Exploring the link between cation exchange capacity and magnetic susceptibility

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Abstract. This study explores the relationship between soil magnetic susceptibility (κ) and cation exchange capacity (*CEC*) across diverse European soils, aiming to enhance pedotransfer functions (PTFs) for soil *CEC* using near-surface electromagnetic geophysics. We hypothesize that soil κ , can improve the prediction of *CEC* by reflecting the soil's mineralogical composition, particularly in sandy soils.

We collected data from 49 soil samples in vertical profiles across Belgium, the Netherlands, and Serbia, including κ in-situ conditions (κ_*), low and high frequency κ in the laboratory, in-site electrical conductivity (σ), iron content, soil texture, humus content, bulk density, water content, water pH, and *CEC*. We used these properties as features to develop univariable and multivariable (in pairs) polynomial regressions to predict *CEC* for sandy and clayey soils.

Results indicate that κ_* significantly improves *CEC* predictions in sandy soils, independent of clay content, with a combined κ_* - σ model achieving the highest predictive performance ($R^2 = 0.94$). In contrast, laboratory-measured κ was less effective, likely due to sample disturbance.

This study presents a novel *CEC* PTF based on σ and κ_* , offering a rapid, cost-effective method for estimating *CEC* in field conditions. While our findings underscore the value of integrating geophysical measurements into soil characterization, further research is needed to refine the κ - *CEC* relationship and develop a more widely applicable model.

25 1 Introduction

Modern strategies for soil characterization are crucial for addressing the global challenges of soil degradation and pollution. Near-surface electromagnetic geophysics, in particular, facilitates rapid quantitative assessment of soils, offering insights

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into subsurface electrical conductivity (σ), dielectric permittivity (ε), and magnetic susceptibility (κ) (Garré et al., 2022; Romero-Ruiz et al., 2018). Data collected on these electromagnetic properties can be used for direct qualitative soil survey interpretations or for more comprehensive quantitative analyses involving pedophysical models and pedotransfer functions.

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Pedophysical models (PMs) link depth-specific geophysical properties with common soil properties. The need for developing such models is growing due to the demand for high-precision soil characterization (Romero-Ruiz et al., 2018; Verhegge et al., 2021; Wunderlich et al., 2013). Pedotransfer functions (PTFs) are models used to predict soil properties that are typically costly to obtain and are therefore determined less frequently than soil attributes that can be characterized more effectively (Van Looy et al., 2017).

An important indicator of soil health and fertility, which is also crucial in most PMs and PTFs, is the cation exchange capacity (CEC) (Glover, 2015; Mendoza Veirana et al., 2023). CEC, which refers to a soil's capacity to retain and exchange positively charged ions (Khaledian et al., 2017), is highly correlated with the soil clay content due to a larger colloid surface for particle exchanges. Furthermore, as it is influenced by the soil's physical (e.g., texture), chemical (e.g., pH, mineralogy), and biological properties (e.g., organic matter); thus, CEC integrates aspects from all three main indicator groups commonly used to assess soil quality (Khaledian et al., 2017). Several PMs for soil σ prediction include CEC due to the significant influence of free charges on σ , despite the contribution of other properties such as water content (θ) and salinity or bulk soil σ (Glover, 2015). Soil charges can be either permanent or variable, depending primarily on soil pH (Chapman, 1965; Sumner & Miller, 2018). The relationship between clay content and CEC can be highly variable due to its dependence on clay mineralogy (ranging from 3-15 meg/100g for kaolinite to 100-150 meg/100g for vermiculite), and the relative proportion of variable and permanent CEC varies among clay minerals (Miller, 1970; Seybold et al., 2005). To standardize CEC measurements under varying soil conditions, it is common to use the CEC in neutral pH conditions (=7) CEC 7. However, conventional analytical methods for measuring CEC, such as the sodium saturation method, are time-consuming and expensive (Busenberg & Clemency, 1973). Due to the critical importance of CEC, its measurement cost, and its correlation with other soil properties, numerous PTFs (Khaledian et al., 2017) and worldwide hybrid models (Poggio et al., 2021) for CEC7 have been developed. Commonly, CEC PTFs are expressed in function of clay content and humus, and less frequently pH and soil depth (Khaledian et al., 2017; Seybold et al., 2005). While commonly CEC PTFs are multivariate polynomial regressions, machine learning methods are used in the last decade when large datasets are available, as artificial neural networks (Ghorbani et al., 2015), and genetic expression programming and multivariate adaptive regression splines (Emangolizadeh et al., 2015). However, when working with small datasets, polynomial regressions are often preferred. Additionally, results have shown that σ and soil κ are independent (Maier et al., 2006), even though they generally correlate well with CEC.

