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Global assessment of climatic responses to the ozone-vegetation interactions

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11 **Global assessment of climatic responses to the ozone-vegetation**
12 **interactions**

13

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22 **Abstract.** The coupling between surface ozone (O₃) and vegetation significantly
23 influences the regional to global climate. O₃ uptake by plant stomata inhibits the
24 photosynthetic rate and stomatal conductance, impacting evapotranspiration through
25 land surface ecosystems. Using a climate-vegetation-chemistry coupled model (the
26 NASA GISS ModelE2 coupled with the Yale Interactive terrestrial Biosphere, or
27 ModelE2-YIBs), we assess the global climatic responses to O₃-vegetation interactions
28 during the boreal summer of the present day (2005-2014). High O₃ pollution reduces
29 stomatal conductance, resulting in warmer and drier conditions worldwide. The most
30 significant responses are found in the eastern U.S. and eastern China, where the surface
31 air temperature increases by +0.33±0.87 °C and +0.56±0.38 °C, respectively. These
32 temperature ~~rises~~increases are accompanied by decreased latent heat and increased
33 sensible heat in both regions. The O₃-vegetation interaction also affects atmospheric
34 pollutants. ~~Surface~~The surface maximum daily 8-hour average O₃ concentrations
35 increase by +1.46±3.02 ppbv in eastern China and +1.15±1.77 ppbv in the eastern U.S.
36 due to the O₃-induced inhibition of stomatal uptake. With reduced atmospheric stability
37 following ~~the~~a warmer climate, increased ~~cloudiness~~cloud cover but decreased relative
38 humidity jointly reduce aerosol optical depth by -0.06±0.01 (-14.67±12.15%) over
39 eastern China. This study suggests that vegetation feedback should be considered for a
40 more accurate assessment of climatic perturbations caused by tropospheric O₃.

41 1 Introduction

42 Tropospheric ozone (O₃), one of the most detrimental air pollutants (Myhre et al.,
43 2013), not only ~~poses threats to~~threatens human health (Norval et al., 2011; Nuvolone
44 et al., 2018) but also induces phytotoxic effects ~~to~~on vegetation (Mills et al., 2007;
45 Pleijel et al., 2007). When plants are exposed to certain levels of O₃, plant
46 photosynthesis and stomatal conductance ~~is~~are inhibited ~~due to~~because of the O₃
47 oxidation of ~~cellular, enzyme~~cells, enzymes, and chlorophyll (Dizengremel, 2001;
48 Fiscus et al., 2005; Jolivet et al., 2016). Consequently, ~~the~~ carbon assimilation ~~of~~in
49 terrestrial ecosystems is limited (Yue and Unger, 2014; Oliver et al., 2018), and the
50 land–air exchange rates of water and heat fluxes are altered (Lombardozzi et al., 2015).

51 Experimental studies have shown that ~~the~~ excessive O₃ exposure ~~reduced~~reduces
52 both plant photosynthesis and stomatal conductance (Ainsworth et al., 2012;
53 Lombardozzi et al., 2013). The reduction rates are dependent on the O₃ stomatal fluxes
54 as well as the ~~damaging~~damage sensitivities ~~that,~~ which vary among different
55 vegetation types (Nussbaum and Fuhrer, 2000; Karlsson et al., 2004; Pleijel et al., 2004).
56 Several exposure-based ~~indexes~~indices, such as accumulated hourly O₃ concentrations
57 over a threshold of 40 ppb (AOT40) and the sum of all hourly average concentrations
58 (SUM00), are used to assess O₃-induced vegetation damage (Fuhrer et al., 1997;
59 Paoletti et al., 2007). In addition, the flux-related POD_y method (phytotoxic O₃ dose
60 above a threshold flux of y) is also widely applied to consider the dynamic adjustment
61 of stomatal conductance (Buker et al., 2015; Sicard et al., 2016). ~~Taking into~~
62 ~~account~~Considering the variability of plant sensitivities, different O₃ damage schemes
63 ~~were~~have been proposed to quantify the ~~O₃-impacts~~ of O₃ on land carbon assimilation
64 ~~from~~at regional to global scales (Anav et al., 2011; Lam et al., 2023; Lei et al., 2020).
65 For example, Sitch et al. (2007) calculated ~~the~~ simultaneous ~~damages~~damage to both
66 photosynthesis and stomatal conductance ~~based on~~ the basis of instantaneous O₃
67 stomatal uptake. In contrast, Lombardozzi et al. (2012) estimated ~~the~~ decoupled
68 reductions in plant photosynthesis and stomatal conductance ~~using~~via different
69 response relationships to ~~the~~ cumulative O₃ stomatal uptake. ~~Applications~~The
70 application of different schemes has resulted in a wide range of reductions in gross

71 primary productivity (GPP) ~~by~~of 2–12% globally, with regional hotspots of up to 20–
72 –30% (Lombardozzi et al., 2015; Unger et al., 2020; Zhou et al., 2024).–

73 ~~The~~ O₃-induced inhibition ~~in~~of stomatal conductance decreases dry deposition and
74 consequently enhances surface O₃ concentrations (Clifton et al., 2020; Wesely and
75 Hicks, 2000; Zhang et al., 2006). Using the Sitch et al. (2007) scheme with high O₃
76 ~~damaging~~damage sensitivities in ~~the~~-ModelE2-YIBs (NASA GISS ModelE2 coupled
77 with the Yale Interactive terrestrial Biosphere model), Gong et al. (2020) revealed that
78 O₃-–vegetation interactions increased regional O₃ concentrations by 1.8 ppbv in the
79 eastern U.S., 1.3 ppbv in Europe, and ~~2.1 ppbv~~ 1 ppbv in eastern China ~~for the year in~~
80 2010. ~~As a~~In comparison, Sadiq et al. (2017) ~~found~~reported consistently stronger
81 feedback on O₃ concentrations in these polluted regions using via the scheme of
82 Lombardozzi et al. (2012) embedded in the Community Earth System Model (CESM).
83 Moreover, the inclusion of online O₃-–vegetation interactions in numerical models will
84 also result in a greater loss of simulated land carbon assimilation due to the feedbacks
85 of both ecosystems and surface O₃. This is attributable to several factors. On the one
86 hand, O₃ ~~damages~~damage to leaf photosynthesis ~~inhibit~~inhibits plant growth and
87 ~~decrease~~decreases the leaf area index (LAI), leading to ~~higher~~ a greater reduction
88 ~~percentage~~ in GPP ~~compared to~~ than in simulations without LAI changes (Yue et al.,
89 2020). On the other hand, ~~the~~ O₃ enhancement due to vegetation feedback may cause
90 additional vegetation damage and result in further GPP losses (Lei et al., 2021). As a
91 result, ~~the~~–O₃-–vegetation interactions should be considered in ~~the~~–global
92 ~~estimate~~estimates of O₃ ~~damages~~damage to ecosystem functions.

93 In addition to affecting surface O₃, ~~the~~–O₃-–vegetation ~~interaction~~interactions can
94 also alter ~~the~~-water and energy exchange between the land and atmosphere through the
95 modulation of stomatal conductance. For example, Lombardozzi et al. (2015) used the
96 Community Land Model (CLM) and estimated that the cumulative uptake of O₃ by ~~the~~
97 leaves resulted in a reduction of 2.2% in transpiration but an increase of 5.4% in runoff
98 globally. Arnold et al. (2018) used CESM and ~~found~~reported that plant exposure to O₃
99 could decrease ~~the~~-land-–air moisture fluxes and atmospheric humidity, which further
100 reduced shortwave cloud forcing in polluted regions and induced widespread surface

101 warming up to +1.5 K. Two recent studies utilized the WRF-~~chem~~Chem model and
102 revealed considerable warming and ~~the~~-associated meteorological perturbations due to
103 ~~the~~-O₃-vegetation interactions in China (Zhu et al., 2022; Jin et al., 2023). However,
104 all these modeling studies applied the same O₃ vegetation damage scheme proposed by
105 Lombardozzi et al. (2012). It's is necessary to assess the climatic responses to O₃-vegetation
106 interactions usingvia different schemes ~~so as~~ to explore ~~the~~-robust responses
107 and ~~the~~-associated uncertainties.—

108 In this study, we quantified the global impacts of O₃-vegetation
109 ~~interaction~~interactions on climatic conditions and surface air pollutants during the
110 2010s ~~using~~-thevia ModelE2-YIBs (Yue and Unger, 2015). This fully coupled
111 framework was implemented with the ~~semi-mechanistic~~semimechanistic O₃ damage
112 scheme proposed by Sitch et al. (2007), which ~~calculated~~calculates aggregated O₃
113 damage to photosynthesis ~~based on~~ the basis of varied sensitivities to instantaneous
114 stomatal O₃ uptake across eight plant functional types (PFTs). We performed sensitivity
115 experiments to quantify the responses of surface air temperature and precipitation to
116 O₃-vegetation ~~interaction~~interactions. The feedbacks to aerosols and O₃
117 concentrations were also examined.—

119 2 Method

120 2.1 Model descriptions

121 ~~The~~-ModelE2-YIBs is a fully coupled climate-carbon-chemistry model
122 ~~combining~~that combines the NASA GISS ModelE2 with the YIBs vegetation model.
123 ModelE2 is a general circulation model with ~~the~~a horizontal resolution of 2°×2.5° in
124 latitude and longitude and 40 vertical layers up to 0.1 hPa. It dynamically simulates
125 gas-phase chemistry (NO_x-HO_x-O_x-CO_x-CH₄-and NMVOCs), aerosols (sulfate,
126 nitrate, black and organic carbon, dust, and sea salt), and their interactions (Menon and
127 Rotstayn, 2006). Both the physical and chemical processes are calculated every 0.5 h,
128 and the radiation module is called every 2.5 h. The radiation module includes direct and
129 indirect aerosol radiative effects and accounts for the absorption of multiple greenhouse
130 gases (GHGs). For cloud optical parameters, ~~it uses~~-Mie scattering, ray tracing, and

matrix theory [are used](#) (Schmidt et al., 2006). The model outperforms 20 other IPCC-class climate models in simulating surface solar radiation (Wild et al., 2013) and has been extensively validated for meteorological and hydrological variables against observations and reanalysis data (Schmidt et al., 2014).-

The YIBs model employs the well-established Farquhar model for leaf photosynthesis and [the](#) Ball-Berry model for stomatal conductance (Farquhar et al., 1980; Ball et al., 1987) as follows:

$$A_{tot} = \min (J_c, J_e, J_s) \quad (1)$$

Here, the total leaf photosynthesis, denoted as A_{tot} ($\mu\text{mol m}^{-2}$ [leaf] s^{-1}), is calculated considering both C_3 (Collatz et al., 1991) and C_4 plants (Collatz et al., 1992). ~~The~~ A_{tot} is derived from the minimum value of the constraints. ~~The~~ [The rate of carboxylation limited by](#) ribulose-1,5-bisphosphate carboxylase (Rubisco) ~~limited rate of carboxylation~~ is J_c :-

$$J_c = \begin{cases} V_{cmax} \left(\frac{c_i - \Gamma_*}{c_i + K_c(1 + O_i/K_o)} \right) & \text{for } C_3 \text{ plant} \\ V_{cmax} & \text{for } C_4 \text{ plant} \end{cases} \quad (2)$$

The carboxylation rate restricted by the availability of light is J_e :

$$J_e = \begin{cases} a_{leaf} \times PAR \times \alpha \times \left(\frac{c_i - \Gamma_*}{c_i + 2\Gamma_*} \right) & \text{for } C_3 \text{ plant} \\ a_{leaf} \times PAR \times \alpha & \text{for } C_4 \text{ plant} \end{cases} \quad (3)$$

The export-limited ~~rate~~ [rates](#) for C_3 plants and the phosphoenolpyruvate carboxylase (PEPC) ~~-~~ limited ~~rate~~ [rates](#) of carboxylation for C_4 plants are represented by J_s :

$$J_s = \begin{cases} 0.5 V_{cmax} & \text{for } C_3 \text{ plant} \\ K_s \times V_{cmax} \times \frac{c_i}{P_{atm}} & \text{for } C_4 \text{ plant} \end{cases} \quad (4)$$

