When and why microbial-explicit soil organic carbon models can be unstable

Erik Schwarz[1], Samia Ghersheen[2], Salim Belyazid[1], and Stefano Manzoni[1]

[1]Department of Physical Geography and Bolin Centre for Climate Research, Stockholm University, Stockholm, Sweden
[2]Department of Soil and Environment, Swedish University of Agricultural Sciences, Uppsala, Sweden

Correspondence: Erik Schwarz (erik.schwarz@natgeo.su.se)

Abstract. Microbial-explicit soil organic carbon (SOC) cycling models are increasingly recognized for their advantages over linear models in describing SOC dynamics. These models are known to exhibit oscillations, but it is not clear when they yield stable vs. unstable equilibrium points (EPs) – i.e. EPs that exist analytically, but are not stable to small perturbations and cannot be reached by transient simulations. Occurrence of such unstable EPs can lead to unexpected model behaviour in transient simulations or unrealistic predictions of steady state soil organic carbon (SOC) stocks. Here we ask when and why unstable EPs can occur in an archetypal microbial-explicit model (representing SOC, dissolved OC [DOC], microbial biomass, and extracellular enzymes) and some simplified versions of it. Further, if a model formulation allows for physically meaningful but unstable EPs, can we find constraints in the model parameters (i.e. environmental conditions and microbial traits) that ensure stability of the EPs? We use analytical, numerical and descriptive tools to answer these questions. We found that instability can occur when the resupply of a growth substrate (DOC) is (via a positive feedback loop) dependent on its abundance. We identified a conservative, sufficient condition on model parameters to ensure stability of EPs. Principally, three distinct strategies can avoid instability: 1) negligence of explicit DOC dynamics, 2) biomass independent uptake rate or 3) correlation between parameter values to obey the stability criterion. While the first two approaches simplify some mechanistic processes, the third approach points to the interactive effects of environmental conditions and parameters describing microbial physiology highlighting the relevance of basic ecological principles for avoidance of unrealistic (i.e. unstable) simulation outcomes. These insights can help to improve applicability of microbial-explicit models, aid our understanding of the dynamics of these models and highlight the relation between mathematical requirements and (*in silico*) microbial ecology.

1 Introduction

Current Earth system models (ESMs) have very simplified representations of soil organic carbon (SOC) dynamics (Bradford et al., 2016; Todd-Brown et al., 2013; Varney et al., 2022). Accuracy in matching observed SOC stocks and turnover times has not significantly improved in the latest ensemble of ESMs used in the Coupled Model Intercomparison Project (CMIP, CMIP 6) (Varney et al., 2022) – with uncertainty about SOC responses to climate change remaining high (Todd-Brown et al., 2013; Varney et al., 2022). Consequently, a need to improve and diversify description of SOC dynamics in ESMs has been identified (Bradford et al., 2016; Todd-Brown et al., 2013; Varney et al., 2022; Wieder et al., 2015, 2018). Current ESMs employ linear

25 degradation kinetics to simulate SOC degradation (Todd-Brown et al., 2013), missing to integrate our current understanding of major controls on SOC fate and to acknowledge the uncertainties in describing these processes (Abs et al., 2023; Bradford et al., 2016; Wieder et al., 2015, 2018). Non-linear, microbial-explicit SOC models can improve model-data agreement (Hararuk et al., 2015; Wieder et al., 2013). These models vary in number and identity of C pools and degree of non-linearity (e.g. Allison et al., 2010; Manzoni and Porporato, 2009; Schimel and Weintraub, 2003; Wang et al., 2013, 2015; Wieder et al., 2014, 2015).

30 Among these models, the AWB (Allison-Wallenstein-Bradford) model (Allison et al., 2010) has emerged as an archetypal model structure to study the influence of soil microbial processes on carbon stocks (e.g. Abs et al., 2022; Calabrese et al., 2022; Georgiou et al., 2017; Hararuk et al., 2015; Tao et al., 2023; Wieder et al., 2015). The AWB model explicitly represents pools of microbial biomass, extracellular enzymes produced by microbes, polymeric SOC that is not available for microbial uptake, and a pool of available dissolved organic carbon (DOC) produced from enzymatic depolymerization of SOC (Allison

35 et al., 2010, Fig. 1a). With only four C pools and commonly two non-linear terms, the AWB model retains a comparably simple structure and remains somewhat analytically tractable.

While better at predicting modern day SOC stocks, microbial-explicit SOC models are known to exhibit oscillatory behaviour (e.g. Georgiou et al., 2017; Manzoni and Porporato, 2007; Sierra and Müller, 2015; Wang et al., 2014, 2016). Such oscillations can represent carbon-microbe dynamics observed at small spatial scales (see e.g. discussion in Manzoni and Porporato, 2007) but are unfavourable for application at larger spatial and temporal scales, where such oscillations are generally not

40 observed (Georgiou et al., 2017; Wang et al., 2014). A few studies have analysed the oscillatory properties of some microbial-explicit SOC models (Georgiou et al., 2017; Manzoni and Porporato, 2007; Wang et al., 2014, 2016). These studies characterized the dynamics exhibited after a perturbation around a model's equilibrium (i.e., when all state variables are at steady state): does a model directly converge back to its previous equilibrium or does it approach the equilibrium with dampened oscillations?

45 Different degrees of oscillatory behaviour have been described, but generally these models were found to be stable (that is, they do converge back to their previous equilibrium) for given parametrizations or if they follow basic principles such as mass conservation and dependence of fluxes on source pools (Sierra and Müller, 2015; Wang et al., 2014, 2016). Stable oscillatory behaviour, however, is only one of the possible dynamics such non-linear models can exhibit. In fact, these models can also be unstable (that is after perturbation a model does not converge back to its previous equilibrium) (Abs et al., 2022; Raupach,

50 2007; Schimel and Weintraub, 2003; Sierra and Müller, 2015), but the occurrence of unstable equilibria in microbial-explicit SOC models remains largely unexplored. While unstable equilibrium points exist analytically, they can never be reached by transient simulations. Thus, model parameterizations that yield unstable equilibria can lead to unpredictable simulation outcomes as amplifying oscillations can occur, expected equilibrium states are not reached (because they are unstable) hindering convergence in model spin-up, or (some) state variables might collapse (e.g. Fig. 1b, yellow line). Further, if C stocks are

55 predicted based on analytical steady state solutions, unstable equilibria might lead to unrealistic predictions, mismatching outcomes from dynamic simulations. To increase reliability of model predictions and model applicability, it is important to understand when and for what reasons microbial-explicit SOC models become unstable.

Here, we study an archetypal microbial-explicit SOC model (based on the AWB model, Allison et al., 2010, and some simplified versions of it) to answer the questions: 1) What mechanisms in microbial-explicit SOC models (model structures,

used kinetic formulations, and parameter values) cause unstable equilibrium points to emerge? and 2) how can we select model structures and/or constrain model parameters to ensure stability of equilibrium points?

2 Methods

2.1 Archetypal microbial-explicit SOC model

We start by defining the C mass balance equations for a system encompassing SOC (S), dissolved organic C (DOC, D), microbial biomass C (MBC, B), and extracellular enzyme C (ENZ, E) (eq. 1 - 4). The C compartments and flows are illustrated in Fig. 1a and symbols for the variables and fluxes are defined in Table 1 and 3. The C mass balance equations are written as a system of ordinary differential equations (ODE), where for convenience the fluxes are aligned vertically according to their meaning (from left to right: external inputs, depolymerization, uptake and metabolism, decay, and finally abiotic losses)

$$\frac{\mathrm{d}S}{\mathrm{d}t} = \quad f_I I - P \qquad\qquad\qquad\qquad +f_D r_B D_B \qquad\qquad -L_S \qquad (1)$$

$$\frac{\mathrm{d}D}{\mathrm{d}t} = \quad (1-f_I)I + P \qquad\qquad -U \qquad +(1-f_D)r_B D_B + D_E \qquad -L_D \qquad (2)$$

$$\frac{\mathrm{d}B}{\mathrm{d}t} = \qquad\qquad\qquad y_B U - R_E \qquad\qquad -D_B \qquad\qquad\qquad (3)$$

$$\frac{\mathrm{d}E}{\mathrm{d}t} = \qquad\qquad (y_m - y_B)U + R_E \qquad\qquad -D_E \qquad -L_E \quad . \qquad (4)$$

Organic matter enters the system with flux I that is partitioned between SOC and DOC depending on the fraction f_I. SOC is depolymerized by extracellular enzymes at rate P and transferred to the DOC pool. DOC is directly available for microbial uptake at rate U. Both P and U are non-linear functions that can take on various forms (Table 2).

Microbes assimilate the substrate with a maximal efficiency $y_m \leq 1$ that is limited by physiological and/or thermodynamic constraints (Chakrawal et al., 2022) and use the substrate either for growth (i.e. biomass production at rate $y_B U$) or to produce extracellular enzymes (at rate $(y_m - y_B)U$). We denote this uptake-dependent pathway of extracellular enzyme production as "inducible" ENZ production. An alternative or complementary mode of ENZ production is the biomass-dependent "constitutive" ENZ production at rate R_E given by

$$R_E = m_E B \quad , \qquad\qquad\qquad\qquad\qquad\qquad (5)$$

where m_E is the rate constant of constitutive ENZ production. In both formulations, we assumed that respiratory costs associated with enzyme production are already included in the growth respiration (proportional to $1 - y_m$). Two limiting cases can be derived from this general description of extracellular enzyme production:

Only constitutive ENZ production: $\qquad\qquad\qquad\qquad\qquad y_B = y_m \qquad\qquad (6)$

Only inducible ENZ production: $\qquad\qquad\qquad\qquad\qquad m_E = 0 \quad . \qquad\qquad (7)$

Both MBC and ENZ are assumed to decay with a linear decay rate D_i

$$D_i = d_i i \quad , (i = B, E), \qquad\qquad\qquad\qquad\qquad\qquad (8)$$

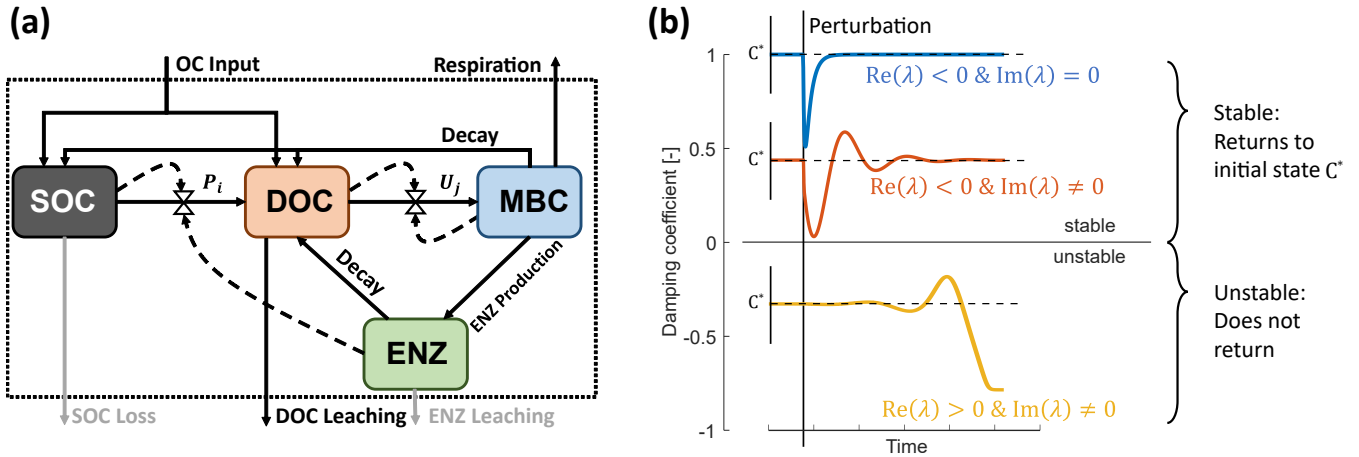



Figure 1. Model schematics of the archetypal microbial-explicit SOC model (a) and its relevant stability behaviours (b). Colored boxes in (a) indicate the state variables soil organic carbon (SOC), dissolved organic carbon (DOC), microbial biomass carbon (MBC), and extracellular enzymes (ENZ). Solid arrows indicate carbon fluxes and dashed arrows connected to valve symbols indicate controls over the non-linear kinetic. Grey arrows indicate processes neglected in some analyses. The dotted box delineates the system's boundary. Colored lines in (b) illustrate the dynamics of a state variable relative to its steady state value (C^*) following a perturbation for a stable node (damping coefficient = 1; blue line), a stable focus (0 <damping coefficient< 1; red line), and an unstable focus (−1 <damping coefficient< 0; yellow line). The subplot axes are centered around the value of their respective damping coefficient.

but we also consider density-dependent microbial decay as an alternative to the linear kinetic ($D'_B = d'_B B^b$ with $1 < b \leq 2$; Georgiou et al., 2017). All decayed ENZ are assumed to return to the DOC pool while only a fraction r_B of the decayed microbial biomass is recycled in the system and partitioned between SOC and DOC according to the factor f_D. In turn, $(1 - r_B)D_B$ represents linear microbial maintenance respiration. SOC, DOC and ENZ can have abiotic losses L_i (e.g. erosion, leaching, ...)

$$L_i = l_i i \quad , (i = S, D, E) \quad . \tag{9}$$

The system of eq. 1 - 4 constitutes a model of SOC cycling of varying complexity depending on the chosen kinetics. We refer to this four-pool model version as the $SDBE$ model, according to the represented state variables. We use this system as a starting point for our analysis, but reduce it to simpler variants to derive specific analytical results (Sect. 2.2).

Commonly, depolymerization of SOC by extracellular enzymes is described by either multiplicative (m), forward Michaelis-Menten (f), reverse Michaelis-Menten (r) (Schimel and Weintraub, 2003) or equilibrium chemistry approximation (ECA, e) (Tang and Riley, 2013) kinetics. Table 2 lists the respective formulations of P_i ($i = m, f, r, e$), where v_i^p is the maximal depolymerization rate coefficient and K_i^p the respective half-saturation constant (if applicable). The uptake of DOC by microbes can be described with similar formulations U_j ($j = m, f$), simply by replacing S for D and E for B (Table 2), where v_j^u is the maximal uptake rate coefficient and K_f^u the respective half-saturation constant.

Table 1. Description of all state variables and fluxes.

