

1 Reefal ostracod assemblages from the Zanzibar Archipelago (Tanzania)
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13

14 **Abstract**

15 Tropical reefs encompass tremendous biodiversity yet are imperiled by increasing
16 natural and anthropogenic disturbances worldwide. Meiobenthic biotas on coral reefs,
17 for example, ostracods, may experience substantial diversity loss and compositional
18 changes even before being examined. In this study, we investigated the reefal ostracod
19 assemblages from the highly diverse and productive ecosystem in the Zanzibar
20 Archipelago (Pemba, Zanzibar, and Mafia islands), Tanzania, to understand how their
21 diversity and faunal structure vary in response to water depth, benthic community type,
22 and human impacts. We characterized four distinct ostracod faunas associated with
23 different benthic habitats, which were deep fore reefs, shallow fringing reefs, degraded
24 fringing reefs, and ~~algae-at~~ covered intertidal flats. We identified typical ostracod
25 associations, i.e., Bairdiidae versus Loxoconchidae-Xestoleberididae, that showed
26 affinities to hard corals or algae on the reef platforms, respectively. Highest diversity
27 was found on shallow fringing reefs where ~~coral-affinedreefal~~ and ~~algae-affinedalgal~~
28 taxa exhibited maximum overlap of their distributional ranges, while the sand flats,
29 mangrove, and marginal reefs within the intertidal zone had much lower diversity with
30 high dominance of euryhaline taxa. Along the western coast of Zanzibar Island, coastal
31 development likely resulted in a unique faunal composition and comparatively low
32 diversity of ostracod assemblages among those in reefal habitats, in conjunction with
33 overall reef ecosystem degradation. This study represents the first large-scale
34 assessment of shallow-marine ostracods in the Zanzibar Archipelago. It lays a solid

35 foundation for future research into the ecological significance of ostracods on coral
36 reefs.

37

38 **1 Introduction**

39 Coral reefs as the most diverse ecosystem in the marine realm hold great ecological and
40 economic values, yet our knowledge of its enormous biodiversity is far from complete.
41 Compared with well-studied, conspicuous macrofauna (Souza et al., 2023), meiofauna
42 on coral reefs are highly under-represented in current research despite being
43 ecologically essential components and contributing significantly to total biodiversity
44 (Leray and Knowlton, 2015; Plaisance et al., 2011). Ostracoda (Crustacea) among all
45 meio-benthos has a tight association with reef environments tracing back to the lower
46 Paleozoic (Whatley and Watson, 1988). It is considered As-a useful model organism in
47 modern and paleo biodiversity research because of its high fossilization potential, high
48 abundance, and ubiquity in almost all marine ecosystems (Yasuhara et al., 2017).
49 However, ostracods on coral reefs are poorly understood. ~~Does~~ ostracod exhibit
50 higher diversity in reefal habitats compared with other soft sediment environments?
51 What are the characteristic ostracod taxa occupying different niches on coral reefs?
52 Answers to these questions are important for a holistic understanding of the reef
53 ecosystem and may hint at the underlying mechanisms that support such extraordinary
54 reef diversity. With intensifying anthropogenic disturbances at local to global scales,
55 the need to examine reefal ostracods before they perish is pressing.

56

57 Studies targeting tropical shallow-marine ostracods on coral reefs are surprisingly
58 deficient. Across the circumtropical belt, the central Indo-Pacific receives the most
59 attention for its diverse reefal ostracods, with pioneering studies identifying distinct
60 faunas associated with depth habitats from the shallow intertidal to deep reefal zones
61 (Whatley and Watson, 1988; Babinot and Degaugue-Michalski, 1996). Apart from
62 bathymetry, the distribution of reefal ostracods seems also related to benthic
63 community type (coral reefs versus seagrass/algal beds), sediment type (i.e., sandy
64 versus muddy deposits), in addition to local hydrology (i.e., exposure to wave energy)
65 (Weissleder et al., 1989; Whatley and Watson, 1988; Babinot and Degaugue-
66 Michalski, 1996; Tabuki, 1990, 1987). However, most of these works are confined to
67 small geographic areas and based on limited (sub)fossil materials. An extensive
68 regional-scale survey of reefal ostracods has never been conducted. More importantly,

69 the focus of previous studies mainly revolved around taxonomy, and biogeography to
70 a lesser degree, while quantitative assessments of biodiversity are largely lacking
71 (Tabuki, 1987, 1990; Mostafawi et al., 2005). The highest species richness ($S=74$) was
72 reported for a reef slope environment in Pulau Seribu, Java (Whatley and Watson, 1988)
73 in contrast to much lower values at lagoons ($S=27-42$) (Babinot and Degaugue-
74 Michalski, 1996; Weissleader et al., 1989) and reef flat ($S=34$) (Mostafawi et al., 2005).

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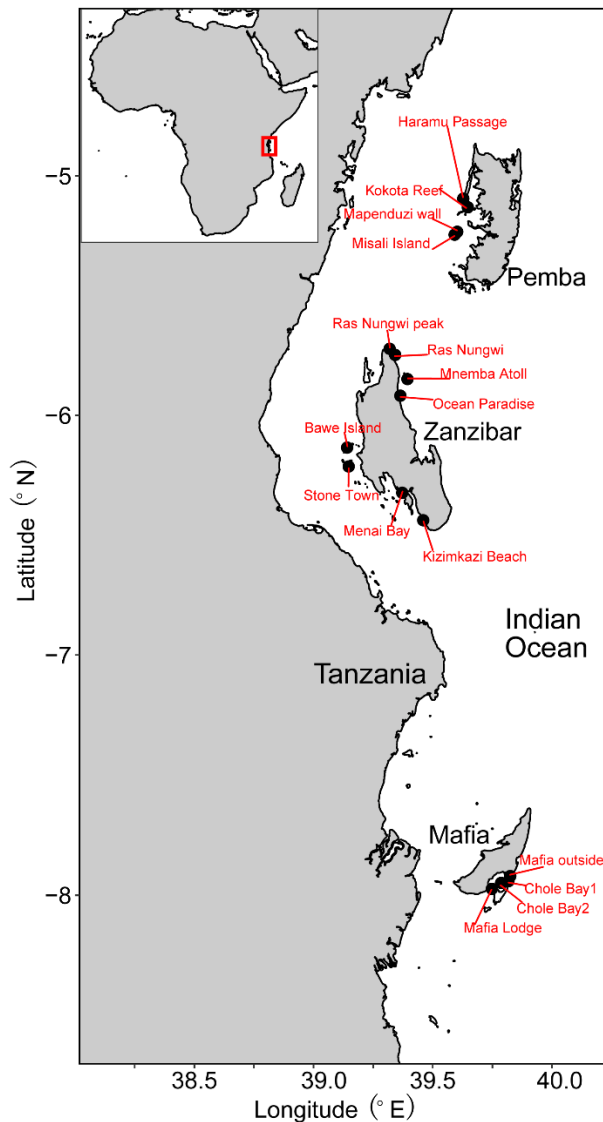
76 Reefal ostracods are even less known in other tropical regions outside of the central
77 Indo-Pacific. Along the eastern coast of Africa, where the reef ecosystem is productive
78 and biodiverse, the only studies on ostracod assemblages are perhaps Hartmann (1974)
79 and Jellinek (1993) that document more than 200 species inhabiting the algae facies
80 and reefal facies across the littoral zone in Kenya. Here we present the first large-scale
81 study on reefal ostracods from the Zanzibar Archipelago, Tanzania, a biodiversity
82 hotspot of great conservation interests and vulnerability to increasing anthropogenic
83 impacts (Grimsditch et al., 2009). We investigated the geographical structure of
84 ostracod diversity and composition in relation to environmental habitats among three
85 major islands of Pemba, Zanzibar, and Mafia. We compared the patterns with those of
86 benthic foraminifera (Thissen and Langer, 2017) to explore complex environmental
87 controls on the two groups of meio-benthos. This study is a major step towards better
88 understanding of tropical shallow-marine ostracods in eastern Africa and provides
89 valuable insight into the ostracod-reef association in general.

90

91 **2 Regional setting**

92 The Zanzibar Archipelago is located along the eastern coast of Tanzania in the Western
93 Indian Ocean (Fig. 1) (Thissen and Langer, 2017). It belongs to the eastern African
94 biogeographic province that stretches from Somalia to the northeastern coast of South
95 Africa (Costello et al., 2017; Obura, 2012). The archipelago is strongly influenced by
96 the warm, westward-flowing South Equatorial Current and the northward-flowing East
97 African Coastal Current (Narayan et al., 2022). The western coastlines are more
98 protected, with generally higher coral coverage, whereas the eastern coastlines are
99 exposed to large physical disturbances and strong wave energy (Thissen and Langer,
100 2017). Tides there are semi-diurnal, with a maximum range of 4.5 m and a neap tidal
101 range of 0.9 m (Thissen and Langer, 2017; Narayan et al., 2022). The islands possess a
102 great variety of benthic habitats from the littoral to open-water zone, with mangroves,

103 vegetated sand flats, and reef complexes. Reefs are mainly fringing reefs that are
104 situated on the narrow continental shelf (Mafia, Zanzibar) or are separated from the
105 African mainland by the deep Pemba channel (Pemba) (Thissen and Langer, 2017).
106 Noticeably, the major islands are subject to very different degrees of human exploration,
107 as Zanzibar is densely populated and highly urbanized while Mafia and Pemba are
108 largely uninhabited (Narayan et al., 2022). Stone Town and Bawe, in particular, are
109 faced with a direct discharge of untreated domestic sewage along the western coast of
110 Zanzibar Island, where moderate levels of reef deterioration have been found with
111 diversity decrease and coral cover loss (Bravo et al., 2021; Larsen et al., 2023).
112 Although extensive long-term monitoring is still lacking, previous studies indicate that
113 the Pemba reefs are likely in pristine conditions with the highest coverage of live hard
114 corals, while the Zanzibar reefs are often dominated by dead corals intermingled with
115 algae and seagrass habitats (Ussi et al., 2019; Larsen et al., 2023; Grimsditch et al.,
116 2009). No quantitative assessment of reef health has been conducted in Mafia Island,
117 unfortunately, but our field observations suggested moderate to good conditions at our
118 sampling sites.



