The contribution of transport emissions to ozone mixing ratios and methane lifetime in 2015 and 2050 in the Shared Socioeconomic Pathways (SSPs)

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Abstract. We quantify the contributions of emissions from the transport sector to tropospheric ozone and the hydroxyl radical (OH) by means of model simulations with a global chemistry-climate model equipped with a source attribution method. For the first time we applied a method which allows to quantify also contributions to OH which is invariant upon disaggregation or recombination and additive. Based on these quantified contributions, we analyse the ozone radiative forcings (RF) and methane

5 lifetime reductions attributable to emissions from the transport sectors. The contributions were analysed for each transport sector separately and for 2015 as well as for 2050 under the Shared Socioeconomic Pathways (SSPs) SSP1-1.9, SSP2-4.5, and SSP3-7.0. In line with previous publications using the source attribution approach, we quantify an ozone RF attributable to emissions from land transport, shipping, and aviation for the year 2015 of 121 mW m^{-2}, 60 mW m^{-2}, and 31 mW m^{-2}, respectively. At the same time, we diagnose a relative reduction in methane lifetime due to transport emissions of 14.3 % (land

10 transport), 8.5 % (shipping), and 3.8 % (aviation). These reductions are significantly larger as reported by previous studies due to the application of the source attribution method. Compared to 2015, only SSP1-1.9 shows a strong decrease in ozone RF and methane lifetime reduction attributable to the entire transport sector in 2050. For the projections SSP2-4.5, we find similar effects of the total transport sector as for 2015, while the effects in SSP3-7.0 increase compared to 2015. This small change of the effects for the two projections compared to 2015 is caused by two main factors. Firstly, aviation emissions are

15 projected to increase in SSP2-4.5 (increase of 107 %) and SSP3-7.0 (+86 %) compared to 2015, resulting in a projected ozone RF of 55 mW m^{-2} (+78 %) and 50 mW m^{-2} (+61 %) for the year 2050 from aviation emissions. Secondly, the non-linear effects of atmospheric chemistry in polluted regions such as Europe and North America lead to rather small reductions in ozone and OH in response to emission reductions, especially from land transport emissions. In addition, the increase in emissions from land transport in other parts of the world, particularly in South Asia, leads to an increased contribution of ozone and

20 OH. In particular, ozone formed by land transport emissions from South Asia causes a strong RF that partially offsets the reductions in Europe and North America. Moreover, our results show that besides the non-linear response, lack of international co-operations, as in the SSP3-7.0 projection, hinder mitigation of ground-level ozone.

1 Introduction

Emissions from the transport sector, i.e., land transport, shipping, and aviation, affect climate and air quality. The climate is influenced by emissions of both, the long-lived greenhouse gas carbon dioxide (CO_2), and short-lived climate forcers such as nitrogen oxides ($NO_x = NO + NO_2$) or particles (Szopa et al., 2021). In contrast to the effects of the CO_2 emissions on the climate, non-CO_2 effects heavily depend on the atmospheric conditions at their emission point and their further transport and chemical transformation in the atmosphere. Examples for such complex processes are the formation of tropospheric ozone or secondary aerosols (Seinfeld and Pandis, 2006) and aerosol-cloud interactions (Bellouin et al., 2020). Tropospheric ozone is a greenhouse gas, which also affects air quality and is harmful to humans and vegetation (Monks et al., 2015; Fleming et al., 2018; Mills et al., 2018). Moreover, tropospheric ozone levels are closely linked to the tropospheric oxidation capacity, which determines the lifetime of methane, which also affects climate (e.g. Stevenson et al., 2013).