Symbol	Description	Unit
State Variables		
S	Soil organic carbon (SOC)	mg C g^{-1}
D	Dissolved organic carbon (DOC)	mg C g^{-1}
B	Microbial biomass (MBC)	mg C g^{-1}
E	Extracellular enzymes (ENZ)	mg C g^{-1}
Fluxes		
I	Organic carbon input	mg C g^{-1} d^{-1}
P	Depolimerization of SOC	mg C g^{-1} d^{-1}
U	Microbial uptake of DOC	mg C g^{-1} d^{-1}
L_S	Loss of SOC	mg C g^{-1} d^{-1}
L_D	Leaching of DOC	mg C g^{-1} d^{-1}
L_E	Leaching of extracellular enzymes	mg C g^{-1} d^{-1}
D_B	Decay of microbial biomass	mg C g^{-1} d^{-1}
D_E	Decay of extracellular enzymes	mg C g^{-1} d^{-1}
R_E	Constitutive production of extracellular enzymes	mg C g^{-1} d^{-1}

Table 2. Employed non-linear kinetics for the depolymerization rate P_i and uptake rate U_j. m: multiplicative, f: forward Michaelis-Menten, r: reverse Michaelis-Menten (Schimel and Weintraub, 2003), e: equilibrium chemistry approximation (Tang and Riley, 2013). U_r and U_e are not used in our analysis.

Kinetic (i|j)	P_i	U_j
m	$v_m^p SE$	$v_m^u DB$
f	$v_f^p \frac{S}{K_f^p + S} E$	$v_f^u \frac{D}{K_f^u + D} B$
r	$v_r^p S \frac{E}{K_r^p + E}$	–
e	$v_e^p \frac{SE}{K_e^p + S + E}$	–

Many combinations of depolymerization and uptake kinetics are possible. For model versions with both non-linear terms we limit our analysis to only a few combinations of depolymerization and uptake kinetics (indicated by the subscript $i \times j$ for the i-th depolymerization kinetic and j-th uptake kinetic), namely: $m \times m$, $f \times f$, and $r \times f$ (summary of analysed scenarios in Table 4). The first combination employing only multiplicative kinetics facilitates analytical tractability; the second combination is commonly used in other models (e.g. Allison et al., 2010; Georgiou et al., 2017; Tao et al., 2023); the third combination is based on conclusions of Tang and Riley (2019) that $r \times f$ might be an appropriate (and analytically tractable) approximation of ECA kinetics.

To improve analytical tractability of the four-pool model we neglect abiotic losses of SOC and ENZ by setting $L_S = L_E = 0$ and limit the analysis only to the case of constitutive ENZ production ($y_B = y_m$, eq. 6) (Table 4). The Jacobian matrix of partial derivatives for the four-pool model $J_{i \times j}^{SDBE}$ is given in eq. A17. The matrix $J_{i \times j}^{SDBE}$ is expressed in terms of general depolymerization and uptake kinetics P_i and U_j which allows for stability analysis of the $SDBE$ model irrespective of the specific kinetics.

2.2 Reduced models for mathematical analysis

To identify which and how structural elements of the four-pool model with two non-linear kinetics affect model stability, we introduce two reduced model versions:

1. the SBE (SOC-MBC-ENZ) model, neglecting DOC dynamics, and

2. the SDB (SOC-DOC-MBC) model, assuming ENZ to be at quasi-steady state.

Both model versions have only three pools, but are different as in the SBE model only one non-linear term remains, while the SDB model still has both non-linear depolymerization and uptake kinetics. We analyze the former, less non-linear model for all depolymerization kinetics listed in Table 2 with both constitutive and inducible ENZ production pathways and including abiotic losses of S and E (Table 4). In contrast, we analyze the latter, more non-linear model, after applying the same simplifying assumptions as for the four-pool model (Table 4) – that is we set $L_S = L_E = 0$, $y_B = y_m$, and only consider three combinations of depolymerization and uptake kinetics ($m \times m$, $f \times f$, and $r \times f$).

2.2.1 SBE model

DOC dynamics are neglected in the SBE (SOC-MBC-ENZ) model. Instead it is assumed that any organic carbon that is made available by depolymerization of SOC is directly taken up by microbes – that is $U = P$. The flux of decayed extracellular enzymes D_E enters the SOC pool and the partitioning factors f_I and f_D are set to 1. The resulting system of equations is given for the i-th kinetic formulation for P (Table 2) by

$$\frac{\mathrm{d}S}{\mathrm{d}t} = I \qquad\qquad -P_i \qquad\qquad +r_B D_B + D_E \qquad\qquad -L_S \tag{10}$$

$$\frac{\mathrm{d}B}{\mathrm{d}t} = \qquad\qquad y_B P_i - R_E \qquad\qquad -D_B \tag{11}$$

$$\frac{\mathrm{d}E}{\mathrm{d}t} = \qquad\qquad (y_m - y_B)P_i + R_E \qquad\qquad -D_E \qquad\qquad -L_E \quad . \tag{12}$$

Note that in this formulation, unless ENZ production is purely constitutive (eq. 6), ENZ production is (partly) independent of microbial biomass (as P_i only depends on E and not on B). Consequently, as soon as there are extracellular enzymes that catalyze depolymerization, further enzyme production entails. The Jacobian matrix of the SBE model J_i^{SBE} is a 3×3 matrix given by eq. A1.

2.2.2 SDB model

In the SDB (SOC-DOC-MBC) model, the extracellular enzyme pool is assumed to be at quasi-steady state, that is $\frac{\mathrm{d}E}{\mathrm{d}t} = 0$. With $L_E = 0$ and $y_B = y_m$ we obtain the quasi-steady state concentration of E from eq. 4, 5 and 8 as

$$E^{qss} = \frac{m_E}{d_E} B \quad . \tag{13}$$

The SDB model is obtained by substituting E^{qss} for E in eq. 1 - 3, which yields after assuming $L_S = 0$

$$\frac{\mathrm{d}S}{\mathrm{d}t} = \qquad f_I I - P_i^{qss} \qquad\qquad\qquad +f_D r_B D_B \tag{14}$$

$$\frac{\mathrm{d}D}{\mathrm{d}t} = \qquad (1-f_I)I + P_i^{qss} \qquad -U_j \qquad +(1-f_D)r_B D_B + D_E^{qss} \qquad -L_D \tag{15}$$

$$\frac{\mathrm{d}B}{\mathrm{d}t} = \qquad\qquad\qquad y_m U_j - R_E \qquad\qquad -D_B \qquad\qquad . \tag{16}$$

In this model version, two non-linearities remain. Substituting E^{qss} for E in D_E and P_i we obtain $D_E^{qss} = R_E$ and $P_i^{qss} = f(S, B)$, respectively. The Jacobian matrix of the SDB model $J_{i \times j}^{SDB}$ is a 3×3 matrix given by eq. A10.

2.3 Stability analysis

Stability behaviour here refers to how a model responds to a small perturbation around an equilibrium point (illustrated in Fig. 1b). The equilibrium points are obtained by assuming that the system behaviour does not change over time – i.e. by setting the ODEs of all state variables C equal to zero ($\frac{\mathrm{d}C}{\mathrm{d}t} = 0$). This yields their steady states C^*. For non-linear systems, stability is determined by the eigenvalues (λ) of the Jacobian matrix J evaluated at an equilibrium (denoted as $J\big|_*$; $\mathrm{eig}(J\big|_*) = \lambda$) (e.g. Argyris et al., 2015). Briefly, if the real parts of *all* λ are negative ($\mathrm{Re}(\lambda) < 0$), the equilibrium is stable: the system will converge back to this equilibrium after a perturbation. Instead, if *one or more* eigenvalues have positive real parts ($\mathrm{Re}(\lambda) > 0$), the equilibrium is unstable and the system will not return to the same state. If the eigenvalues additionally have non-zero imaginary parts ($\mathrm{Im}(\lambda) \neq 0$), oscillations around the equilibrium occur (Fig. 1b). Stability analysis is described in more details in e.g., Argyris et al. (2015).

2.3.1 Analytical approach

Because the eigenvalues of the Jacobian matrix can be analytically cumbersome even in the comparably compact three-pool models, we instead evaluate the Routh-Hurwitz criterion (e.g. Argyris et al., 2015; Horn and Johnson, 1991) for $J\big|_*$. The Routh-Hurwitz criterion states that all $\mathrm{Re}(\lambda)$ have negative signs, if, and only if:

1. all coefficients a_i of the characteristic polynomial $\det(J\big|_* - \mathbf{1}\lambda) = 0$ (where $\mathbf{1}$ is the identity matrix; eq. 17 & 18) are positive (i.e. $a_i > 0$), and

2. $a_1 a_2 - a_3 > 0$ (if $J\big|_*$ is a 3×3 matrix) or $a_1 a_2 a_3 - a_3^2 - a_1^2 a_4 > 0$ (if $J\big|_*$ is a 4×4 matrix).

Thus, by applying the Routh-Hurwitz criterion we can analytically evaluate the stability around the equilibrium points of the non-linear systems given by the three- and four-pool models without directly evaluating λ analytically. The characteristic polynomial for a 3×3 matrix is given by

$$\lambda^3 + a_1\lambda^2 + a_2\lambda + a_3 = 0 \tag{17}$$

and for a 4×4 matrix is

$$\lambda^4 + a_1\lambda^3 + a_2\lambda^2 + a_3\lambda + a_4 = 0 \quad . \tag{18}$$

In both cases, a_1 is the negative trace of $J\big|_* $ $(a_1 = -\mathrm{tr}(J\big|_*))$.

2.3.2 Numerical simulations

We also compute λ and the steady state values of the state variables numerically. If not otherwise specified, $100\,000$ Monte Carlo simulations were produced by randomly drawing parameter values from (log-)uniform distributions using a latin-hybercube sampling algorithm (MATLAB R2022b's `lhsdesign` function; The MathWorks Inc., 2022). All parameters and their respective ranges are listed in Table 3. Partitioning coefficients were sampled from uniform distributions while rate constants were $\log_{10}$-transformed before sampling.

Following Georgiou et al. (2017) and Sierra and Müller (2015), the stability of equilibria was evaluated using the "damping coefficient" given by

$$\zeta = \min\left[\frac{-\mathrm{Re}(\lambda)}{\sqrt{\mathrm{Re}(\lambda)^2 + \mathrm{Im}(\lambda)^2}}\right] \quad , \tag{19}$$

which ranges between -1 and 1. ζ has positive values only if all $\mathrm{Re}(\lambda) < 0$ indicating a stable equilibrium and negative values if any $\mathrm{Re}(\lambda) > 0$ indicating an unstable equilibrium. For $\mathrm{Im}(\lambda) = 0$, ζ is either 1 or -1 indicating no oscillations, while $-1 < \zeta < 1$ for $\mathrm{Im}(\lambda) \neq 0$ indicates that oscillations occur.

Numerical simulations were carried out in MATLAB 2022b (The MathWorks Inc., 2022).

2.3.3 Classification of equilibrium points

Our analyses only consider physically meaningful equilibrium points - that is, only equilibrium points for which all state variables are simultaneously positive and real. Within the physically meaningful equilibrium points we distinguish three categories:

1. stable: all physically meaningful equilibrium points that are stable (i.e., stable node or focus points, Argyris et al., 2015),

2. stable and plausible: all physically meaningful equilibrium points that are stable and also give plausible numerical results

3. unstable: all physically meaningful equilibrium points that are not stable (i.e., unstable node or focus points, Argyris et al., 2015).

Table 3. Description of all parameters, their units, and used ranges for Monte Carlo simulations. Where applicable, parametrizations of limiting cases are separated by |. Parameter ranges where derived from Hararuk et al. (2015); Tao et al. (2023) and Cotrufo and Lavallee (2022). Baseline values were based on "conventional" values defined by Tao et al. (2023). See Supplemental Information (SI) Sect. 1 for derivation of parameter ranges for m, r, and l kinetics as well as for d'_B and I.

Symbol	Description	Unit	Baseline	Range
	Rate Constants			
I	Organic C input rate	$\mathrm{mg\,C\,g^{-1}\,d^{-1}}$	varied	$1.88\cdot10^{-4} - 2.43\cdot10^{-2}$
v_m^p	Depolymerization rate coefficient (m)	$\mathrm{g\,mg\,C^{-1}\,d^{-1}}$	$1.99\cdot10^{-1}$	$9.13\cdot10^{-3} - 5.48\cdot10^{3}$
v_f^p	Depolymerization rate coefficient (f)	$\mathrm{d^{-1}}$	$5.93\cdot10^{1}$	$9.13 - 2.74\cdot10^{5}$
v_r^p	Depolymerization rate coefficient (r)	$\mathrm{d^{-1}}$	$2.49\cdot10^{-1}$	$9.13\cdot10^{-2} - 2.74$
K_f^p	Depolymerization half-saturation constant (f)	$\mathrm{mg\,C\,g^{-1}}$	$3.00\cdot10^{2}$	$5.00\cdot10^{1} - 1.00\cdot10^{3}$
K_r^p	Depolymerization half-saturation constant (r)	$\mathrm{mg\,C\,g^{-1}}$	$2.00\cdot10^{-1}$	$2.50\cdot10^{-2} - 3.00$
v_m^u	Uptake rate coefficient (m)	$\mathrm{g\,mg\,C^{-1}\,d^{-1}}$	1.25	$3.04\cdot10^{-2} - 1.10\cdot10^{2}$
v_f^u	Uptake rate coefficient (f)	$\mathrm{d^{-1}}$	$2.49\cdot10^{-1}$	$9.13\cdot10^{-2} - 2.74$
v_l^u	Linear uptake rate coefficient	$\mathrm{d^{-1}}$	1.25	$3.04\cdot10^{-3} - 1.10\cdot10^{1}$
K_f^u	Uptake half-saturation constant (f)	$\mathrm{mg\,C\,g^{-1}}$	$2.00\cdot10^{-1}$	$2.50\cdot10^{-2} - 3.00$
l_S	Loss rate coefficient of SOC	$\mathrm{d^{-1}}$	0	–
l_D	Leaching rate coefficient of DOC	$\mathrm{d^{-1}}$	varied $\mid 0$	$2.74\cdot10^{-4} - 2.74\cdot10^{-1} \mid 0$
l_E	Leaching rate coefficient of extracellular enzymes	$\mathrm{d^{-1}}$	0	–
d_B	Decay rate coefficient of biomass	$\mathrm{d^{-1}}$	$4.81\cdot10^{-3}$	$1.37\cdot10^{-3} - 2.74\cdot10^{-1}$
d'_B	Density-dependent d_B	$\mathrm{g\,mg\,C^{-1}\,d^{-1}}$	–	$1.37\cdot10^{-2} - 2.74$
d_E	Decay rate coefficient of extracellular enzymes	$\mathrm{d^{-1}}$	$2.49\cdot10^{-2}$	$2.74\cdot10^{-3} - 2.74$
m_E	Constitutive enzyme production rate coefficient	$\mathrm{d^{-1}}$	$1.25\cdot10^{-4}$	$8.22\cdot10^{-5} - 1.83\cdot10^{-4} \mid 0$
	Partitioning Coefficients			
r_B	Recycling efficiency of decayed biomass	1	1.00	$0.20 - 1.00$
y_m	Maximal yield	1	0.60	$0.01 - 0.80$
y_B	Fraction of uptake going to biomass production	1	y_m	$0 < y_B \leq y_m \mid y_m$
f_I	Fraction of input going to SOC	1	0.90	$0.50 - 1.00$
f_D	Fraction of decayed biomass going to SOC	1	0.50	$0.50 - 1.00$
	Parameter Groups			
α	Extracellular enzyme turnover	$\mathrm{d^{-1}}$		$\alpha = d_E + l_E$
β	Microbial biomass turnover	$\mathrm{d^{-1}}$		$\beta = d_B + m_E$
η	Parameter group 1	$\mathrm{d^{-1}}$		$\eta = (y_m - y_B)\,d_B + y_m m_E \geq 0$
ω	Parameter group 2	$\mathrm{d^{-2}}$		$\omega = \alpha\beta - \alpha y_B r_B d_B - \eta d_E \geq 0$

Based on data synthesized by Wang et al. (2013) and educated guesses we applied the following conditions for considering results as "plausible": tOC=SOC+DOC+MBC+ENZ $\leq$ 500 mgC g^{-1} (=50 %), DOC/tOC < 0.01, MBC/tOC < 0.05, and ENZ/MBC < 0.1, where tOC indicates the total organic carbon content (the sum of all four carbon pools).

