality of C inputs above- but also below-ground that were retained to the soils. In our experiments, the annual biomass C inputs retained in the NT soils ranged from 3.61 to 6.07 Mg ha<sup>-1</sup> yr<sup>-1</sup>, 646 versus 1.36 to 2.20 Mg ha<sup>-1</sup> yr<sup>-1</sup> under CTM (Table 1). In a clayey Oxisol of Brazil, a 16-year-old 647 648 experiment revealed that NT was more effective than CT at converting biomass-C inputs into SOC, with 649 a C conversion ratio in 0-40 cm depth of 0.35 compared to 0.07 under NT and CT, respectively (Sá et al., 650 2014). In addition, integration of cover crops into the crop production system led to a significant increase in SOC. From the observation of 139 plots at 37 sites from the tropics and temperate zone and diverse 651 652 soil types, Poeplau and Don (2015) reported that the use of cover crops led to an average SOC accumulation rate of 0.32 Mg C ha<sup>-1</sup> yr<sup>-1</sup> at 22 cm depth. Association of tropical legume cover crops in 653 maize production led to increased SOC stock in the surface as well as the whole soil profile. Diekow et 654

655 al. (2005) found that the SOC accumulation rate in the legume-based cropping systems was 0.83 and 1.42 Mg C ha<sup>-1</sup> yr<sup>-1</sup> in the 0–17.5 cm and 0–107.5 cm layer, respectively after 17 years of NT adoption in a 656 657 Brazilian Acrisol, From a 30-year-old experiment in a Brazilian Acrisol, legume cover crops were twice 658 as effective in storing C as mineral N fertilization, with 1 kg of residue C input being transformed to 0.15 659 kg of SOC (Veloso et al., 2018). 660 Considering the challenges faced by smallholder farmers in Cambodia with low financial resources and/or high level of indebtedness, the main strategy should focus on enhancing nutrients cycling through 661 662 continuous biomass-C inputs under no-till cropping systems plus a combination of actions to reduce 663 nutrient removal from cassava fields through the non-removal of leaves and of a proportion of stalks, that 664 may also help to reduce the impact of nutrients deficiency. In addition, the tolerance of cassava to acidic 665 soil, its ability to grow on depleted and degraded soils related to the occurrence and synergistic effects of 666 arbuscular mycorrhizal fungi (Howeler et al., 1982; Howeler and Sieverding, 1983), and plant growthpromoting rhizobacteria (Balota et al., 1999), its nutrient recycling ability through leaf litter and when the 667 668 stalks are not used as planting materials and kept into the field, could be used to improve soil and cropping 669 system sustainability (Fermont et al., 2008). This possible use by farmers of cassava cropping systems as 670 a strategy for regenerating soil fertility was also emphasized by Saïdou et al., (2004) and Adjei-Nsiah et 671 al., (2007) in Benin and Ghana, respectively. 672 Long-term NT adoption has been shown to significantly improve soil structure, soil porosity and pore 673 connectivity (Cooper et al., 2021) contributing to the improvement of water infiltration, gas exchanges 674 and microbial activities, and roots development to deeper soil profile (Rosolem et al., 2016). In addition, 675 aerobic condition of soil aggregates would enhance SOC stability in unsaturated soils (Zhang et al., 2021).

676 Sisti et al. (2004) showed that increased C accumulation in NT soil below 30 cm depth could be explained 677 by greater root density when compared with CT. Another possibility is that organic residues from upper 678 layers were transported downward by soil meso- and macro-fauna organisms, which could have been 679 favoured by better environmental conditions provided by the continuous C flow and soil structure 680 enhancement under NT systems (Lavelle et al., 2016). 681 In our study, the SOC stock in the whole soil profile (0–100 cm) under CTM and for the three experiments 682 remained stable, which could be attributed to the fully retained crop residues (i.e., mungbean, rice, and 683 maize) in MaiEx and (i.e., sesame and soybean) in SoyEx and partially retained cassava's fallen leaves 684 and stalks in CasEx (Table 1), indicating that it reached a new SOC equilibrium. The high clay content 685 also contributed to the stabilisation of the MAOC that accounted, in 2021, for 97.2% of the carbon stock 686 along the soil profile. 687 Under a synchronic approach, considering CTM as the reference, the SOC stock change rates in 2021 under NT systems ranged from 0.13 to 0.60, -0.50 to 0.43, and 0.10 to 0.46 Mg C ha<sup>-1</sup> yr<sup>-1</sup> in MaiEx, 688 689 SoyEx, and CasEx, respectively (Fig. 6). When compared with the diachronic approach, this corresponds 690 to an underestimation of 146 to 536%, 51 to 347%, and 51 to 997% in MaiEx, SoyEx, and CasEx, 691 respectively (Table 3). In Brazil, from on-farm assessments of the SOC dynamics under long-term NT 692 systems in tropical heavy clayey soils Neto et (al., 2010) and Junior et al., (2013) reported that synchronic approach led to biased annual SOC accumulation rates under NT systems when compared with diachronic 693 694 approach. The main factors associated with the errors could be the underlying heterogeneities of the soil 695 conditions prior to the conversion to NT systems that are hard to capture despite all the steps of the 696 methodologically precautious measurements being implemented properly (Neto et al., 2010; Junior et al.,

- 697 2013). Our findings clearly emphasize the importance of the diachronic approach in accurately estimating
- the effects of long-term NT systems on SOC storage.



**Figure 6.** Comparison between the diachronic and synchronic approaches to estimate SOC stock change rate (0–100 cm) from 2011 to 2021 under NT systems in the tropical red Oxisol of Cambodia (n = 3; error bars = SE). A: MaiEx (maize-based trial); B: SoyEx (soybean-based trial); and C: CasEx (cassava-based experiments). CAM: monocropping under conservation agriculture with no-till mulch-based cropping systems associated with different crop sequences, and CAR1 and CAR2 refer to bi-annual crop rotational systems under conservation agriculture with no-till mulch-based cropping systems associated with different crop sequences described in Table 1. The stock change rates under diachronic were calculated by subtracting the stock of the same treatment in 2021 from the stock in 2011 and dividing by the number of years between the 1st and 2nd samplings (10 years), while the stock change rates of CA systems in 2021 under synchronic were calculated by subtracting the stock of each CA treatment from the stock of CTM in 2021, considered the control, and dividing by the number of years between the 1st and 2nd

samplings (10 years). (\*) indicates a significant difference (Tukey's test; p < 0.05) in SOC stock between 2011 and 2021. Positive values indicate SOC stock accumulation; negative values indicate SOC loss.