119

120 Fig. 1. Locality map showing three major islands of the Zanzibar Archipelago with
 121 sample sites.

122

123 3 Materials and methods

124 3.1 Samples

125 26 surface sediment samples were collected from 16 sites during two field campaigns
 126 in 2005 at the islands of Zanzibar and Pemba, and in 2012 at Mafia Island
 127 ([supplementary tTable 1](#)). Depositional depths of all samples range from 0 to 42 m
 128 across the intertidal and subtidal zones. The selected sampling sites cover all major
 129 types of benthic habitats, including [a nearshore mangroves](#), coastal sand flats, and
 130 fringing-, fore-, and back-reefs ([Table 1](#)). [Note that the mangrove habitat may be](#)
 131 [underrepresented in current study as we have only one such site, however.](#) Samples
 132 were collected by SCUBA diving to [scrape along the seabed and](#) fill plastic [containers](#)

133 ~~bags~~ with surface sediments from the top 2 cm, in order to avoid the loss of finer
 134 particles due to suspension.

136 Most sampling sites were fine to medium-grained carbonate-rich, bioclastic sands and
 137 deposits with some reef rubble. Sediments were washed through a 63 µm sieve and
 138 oven dried at 50 °C. The residue was dry sieved over a 150 µm mesh sieve and ostracods
 139 were picked from the >150 µm size fraction, because smaller individuals are usually
 140 early juveniles that are not preserved and/or difficult to identify (Yasuhara et al., 2017).

141 ~~Sediment rich~~ Large volume samples were split into aliquot fractions using a
 142 microsplitter. The sample materials were primarily death assemblages though a very
 143 small number of specimens were preserved with soft parts (less than 1% among all
 144 observed individuals), indicating they were alive at the time of collection. Both live and
 145 dead specimens were included in the total count to represent time-averaged
 146 assemblages, which method effectively defines reef habitats and provides general
 147 environmental and diversity data useful in paleoecology (Glenn-Sullivan and Evans,
 148 2001; Langer and Lipps, 2003). A single valve or a carapace was considered as one
 149 individual, which is a standard counting method in ostracod research (Yasuhara et al.,
 150 2017). Selected specimens were imaged using a Scanning Electron Microscope (SEM).

156 Table 1. Ostracod assemblage information including raw species richness, number of
 157 counted individuals, and abundance per gram sediment, in addition to a characterization
 158 of benthic habitat in terms of sediment type and algae coverage in each location.

Sample	Species richness	<u>No. individuals</u>	<u>Abundance (per g)</u>	Habitat	<u>Sediment type</u>	<u>Algae coverage</u>
Haramu Passage20	37	69	<u>1.645</u>	fore reef	<u>bioclastic sand</u>	<u>Low</u>
Haramu Passage30	35	60	<u>4.212</u>	fore reef	<u>bioclastic sand</u>	<u>Low</u>
Kokota Reef25	64	235	<u>4.176</u>	fringing reef	<u>bioclastic sand</u>	<u>Low</u>
Kokota Reef16	78	364	<u>50.845</u>	fringing reef	<u>bioclastic sand</u>	<u>Low</u>
Mapenduzi wall40	60	235	<u>37.337</u>	fore reef	<u>bioclastic sand</u>	<u>Low</u>

Mapenduzi wall42	55	188	<u>22.212</u>	fore reef	<u>bioclastic sand</u>	<u>Low</u>
Misali Island20	65	254	<u>29.480</u>	fore reef	<u>bioclastic sand</u>	<u>Low</u>
Ras Nungwi peak12	56	296	<u>14.775</u>	fringing reef	<u>bioclastic sand</u>	<u>Medium</u>
Ras Nungwi peak12-14	46	116	<u>7.635</u>	fringing reef	<u>bioclastic sand</u>	<u>Medium</u>
Ras Nungwi peak20	81	311	<u>67.845</u>	fringing reef	<u>bioclastic sand</u>	<u>Low</u>
Ras Nungwi16	92	408	<u>40.674</u>	fringing reef	<u>bioclastic sand</u>	<u>Medium</u>
Ras Nungwi20	37	76	<u>16.497</u>	fringing reef	<u>bioclastic sand</u>	<u>Low</u>
Mnemba Atoll30	33	87	<u>45.218</u>	sand flat	<u>bioclastic sand</u>	<u>Medium</u>
Ocean Paradise3	46	231	<u>57.750</u>	back reef	<u>bioclastic sand</u>	<u>High</u>
Bawe Island9-30	80	410	<u>102.015</u>	fringing reef	<u>bioclastic sand</u>	<u>High</u>
Bawe Island grob	64	308	<u>13.077</u>	fringing reef	<u>bioclastic sand</u>	<u>High</u>
Stone Town12	77	519	<u>176.291</u>	fringing reef	<u>bioclastic sand</u>	<u>High</u>
Stone Town20	66	361	<u>158.542</u>	fringing reef	<u>bioclastic sand</u>	<u>High</u>
Menai Bay1	36	241	<u>21.294</u>	mangrove	<u>fine-grained sand</u>	<u>High</u>
Kizimkazi Beach1	24	59	<u>27.949</u>	sand flat	<u>fine-grained sand</u>	<u>High</u>
Mafia outside21	44	94	<u>20.764</u>	fore reef	<u>bioclastic sand</u>	<u>Medium</u>
Mafia outside20	82	347	<u>96.657</u>	fore reef	<u>bioclastic sand</u>	<u>Medium</u>
Chole Bay 1 (18-21)	27	74	<u>3.664</u>	Back reef	<u>bioclastic sand</u>	<u>Medium</u>
Chole Bay 2 (15-18)	77	241	<u>55.658</u>	fringing reef	<u>bioclastic sand</u>	<u>Medium</u>
Chole Bay 2 (20)	72	281	<u>69.383</u>	fringing reef	<u>bioclastic sand</u>	<u>Medium</u>
Mafia Lodge (0-3)	62	397	<u>65.576</u>	fringing reef	<u>fine-grained sand</u>	<u>High</u>

159

160 3.2 Quantitative analysis

161 We used Hill numbers (i.e., the effective number of equally abundant species)
162 parameterized by a diversity order q to estimate ostracod diversity in each sample and
163 island (Hill, 1973). Hill numbers have several major advantages over other diversity
164 indices and are increasingly adopted by ecologists (Chao et al., 2020). For example, the
165 Hill numbers will double when combining two identically distributed but distinct
166 communities, so they obey the “doubling property” and behave like species richness
167 (Chao et al., 2014b). In other words, the unit of Hill numbers is also “species” and thus
168 is more ecologically meaningful than other traditional diversity indices. Also, the order

169 q of the Hill numbers controls the sensitivity of the diversity metric to species relative
170 abundance. When the order $q=0$, Hill number (0D) reduces to species richness; when
171 the order $q=1$, Hill number (1D) measures the diversity of the abundant species; when
172 the order $q=2$, Hill number (2D) measures the diversity of dominant species (Chao et
173 al., 2014b). Therefore, besides species richness, the Hill numbers also estimate the
174 effective (or hypothetical) numbers of abundant and dominant species. Coincidentally,
175 the Hill numbers 1D and 2D are equivalent to the exponential of Shannon entropy and
176 Simpson index (hereafter referred to as Shannon and Simpson diversity), respectively
177 (Chao et al., 2014b), making them conceptually easy to understand by ecologists. To
178 make a fair comparison among multiple assemblages, we standardized the Hill numbers
179 with rarefaction or extrapolation to the largest sample completeness possible across
180 samples (82.5%) and across islands (98.6%) (Chao et al., 2020). The standard error and
181 95% confidence intervals of the Hill numbers were estimated by bootstrap resampling,
182 which was repeated 1000 times. Species evenness, ${}^qE_3(p) = ({}^qD - 1)/(S - 1)$, where qD
183 denotes Hill numbers of order q , and S denotes species richness, was quantified using
184 the continuous profiles of Hill numbers as functions of order q (Chao and Ricotta, 2019).
185 A gradual profile suggests a more even community in which the species richness and
186 number of abundant and dominant species are similar. In contrast, a steep profile
187 indicates an uneven community comprised of one or a few dominant species (Mamo et
188 al., 2023).

189

190 To distinguish biofacies associated with different benthic habitats, we conducted
191 hierarchical cluster analysis based on Ward's minimum variance and three Hill number-
192 based dissimilarity indices, including Sørensen ($q=0$), Horn ($q=1$), and Morisita-Horn
193 ($q=2$), to estimate the effective proportion of un-shared species in the ostracod
194 assemblages (Chao et al., 2014a). Similarly, the order q controls the sensitivity of the
195 Hill number-based dissimilarities to species relative abundance. While the classic
196 Sørensen dissimilarity is presence-absence based, the latter two indices are designed to
197 quantify the compositional dissimilarities of abundant and dominant species,
198 respectively. The Ward's algorithm is preferred for delineating biofacies because it
199 minimizes the error sum of squares within clusters and generates more balanced clusters.
200 The number of clusters was determined by considering both the structure of the
201 dendrograms and the average silhouette width, with a higher value indicating greater
202 cohesion and separation of clusters. We also performed a non-Metric Multidimensional

203 Scaling (nMDS) to visualize and summarize faunal similarities among ostracod
204 assemblages in two-dimensional space. Stress values were calculated to quantitatively
205 weigh the ‘goodness of fit’ between the original input data matrix and the ultrametric
206 matrix of the resultant nMDS scatter plots (Hong et al., 2022; Kruskal, 1964). We used
207 a compositional heat map to illustrate the relationships between samples by Horn
208 dissimilarities and between species by Hellenger distances.

209

210 All analyses were implemented in RStudio. We used the package ‘iNEXT’ to estimate
211 diversity (Chao et al., 2014a; Hsieh et al., 2016) and ‘vegan’ for our multivariate
212 analyses (Oksanen et al., 2020). Figures and maps were constructed using ‘ggplot2’
213 (Wickham, 2020).