2.3.4 Causal loop analysis

Additionally to the mathematical analysis of equilibrium points and their stability, we present causal loop diagrams that qualitatively summarize causal links in a system and the feedbacks they create (Haraldsson, 2004). This analysis can help to understand the behaviour a system exhibits after a perturbation around an equilibrium point. In a causal loop diagram, causal connections are depicted by arrows, tying a cause (at the tail of the arrow) to its direct effect (at the head of the arrow) (Haraldsson, 2004). The sign of the causal relation (+ or −) depends on whether an isolated change in one element causes another to change in the same (+) or opposite (−) direction of the initial change (relative to the unchanged state) (Haraldsson, 2004; Richardson, 1986). For example, a decrease in the microbial uptake rate would lead to relatively less microbial biomass (compared to the case that the uptake rate had not changed), describing a positive causal relation. Closed loops with zero or an even number of negative interactions are denoted as positive or reinforcing feedback loops R, and closed loops with a odd number of negative interactions as a negative or balancing feedback loops B (Haraldsson, 2004; Richardson, 1986).

3 Results

Table 4. Summary of analysed models, respective simplifying assumptions, and type of analysis (ana. = analytical, num. = numerical).

Model	# Pools	Kinetics	Simplifying Assumptions		Analysis
SBE	3	m, f, r, e	none		ana.
SDB	3	$m \times m, f \times f, r \times f$	$L_S = L_E = 0;$	$y_B = y_m$	ana.
$SDBE$	4	$m \times m, f \times f, r \times f$	$L_S = L_E = 0;$ $\begin{cases} y_B = y_m \\ m_E = 0 \end{cases}$		ana. + num. / num.

We analyzed three model versions with different model structures (number of state variables and/or non-linearites; Table 4, Sect. 2.2). We first present analytical results on the simpler three-pool models with one (SBE) and two non-linear terms (SDB), followed by analytical and numerical results on the four-pool $SDBE$ model (Table 4). We use causal loop diagrams to qualitatively interpret these results.

3.1 SBE model: neglecting DOC dynamics

3.1.1 Steady state solutions

Table 5. Summary of steady state solutions of the three-pool SBE model for different kinetics of depolymerization. The "biotic" equilibrium solutions for microbial biomass and extracellular enzymes have the same form for any chosen kinetic.

	Kinetic (i)	$S_{k,i}^*$	$B_{k,i}^*$	$E_{k,i}^*$
abiotic ($k = 0$)	i	$\frac{I}{l_S}$	0	0
biotic ($k = 1$)	m	$\frac{\alpha\beta}{v_m^p \eta}$		
	f	$\frac{\alpha\beta}{v_f^p \eta} \frac{K_f^p}{1 - \frac{\alpha\beta}{v_f^p \eta}}$	$\frac{\alpha y_B}{\omega} l_S \left(S_0^* - S_{1,i}^* \right)$	$\frac{\eta}{\alpha y_B} B_{1,i}^*$
	r	$\frac{\alpha\beta}{v_r^p \eta} \frac{K_r^p \omega + I\eta}{\omega + l_S \frac{\alpha\beta}{v_r^p}}$		
	e	$\frac{\alpha\beta}{v_e^p \eta} \frac{K_e^p \omega + I\eta}{\omega - \omega \frac{\alpha\beta}{v_e^p \eta} + l_S \frac{\alpha\beta}{v_e^p}}$		

For all kinetic descriptions of the depolymerization rate (Table 2), the three-pool SBE model has two equilibrium points (EPs). Of these, one is an "abiotic" equilibrium Q_0, where only SOC exists and microbial biomass and extracellular enzymes are zero, i.e. $Q_0 = (S_0^*, 0, 0)$. Here the asterisk indicates a state variable at steady state and the subscript 0 signifies the "abiotic" solution. In turn, for each kinetic there exists an alternative "biotic" equilibrium point with non-zero microbial biomass and extracellular enzymes, $Q_{1,i} = (S_{1,i}^*, B_{1,i}^*, E_{1,i}^*)$. The steady state solutions for these equilibria depend on the i-th formulation used to describe P_i. All solutions are reported in Table 5, where for convenience parameters have been combined into parameter groups: extracellular enzyme turnover $\alpha > 0$, microbial biomass turnover $\beta > 0$ and parameter groups $\eta \geq 0$ and $\omega \geq 0$ (Table 3).

While the "abiotic" equilibrium is always positive, the "biotic" one can only be positive (and thus physically meaningful) if

$$S_{1,i}^* < S_0^* \quad \rightarrow \quad l_S S_{1,i}^* < I, \tag{20}$$

i.e. if the linear SOC loss rate is smaller than SOC inputs. Note that for f and e kinetics additional conditions apply for positivity.

If abiotic loss of SOC is neglected (i.e. $l_S = 0$), the "abiotic" equilibrium does not exist (SOC would accumulate at the constant rate I) and by eq. 20 the "biotic" equilibrium is always physically meaningful.

3.1.2 Stability analysis

To analyze whether a physically meaningful equilibrium point is also stable we apply the Routh-Hurwitz criterion to the Jacobian matrix J_i^{SBE} (eq. A1) evaluated at the k-th equilibrium point ($J_i^{SBE}\big|_{*,k}$) – either the "abiotic" ($k = 0$) or the "biotic" equilibrium ($k = 1$) (detailed in Appendix Sect. A1 and SI Sect. 2.1). From this we find that the stability of a physically

meaningful "abiotic" equilibrium is conditional on

$$\left.\frac{\partial P_i}{\partial E}\right|_{*,0} < \frac{\alpha\beta}{\eta} \quad , \tag{21}$$

whereas all physically meaningful "biotic" equilibrium points of the three-pool SBE model are also stable.

Fig. 2a shows a simplified causal loop diagram of the SBE model (sparing all loss and decay terms) that can help to understand the dynamic behaviour of the model after a perturbation around an equilibrium. The reinforcing loop R_1 describes the increase in microbial biomass with increasing depolymerization rate ($\propto$ uptake rate), leading to increased ENZ production rate, ENZ concentration, and consequently a further increasing depolymerization rate. This reinforcing effect is dampened by the balancing loops B_1 (the depletion of SOC by depolymerization) and B_2 (the carbon cost of ENZ production). The reinforcing loop R_2 exists only if inducible ENZ production is considered ($y_B < y_m$; higher depolymerization stipulates the production of more extracellular enzymes, which promote depolymerization). R_1 and R_2 are not independent of each other and have to obey mass balance — i.e. per unit of uptake an increase in inducible ENZ production ($\propto y_m - y_B$) can only be achieved by reducing the built-up of microbial biomass (lowering y_B). An extreme case of this is $y_B = 0$; in this case, no microbial biomass is produced and only B_1 and R_2 remain. From the stability analysis (Appendix Sect. A1 and SI Sect. 2.1), we obtained no condition on stability of physically meaningful equilibrium points in the SBE model. Thus, for any proportion of constitutive vs. inducible ENZ production, all physically meaningful "biotic" equilibria are also stable. That is, the dynamic behaviour of the model after a perturbation around its equilibrium is dominated by the balancing feedbacks, ensuring a convergence back to the equilibrium.

3.1.3 Exclusive stability of either "abiotic" or "biotic" equilibrium

We recall that for the "biotic" equilibrium to be physically meaningful it is required that $S_{1,i}^* < S_0^*$ (eq. 20); whereas for the "abiotic" equilibrium to be stable it is required by eq. 21 that $\left.\frac{\partial P_i}{\partial E}\right|_{*,0} < \frac{\alpha\beta}{\eta}$. This condition translates, e.g. for multiplicative kinetics to:

$$\left.\frac{\partial P_m}{\partial E}\right|_{*,0} = v_m^p S_0^* < \frac{\alpha\beta}{\eta} \quad \rightarrow \quad S_0^* < \frac{\alpha\beta}{v_m^p \eta} = S_{1,m}^* \quad . \tag{22}$$

This means that when the "biotic" equilibrium is physically meaningful the "abiotic" equilibrium is unstable and vice versa. Therefore, no region in the parameter space yields a physically meaningful bi-stability in which "biotic" and "abiotic" equilibira are simultaneously physically meaningful and stable. This holds for all evaluated kinetics (see SI Sect. 2.1 for the remaining analytical derivations).

3.2 SDB model: neglecting ENZ dynamics

3.2.1 Steady state solutions

Table 6. Summary of steady states for the three-pool SDB and four-pool $SDBE$ model for different kinetics of depolymerization and uptake. The "abiotic" steady state is only defined for $l_S > 0$. "Biotic" steady states are given for $l_E = l_S = 0$ and $y_B = y_m$. $E^{\dagger}_{k,i\times j}$ signifies $E^{qss}_{k,i\times j}$ or $E^{*}_{k,i\times j}$ in the SDB respectively $SDBE$ model. The "biotic" equilibrium solutions for microbial biomass and enzymes have the same form for all chosen kinetics.

	Kinetics	$S^{*}_{k,i\times j}$	$D^{*}_{k,i\times j}$	$B^{*}_{k,i\times j}$	$E^{\dagger}_{k,i\times j}$
abiotic ($k=0$)	$i\times j$	$f_I \dfrac{I}{l_S}$	$(1-f_I)\dfrac{I}{l_D}$	0	0
	$m\times m$	$\dfrac{d_E}{y_m v_m^p m_E}\gamma_{m\times m}$	$\dfrac{\beta}{y_m v_m^u}$		
biotic ($k=1$)	$f\times f$	$\dfrac{d_E}{y_m v_f^p m_E}\gamma_{f\times f}\dfrac{K_f^p}{1-\frac{d_E}{y_m v_f^p m_E}\gamma_{f\times f}}$	$K_f^u \dfrac{\beta}{y_m v_f^u - \beta}$	$\dfrac{y_m}{\pi}\left(I - l_D D^{*}_{1,i\times j}\right)$	$\dfrac{m_E}{d_E}B^{*}_{1,i\times j}$
	$r\times f$	$\dfrac{d_E}{y_m v_r^p m_E}\gamma_{r\times f}\left(K_r^p + E^{\dagger}_{1,r\times f}\right)$	$K_f^u \dfrac{\beta}{y_m v_f^u - \beta}$		

260 Table 6 reports the steady state solution of the three-pool SDB model, where for convenience parameters were further grouped in

$$\gamma_{i\times j} = f_D r_B d_B y_m + \pi f_I \frac{I}{I - l_D D^{*}_{1,i\times j}} \tag{23}$$

and

$$\pi = \frac{1}{d_E}\omega(l_E = 0, y_B = y_m) = \underbrace{(1 - y_m)}_{>0} m_E + \underbrace{(1 - y_m r_B)}_{>0} d_B > 0 \quad . \tag{24}$$

265 Because an "abiotic" equilibrium exists only for $l_S > 0$ (Table 6) we only evaluate the stability of the "biotic" equilibrium and drop "1" from the subscript for conciseness. "Biotic" steady states can only be physically meaningful for $I > l_D D^{*}_{i\times j}$; meaning the DOC leaching flux can not be larger than the total OC input flux. With $f \times f$ and $r \times f$ kinetics it is additionally required that $\beta < y_m v_f^u$, implying that the maximal per-biomass assimilation rate $y_m v_f^u$ must be larger than microbial biomass turnover ($\beta = m_E + d_B$; Table 3). For $f \times f$ kinetics it is additionally required that $\frac{d_E}{y_m v_f^p m_E}\gamma_{f\times f} < 1$ for steady states to be 270 positive.

In absence of DOC leaching (for $l_D = 0$), with $m \times m$ and $f \times f$ kinetics $S^{*}_{i\times j}$ becomes independent of I (by eq. 23) while $B^{*}_{i\times j}$ and $E^{*}_{i\times j}$ are linear functions of I. In contrast, for $l_D = 0$, $S^{*}_{r\times f}$ is linearly dependent on OC input I. For $l_D > 0$, $S^{*}_{m\times m}$ is a function of $\frac{I}{I - l_D D^{*}_{m\times m}}$ (eq. 23) causing $S^{*}_{m\times m}$ to decline with increasing inputs as $\frac{I}{I - l_D D^{*}_{m\times m}} \to 1$ for $I >> l_D D^{*}_{m\times m}$. Only in $S^{*}_{r\times f}$ does I still appear in a linear term also for $l_D > 0$.

3.2.2 Stability analysis

The Jacobian matrix of the SDB model around the "biotic" equilibrium $J^{SDB}_{i\times j}\big|_{*}$ is given by eq. A10. Evaluating the respective coefficients of the characteristic polynomial and the requirement that $a_1 a_2 - a_3 > 0$ yields a cumbersome sufficient and

necessary condition for stability of the "biotic" EPs of the SDB model (eq. A11-A13, see details in Appendix Sect. A2 and additionally SI Sect. 2.2). The appearance of such a conditional statement means that in contrast to the SBE model, the SDB

280 model can have physically meaningful but unstable EPs (i.e. if the conditions described by eq. A11-A13 do not hold). A perturbation around such an unstable EP will cause the system to diverge from the EP. In this case, the biotic pools (MBC and quasi-steady state ENZ) will collapse, while DOC will reach a steady state as $D^*_{0,i\times j}$ (for $l_D > 0$), and SOC will accumulate indefinitely.

The obtained sufficient and necessary condition given by eq. A11-A13 does not allow for easy interpretation or application.