### 4.2 Change in N stock

699

700 In addition to increasing SOC stock in the surface and the whole soil profile, Diekow et al. (2005) found 701 that soil TN stock was significantly increased by an average of 27% in the surface (0–17.5 cm) and by 702 6% in the whole profile (0–107.5 cm) after 17 years of NT maize and tropical legume intercropping and 703 N fertilization compared with its original state under native grassland of Brazilian Acrisol. Sá et al., (2014) reported a significant correlation ( $R^2 = 0.89$ , P < 0.0002) between soil N and SOC stock accumulation. 704 Each unit of N stock accumulation contributed to the sequestration of 10.2 Mg C ha<sup>-1</sup> at the top 0–10 cm 705 706 under long-term (16-year) continuous NT maize-based production of Brazilian's Oxisol. 707 However, the diachronic assessment in our study showed that TN stock under NT systems significantly 708 increased only in the topsoil (0-5 cm) in MaiEx and SoyEx, while the stock remained stable in CasEx 709 (Table 3). The significant decline of TN stock under NT systems, although with variability across the NT 710 systems and the experiments, was detected below 20 cm. When considering the whole profile (0-100 711 cm), significant depletion of N stock was observed under the NT monocropping systems, with a loss rate at -0.10 and -0.17 Mg N ha<sup>-1</sup> yr<sup>-1</sup> in SoyEx (NTM-Sb) and CasEx (NTM-Cs), respectively. Under NT 712 713 crop rotation systems, despite non-significant, TN stock tended to decrease across the three experiments, with a depletion rate ranging from -0.03 to -0.09 Mg N ha<sup>-1</sup> yr<sup>-1</sup> (Table 4). 714 715 The depletion of TN stock under NT was reported from short- (Wuaden et al., 2020) to longer-term NT 716 adoption (Delgado, 2023). From a short-term (5-year) conversion of native grassland to cropland under 717 NT adoption with a double cropping with maize as a cash crop followed by black oat as a cover crop in

Brazil's Rhodic Nitosol, NT soils had significant losses of soil total N in comparison with the original 718 719 stocks under grassland throughout the soil profile, with the exception of the 0–5 and 10–20 cm soil layers. Considering the whole profile (0–60 cm), soil total N was depleted by -1.7 Mg N ha<sup>-1</sup>, equivalent to an 720 annual loss rate of -0.34 Mg N ha<sup>-1</sup> yr<sup>-1</sup> after 5 years of grassland conversion to NT (Wuaden et al., 2020). 721 722 Results from a 12-year experiment in the US (0-120 cm depth) in an irrigated NT continuous maize 723 rotation where mineral N were applied at different rates indicated that even NT could potentially have 724 significant net N loss with an average loss of -15 kg N ha<sup>-1</sup> yr<sup>-1</sup> at the top 30 cm of soil regardless of N 725 application rate (Delgado, 2023). In our study, it is a rather surprising finding to observe an increase in SOC and a simultaneous soil TN 726 727 depletion. Associating legume cover crops in the cropping system did not enhance soil N through 728 biological N fixation (Rosolem et al., 2016). In general, N is an important factor contributing to SOC 729 storage (De Vries, 2014; Kirkby et al., 2014). Nutrient reserves are among other factors that determine 730 soil C storage capacity (Lal, 2018). Therefore, more studies on nutrient availability and their 731 stoichiometry relationship including in deeper layers (>100 cm), on the N use efficiency and N cycling 732 are needed to understand the driving mechanisms of the N dynamics under these NT systems. 733 Nitrogen uptake and/or N priming effects from the cover crops, among other factors, could possibly have resulted in N loss. Priming effects are short-term changes in the turnover of soil N caused by the addition 734 735 of organic or mineral fertilizer, the mechanical treatment of soil, its drying and rewetting (Kuzyakov et 736 al., 2000), and the exudation of organic substances in the rhizosphere by living plants (Kuzyakov, 2002). 737 These effects can occur immediately or very shortly after the addition of a specific substance to the soil 738 and are larger in soils rich in C and N than those in poor soils (Kuzyakov et al., 2000). In our experiments, under CA systems, the soils are year-round protected by the cover crops established through association or succession with the main crops (maize, soybean, and cassava) and continue to grow after the main crop harvest. Several species of drought-tolerant and fast-growing cover crops (stylo, brachiaria, cowpea, sorghum, pearl millet, and sunhemp), which are commonly used in our experiments as a single or mixture (Table 1), are good examples of remaining green throughout the dry season with root exudates that may have enhanced the priming effect. In addition, the symbiosis relationship between the cover crops and rhizobia during the dry season could also be low due to low soil moisture content, therefore resulting in high N uptake from the soil by those cover crops. Their drought-tolerant characteristics allow these species to cross the dry season, even with little or no rain for more than 4 months in the dry season. Their fast-growing characteristics, along with the species diversity, produced a large amount of biomass annually and were retained in the soil at the termination of the cultivation of the main crops (Table 1), which may create conditions for the N uptake or N priming effects to happen. Measurement of N content and the estimation of biological nitrogen fixation by the legume cover crops using the <sup>15</sup>N isotopic technique should be conducted to better understand N dynamics in the different systems. To date, few studies on the impact of no-till systems on N gains or N losses, N use efficiency and changes in soil N have been conducted, and the results vary depending on the period of NT adoption and sampling depths (Congreves et al., 2017; Delgado, 2023). Further research is needed to understand the driving mechanism of the N dynamics under NT systems by considering deeper layers (>100 cm) for making informed decisions regarding sustainable soil fertility management and crop production systems. Positive accumulation rates of SOC stock, recorded under NT systems, could not be sustained on the long-term as the depletion of the TN stock may lead to nutrient scarcity of other nutrients (P, S, Ca<sup>2+</sup> and Mg<sup>2+</sup>) that is

739

740

741

742

743

744

745

746

747

748

749

750

751

752

753

754

755

756

757

758

the driving force limiting SOC accumulation. Further analysis is needed to assess the coming changes in SOC and N stocks along with the content of nutrients and its potential impact on the rate of SOC accumulation or depletion (Kirkby et al., 2013).

### 3.4 Carbon and N in size fractions and stabilization processes in NT systems

763

764

765

766

767

768

769

770

771

772

773

774

775

776

777

778

779

Particulate organic matter (POM) and mineral-associated organic matter (MAOM) are the two main fractions of the SOC pools. They differ in physical and chemical characteristics as well as their turnover rates. POM is more sensitive to soil tillage and land use than MAOM and total SOC (Blanco-Moure et al., 2013; Kan et al., 2021). In all the experiments, NT systems significantly increased C in both POM and MAOM fractions in the topsoil layer (Fig. 4). These increases could be attributed to the continuous supply of large amounts and diverse biomass-C inputs to the soil surface, through the diversity of the root systems along with the low level of soil disturbance under NT systems (Sá et al., 2014; Briedis et al., 2018). During the decomposition process, microbial communities use the rapidly decomposable materials as energy sources, while the recalcitrant and other labile compounds materials act as the glue to bind soil mineral particles together (Witzgall et al., 2021). This process is a pathway for the formation of soil micro aggregates (Bot and Benites, 2005). The continuous supply of biomass C inputs to the soil associated with microbial decomposition without soil mechanical disturbance creates a favourable environment for the emergence of soil macroaggregates (Crews and Rumsey, 2017). Organic carbon inside soil aggregates is physically protected from microbial oxidation as well as strongly associated with the organo-minerals, leading to SOC stabilization over time (Powlson et al., 1987; Lützow et al., 2006). In the same