Tropospheric ozone is formed by photochemical reactions from precursor emissions such as NO_x, carbon monoxide (CO), methane (CH_4) and Volatile Organic Compounds (VOC). The formation of tropospheric ozone depends on the meteorological conditions and on the ratio of NO_x to VOC concentrations in a non-linear way (e.g. Sillman, 1995; Archibald et al., 2020). Typically, two ozone formation regimes are distinguished: the so-called NO_x-limited regime, in which ozone production increases with increasing NO_x concentrations, and the VOC-limited regime, in which ozone production increases with increasing VOC-concentrations but decreases with increasing NO_x-concentrations. Due to this non-linear behaviour, a relatively large reduction in precursor emissions will often only result in a relatively small reduction in ozone concentration. Depending on the prevailing chemical regime, emission reductions can even lead to an increase of O_3 in specific regions. Examples of this phenomenon could be observed during the emission reductions as a consequence of the countermeasures against the COVID-19 pandemic: although NO_x emissions were greatly reduced, measurements and results of model simulations showed only a slight decrease in O_3 or, in some cases, regionally increasing O_3 values (e.g. Mertens et al., 2021; Matthias et al., 2021; Grange et al., 2021; Putero et al., 2023).

These non-linear processes in atmospheric chemistry need to be considered when designing mitigation measures for the transport sector. In this study, we therefore analyse the contributions of emissions from the three transport sectors – land transport (including road, rail and inland-shipping), aviation, and shipping – to the mixing ratios of ozone and the hydroxyl radical (OH) abundance in the troposphere. We analyse the present-day (2015) conditions and three future projections for 2050 and focus on non-linear changes of ozone and OH caused by emission changes. The year 2050 is chosen because it reflects a time horizon over which large changes of the transport sector are plausible, without exceedingly large speculations into the far future. For the future projections we selected three scenarios among the Shared Socioeconomic Pathways (SSPs Riahi et al., 2017): a 'green' scenario (SSP1; van Vuuren et al., 2017), a middle-of-the-road scenario (SSP2; Fricko et al., 2017) and a scenario of regional rivalry with low mitigative capacities (SSP3; Fujimori et al., 2017). We choose these scenarios as they span a broad range of possible future emission developments. In addition, the differences of the emission totals are large enough to yield different chemical regimes. Hence, we will not focus on the details of the socioeconomic backgrounds and technological pathways behind the chosen scenarios. Rather, we will use the scenarios as a tool to study the chemical effects

of transport emissions in different emission scenarios and to derive possible consequences for mitigation strategies. Here, we want to focus on three scientific questions:

– How large are the contributions of transport emissions to tropospheric ozone and what is their effect on climate?

60 – How do transport emissions influence the atmospheric oxidation capacity and with this the methane lifetime?

– How do these effects change under different future emission scenarios?

For the analyses, we apply the chemistry-climate model ECHAM/MESSy Atmospheric Chemistry (EMAC) which is equipped with a tagging technique for source attribution of ozone and ozone precursors including HO_x (Grewe et al., 2017; Rieger et al., 2018). We performed simulations for 2015 and for the three considered SSPs.

65 This manuscript adds to the findings of Mertens et al. (2018). In the aforementioned article, a source appointment method, considering NO_x and VOC at the same time, was applied for the first time to quantify the contribution of land transport and shipping emissions to tropospheric ozone. Here, we expand on those findings by adding the following important updates:

– We use a more recent emissions inventory including also consistent projections of the emissions for 2050.

– We integrate the aviation sector to compare the effects of all transport sectors in a consistent way using the same methods.

70 – We analyse the effects from land transport emissions of different regions.

– We apply a new tagging approach of OH and the hydroperoxyl radical HO_2 (Rieger et al., 2018) to also account for the contributions of transport emissions to OH and their effects on the methane lifetime.

Accordingly, the manuscript presents, to our knowledge, the first consistent analyses of the effects of transport emissions on tropospheric ozone using a source attribution approach considering all transport sectors and the effects on OH and the methane

75 lifetime.

In this manuscript, we first describe the model system, the model set-up, and the applied source attribution method in detail in Sect. 2, followed by a summary of the transport emissions in the SSPs in Sect. 3. Subsequently, we discuss the contribution of the emissions from the three transport sectors on tropospheric ozone (in Sect. 4) and the transport-related ozone radiative forcings (Sect. 5). In Sect. 6, we investigate the contributions of the transport emissions to OH and to the methane lifetime.