285 We thus propose a sufficient (i.e. a more conservative or strict) condition for stability that is easier to trace analytically as

$$\frac{\partial P_i^{qss}}{\partial S}\bigg|_* + y_m f_D r_B d_B \geq y_m \frac{\partial P_i^{qss}}{\partial B}\bigg|_* \quad . \tag{25}$$

This sufficient condition for stability of "biotic" equilibrium points of the SDB model holds for all cases relevant to our analysis (SI Sect. 2.2). Described in words, this condition requires the depolymerization rate to be less sensitive to a change in microbial biomass than to a proportional change in SOC. Fig. 2b illustrates this relation in a simplified causal loop diagram.

290 The reinforcing loop R_1 causes the depolymerization rate (P_i^{qss}) to increase as the (quasi-steady state) ENZ concentration increases (quantified by $\frac{\partial P_i^{qss}}{\partial B}\big|_*$). This then increases DOC concentration, uptake, and ultimately causes a further increase in microbial biomass and (quasi-steady state) ENZ. This positive feedback is accelerated by an additional reinforcing feedback loop (R_2) that causes uptake to further increase as microbial biomass increases. The balancing loops B_2 and B_3 respectively describe the depletion of DOC with increasing uptake and the reduction of biomass as more extracellular enzymes are being

295 produced. Lastly, the balancing loop B_1 causes SOC to change in the opposite direction than the depolymerization rate (i.e., SOC is depleted as depolymerization increases and relatively more SOC remains as depolymerization decreases), counteracting the initial change in P_i^{qss} (quantified by $\frac{\partial P_i^{qss}}{\partial S}\big|_*$). Therefore, the sufficient stability condition in eq. 25 can be interpreted in the sense that the negative feedback B_1 must be quantitatively stronger than the positive feedback R_1 (by some factor y_m and buffered by a constant term; eq. 25).

300 In essence, the positive feedback R_1 can drive the system to overshoot or collapse: e.g., if microbial biomass or (quasi-steady state) ENZ concentration happens to decrease due to a perturbation, this will reduce the depolymerization rate, and following the positive feedback, result in further reduced MBC and (quasi-steady state) ENZ concentration. The biotic pools would collapse and the system would not be able to recover to its initial equilibrium (i.e. be unstable). Only if the entailing accumulation of SOC increases the depolymerization rate more than it is reduced by the depletion of (quasi-steady state) ENZ,

305 will the system be able to recover and retain the biotic components (i.e. be stable).

We note that for linear uptake kinetics (i.e. $U_l = v_l^u D$), additional terms in the necessary and sufficient condition can help to fulfill the Routh-Hurwitz criterion for stability (see details in Appendix Sect. A2). In the causal loop diagram, linear uptake kinetics remove the positive feedback between uptake rate and biomass (R_2 in Fig. 2b vanishes). Although we could not show this analytically, numerical evaluation showed that for the chosen parameter spaces (Table 3) with linear uptake kinetics U_l,

310 the physically meaningful EPs of the SDB model were always stable (SI Fig. S3c-d).

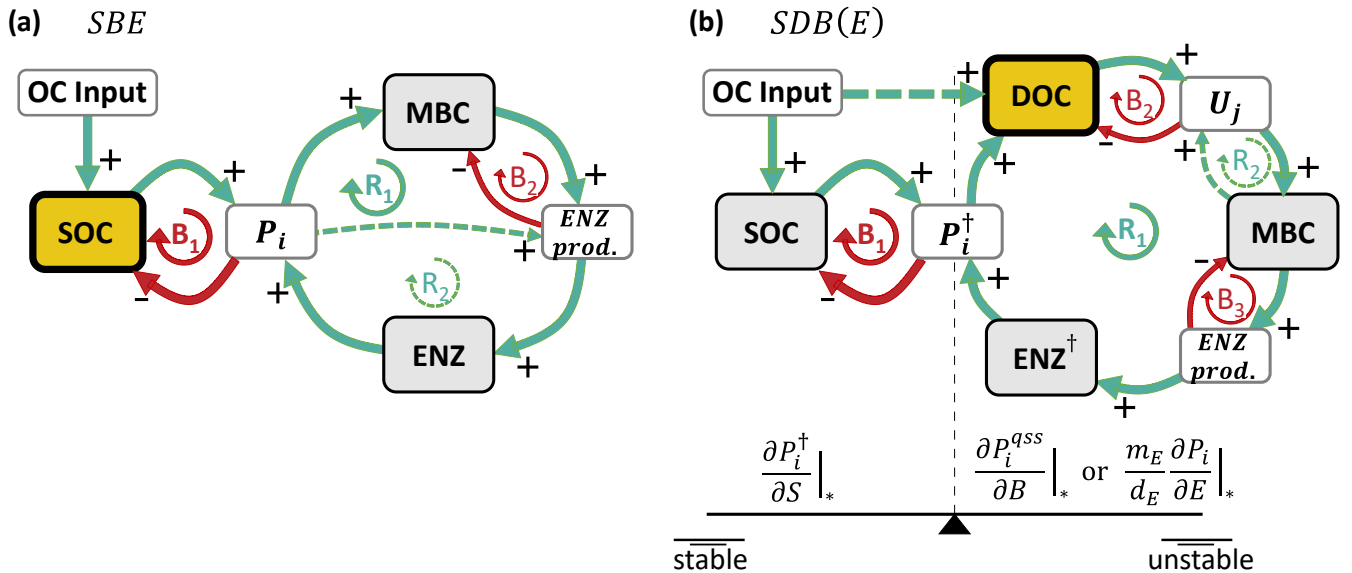



Figure 2. Simplified causal loop diagrams of the SBE (a) and SDB and $SDBE$ models (b). Green arrows marked with "+" indicate positive and red arrows marked with "−" negative interactions. R signifies reinforcing and B balancing loops. Pools are marked by grey, rates by white boxes. The microbial growth substrate is highlighted in yellow with a thick outline. In (a), the dashed green line (at R_2) indicates the effect of inducible ENZ production and vanishes if only constitutive ENZ production is considered. In (b), the dashed green line (at R_2) indicates the effect of biomass-dependent non-linear uptake kinetics and vanishes for biomass-independent uptake kinetics. In the SDB ($SDBE$) model $P_i^{\dagger}$ and ENZ† signify P_i^{qss} (P_i) and $E_{i\times j}^{qss}$ ($E_{i\times j}^*$). The seesaw in (b) illustrates the balance between the partial derivatives in eq. 25 & 26 and how it affects stability. See Sect. 2.3.4 for details on how to read the causal loop diagrams.

3.3 $SDBE$ model: full archetypal model

The four-pool $SDBE$ model with $L_S = L_E = 0$ and $y_B = y_m$ has the same steady state solutions as the three-pool SDB model, but now the solution for the ENZ pool is denoted as $E_{i\times j}^*$ because ENZ is not considered to be at quasi-steady state as in the SDB model (Table 6).

3.3.1 Analytical stability analysis

In the four-pool $SDBE$ model, the coefficients of the characteristic polynomial of $J_{i\times j}^{SDBE}\big|_*$ (eq. A17) remain analytically tractable (SI Sect. 2.3). The trace of $J_{i\times j}^{SDBE}\big|_*$ is always negative (and thus $-\mathrm{tr}(J_{i\times j}^{SDBE}\big|_*) = a_1 > 0$) and its determinant is always positive ($\det(J_{i\times j}^{SDBE}\big|_*) = a_4 > 0$). However, the additional Routh-Hurwitz criterion for the 4×4 matrix $J_{i\times j}^{SDBE}\big|_*$ (given by $a_1 a_2 a_3 - a_3^2 - a_1^2 a_4 > 0$) becomes analytically intractable. Despite this additional complexity, we can still draw some conclusions based on similarities between the SDB and $SDBE$ models. Considering that in the $SDBE$ model ENZ dynamics are explicitly represented (and thus e.g. $\frac{\partial P_i^{qss}}{\partial B} \to \frac{m_E}{d_E}\frac{\partial P_i}{\partial E}$) similar conditions emerge for positivity of the coefficients of the characteristic polynomial as in the SDB model (SI Sect. 2.3). Based on these similarities, we propose that the sufficient

condition for stability found for the three-pool SDB model might also hold in the $SDBE$ model. This proposed sufficient condition is given by

$$325 \quad Z_{i\times j} = \left.\frac{\partial P_i}{\partial S}\right|_* + y_m f_D r_B d_B - y_m \frac{m_E}{d_E}\left.\frac{\partial P_i}{\partial E}\right|_* \quad ; \quad Z_{i\times j} \geq 0 \quad . \tag{26}$$

The simplified causal loop diagram of the $SDBE$ model in Fig. 2b gives rise to the same interpretation of this condition as in the SDB model. In the following we confirm that this condition holds in the $SDBE$ model via numerical analysis.

3.3.2 Numerical stability analysis

Testing the sufficient condition for stability

330 We produced $100\,000$ Monte Carlo simulations and computed the damping coefficient ζ (eq. 19) to numerically evaluate the stability of equilibrium points in the $SDBE$ model. Within the sampled parameter space (Table 3), physically meaningful equilibrium points for which $Z_{i\times j} \geq 0$ (eq. 26) always also had $\zeta > 0$ and were thus stable (Fig. 3a-b, SI Fig. S1). Damping coefficients with $\zeta \leq 0$ were only observed when $Z_{i\times j} < 0$ (that is for points above the black line in Fig. 3a or red points in Fig. 3c). While a total of 46 932 evaluated equilibrium points were physically meaningful and stable, less than half of these
335 (22 386) also fulfilled the condition $Z_{i\times j} \geq 0$. In turn, the majority of these equilibrium points (24 546) were stable despite contradicting this condition – i.e. the condition given by $Z_{i\times j} \geq 0$ is very conservative. In case the condition is fulfilled, the value of $Z_{i\times j}$ correlates well with the value of ζ meaning that for larger $Z_{i\times j}$ oscillations are generally more dampened (Fig. 3b).

Changing environmental conditions, microbial physiology, and stability

340 Even for the simplified sufficient condition $Z_{i\times j} \geq 0$ analysis is cumbersome for kinetics other than $m \times m$ (SI Sect. 2.3). We thus varied specific parameters individually and evaluated their effect on the numerically computed damping coefficient ζ.

Keeping all microbial and enzymatic parameters constant (set to their baseline value, that is for constitutive ENZ production, Table 3), stability depends on the environmental control parameters l_D and I (Fig. 4a). At baseline parameter values, most environmental conditions yield stable EPs but strong oscillations around these EPs occur (damping coefficient $\zeta < 1$). As
345 conditions become less favourable and either l_D increases and/ or I decreases, equilibrium points can become unstable ($\zeta < 0$).

Next, we analysed the influence of individual microbial parameter values on the stability of EPs for a number of scenarios defined by combinations of I and l_D (Fig. 4b-d; Table 7). Generally, if DOC leaching is neglected (solid lines, Fig. 4b-d), the variation in just one parameter rarely leads to unstable equilibria (only at very high microbial decay rates d_B). In contrast, if DOC leaching occurs, variation in key physiological parameters can lead to a transition from stable to unstable EPs. This
350 happens e.g. as y_m becomes too low or d_B too high – i.e. for specific environmental conditions there are lower threshold values for y_m and upper thresholds for d_B beyond which equilibria become unstable (Fig. 4b-c, Table 7). With the exception of f_D, the partitioning of decayed microbial biomass between SOC and DOC, all parameters show such a threshold within the explored ranges (SI Fig. S2 and SI Sect. 2.3).

Using the alternative description of inducible production of extracellular enzymes, the stability behaviour with respect to
355 changes in y_B is more varied (Fig, 4d). As for y_m, there are lower y_B threshold values below which steady states become

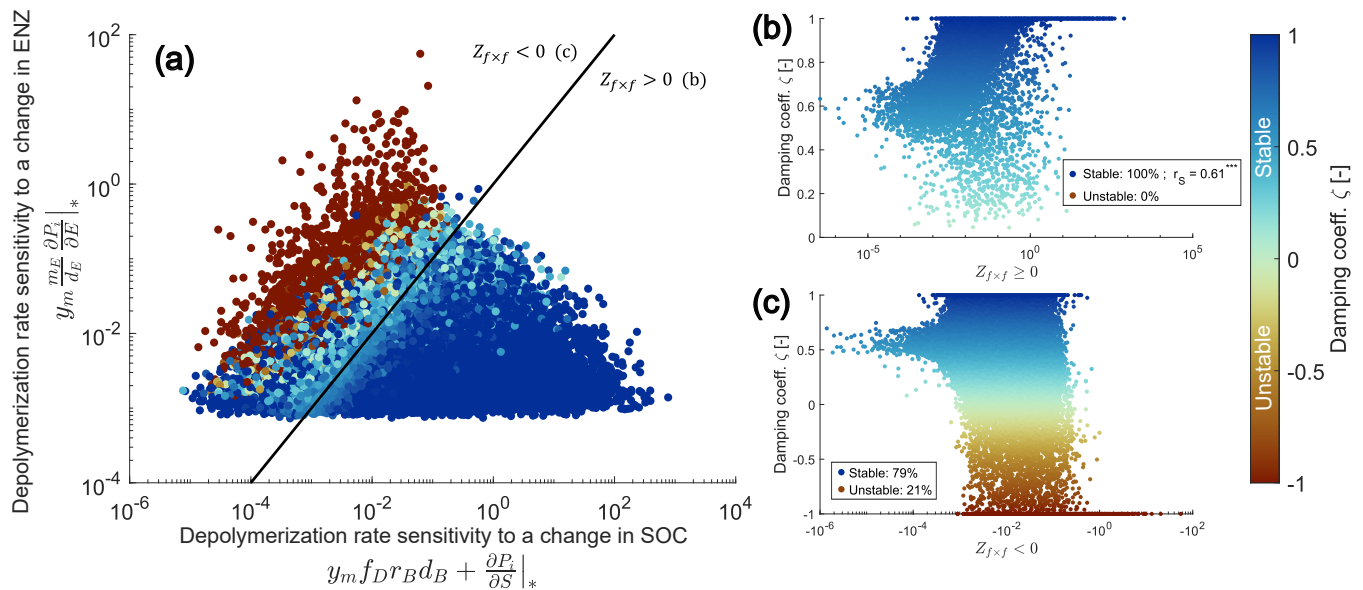



Figure 3. Numerical evaluation of the proposed sufficient condition for stability of the $SDBE$ model with $f \times f$ kinetics and constitutive ENZ production. The proposed stability condition $Z_{f \times f} \geq 0$ (eq. 26) is derived from an analytical evaluation of the simpler SDB model. 100 000 Monte Carlo calculations of equilibrium points were produced sampling the parameter space in Table 3. (a) illustrates the separation of all physically meaningful equilibrium points by the positive and negative terms of $Z_{f \times f}$. Points on and below the black 1:1 line (indicating $Z_{f \times f} = 0$) fulfill the condition $Z_{f \times f} \geq 0$. The color-code indicates the value of the damping coefficient ζ. (b) and (c) show values of ζ vs. values of $Z_{f \times f}$ (for $Z_{f \times f} \geq 0$ in (b), and $Z_{f \times f} < 0$ in (c)). Legends in (b) and (c) state the percentages of stable and unstable equilibrium points; r_S in (b) gives the Spearman rank correlation coefficient (***: significant at $p < 0.005$).

unstable. However, there can also be upper thresholds for y_B above which too few enzymes are being produced to ensure sufficient C acquisition.