experiments as in our study but after 3 years of CA adoption, Hok et al., (2021) reported that soil aggregation was one of the main stabilization mechanisms, providing physical protection to the newly derived C into the soil microaggregates protected by macroaggregates. From our knowledge of the literature, the high SOC accumulation rate recorded under cassava-based CA cropping systems is relatively unique and, in addition to the residues of cover crops and maize under the bi-annual crop rotation system, the nature of the cassava residues that was retained into the field with high cellulose and lignin contents may explain this result (Veiga et al., 2016). From an incubation of labelled litter, Witzgall et al., (2021) found that the occlusion of organic matter into aggregates and the formation of organo-mineral associations occurs concurrently on fresh litter surfaces regardless of soil structure. In addition, the increase in C-MAOM is attributed to C transfer from POM and other labile C pools. Over time, these compounds are transferred to more stable pools, creating associations with mineral colloids, with MAOM being more stabilized (Briedis et al., 2018). Rosolem et al. (2016) conducted 3-year successive experiments to assess the above- and belowground effects of a wide range of tropical grasses and legume cover crops, which were the same species that were used under the NT systems in our experiments, in combination with no-till soybean-based cropping systems in Brazilian tropical clayey Rhodic Ferralsol on total organic C and N stocks and in POM fraction. They reported that the presence of C4 deep-root grass cover crops during the fallow period significantly increased total organic C and POM. Furthermore, legume cover crops contributed to maintaining the C/N ratio in the topsoil layers, which could keep increasing C over time. Beside the aboveground biomass, root systems of cover crops are also an important C inputs for SOC accumulation due to their capacity to grow deeper in the soil profile, during the dry season for some, exploring large volume of soil (Rosolem

780

781

782

783

784

785

786

787

788

789

790

791

792

793

794

795

796

797

798

799

et al., 2016; Sokol et al., 2019), releasing large quantities of roots exudates, and recycling nutrients (Rosolem et al., 2005). The increase of C stock in POM and MAOM shows that NT systems with the use of cover crops is a key strategy to promote both SOC storage and long-term SOC stabilization. In contrast to C-POM and C-MAOM, although the significant increase of N-POM stock in the top 0–10 cm was observed under all the NT systems in MaiEx and SoyEx, the depletion of N-MAOM stock was observed below 5 cm in CasEx and below 40 cm in MaiEx and SoyEx. This raises questions about the N dynamics and N supplies through the use of mineral fertilizers, as well as N fixation through the use of legume crops in the NT cropping systems. Therefore, there is a need to conduct further research on N use efficiency, N cycles, and nutrient availability and their stoichiometry relationship by considering deeper layers (> 100 cm) to understand the mechanism driving N loss under NT systems in these long-term experiments. 

#### 5 Conclusion

The present study showed that, over 10 years, effects of NT systems on SOC and TN stocks and fractions varied across the three NT systems and the experiments. All the NT cropping systems significantly increased SOC stock in the surface layers in SoyEx and in deeper soil layers under MaiEx and CasEx. When considering the whole profile (0–100 cm), the annual SOC accumulation rates in NT systems ranged from 0.86–1.47 and 0.70–1.07 Mg C ha<sup>-1</sup> yr<sup>-1</sup> in MaiEx and CasEx, respectively. Similarly, under all NT systems, increases in C-POM and C-MAOM stocks were observed in the topsoil layers in MaiEx and SoyEx and in C-MAOM stock in 0–40 cm in CasEx. However, under all the NT systems, N-POM

820 stock only increased in the surface 0–10 cm layer, but N-MAOM stock decreased below 5 cm in CasEx 821 and below 40 cm in MaiEx and SoyEx. 822 Overall, our findings reveal that diachronic sampling is crucial for proper measurements of the impacts 823 of NT systems on SOC dynamics with time. Long-term adoption of NT cropping systems accompanied 824 by diversified crop and cover crop species significantly increased SOC stock and fractions in the tropical 825 red Oxisol of Cambodia. The study highlights the potential of NT cropping systems for SOC accumulation 826 and stabilization over time, even for cassava, which is known to induce soil degradation, but raises 827 questions about soil N dynamics. Further research on the N dynamics is needed to understand the 828 mechanism driving N loss in NT systems for making informed decisions regarding sustainable soil 829 fertility management and crop production systems.

# 6 Code and data availability

- All data are freely available on the CIRAD data repository https://doi.org/10.18167/DVN1/NNBBAQ
- 832 (Leng et al., 2024).

830

833

834

835

836

837

#### 7 Author contributions

VL co-established and managed the experiments, carried out the fieldworks, managed all sample collection campaigns and laboratory works, and wrote the first manuscript draft. FT co-managed the experiments, provided support to the team for the last 8 years and raised funds along with GDA/DALRM to sustain these long-term experiments. LT, RC, and FT conceptualized the research, supervised the field

operations, analytical procedure, data computation, and reviewed the manuscript. VS, PL, LH, JS and CB gave advices for the analytical procedures, data calculation and manuscript improvement. PM significantly contributed to the implementation of the field operations, sample collection and lab works. TF performed the statistical analysis. SB raised funds, designed, established and managed the experiments during the first years and contributed to the manuscript improvement.

### Competing interests

One co-author is a member of the editorial board of SOIL.

# 9 Acknowledgements

This research has been fully funded by the Agroecology and Safe Food System Transitions in Southeast Asia (ASSET) project. The project is co-financed by French Development Agency (AFD), European Union (EU), and French Facility for Global Environment (FFEM) under the conventions CZZ2453 and CZZ2868. We thank the DALRM/GDA/MAFF for providing access to the experiments and supporting the paperwork authorization for sample shipments to Switzerland. We further acknowledge CIRAD, France, for its strong collaboration with GDA and for continuing to provide the technical support for maintaining these long-term experiments. Additional thanks to CARDEC's team, who manages the field operations. Special appreciation to Britta Jahn-Humphrey of the Sustainable Agroecosystem, ETH Zurich, for processing paperwork and overseeing the measurement of all the SOC and TN.