In these functions, V_{cmax} ($\mu\text{mol m}^{-2} \text{s}^{-1}$) is the maximum carboxylation capacity. c_i and O_i (Pa) represent the internal leaf CO_2 and oxygen partial pressure, [respectively](#). Γ_* (Pa) denotes the CO_2 compensation point, ~~while~~ [whereas](#) K_c and K_o (Pa) are Michaelis-Menten constants for the carboxylation and oxygenation of Rubisco, respectively. The parameters Γ_* , K_c , and K_o vary with temperature ~~based on~~ [the basis of](#) the sensitivity of the vegetation to temperature (Q_{10} coefficient). PAR ($\mu\text{mol m}^{-2} \text{s}^{-1}$) is the absorbed photosynthetically active radiation, a_{leaf} is [the](#) leaf-~~specific~~ [specific](#) light absorbance that considers sunlit and shaded leaves, and α is [the](#)

154 quantum efficiency. P_{atm} (Pa) represents the ambient pressure. K_s is set to 4000 as a
 155 constant following Oleson et al. (2010), to limit photosynthesis of C₄ plants ~~get that~~
 156 become saturated at lower CO₂ concentrations.

$$g_s = m \frac{(A_{tot} - R_d) \times RH}{c_s} + b \quad (5)$$

157 The stomatal conductance (g_s , mol [H₂O] m⁻² s⁻¹) is linked to the variations ~~of in~~
 158 A_{tot} with several parameters, such as the dark respiration rate (R_d , μmol m⁻² s⁻¹),
 159 relative humidity (RH), and CO₂ concentration at the leaf surface (c_s):

$$g_s = m \frac{(A_{tot} - R_d) \times RH}{c_s} + b \quad (5)$$

160 The model simulates the biophysical processes of eight PFTs, including tundra, C₃/C₄
 161 grass, shrubland, deciduous broadleaf forest, evergreen broadleaf forest, evergreen
 162 needleleaf forest, and cropland. Different values are assigned to parameters m and b for
 163 each PFT (Table S1). The carbon ~~uptake~~ taken up by the leaf ~~is~~ then
 164 ~~accumulated~~ accumulates and is allocated to different organs to support ~~the~~ plant
 165 development ~~with dynamical~~, resulting in dynamic changes in the LAI and tree growth.

167 2.2 ~~The O₃~~ vegetation damage scheme

168 The YIBs model employs a ~~semi-mechanistic~~ semimechanistic parameterization
 169 proposed by Sitch et al. (2007) to estimate the impact of O₃ on photosynthesis through
 170 stomatal uptake. The scheme applies an undamaged factor (F) (nmol m⁻² s⁻¹) to both
 171 A_{tot} and g_s as follows:-

$$A_{totd} = A_{tot} \cdot F \quad (6)$$

$$g_{sd} = g_s \cdot F \quad (7)$$

172 where A_{totd} and g_{sd} are the unaffected photosynthesis and stomatal conductance
 173 ~~separately~~, respectively. The factor F is defined as follows:

$$F = 1 - a_h \cdot \max [F_{O_3} - F_{O_3,crit}, 0.0] \quad (8)$$

174 where a_h (mmol m⁻² s⁻¹) is the high O₃ sensitivity coefficient, calibrated by Sitch et
 175 al. (2007) on data from field observations by Karlsson et al. (2004) and Pleijel et al.
 176 (2004) to represent the ‘high’ sensitivity of the relative species of each PFT. $F_{O_3,crit}$
 177 (nmol m⁻² s⁻¹) is the specific threshold for O₃ ~~damages~~ damage, both of which

178 ~~varies vary~~ with vegetation ~~type type~~ (Table S1 ~~-~~):

$$F_{O_3} = \frac{[O_3]}{R_a + \frac{k_{O_3}}{g_{sd}}}, \quad (9)$$

179 where $[O_3]$ represents surface O_3 concentrations, and R_a ($s\ m^{-1}$) ~~stands for~~ represents
180 aerodynamic resistance, which expresses the turbulent transport efficiency in
181 transferring sensible heat and water vapor between the land surface and a reference
182 height. The constant $k_{O_3}=1.67$ is the ratio of stomatal resistance ~~for~~ to O_3 , which is
183 estimated ~~based on~~ the basis of the theoretical stomatal resistance to water (Laisk et al.,
184 1989). When plants are exposed to $[O_3]$ (Eq. 9), A_{tot} and g_s ~~will~~ decrease (Eq. 6 and
185 Eq. 7) if the excess O_3 enters the leaves (Eq. 8). The increased stomatal resistance ~~aets~~
186 ~~to protect~~ protects plants by reducing the O_3 uptake of stomata. Consequently, the
187 damage scheme describes both changes in the photosynthetic rate and stomatal
188 conductance.

190 2.3 Experiments

191 To explore the coupled O_3 -vegetation effect, we performed two simulations using
192 the ModelE2-YIBs model. The control experiment “O3_offline” was conducted
193 without ~~the~~ O_3 ~~damages~~ damage to vegetation. ~~As a~~ For comparison, the sensitivity
194 experiment “O3_online” ~~contained~~ included online O_3 -vegetation
195 ~~interaction~~ interactions with high O_3 sensitivity. For both experiments, the
196 anthropogenic emissions ~~of~~ from 2010 (the average of 2005-2014) for 8 species (BC,
197 OC, CO, NH_3 , NO_x , SO_2 , ~~Alkenes~~ alkenes, and ~~Paraffin~~ paraffin) from 8 economic
198 sources (agriculture, energy, industry, transportation, ~~resident~~ residential, solvent, waste,
199 and international shipping) and biomass burning ~~source~~ sources were collected from the
200 Coupled Model Intercomparison Project phase 6 (CMIP6) (van Marle et al., 2017;
201 Hoesly et al., 2018). The ensemble ~~mean~~ means of the monthly sea surface temperature
202 (SST) and sea ice ~~fraction~~ concentration (SIC) simulated by 21 CMIP6 models during
203 the time period of 2005-2014 ~~was~~ were employed as the boundary conditions. The
204 cover ~~fraction~~ fractions of 8 PFTs (Fig. S1) fixed at 2010 were adopted from the land
205 use harmonization (LUH2) dataset (Hurt et al., 2020). For each time-slice simulation,

206 the model was run for 30 years with all the input data fixed, and the first 10 years
207 ~~are~~ were used as the spin-up period. We calculated the average of the last 20 years and
208 focused on the boreal summer season (June–July–August, JJA), when the interaction
209 of vegetation and surface O₃ reaches ~~the~~ its annual maximum ~~in one year~~ (~~fig.~~ (Fig. S3)).
210 ~~In order to~~ To show the uncertainty introduced by the internal variability of the model,
211 all the related global/regional values are denoted as “mean/sum ± standard deviation of
212 the last 20 model years”. We explored the climatic responses to O₃–vegetation
213 interactions as the differences between “O₃_online” and “O₃_offline” ~~on~~ at the global
214 scale, with ~~the~~ a focus ~~over the~~ on hotspot regions, such as the eastern U.S. (30–40° 40°N,
215 80–90° °W) and eastern China (22.5–38° 38°N, 106–122° 122°E).

217 2.4 Data for model evaluation

218 We evaluated the simulated air pollutants, carbon fluxes, and meteorological
219 variables from the ‘O₃_offline’ ~~run using simulation against~~ observational and
220 reanalysis datasets. The worldwide observations of the maximum daily 8-hour average
221 O₃ (MDA8 O₃) concentrations were ~~mainly~~ collected from three regional networks: the
222 Air Quality Monitoring Network operated by the Ministry of Ecology and Environment
223 (AQMN-MEE) in China, the Clean Air Status and Trends Network (CASTNET) in the
224 U.S., and the European Monitoring and Evaluation Programme (EMEP) in Europe.
225 ~~Observations~~ The observations used for validation beyond China, sourced from Sofen
226 et al. (2016), are averaged over the period 2005–2014. This dataset encompasses 7288
227 station records worldwide and excludes the uncertainty associated with high mountain-
228 top sites. For AQMN-MEE, the mean value of 2014–2018 was used ~~due to its~~
229 ~~establishment~~ because it was established in 2013. The simulated aerosol optical depth
230 (AOD) and LAI were validated using satellite-based data from the Moderate Resolution
231 Imaging Spectroradiometer (MODIS) retrievals collection 5 (Remer et al., 2005)
232 (<http://modis.gsfc.nasa.gov/>) averaged ~~for~~ over the years 2005–2014. The simulated
233 GPP was evaluated against the data product upscaled from the FLUXNET eddy
234 covariance measurements for 2009–2011 (Jung et al., 2011). The daily temperature at
235 2m m (T_{2m}) ~~in~~ from 2005–2014 was obtained from the National Centers for

236 Environmental Prediction/National Center for Atmospheric Research (NCEP/NCAR)
 237 reanalysis 1 (NCEP1) (Kalnay et al., 1996). For precipitation, we used the monthly data
 238 averaged ~~in~~from 2005–2014 from ~~the~~ Global Precipitation Climatology Project (GPCP)
 239 (Huffman et al., 1997; Adler et al., 2018). All these datasets were interpolated to the
 240 same resolution as ~~the~~ ModelE2-YIBs model. ~~Root~~The root-mean-square-error (RMSE)
 241 and normalized mean ~~biases~~ (NMBs) ~~bias~~ (NMB) were applied to quantify the deviations
 242 of ~~the~~ simulations from ~~the~~ observations:

$$243 \quad RMSE = \sqrt{\frac{1}{n} \sum_{i=1}^n (S_i - O_i)^2} \quad (10)$$

$$NMB = \sum_{i=1}^n (S_i - O_i) / \sum_{i=1}^n O_i \times 100\% \quad (11)$$

244 Here, S_i and O_i represent the simulated and observed values, respectively; n denotes
 245 the total ~~grid~~ number ~~of grid points~~ used in the comparisons.

246

247 3. Results

248 3.1 ~~The control~~Control simulation and model evaluations

249 We first evaluated the air pollutants simulated by the control simulation
 250 (~~O3_offline~~) of ~~the~~ ModelE2-YIBs model (Fig. 1). Over a total of 503 grids with site-
 251 level O_3 measurements (Fig. 1b), the model replicated both the magnitude and spatial
 252 distribution of MDA8 O_3 , with ~~a~~ correlation coefficient (r) of 0.59 and NMB of -2.54%
 253 (Fig. 1c). ~~Simulated~~The simulated summertime surface MDA8 O_3 ~~concentration~~ was
 254 high in regions with ~~large~~high anthropogenic emissions, such as western Europe and
 255 eastern China (Ohara et al., 2007), as well as in central Africa ~~with~~, ~~which has~~ frequent
 256 fire emissions (van der Werf et al., 2017). ~~On~~At the global scale, the model yielded an
 257 average MDA8 O_3 ~~concentration~~ of 43.93 ppbv, and observations ~~showed~~revealed an
 258 average of 44.72 ppbv over the same grids. However, the ~~modeled result is~~model
 259 overestimated ~~the concentrations~~ over the North China Plain and slightly
 260 underestimated ~~them~~ over the U.S., likely due to ~~the~~biases in the emission inventories
 261 and ~~the~~ predicted climate that drive ~~the~~ O_3 production. ~~Simulated~~The simulated AOD
 262 at 550 nm by ~~O3_offline~~ (Fig. 1d) showed ~~similar~~a spatial pattern ~~as similar to that of~~
 263 the satellite retrievals (Fig. 1e)), with ~~R=r~~ $r = 0.75$ and ~~an~~ NMB of -7.35% globally (Fig.