80 Finally, limitations of the study are discussed in Sect. 7.

2 Model description, emissions, and numerical experiments

In the following sections, the EMAC model, the source attribution model and the numerical experiments will be described in detail. In addition, we will discuss the transport emissions of the three considered SSP narratives.

2.1 The chemistry-climate model EMAC

85 For the present study, we applied the EMAC chemistry-climate model (Jöckel et al., 2006, 2010, 2016). EMAC uses the second version of the Modular Earth Submodel System (MESSy2) to link multi-institutional computer codes. The core atmospheric model is the 5th generation European Centre Hamburg general circulation model (ECHAM5, Roeckner et al., 2006). We utilized EMAC (ECHAM5 version 5.3.02, MESSy version d2.54.0.3-pre2.55-02-2077) in the T42L90MA-resolution, i.e., with a spherical truncation of T42 (corresponding to a quadratic Gaussian grid of approx. 2.8 by 2.8 degrees in latitude and 90 longitude) with 90 vertical hybrid pressure levels up to 0.01 hPa.

The simulation set-up is very similar to the simulation *RC1SD-base-10a* described in detail by Jöckel et al. (2016) alongside with an evaluation of the resulting model simulations. Therefore, we describe here only the most important details of the set-up and differences compared to *RC1SD-base-10a*:

- The applied gas phase mechanism in the submodel MECCA (Module Efficiently Calculating the Chemistry of the At-
95 mosphere, Sander et al., 2011) incorporates the chemistry of ozone, methane, and odd nitrogen. Alkanes and alkenes are considered up to C_4, while the oxidation of isoprene (C_5H_8) and some non-methane hydrocarbons (NMHCs) are described with the Mainz Isopren Mechanism, version 1 (von Kuhlmann et al., 2004). The mechanisms of MECCA is part of the Supplement.

- Heterogeneous reactions in the stratosphere (submodel MSBM, Jöckel et al., 2010) as well as aqueous phase chemistry
100 and scavenging (submodel SCAV, Tost et al., 2006) are included.

- Emissions of methane (CH_4), the long lived greenhouse-gases (e.g. CO_2, N_2O, etc.) and the ozone depleting substances (ODS) are not considered explicitly. Instead, pseudo-emissions of these emissions are calculated using the submodel TNUDGE (Kerkweg et al., 2006). TNUDGE relaxes mixing ratios in the lowest model layer towards prescribed mixing ratios using Newtonian relaxation (see also Jöckel et al., 2016).

105 - The emissions of lightning NO_x depend on the meteorology and are parameterised after Grewe et al. (2001), resulting in global total emissions of $\approx 4.5 \, \mathrm{Tg(N)\,a^{-1}}$. This value is well in the range of the best estimate given by Schumann and Huntrieser (2007) that is based on a review of observational an modelling approaches.

- Emissions of NO_x from soil and biogenic C_5H_8 emissions also depend on the meteorology and are calculated using the MESSy submodel ONEMIS (Kerkweg et al., 2006), following the parametrisations by Yienger and Levy (1995) for
110 soil-NO_x and Guenther et al. (1995) for biogenic C_5H_8.

2.2 Source attribution method

The TAGGING method has been applied in numerous studies to quantify the effects of the transport sectors on tropospheric ozone (Mertens, 2017; Mertens et al., 2018, 2020a, b, 2021; Maruhashi et al., 2024). The concept of the source attribution

methods (also known as tagging or labelling methods) fundamentally differs from the concept of perturbation methods (also
115 known as zero-out, brute force etc., Wang et al., 2009; Grewe et al., 2010; Clappier et al., 2017; Mertens et al., 2020b).