# 855 10 References

- ADB: Cambodia Agriculture, Natural Resources, and Rural Development Sector Assessment, Strategy, and Road Map, Asian Development Bank, Manila, Philippines, https://doi.org/10.22617/TCS210256-2, 2021.
- Adjei-Nsiah, S., Kuyper, T. W., Leeuwis, C., Abekoe, M. K., and Giller, K. E.: Evaluating sustainable and profitable cropping sequences with cassava and four legume crops: Effects on soil fertility and maize yields in the forest/savannah transitional agro-ecological zone of Ghana, Field Crops Research, 103, 87–97, https://doi.org/10.1016/j.fcr.2007.05.001, 2007.
- Angers, D. A. and Eriksen-Hamel, N. S.: Full-Inversion Tillage and Organic Carbon Distribution in Soil
  Profiles: A Meta-Analysis, Soil Science Soc of Amer J, 72, 1370–1374,
  https://doi.org/10.2136/sssaj2007.0342, 2008.
- Balota, E. L., Lopes, E. S., Hungria, M., and Döbereiner, J.: Occurence of diazotrophic and arbuscular mycorhizal fungi on the cassava crop, Pesq. agropec. bras., 34, 1265–1276, https://doi.org/10.1590/S0100-204X1999000700020, 1999.
- Barbier, E. B. and Di Falco, S.: Rural Populations, Land Degradation, and Living Standards in
   Developing Countries, Review of Environmental Economics and Policy, 15, 115–133,
   https://doi.org/10.1086/713152, 2021.
- 872 Barbier, E. B. and Hochard, J. P.: Land degradation and poverty, Nat Sustain, 1, 623–631, https://doi.org/10.1038/s41893-018-0155-4, 2018.
- Bayer, C., Martin-Neto, L., Mielniczuk, J., Pavinato, A., and Dieckow, J.: Carbon sequestration in two
  Brazilian Cerrado soils under no-till, Soil and Tillage Research, 86, 237–245,
  https://doi.org/10.1016/j.still.2005.02.023, 2006.
- Beillouin, D., Corbeels, M., Demenois, J., Berre, D., Boyer, A., Fallot, A., Feder, F., and Cardinael, R.:
  A global meta-analysis of soil organic carbon in the Anthropocene, Nat Commun, 14, 3700,
  https://doi.org/10.1038/s41467-023-39338-z, 2023.
- Bernoux, M., Feller, C., Cerri, C. C., Eschenbrenner, V., and Cerri, C. E. P.: Soil Carbon Sequestration,
   in: SOIL EROSION AND CARBON DYNAMICS, CRC Press, Boca Raton, 13–22, 2005.
- Blanco-Canqui, H. and Lal, R.: No-Tillage and Soil-Profile Carbon Sequestration: An On-Farm
  Assessment, Soil Science Soc of Amer J, 72, 693–701, https://doi.org/10.2136/sssaj2007.0233,
  2008.
- Blanco-Canqui, H., Schlegel, A. J., and Heer, W. F.: Soil-profile distribution of carbon and associated properties in no-till along a precipitation gradient in the central Great Plain, Agriculture, Ecosystems & Environment, 144, 107–116, https://doi.org/10.1016/j.agee.2011.07.004, 2011.
- Blanco-Moure, N., Gracia, R., Bielsa, C., and Lopez, M. V.: Long-term no-tillage effects on particulate and mineral-associated soil organic matter under rainfed Mediterranean conditions, Soil Use and Management, 29, 250–259, https://doi.org/doi: 10.1111/sum.12039, 2013.
- Boddey, R. M., Jantalia, C. P., Conceição, P. C., Zanatta, J. A., Bayer, C., Mielniczuk, J., Dieckow, J.,
- Dos Santos, H. P., Denardin, J. E., Aita, C., Giacomini, S. J., Alves, B. J. R., and Urquiaga, S.:
  Carbon accumulation at depth in Ferralsols under zero-till subtropical agriculture, Global Change
- Biology, 16, 784–795, https://doi.org/10.1111/j.1365-2486.2009.02020.x, 2010.

- 895 Bot, A. and Benites, J.: The importance of soil organic matter: key to drought-resistant soil and sustained food and production, Food and Agriculture Organization of the United Nations, Rome, 78 pp., 2005.
- Briedis, C., De Moraes Sá, J. C., Lal, R., Tivet, F., Franchini, J. C., De Oliveira Ferreira, A., Da Cruz
  Hartman, D., Schimiguel, R., Bressan, P. T., Inagaki, T. M., Romaniw, J., and Gonçalves, D. R.
  P.: How does no-till deliver carbon stabilization and saturation in highly weathered soils?
  CATENA, 163, 13–23, https://doi.org/10.1016/j.catena.2017.12.003, 2018.
- 902 CARDI: Annual report 2016, Cambodian Agricultural Research and Development Institue (CARDI), 903 Phnom Penh, Cambodia, 2017.
- Castro, G. S. A., Crusciol, C. A. C., Calonego, J. C., and Rosolem, C. A.: Management Impacts on Soil
   Organic Matter of Tropical Soils, Vadose Zone Journal, 14, vzj2014.07.0093,
   https://doi.org/10.2136/vzj2014.07.0093, 2015.
- Congreves, K. A., Hooker, D. C., Hayes, A., Verhallen, E. A., and Van Eerd, L. L.: Interaction of longterm nitrogen fertilizer application, crop rotation, and tillage system on soil carbon and nitrogen dynamics, Plant Soil, 410, 113–127, https://doi.org/10.1007/s11104-016-2986-y, 2017.
- Cooper, H. V., Sjögersten, S., Lark, R. M., Girkin, N. T., Vane, C. H., Calonego, J. C., Rosolem, C.,
  and Mooney, S. J.: Long-term zero-tillage enhances the protection of soil carbon in tropical
  agriculture, European journal of soil science, 72, 2477–2492, https://doi.org/10.1111/ejss.13111,
  2021.
- Corbeels, M., Cardinael, R., Naudin, K., Guibert, H., and Torquebiau, E.: The 4 per 1000 goal and soil carbon storage under agroforestry and conservation agriculture systems in sub-Saharan Africa, Soil and Tillage Research, 188, 16–26, https://doi.org/10.1016/j.still.2018.02.015, 2019.
- Craswell, E. T. and Lefroy, R. D. B.: The role and function of organic matter in tropical soils, in:
  Managing Organic Matter in Tropical Soils: Scope and Limitations, edited by: Martius, C.,
  Tiessen, H., and Vlek, P. L. G., Springer Netherlands, Dordrecht, 7–18,
  https://doi.org/10.1007/978-94-017-2172-1 2, 2001.
- Crews, T. E. and Rumsey, B. E.: What Agriculture Can Learn from Native Ecosystems in Building Soil Organic Matter: A Review, Sustainability, 9, https://doi.org/10.3390/su9040578, 2017.
  - De Vries, W.: Nutrients trigger carbon storage, Nature Clim Change, 4, 425–426, https://doi.org/10.1038/nclimate2255, 2014.