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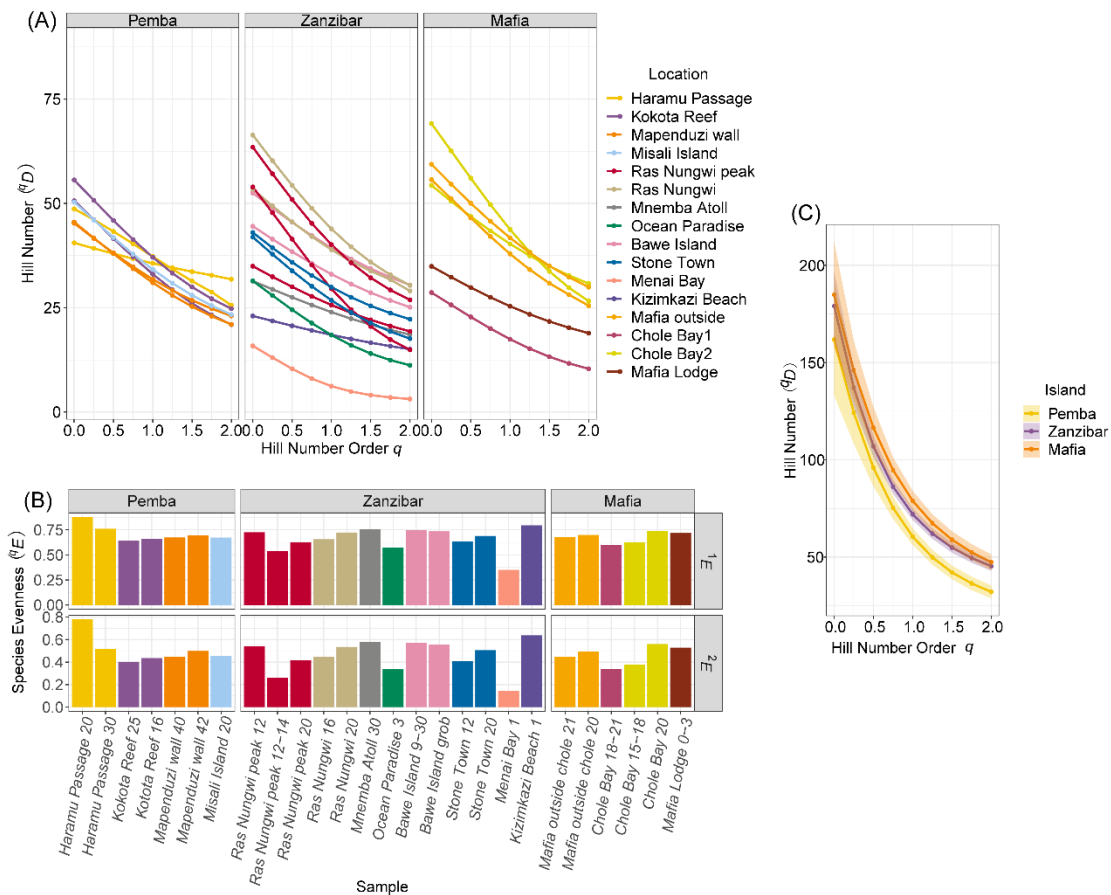
215 **4 Results**

216 4.1 Diversity

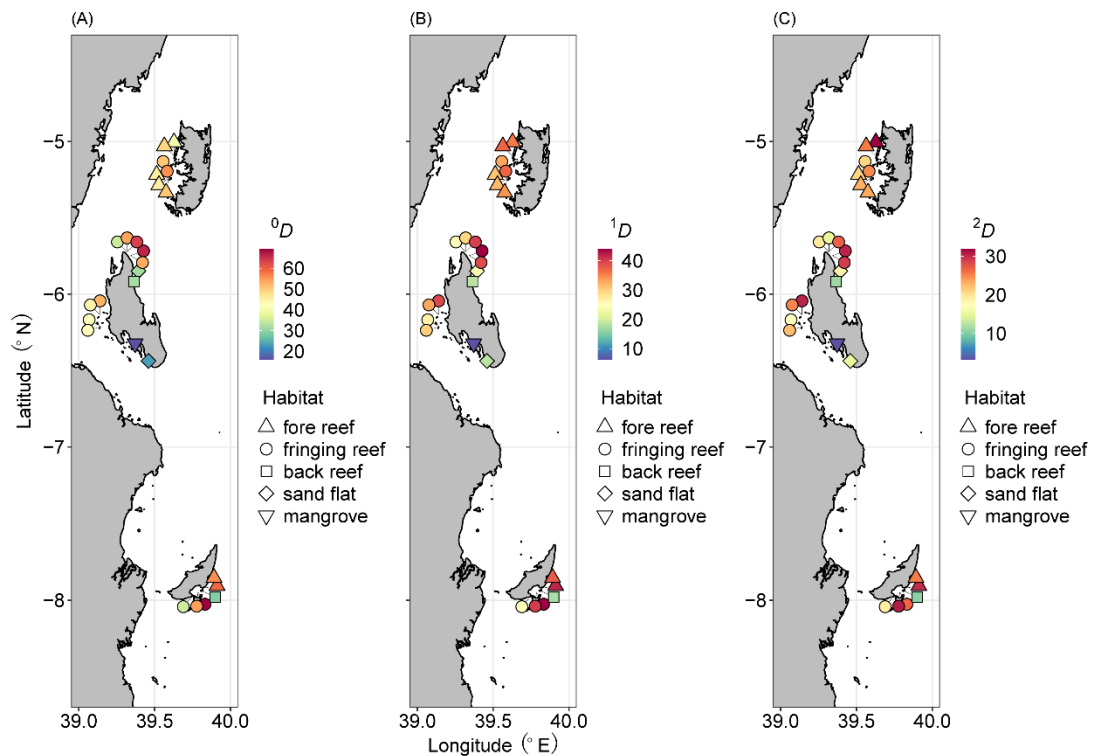
217 A total of 6262 ostracods were recovered from 26 samples at 16 locations around the
218 Zanzibar Archipelago. They represent remarkably diverse ostracod assemblages
219 comprised of 235 species under 77 genera. Exceedingly high abundance was found at
220 Stone Town, while sites at Bawe Island and Mafia outside were also abundant, in
221 contrast to the lowest abundance at Haramu Passage and Chole Bay 1 (Table 1).

222 Considering the alpha diversity of individual sample as measured by Hill number of
223 different order q , the spatial diversity patterns were relatively consistent for rare (i.e.,
224 species richness, 0D) and abundant (1D) species. The highest values were recorded for
225 fringing reefs at Chole Bay 2 and Ras Nungwi, followed by fringing reefs at Mafia
226 outside and Ras Nungwi peak (Figs. 2A, 3A-B, 2A, 3, S1-S2). Moderately high levels
227 of diversity were observed at fore reef sites in Pemba Island and fringing reefs at Bawe,
228 Stone Town, and Mafia Lodge. In terms of the diversity of dominant (2D) species, there
229 was a more homogenous distribution with similarly high values found at various
230 fringing and fore reefs, including Chole Bay 2, Mafia outside, Haramu Passage, Bawe
231 Island, Ras Nungwi and Ras Nungwi peak (Figs. 2A, 3C). All remaining localities
232 (Chole Bay 1, Mnemba Atoll, Ocean Paradise and Kizimkazi Beach) characterized by
233 sand flat and back reef habitats had consistently low diversity across all order q ,
234 especially Menai Bay that was lined with mangrove stands (Figs. 2A, 3). Evenness was
235 highest at Haramu Passage and lowest at Menai Bay for both orders $q=1$ and $q=2$ (Figs.
236 2B, S13). With respect to the gamma diversity of each island, Mafia and Zanzibar were

237 almost equally diverse across all order q , while Pemba had significantly lower diversity
 238 for abundant and dominant species (Figs. 2C, S24).
 239



240
 241 Fig. 2. Diversity results of the Zanzibar Archipelago ostracods. (A) Alpha diversity of
 242 each sample shown by Hill number profile based on 82.5% sample coverage. The
 243 overall elevation of the profile indicates the diversity based on hill number across
 244 different order q . The levelness of the line indicates species evenness of the assemblage,
 245 because a complete leveled diversity profile would suggest that the numbers of total,
 246 common and dominant species are all the same. (B) Evenness of each sample as the
 247 normalized slope of Hill number profile for order $q=1$ and $q=2$ based on 82.5% sample
 248 coverage. (C) Gamma diversity of each island shown by Hill number profile based on
 249 98.6% sample coverage. The shade area shows 95% confidence interval of the profile.
 250



251

252 Fig. 3. Diversity maps of the Zanzibar archipelago ostracod. Distributions of Hill
 253 numbers 0D (A: $q = 0$, species richness), 1D (B: $q = 1$, exponential Shannon), and 2D
 254 (C: $q = 2$, inversed Simpson). We used 82.5% sample coverage to standardize the Hill
 255 number estimates. Diversity and habitat are represented by color and shape as in the
 256 legends, respectively. Diversity maps of Zanzibar archipelago ostracod. Distributions
 257 of Hill numbers 0D (A: $q = 0$, species richness), 1D (B: $q = 1$, exponential Shannon);
 258 and 2D (C: $q = 2$, inversed Simpson). We used 82.5% sample coverage to standardize
 259 the Hill number estimates. Diversity and habitat are represented by color and shape as
 260 in the legends, respectively.