In this study, ozone and its precursors are attributed to 18 source categories which are listed in Table 1. Most of these
categories refer to specific emission sectors, but some others (for example production of ozone in the stratosphere) do not
refer to such sectors. For land transport and anthropogenic non-traffic emissions we separate five different geographical source
regions (see Sect. S8 in the Supplement for detailed definition). These are:

120   – Europe (EU)

    – North America (NA)

    – South Asia (SA)

    – East Asia (EA)

    – rest of the world (ROW)

125   For better readability, we denote absolute contributions as superscripted abbreviations. For example, O_3^{SHP} and OH^{SHP}
denotes the absolute contributions of shipping emissions to O_3 and OH mixing ratios, respectively. In addition to the abbreviations given in Table 1, we define the abbreviation *tra* indicating the contribution of land transport emissions from all regions
(i.e., the sum of land transport from Europe, North America, East Asia, South Asia and the rest of the world). Relative contributions are defined as the mixing ratios of a specific category (e.g., O_3^{SHP}) divided by total mixing ratio of the species (e.g., O_3).
130 Since the budget of the tagged species is closed, the relative contribution of the sum over all categories for each considered
species yields 100%. For example, the relative contribution of shipping emissions to ozone, denoted as $C^{SHP}(O_3)$, in percent,
is defined as:

$$C^{SHP}(O_3) = 100 \cdot \frac{O_3^{SHP}}{O_3}. \tag{1}$$

2.3 Radiative forcing calculations

135 Based on the diagnosed ozone contributions, we calculate the contribution of each category to the radiative flux (stratospherically adjusted) at the top of the atmosphere using multiple calls of the radiation routines (the submodel RAD, Dietmüller
et al., 2016). In the following, we refer to this contribution to the radiative flux as the "radiative forcing (denoted as RF) of
the respective emission category". As discussed in detail by Mertens et al. (2018) our definition of the RF does not take into
account the change of ozone since pre-industrial times (see Forster et al., 2021), as we perform no pre-industrial simulation.
140 Instead, the ozone RF in each simulation is decomposed into the respective contributions from the 18 source categories for
present day/future conditions (see also Dahlmann et al., 2011). The change of ozone attributed to natural sources between preindustrial and present-day caused by the increase of anthropogenic emissions and the respective effects on the ozone chemistry
are not considered (see detailed discussion in the Supplement of Mertens et al., 2018).

Table 1. Overview of the categories for the tagging source attribution applied in this study. Detailed definitions of the categories are given by Grewe et al. (2017). [1] The abbreviation IN stands for industry, but anthropogenic non traffic contains all non-traffic related activities i.e. industry, households, energy and agricultural including agricultural waste burning

name	abbreviation	description
anthropogenic non-traffic Europe	IN[1]-EU	anthropogenic emissions not related to transport from Europe
anthropogenic non-traffic North America	IN-NA	anthropogenic emissions not related to transport from North America
anthropogenic non-traffic South Asia	IN-SA	anthropogenic emissions not related to transport from South Asia
anthropogenic non-traffic East Asia	IN-EA	anthropogenic emissions not related to transport from East Asia
anthropogenic non-traffic ROW	IN-ROW	anthropogenic emissions not related to transport from the rest of the world (ROW)
land transport Europe	LT-EU	land transport emissions from Europe
land transport North America	LT-NA	land transport emissions from North America
land transport South Asia	LT-SA	land transport emissions from South Asia
land transport East Asia	LT-EA	land transport emissions from East Asia
land transport ROW	LT-ROW	land transport (IPCC codes 1A3b_c_e) emissions from rest of the world
land transport total	TRA	The category is not tagged individually, but calculated as the sum of LT-EU, LT-NA, LT-SA, LT-EA and LT-ROW
shipping	SHP	Global shipping emissions (IPCC code 1A3d)
aviation	AIR	Global aviation emissions
stratosphere	STR	downward transport from the stratosphere (more specifically, production of ozone by O_2)
biogenic	SOIL	biogenic emissions and soil-NO_x
biomass burning	BB	emissions from biomass burning
CH4	CH4	degradation of CH_4 as source of NMHCs
N2O	N2O	degradation of N_2O as source of NO_y
lightning	LIG	emissions of lightning NO_x

A detailed description of the procedure to calculate the RF is given by Mertens et al. (2018). Here we recall only the most
145 important aspects, exemplarily for the shipping (SHP) sector:

1. $\Delta O_3{}^{\mathrm{SHP}} = O_3 - O_3^{\mathrm{SHP}}$ is calculated;

2. Stratospheric adjusted radiative fluxes are calculated at every radiation time step (i.e. every third model time step) for O_3
   and $\Delta O_3{}^{\mathrm{SHP}}$.