In summary, by varying only individual parameters, instabilities can arise when assimilation, depolymerization, or ENZ production are too low; or when abiotic C losses are too high (Fig. 4b-d; Fig. S2; Table 7). These results are in line with the

360 analytic analysis of the sufficient condition (eq. 26) for $m \times m$ kinetics (SI Sect. 2.3).

Density-dependent mortality

Georgiou et al. (2017) proposed a density-dependent formulation of the microbial decay rate ($D'_B = d'_B B^b$ with $1 < b \leq$ 2) as an alternative to the conventional linear decay term that yields mostly stable non-oscillatory behaviour (note that this formulation causes both microbial mortality and maintenance respiration to be density-dependent). For our $SDBE$ model

365 with DOC leaching, we could only find an analytical steady state solution for $m \times m$ kinetics and $b = 2$. In this case, the density dependent formulation could vastly alleviate the previously observed instability and resulted in damping coefficients for plausible equilibrium points close to 1 for most of the explored parameter spaces (SI Fig. S4a). However, some physically

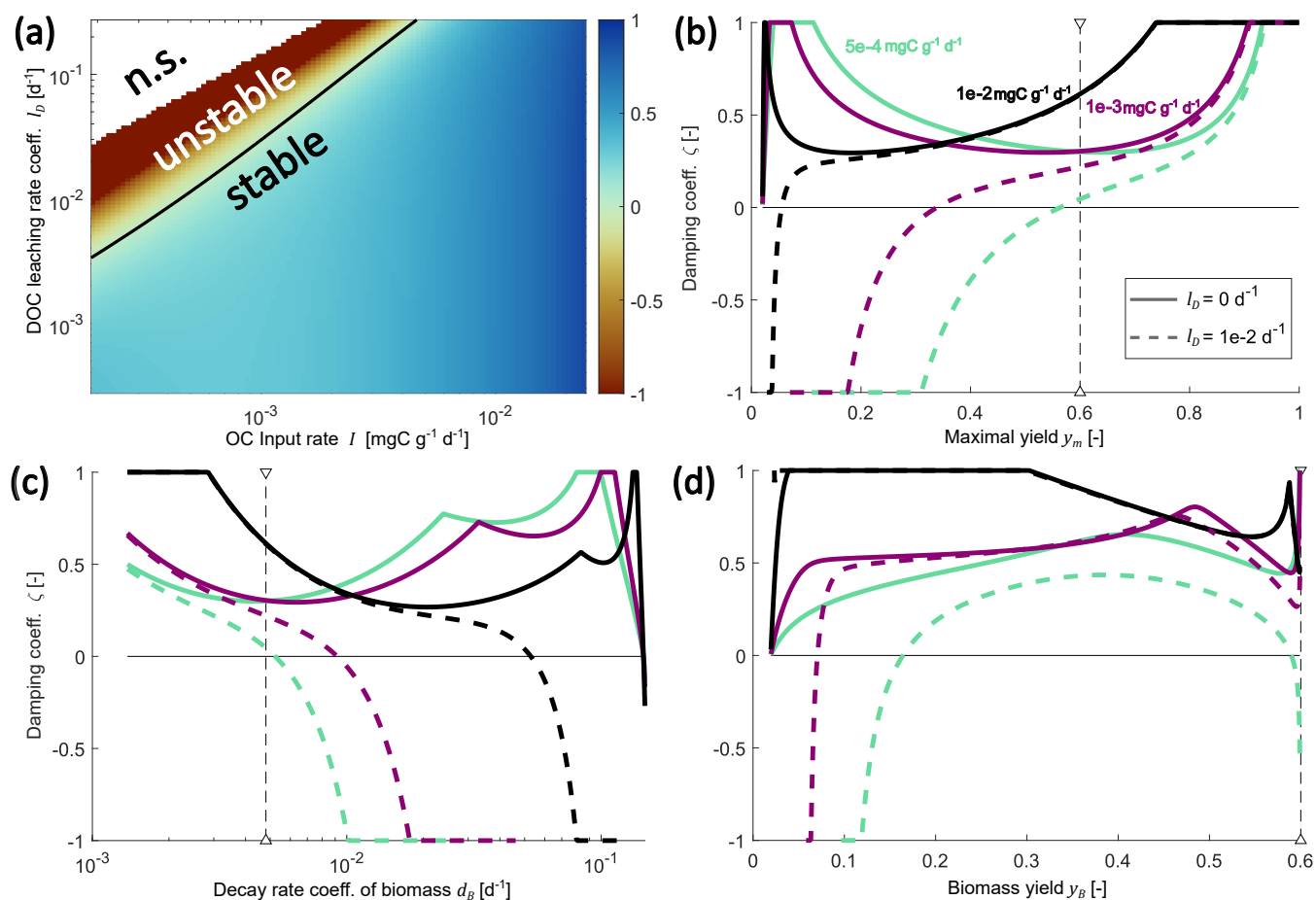
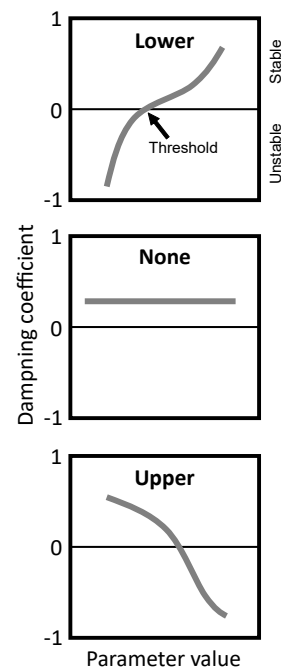



Figure 4. Changes in the damping coefficient with changes in environmental controls (a) and some microbial physiology parameters (b-d) in the $SDBE$ model with $f \times f$ kinetics. (a) shows the dampening coefficient as a function of I and l_D with all other parameters held at their baseline values (Table 3). The black line indicates a $\zeta = 0$ (n.s. = no physically meaningful solution). (b-d) show variation in the damping coefficient for different combinations of I and l_D values. Different line styles indicate scenarios with different DOC leaching rate coefficients: solid lines indicate $l_D = 0$ and dashed lines $l_D = 1 \cdot 10^{-2} \, \mathrm{d}^{-1}$. Different line colors indicate scenarios with different OC input rates: turquoise lines indicate $I = 5 \cdot 10^{-4} \, \mathrm{mg \, C \, g^{-1} \, d^{-1}}$, violet lines $I = 1 \cdot 10^{-3} \, \mathrm{mg \, C \, g^{-1} \, d^{-1}}$, and black lines $I = 1 \cdot 10^{-2} \, \mathrm{mg \, C \, g^{-1} \, d^{-1}}$. In (b) and (c) the baseline model with constitutive production of extracellular enzymes is used and y_m and d_B varied respectively. Baseline parameter values are indicated by vertical dashed lines. In (d) instead only inducible production of extracellular enzymes is considered and y_B varied. Note the log x- and y-axis in panel (a) and the log x-axis in panel (c).

Table 7. Indication of thresholds in parameter values for stability of the $SDBE$ model. Analysis is based on exclusively varying one parameter while keeping all others at their baseline value (Table 3; Fig. 4, SI Fig. S2). The meaning of the different thresholds is illustrated in the schematics on the right. This analysis applies to all evaluated combinations of kinetics ($m \times m$, $f \times f$, $r \times f$) using constitutive ENZ production; except for analysis of y_B where we considered only inducible ENZ production. ($i = m, f, r$), ($j = m, f$).

Process	Parameter	Threshold		
		Lower	Upper	None
Depolymerization	v_i^p	X		
	K_i^p		X	
Assimilation	v_j^u	X		
	K_f^u		X	
	y_m	X		
ENZ production	m_E	X		
	y_B	X	X	
Decay	d_B		X	
	d_E		X	
Mass balance losses	r_B	X		
	l_D		X	
OC Input	I	X		
SOC-DOC partition	f_I		X	
	f_D			X

meaningful but unstable EPs were still observed. Only with negligible DOC leaching did physically meaningful but unstable equilibrium points vanish completely. This was numerically tested for $m \times m$, $f \times f$, and $r \times f$ kinetics (SI Fig. S4b-d).

370   *Instability and predicted organic carbon pools*

Fig. 5a-b illustrates the joint distributions of physically meaningful SOC and MBC pools in the $f \times f$ model for scenarios where DOC leaching is either considered (Fig. 5a) or neglected (Fig. 5b) (for 10 000 Monte Carlo simulations within the parameter ranges given in Table 3). By accounting for DOC leaching, just about half of the simulations yield physically meaningful EPs – of which most (4 698) were also stable, but more than 10 % (687) were unstable. Neglecting DOC leaching

375 increases the total number of physically meaningful results (6 896) and simultaneously reduces the relative share of physically meaningful but unstable results to < 5 % (312). In both scenarios, most simulations yield implausible results for steady state C stocks – e.g. MBC being larger than SOC. However, in both scenarios, unstable EPs largely overlap with stable and plausible outcomes in the SOC-MBC solution space. This is also evident in the empirical probability density functions of all the four state variables (Fig. 5c-d). Especially if DOC leaching is considered ($l_D > 0$, Fig. 5c), values of plausible and unstable steady

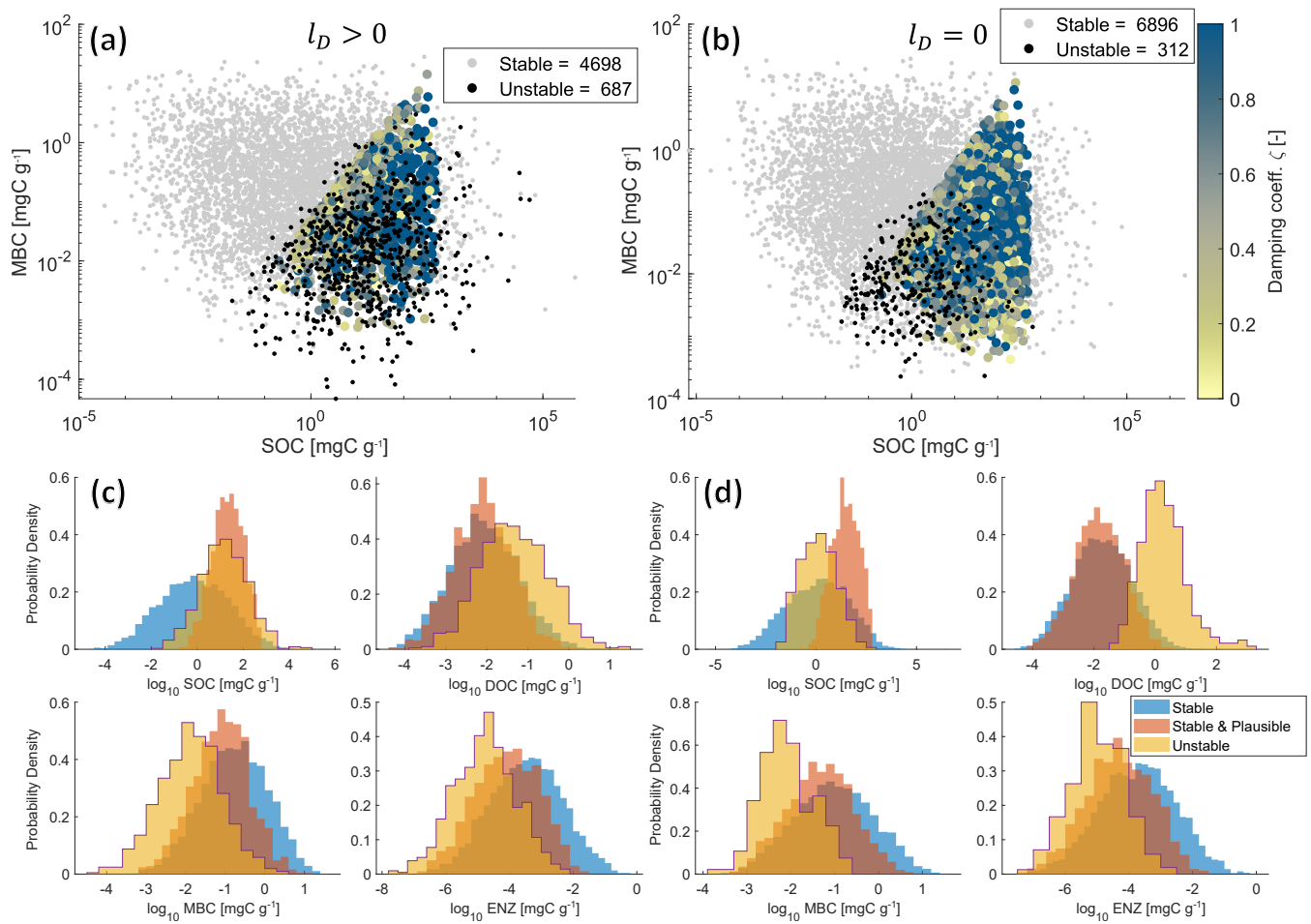



Figure 5. Physically meaningful (positive & real) steady state solutions from 10 000 Monte Carlo simulations of the $SDBE$ model with $f \times f$ kinetics and constitutive ENZ production. Scatter plots of MBC vs. SOC concentrations are shown for $l_D > 0$ (a) and $l_D = 0$ (b). Color-coded points are stable and plausible steady state solutions, the color code indicating the value of the damping coefficient. Grey points are stable but not plausible steady state solutions, and black points are physically meaningful but unstable steady state solutions. Legends indicate the numbers of physically meaningful and stable (stable + stable & plausible) or unstable EPs. Plot groups (c) and (d) show empirical probability density functions of each state variable for stable, stable & plausible, and unstable physically meaningful steady state solutions for $l_D > 0$ (c) and $l_D = 0$ (d), respectively. Note that SOC contents > 1000 mg C g^{-1} are mathematically possible but unphysical model outcomes, as we neglect soil volume changes.

380   state SOC pools largely overlap. In contrast, the distributions of plausible and unstable steady state pool sizes of MBC, DOC, and ENZ do not overlap as closely as those for SOC. These distinctions are amplified in the cases where DOC leaching is neglected ($l_D = 0$, Fig. 5d), in particular for DOC.