Soil magnetic susceptibility has been correlated positively with *CEC* in studies focusing on soil type identification (Mello et al., 2020) (Pearson's correlation 0.4), soil characterization (Siqueira et al., 2010) (Pearson's correlation 0.68), paleoclimatic

reconstruction (Maher, 1998) (Pearson's correlation 0.95 for Podsol and 0.73 for Cambisol samples), and electromagnetic induction applications (McLachlan et al., 2022). κ describes a material's ability to become magnetized when subjected to an external magnetic field. It quantifies the degree of magnetization induced in the substance relative to the strength of the applied magnetic field. The composition of the parent material, and consequently the mineralogy of the rocks and sediments that formed the soils, are the main parameters influencing soil magnetic properties (Jordanova, 2017). Formally, κ is defined
as κ = H/M, where M is the induced magnetization of the material and H is the applied magnetic field. Soil and sediments κ measurements are widely used to detect the presence of pedogenic ferrimagnetic minerals (Dearing et al., 1996).

Soil clay content and soil κ are correlated positively due to the presence of ferrimagnetic minerals (such as maghemite) in the clay fraction, originating either from the parent material or through pedogenesis (Mello et al., 2020) (Pearson's correlation 0.26). Consequently, it has been suggested that the observed correlation between κ and CEC is actually due to their mutual correlation with clay content, indicating that there may not be a direct effect of κ on CEC (Mello et al., 2020).

To the best of our knowledge, the κ - *CEC* relationship has not been studied beyond the site level (Siqueira et al., 2010). This limited scope represents a significant research gap, as the broader applicability of κ for *CEC* prediction remain largely unexplored across diverse soil types and conditions.

The main hypothesis is that soil κ can support characterizing soil mineralogy, which also influences the permanent component of *CEC*. Therefore soil κ may significantly enhance the accuracy of *CEC* PTFs. This study directly addresses the identified gap by systematically examining the κ – *CEC* relationship using a new comprehensive dataset. The potential to develop more robust, widely applicable *CEC* PTFs underscores the significance of this work, with implications for sustainable land management, precision agriculture, and environmental monitoring.

To improve predictions of field *CEC* by integrating soil κ , this study focuses on developing and test uni- and multivariate polynomial PTFs based on data of diverse soil types sampled in Europe. In addition, we explore soil κ measured in-situ and in laboratory at different frequencies to give insights into the κ -*CEC* relationship and investigate how clay content affects the relationship between κ and *CEC*. While the methodology of this study focusses on soil and geophysical data collection, data analysis and model development, delving into the underlying physicochemical mechanisms of soil mineralogy that would link κ and *CEC* are out of our scope but is highlighted as an important direction for future research. To ensure transparency and reproducibility, all the collected data and developed code for this work is publicly available in an open source Python software: Mendoza Veirana, 2024.