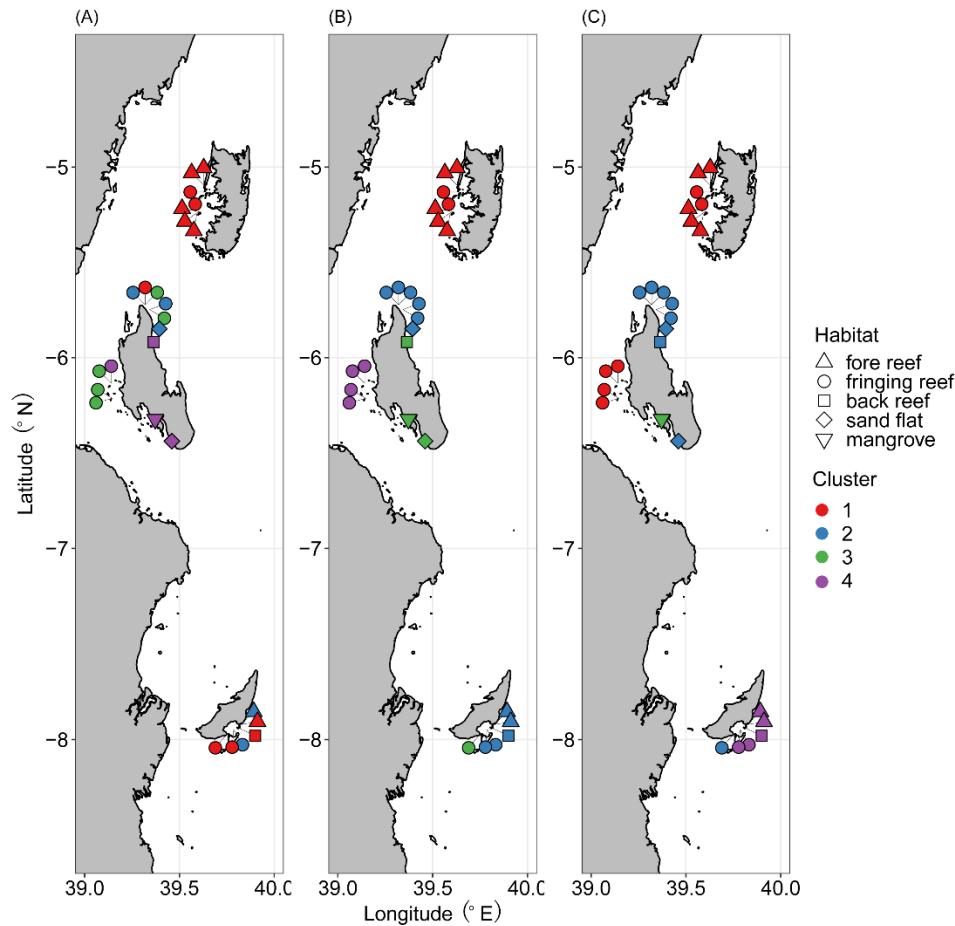
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263 4.2 Multivariate analysis

264 First, cluster analyses based on Sørensen, Horn, and Morisita-Horn dissimilarities
 265 delineated biofacies considering faunal composition in terms of species occurrence,
 266 relative abundance of abundant species, and relative abundance of dominant species,
 267 respectively. The greatest average silhouette width suggested the division of samples
 268 into ten clusters for all three dissimilarity measures; however, it is beyond interpretable
 269 to have too many clusters, given the size of our dataset. We, therefore, referred to the
 270 structure of the dendrograms based on three dissimilarity measures to determine the

271 optimum number of clusters to be four (Fig. S35). The NMDS results showed a clear
272 separation of four biofacies based on Horn and Morisita-Horn dissimilarities, but not
273 Sørensen dissimilarity, which was calculated with a relatively high stress value (0.26)
274 (Fig. S44). Ostracod faunas in Pemba Island constituted a distinct group across all levels
275 of faunal composition from presence/absence to relative abundance (Biofacies 1; Fig.
276 4-5). Ras Nungwi, Ras Nungwi peak, and nearby Menemba Atoll were congregated
277 with different sites around Zanzibar and Mafia in Biofacies 2, including Mafia outside
278 and Chole Bay 2 in Sørensen, Mafia outside, Chole Bay 1 and Chole Bay 2 in Horn,
279 Ocean Paradise, Kizimkazi Beach and Mafia Lodge in Morisita-Horn analysis (Fig. 45).
280 Samples assigned to Biofacies 3 and 4 strongly varied depending on the dissimilarity
281 matrix used, indicating these biofacies have different ecological meaning among three
282 cluster analyses (Fig. 4). Specifically, they scattered around the entire Zanzibar Island
283 based on Sørensen dissimilarity. Biofacies 4 was distributed along the western coast of
284 Zanzibar, including Stone Town and Bawe, and Biofacies 3 covered the remaining
285 Zanzibar locations (Menai Bay, Ocean Paradise and Kizimkazi Beach) in addition to
286 Mafia Lodge based on Horn dissimilarity. On the other hand, when Morisita-Horn
287 dissimilarity was applied, Menai Bay was different from all other sites as a distinctive
288 Biofacies 3 while most Mafia sites (Mafia outside, Chole Bay 1, and Chole Bay 2)
289 aggregated in Biofacies 4. Considering the performance of multivariate analyses to
290 reflect and interpret biological patterns, we think that cluster and NMDS results based
291 on Horn dissimilarity most reasonably captured the underlying ecological significance
292 of reefal versus non-reefal facies as determined by benthic community, depth, and
293 possibly anthropogenic disturbances (see the Discussion section). We therefore focus
294 on the four biofacies as divided by Horn-based analysis to scrutinize their diversity and
295 compositional structure in relation to a set of environmental variables.
296
297



298

299 Fig. 45. Distribution of ostracod Biofacies 1-4 based on (A) Sørensen, (B) Horn, and (C)
 300 Morisita-Horn dissimilarities and Ward's minimum variance cluster analysis. Note that
 301 the color schemes are independent among panels; thus, the biofacies based on different
 302 dissimilarities are unrelated. Cluster and habitat are represented by color and shape as
 303 in the legends, respectively.

304

305 Each biofacies based on Horn dissimilarity index was demonstrated with the top 10
 306 species of highest mean relative abundance as shown in Table 2 and Figure 5s-6-8.
 307 Noticeably, the Pemba fauna in Biofacies 1 was dominated by genus *Neonesidea* (*N.*
 308 *cf. crepidula* and *N. schulzi*) and *Bosasella* (*B. profunda* and *B. elongate*), together with
 309 *Paracytheridea tschoppi* (Fig. 69; Table 2). Biofacies 2 included the most diverse sites
 310 in Zanzibar and Mafia, which all shared similar faunal structures with a high abundance
 311 of *Loxocorniculum* sp. 2, *Xestoleberis rotunda*, *Paracytheridea albatros* and
 312 *Loxoconcha* sp. 3. Biofacies 3 composed of low-diversity sites in Zanzibar and Mafia
 313 was distinguished by highly abundant *Perissocytheridea estuaria-sp.1*, *Xestoleberis*
 314 *hanaii*, as well as three *Loxoconcha* species (*L.* sp. 3, *L. ghardaqensis* and *L.*

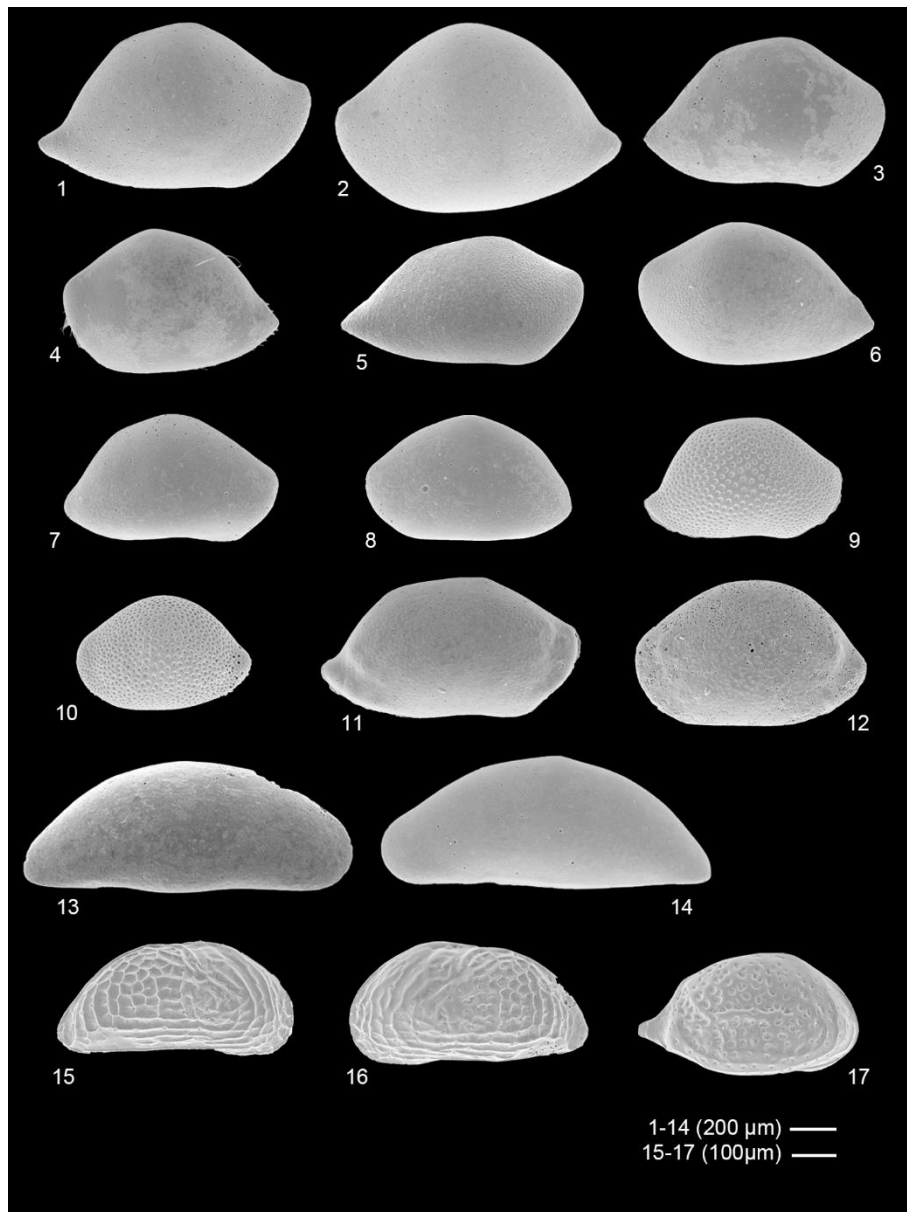
315 *lilljeborgii*). Finally, the faunal structure of Biofacies 4 in western Zanzibar showed
 316 some similarities to that of Biofacies 1 in Pemba with many common species, however,
 317 they clearly differed by the dominance of *Xestoleberis hanaii* and *Patrizia*
 318 *nucleuspersici* in Biofacies 4.

319

320 Table 2. List of top 10 species of highest % mean relative abundance for Biofacies 1-4
 321 based on Horn dissimilarity.