264 1f). Both ~~the~~ simulations and observations ~~showed~~revealed AOD hotspots over North
265 Africa and the Middle East, where dust emissions dominate, and in northern India and
266 eastern China, where anthropogenic emissions are ~~large~~high (Feng et al., 2020).—

267 We then evaluated the simulated GPP and LAI ~~by the~~via a control experiment for
268 the boreal summer period (Fig. 2). ~~Observations showed~~The observations revealed GPP
269 hotspots over boreal forests, such as those in the eastern U.S., Eurasia, and East Asia,
270 and ~~the~~ tropical forests, such as those in the Amazon, ~~central~~Central Africa, and
271 Indonesia (Fig. 2b). The seasonal total GPP was estimated to be 41.63 Pg[C],
272 which accounted for 35% of the annual amount. ~~Simulations~~The simulations captured
273 the observed GPP pattern ~~on~~at the global scale, with $r = 0.64$ and $NMB = -7.81\%$
274 ~~over~~across 2581 grids (Fig. 2c), with ~~underestimation~~underestimations in the tundra
275 area and slight ~~overestimation~~overestimations in the tropical ~~rain forest~~rainforest and
276 evergreen forest regions. The model simulated a seasonal total GPP of 38.69 Pg[C],
277 equivalent to 34% of the annual amount. ~~Simulated~~The simulated LAI showed ~~similar~~
278 patterns ~~as similar to those of~~GPP (Fig. 2d) and resembled the observed LAI (Fig. 2e),
279 with a spatial correlation of $r = 0.79$ and a low $NMB = -5.43\%$ ~~over~~across 4435 grids
280 globally (Fig. 2f).—

281 We further validated the simulated meteorology from O3_offline (Fig. S2). For the
282 surface air temperature, the model (Fig. S2a) reproduced the observed (Fig. S2b) pattern,
283 with an RMSE of 3.21 °C and an r of 0.99 ~~against~~compared with the observations (Fig.
284 S2c). For precipitation, the simulation (Fig. S2d) ~~captures~~captured the observed spatial
285 pattern (Fig. S2e), with $NMB = 17.26\%$ and $r = 0.75$ (Fig. S2f). Overall, the model
286 ~~captures~~captured the spatial characteristics and magnitudes of air pollutants, biospheric
287 parameters, and meteorological fields, making it a valuable tool for studying O₃—
288 vegetation interactions.—

290 3.2 O₃ damage to terrestrial ecosystems

291 We assessed the damaging effects of surface O₃ ~~to~~on ecosystems due to online O₃—
292 —vegetation interactions (Fig. 3). The impacts of O₃ on biospheric variables were
293 mainly located in regions characterized by abundant vegetation cover and elevated O₃

294 concentrations. On the global scale, O₃ induced ~~the~~ GPP reduction of -1.80 ± 0.61 PgC
295 yr⁻¹ ($-4.69 \pm 1.56\%$, Fig. 3a). This deleterious effect was more pronounced in specific
296 regions, notably eastern China and eastern U.S., with significant GPP declines of -
297 $25.40 \pm 1.90\%$ and $-20.14 \pm 5.02\%$, respectively, under high O₃ sensitivity conditions (Fig.
298 3a and Table S2). ~~Meanwhile~~ ~~Moreover~~, stomatal conductance significantly decreased
299 in the middle latitudes of ~~the~~ Northern Hemisphere (Fig. 3b). The most substantial
300 relative change of $-30.62 \pm 4.30\%$ was observed in eastern China, followed by -
301 $25.65 \pm 9.32\%$ in the eastern U.S. (Fig. 3b and Table S2). ~~Though~~ ~~Although~~ there are
302 positive responses in some regions, they are not dominant and ~~are~~ hardly significant.
303 These values were stronger than ~~that~~ ~~those~~ for GPP (Fig. 3a), likely ~~due to~~ ~~because of~~ the
304 climatic feedback to O₃-~~vegetation~~ interactions. The opening of ~~the~~ plant stoma plays
305 a crucial role in regulating ~~the~~ energy and water exchange between ~~the~~ land surface and
306 the atmosphere. The inhibition of stomatal conductance by surface O₃ leads to ~~the~~
307 warmer (Fig. 4a) and drier (Fig. 4b) climate in those hotspot regions, resulting in even
308 stronger ~~inhibition~~ ~~inhibitory~~ effects on stomatal conductance. Following the changes in
309 GPP, ~~the~~ global LAI on average decreased by 0.01 ± 0.01 m² m⁻² ($-0.62 \pm 0.84\%$), with
310 regional ~~maximums~~ ~~maxima~~ of $-4.53 \pm 1.14\%$ in eastern China and $-5.87 \pm 3.11\%$ in ~~the~~
311 eastern U.S. (Table S2).-

313 3.3 Global climatic responses to O₃-~~vegetation~~ interactions

314 In response to the O₃-induced inhibition of stomatal conductance, surface air
315 temperature increased by 0.05 ± 0.20 °C (Fig. 4a) ~~while~~, ~~whereas~~ precipitation
316 decreased by -0.01 ± 0.03 mm day⁻¹ (Fig. 4b) ~~on~~ ~~at~~ the global scale. The most significant
317 change was the warming of 0.56 ± 0.38 °C and precipitation reduction of -0.79 ± 1.05 mm
318 day⁻¹ ($-16.18 \pm 20.38\%$) in eastern China (Table S3), ~~following~~ ~~followed by~~ the
319 ~~largest~~ ~~greatest~~ inhibition ~~to~~ ~~of~~ stomatal conductance (Fig. 3b). Such warming and
320 rainfall ~~deficit~~ ~~deficits~~ also appeared in ~~the~~ eastern U.S. and western Europe, where the
321 O₃-~~vegetation~~ interactions were notable. The O₃-induced inhibition ~~to~~ ~~of~~ stomatal
322 conductance decreased ~~the~~ latent heat flux (Fig. 4e) and ~~the~~ consequent precipitation

(Fig. 4b) in those hotspot regions. ~~Meanwhile~~ ~~Moreover~~, the ~~reduction of~~ ~~reduced~~ latent heat flux ~~promotes~~ ~~promoted~~ ~~higher~~ surface air ~~temperature~~ ~~temperatures~~ (Fig. 4a), resulting in ~~the~~ ~~an~~ increase ~~of~~ ~~in~~ ~~the~~ sensible heat flux (Fig. 4f). Such warming ~~was~~ ~~has~~ also ~~been~~ reported in field experiments, where relatively high O₃ exposure resulted in noticeable increases ~~of~~ ~~in~~ canopy temperature along with reductions ~~of~~ ~~in~~ transpiration (Bernacchi et al., 2011; VanLoocke et al., 2012). Globally, temperature and precipitation showed patchy responses with both positive and negative anomalies, suggesting that the regional hotspots of O₃-induced meteorological changes propagate to surrounding areas through atmospheric perturbations. –

We further examined the changes in air humidity and ~~cloudiness~~. ~~Surface~~ ~~cloud~~ ~~cover~~. ~~The~~ ~~surface~~ relative humidity decreased by $-0.18 \pm 0.53\%$ globally, with a similar pattern as that of precipitation (Fig. 4c). The most significant reductions ~~were~~ ~~occurred~~ over eastern China and ~~the~~ eastern U.S., where both ~~the~~-warming (Fig. 4a) and rainfall deficit (Fig. 4b) contributed to ~~the~~-drought. However, in ~~the~~-adjacent regions, such as northern China and ~~the~~ central U.S., both rainfall and surface relative humidity ~~showed~~ ~~certain~~ ~~enhancement~~. ~~increased~~. These changes were associated with the regional increase ~~of~~ ~~in~~ cloud cover (Fig. 4d). The sensible heat flux increased by $6.3 \pm 5.4 \text{ W m}^{-2}$ ($16.54 \pm 15.59\%$) and $7.12 \pm 3.86 \text{ W m}^{-2}$ ($25.46 \pm 14.71\%$) in ~~the~~ eastern U.S. and eastern China, respectively, suggesting ~~a~~ ~~the~~ transfer of thermal energy from ~~the~~ land to the atmosphere ~~by~~ ~~via~~ O₃-vegetation interactions (Fig. 4f and Table S3). The warming effect further triggered anomalous updrafts in the lower troposphere, represented by ~~the~~ changes in vertical velocity (Fig. 5), leading to enhanced convection, reduced atmospheric stability, and consequently an increase in low-level ~~cloudiness~~ ~~cloud~~ ~~cover~~ (Fig. 4d). However, despite the usual cooling effect associated with increased cloud cover due to reductions in radiation, in regions predominantly influenced by O₃-vegetation interactions, this cooling effect was outweighed by ~~the~~-O₃-induced warming through ~~the~~ inhibition of stomatal conductance. Therefore, temperatures exhibited an overall increase of $0.56 \pm 0.38 \text{ }^\circ\text{C}$ in eastern China and $0.33 \pm 0.87 \text{ }^\circ\text{C}$ in the eastern U.S. (Table S3).

3.4 Changes ~~of~~in air pollution caused by O₃-vegetation interactions

Changes in surface water and heat fluxes induced by O₃-vegetation interactions could feed back to affect air pollutants, such as O₃ and aerosols. As shown in Fig. 6a and Table S4-~~show~~, surface MDA8 O₃ concentrations ~~enhanced~~increased by 1.46±3.02 ppbv in eastern China and 1.15±1.77 ppbv in the eastern U.S. due to ~~the~~ decreased dry deposition following O₃ inhibition ~~on~~of stomatal conductance. ~~It~~This indicates that ~~the~~ high contemporary O₃ pollution may worsen air quality through O₃-vegetation interactions. However, negative O₃ changes were predicted in the central U.S. and western China, where ~~the~~ increased rainfall dampened O₃ through chemical reactions and wet deposition. On a global scale, surface MDA8 O₃ concentrations showed a limited increase of 0.03±0.4 ppbv ~~due to~~because of the offset between positive and negative feedbacks. The ~~enhancement of~~increase in O₃ concentrations in polluted regions may exacerbate the warming effect of O₃ as a greenhouse gas and cause additional ~~damages~~damage to vegetation. For ~~instance, example, the effects of~~ offline O₃ ~~damages~~damage on GPP in eastern China and the eastern US ~~are~~were simulated to be -0.52±0.03 Pg[C] (-24.98±0.91%) and -0.17±0.02 Pg[C] (-16.71±1.16%), respectively, which are smaller than those induced by O₃-vegetation interactions (Table S2).

Aerosols also exhibited evident changes ~~by the~~in response to O₃-vegetation interactions. The AOD ~~showed significant reductions~~significantly decreased over ~~the~~ hotspot regions, such as eastern China and the eastern U.S. (Fig. 6b). In the ModelE2-YIBs model, sulfate was especially sensitive to ~~cloud~~clouds, which could enhance ~~the~~ aerosol scavenging through cloud water precipitation (Koch et al., 2006). The large ~~enhancement of cloudiness~~increase in cloud cover removed sulfate more efficiently than the other aerosol species did, leading to an average ~~decline~~decrease of -1.94±1.67 μg m⁻³ (-8.52±6.88%) in the PM_{2.5} loading over eastern China (Fig. S4 and Table S4). ~~Meanwhile~~Moreover, the reduction ~~of~~in surface relative humidity (Fig. 4c) in the regions with strong O₃-vegetation interactions limited the hygroscopic growth of aerosols, leading to a more noticeable decrease in AOD (Petters and Kreidenweis, 2007; Pitchford et al., 2007) by -0.06±0.05 (-14.67±16.75%) in eastern China (Table S4). ~~The~~

383 ~~similar~~ Similar aerosol changes were found in ~~the~~ eastern U.S. but with smaller
384 reductions ~~of~~in PM_{2.5} by $-0.27 \pm 0.36 \mu\text{g m}^{-3}$ ($-6.01 \pm 7.9\%$) and AOD by -0.01 ± 0.01 ($-$
385 $8.15 \pm 9.38\%$) (Table S4). ~~Beyond~~In addition to the key O₃-vegetation coupling regions,
386 positive but insignificant changes in AOD were predicted, leading to ~~the~~ moderate AOD
387 changes ~~on~~at the global scale (Fig. 6b).-