2 Methods

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2.1 Study area, field measurements, and soil analysis

From 8 sites in Belgium, the Netherlands and Serbia, 49 soil samples were collected across a wide range of USDA soil textures from sand to clay, and WRB (IUSS Working Group WRB, 2022) soil types (Figure 1 and Table 1, see also Mendoza Veirana et al. (2023)). Specifically, 6 sites in Belgium contributed 38 samples, one site in the Netherlands contributed 6 samples, and one site in Serbia contributed 5 samples. This distribution ensures representation of diverse soil types and textures across the three countries. At each site, test pits were dug to identify and sample different soil horizons. For each soil horizon and within a vertical soil profile, soil field κ was measured (κ) (5 to 11 measurements per site, 49 in total) using a kappa meter SM30 (ZH Instruments, Brno, Czech Republic) at 8 kHz. The sensor measures soil κ with a penetration depth of 2 cm and a sensitivity of 10^{-7} SI units. First, the sensor was placed against the soil's profile wall for a measurement, followed by an additional measurement taken in open air away from the profile to obtain a reference zero κ value for measurement calibration (ZH instruments, 2022). Additionally, a HydraProbe sensor (Stevens, Water Monitoring Systems) was employed to measure σ along the profile wall. The correction proposed by Logsdon et al. (2010) was applied to improve the quality of these readings.

Undisturbed soil samples (100 cm³) were collected manually, by pushing standard steel rings horizontally into the soil profile wall at the same locations where κ_* was measured. After the cores were weighed fresh and oven-dried for 24 h at 105 °C, volumetric water content (θ) was calculated from the water-mass loss divided by the core volume, and bulk density (b_d) from the oven-dry mass divided by the same volume (Grossman and Reinsch, 2002).

Disturbed soil samples of about 250 g were collected around the undisturbed samples. They were air-dried, homogenized in an agate mortar, and sieved using a 2 mm mesh for determination of texture, and chemical and magnetic soil properties. Clay, silt and sand content (denoted as Clay, Silt, Sand, respectively, expressed in %) was measured following the pipette method (NF X31-107, 2003) after sieving at 2 mm, content of humus, *CEC* was determined by CoHex method (Ciesielski et al., 1997a, 1997b).

												CEC
			Depth	Sand	Clay	κ_*	κ_{lf}		Fe	Humus	σ	[meq/
	Samples	Soil type	[cm]	[%]	[%]	[10-5]	[10-5]	κ_{fd} [-]	[ppm]	[%]	[mS/m]	100g]
A	5	Luvisols	[4,106]	[9,16]	[9,13]	[14,32]	[14,40]	[5.2,8.1]	[14,21]	[0.1,2.3]	[12,35]	[6,9]
DREN	5		[72,252]	[18,35]	[27,39]	[47,66]	[51,70]	[6.2,9.1]	[31,38]	[0.6,1.6]	[30,53]	[20,25]
Е	5	Cambisols	[20,110]	[24,35]	[20,25]	[7.2,14]	[8,15]	[3.5,5.5]	[19,25]	[0.8,2.6]	[38,50]	[9,12]
EH2	5	Phaeozems	[20,94]	[10,54]	[17,53]	[12,20]	[13,38]	[3.2,6.7]	[42,50]	[0.3,5.7]	[55,66]	[16,39]
НОЕКЕ	11	Cambisols	[28,258]	[9,45]	[16,32]	[4.5,116]	[6.8,127]	[4.4,8.1]	[17,34]	[0.5,11]	[27,59]	[8,30]
P	7	Retisols	[32,144]	[42,80]	[8,11]	[2.6,12]	[4.5,16]	[2.9,8.5]	[5.9,12]	[0,2]	[8,17]	[1.6,11]
S	5	Arenosols	[28,130]	[83,93]	[5,7]	[3,20]	[2.9,73]	[3.3,9.2]	[3.6,16]	[0,2]	[11,29]	[2,5]
VALTHE	6	Podzol	[10,60]	[91,95]	[3,4]	[0.8,12]	[1.2,19]	[3.9,9.7]	[1,2.5]	[0,2.2]	[0.5,1]	[1.6]
All	49		[4,252]	[9,95]	[3,53]	[0.8,116]	[1.2,127]	[2.9,9.7]	[1,50]	[0,11]	[0,66]	[1.6,39]

Table 1 Minimum and maximum intervals of soil and magnetic properties for each explored site. Data ranges reflect the diversity of soil types and conditions across various sites.