923

- Delgado, J. A.: Long-term nitrogen balance of an irrigated no-till soil-corn system, Nutr Cycl
   Agroecosyst, 126, 229–243, https://doi.org/10.1007/s10705-023-10287-9, 2023.
- Diekow, J., Mielniczuk, J., Knicker, H., Bayer, C., Dick, D. P., and Kögel-Knabner, I.: Soil C and N stocks as affected by cropping systems and nitrogen fertilisation in a southern Brazil Acrisol managed under no-tillage for 17 years, Soil and Tillage Research, 81, 87–95, https://doi.org/10.1016/j.still.2004.05.003, 2005.
- 931 Dragović, N. and Vulević, T.: Soil Degradation Processes, Causes, and Assessment Approaches, in:
- Life on Land, edited by: Leal Filho, W., Azul, A. M., Brandli, L., Lange Salvia, A., and Wall, T.,
- 933 Springer International Publishing, Cham, 1–12, https://doi.org/10.1007/978-3-319-71065-5\_86-1, 2020.
- 935 Du, Z., Angers, D. A., Ren, T., Zhang, Q., and Li, G.: The effect of no-till on organic C storage in

- Chinese soils should not be overemphasized: A meta-analysis, Agriculture, Ecosystems & Environment, 236, 1–11, https://doi.org/10.1016/j.agee.2016.11.007, 2017.
- Ellert, B. H. and Bettany, J. R.: Calculation of organic matter and nutrients stored in soils under contrasting management regimes, Can. J. Soil. Sci., 75, 529–538, https://doi.org/10.4141/cjss95-075, 1995.
- Fermont, A. M., Van Asten, P. J. A., and Giller, K. E.: Increasing land pressure in East Africa: The changing role of cassava and consequences for sustainability of farming systems, Agriculture, Ecosystems & Environment, 128, 239–250, https://doi.org/10.1016/j.agee.2008.06.009, 2008.
- Fowler, A. F., Basso, B., Millar, N., and Brinton, W. F.: A simple soil mass correction for a more
   accurate determination of soil carbon stock changes, Sci Rep, 13, 2242,
   https://doi.org/10.1038/s41598-023-29289-2, 2023.
- Frasier, I., Noellemeyer, E., Figuerola, E., Erijman, L., Permingeat, H., and Quiroga, A.: High quality residues from cover crops favor changes in microbial community and enhance C and N sequestration, Global Ecology and Conservation, 6, 242–256, https://doi.org/10.1016/j.gecco.2016.03.009, 2016.
- Fujisaki, K., Chevallier, T., Chapuis-Lardy, L., Albrecht, A., Razafimbelo, T., Masse, D., Ndour, Y. B.,
   and Chotte, J.-L.: Soil carbon stock changes in tropical croplands are mainly driven by carbon
   inputs: A synthesis, Agriculture, Ecosystems & Environment, 259, 147–158,
   https://doi.org/10.1016/j.agee.2017.12.008, 2018.
- Hok, L., De Moraes Sá, J. C., Boulakia, S., Reyes, M., Leng, V., Kong, R., Tivet, F. E., Briedis, C.,
   Hartman, D., Ferreira, L. A., Magno, T., and Pheav, S.: Short-term conservation agriculture and
   biomass-C input impacts on soil C dynamics in a savanna ecosystem in Cambodia, Agriculture,
   Ecosystems & Environment, 214, 54–67, https://doi.org/10.1016/j.agee.2015.08.013, 2015.
- Hok, L., De Moraes Sá, J. C., Reyes, M., Boulakia, S., Tivet, F., Leng, V., Kong, R., Briedis, C., Da
  Cruz Hartman, D., Ferreira, L. A., Inagaki, T. M., Gonçalves, D. R. P., and Bressan, P. T.:
  Enzymes and C pools as indicators of C build up in short-term conservation agriculture in a
  savanna ecosystem in Cambodia, Soil and Tillage Research, 177, 125–133,
  https://doi.org/10.1016/j.still.2017.11.015, 2018.
- Hok, L., De Moraes Sá, J. C., Boulakia, S., Reyes, M., De Oliveira Ferreira, A., Elie Tivet, F., Saab, S.,
   Auccaise, R., Massao Inagaki, T., Schimiguel, R., Aparecida Ferreira, L., Briedis, C., Santos
   Canalli, L. B., Kong, R., and Leng, V.: Dynamics of soil aggregate-associated organic carbon
   based on diversity and high biomass-C input under conservation agriculture in a savanna
   ecosystem in Cambodia, CATENA, 198, 105065, https://doi.org/10.1016/j.catena.2020.105065,
   2021.
- Howeler, R. H. and Sieverding, E.: Potentials and limitations of mycorrhizal inoculation illustrated by
   experiments with field-grown cassava, Plant Soil, 75, 245–261,
   https://doi.org/10.1007/BF02375570, 1983.
- Howeler, R. H., Cadavid, L. F., and Burckhardt, E.: Response of cassava to VA mycorrhizal inoculation and phosphorus application in greenhouse and field experiments, Plant Soil, 69, 327–339, https://doi.org/10.1007/BF02372454, 1982.
- 976 IUSS Working Group WRB: World reference base for soil resources 2014: international soil

- classification system for naming soils and creating legends for soil maps, FAO, Rome, 2015.
- 978 Junior, C. C., Corbeels, M., Bernoux, M., Píccolo, M. C., Siqueira Neto, M., Feigl, B. J., Cerri, C. E. P.,
- 979 Cerri, C. C., Scopel, E., and Lal, R.: Assessing soil carbon storage rates under no-tillage:
- Comparing the synchronic and diachronic approaches, Soil and Tillage Research, 134, 207–212, https://doi.org/10.1016/j.still.2013.08.010, 2013.
- Kan, Z.-R., Liu, W.-Z., Liu, W.-S., Lal, R., Dang, Y. P., Zhao, X., and Zhang, H.-L.: Mechanisms of
   soil organic carbon stability and its response to no-till: A global synthesis and perspective, Global
   Change Biology, 00, 1–18, https://doi.org/DOI: 10.1111/gcb.15968, 2021.
- Ken, S., Sasaki, N., Entani, T., Ma, H. O., Thuch, P., and Tsusaka, T. W.: Assessment of the Local
   Perceptions on the Drivers of Deforestation and Forest Degradation, Agents of Drivers, and
   Appropriate Activities in Cambodia, Sustainability, 12, 9987, https://doi.org/10.3390/su12239987,
   2020.
- Kirkby, C. A., Richardson, A. E., Wade, L. J., Batten, G. D., Blanchard, C., and Kirkegaard, J. A.:
   Carbon-nutrient stoichiometry to increase soil carbon sequestration, Soil Biology and
   Biochemistry, 60, 77–86, https://doi.org/10.1016/j.soilbio.2013.01.011, 2013.
- Kirkby, C. A., Richardson, A. E., Wade, L. J., Passioura, J. B., Batten, G. D., Blanchard, C., and
   Kirkegaard, J. A.: Nutrient availability limits carbon sequestration in arable soils, Soil Biology
   and Biochemistry, 68, 402–409, https://doi.org/10.1016/j.soilbio.2013.09.032, 2014.
- Koun, P., Vernet, P., Filloux, T., Sar, V., Seng, V., Srimongkol, P., Tantachasatid, P., Sen, R., Pheap,
   S., Tivet, F., and Thoumazeau, A.: Early effects of conservation agriculture on soil organic carbon dynamics of Mollisols in Cambodia, Soil Use and Management, sum.12922,
   https://doi.org/10.1111/sum.12922, 2023.
- 999 Kuzyakov, Y.: Review: Factors affecting rhizosphere priming effects, Z. Pflanzenernähr. Bodenk., 165, 1000 382–396, <a href="https://doi.org/10.1002/1522-2624(200208)165:4<382">https://doi.org/10.1002/1522-2624(200208)165:4<382</a>: AID-JPLN382>3.0.CO;2-#, 2002.
- 1002 Kuzyakov, Y., Friedel, J. K., and Stahr, K.: Review of mechanisms and quanti®cation of priming effects, Soil Biology, 2000.
- Lal, R.: Restoring Soil Quality to Mitigate Soil Degradation, Sustainability, 7, 5875–5895, https://doi.org/10.3390/su7055875, 2015a.
- Lal, R.: Sequestering carbon and increasing productivity by conservation agriculture, Journal of Soil and Water Conservation, 70, 55A-62A, https://doi.org/10.2489/jswc.70.3.55A, 2015b.
- Lal, R.: Digging deeper: A holistic perspective of factors affecting soil organic carbon sequestration in agroecosystems, Global Change Biology, 24, 3285–3301, https://doi.org/10.1111/gcb.14054, 2018.
- Lavallee, J. M., Soong, J. L., and Cotrufo, M. F.: Conceptualizing soil organic matter into particulate and mineral-associated forms to address global change in the 21st century, Global Change Biology, 26, 261–273, https://doi.org/DOI: 10.1111/gcb.14859, 2020.
- Lavelle, P., Spain, A., Blouin, M., Brown, G., Decaëns, T., Grimaldi, M., Jiménez, J. J., McKey, D.,
- Mathieu, J., Velasquez, E., and Zangerlé, A.: Ecosystem Engineers in a Self-organized Soil: A
- 1016 Review of Concepts and Future Research Questions, Soil Science, 181, 91–109,
- 1017 https://doi.org/10.1097/SS.000000000000155, 2016.