3. In post-processing, the radiative contribution is calculated by subtracting the radiative fluxes (rflux) at the top of the
150 atmosphere:

$$RF_{O3}^{\mathrm{SHP}} = \mathtt{rflux}(O_3) - \mathtt{rflux}(\Delta O_3{}^{\mathrm{SHP}}), \tag{2}$$

2.4 Numerical experiments and emissions

For the present study, four different numerical experiments were performed as summarised in Table 2. Each simulation has been
conducted for the period $07/2012-12/2017$. The year 2012 is considered as spin-up and the years $2013-2017$ are analysed.
155 The chemical tracers, the tagging tracers, and water vapour are initialised in all simulations from a 1.5-years-long spin-up
simulation. This initialisation is the same in each simulation.

Table 2. Overview of the performed numerical simulations.

name	emission scenario	emission year	meteorology
PD	SSP2-4.5	2015	07/2012 – 12/2017
SSP1-1.9	SSP2-4.5	2050	07/2012 – 12/2017
SSP2-4.5	SSP1-1.9	2050	07/2012 – 12/2017
SSP3-7.0	SSP3-7.0	2050	07/2012 – 12/2017

EMAC is nudged by Newtonian relaxation of temperature, divergence, vorticity and the logarithm of surface pressure (Jöckel
et al., 2006) towards ERA-Interim (Dee et al., 2011) reanalysis data. Furthermore, the sea surface temperature and sea ice
coverage are prescribed as transient time-series from ERA-Interim.
160 To allow for identical, even binary identical, meteorological conditions in all simulations, the Quasi Chemistry Transport
Model mode (QCTM mode; Deckert et al., 2011) of EMAC is used. In this mode, mixing ratios of the radiatively active trace
gases are prescribed for the radiation calculations. Furthermore, mixing ratios of the chemical species are prescribed for the
processes responsible for the coupling of the chemistry and with the hydrological cycle. The prescribed mixing ratios are
monthly averages taken from a previous model simulation with the same set-up of the *PD* simulation of this study.
165 The simulations are performed as time slices, meaning that prescribed monthly emissions for 2015 or 2050 are repeated
every year. For each simulation, the emissions of short-lived chemical species (NO_x, NH_3, SO_2, CO and VOC) from biomass
burning, anthropogenic non-traffic sources, land transport, shipping, and aviation sectors are taken from the respective emission

scenarios. Emissions of all other species, including the longer lived greenhouse gases and CH_4, are kept at the 2015 level in all simulations.

170 Emissions of soil NO_x, biogenic isoprene, and lightning NO_x are calculated online, driven by the meteorological conditions, and are therefore identical in each simulation. The emissions of longer-lived species, which are nudged as described in Sect. 2.1, are taken from the SSP2-4.5 emission inventory for 2015. The transport emissions in the simulations are discussed in detail in Sect. 3, including an analysis of the NO_x emission totals. More information on the applied emissions (also for CO and VOC) are given in Sect. S1 in the Supplement.

3 Transport emissions in the SSP projections

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In our study, we consider three specific emission scenarios among the various combinations of SSP projections and forcing pathways (Gidden et al., 2019):

- SSP1-1.9: A version of the SSP1 narrative with a target radiative forcing from anthropogenic emissions of 1.9 W m^{-2} in 2100;

180 - SSP2-4.5: A version of the SSP2 narrative with a target radiative forcing from anthropogenic emissions of 4.5 W m^{-2} in 2100;

- SSP3-7.0: A version of the SSP3 narrative with a target radiative forcing from anthropogenic emissions of 7.0 W m^{-2} in 2100.

The narratives of these the SSPs can be summarized as following (O'Neill et al., 2017):

185 - SSP1 (Sustainability): SSP1 features a relatively low population growth with a medium-high economic growth per capita. There is moderate international trade, effective international cooperation and policy focuses on a sustainable development. Technological development is rapid with a strong change away from fossil fuels and towards efficiency and renewable energy.