4 Discussion

4.1 Model structure matters: standard microbial-explicit SOC models can have unstable equilibria

385 Microbial-implicit models with linear decomposition kinetics are stable as long as there is no inert pool, mass conservation applies and rates are proportional to the amount of carbon in the donor pool (Sierra and Müller, 2015). It is more difficult to define general stability criteria for nonlinear models, whose structure and type of nonlinearity affects the model behaviour around the equilibrium. Manzoni and Porporato (2007) and Raupach (2007) showed analytically that the non-trivial steady state of two-pool models, consisting of a substrate pool and a microbial pool (denoted as "harvester" system in Raupach,

390 2007), are always stable for multiplicative and forward Michaelis-Menten kinetics (but only under the assumption that the input to the substrate (I) is a constant; Raupach, 2007). We show here that the same is true for all physically meaningful, non-trivial ("biotic") equilibrium points also if a third pool representing extracellular enzymes is added (SBE model). This result holds irrespective of the kinetic laws used to describe SOC depolymerization and whether ENZ production is considered to be constitutive, inducible, or a combination of both.

395 Interestingly, by introducing a second non-linear term Raupach (2007) found that unstable equilibria could emerge in their two-pool model. In contrast, Wang et al. (2014, 2016) demonstrated for several versions of a three-pool (litter-SOC-microbes) model with two non-linearities (microbial degradation and subsequent uptake of litter and SOC) that the equilibrium points of these models were always stable. An underlying assumption in these models was that the available substrate pool (similar to what we described as DOC) was at quasi-steady state. Our derivation of the SBE model follows a similar simplification –

400 and also does not yield unstable behaviour. By contrast, unstable equilibrium points are possible in our three- and four-pool model versions with two non-linearities that explicitly consider DOC (SDB and $SDBE$ model). Whether equilibrium points in microbial-explicit SOC models can become unstable is thus not dependent on the number of pools or the number of non-linearities *per se*, but rather on the combination of non-linearities, the coupling of different pools and rates, and what feedbacks they create.

405 Comparing our three- and four-pool models to the simpler two-pool model analysed by Raupach (2007) can help to understand why instability can occur in these models. Briefly, their model describes human consumption of a food resource, but is in structure similar to our models (analogous terms in our models are given in brackets): the resource (SOC in SBE, DOC in SDB and $SDBE$ models), is taken up by the human consumer (microbes), and thereby depleted. The uptake process is always described as a non-linear term, equivalent to our description of U_j. Raupach (2007) analysed two different cases with respect

410 to the resupply of the resource (I in SBE, $(1 - f_I)I + P_i^\dagger$ in SDB and $SDBE$): 1) resupply is independent of the available resource and 2) resupply is dependent on the resource itself. The first case is similar to our SBE model, where the resource SOC is replenished by the external input I and is thus independent of the SOC availability itself. In these cases the "biotic" (respectively resource-human coexistance in Raupach, 2007) equilibrium is always stable if it is physically meaningful. In turn, the second case can be compared to our SDB and $SDBE$ models: unless the external input to DOC is very high (for low

415 f_I), the replenishment of the resource DOC is dominated by the depolymerization rate $P_i^\dagger$ – which via the positive feedback loop R_1 is dependent on the available DOC (compare Fig. 2b). These models can have unstable "biotic" equilibria (Raupach,

2007). Therefore, the (more or less direct) dependency of the resource resupply on the abundance of the resource itself can be identified as the root cause that allows for instability in these models.

We further tested this hypothesis by setting $f_I = 0$ in the $SDBE$ model (i.e., all external input goes into DOC directly), bypassing the dependency of DOC replenishment on depolymerization. In line with our expectation (and the analytical solution of $Z_{m \times m}$ and $Z_{f \times f}$ for $f_I = 0$ in SI Sect. 2.3) this effectively prevented the occurrence of unstable EPs (SI Fig. S5).

4.2 Avoiding instability

Our analysis of different model structures and their stability behaviour points to three approaches to avoid unstable EPs in microbial-explicit SOC models:

1. model structure – avoid positive feedback coupling between microbial growth substrate (here DOC) available for uptake and its resupply

2. kinetic formulations – avoid accelerated depletion of DOC by reducing the dependency of uptake on microbial biomass

3. parameter values – choose parameter values so that the sufficient and/or necessary conditions for stability are met.

The first approach is commonly taken in models that assume DOC to be in quasi-steady state (e.g. Wang et al., 2014, 2016), but might have shortcomings in cases where DOC dynamics become important e.g. if drying-rewetting dynamics or leaching are relevant. If DOC leaching is not considered to be a relevant process, neglecting this process but keeping a dynamic description of DOC can already considerably reduce the likelihood of unstable EPs.

The second approach is used e.g. in models that assume DOC uptake as independent of microbial biomass, but dependent on the availability of DOC respectively its diffusive flux to a cell (e.g. Manzoni et al., 2014). Alternatively, using for instance reverse Michaelis-Menten kinetics to describe microbial uptake can dampen oscillations (Wang et al., 2016). Since uptake kinetics using reverse Michaelis-Menten or the ECA formulation become similar to linear uptake kinetics at relatively high concentrations of microbial biomass they could also help to alleviate instability issues under these conditions. However, when using forward Michaelis-Menten kinetics for decomposition and DOC assimilation, the half saturation constants can attain large values after data assimilation, suggesting that decomposition kinetics might be approximated by multiplicative rather than linear kinetics, at least in large scale model applications (Tao et al., 2024b).

Lastly, the third approach might seem straight forward as we could expect parameter values calibrated with measurement data to yield both stable and plausible EPs. However, our numerical simulations indicated that especially if DOC leaching is considered, calibrating parameter values with SOC and microbial biomass data alone could still lead to plausible yet unstable EPs (Fig. 5). While data on carbon contents in the extracellular enzyme pool are still not available, combining microbial biomass data with quantitative data on DOC pools (as e.g. in Wang et al., 2013) could help to avoid calibration to parameter values that lead to unstable EPs. Moreover, stability criteria can be obeyed in various other ways, e.g. by considering correlations between parameter values or introducing additional constraints on microbial physiology.

4.2.1 Correlations between parameter values

Correlations between parameter values could effectively alleviate the occurrence of unstable EPs by simultaneously changing
parameter values that appear on both sides of the inequality given by eq. 25 or 26, thereby ensuring that these conditions are
always fulfilled even as parameter values change. Some evidence for this to be effective is provided e.g. by Hararuk et al. (2015),
who used the four-pool AWB model (Allison et al., 2010) (similar to the $SDBE$ model) for predictions of global carbon stocks.
They prescribed e.g. the uptake and depolymerization rate coefficients and the respective half-saturation constants to positively
correlate with temperature. Thus, with the same directional change in temperature, these parameter values change in opposite
directions with respect to their threshold values for stability (Table 7) – i.e., with a decrease in temperature v_f^p decreases,
moving closer to its threshold, but simultaneously also K_f^p decreases, moving further away from its threshold. Consequently,
for a wide range of parameter values the conditions for stability could be fulfilled. Indeed, Hararuk et al. (2015) reported that
they did not observe any unstable equilibria with maximum-likelihood parameters in their global study.

Beyond the qualitative assessment of parameter thresholds (Table 7), explicit analytical expressions of the necessary or suf-
ficient conditions for stability (as for $Z_{m \times m}$ in SI Sect. 2.3) could be used to quantitatively assess what parameter correlations
are required to ensure stability of equilibria across reasonable parameter ranges. However, for other kinetic formulations than
$m \times m$ these terms might become difficult to trace analytically.

4.2.2 Eco-evolutionary constraints on microbial traits

The observation that stability thresholds for parameters shift as environmental conditions change can further be interpreted
in the light of expected variations in microbial functional traits. While some kinetic rate parameters might be correlated due
to thermodynamics (e.g. temperature response of rate parameters), correlations among other parameters, like the investment
into growth or extracellular enzyme production, might rather emerge as outcomes of eco-evolutionary processes that select
specific combination of traits in a given environment (Abs et al., 2023, 2024). These combinations of traits would manifest
themselves as microbial life-history strategies under different environmental conditions (e.g. Malik et al., 2020). Following
this logic, changing environmental conditions could constrain the space for microbial physiological adaptation because mi-
crobial traits would need to ensure stability. For example, very inefficient microbes (having a low y_m) could not establish a
stable equilibrium under very unfavourable conditions (low OC input and/or high DOC leaching, Fig. 4b) – unless other traits
change simultaneously. This reflects a basic principle of ecology: that organisms have to be adapted to the environment they
inhabit, and such adaptation is expressed through sets of coordinated traits. Currently, this basic principle is not integrated in
microbial-explicit SOC models (but see Abs et al., 2022, for a recent attempt at addressing this challange), which can lead to
matching specific environmental conditions with a (modelled) microbial population that is not able to sustain itself under those
conditions.

While our results are restricted to equilibrium conditions in the absence of external perturbations (e.g., fluctuating environ-
mental conditions), we can speculate that trait coordination should emerge also in a fluctuating environment (for an example of
plant trait coordination under stochastic soil moisture, see Bassiouni et al., 2023). In fact, changes in the microbial community

with relevance for SOC cycling can happen on different time-scales, from hours (stress response), and weeks (changes in community composition), to years and decades (mutations) (Abs et al., 2024). Environmental fluctuations occur at all these scales, and microbial communities are well equipped to adapt to these stochastically changing environments.

Integrating soil microbial ecological understanding into microbial-explicit SOC models could thus yield alternative mathematical descriptions or parameter relations that could prevent mismatching between parameter values and environmental conditions, and ultimately improve model applicability (Georgiou et al., 2017). Evidence on the importance of microbial ecology and evolution for SOC cycling is accumulating (Abs et al., 2022, 2023). For instance, microbes have been found to invest more into production of extracellular enzymes in soils with lower SOC contents (Calabrese et al., 2022; Malik et al., 2019) and density-dependent microbial mortality, a concept derived from ecological considerations, can effectively alleviate oscillatory behaviour (Georgiou et al., 2017).

Yet, revisiting the proposed stability criterion from an eco-evolutionary perspective prompts a paradox. In a model that obeys the proposed stability criterion (eq. 25 and 26), a reduction in enzyme production will always lead to an increase of the depolymerization rate — a strong disincentive for microbial investment into enzyme production. We can resolve this paradox by realizing that the SDB and $SDBE$ models as analyzed (and applied, e.g. by Hararuk et al. (2015) but also the original model formulation by Allison et al. (2010)), there is no process (such as competition with other (micro-)organisms or abiotic removal of (accessible) SOC through e.g. leaching or occlusion) that competes with microbes for SOC. Consequently, the only alternative to SOC degradation is its accumulation. In turn this allows for high depolymerization rates at minimal microbial investment into enzyme production. This outcome emerges also from a mathematical analysis of optimal substrate utilization – without losses of substrate due to abiotic processes or competition, decomposers do not have any reason to invest in resource acquisition (Manzoni et al., 2023). Thus, a rigorous eco-evolutionary optimization approach (that for instance aims to maximize microbial growth rate) cannot be readily implemented with the current model structure but would require extension of the model to account for competition for SOC. Our analysis instead demonstrates how (given the model structure) microbes might adapt to environmental conditions not to maximize their fitness, but to attain a stable population. Whether stability and evolutionary fitness maximization are convergent (i.e., an organism at an evolutionary fitness maximum would also establish a stable population) poses an interesting endeavor for future research.

4.2.3 Comparing approaches to avoid instability

It is important to note that while our analysis indicates all of the three approaches listed in Sect. 4.2 as feasible means to avoid instability, it gives no indication on which of the approaches should be preferred. As different objectives and research questions have differing requirements (e.g. whether occurrence of instabilities should globally be avoided by removing the positive DOC feedback or whether it suffices to constrain the available parameter space) we cannot give a general recommendation. However, an important distinction between the different approaches is that while the first two (removal of an explicit DOC representation respectively the dependency of uptake on microbial biomass) are simplifications of the system that might require further justification for specific use cases, the third approach can add realism to the model by explicitly considering the interaction between environmental conditions and the microbial community. In line with Abs et al. (2022, 2023, 2024) and Georgiou et al.

515 (2017) we want to highlight the potential of using such ecologically consistent mathematical descriptions to improve current model formulations. In other words, we cannot simply add a biotic component to models without acknowledging that this component has to be "adapted" (as species and communities are in the real world) to the environmental conditions it is exposed to.

4.3 Implications

4.3.1 Mathematical insights on microbial mediated SOC cycling

520 Currently, the debate on the implications of different model formulations for microbial mediated SOC cycling is ongoing (He et al., 2024; Lennon et al., 2024; Tao et al., 2023, 2024a). Our contribution provides an additional perspective to this discussion, leveraging less mechanistic arguments but rather mathematical requirements. Based on our analysis, we cannot draw any conclusions on which (if any) model version is most suitable to realistically represent microbial mediated SOC cycling at the 525 large scale. Microbial-explicit models including a DOC pool (SDB and $SDBE$ models) create a positive feedback loop that allows for instabilities to occur, but this finding by no means indicates a shortcoming of such models. In fact, the positive feedback loop is a direct consequence of our conceptual understanding of microbial mediated, enzymatic driven degradation of organic carbon substrates in soils (Kuzyakov et al., 2000; Kuzyakov, 2010). At small spatial scales oscillatory behaviour can be realistic (Manzoni and Porporato, 2007) and also collapse of a local microbial population is not difficult to imagine (the 530 alternative and perhaps more realistic outcome that microbes turn dormant cannot be described by these models). However, at large spatial scales these behaviours are not observed, indicating that SOC cycling is more stable at these scales (Wang et al., 2014; Georgiou et al., 2017). This can be seen as indicative of a scaling problem – the same process cannot be described by the same mathematical formulations across scales (Chakrawal et al., 2020; Wilson and Gerber, 2021). The proposed stability criterion as well as the outlined approaches to avoid instability could in turn guide development of upscaled model formulations 535 – suitable upscaled kinetics might employ one of the proposed approaches to avoid instability or could be designed to obey the proposed (note: *very conservative*, see further discussion below) stability criteria.

4.3.2 Applicability of the proposed stability criterion

We could identify a sufficient and necessary condition for stability of the SDB model (eq. A11-A13). However, the condition we found is difficult to interpret and apply. We thus proposed a stricter but simpler sufficient condition for stability (eq. 25). 540 By comparing the SDB and $SDBE$ model we proposed that a similar constraint ($Z_{i \times j} \geq 0$, eq. 26) would also hold as a sufficient condition for the $SDBE$ model, despite the Routh-Hurwitz stability criterion being not fully tractable analytically for this model version. Numerical analyses confirmed that the proposed sufficient condition ensures stability of the $SDBE$ model within the vast parameter space we explored. However, these sufficient conditions are very conservative and can exclude a substantial fraction of the physically meaningful and stable equilibrium points. Further, despite a clear correlation between 545 $Z_{i \times j}$ and the damping coefficient ζ, the stability condition does not give direct insights into the oscillation behaviour. How useful the stricter sufficient and necessary conditions would be in constraining model parameters – as compared to the simpler

sufficient conditions – might depend on the specific model applications. Despite the potential challenges in evaluating these conditions, they can still be useful to understand the processes or parameter interactions that cause unstable EPs to occur and can guide ecology-informed model developments.