Species	Biofacies1	Biofacies2	Biofacies3	Biofacies4
<i>Neonesidea cf. crepidula</i>	0.085857	NA	NA	NA
<i>Bosasella profunda</i>	0.079436	NA	NA	0.040846
<i>Neonesidea schulzi</i>	0.075285	0.032551	0.024322	0.041291
<i>Paracytheridea tschoppi</i>	0.035779	NA	NA	0.028826
<i>Loxocorniculum sp. 2</i>	0.030562	0.063399	NA	NA
<i>Xestoleberis hanaii</i>	0.028593	0.039954	0.084378	0.071834
<i>Patrizia nucleuspersici</i>	0.02842	NA	NA	0.057965
<i>Paranesidea cf. spongicola</i>	0.026203	NA	NA	0.029754
<i>Xestoleberis sp. 1</i>	0.023801	NA	NA	NA
<i>Bosasella elongata</i>	0.023369	NA	0.017579	NA
<i>Xestoleberis rotunda</i>	NA	0.061861	NA	NA
<i>Paracytheridea albatros</i>	NA	0.045056	0.037464	NA
<i>Loxoconcha sp. 3</i>	NA	0.041327	0.110386	NA
<i>Bosasella sp. 1</i>	NA	0.040122	NA	NA
<i>Macrocyprina maddocksae</i>	NA	0.039264	NA	NA
<i>Caudites exmouthensis</i>	NA	0.027832	NA	NA
<i>Paranesidea sp. 1</i>	NA	0.025497	NA	NA
<i>Perissocytheridea estuaria-sp.1</i>	NA	NA	0.157932	NA
<i>Loxoconcha ghardagensis</i>	NA	NA	0.073153	NA
<i>Hiltermannicythere rubrimaris</i>	NA	NA	0.04805	NA
<i>Loxoconcha lilljeborgii</i>	NA	NA	0.033061	NA
<i>Neohornibrookella lactea</i>	NA	NA	0.018616	NA
<i>Neonesidea sp. 3</i>	NA	NA	NA	0.048331
<i>Neonesidea paiki</i>	NA	NA	NA	0.042016
<i>Loxoconcha cf. gisellae</i>	NA	NA	NA	0.035319
<i>Perissocytheridea? sp. 2</i>	NA	NA	NA	0.029391

322

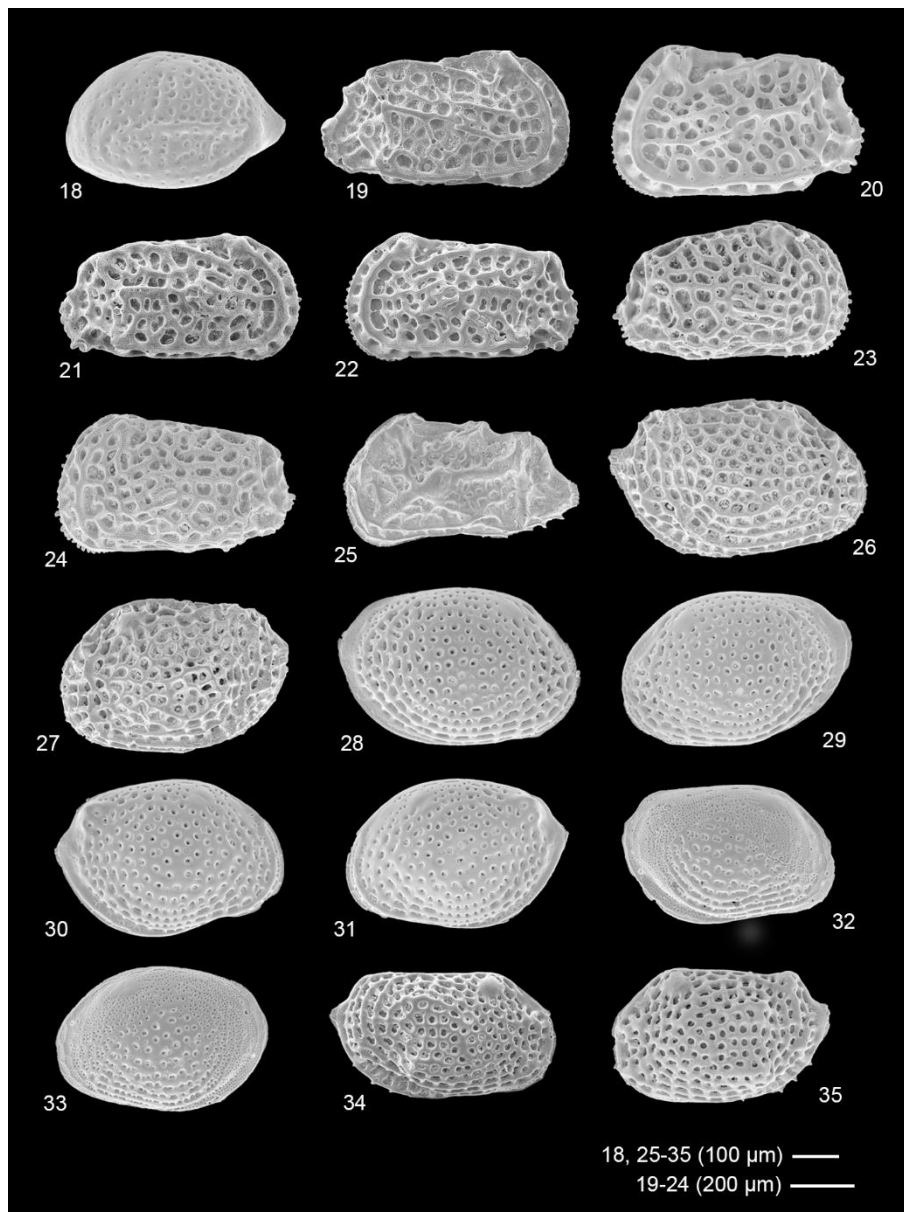


323

324 Fig. 56. Scanning electron microscopy images of the top 10 ostracod species of
 325 highest % mean relative abundance for Biofacies 1-4 based on Horn dissimilarity. 1,
 326 *Neonesidea* cf. *crepidula*, RV, [Kokota reef25](#); 2, *Neonesidea* cf. *crepidula*, LV, [Haramu](#)
 327 [Passage30](#); 3, *Neonesidea paiki*, RV, [Bawe island9-30](#); 4, *Neonesidea paiki*, LV,
 328 [Kokota reef25](#); 5, *Neonesidea schulzi*, RV, [Mapenduzi wall42](#); 6, *Neonesidea schulzi*,
 329 LV, [Kokota reef25](#); 7, *Neonesidea* sp. 3, RV, [Bawe island9-30](#); 8, *Neonesidea* sp. 3,
 330 LV, [Bawe island9-30](#); 9, *Paranesidea* cf. *spongicola*, RV, [Bawe island9-30](#); 10,
 331 *Paranesidea* cf. *spongicola*, LV, [Kokota reef25](#); 11, *Paranesidea* sp. 1, RV, [Chole](#)
 332 [bay18-21](#); 12, *Paranesidea* sp. 1, LV, [Chole bay18-21](#); 13, *Macrocyprina maddocksae*,
 333 RV, [Haramu Passage20](#); 14, *Macrocyprina maddocksae*, LV, [Kotota reef16](#); 15,
 334 *Perissocytheridea estuariasp.1*, RV, [Menai bay1](#); 16, *Perissocytheridea estuariasp.1*,