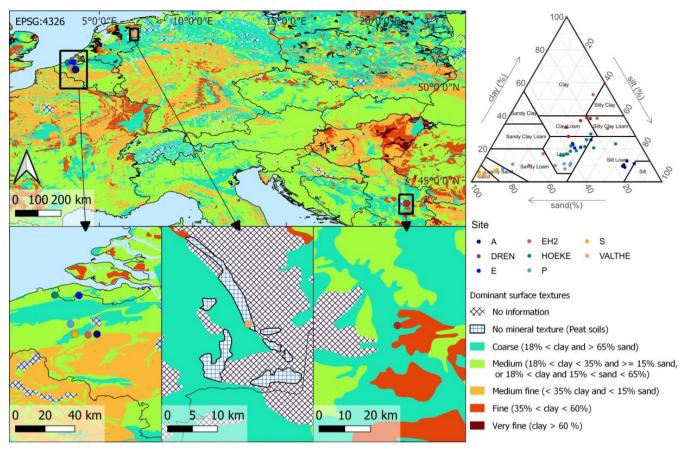


Figure 1 Locations of the study sites. Background shows dominant surface texture (European Soil Database v2.0, 2004). Colors represent the texture of the sites: sandy in yellowish, silty in blueish and clayey in reddish. The United States Department of Agriculture (USDA) texture triangle showing the particle size distribution categorized by sampling site. The samples presented in Table 1 are represented by triangles (adapted from (Mendoza Veirana et al., 2023)).

Magnetic susceptibility κ was measured using a Kappabridge MKF1-FA (AGICO Instruments, Brno, Czech Republic), in addition to the field κ measurements. Prior to these laboratory measurements, corrections were made to eliminate the influence of the diamagnetic sample holder. Samples were placed in 10 cm³ plastic holders, and the in-phase κ was recorded at both low frequency (κ_{lf}) and high frequency (κ_{hf}) (976 Hz and 15616 Hz, respectively). Thus, the percentage frequency dependent magnetic susceptibility (κ_{fd}) was calculated as:

$$\kappa_{fd} = \frac{\kappa_{lf} - \kappa_{hf}}{\kappa_{lf}} \times 100 \text{ [\%]}$$

Equation 1

Additionally, the absolute difference ($\kappa_{fd\,abs} = \kappa_{lf} - \kappa_{hf}$) was calculated. Such measurements are used to detect the presence of superparamagnetic ferrimagnetic minerals occurring as ultrafine (<0.03 µm) crystals (mostly ultrafine magnetite and maghemite) produced largely by pedogenic biochemical processes in soil (Dearing, 1994). Samples where ultrafine minerals are present will show increased frequency dependent magnetic susceptibility; samples without such minerals will show identical κ values at the two frequencies.

To support magnetic observations, elemental analyses were performed to evaluate the total iron content concentration (Fe) of all samples through X-ray fluorescence (XRF, Niton XL3t GOLDD+, Thermo Fisher Scientific Inc., USA) on samples that passed a 0.5 mm sieve and placed in capsules covered with 4⁻⁶ m cellophane. Three consecutive XRF measurements were performed and averaged for each sample.

2.2 Model development

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The absence of previous attempts at developing a *CEC* PTF using soil κ data that can be generalized beyond site-specific highlights the importance of thorough data exploration. We chose to build polynomial models due to their interpretability, simplicity, and the use of only one tuning parameter (polynomial degree), which is suitable given the relatively small dataset (n=49).

Field and laboratory measured soil properties were used as features for predicting *CEC* (target variable), these are: soil depth, water pH, *Humus*, *Clay*, *Silt*, and *Sand*, b_d , σ , κ_* , κ_{lf} , κ_{fd} , κ_{fdabs} , and *Fe*. All (13) features were used to develop univariable polynomial regressions, and multivariable models were created by combining features in pairs, resulting in 91 feature combinations. The top four combinations in terms of test performance were compared to the standard combination of *Clay* and *Humus* content, also, single features were considered (*Clay*, σ , and κ_*).