- Leng, V., Cardinael, R., Tivet, F. E., Seng, V., Mark, P., Lienhard, P., Filloux, T., Six, J., Hok, L.,
- Boulakia, S., Briedis, C., Sá, D. M. J. C., and Thuriès, L.: Data for "Diachronic assessment of soil
- organic C and N dynamics under long-term no-till cropping systems in the tropical upland of Cambodia," https://doi.org/10.18167/DVN1/NNBBAQ, 2024.
- Li, G., Yu, C., Shen, P., Hou, Y., Ren, Z., Li, N., Liao, Y., Li, T., and Wen, X.: Crop diversification
- promotes soil aggregation and carbon accumulation in global agroecosystems: A meta-analysis, Journal of Environmental Management, 350, 119661,
- 1025 https://doi.org/10.1016/j.jenvman.2023.119661, 2024.
- Liang, C., Cheng, G., Wixon, D. L., and Balser, T. C.: An Absorbing Markov Chain approach to understanding the microbial role in soil carbon stabilization, Biogeochemistry, 106, 303–309, https://doi.org/DOI 10.1007/s10533-010-9525-3, 2011.
- Lienhard, P., Tivet, F., Chabanne, A., Dequiedt, S., Lelièvre, M., Sayphoummie, S., Leudphanane, B.,
  Prévost-Bouré, N. C., Séguy, L., Maron, P.-A., and Ranjard, L.: No-till and cover crops shift soil
  microbial abundance and diversity in Laos tropical grasslands, Agron. Sustain. Dev., 33, 375–384,
  https://doi.org/10.1007/s13593-012-0099-4, 2013.
- Liu, X., Wu, X., Liang, G., Zhang, M., Li, S., and Zheng, F.: A global meta-analysis of the impacts of no-tillage on soil aggregation and aggregate-associated organic carbon, Land Degradation & Development, 32, 5292–5305, https://doi.org/DOI: 10.1002/ldr.4109, 2021.
- Lorenz, K. and Lal, R.: The Depth Distribution of Soil Organic Carbon in Relation to Land Use and Management and the Potential of Carbon Sequestration in Subsoil Horizons, in: Advances in Agronomy, vol. 88, Elsevier, 35–66, https://doi.org/10.1016/S0065-2113(05)88002-2, 2005.
- Luo, Z., Wang, E., and Sun, O. J.: Can no-tillage stimulate carbon sequestration in agricultural soils? A meta-analysis of paired experiments, Agriculture, Ecosystems & Environment, 139, 224–231, https://doi.org/10.1016/j.agee.2010.08.006, 2010.
- Lützow, M. V., Kögel-Knabner, I., Ekschmitt, K., Matzner, E., Guggenberger, G., Marschner, B., and Flessa, H.: Stabilization of organic matter in temperate soils: mechanisms and their relevance under different soil conditions a review, European J Soil Science, 57, 426–445, https://doi.org/10.1111/j.1365-2389.2006.00809.x, 2006.
- 1046 MAFF: Land Degradation Neutrality Targets, 2018.
- Morugán-Coronado, A.: The impact of crop diversification, tillage and fertilization type on soil total microbial, fungal and bacterial abundance: A worldwide meta-analysis of agricultural sites, 2022.
- Neto, M. S., Scopel, E., Corbeels, M., Cardoso, A. N., Douzet, J.-M., Feller, C., Piccolo, M. D. C.,
  Cerri, C. C., and Bernoux, M.: Soil carbon stocks under no-tillage mulch-based cropping systems
- in the Brazilian Cerrado: An on-farm synchronic assessment, Soil and Tillage Research, 110, 187–195, https://doi.org/10.1016/j.still.2010.07.010, 2010.
- Nkonya, E., Mirzabaev, A., and Von Braun, J. (Eds.): Economics of Land Degradation and Improvement – A Global Assessment for Sustainable Development, Springer International Publishing, Cham, https://doi.org/10.1007/978-3-319-19168-3, 2016.
- Nut, N., Mihara, M., Jeong, J., Ngo, B., Sigua, G., Prasad, P. V. V., and Reyes, M. R.: Land Use and
- Land Cover Changes and Its Impact on Soil Erosion in Stung Sangkae Catchment of Cambodia,
- 1058 Sustainability, 13, 9276, https://doi.org/10.3390/su13169276, 2021.

- 1059 Obalum, S. E., Chibuike, G. U., Peth, S., and Ouyang, Y.: Soil organic matter as sole indicator of soil degradation, Environ Monit Assess, 189, 176, https://doi.org/10.1007/s10661-017-5881-v, 2017. 1060
- 1061 Ogle, S. M., Swan, A., and Paustian, K.: No-till management impacts on crop productivity, carbon input 1062 and soil carbon sequestration, Agriculture, Ecosystems & Environment, 149, 37–49, 1063 https://doi.org/10.1016/j.agee.2011.12.010, 2012.
- 1064 Oliveira, F. C. C., Ferreira, G. W. D., Souza, J. L. S., Vieira, M. E. O., and Pedrotti, A.: Soil physical 1065 properties and soil organic carbon content in northeast Brazil: long-term tillage systems effects, 1066 Sci. agric. (Piracicaba, Braz.), 77, e20180166, https://doi.org/10.1590/1678-992x-2018-0166. 1067 2020.