335 LV, Menai bay1; 17, *Perissocytheridea?* sp. 2, RV, Bawe island grob. All adults and
 336 lateral views.

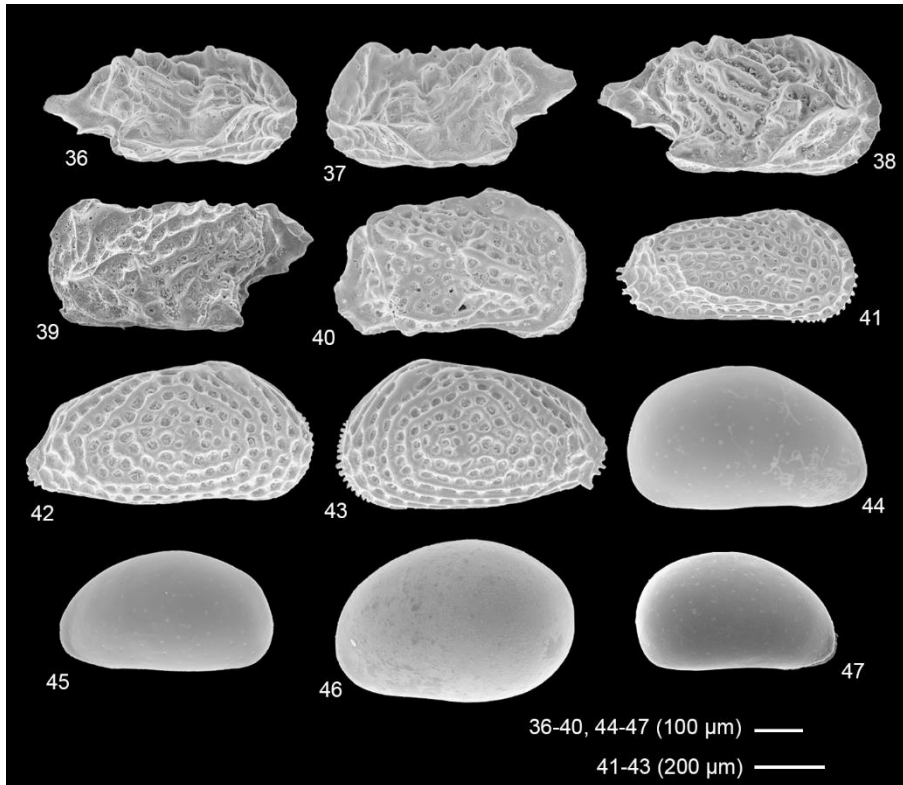
337



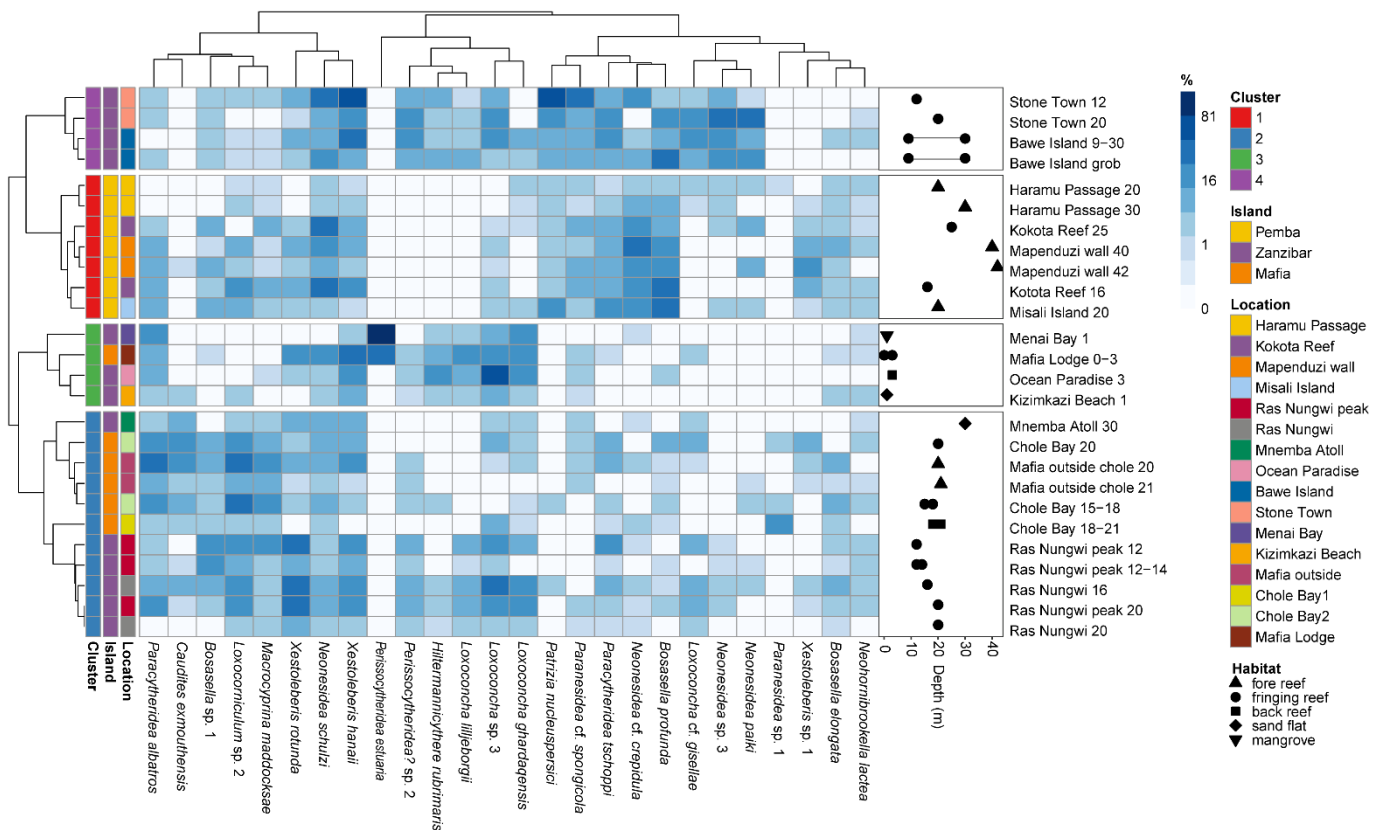
338

339 Fig. 57. Continued. Scanning electron microscopy images of the top 10 ostracod
 340 species of highest % mean relative abundance for Biofacies 1-4 based on Horn
 341 dissimilarity. 18, *Perissocytheridea?* sp. 2, LV, Bawe island9-30; 192, *Bosasella*
 342 *elongate*, RV, Haramu Passage30; 203, *Bosasella elongate*, LV, Mapenduzi wall42;
 343 214, *Bosasella profunda*, RV, Haramu Passage20; 225, *Bosasella profunda*, LV,
 344 Mapenduzi wall42; 236, *Bosasella* sp. 1, RV, Kokota reef25; 247, *Bosasella* sp. 1, LV,
 345 Kotota reef16; 258, *Caudites exmouthensis*, LV, RusNungwi16; 269, *Loxoconcha*
 346 *ghardaqensis*, RV, Mnemba atoll30; 2710, *Loxoconcha ghardaqensis*, LV, Ras

347 [Nungwi peak12](#); [2811](#), *Loxoconcha* cf. *gisellae*, RV, [Bawe island9-30](#); [2912](#),
 348 *Loxoconcha* cf. *gisellae*, LV, [Bawe island9-30](#); [3013](#), *Loxoconcha lilljeborgii*, RV,
 349 [Bawe island9-30](#); [3114](#), *Loxoconcha lilljeborgii*, LV, [Bawe island grob](#); [3215](#),
 350 *Loxoconcha* sp. 3, RV, [Stone town20](#); [3316](#), *Loxoconcha* sp. 3, LV, [Stone town20](#); [3417](#),
 351 *Loxocorniculum* sp. 2, RV, [Haramu Passage30](#); [3518](#), *Loxocorniculum* sp. 2, LV,
 352 [Kotota reef16](#). All adults and lateral views.
 353



354
 355 Fig. 58. Continued. Scanning electron microscopy images of the top 10 ostracod
 356 species of highest % mean relative abundance for Biofacies 1-4 based on Horn
 357 dissimilarity. [361](#), *Paracytheridea albatross*, RV, [Kokota reef25](#); [372](#), *Paracytheridea*
 358 *albatross*, LV, [Kotota reef16](#); [38](#), *Paracytheridea tschoppi*, RV, [Kokota reef25](#); [394](#),
 359 *Paracytheridea tschoppi*, LV, [Mapenduzi wall42](#); [405](#), *Neohornibrookella lactea*, RV,
 360 [Misali island20](#); [416](#), *Hiltermannicythere rubrimaris*, RV, [Stone town20](#); [427](#), *Patrizia*
 361 *nucleuspersici*, RV, [Stone town20](#); [438](#), *Patrizia nucleuspersici*, LV, [Stone town12](#);
 362 [449](#), *Xestoleberis hanaii*, RV, [Bawe island9-30](#); [4510](#), *Xestoleberis hanaii*, LV, [Kokota](#)
 363 [reef25](#); [4611](#), *Xestoleberis rotunda*, LV, [Ras Nungwi peak12](#); [4712](#), *Xestoleberis* sp. 1,
 364 RV, [Mapenduzi wall42](#). All adults and lateral views.
 365



366

367 Fig. 69. Dendrograms based on Horn dissimilarity between samples and Hellenger
 368 distances between top 10 species of highest mean relative abundance in each cluster.
 369 The blue heatmap indicates the relative (%) abundance of each species in each sample.
 370 The side panel shows water depth and habitat type of each sample (note that several
 371 samples are shown by their corresponding depth ranges).

372

373 5 Discussion

374 Through Hill number profile and multivariate analyses, we quantified a highly diverse
 375 ostracod fauna in the Zanzibar Archipelago composed of four distinct biofacies. The
 376 delineation of biofacies varied considerably depending on the dissimilarity matrix used,
 377 indicating inconsistent faunal structures across different levels of species information
 378 from occurrence to relative abundance (Fig. 45). In terms of the presence/absence of
 379 species (Sørensen dissimilarity), all Pemba sites united in Biofacies 1 but the
 380 assignment of Zanzibar and Mafia sites into Biofacies 1-4 seemingly conformed to a
 381 noisy pattern (Fig. 4A). Accordingly, four biofacies intersected with each other in
 382 nMDS space with relatively high stress value (Fig. S4A). A possible explanation is that
 383 the occurrence of individual species may be homogenous among sites in similar
 384 environmental conditions within a finite geographic region. Many species are likely to

385 be ubiquitous across the entire neritic zone despite showing certain ecological
386 preferences, and the redeposition processes may further facilitate the mixing of death
387 assemblages to blur the spatial signal at a local scale (Frenzel and Boomer, 2005).
388 Consequently, species presence in all available habitats may translate to considerable
389 faunal similarities among biofacies as measured by Sørensen index. When considering
390 the composition of abundant species (Horn dissimilarity) (Figs. 4B ~~and 5B~~), the
391 identification of four biofacies instead reflected significant changes in ostracod
392 assemblages along two important environmental gradients, which are benthic
393 community type and water depth. Specifically, Biofacies 1 and 2 characterize typical
394 fore reefs in deep subtidal (sampling depth 16-42 m) and fringing reefs in shallow
395 subtidal (12-30 m), respectively (Fig. 69). Biofacies 3 indicates intertidal habitats with
396 plant cover (0-30 m), and finally Biofacies 4 features degraded fringing reefs in shallow
397 subtidal (9-30 m) (see discussion below).