Since model performance was largely dependent on clay content, the samples were divided into sandy (n=25) and clayey (n=24) groups, using the median clay content (16.1%) as a threshold. Both input datasets were split randomly into training (70%) and testing (30%) subsets, without consideration of any soil characteristics. This approach ensured that samples from different sites, soil horizons, and physicochemical properties were mixed during data splitting. To further ensure an unbiased model evaluation, the training and testing process was repeated 100 times. The best polynomial degree (linear or quadratic) was determined by the highest median of the R^2 test scores over the 100 repetitions (Tibshirani et al., 2001). Finally, model implementation was performed after tuning and feature selection using all the samples of each subset.

2.3 Statistical analysis

Predicting *CEC* based on soil properties, particularly focusing on magnetic characteristics is a multivariate problem. Commonly, many variables are linearly correlated with *CEC*, such as *Clay*, κ , *Fe*, *Humus*, and σ . The challenge lies in distinguishing between independent and masked effects. Importantly, the positive correlation between κ and *CEC* may be due to the strong correlation between clay content and both *CEC* and κ (Mello et al., 2020).

To address this, we quantified the independent correlation between κ and CEC, irrespective of the effects of clay by calculating the partial correlation, that is the correlation between the residuals of the linear fitting of the covariable (Clay) with the variables (κ and CEC).

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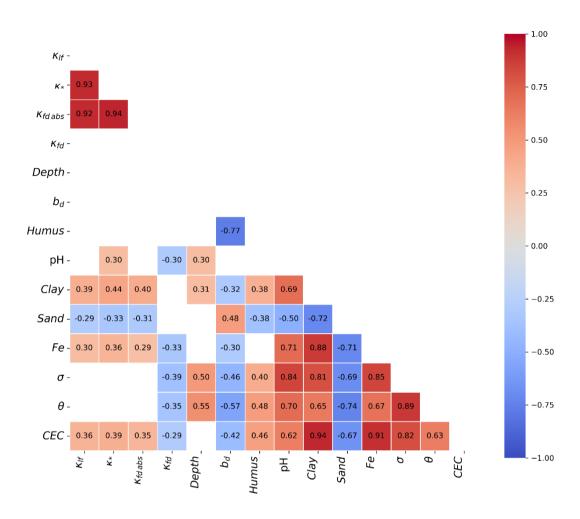
3 Results and discussion

3.1 Data exploration

A general variable exploration analysis is presented in this section, which is fundamental for the model development in the next section. Spearman's rank correlations between all the features mentioned and target can be seen in

Figure 2. As expected, soil κ is less correlated to *CEC* than common soil properties such as *Clay* and *Sand*, water pH and θ , while Fe and σ correlate strongly with *CEC*. Additionally, consistent with the findings of Mello et al. (2020) and Ayoubi et al. (2018), there is a positive correlation between *Clay* and κ (both κ_* and κ_{lf}). Also in line with Maier et al. (2006), σ is not correlated significantly to κ . Conversely, *Sand* correlates negatively with both κ and Fe.

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Figure 2 Spearman rank correlation heatmap showing significant P-values≤0.005 for the 49 soil samples, missing correlations have P-values>0.005.

Comparing soil κ_* to κ_{lf} reveals a similar trend across the entire range of observations (10⁻⁵ to 10⁻³) (see Figure 3). This trend persists despite κ_* being measured in undisturbed soil structures with field bulk density, while κ_{lf} was obtained from repacked samples that do not preserve the field structure and density. Additionally, the measurement frequency for κ_* (8 kHz) differs from that of κ_{lf} (~1 kHz).