389 4. Discussion and conclusions

390 We examined ~~the~~O₃-vegetation feedback to climate and air pollution in the 2010s
391 ~~using~~via the fully coupled climate-carbon-chemistry model ModelE2-YIBs. During
392 boreal summer, surface O₃ resulted in strong ~~damages~~damage to GPP and ~~inhibitions~~
393 ~~to~~inhibited stomatal conductance with regional hotspots over eastern China and ~~the~~
394 eastern U.S. Consequently, surface transpiration was weakened, leading to decreased
395 latent heat fluxes and relative humidity but increased surface air temperature.
396 ~~Meanwhile,~~~~the~~Moreover, surface warming increased cloud cover by reducing
397 atmospheric stability. However, the ~~enhancement of cloudiness~~increase in cloud cover
398 decreased ~~the~~ surface temperature and promoted precipitation outside the key regions
399 with intense O₃-vegetation interactions. The O₃-induced inhibition ~~to~~of stomatal
400 conductance resulted in a localized increase in O₃ ~~concentrations~~concentration. In
401 contrast, the increased cloud cover and decreased relative humidity jointly reduced ~~the~~
402 AOD in hotspot regions. ~~On~~At the global scale, the mean changes ~~of~~in both climate and
403 air pollution were moderate ~~due to~~because of the offset between the changes with
404 opposite signs.-

405 Our ~~predicted~~predictions of the changes in water/heat fluxes ~~caused~~ by O₃-
406 vegetation interactions were consistent with ~~those of~~ previous studies (Lombardozzi et
407 al., 2015; Arnold et al., 2018; Gong et al., 2020). For example, ~~the~~ simulations by
408 Lombardozzi et al. (2015) revealed that surface O₃ reduces global GPP by 8% - 12%
409 and transpiration by 2% - 2.4% , with regional reductions ~~of~~up to 20% for GPP and 15%
410 for transpiration in eastern China and ~~the~~ U.S. These changes ~~were in general~~are
411 ~~generally~~ consistent with our results ~~though,~~ although we predicted ~~larger~~greater
412 reductions in transpiration than ~~in~~ GPP due to O₃-vegetation interactions. Using the

413 same scheme as Lombardozzi et al. (2015), Sadiq et al. (2017) ~~showed~~reported that O₃-
414 -vegetation coupling induced ~~the~~ surface warming of 0.5-~~1~~0.5-1 °C and O₃
415 enhancement of 4-6 ppbv in eastern China and eastern U.S. The magnitude of these
416 responses was much stronger than our predictions, likely because they considered the
417 accumulation effect of O₃. In contrast, ~~the~~ regional simulations by Jin et al. (2023)
418 revealed that O₃--vegetation coupling led to ~~the~~ increases of temperature of up to
419 0.16 °C and surface O₃ of up to 0.6 ppbv in eastern China, both of which were smaller
420 than our predictions. The damage scheme they use, which depends on cumulative O₃
421 uptake, omits the difference in impact on sunlit or shaded leaves and ~~will~~
422 ~~overestimate~~overestimates the O₃ damage onto GPP compared ~~to~~with the scheme we
423 use, which considers transient O₃ flux (Cao et al., 2024). The discrepancies of O₃--
424 vegetation ~~feedback using~~feedbacks when the same O₃ damage schemes were used
425 revealed ~~the~~ uncertainties from in the climate and chemistry models. Our predictions
426 were within the range of previous estimates for both climatic and O₃ changes.—

427 There were ~~some~~several limitations in our simulated O₃--vegetation interactions.
428 First, the ~~semi-meehanistic~~semimechanistic O₃ damage scheme we used in the present
429 study linked ~~the damages~~damage to photosynthesis with ~~those~~damage to stomatal
430 conductance (Sitch et al., 2007), leading to ~~stronger~~a greater percentage of inhibition
431 percentage in of stomatal conductance than ~~that in of~~ photosynthesis, considering ~~the~~ O₃-
432 -vegetation ~~feedback~~feedbacks. However, some observations ~~showed~~have shown that
433 ~~the~~ damage to stomatal conductance ~~occurred~~occurs more slowly and might not be
434 proportional to the decline of photosynthetic rates (Gregg et al., 2006; Lombardozzi
435 et al., 2012). Second, observations have shown large variability of plant
436 ~~sensitivities~~sensitivity to O₃ ~~damages~~damage. The Sitch et al. (2007) scheme ~~employed~~
437 ~~the~~employs low to high ranges of sensitivity to indicate ~~the inter-specific~~interspecific
438 variabilities. In this study, we employed only ~~the~~ high O₃ sensitivity to explore the
439 maximum responses. The possible uncertainties due to varied O₃ damage sensitivities
440 ~~deserved~~deserve further ~~investigations~~investigation. Third, large-scale observations
441 were not available to validate the simulated regional to global responses of climate and

442 air pollutants. The O₃ vegetation damage scheme ~~was~~has been extensively validated
443 against site-level measurements of both photosynthesis (Yue and Unger, 2018) and
444 stomatal conductance (Yue et al., 2016). However, we were conservative about the
445 derived global responses given that previous studies ~~showed~~have shown large
446 discrepancies ~~using~~when the same O₃ damage scheme is used but have been
447 implemented in different climate and/or chemistry models (-Lombardozzi et al., 2015;
448 Sadiq et al., 2017; Jin et al., 2023). Furthermore, the 2°×2.5° resolution of the current
449 version of the ModelE2-YIBs model has ~~limitation~~limitations due to the high
450 computational demands. However, high-resolution models exhibit improved
451 simulations of extreme events (Chang et al., 2020; Ban et al., 2021), which have certain
452 ~~effect~~effects on O₃-vegetation interactions (Mills et al., 2016; Lin et al., 2020). While
453 chemical transport models with relatively coarse ~~resolution~~resolutions can ~~raise~~increase
454 biases in simulated air pollutants, they still capture large-scale patterns similar to those
455 of fine-resolution results and ~~is reasonable~~~~compared~~compare reasonably well to
456 observational data (Wang et al., 2013; Li et al., 2016; Lei et al., 2020). Moreover, we
457 omit the slow climatic feedback caused by air-sea ~~interaction~~interactions in the
458 simulations. Studies have revealed that these interactions may result in different
459 climatic perturbations from those in simulations with fast responses of the land surface
460 alone (Yue et al., 2011). A dynamic ocean model ~~is considered to~~would enrich ~~the~~ future
461 research. ~~Meanwhile~~Moreover, this study ~~does~~did not isolate the different impacts of
462 aerosols, even though the radiation module ~~includes~~included both direct and indirect
463 radiative effects. We will investigate this further in the future by identifying the main
464 processes involved.

465 Despite these uncertainties, our simulations revealed considerable changes ~~of~~in
466 both climate and air pollutants in response to O₃-vegetation interactions. The most
467 intense warming, dryness, and O₃ enhancement were predicted in eastern China and the
468 eastern U.S., affecting the regional climate and threatening public health for these top
469 two economic centers. In contrast, ~~we~~ for the first time, we revealed ~~the~~a reduction ~~of~~in
470 aerosol loading in ~~these~~ hotspot regions, suggesting both positive and negative effects
471 ~~to~~on air pollutants ~~by~~via O₃-vegetation feedback. Such interactions should be

472 considered in ~~the~~ Earth system models ~~so as~~ to better project future changes in climate
473 and air pollutants following ~~the~~ anthropogenic interventions to both O₃ precursor
474 emissions and ecosystem functions.—

475

476 **Data Availability**

477 The observational data and model outputs that support the findings in this study are
478 available from [the](#) corresponding authors upon reasonable request.

479

480 **Author contributions**

481 XY conceived the project. XZ performed the model simulations, conducted ~~results~~[the](#)
482 analysis and wrote the draft manuscript. XY, CT and XL assisted in the interpretation
483 of the results and contributed to the discussion and improvement of the paper.

484

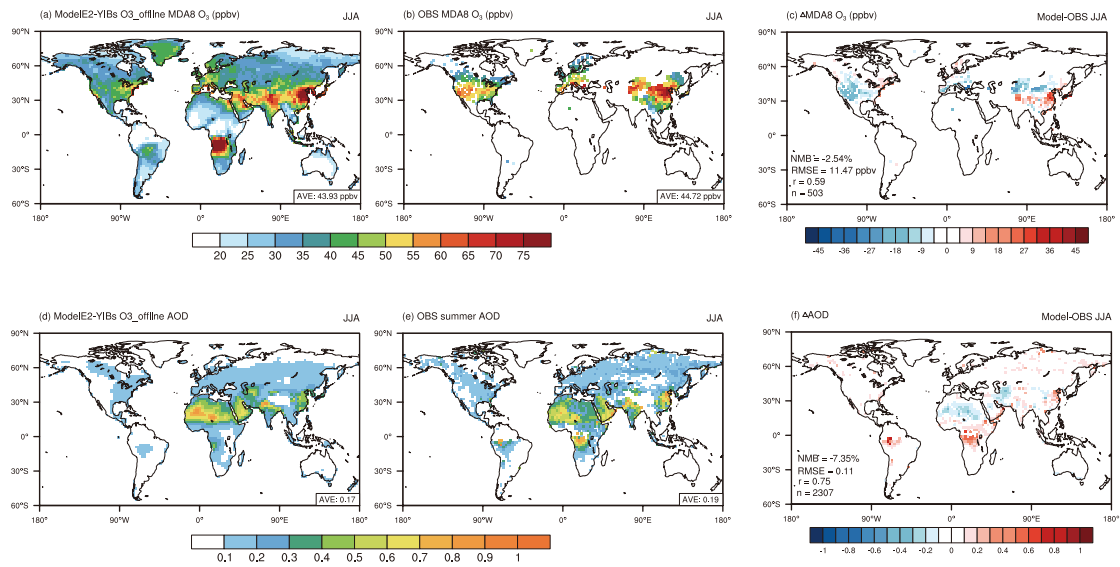
485 **Competing interests**

486 The authors declare that they have no ~~conflict~~[conflicts](#) of interest.

487

488 **Acknowledgments**

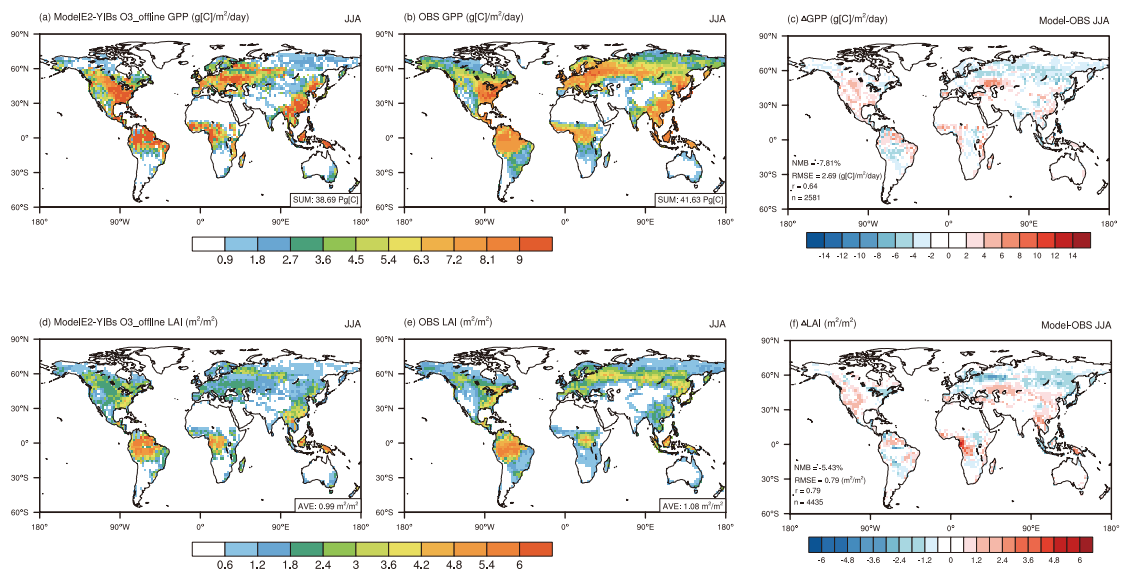
489 This study was jointly funded by [the](#) National Key Research and Development Program
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491 (no. 42293323).