398

399 We summarized the ecological preferences of dominant genera in each biofacies based
400 on Horn dissimilarity (Table 3 and Fig. 69) and investigated how key environmental
401 factors (benthic community type, water depth, and anthropogenic disturbance) may
402 control the distribution and diversity of reefal ostracod assemblages. First of all,
403 *Neonesidea* and *Paranesidea* (family Bairdiidae) are typical reefal genera that reach
404 their maximum diversity and incidence on reefs and reef-associated habitats in tropical
405 shallow-marine environments (Whatley and Watson, 1988; Maddocks, 2013; Titterton
406 and Whatley, 1988). Their dominance in Biofacies 1 is consistent with our background
407 understanding that the Pemba reefs were pristine and healthy (Ussi et al., 2019;
408 Grimsditch et al., 2009). However, it should be noted that individual species of these
409 genera likely have different environmental tolerance. For example, *N. cf. crepidula*
410 were restricted to Biofacies 1 while *N. schulzi* were widespread among four biofacies
411 inhabiting both reef and algae habitats (Fig. 69) (Mostafawi et al., 2005). *Bosasella* as
412 another prominent component of Biofacies 1 is also known to occur on coral reefs in
413 the western Indian Ocean (Munef et al., 2012; Jellinek, 1993). *Paracytheridea* and
414 *Caudites* on the other hand are loosely categorized as reefal genera, as their dominance
415 on coral reefs was reported but not studied in detail (Whatley and Watson, 1988; Keyser
416 and Mohammed, 2021). In this study, they were common on fore- and fringing-reefs in
417 Biofacies 1 and 2 (Fig. 69). *Loxoconcha* and *Loxocorniculum* (family Loxoconchidae)
418 as two phylogenetically related and ecologically similar genera exhibited ubiquitous

419 distribution around the Zanzibar Archipelago with highest relative abundance in
420 Biofacies 3 followed by Biofacies 2. As generalists, they thrive on a wide variety of
421 benthic habitats across the neritic zone and show affinities to plant substrates (algae and
422 seagrass beds) in particular (Munef et al., 2012; Keyser and Mohammed, 2021; Kamiya,
423 1988). The ecology of *Xestoleberis* is very similar to that of Loxoconchidae, living both
424 on coral reefs and algal flats (Keyser and Mohammed, 2021; Munef et al., 2012;
425 Whatley and Watson, 1988; Kamiya, 1988). This genus was almost equally weighted
426 in all biofacies, although individual species clearly preferred different environments, as
427 *X. hanaii* prevailed in Biofacies 3 and 4 while *X. rotunda* only in Biofacies 2 (Fig. 69).
428 *Patrizia* is documented as a reefal genus in lower littoral zone along the eastern coast
429 of tropical Africa (Jellinek, 1993). It dominated the relatively deep fringing-reef faunas
430 of Biofacies 4, which were subject to sewage-derived nutrient and trace metal pollution
431 from the-Zanzibar Town (Narayan et al., 2022; Bravo et al., 2021). Different from all
432 the above-discussed genera, *Hiltermannicythere* and *Perissocytheridea* are restricted to
433 shallow intertidal environments as phytal and sediment-dwelling taxa, respectively
434 (Jellinek, 1993), which explains their abundance in our Biofacies 3. *Perissocytheridea*
435 is especially considered a bioindicator of brackish water facies, adapting to euryhaline
436 conditions (Nogueira and Ramos, 2016; Keyser, 1977). Furthermore, we revealed a
437 more generalized pattern of the compositional differences among biofacies with the top
438 5 families of highest mean relative abundance in each biofacies (Fig. 740).

439

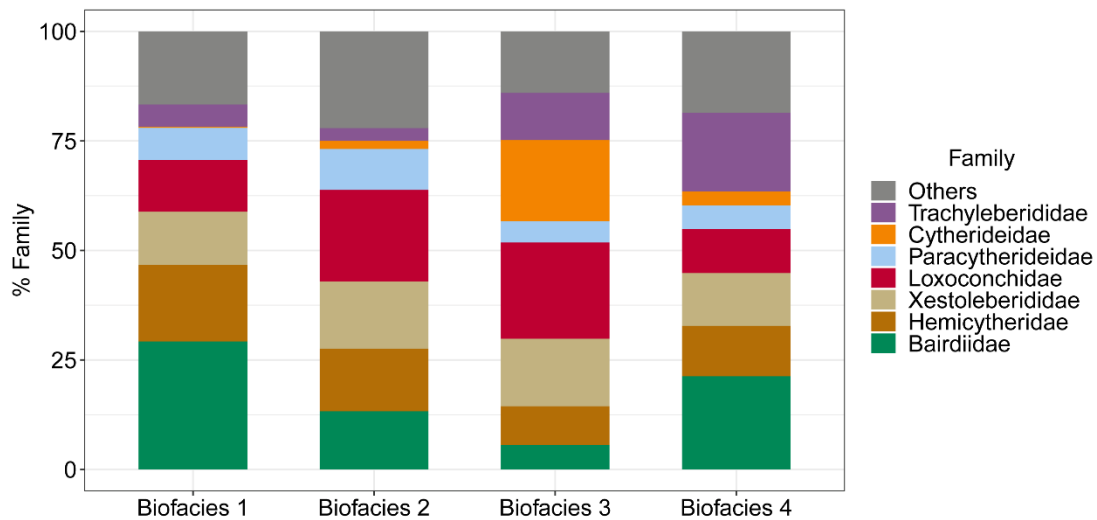
440 Table 3. Autoecology summary of important ostracod genera.

Genus	Predominant habitats	References
<i>Neonesidea</i>	Coral reef	Whatley and Watson (1988); Maddocks (2013); Titterton and Whatley (1988); Maddocks (1969)
<i>Paranesidea</i>	Coral reef	Titterton and Whatley (1988); Whatley and Watson (1988); Maddocks (1969)
<i>Bosasella</i>	Coral reef	Munef et al. (2012)
<i>Loxoconcha</i>	Algal mat and reef	Keyser and Mohammed (2021); Whatley and Watson (1988); Munef et al. (2012); Kamiya (1988)
<i>Loxocorniculum</i>	Algal mat and reef	Munef et al. (2012); Kamiya (1988)
<i>Xestoleberis</i>	Algal mat and reef	Keyser and Mohammed (2021); Whatley and Watson (1988); Munef et al. (2012); Kamiya (1988)
<i>Patrizia</i>	Coral reef	Jellinek (1993)

<i>Hiltermannicythere</i>	Intertidal algal mat	Jellinek (1993); Keyser and Mohammed (2021)
<i>Paracytheridea</i>	Coral reef	Whatley and Watson (1988)
<i>Caudites</i>	Coral reef	Whatley and Watson (1988); Keyser and Mohammed (2021)
<i>Perissocytheridea</i>	Intertidal sand flat, <u>eur haline</u> brackish <u>water</u>	Nogueira and Ramos (2016); Keyser (1977)

441

442



443

444 Fig. 710. Family composition of Biofacies 1-4 based on Horn dissimilarity. The top 5
 445 families of the highest % relative abundance in each biofacies are shown.

446

447 Thus, our study indicates that the distribution of shallow-marine ostracods in the the
 448 Zanzibar Archipelago is characterized by three reefal facies and one intertidal facies.
 449 Yet slight differences in bathymetry, benthic community type, and anthropogenic
 450 impacts likely contributed to subtle faunal changes among the reefal Biofacies 1, 2, and
 451 4. The fore reefs in Pemba (Biofacies 1) were deepest with high incidence and diversity
 452 of live hard corals (Gavrilets and Losos, 2009; Ussi et al., 2019), which accounted for
 453 the definite dominance of ostracod reefal taxa (Bairdiidae and *Bosasella*) over algal
 454 taxa (Loxoconchidae and Xestoleberididae) (Figs. 69-710). Moderately high levels of
 455 diversity in terms of rare, abundant, and dominant species were observed for these
 456 ostracod assemblages (Fig. s. 2-3). The Pemba reefs are thereby considered the most
 457 mature and authentic reef ecosystem, serving as a natural reference for comparing with
 458 other sites. The fringing-reef fauna of western Zanzibar (Stone Town and Bawe,
 459 Biofacies 4) exhibited certain similarities with the Pemba fauna as indicated by the

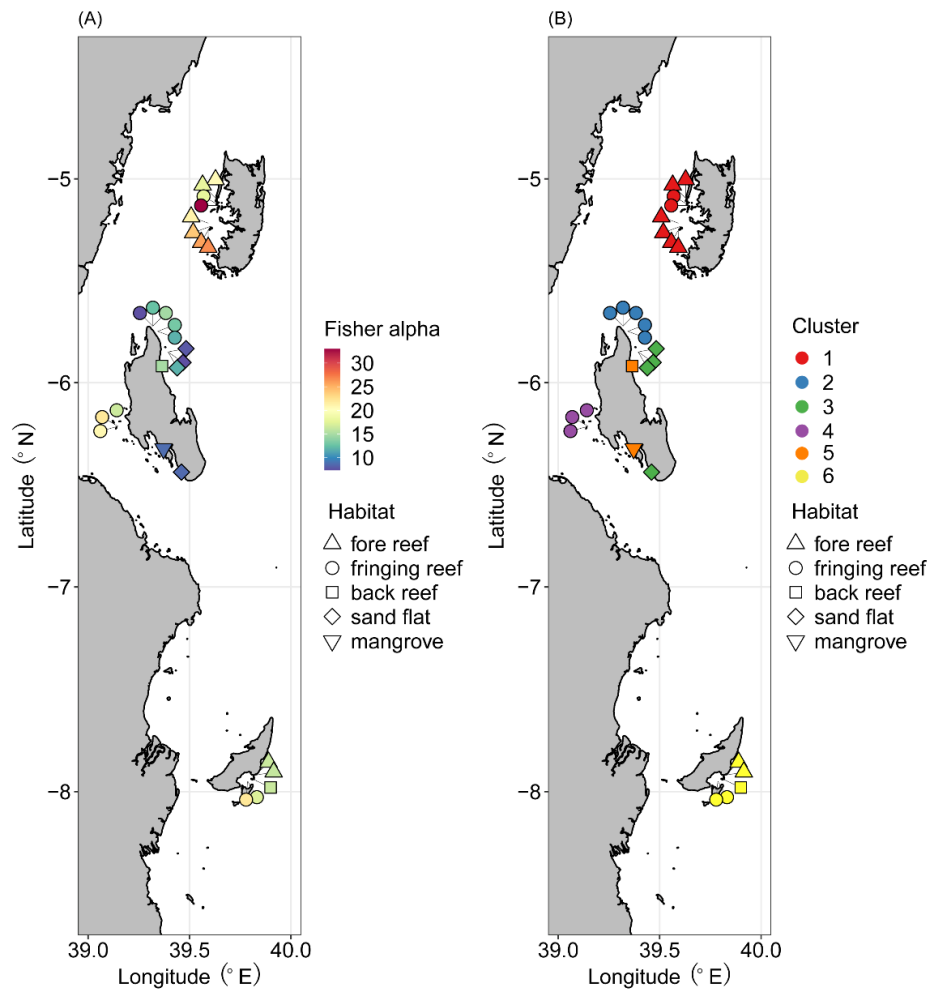
460 prevalence of *Bosasella profunda*, *Paracytheridea tschoppi*, and *Paranesidea* cf.
461 *spongicola* in both facies (Fig. 69). Indeed, they were grouped together based on the
462 composition of dominant species (Morisita-Horn analysis) (Fig. 45C). Faunal
463 similarities between Pemba and Stone Town make sense as they are in comparable
464 baseline conditions of water depths and hydrology along the protected western coast of
465 the Zanzibar Archipelago, in contrast to Ras Nungwi and Chole Bay that are exposed
466 to oceanic disturbances from the east (Fig. 1). However, Biofacies 4 was differentiated
467 from Biofacies 1 by the dominance of *Patrizia* in conjunction with the absence of
468 *Neonesidea* cf. *crepidula*. It also had the highest relative abundance of
469 Trachyleberididae genera among all facies (Fig. 710), for example, *Adencythere*,
470 *Strobilocythere*, *Bradyon*, and *Actinocythereis*, but their ecologies are not well
471 understood. Stressful environmental conditions in terms of overexploitation, tourism
472 and coastal pollution offer the most possible explanation for such a unique faunal
473 composition and comparatively low diversity of Biofacies 4 (Fig. s-3-2-3) (Bravo et al.,
474 2021; Larsen et al., 2023). Consistently, foraminifera and coral surveys indicated early
475 stages of reef degradation there (Narayan et al., 2022; Bravo et al., 2021; Thissen and
476 Langer, 2017) It is possible that ongoing anthropogenic disturbances near the Stone
477 Town will eventually exceed the critical threshold levels to cause more pronounced
478 changes in ostracod faunal structures in terms of dominant species through a shift in
479 benthic habitat (Narayan et al., 2022; Hong et al., 2022). Other than Biofacies 1 and 4,
480 Biofacies 2 mostly represented a different type of relatively shallow (12-20 m) fringing-
481 back reef reefal-habitats of Ras Nungwi, Chole Bay, and Mafia outside, in addition to
482 a deeper (30 m) sand flat of Mnemba Atoll, and Mnemba Atoll (Fig. 65B). Algal taxa
483 (Loxoconchidae and Xestoleberididae) and reefal taxa (Bairdiidae, *Bosasella*,
484 *Paracytheridea*, and *Caudites*) reached equally high levels of relative abundance there
485 (Figs. 69-710). ~~Most sites in Biofacies 2 were relatively shallow (12-21 m) except for~~
486 ~~Mnemba Atoll (30 m), and they covered the transitional zone from intertidal-sandy~~
487 ~~bottom to subtidal true reefs.~~ Microhabitats on the reef platforms of Biofacies 2 are
488 believed to be diverse and heterogenous with interlaced live and dead corals, algae and
489 seagrass, calcareous sands, as well as bare substrate rock (Ussi et al., 2019; Larsen et
490 al., 2023), which facilitated the coexistence of reefal and algal ostracods and
491 consequently the highest diversity of local assemblages (Fig. s-3-2-3). The remaining
492 Biofacies 3 corresponded to the shallowest intertidal habitats with various benthic
493 communities, including a marginal back reef, a marginal fringing reef, a sand flat, and