1068 1069

1081

- Olson, K., Ebelhar, S. A., and Lang, J. M.: Long-Term Effects of Cover Crops on Crop Yields, Soil Organic Carbon Stocks and Sequestration, OJSS, 04, 284–292, https://doi.org/10.4236/ojss.2014.48030, 2014. 1070
- 1071 Paustian, K., Andrén, O., Janzen, H. H., Lal, R., Smith, P., Tian, G., Tiessen, H., Van Noordwijk, M., 1072 and Woomer, P. L.: Agricultural soils as a sink to mitigate CO 2 emissions, Soil Use and 1073 Management, 13, 230–244, https://doi.org/10.1111/j.1475-2743.1997.tb00594.x, 1997.
- 1074 Pheap, S., Lefèvre, C., Thoumazeau, A., Leng, V., Boulakia, S., Koy, R., Hok, L., Lienhard, P., 1075 Brauman, A., and Tivet, F.: Multi-functional assessment of soil health under Conservation 1076 Agriculture in Cambodia, Soil and Tillage Research, 194, 104349, 1077 https://doi.org/10.1016/j.still.2019.104349, 2019.
- 1078 Poeplau, C. and Don, A.: Carbon sequestration in agricultural soils via cultivation of cover crops – A 1079 meta-analysis, Agriculture, Ecosystems & Environment, 200, 33–41, https://doi.org/10.1016/j.agee.2014.10.024, 2015. 1080
  - Polyakov, V. and Lal, R.: Modeling soil organic matter dynamics as affected by soil water erosion, Environment International, 30, 547–556, https://doi.org/10.1016/j.envint.2003.10.011, 2004.
- 1083 Powlson, D. S., Prookes, P. C., and Christensen, B. T.: Measurement of soil microbial biomass provides 1084 an early indication of changes in total soil organic matter due to straw incorporation, Soil Biology and Biochemistry, 19, 159–164, https://doi.org/10.1016/0038-0717(87)90076-9, 1987. 1085
- Powlson, D. S., Stirling, C. M., Thierfelder, C., White, R. P., and Jat, M. L.: Does conservation 1086 1087 agriculture deliver climate change mitigation through soil carbon sequestration in tropical agro-1088 ecosystems? Agriculture, Ecosystems & Environment, 220, 164–174, 1089 https://doi.org/10.1016/j.agee.2016.01.005, 2016.
- 1090 Prommer, J., Walker, T. W. N., Wanek, W., Braun, J., Zezula, D., Hu, Y., Hofhansl, F., and Richter, A.: Increased microbial growth, biomass, and turnover drive soil organic carbon accumulation at 1091 1092 higher plant diversity, Global Change Biology, 26, 669–681, https://doi.org/DOI: 1093 10.1111/gcb.14777, 2019.
- Rodrigues, A. T. L., Giacomini, S. J., Dieckow, J., Cherubin, M. R., Sangiovo Ottonelli, A., and Bayer, 1094 1095 C.: Carbon saturation deficit and litter quality drive the stabilization of litter-derived C in mineralassociated organic matter in long-term no-till soil, CATENA, 219, 106590, 1096 1097 https://doi.org/10.1016/j.catena.2022.106590, 2022.
- 1098 Rosolem, C. A., Li, Y., and Garcia, R. A.: Soil carbon as affected by cover crops under no-till under tropical climate, Soil Use and Management, 32, 495–503, https://doi.org/10.1111/sum.12309, 1099

1100 2016.

1114

1115

- 1101 Sá, D. M. J. C., Tivet, F., Lal, R., Briedis, C., Hartman, D. C., Dos Santos, J. Z., and Dos Santos, J. B.:
- Long-term tillage systems impacts on soil C dynamics, soil resilience and agronomic productivity
- of a Brazilian Oxisol, Soil and Tillage Research, 136, 38–50,
- https://doi.org/10.1016/j.still.2013.09.010, 2014.
- Saïdou, A., Kuyper, T. W., Kossou, D. K., Tossou, R., and Richards, P.: Sustainable soil fertility management in Benin: learning from farmers, NJAS: Wageningen Journal of Life Sciences, 52, 349–369, https://doi.org/10.1016/S1573-5214(04)80021-6, 2004.
- Sar, V.: Biofunctool® Approach Assessing Soil Quality under Conservation Agriculture and Conventional Tillage for Rain- fed Lowland Rice Systems in Cambodia, 2021.
- Séguy, L., Bouzinac, S., and Husson, O.: Direct-seeded tropical soil systems with permanent soil cover: Learning from Brazilian experience, in: Biological approaches to sustainable soil systems, edited by: Uphoff, N., Ball, S. A., Fernandes, E., Herren, H., Husson, O., Laing, M., Palm, C., Pretty, J., Sanchez, P., Sanginga, N., and Thies, J., Taylor and Francis, New York, USA, 323–342, 2006.
  - Schmidt, R., Mitchell, J., and Scow, K.: Cover cropping and no-till increase diversity and symbiotroph:saprotroph ratios of soil fungal communities, Soil Biology and Biochemistry, 129, 99–109, https://doi.org/10.1016/j.soilbio.2018.11.010, 2019.
- Shumba, A., Chikowo, R., Thierfelder, C., Corbeels, M., Six, J., and Cardinael, R.: Mulch application as the overarching factor explaining increase in soil organic carbon stocks under conservation agriculture in two 8-year-old experiments in Zimbabwe, SOIL, 10, 151–165, https://doi.org/10.5194/soil-10-151-2024, 2024.
- Sisti, C. P. J., Dos Santos, H. P., Kohhann, R., Alves, B. J. R., Urquiaga, S., and Boddey, R. M.: Change in carbon and nitrogen stocks in soil under 13 years of conventional or zero tillage in southern Brazil, Soil and Tillage Research, 76, 39–58, https://doi.org/10.1016/j.still.2003.08.007, 2004.
- Sithole, N. J., Magwaza, L. S., and Thibaud, G. R.: Long-term impact of no-till conservation agriculture and N-fertilizer on soil aggregate stability, infiltration and distribution of C in different size fractions, Soil and Tillage Research, 190, 147–156, https://doi.org/10.1016/j.still.2019.03.004, 2019.
- Six, J., Conant, R. T., Paul, E. A., and Paustian, K.: Stabilization mechanisms of soil organic matter: Implications for C-saturation of soils, Plant and soil, 241, 155–176, 2002.
- Six, J., Bossuyt, H., Degryze, S., and Denef, K.: A history of research on the link between (micro)aggregates, soil biota, and soil organic matter dynamics, 2004.
- Sokol, N. W., Kuebbing, Sara. E., Karlsen-Ayala, E., and Bradford, M. A.: Evidence for the primacy of living root inputs, not root or shoot litter, in forming soil organic carbon, New Phytologist, 221, 233–246, https://doi.org/10.1111/nph.15361, 2019.
- Sourn, T., Pok, S., Chou, P., Nut, N., Theng, D., and Prasad, P. V. V.: Assessment of Land Use and Land Cover Changes on Soil Erosion Using Remote Sensing, GIS and RUSLE Model: A Case Study of Battambang Province, Cambodia, Sustainability, 14, 4066, https://doi.org/10.3390/su14074066, 2022.
- Sourn, T., Pok, S., Chou, P., Nut, N., Theng, D., and Hin, L.: Assessing Land Use and Land Cover (LULC) Change and Factors Affecting Agricultural Land: Case Study in Battambang Province,