492

493 **Figure 1.** Evaluation of the present-day boreal summertime (June–August) air
 494 pollutants at the present day simulated by the ModelE2-YIBs model. Surface
 495 the daily maximum 8-hour ozone (MDA8 O₃; (a-c) and aerosol optical depth (AOD; (d-f)
 496 from the O3 offline simulation O3 offline (a & d) and observations (b & e) are
 497 compared. The correlation coefficients coefficient (r), root-mean-square error (RMSE),
 498 normalized mean bias (NMB), and number of grid cells (n) for the comparisons are
 499 listed on the mean bias maps (c & f).

500

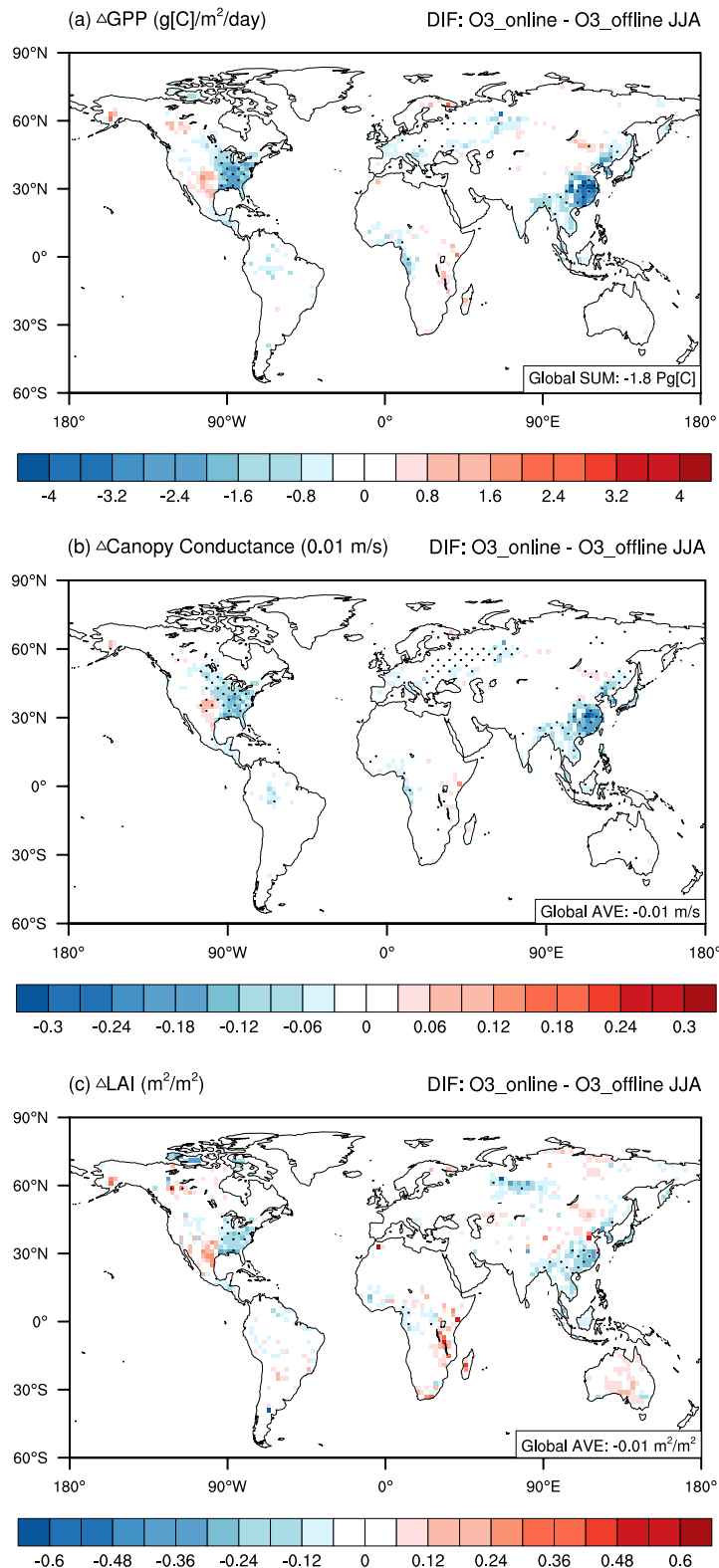


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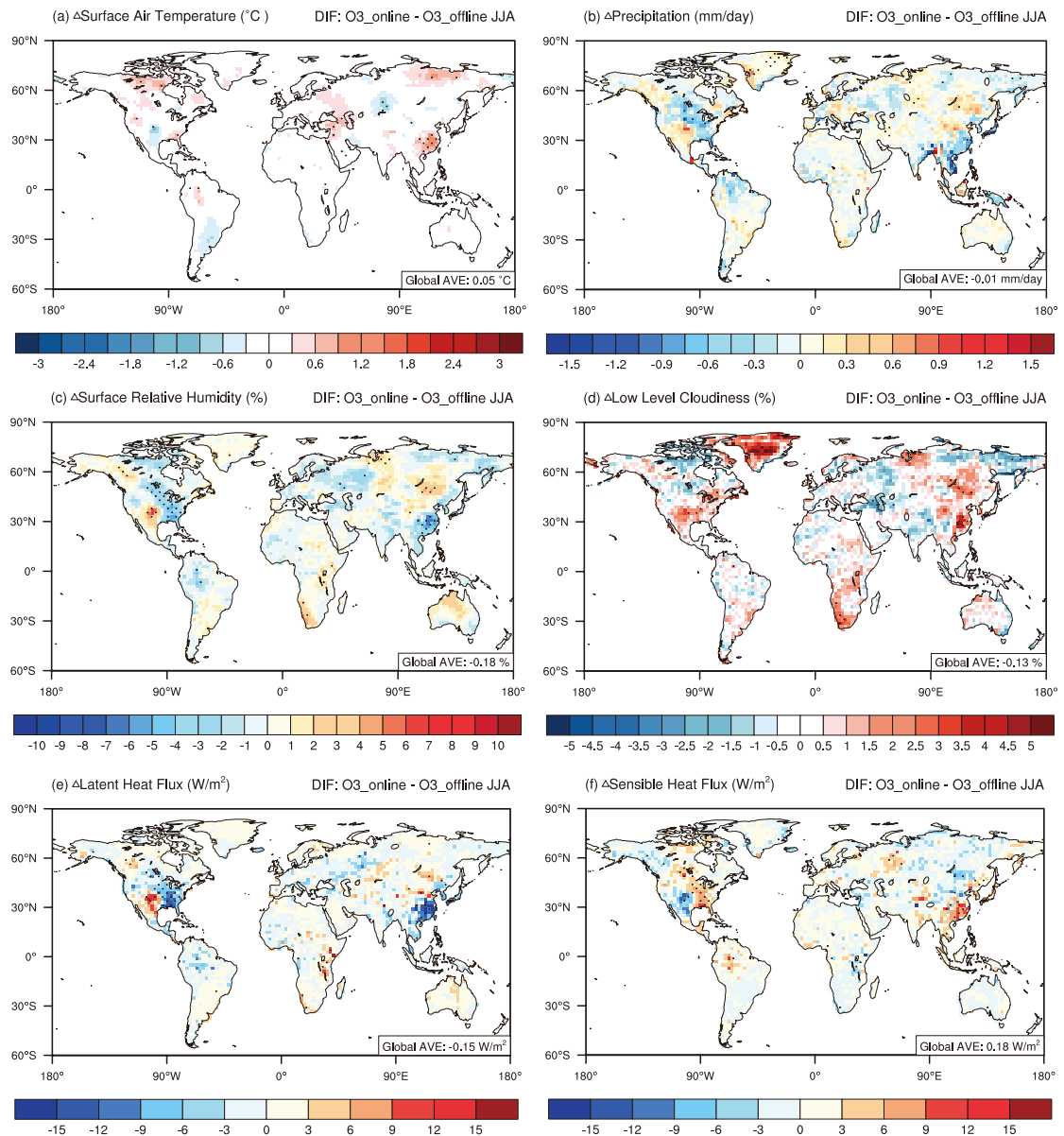
503

Figure 2. The same Same as in Fig. 1 but for gross primary productivity (GPP; a-c) and the leaf area index (LAI; d-f).



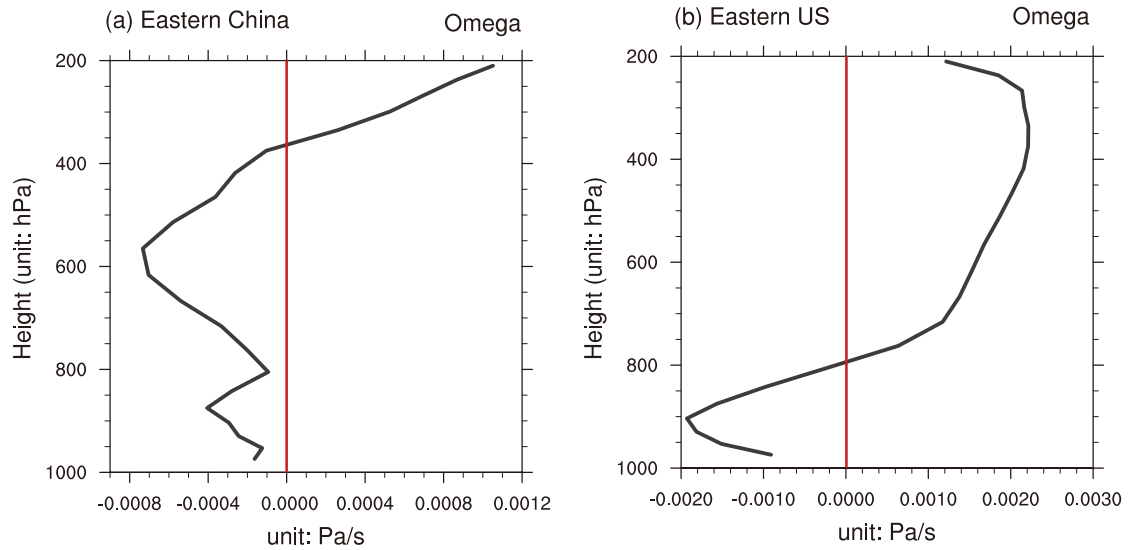
504

505 **Figure 3.** Changes ~~of~~ in present-day boreal summertime biospheric variables induced
 506 by O₃-vegetation interactions at the present day. Results. The results shown are the
 507 changes ~~of~~ in (a) GPP, (b) canopy conductance, and (c) LAI between simulationsthe
 508 O3_online and O3_offline. ~~Black~~ simulations. The black dots denote areas with
 509 significant changes ($p < 0.1$). ~~—~~