494 a mangrove (Fig. ~~65B~~). Typical reefal taxa (Bairdiidae and *Bosasella*) dropped to their
495 lowest relative abundance in this facies, replaced by large numbers of Loxoconchidae,
496 *Perissocytheridea*, and *Hiltermannicythere* that well adapted to shallow euryhaline
497 conditions (Figs. ~~69-710~~). Not surprisingly, the diversity of Biofacies 3 was much lower
498 than that of open-ocean reefal facies, as drastic changes in temperature, salinity,
499 dissolved oxygen, and wave energy in the intertidal zone may be too challenging for
500 many marine taxa (Fig. ~~s-32-3~~) (Morley and Hayward, 2007; Frenzel and Boomer,
501 2005). The mangrove habitat at Menai Bay was unique concerning the absolute
502 dominance of *Perissocytheridea* in line with its lowest diversity and evenness (Figs.
503 ~~2B, 3-2-3~~). It indeed constituted an independent biofacies based on Morisita-Horn
504 analysis (Fig. ~~45C~~).

505

506 The division scheme of four biofacies based on Horn dissimilarity explicitly revealed
507 spatial patterns of ostracod distribution in aspect of diversity and composition, as
508 discussed above. Our results are generally concordant with a previous study on benthic
509 foraminifera, which separated six clusters of Pemba, Stone Town, Mafia Bay, Ras
510 Nungwi, Mnemba Atoll, and Menai Bay, respectively (Fig. ~~811B~~) (Thissen and Langer,
511 2017). Each of these foraminifera clusters corresponded to major habitat types, as
512 argued by the authors (Thissen and Langer, 2017), and we accordingly pointed out the
513 consistent role of habitat factors in shaping the biogeography of both ostracod and
514 foraminifera biotas. However, the diversity patterns of these two groups were
515 apparently different among reefal habitats (Figs. 3, ~~811A~~). High, moderate, and low
516 levels of diversity were recorded on fore reefs (Pemba), fringing reefs (Mafia and
517 Zanzibar), and intertidal (Zanzibar) for foraminifera, in contrast to fringing reefs (Mafia
518 and Zanzibar), fore reefs (Pemba), and intertidal (Zanzibar) for ostracods, respectively.
519 Such discrepancies may imply a tight association of foraminifera with reef ecosystem
520 and their ultra-sensitivity to reef health, since their diversity generally decreased from
521 pristine, mature reefs to degraded, marginal reefs. Ostracods, on the other hand, may
522 be less confined or specific to reef habitats. The occupation of coral and algae substrate
523 by distinct faunal groups allows them to thrive in the transitional zone between marginal
524 and true reefs. Another important factor accounting for the different distributional
525 patterns between ostracods and foraminifera is likely their tolerance to eutrophication
526 and pollution. Previous studies indicate that an intermediate level of eutrophication is
527 beneficial to ostracods and many other soft sediment benthos, which are also resistant

528 to heavy metal pollution in highly urbanized areas (Hong et al., 2022). Consistently,
 529 our sampling sites at Stone Town reported the highest abundance of ostracods (Table
 530 1). Foraminifera on the other hand are susceptible to environmental stressors, as shown
 531 by low taxonomic richness and high dominance of the faunas in eutrophic conditions
 532 (Mamo et al., 2023). In our case, it makes sense that the highest diversity of
 533 foraminifera was found in pristine and oligotrophic Pemba waters.



534
 535 Fig. 811. Distributions of benthic foraminifera (A) diversity measured as Fisher alpha
 536 index; (B) cluster groups based on Q-mode cluster analysis. Modified from Thissen and
 537 Langer (2017). Diversity/cluster and habitat are represented by color and shape as in
 538 the legends, respectively.

539
 540 Most importantly, this study established a clear benthic community axis along which
 541 the composition and diversity of ostracod assemblage vary, i.e., from coral reefs to
 542 algae turfs. We identified typical reefal association (*Bairdiidae-Bosasella*) versus algal
 543 association (*Loxoconchidae-Xestoleberididae*) (Fig. 710), and their relative dominance

544 may be used as a direct indication of benthic community type. As there is a growing
545 interest to monitor the degradation of reef ecosystems from the coral-dominated phase
546 to the algae-dominated phase (Roth et al., 2018; Knowlton and Jackson, 2008;
547 Knowlton, 2012), our finding is of potential conservation value. Ostracod species
548 diversity was higher on shallow fringing reefs than on deep fore reefs, as the former
549 ecosystem harbored evenly weighted reefal and algal taxa within a dynamic mosaic of
550 microhabitats. Our results thus strongly indicate the importance of coral reefs in
551 harboring conspicuously high levels of meiobenthic biodiversity, likely through finer
552 niche partitioning (Kohn et al., 1997; Fox and Bellwood, 2013). Along with the benthic
553 community factor, we quantified prominent changes in faunal structure and diversity
554 along a depth gradient, as the intertidal euryhaline assemblages transited to subtidal
555 fully marine assemblages. It is widely recognized that shallow-marine biotas are
556 especially susceptible to depth associated changes, such as temperature, salinity, wave
557 action, and light penetration (Carvalho et al., 2012; Tian et al., 2022). This study
558 showed that a narrow depth zone across the intertidal and subtidal (~40 m) was further
559 divided and occupied by distinct biofacies. Such a finely tuned vertical gradient of
560 diversity and faunal composition added to an exceedingly large regional species pool
561 (235 species) in this tropical shallow-marine setting. Last but not least, it should be
562 aware that the effects of depth and benthic community type are often intertwined with
563 each other in determining ostracod assemblages, as the habitat-building corals and algae
564 essentially exhibit depth distributions. At a regional scale like the Zanzibar Archipelago,
565 the combined effects of water depth and benthic community characteristics should be
566 considered in studying the spatial patterns of benthic organisms.

567

568 **6 Conclusion**

569 In conclusion, this study showed that the diversity and faunal composition of reefal
570 ostracod assemblages vary along benthic community and bathymetric gradients, which
571 may also be altered by local anthropogenic disturbances. Ostracod faunas on shallow
572 fringing reefs were especially diverse, which may be explained by high levels of habitat
573 complexity and heterogeneity. The relative dominance of reefal taxa (Bairdiidae)
574 versus algal taxa (Loxoconchidae-Xestoleberididae) is likely determined by the
575 proportion of coral versus algae cover on the reef platforms, though more extensive
576 studies beyond this region are needed to confirm the universality of this pattern. Coral
577 reefs worldwide are vulnerable to ongoing climate changes and other human impacts at

578 local to global scales, and many reefal species are at risk of extinction. It is of great
579 importance that we inspect and understand the immense biodiversity of meiobenthos
580 on coral reefs as an indispensable part of the ecosystem.

581

582 **Data availability**

583 Ostracod census data is available in the online supplementary materials. ~~will be~~
584 ~~deposited into The Paleobiology Database. DOI will be added later.~~

585

586 **Author contributions**

587 Each named author has participated sufficiently in the work to take public responsibility
588 for the content. SYT and ML developed the concept. ML collected the samples. SYT
589 performed the research ~~carried out the experiments~~ and collected the data. SYT and
590 CLW analysed the data ~~performed the data analyses~~. SYT drafted the manuscript. ML,
591 MY, and CLW reviewed and edited the manuscript.

592

593 **Competing interests**

594 The authors declare that they have no competing interests.

595

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604

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