- 1141 Cambodia, Res. World Agric. Econ., 4, 41–54, https://doi.org/10.36956/rwae.v4i4.925, 2023.
- 1142 Stavi, I. and Lal, R.: Achieving Zero Net Land Degradation: Challenges and opportunities, Journal of 1143 Arid Environments, 112, 44–51, https://doi.org/10.1016/j.jaridenv.2014.01.016, 2015.
- 1144 Suong, M., Chapuis, E., Leng, V., Tivet, F., Waele, D. D., Thi, H. N., and Bellafiore, S.: Impact of a 1145 conservation agriculture system on soil characteristics, rice yield, and root-parasitic nematodes in 1146 a Cambodian lowland rice field. Journal of Nematology, 51, 1–15. 1147 https://doi.org/10.21307/jofnem-2019-085, 2019.
- TerAvest, D., Carpenter-Boggs, L., Thierfelder, C., and Reganold, J. P.: Crop production and soil water 1148 management in conservation agriculture, no-till, and conventional tillage systems in Malawi, 1149 1150 Agriculture, Ecosystems & Environment, 212, 285–296. https://doi.org/10.1016/j.agee.2015.07.011, 2015.
- Tiecher, T., Gubiani, E., Santanna, M. A., Veloso, M. G., Calegari, A., Canalli, L. B. D. S., Finckh, M. 1152 1153 R., Caner, L., and Rheinheimer, D. D. S.: Effect of 26-years of soil tillage systems and winter 1154 cover crops on C and N stocks in a Southern Brazilian Oxisol, Revista Brasileira de Ciência do Solo, 44, e0200029, https://doi.org/10.36783/18069657rbcs20200029, 2020. 1155
- 1156 Tiemann, L. K., Grandy, A. S., Atkinson, E. E., Marin-Spiotta, E., and McDaniel, M. D.: Crop 1157 rotational diversity enhances belowground communities and functions in an agroecosystem, 2015.
- Tivet, F., De Moraes Sá, J. C., Lal, R., Briedis, C., Borszowskei, P. R., Dos Santos, J. B., Farias, A., 1158 Eurich, G., Hartman, D. D. C., Nadolny Junior, M., Bouzinac, S., and Séguy, L.: Aggregate C 1159 1160 depletion by plowing and its restoration by diverse biomass-C inputs under no-till in sub-tropical and tropical regions of Brazil, Soil and Tillage Research, 126, 203-218, 1161 1162 https://doi.org/10.1016/j.still.2012.09.004, 2013.
  - UNCCD: Country profile of Cambodia. Investing in land degradation neutrality: Making te case. An overview of indicator and assessments, Global mechanism of the UNCCD, Bonn, Germany, 2018.
  - Veiga, J. P. S., Valle, T. L., Feltran, J. C., and Bizzo, W. A.: Characterization and productivity of cassava waste and its use as an energy source, Renewable Energy, 93, 691–699, https://doi.org/10.1016/j.renene.2016.02.078, 2016.
- 1168 Veloso, M. G., Angers, D. A., Tiecher, T., Giacomini, S., Dieckow, J., and Bayer, C.: High carbon 1169 storage in a previously degraded subtropical soil under no-tillage with legume cover crops, 1170 Agriculture, Ecosystems & Environment, 268, 15–23, https://doi.org/10.1016/j.agee.2018.08.024, 1171 2018.
- 1172 Virto, I., Barré, P., Burlot, A., and Chenu, C.: Carbon input differences as the main factor explaining the variability in soil organic C storage in no-tilled compared to inversion tilled agrosvstems. 1173 1174 Biogeochemistry, 108, 17–26, https://doi.org/10.1007/s10533-011-9600-4, 2012.
- Von Haden, A. C., Yang, W. H., and DeLucia, E. H.: Soils' dirty little secret: Depth-based comparisons 1175 1176 can be inadequate for quantifying changes in soil organic carbon and other mineral soil properties, 1177 Global Change Biology, 26, 3759–3770, https://doi.org/10.1111/gcb.15124, 2020.
- Witzgall, K., Vidal, A., Schubert, D. I., Höschen, C., Schweizer, S. A., Buegger, F., Pouteau, V., 1178
- 1179 Chenu, C., and Mueller, C. W.: Particulate organic matter as a functional soil component for
- 1180 persistent soil organic carbon, Nat Commun, 12, 4115, https://doi.org/10.1038/s41467-021-
- 24192-8, 2021. 1181

1151

1163

1164 1165

1166

- World Bank Group: Cambodia country climate and development report, The World Bank, Washinton DC, The US, https://doi.org/10.1596/978-1-4648-0483-0 ch2 EAP, 2023.
- Wuaden, C. R., Nicoloso, R. S., Barros, E. C., and Grave, R. A.: Early adoption of no-till mitigates soil organic carbon and nitrogen losses due to land use change, Soil and Tillage Research, 204, 104728, https://doi.org/10.1016/j.still.2020.104728, 2020.
- Xiao, L., Zhou, S., Zhao, R., Greenwood, P., and Kuhn, N. J.: Evaluating soil organic carbon stock
   changes induced by no-tillage based on fixed depth and equivalent soil mass approaches,
   Agriculture, Ecosystems & Environment, 300, 106982,
   https://doi.org/10.1016/j.agee.2020.106982, 2020.
- Zhang, W.-P., Fornara, D., Yang, H., Yu, R.-P., Callaway, R. M., and Li, L.: Plant litter strengthens
   positive biodiversity-ecosystem functioning relationships over time, Trends in Ecology &
   Evolution, 38, 473–484, https://doi.org/10.1016/j.tree.2022.12.008, 2023.
- Zhang, X., Gregory, A. S., Whalley, W. R., Coleman, K., Neal, A. L., Bacq-Labreuil, A., Mooney, S. J.,
   Crawford, J. W., Soga, K., and Illangasekare, T. H.: Relationship between soil carbon
   sequestration and the ability of soil aggregates to transport dissolved oxygen, Geoderma, 403,
   115370, https://doi.org/10.1016/j.geoderma.2021.115370, 2021.