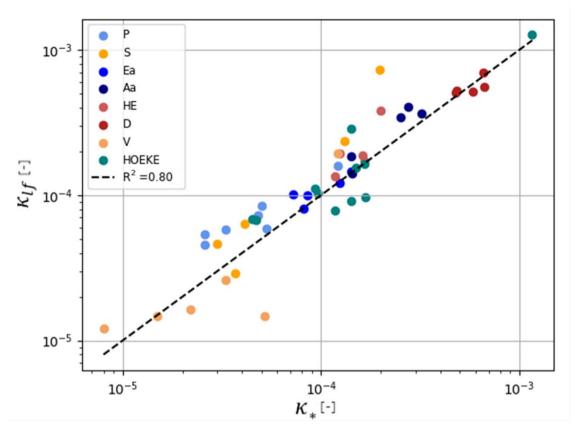


Figure 3 Logarithmic scatter plot showing the in-situ obtained magnetic susceptibility (κ_*) vs the laboratory observed low frequency magnetic susceptibility (κ_{lf}).

190 **3.2 CEC** modelling

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Model training and testing was performed in sandy and clayey groups independently. The predictors for CEC, with the best overall model performance, turned out to be highly dependent on the group. Notably, using σ and κ_* provided the best prediction results on the sandy group, with training and testing median R^2 values of 0.95 and 0.85, respectively (see Figure 4). This performance is significantly higher than that achieved with the commonly used features, such as Clay and Humus content, which had a median test R^2 =0.38. Additionally, the combination of Sand and pH performed equally well, followed closely by combinations of κ_* and κ_{lf} features.

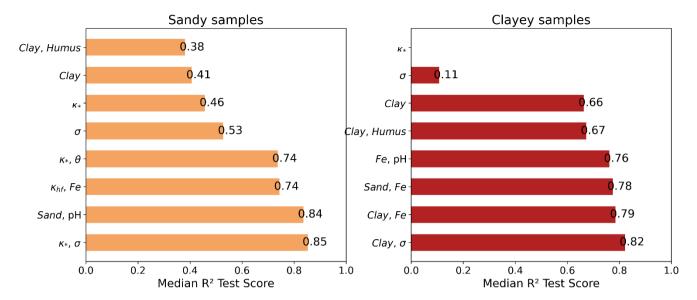


Figure 4 Horizontal bar plot showing test model performances of *CEC* prediction based on different features (vertical axis) for sandy and clayey samples. Features shown here are the top four in performance (bottom), and κ_* , σ , Clay, and the pair Clay, Humus.

The strong performance of σ and κ_* as predictors of *CEC* in sandy soils (median test $R^2 = 0.85$) is particularly noteworthy. σ is known to be influenced by several factors including soil water content, salinity, and the concentration of dissolved ions, which collectively can reflect the variable component of *CEC* (Glover, 2015). In sandy soils, which typically have lower water and nutrient retention capacities, σ can provide a dynamic measure of the available exchangeable cations at a given time. Concurrently, the strong predictive capacity of κ_* suggests it captures a different, yet complementary, aspect of *CEC*. In soils with low clay content, and therefore limited colloid surface area, the permanent component of *CEC* is likely more affected by minerals. The fact that κ_* , measured in-situ, performed better than laboratory κ suggests that the undisturbed soil structure and field conditions are crucial for this relationship, possibly reflecting the spatial arrangement and contact of these minerals within the soil matrix. Additionally, both σ and κ_* can be quickly measured in field conditions without the need for invasive sampling. Therefore, after implementing the best *CEC* PTF for sandy samples (see Figure 5):