4.3.3 Conditions leading to instability

Our numerical analysis of the $SDBE$ model indicates that instability of equilibrium points becomes more likely with decreasing carbon inputs, increasing DOC leaching, and low process rates (Fig. 4, Table 7). All these conditions are most likely to be met in high altitude and/or latitude environments. This is in line with Hararuk et al. (2015), who observed strongest oscillations (longest time to dampen oscillations, indicative of diminishing real parts of the eigenvalues) of their calibrated four-pool model in these regions. Therefore, analytical steady states of microbe-explicit SOC models applied in high altitude and/or latitude environments could be unstable and analytical steady state solutions could thus not reliably be used for initialization of simulation runs or prediction of SOC stocks.

5 Conclusions

Microbial-explicit SOC models aim to improve the representation of SOC dynamics by accounting for its biotic control. At very small spatial and temporal scales their oscillatory behaviour and potential for instability can reflect relevant (micro-)ecosystem processes (Manzoni and Porporato, 2007). However, if applied at larger scales such as in Earth-system models, these properties can result in unrealistic simulation outcomes (Georgiou et al., 2017; Wang et al., 2014). Here we analyzed what processes can lead to instability in these models. By comparing the stability behaviour of an archetypal microbial-explicit SOC model (the AWB model; Allison et al., 2010) with some reduced model versions and stability analysis of similar models in the literature, we found that instability can occur in models that assume a positive feedback between the resupply of a microbial growth substrate (i.e. DOC) and its abundance. We found that stability is (sufficiently) conditional on the balance between the sensitivity of the depolymerization rate to changes in extracellular enzyme vs. SOC concentration. Based on these analyses, we suggest that instability can be avoided by selecting specific 1) model structures, 2) kinetic formulations, and/or 3) parameter relations or values. While these approaches can vastly differ, an emerging common theme is that acknowledging ecological principles and processes can be leveraged to improve model applicability. These findings have implications for further development of microbial-explict models and potential upscaling approaches, calling for ecological consistent model formulations and rigorous mathematical analysis of newly introduced models.

Appendix A: Details on stability analyses

A1 SBE model

575 The Jacobian matrix of the SBE model is given by

$$J_i^{SBE} = \begin{bmatrix} \frac{\partial \dot{S}}{\partial S} & \frac{\partial \dot{S}}{\partial B} & \frac{\partial \dot{S}}{\partial E} \\ \frac{\partial \dot{B}}{\partial S} & \frac{\partial \dot{B}}{\partial B} & \frac{\partial \dot{B}}{\partial E} \\ \frac{\partial \dot{E}}{\partial S} & \frac{\partial \dot{E}}{\partial B} & \frac{\partial \dot{E}}{\partial E} \end{bmatrix} = \begin{bmatrix} -l_s - \frac{\partial P_i}{\partial S} & r_B d_B & d_E - \frac{\partial P_i}{\partial E} \\ y_B \frac{\partial P_i}{\partial S} & -(d_B + m_E) & y_B \frac{\partial P_i}{\partial E} \\ (y_m - y_B)\frac{\partial P_i}{\partial S} & m_E & (y_m - y_B)\frac{\partial P_i}{\partial E} - (d_E + l_E) \end{bmatrix} \quad , \tag{A1}$$

where $\frac{\partial x}{\partial y}$ is the partial derivative of x with respect to y and $\dot{x}$ is used to denote $\frac{\mathrm{d}x}{\mathrm{d}t}$ so that e.g. $\frac{\partial \dot{S}}{\partial S} = \frac{\partial}{\partial S}\left(\frac{\mathrm{d}S}{\mathrm{d}t}\right)$.

A1.1 Stability analysis of the "abiotic" equilibrium

First, we evaluate the parameter space in which the "abiotic" equilibrium is stable. Substituting the steady state solutions for
580 Q_0 given in Table 5 into J_i^{SBE} (eq. A1) yields

$$J_i^{SBE}\big|_{*,0} = \begin{bmatrix} -l_s & r_B d_B & d_E - \frac{\partial P_i}{\partial E}\big|_{*,0} \\ 0 & -\beta & y_B \frac{\partial P_i}{\partial E}\big|_{*,0} \\ 0 & m_E & (y_m - y_B)\frac{\partial P_i}{\partial E}\big|_{*,0} - \alpha \end{bmatrix} \quad . \tag{A2}$$

For Q_0 to be stable by the Routh-Hurwitz criterion, it is required that all the coefficients a_i of the characteristic polynomial of $J_i^{SBE}\big|_{*,0}$ and additionally $a_1 a_2 - a_3$ be positive. By this we find that stability of Q_0 is conditional on the sufficient and necessary condition (SI Sect. 2.1)

585
$$\frac{\partial P_i}{\partial E}\bigg|_{*,0} < \frac{\alpha\beta}{\eta} \quad . \tag{A3}$$

A1.2 Stability analysis of the "biotic" equilibrium

Next, we analyze the stability of the "biotic" equilibrium, by evaluating the Jacobian matrix J_i^{SBE} (eq. A1) around its "biotic" steady states $Q_{1,i}$ (Table 5)

$$J_i^{SBE}\big|_{*,1} = \begin{bmatrix} -l_s - \frac{\partial P_i}{\partial S}\big|_{*,1} & r_B d_B & d_E - \frac{\partial P_i}{\partial E}\big|_{*,1} \\ y_B \frac{\partial P_i}{\partial S}\big|_{*,1} & -\beta & y_B \frac{\partial P_i}{\partial E}\big|_{*,1} \\ (y_m - y_B)\frac{\partial P_i}{\partial S}\big|_{*,1} & m_E & (y_m - y_B)\frac{\partial P_i}{\partial E}\big|_{*,1} - \alpha \end{bmatrix} \quad . \tag{A4}$$

590 To evaluate the Routh-Hurwitz criterion it is convenient to re-express P_i in terms of $\frac{\partial P_i}{\partial E}$ as

$$P_i = \frac{\partial P_i}{\partial E} x_i^{-1} E \quad , \tag{A5}$$

where the factor x_i is introduced to maintain generality of this substitution for any P_i as defined in Table 2:

$$x_i = \begin{cases} 1 & \text{if } (i = m, f) \\ 1 - \frac{E}{K_r^p + E} & \text{if } i = r \\ 1 - \frac{E}{K_e^p + S + E} & \text{if } i = e \end{cases} \quad . \tag{A6}$$

x_i has the convenient property $0 < x_i \leq 1$. The ensuing derivations hold for all x_i as long as $0 < x_i \leq 1$ and also for other kinetics P_j not explored here for which a $0 < x_j \leq 1$ can be found that satisfies eq. A5.

Substituting eq. A5 and $B_{1,i}^* = \frac{\alpha y_B}{\eta} E_{1,i}^*$ (Table 5) into eq. 12 evaluated at steady state yields

$$\frac{dE_i}{dt} = (y_m - y_B) \left.\frac{\partial P_i}{\partial E}\right|_{*,1} x_i^{-1}\big|_{*,1} E_{1,i}^* + m_E \frac{\alpha y_B}{\eta} E_{1,i}^* - d_E E_{1,i}^* - l_E E_{1,i}^* = 0 \quad , \tag{A7}$$

from which, for $E_{1,i}^* \neq 0$ we find

$$\left.\frac{\partial P_i}{\partial E}\right|_{*,1} = \alpha x_i\big|_{*,1} \frac{1 - \frac{m_E y_B}{\eta}}{y_m - y_B} = \frac{\alpha \beta}{\eta} x_i\big|_{*,1} \quad . \tag{A8}$$

With this definition we obtain

$$J_i^{SBE}\big|_{*,1} = \begin{bmatrix} -l_s - \frac{\partial P_i}{\partial S}\big|_{*,1} & r_B d_B & d_E - \frac{\alpha\beta}{\eta} x_i\big|_{*,1} \\ y_B \frac{\partial P_i}{\partial S}\big|_{*,1} & -\beta & y_B \frac{\alpha\beta}{\eta} x_i\big|_{*,1} \\ (y_m - y_B)\frac{\partial P_i}{\partial S}\big|_{*,1} & m_E & \alpha\left(x_i\big|_{*,1} - \frac{m_E y_B}{\eta} x_i\big|_{*,1} - 1\right) \end{bmatrix} \quad . \tag{A9}$$

From this it can be seen that the trace of $J_i^{SBE}\big|_{*,1}$ (the sum of the diagonal entries) is always negative, since $\frac{\partial P_i}{\partial S}\big|_{*,1} > 0$ and $x_i\big|_{*,1} \leq 1$; and thus $a_1 > 0$. Likewise it can be shown that all remaining coefficients of the characteristic polynomial are always positive and that $a_1 a_2 - a_3 > 0$ (see detailed analytical derivations in SI Sect. 2.1). Thus, all physically meaningful "biotic" equilibrium points of the three-pool SBE model are stable.

A2 SBD model

The Jacobian matrix $J_{i\times j}^{SDB}$ of the SDB model evaluated around the biotic equilibrium is given by

$$J_{i\times j}^{SDB}\big|_* = \begin{bmatrix} -\frac{\partial P_i^{qss}}{\partial S}\big|_* & 0 & -\frac{\partial P_i^{qss}}{\partial B}\big|_* + f_D r_B d_B \\ \frac{\partial P_i^{qss}}{\partial S}\big|_* & -\frac{\partial U_j}{\partial D}\big|_* - l_D & \frac{\partial P_i^{qss}}{\partial B}\big|_* - \frac{\partial U_j}{\partial B}\big|_* + (1 - f_D) r_B d_B + m_E \\ 0 & y_m \frac{\partial U_j}{\partial D}\big|_* & y_m \frac{\partial U_j}{\partial B}\big|_* - m_E - d_B \end{bmatrix} \quad . \tag{A10}$$

Since only the biotic EP of the SDB model is analysed, we dropped the subscript 1 here.

Evaluating the Routh-Hurwitz criterion for the SDB model with the chosen kinetic formulations (Table 2) yields the sufficient and necessary condition for stability as

$$a_1 a_2 - a_3 = X_{i\times j} + Y_{i\times j} > 0 \tag{A11}$$

with

$$X_{i\times j} = \frac{\partial U_j}{\partial D}\bigg|_* \left\{ \left(\frac{\partial U_j}{\partial D}\bigg|_* + \frac{\partial P_i^{qss}}{\partial S}\bigg|_* \right) \underbrace{\left(\frac{\partial P_i^{qss}}{\partial S}\bigg|_* + y_m f_D r_B d_B - y_m \frac{\partial P_i^{qss}}{\partial B}\bigg|_* \right)}_{\text{Term giving rise to the sufficient condition (eq. 25)}} + \pi \frac{\partial U_j}{\partial D}\bigg|_* \right\} \tag{A12}$$

and

$$Y_{i\times j} = -\frac{\partial \dot{B}}{\partial B}\bigg|_{*,j} \left\{ \frac{\partial P_i^{qss}}{\partial S}\bigg|_* \left(\frac{\partial P_i^{qss}}{\partial S}\bigg|_* - \frac{\partial \dot{B}}{\partial B}\bigg|_{*,j} \right) + \frac{\partial U_j}{\partial D}\bigg|_* \left(2 \cdot \frac{\partial P_i^{qss}}{\partial S}\bigg|_* + y_m f_D r_B d_B - y_m \frac{\partial P_i^{qss}}{\partial B}\bigg|_* + \pi \right) \right\} \quad , \tag{A13}$$

both shown for $l_D = 0$ for conciseness (see full expressions with $l_D \geq 0$ and detailed analysis of all coefficients of the characteristic polynomial in SI Sect. 2.2). The sufficient condition for stability given by eq. 25 in the main text holds for any $l_D \geq 0$. Beyond this sufficient condition for stability, the additional positive term $\pi \frac{\partial U_j}{\partial D}\big|_*$ in eq. A12 might indicate the stabilizing influence of the balancing loop B_2 (Fig. 2b) and the recycling of ENZ and MBC to SOC (compare eq. 24).

$\frac{\partial \dot{B}}{\partial B}\big|_{*,j}$ in eq. A13 is the lower right entry of $J_{i\times j}^{SDB}\big|_*$ (eq. A10) and is given by

$$\frac{\partial \dot{B}}{\partial B}\bigg|_{*,j} = \frac{\partial}{\partial B}\frac{dB}{dt}\bigg|_{*,j} = y_m \frac{\partial U_j}{\partial B}\bigg|_* - m_E - d_B \quad . \tag{A14}$$

For any choice of U_j that is linear in B (as is the case for U_m and U_f – compare Table 2) we find from solving $\frac{dB}{dt}$ (eq. 16) at steady state that $y_m \frac{\partial U_{(m,f)}}{\partial B}\big|_* = m_E + d_B$ and thus $\frac{\partial \dot{B}}{\partial B}\big|_{*,(m,f)} = 0$, so that $Y_{i\times(m,f)} = 0$ (eq. A13). The only necessary and sufficient condition for stability of the SDB model in these cases is thus $X_{i\times j} > 0$ (eq. A12) (for $l_D = 0$, see SI Sect. 2.2 for the corresponding necessary condition for $l_D > 0$).

For linear uptake kinetics U_l, that is

$$U_l = v_l^u D \quad , \tag{A15}$$

$Y_{i\times l}$ (eq. A13) does not vanish from eq. A11 since $\frac{\partial U_l}{\partial B}\big|_* = 0$ and consequently from eq. A14

$$-\frac{\partial \dot{B}}{\partial B}\bigg|_{*,l} = m_E + d_B > 0 \quad . \tag{A16}$$

Because of its additional positive components, $Y_{i\times j}$ can be positive even if the sufficient condition of eq. 25 is not fulfilled. Thus, using U_l can help to ensure positivity of all coefficients of the characteristic polynomial and $a_1 a_2 - a_3 > 0$.

A3 $SDBE$ model

The Jacobian matrix of the four-pool $SDBE$ model with $L_S = L_E = 0$ and $y_B = y_m$ (eq. 6) (Table 4) and evaluated around the biotic equilibrium is given by

$$J_{i\times j}^{SDBE}\bigg|_* = \begin{bmatrix} -\frac{\partial P_i}{\partial S}\big|_* & 0 & f_D r_B d_B & -\frac{\partial P_i}{\partial E}\big|_* \\ \frac{\partial P_i}{\partial S}\big|_* & -l_D - \frac{\partial U_j}{\partial D}\big|_* & (1-f_D) r_B d_B - \frac{\partial U_j}{\partial B}\big|_* & d_E + \frac{\partial P_i}{\partial E}\big|_* \\ 0 & y_m \frac{\partial U_j}{\partial D}\big|_* & y_m \frac{\partial U_j}{\partial B}\big|_* - (m_E + d_B) & 0 \\ 0 & 0 & m_E & -d_E \end{bmatrix} \quad . \tag{A17}$$

Author contributions. ES and SM conceptualized the study. ES and SG lead and SM and SB assisted the formal analysis and investigation. All authors discussed results together. ES wrote the original manuscript draft and produced the figures, with feedback from SM. All authors reviewed and commented on the manuscript draft and revisions.