510

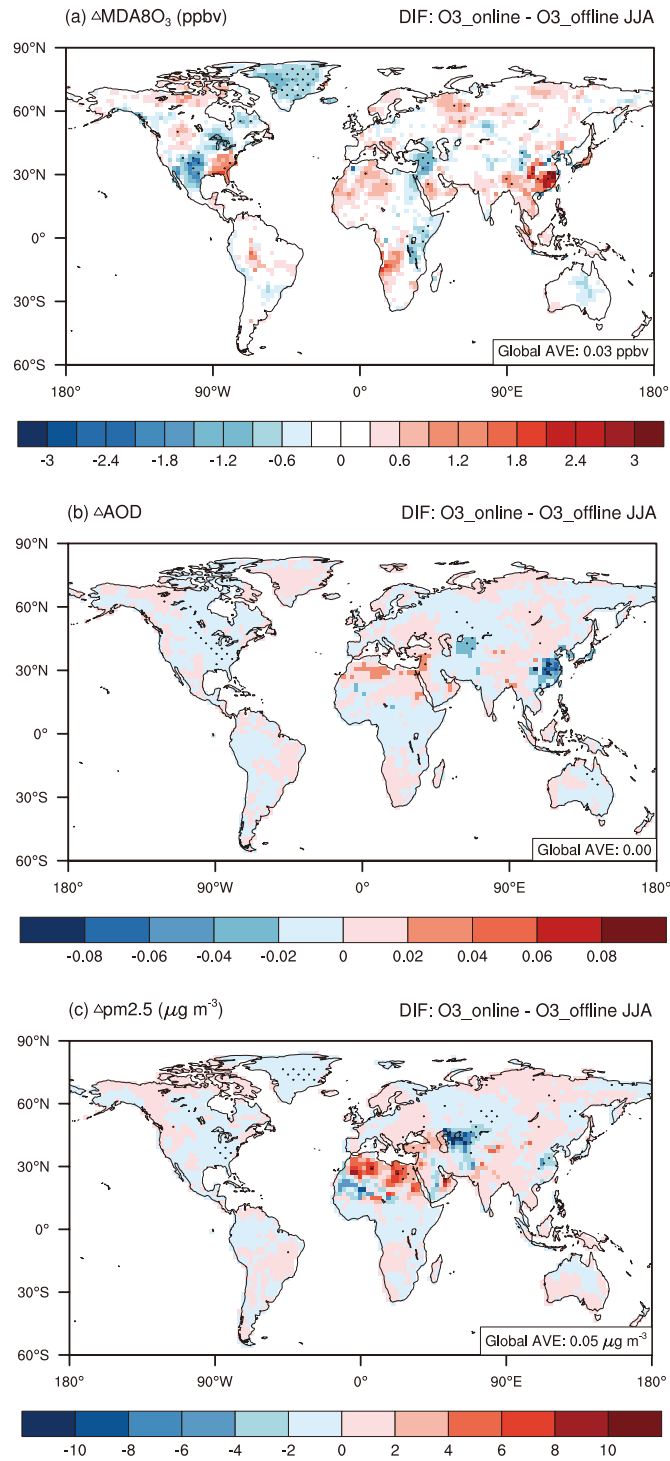
511 **Figure 4.** Changes ~~of~~in present-day boreal summertime meteorological fields ~~caused~~
 512 by O₃-vegetation interactions ~~at the present day.~~ **Results.** ~~The results~~ shown are
 513 changes ~~of~~in (a) surface air temperature, (b) precipitation, (c) surface relative humidity,
 514 (d) low-level cloudiness, (e) latent heat flux, and (f) sensible heat flux between
 515 ~~simulations~~the O₃_online and O₃_offline ~~simulations~~. For heat fluxes, positive values
 516 (shaded in red ~~color~~) indicate ~~that~~ the upward fluxes change. ~~Black~~The black
 517 dots denote areas with significant changes ($p < 0.1$).-



519

520 **Figure 5.** Vertical profile of the vertical velocity. ~~Results~~ The results shown are changes
 521 of in the vertical velocity in (a) ~~Eastern~~ eastern China and (b) ~~Eastern US~~ the eastern U.S.
 522 between ~~simulations~~ the O3_online and O3_offline. ~~Solid~~ simulations. The solid red line
 523 denotes ~~the value~~ 0. Please ~~notice~~ note the differences in the scales.

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Fig. 6. Changes of in present-day summertime atmospheric pollution caused by O₃-vegetation interactions at present day. **Results.** The results shown are the changes of in (a) O₃, (b) AOD, and (c) PM_{2.5} between the O3_online and O3_offline. **Black simulations.** The black dots denote areas with significant changes ($p < 0.1$).-

531 **References**

- 532 Adler, R. F., Sapiano, M. R. P., Huffman, G. J., Wang, J.-J., Gu, G., Bolvin, D., Chiu,
533 L., Schneider, U., Becker, A., Nelkin, E., Xie, P., Ferraro, R., and Shin, D.-B.: The
534 Global Precipitation Climatology Project (GPCP) Monthly Analysis (New Version
535 2.3) and a Review of 2017 Global Precipitation, *Atmosphere*, 9, 138,
536 <https://doi.org/10.3390/atmos9040138>, 2018.
- 537 Ainsworth, E. A., Yendrek, C. R., Sitch, S., Collins, W. J., and Emberson, L. D.: The
538 effects of tropospheric ozone on net primary productivity and implications for
539 climate change, *Annu. Rev. Plant. Biol.*, 63, 637–661,
540 <https://doi.org/10.1146/annurev-arplant-042110-103829>, 2012.
- 541 Anav, A., Menut, L., Khvorostyanov, D., and Viovy, N.: Impact of tropospheric ozone
542 on the Euro-Mediterranean vegetation, *Glob. Change. Biol.*, 17, 2342–2359,
543 <https://doi.org/10.1111/j.1365-2486.2010.02387.x>, 2011.
- 544 Arnold, S. R., Lombardozzi, D., Lamarque, J. -F., Richardson, T., Emmons, L. K.,
545 Tilmes, S., Sitch, S. A., Folberth, G., Hollaway, M. J., and Val Martin, M.:
546 Simulated Global Climate Response to Tropospheric Ozone-Induced Changes in
547 Plant Transpiration, *Geophys. Res. Lett.*, 45, 13070–13079,
548 <https://doi.org/10.1029/2018GL079938>, 2018.
- 549 Ball, J. T., Woodrow, I. E., and Berry, J. A.: A Model Predicting Stomatal Conductance
550 and its Contribution to the Control of Photosynthesis under Different
551 Environmental Conditions, in: *Progress in Photosynthesis Research*, edited by:
552 Biggins, J., Springer Netherlands, Dordrecht, 221–224,
553 https://doi.org/10.1007/978-94-017-0519-6_48, 1987.
- 554 Ban, N., Caillaud, C., Coppola, E., Pichelli, E., Sobolowski, S., Adinolfi, M., Ahrens,
555 B., Alias, A., Anders, I., Bastin, S. and Belušić, D.: The first multi-model ensemble
556 of regional climate simulations at kilometer-scale resolution, part I: evaluation of
557 precipitation, *Clim. Dynam.*, 57, 275-302, [https://doi.org/10.1007/s00382-021-](https://doi.org/10.1007/s00382-021-05708-w)
558 [05708-w](https://doi.org/10.1007/s00382-021-05708-w), 2021.
- 559 Buker, P., Feng, Z., Uddling, J., Briolat, A., Alonso, R., Braun, S., Elvira, S., Gerosa,
560 G., Karlsson, P. E., Le Thiec, D., Marzuoli, R., Mills, G., Oksanen, E., Wieser, G.,
561 Wilkinson, M., and Emberson, L. D.: New flux based dose-response relationships
562 for ozone for European forest tree species, *Environ. Pollut.*, 206, 163–174,
563 <https://doi.org/10.1016/j.envpol.2015.06.033>, 2015.
- 564 Bernacchi, C. J., Leakey, A. D. B., Kimball, B. A., and Ort, D. R.: Growth of soybean
565 at future tropospheric ozone concentrations decreases canopy evapotranspiration
566 and soil water depletion, *Environ. Pollut.*, 159, 1464–1472,
567 <https://doi.org/10.1016/j.envpol.2011.03.011>, 2011.
- 568 Cao, J., Yue, X. and Ma, M.: Simulation of ozone-vegetation coupling and feedback in

- 569 China using multiple ozone damage schemes, *Atmos. Chem. Phys.*, 24(7), 3973-
570 3987, <https://doi.org/10.5194/acp-24-3973-2024>, 2024.
- 571 Chang, P., Zhang, S., Danabasoglu, G., Yeager, S.G., Fu, H., Wang, H., Castruccio, F.S.,
572 Chen, Y., Edwards, J., Fu, D. and Jia, Y.: An unprecedented set of high-resolution
573 earth system simulations for understanding multiscale interactions in climate
574 variability and change, *J. Adv. Model. Earth. Sy.*, 12,
575 <https://doi.org/10.1029/2020MS002298>, 2020.
- 576 Clifton, O. E., Paulot, F., Fiore, A. M., Horowitz, L. W., Correa, G., Baublitz, C. B.,
577 Fares, S., Goded, I., Goldstein, A. H., Gruening, C., Hogg, A. J., Loubet, B.,
578 Mammarella, I., Munger, J. W., Neil, L., Stella, P., Uddling, J., Vesala, T., and
579 Weng, E.: Influence of Dynamic Ozone Dry Deposition on Ozone Pollution, *J.*
580 *Geophys. Res-Atmos.*, 125, e2020JD032398,
581 <https://doi.org/10.1029/2020JD032398>, 2020.
- 582 Collatz, G. J., Ball, J. T., Grivet, C., and Berry, J. A.: Physiological and Environmental-
583 Regulation of Stomatal Conductance, Photosynthesis and Transpiration – a Model
584 That Includes a Laminar Boundary-Layer, *Agr. Forest Meteorol.*, 54, 107–136,
585 doi:10.1016/0168-1923(91)90002-8, 1991.
- 586 Collatz, G. J., Ribas-Carbo, M., and Berry, J. A.: Coupled Photosynthesis-Stomatal
587 Conductance Model for Leaves of C4 Plants, *Aust. J. Plant Physiol.*, 19, 519–538,
588 <https://doi.org/10.1071/PP9920519>, 1992.
- 589 Dizengremel, P.: Effects of ozone on the carbon metabolism of forest trees, *Plant.*
590 *Physiol. Bioch.*, 39, 729–742, [https://doi.org/10.1016/S0981-9428\(01\)01291-8](https://doi.org/10.1016/S0981-9428(01)01291-8),
591 2001.
- 592 Farquhar, G. D., von Caemmerer, S., and Berry, J. A.: A biochemical model of
593 photosynthetic CO₂ assimilation in leaves of C₃ species, *Planta*, 149, 78–90,
594 <https://doi.org/10.1007/BF00386231>, 1980.
- 595 Feng, L., Smith, S. J., Braun, C., Crippa, M., Gidden, M. J., Hoesly, R., Klimont, Z.,
596 van Marle, M., van den Berg, M., and van der Werf, G. R.: The generation of
597 gridded emissions data for CMIP6, *Geosci. Model Dev.*, 13, 461–482,
598 <https://doi.org/10.5194/gmd-13-461-2020>, 2020.
- 599 Fiscus, E. L., Booker, F. L., and Burkey, K. O.: Crop responses to ozone: uptake, modes
600 of action, carbon assimilation and partitioning, *Plant. Cell. Environ.*, 28, 997–1011,
601 <https://doi.org/10.1111/j.1365-3040.2005.01349.x>, 2005.
- 602 Fuhrer, J., Skärby, L. and Ashmore, M.R.: Critical levels for ozone effects on vegetation
603 in Europe, *Environ. Pollut.*, 97, 91-106, [https://doi.org/10.1016/S0269-](https://doi.org/10.1016/S0269-7491(97)00067-5)
604 [7491\(97\)00067-5](https://doi.org/10.1016/S0269-7491(97)00067-5), 1997.
- 605 Gong, C., Lei, Y., Ma, Y., Yue, X., and Liao, H.: Ozone–vegetation feedback through

606 dry deposition and isoprene emissions in a global chemistry–carbon–climate
607 model, *Atmos. Chem. Phys.*, 20, 3841–3857, [https://doi.org/10.5194/acp-20-](https://doi.org/10.5194/acp-20-3841-2020)
608 3841-2020, 2020.

609 Gregg, J. W., Jones, C. G., and Dawson, T. E.: Physiological and Developmental Effects
610 of O₃ on Cottonwood Growth in Urban and Rural Sites, *Ecol. Appl.*, 16, 2368–
611 2381, [https://doi.org/10.1890/1051-0761\(2006\)016\[2368:PADEOO\]2.0.CO;2](https://doi.org/10.1890/1051-0761(2006)016[2368:PADEOO]2.0.CO;2),
612 2006.

613 Hoesly, R. M., Smith, S. J., Feng, L., Klimont, Z., Janssens-Maenhout, G., Pitkanen, T.,
614 Seibert, J. J., Vu, L., Andres, R. J., Bolt, R. M., Bond, T. C., Dawidowski, L.,
615 Kholod, N., Kurokawa, J., Li, M., Liu, L., Lu, Z., Moura, M. C. P., O'Rourke, P.
616 R., and Zhang, Q.: Historical (1750–2014) anthropogenic emissions of reactive
617 gases and aerosols from the Community Emissions Data System (CEDS), *Geosci.*
618 *Model. Dev.*, 11, 369–408, <https://doi.org/10.5194/gmd-11-369-2018>, 2018.