$$\textit{CEC} = 1.233 + 14000 \cdot \kappa_* - 0.00861 \cdot \sigma - 5.91 \cdot 10^7 \cdot \kappa_*^2 + 1350 \cdot \kappa_* \cdot \sigma + 0.000624 \cdot \sigma^2; \\ R^2 = 0.94 \cdot 10^7 \cdot \kappa_*^2 + 1350 \cdot \kappa_* \cdot \sigma + 0.000624 \cdot \sigma^2; \\ R^2 = 0.94 \cdot 10^7 \cdot \kappa_*^2 + 1350 \cdot \kappa_* \cdot \sigma + 0.000624 \cdot \sigma^2; \\ R^2 = 0.94 \cdot 10^7 \cdot \kappa_*^2 + 1350 \cdot \kappa_* \cdot \sigma + 0.000624 \cdot \sigma^2; \\ R^2 = 0.94 \cdot 10^7 \cdot \kappa_*^2 + 1350 \cdot \kappa_* \cdot \sigma + 0.000624 \cdot \sigma^2; \\ R^2 = 0.94 \cdot 10^7 \cdot \kappa_*^2 + 1350 \cdot \kappa_*^2 \cdot \sigma + 0.000624 \cdot \sigma^2; \\ R^2 = 0.94 \cdot 10^7 \cdot \kappa_*^2 + 1350 \cdot \kappa_*^2 \cdot \sigma + 0.000624 \cdot \sigma^2; \\ R^2 = 0.94 \cdot 10^7 \cdot \kappa_*^2 + 1350 \cdot \kappa_*^2 \cdot \sigma + 0.000624 \cdot \sigma^2; \\ R^2 = 0.94 \cdot 10^7 \cdot \kappa_*^2 + 1350 \cdot \kappa_*^2 \cdot \sigma + 0.000624 \cdot \sigma^2; \\ R^2 = 0.94 \cdot 10^7 \cdot \kappa_*^2 + 1350 \cdot \kappa_*^2 \cdot \sigma + 0.000624 \cdot \sigma^2; \\ R^2 = 0.94 \cdot 10^7 \cdot \kappa_*^2 + 1350 \cdot \kappa_*^2 \cdot \sigma + 0.000624 \cdot \sigma^2; \\ R^2 = 0.94 \cdot 10^7 \cdot \kappa_*^2 \cdot \sigma + 0.000624 \cdot \sigma^2; \\ R^2 = 0.94 \cdot 10^7 \cdot \kappa_*^2 \cdot \sigma + 0.000624 \cdot \sigma^2; \\ R^2 = 0.94 \cdot 10^7 \cdot \kappa_*^2 \cdot \sigma + 0.000624 \cdot \sigma^2; \\ R^2 = 0.94 \cdot 10^7 \cdot \kappa_*^2 \cdot \sigma + 0.000624 \cdot \sigma^2; \\ R^2 = 0.94 \cdot 10^7 \cdot \kappa_*^2 \cdot \sigma + 0.000624 \cdot \sigma^2; \\ R^2 = 0.94 \cdot 10^7 \cdot \kappa_*^2 \cdot \sigma^2 \cdot \sigma^$$

Equation 2

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Where κ_* is unitless, σ is in mS/m and CEC in meq/100g.

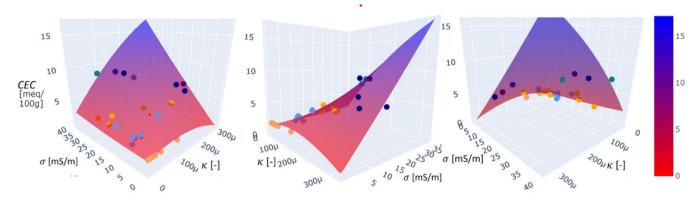


Figure 5 Implemented *CEC* PTF for sandy samples (clay<16.1) (Equation 2) with a R^2 =0.94. The model is colored in vertical axis from red to blue to visualize its shape. Colored dots represent the samples used in the sandy group, belonging to different sites that match the colors in Figure 1.

For clayey samples, σ and Clay features resulted in the best performance with a training and testing R^2 equal to 0.89 and 0.82, respectively. This result is in line with the literature since the link between CEC, σ and Clay is well documented (Glover, 2015; Wunderlich et al., 2013). For clayey samples, κ was not an outstanding feature, likely due to the influence of larger colloid surface that may not be effectively characterized by κ .

3.3 κ_* - CEC statistics

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The partial correlation between κ_* and CEC, while controlling for (removing the effect of) Clay, was found to be 0.61 for sandy samples, and -0.14 for clayey samples. This indicates that in sandy samples, κ_* is just partially influenced by Clay. For clayey samples, however, κ_* is heavily influenced by the soil's clay content, making the correlation between κ_* and CEC minor if the effect of Clay is removed. Consequently, predicting CEC using Clay alone is as effective as using both Clay and κ_* (median testing R^2 of 0.66 and 0.64, respectively), while κ_* alone is a poor predictor (see Figure 4).