640 *Competing interests.* The authors declare that they have no conflict of interest.

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References

Abs, E., Saleska, S., and Ferriere, R.: Microbial eco-evolutionary responses amplify global soil carbon loss with climate warming, Research Sqaure [preprint], https://doi.org/10.21203/rs.3.rs-1984500/v1, 2022.

Abs, E., Chase, A. B., and Allison, S. D.: How do soil microbes shape ecosystem biogeochemistry in the context of global change?, Environmental Microbiology, 25, 780–785, https://doi.org/10.1111/1462-2920.16331, 2023.

Abs, E., Chase, A. B., Manzoni, S., Ciais, P., and Allison, S. D.: Microbial evolution—An under-appreciated driver of soil carbon cycling, Global Change Biology, 30, e17 268, https://doi.org/https://doi.org/10.1111/gcb.17268, 2024.

Allison, S. D., Wallenstein, M. D., and Bradford, M. A.: Soil-carbon response to warming dependent on microbial physiology, Nature Geoscience, 3, 336–340, https://doi.org/10.1038/ngeo846, 2010.

Argyris, J. H., Faust, G., Haase, M., and Friedrich, R.: An Exploration of Dynamical Systems and Chaos: Completely Revised and Enlarged Second Edition, Springer Berlin, Heidelberg, ISBN 978-3-662-46042-9, https://doi.org/10.1007/978-3-662-46042-9, 2015.

Bassiouni, M., Manzoni, S., and Vico, G.: Optimal plant water use strategies explain soil moisture variability, Advances in Water Resources, 173, 104 405, https://doi.org/https://doi.org/10.1016/j.advwatres.2023.104405, 2023.

Bradford, M. A., Wieder, W. R., Bonan, G. B., Fierer, N., Raymond, P. A., and Crowther, T. W.: Managing uncertainty in soil carbon feedbacks to climate change, Nature Climate Change, 6, 751–758, https://doi.org/10.1038/nclimate3071, 2016.

Calabrese, S., Mohanty, B. P., and Malik, A. A.: Soil microorganisms regulate extracellular enzyme production to maximize their growth rate, Biogeochemistry, 158, 303–312, https://doi.org/10.1007/s10533-022-00899-8, 2022.

Chakrawal, A., Herrmann, A. M., Koestel, J., Jarsjö, J., Nunan, N., Kätterer, T., and Manzoni, S.: Dynamic upscaling of decomposition kinetics for carbon cycling models, Geoscientific Model Development, 13, 1399–1429, https://doi.org/10.5194/gmd-13-1399-2020, 2020.

Chakrawal, A., Calabrese, S., Herrmann, A. M., and Manzoni, S.: Interacting Bioenergetic and Stoichiometric Controls on Microbial Growth, Frontiers in Microbiology, 13, 859 063, https://doi.org/10.3389/fmicb.2022.859063, 2022.

Cotrufo, M. F. and Lavallee, J. M.: Soil organic matter formation, persistence, and functioning: A synthesis of current understanding to inform its conservation and regeneration, vol. 172 of *Advances in Agronomy*, pp. 1–66, Academic Press, ISBN 978-0-323-98953-4, https://doi.org/10.1016/bs.agron.2021.11.002, 2022.

Georgiou, K., Abramoff, R. Z., Harte, J., Riley, W. J., and Torn, M. S.: Microbial community-level regulation explains soil carbon responses to long-term litter manipulations, Nature Communications, 8, 1223, https://doi.org/10.1038/s41467-017-01116-z, 2017.

Haraldsson, H. V.: Introduction to System Thinking and Causal Loop Diagrams, Lund University, Department of Chemical Engineering, 2004.

Hararuk, O., Smith, M. J., and Luo, Y.: Microbial models with data-driven parameters predict stronger soil carbon responses to climate change, Global Change Biology, 21, 2439–2453, https://doi.org/10.1111/gcb.12827, 2015.

He, X., Abramoff, R. Z., Abs, E., Georgiou, K., Zhang, H., and Goll, D. S.: Model uncertainty obscures major driver of soil carbon, Nature, 627, E1–E3, https://doi.org/10.1038/s41586-023-06999-1, 2024.

Horn, R. A. and Johnson, C. R.: Topics in Matrix Analysis, Cambrigde University Press, 1991.

Kuzyakov, Y.: Priming effects: Interactions between living and dead organic matter, Soil Biology and Biochemistry, 42, 1363–1371, https://doi.org/https://doi.org/10.1016/j.soilbio.2010.04.003, 2010.

Kuzyakov, Y., Friedel, J., and Stahr, K.: Review of mechanisms and quantification of priming effects, Soil Biology and Biochemistry, 32, 1485–1498, https://doi.org/https://doi.org/10.1016/S0038-0717(00)00084-5, 2000.

Lennon, J. T., Abramoff, R. Z., Allison, S. D., Burckhardt, R. M., DeAngelis, K. M., Dunne, J. P., Frey, S. D., Friedlingstein, P., Hawkes, C. V., Hungate, B. A., Khurana, S., Kivlin, S. N., Levine, N. M., Manzoni, S., Martiny, A. C., Martiny, J. B. H., Nguyen, N. K., Rawat, M., Talmy, D., Todd-Brown, K., Vogt, M., Wieder, W. R., and Zakem, E. J.: Priorities, opportunities, and challenges for integrating microorganisms into Earth system models for climate change prediction, mBio, 15, e00 455–24, https://doi.org/10.1128/mbio.00455-24, 2024.

Malik, A. A., Puissant, J., Goodall, T., Allison, S. D., and Griffiths, R. I.: Soil microbial communities with greater investment in resource acquisition have lower growth yield, Soil Biology and Biochemistry, 132, 36–39, https://doi.org/10.1016/j.soilbio.2019.01.025, 2019.

Malik, A. A., Martiny, J. B. H., Brodie, E. L., Martiny, A. C., Treseder, K. K., and Allison, S. D.: Defining trait-based microbial strategies with consequences for soil carbon cycling under climate change, The ISME Journal, 14, 1–9, https://doi.org/10.1038/s41396-019-0510-0, 2020.

Manzoni, S. and Porporato, A.: A theoretical analysis of nonlinearities and feedbacks in soil carbon and nitrogen cycles, Soil Biology and Biochemistry, 39, 1542–1556, https://doi.org/10.1016/j.soilbio.2007.01.006, 2007.

Manzoni, S. and Porporato, A.: Soil carbon and nitrogen mineralization: Theory and models across scales, Soil Biology and Biochemistry, 41, 1355–1379, https://doi.org/10.1016/j.soilbio.2009.02.031, 2009.

Manzoni, S., Schaeffer, S., Katul, G., Porporato, A., and Schimel, J.: A theoretical analysis of microbial eco-physiological and diffusion limitations to carbon cycling in drying soils, Soil Biology and Biochemistry, 73, 69–83, https://doi.org/10.1016/j.soilbio.2014.02.008, 2014.

Manzoni, S., Chakrawal, A., and Ledder, G.: Decomposition rate as an emergent property of optimal microbial foraging, Frontiers in Ecology and Evolution, 11, https://doi.org/10.3389/fevo.2023.1094269, 2023.

Raupach, M. R.: Dynamics of resource production and utilisation in two-component biosphere-human and terrestrial carbon systems, Hydrology and Earth System Sciences, 11, 875–889, https://doi.org/10.5194/hess-11-875-2007, 2007.

Richardson, G. P.: Problems with causal-loop diagrams, System Dynamics Review, 2, 158–170, https://doi.org/10.1002/sdr.4260020207, 1986.

Schimel, J. P. and Weintraub, M. N.: The implications of exoenzyme activity on microbial carbon and nitrogen limitation in soil: a theoretical model, Soil Biology and Biochemistry, 35, 549–563, https://doi.org/10.1016/S0038-0717(03)00015-4, 2003.

Sierra, C. A. and Müller, M.: A general mathematical framework for representing soil organic matter dynamics, Ecological Monographs, 85, 505–524, https://doi.org/10.1890/15-0361.1, 2015.

Tang, J. and Riley, W. J.: Competitor and substrate sizes and diffusion together define enzymatic depolymerization and microbial substrate uptake rates, Soil Biology and Biochemistry, 139, 107 624, https://doi.org/10.1016/j.soilbio.2019.107624, 2019.

Tang, J. Y. and Riley, W. J.: A total quasi-steady-state formulation of substrate uptake kinetics in complex networks and an example application to microbial litter decomposition, Biogeosciences, 10, 8329–8351, https://doi.org/10.5194/bg-10-8329-2013, 2013.

Tao, F., Huang, Y., Hungate, B. A., Manzoni, S., Frey, S. D., Schmidt, M. W. I., Reichstein, M., Carvalhais, N., Ciais, P., Jiang, L., Lehmann, J., Wang, Y.-P., Houlton, B. Z., Ahrens, B., Mishra, U., Hugelius, G., Hocking, T. D., Lu, X., Shi, Z., Viatkin, K., Vargas, R., Yigini, Y., Omuto, C., Malik, A. A., Peralta, G., Cuevas-Corona, R., Di Paolo, L. E., Luotto, I., Liao, C., Liang, Y.-S., Saynes, V. S., Huang, X., and Luo, Y.: Microbial carbon use efficiency promotes global soil carbon storage, Nature, 618, 981–985, https://doi.org/10.1038/s41586-023-06042-3, 2023.

Tao, F., Houlton, B. Z., Frey, S. D., Lehmann, J., Manzoni, S., Huang, Y., Jiang, L., Mishra, U., Hungate, B. A., Schmidt, M. W. I., Reichstein, M., Carvalhais, N., Ciais, P., Wang, Y.-P., Ahrens, B., Hugelius, G., Hocking, T. D., Lu, X., Shi, Z., Viatkin, K., Vargas, R., Yigini, Y.,

Omuto, C., Malik, A. A., Peralta, G., Cuevas-Corona, R., Di Paolo, L. E., Luotto, I., Liao, C., Liang, Y.-S., Saynes, V. S., Huang, X., and Luo, Y.: Reply to: Model uncertainty obscures major driver of soil carbon, Nature, 627, E4–E6, https://doi.org/10.1038/s41586-023-07000-9, 2024a.

Tao, F., Houlton, B. Z., Huang, Y., Wang, Y.-P., Manzoni, S., Ahrens, B., Mishra, U., Jiang, L., Huang, X., and Luo, Y.: Convergence in simulating global soil organic carbon by structurally different models after data assimilation, Global Change Biology, 30, e17297, https://doi.org/https://doi.org/10.1111/gcb.17297, 2024b.

The MathWorks Inc.: MATLAB version: 9.13.0.2105380 (R2022b) Update 2, Tech. rep., Natick, Massachusetts, United States, https://www.mathworks.com, 2022.

Todd-Brown, K. E. O., Randerson, J. T., Post, W. M., Hoffman, F. M., Tarnocai, C., Schuur, E. A. G., and Allison, S. D.: Causes of variation in soil carbon simulations from CMIP5 Earth system models and comparison with observations, Biogeosciences, 10, 1717–1736, https://doi.org/10.5194/bg-10-1717-2013, 2013.

Varney, R. M., Chadburn, S. E., Burke, E. J., and Cox, P. M.: Evaluation of soil carbon simulation in CMIP6 Earth system models, Biogeosciences, 19, 4671–4704, https://doi.org/10.5194/bg-19-4671-2022, 2022.

Wang, G., Post, W. M., and Mayes, M. A.: Development of microbial-enzyme-mediated decomposition model parameters through steady-state and dynamic analyses, Ecological Applications, 23, 255–272, https://doi.org/10.1890/12-0681.1, 2013.

Wang, G., Jagadamma, S., Mayes, M. A., Schadt, C. W., Megan Steinweg, J., Gu, L., and Post, W. M.: Microbial dormancy improves development and experimental validation of ecosystem model, The ISME Journal, 9, 226–237, https://doi.org/10.1038/ismej.2014.120, 2015.

Wang, Y. P., Chen, B. C., Wieder, W. R., Leite, M., Medlyn, B. E., Rasmussen, M., Smith, M. J., Agusto, F. B., Hoffman, F., and Luo, Y. Q.: Oscillatory behavior of two nonlinear microbial models of soil carbon decomposition, Biogeosciences, 11, 1817–1831, https://doi.org/10.5194/bg-11-1817-2014, 2014.

Wang, Y. P., Jiang, J., Chen-Charpentier, B., Agusto, F. B., Hastings, A., Hoffman, F., Rasmussen, M., Smith, M. J., Todd-Brown, K., Wang, Y., Xu, X., and Luo, Y. Q.: Responses of two nonlinear microbial models to warming and increased carbon input, Biogeosciences, 13, 887–902, https://doi.org/10.5194/bg-13-887-2016, 2016.

Wieder, W. R., Bonan, G. B., and Allison, S. D.: Global soil carbon projections are improved by modelling microbial processes, Nature Climate Change, 3, 909–912, https://doi.org/10.1038/nclimate1951, 2013.

Wieder, W. R., Grandy, A. S., Kallenbach, C. M., and Bonan, G. B.: Integrating microbial physiology and physio-chemical principles in soils with the MIcrobial-MIneral Carbon Stabilization (MIMICS) model, Biogeosciences, 11, 3899–3917, https://doi.org/10.5194/bg-11-3899-2014, 2014.

Wieder, W. R., Allison, S. D., Davidson, E. A., Georgiou, K., Hararuk, O., He, Y., Hopkins, F., Luo, Y., Smith, M. J., Sulman, B., Todd-Brown, K., Wang, Y., Xia, J., and Xu, X.: Explicitly representing soil microbial processes in Earth system models, Global Biogeochemical Cycles, 29, 1782–1800, https://doi.org/10.1002/2015GB005188, 2015.

Wieder, W. R., Hartman, M. D., Sulman, B. N., Wang, Y., Koven, C. D., and Bonan, G. B.: Carbon cycle confidence and uncertainty: Exploring variation among soil biogeochemical models, Global Change Biology, 24, 1563–1579, https://doi.org/10.1111/gcb.13979, 2018.

Wilson, C. H. and Gerber, S.: Theoretical insights from upscaling Michaelis–Menten microbial dynamics in biogeochemical models: a dimensionless approach, Biogeosciences, 18, 5669–5679, https://doi.org/10.5194/bg-18-5669-2021, 2021.