619 Huffman, G. J., Adler, R. F., Arkin, P., Chang, A., Ferraro, R., Gruber, A., Janowiak, J.,
620 McNab, A., Rudolf, B., and Schneider, U.: The Global Precipitation Climatology
621 Project (GPCP) Combined Precipitation Dataset, *B. Am. Meteorol. Soc.*, 78, 5–20,
622 [https://doi.org/10.1175/1520-0477\(1997\)078<0005:TGPCPG>2.0.CO;2](https://doi.org/10.1175/1520-0477(1997)078<0005:TGPCPG>2.0.CO;2), 1997.

623 Hurtt, G. C., Chini, L., Sahajpal, R., Frohling, S., Bodirsky, B. L., Calvin, K., Doelman,
624 J. C., Fisk, J., Fujimori, S., Klein Goldewijk, K., Hasegawa, T., Havlik, P.,
625 Heinemann, A., Humpenöder, F., Jungclaus, J., Kaplan, J. O., Kennedy, J., Krisztin,
626 T., Lawrence, D., Lawrence, P., Ma, L., Mertz, O., Pongratz, J., Popp, A., Poulter,
627 B., Riahi, K., Shevliakova, E., Stehfest, E., Thornton, P., Tubiello, F. N., van
628 Vuuren, D. P., and Zhang, X.: Harmonization of global land use change and
629 management for the period 850–2100 (LUH2) for CMIP6, *Geosci. Model. Dev.*,
630 13, 5425–5464, <https://doi.org/10.5194/gmd-13-5425-2020>, 2020.

631 Ito, G., Romanou, A., Kiang, N.Y., Faluvegi, G., Aleinov, I., Ruedy, R., Russell, G.,
632 Lerner, P., Kelley, M. and Lo, K.: Global carbon cycle and climate feedbacks in
633 the NASA GISS ModelE2, *J. Adv. Model. Earth. Sy.*, 12, [https://doi.org/](https://doi.org/10.1029/2019MS002030)
634 10.1029/2019MS002030, 2020.

635 Jin, Z., Yan, D., Zhang, Z., Li, M., Wang, T., Huang, X., Xie, M., Li, S., and Zhuang,
636 B.: Effects of Elevated Ozone Exposure on Regional Meteorology and Air Quality
637 in China Through Ozone-Vegetation Coupling, *J. Geophys. Res-Atmos.*, 128,
638 e2022JD038119, <https://doi.org/10.1029/2022JD038119>, 2023.

639 Jolivet, Y., Bagard, M., Cabané, M., Vaultier, M.-N., Gandin, A., Afif, D., Dizengremel,
640 P., and Le Thiec, D.: Deciphering the ozone-induced changes in cellular processes:
641 a prerequisite for ozone risk assessment at the tree and forest levels, *Ann. For. Sci.*,
642 73, 923–943, <https://doi.org/10.1007/s13595-016-0580-3>, 2016.

643 Jung, M., Reichstein, M., Margolis, H.A., Cescatti, A., Richardson, A.D., Arain, M.A.,

- 644 Arneth, A., Bernhofer, C., Bonal, D., Chen, J. and Gianelle, D.: Global patterns of
645 land-atmosphere fluxes of carbon dioxide, latent heat, and sensible heat derived
646 from eddy covariance, satellite, and meteorological observations. *J. Geophys. Res-*
647 *Biogeosci.*, 116(G3), <https://doi.org/10.1029/2010JG001566>, 2011.
- 648 Kalnay, E., Kanamitsu, M., Kistler, R., Collins, W., Deaven, D., Gandin, L., Iredell, M.,
649 Saha, S., White, G., Woollen, J., Zhu, Y., Chelliah, M., Ebisuzaki, W., Higgins, W.,
650 Janowiak, J., Mo, K. C., Ropelewski, C., Wang, J., Leetmaa, A., Reynolds, R.,
651 Jenne, R., and Joseph, D.: The NCEP/NCAR 40-Year Reanalysis Project, *B. Am.*
652 *Meteorol. Soc.*, 77, 437–472, <https://doi.org/10.1175/1520->
653 [0477\(1996\)077<0437:TNYRP>2.0.CO;2](https://doi.org/10.1175/1520-0477(1996)077<0437:TNYRP>2.0.CO;2), 1996.
- 654 Karlsson, P., Uddling, J., Braun, S., Broadmeadow, M., Elvira, S., Gimeno, B., Le Thiec,
655 D., Oksanen, E., Vandermeiren, K., Wilkinson, M., and Emberson, L.: New critical
656 levels for ozone effects on young trees based on AOT40 and simulated cumulative
657 leaf uptake of ozone, *Atmos. Environ.*, 38, 2283–2294,
658 <https://doi.org/10.1016/j.atmosenv.2004.01.027>, 2004.
- 659 Koch, D., Schmidt, G. A., and Field, C. V.: Sulfur, sea salt, and radionuclide aerosols
660 in GISS ModelE, *J. Geophys. Res-Atmos.*, 111,
661 <https://doi.org/10.1029/2004JD005550>, 2006.
- 662 Laisk, A., Kull, O., & Moldau, H.: Ozone concentration in leaf intercellular air spaces
663 is close to zero. *Plant Physiol.*, 90(3), 1163–1167,
664 <https://doi.org/10.1104/pp.90.3.1163>, 1989.
- 665 Lam, J. C. Y., Tai, A. P. K., Ducker, J. A., and Holmes, C. D.: Development of an
666 ecophysiology module in the GEOS-Chem chemical transport model version
667 12.2.0 to represent biosphere–atmosphere fluxes relevant for ozone air quality,
668 *Geosci. Model. Dev.*, 16, 2323–2342, <https://doi.org/10.5194/gmd-16-2323-2023>,
669 2023.
- 670 Lei, Y., Yue, X., Liao, H., Gong, C., and Zhang, L.: Implementation of Yale Interactive
671 terrestrial Biosphere model v1.0 into GEOS-Chem v12.0.0: a tool for biosphere–
672 chemistry interactions, *Geosci. Model. Dev.*, 13, 1137–1153,
673 <https://doi.org/10.5194/gmd-13-1137-2020>, 2020.
- 674 Lei, Y., Yue, X., Liao, H., Zhang, L., Yang, Y., Zhou, H., Tian, C., Gong, C., Ma, Y., and
675 Gao, L.: Indirect contributions of global fires to surface ozone through ozone–
676 vegetation feedback, *Atmos. Chem. Phys.*, 21, 11531–11543,
677 <https://doi.org/10.5194/acp-21-11531-2021>, 2021.
- 678 Li, Y., Henze, D.K., Jack, D.: The influence of air quality model resolution on health
679 impact assessment for fine particulate matter and its components, *Air. Qual. Atmos.*
680 *Hlth.*, 9, 51–68, <https://doi.org/10.1007/s11869-015-0321-z>, 2016.

- 681 Lin, M., Horowitz, L.W., Xie, Y., Paulot, F., Malyshev, S., Shevliakova, E., Finco, A.,
682 Gerosa, G., Kubistin, D. and Pilegaard, K.: Vegetation feedbacks during drought
683 exacerbate ozone air pollution extremes in Europe, *Nat. Clim. Change.*, 10,444-
684 451, <https://doi.org/10.1038/s41558-020-0743-y>, 2020.
- 685 Lombardozi, D., Levis, S., Bonan, G., and Sparks, J. P.: Predicting photosynthesis and
686 transpiration responses to ozone: decoupling modeled photosynthesis and stomatal
687 conductance, *Biogeosciences*, 9, 3113–3130, [https://doi.org/10.5194/bg-9-3113-](https://doi.org/10.5194/bg-9-3113-2012)
688 2012, 2012.
- 689 Lombardozi, D., Sparks, J. P., and Bonan, G.: Integrating O₃ influences on terrestrial
690 processes: photosynthetic and stomatal response data available for regional and
691 global modeling, *Biogeosciences*, 10, 6815–6831, [https://doi.org/10.5194/bg-10-](https://doi.org/10.5194/bg-10-6815-2013)
692 6815-2013, 2013.
- 693 Lombardozi, D., Levis, S., Bonan, G., Hess, P. G., and Sparks, J. P.: The Influence of
694 Chronic Ozone Exposure on Global Carbon and Water Cycles, *J. Climate.*, 28,
695 292–305, <https://doi.org/10.1175/JCLI-D-14-00223.1>, 2015.
- 696 van Marle, M. J. E., Kloster, S., Magi, B. I., Marlon, J. R., Daniiau, A.-L., Field, R. D.,
697 Arneeth, A., Forrest, M., Hantson, S., Kehrwald, N. M., Knorr, W., Lasslop, G., Li,
698 F., Mangeon, S., Yue, C., Kaiser, J. W., and van der Werf, G. R.: Historic global
699 biomass burning emissions for CMIP6 (BB4CMIP) based on merging satellite
700 observations with proxies and fire models (1750–2015), *Geosci. Model. Dev.*, 10,
701 3329–3357, <https://doi.org/10.5194/gmd-10-3329-2017>, 2017.
- 702 Menon, S. and Rotstayn, L.: The radiative influence of aerosol effects on liquid-phase
703 cumulus and stratiform clouds based on sensitivity studies with two climate
704 models, *Clim. Dyn.*, 27, 345–356, <https://doi.org/10.1007/s00382-006-0139-3>,
705 2006.
- 706 Mills, G., Buse, A., Gimeno, B., Bermejo, V., Holland, M., Emberson, L., and Pleijel,
707 H.: A synthesis of AOT40-based response functions and critical levels of ozone
708 for agricultural and horticultural crops, *Atmos. Environ.*, 41, 2630–2643,
709 <https://doi.org/10.1016/j.atmosenv.2006.11.016> , 2007.
- 710 Mills, G., Harmens, H., Wagg, S., Sharps, K., Hayes, F., Fowler, D., Sutton, M. and
711 Davies, B.: Ozone impacts on vegetation in a nitrogen enriched and changing
712 climate, *Environ. Pollut.*, 208, 898-908,
713 <https://doi.org/10.1016/j.envpol.2015.09.038>, 2016.
- 714 Myhre, G., Shindell, D., Breion, F.-M., Collins, W., Fuglestedt, J., Huang, J., Koch,
715 D., Lamarque, J.-F., Lee, D., Mendoza, B., Nakajima, T., Robock, A., Stephens,
716 G., Takemura, T., and Zhang, H., Anthropogenic and Natural Radiative Forcing,
717 in: *Climate Change 2013: The Physical Science Basis. Contribution of Working*
718 *Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate*