235 4 Limitations and further directions

The current study, while providing novel insights, has several limitations that also point towards important future research directions.

Firstly, the main limitation of the analyzed results is related to the dataset size, although diverse in terms of European soil types, is relatively small. A larger sample size could improve the statistical relevance of the findings and improve the robustness and generalizability of the developed PTFs (Van Looy et al., 2017). Future work should aim to expand the database with more samples covering an even wider range of soil properties and parent materials.

Secondly, all collected samples come from non-tropical regions, where organic matter content and bacterial activity do not significantly influence soil κ . In contrast, these factors may contribute substantially to higher soil *CEC* in other environments (Seybold et al., 2005). Therefore, the results are valid for the sampled sites that belong to European soils, and applications to scenarios beyond this range of soils should be approached with caution.

Thirdly, a significant limitation is the lack of direct mineralogical analysis, especially for clay and iron oxide fractions. While κ offers an indirect proxy for ferrimagnetic mineralogy, detailed characterization (e.g., via X-ray diffraction) is needed for a causal understanding of the κ - *CEC* link. Identifying specific clay minerals (like kaolinite vs. smectite) and their abundance would clarify their *CEC* contributions and interactions with magnetic minerals. This is a crucial step to move beyond empirical correlations towards a process-based understanding

Fourthly, while field-measured κ proved useful, the reasons for its superiority over laboratory-measured κ_{lf} or κ_{fd} in the PTFs warrant further exploration. This could involve investigating the effects of soil structure, moisture content (which are preserved in in-situ κ measurements). A deeper understanding of how these factors influence different κ measurements could lead to optimized measurement strategies.

Finally, the model shown in Equation 2 is valid for samples with clay content between 2.9% to 16.1%, σ between 0.55 mS/m to 39 mS/m, κ*, between 8 to 320 μ, and CEC between 1.6 meq/100g to 8.7 meq/100g. As larger and more comprehensive datasets become available, exploring advanced modelling techniques, such as machine learning algorithms, may capture more complex, non-linear relationships. Assessing the scalability of the κ-CEC relationship from point measurements to field-scale predictions using proximal sensing platforms, for example, vehicle-mounted EMI sensors providing dense κ data (McLachlan et al., 2022), would be beneficial.

5 Conclusions

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For the first time, the link between soil κ and CEC has been explored using data that extends beyond the site level. By analyzing soil samples across Europe, encompassing a range of diverse soil physicochemical properties, we found that κ_* significantly contributes to predicting soil CEC, particularly in sandy samples, and this contribution is linearly independent of the soil clay content. Conversely, soil κ measured in the laboratory was less effective, likely due to the disturbance of soil structure and soil density.

Based on these findings, we proposed a novel PTF for *CEC* in sandy samples, with a R^2 of 0.94, based on σ and κ_* , which likely relate to the variable and permanent components of *CEC*, respectively. This PTF is valuable because both σ and κ_* are

quick and inexpensive to measure in the field, making it straightforward to predict CEC under field conditions. For instance,

270 it can be used to quickly assess the fertility of sandy soils across agricultural fields.

Further research, along with expanding the existing database, could enhance CEC modeling and provide deeper insights into

the κ_* - CEC relationship. These advances could help integrate independent geophysical properties such as σ and κ , to

quantify key soil properties like CEC, advancing a more holistic approach towards soil characterization.

275 Data availability

10.5281/zenodo.13971643

Code availability

https://github.com/orbit-ugent/kappa-CEC

280 Author contribution

PS, MV and JV designed the surveys and carried them out. MV developed the model code and performed the simulations

and contributed to statistical formal analysis along HG. JV, HG, PS and WC gave writing advice. MV prepared the

manuscript with contributions from all co-authors.

285 Competing interests

The authors declare that they have no conflict of interest.

Special issue statement

Agrogeophysics: illuminating soil's hidden dimensions

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