- 719 Change, edited by: Stocker, T. F., Qin, D., Plattner, G.-K., Tignor, M., Allen, S. K.,
720 Boschung, J., Nauels, A., Xia, Y., Bex, V., and Midgley, P. M., Cambridge
721 University Press, Cambridge, UK and New York, NY, USA, 2013.
- 722 Norval, M., Lucas, R. M., Cullen, A. P., De Gruijl, F. R., Longstreth, J., Takizawa, Y.,
723 and Van Der Leun, J. C.: The human health effects of ozone depletion and
724 interactions with climate change, *Photoch. Photobio. Sci.*, 10, 199–225,
725 <https://doi.org/10.1039/C0PP90044C>, 2011.
- 726 Nussbaum, S. and Fuhrer, J.: Difference in ozone uptake in grassland species between
727 open-top chambers and ambient air, *Environ. Pollut.*, 109, 463–471,
728 [https://doi.org/10.1016/S0269-7491\(00\)00049-X](https://doi.org/10.1016/S0269-7491(00)00049-X), 2000.
- 729 Nuvolone, D., Petri, D., and Voller, F.: The effects of ozone on human health, *Environ.*
730 *Sci. Pollut. R.*, 25, 8074–8088, <https://doi.org/10.1007/s11356-017-9239-3>, 2018.
- 731 Ohara, T., Akimoto, H., Kurokawa, J., Horii, N., Yamaji, K., Yan, X., and Hayasaka, T.:
732 An Asian emission inventory of anthropogenic emission sources for the period
733 1980–2020, *Atmos. Chem. Phys.*, 7, 4419–4444, [https://doi.org/10.5194/acp-7-](https://doi.org/10.5194/acp-7-4419-2007)
734 [4419-2007](https://doi.org/10.5194/acp-7-4419-2007), 2007.
- 735 Oleson, K. W., Lawrence, D. M., Bonan, G. B., Flanne, M. G., Kluzek, E., Lawrence,
736 P. J., Levis, S., Swenson, S. C., and Thornton, P. E.: Technical Description of
737 version 4.0 of the Community Land Model (CLM), National Center for
738 Atmospheric Research, Boulder, USA, CONCAR/TN-478+STR, 2010.
- 739 Oliver, R. J., Mercado, L. M., Sitch, S., Simpson, D., Medlyn, B. E., Lin, Y.-S., and
740 Folberth, G. A.: Large but decreasing effect of ozone on the European carbon sink,
741 *Biogeosciences*, 15, 4245–4269, 2018.
- 742 Petters, M. D. and Kreidenweis, S. M.: A single parameter representation of
743 hygroscopic growth and cloud condensation nucleus activity, *Atmos. Chem. Phys.*,
744 7, 1961–1971, <https://doi.org/10.5194/acp-7-1961-2007>, 2007.
- 745 Paoletti, E., De Marco, A. and Rcalbutto, S.: Why should we calculate complex indices
746 of ozone exposure? Results from Mediterranean background sites. *Environ. Monit.*
747 *Assess.*, 128, pp.19-30, <https://doi.org/10.1007/s10661-006-9412-5>, 2007.
- 748 Pitchford, M., Malm, W., Schichtel, B., Kumar, N., Lowenthal, D., and Hand, J.:
749 Revised Algorithm for Estimating Light Extinction from IMPROVE Particle
750 Speciation Data, *Japca. J. Air. Waste. Ma.*, 57 (11), 1326–1336,
751 <https://doi.org/10.3155/1047-3289.57.11.1326>, 2007.
- 752 Pleijel, H., Danielsson, H., Ojanperä, K., Temmerman, L. D., Högy, P., Badiani, M.,
753 and Karlsson, P. E.: Relationships between ozone exposure and yield loss in
754 European wheat and potato—a comparison of concentration- and flux-based
755 exposure indices, *Atmos. Environ.*, 38, 2259–2269,

- 756 <https://doi.org/10.1016/j.atmosenv.2003.09.076>, 2004.
- 757 Pleijel, H., Danielsson, H., Emberson, L., Ashmore, M. R., and Mills, G.: Ozone risk
758 assessment for agricultural crops in Europe: further development of stomatal flux
759 and flux–response relationships for European wheat and potato, *Atmos. Environ.*,
760 41, 3022–3040, <https://doi.org/10.1016/j.atmosenv.2006.12.002>, 2007.
- 761 Remer, L. A., Kaufman, Y. J., Tanré, D., Mattoo, S., Chu, D. A., Martins, J. V., Li, R.-
762 R., Ichoku, C., Levy, R. C., and Kleidman, R. G.: The MODIS aerosol algorithm,
763 products, and validation, *J. Atmos. Sci.*, 62, 947–973,
764 <https://doi.org/10.1175/JAS3385.1>, 2005.
- 765 Sadiq, M., Tai, A. P., Lombardozzi, D., and Val Martin, M.: Effects of ozone–vegetation
766 coupling on surface ozone air quality via biogeochemical and meteorological
767 feedbacks, *Atmos. Chem. Phys.*, 17, 3055–3066, <https://doi.org/10.5194/acp-17-3055-2017>, 2017.
- 769 Schmidt, G. A., Ruedy, R., Hansen, J. E., Aleinov, I., Bell, N., Bauer, M., Bauer, S.,
770 Cairns, B., Canuto, V., Cheng, Y., Genio, A. D., Faluvegi, G., Friend, A. D., Hall,
771 T. M., Hu, Y., Kelley, M., Kiang, N. Y., Koch, D., Lacis, A. A., Lerner, J., Lo, K.
772 K., Miller, R. L., Nazarenko, L., Oinas, V., Perlwitz, J., Perlwitz, J., Rind, D.,
773 Romanou, A., Russell, G. L., Sato, M., Shindell, D. T., Stone, P. H., Sun, S.,
774 Tausnev, N., Thresher, D., and Yao, M.-S.: Present-Day Atmospheric Simulations
775 Using GISS ModelE: Comparison to In Situ, Satellite, and Reanalysis Data, *J.*
776 *Climate.*, 19, 153–192, <https://doi.org/10.1175/JCLI3612.1>, 2006.
- 777 Schmidt, G. A., Kelley, M., Nazarenko, L., Ruedy, R., Russell, G. L., Aleinov, I., Bauer,
778 M., Bauer, S. E., Bhat, M. K., Bleck, R., Canuto, V., Chen, Y.-H., Cheng, Y., Clune,
779 T. L., Del Genio, A., de Fainchtein, R., Faluvegi, G., Hansen, J. E., Healy, R. J.,
780 Kiang, N. Y., Koch, D., Lacis, A. A., LeGrande, A. N., Lerner, J., Lo, K. K.,
781 Matthews, E. E., Menon, S., Miller, R. L., Oinas, V., Olosó, A. O., Perlwitz, J. P.,
782 Puma, M. J., Putman, W. M., Rind, D., Romanou, A., Sato, M., Shindell, D. T.,
783 Sun, S., Syed, R. A., Tausnev, N., Tsigaridis, K., Unger, N., Voulgarakis, A., Yao,
784 M.-S., and Zhang, J.: Configuration and assessment of the GISS ModelE2
785 contributions to the CMIP5 archive: GISS MODEL-E2 CMIP5 SIMULATIONS,
786 *J. Adv. Model. Earth Syst.*, 6, 141–184, <https://doi.org/10.1002/2013MS000265>,
787 2014.
- 788 Sicard, P., De Marco, A., Dalstein-Richier, L., Tagliaferro, F., Renou, C., Paoletti, Elena,
789 2016. An epidemiological assessment of stomatal ozone flux-based critical levels
790 for visible ozone injury in southern European forests. *Sci. Total. Environ.*, 541,
791 729-741.
- 792 Sitch, S., Cox, P. M., Collins, W. J., and Huntingford, C.: Indirect radiative forcing of
793 climate change through ozone effects on the land-carbon sink, *Nature*, 448, 791–
794 794, <https://doi.org/10.1038/nature06059>, 2007.

- 795 Sofen, E. D., Bowdalo, D., Evans, M. J., Apadula, F., Bonasoni, P., Cupeiro, M., Ellul,
796 R., Galbally, I. E., Girgzdiene, R., Luppó, S., Mimouni, M., Nahas, A. C., Saliba,
797 M., and Tørseth, K.: Gridded global surface ozone metrics for atmospheric
798 chemistry model evaluation, *Earth Syst. Sci. Data*, 8, 41–59,
799 <https://doi.org/10.5194/essd-8-41-2016>, 2016.
- 800 Unger, N., Zheng, Y., Yue, X., and Harper, K. L.: Mitigation of ozone damage to the
801 world’s land ecosystems by source sector, *Nat. Clim. Chang.*, 10, 134–137,
802 <https://doi.org/10.1038/s41558-019-0678-3>, 2020.
- 803 VanLoocke, A., Betzelberger, A. M., Ainsworth, E. A., and Bernacchi, C. J.: Rising
804 ozone concentrations decrease soybean evapotranspiration and water use
805 efficiency whilst increasing canopy temperature, *New. Phytol.*, 195, 164–171,
806 <https://doi.org/10.1111/j.1469-8137.2012.04152.x>, 2012.
- 807 van der Werf, G. R., Randerson, J. T., Giglio, L., van Leeuwen, T. T., Chen, Y., Rogers,
808 B. M., Mu, M., van Marle, M. J. E., Morton, D. C., Collatz, G. J., Yokelson, R. J.,
809 and Kasibhatla, P. S.: Global fire emissions estimates during 1997–2016, *Earth*
810 *Syst. Sci. Data*, 9, 697–720, <https://doi.org/10.5194/essd-9-697-2017>, 2017.
- 811 Wang, Y., Shen, L., Wu, S., Mickley, L., He, J. and Hao, J.: Sensitivity of surface ozone
812 over China to 2000–2050 global changes of climate and emissions, *Atmos.*
813 *Environ.*, 75, 374–382, <https://doi.org/10.1016/j.atmosenv.2013.04.045>, 2013.
- 814 Wesely, M. L. and Hicks, B. B.: A review of the current status of knowledge on dry
815 deposition, *Atmos. Environ.*, 34, 2261–2282, [https://doi.org/10.1016/S1352-2310\(99\)00467-7](https://doi.org/10.1016/S1352-2310(99)00467-7), 2000.
- 817 Wild, M., Folini, D., Schär, C., Loeb, N., Dutton, E. G., and König-Langlo, G.: The
818 global energy balance from a surface perspective, *Clim. Dyn.*, 40, 3107–3134,
819 <https://doi.org/10.1007/s00382-012-1569-8>, 2013.
- 820 Yue, X., Liao, H., Wang, H. J., Li, S. L., and Tang, J. P.: Role of sea surface temperature
821 responses in simulation of the climatic effect of mineral dust aerosol, *Atmos.*
822 *Chem. Phys.*, 11, 6049–6062, <https://doi.org/10.5194/acp-11-6049-2011>, 2011.
- 823 Yue, X. and Unger, N.: Ozone vegetation damage effects on gross primary productivity
824 in the United States, *Atmos. Chem. Phys.*, 14, 9137–9153,
825 <https://doi.org/10.5194/acp-14-9137-2014>, 2014.
- 826 Yue, X. and Unger, N.: The Yale Interactive terrestrial Biosphere model version 1.0:
827 description, evaluation and implementation into NASA GISS ModelE2, *Geosci.*
828 *Model Dev.*, 8, 2399–2417, <https://doi.org/10.5194/gmd-8-2399-2015>, 2015.
- 829 Yue, X. and Unger, N.: Fire air pollution reduces global terrestrial productivity, *Nat.*
830 *Commun.*, 9, 5413, <https://doi.org/10.1038/s41467-018-07921-4>, 2018.

- 831 Yue, X., Keenan, T. F., Munger, W., and Unger, N.: Limited effect of ozone reductions
832 on the 20-year photosynthesis trend at Harvard forest, *Glob. Change Biol.*, 22,
833 3750–3759, <https://doi.org/10.1111/gcb.13300>, 2016.
- 834 Yue, X., Liao, H., Wang, H., Zhang, T., Unger, N., Sitch, S., Feng, Z., and Yang, J.:
835 Pathway dependence of ecosystem responses in China to 1.5°C global warming,
836 *Atmos. Chem. Phys.*, 20, 2353–2366, <https://doi.org/10.5194/acp-20-2353-2020>,
837 2020.
- 838 Zhang, L., Vet, R., Brook, J. R., and Legge, A. H.: Factors affecting stomatal uptake of
839 ozone by different canopies and a comparison between dose and exposure, *Sci.*
840 *Total Environ.*, 370, 117–132, <https://doi.org/10.1016/j.scitotenv.2006.06.004>,
841 2006.
- 842 Zhou, X., Yue, X., and Tian, C.: Responses of Ecosystem Productivity to Anthropogenic
843 Ozone and Aerosols at the 2060, *Earths Future*, 12, e2023EF003781,
844 <https://doi.org/10.1029/2023EF003781>, 2024.
- 845 Zhu, J., Tai, A. P. K., and Hung Lam Yim, S.: Effects of ozone–vegetation interactions
846 on meteorology and air quality in China using a two-way coupled land–
847 atmosphere model, *Atmos. Chem. Phys.*, 22, 765–782,
848 <https://doi.org/10.5194/acp-22-765-2022>, 2022.
- 849