- SSP2 (Middle of the road): SSP2 features a medium population growth and a medium economical growth per capita.
190 International cooperation is relatively weak and international trade is moderate. The environmental policy has a concern for local pollutants, but there is only a weak focus on sustainability in the policy orientation. Accordingly, there is only some investment in renewable energy and fossil fuel continues to play an important role. Technology developments is medium and distributed unevenly, with a slow transfer of technologies.

- SSP3 (Regional Rivalry): SSP3 features a high growth of the population in the group of high- and low-fertility countries
195 (see KC and Lutz, 2017, for more details) and a low growth in rich OECD (Organization for Economic Co-operation and Development) countries. The economic growth per capita is slow and there is a high inequality across countries. There is de-globalisation and a strongly constrained international trade. International cooperation's are weak and there is

a low priority for environmental issues. Technological developments are slow with a low technological change directed towards domestic energy sources.

200   The traffic emissions of air pollutants between the three projections differ strongly, as they depend on the mobility, global trade, as well as on the applied technologies. Overall, in 2050, passenger travel demand is largest in SSP2, followed by SSP1 and SSP3. Cars are an important transportation mode in all three SSPs, with the largest passenger travel demand in SSP2, followed by SSP3 and SSP1. In SSP1, public transport (bus and train) plays a more important role than in SSP2 and SSP3. Moreover, SSP1 features the lowest demand for aviation, followed by SSP3 and SSP2. High-speed trains become more 205 important as an alternative mode of transportation in SSP1 (see Fig. 5 in van Vuuren et al., 2017). The SSP1 has the overall lowest energy demand for the transport sector, followed by SSP2 and SSP3. Even though the energy demand of the transport sector in SSP1 is lower than in SSP3, SSP1 has the larger transport demand. This shows that a decoupling of energy demand and passenger demand is archiveable in a sustainable scenario. In all SSPs, liquid fuels dominate the overall energy demand for transportation in all scenarios until ~2040. Only in the SSP1, electrification of cars increase compared to present day already 210 before 2040. In this scenario, hydrogen also becomes an important energy carrier in the transport sector after 2050. In the SSP3 projection, even in 2050, only very little electrification is projected in the transport sector, while the SSP2 projection assumes a moderate amount of electrification (see Fig. S13 in the Supplement of Bauer et al., 2017). Generally, SSP1 features a maximum electrification of 75 % of total transport, while SSP2 and SSP3 feature 50 % and 10 % maximum electrification, respectively (Fricko et al., 2017).

215   In addition to the adaptation of alternative energy sources, such as electricity, in the transport sector, air pollutant emissions depend heavily on the regulations that are enacted. Different pollution control levels have been defined which are linked with the SSP narratives (Rao et al., 2017):

    – SSP1: Policy makers aim for much lower air pollution levels over the 21st century than the current legislation targets. There is a strong concern for negative effects of air pollutants on health and ecosystems. Moreover, medium and low 220   income countries catch up relatively quickly with industrialized countries in terms of pollution control.

    – SSP2: The general trend of emission reductions is continued worldwide and the pollution concentration targets become more ambitious with increasing income over the century. There is increasing concern for effects of pollutants on human health and the ecosystems. Due to diffusion of technologies and knowledge, developing countries reach levels of emission control relatively early compared to historical developments in OECD countries. Regions with large population densities 225   may not reach the targeted air pollution levels.

    – SSP3: Compared to SSP2, the implementation of pollution control is delayed. Moreover, due to less international cooperation and slow deployment of international laws, technologies are developed less quickly and trans-boundary pollution (e.g. hemispheric transport of air pollutants) is increasing.

    The integrated assessment models (IAMs) behind the SSPs do not apply complex air pollution control technologies. There-230 fore, the pollution control scenarios are implemented in a simplified approach based on trajectories of the emission factors (Rao

et al., 2017). The trajectories are based on two sets of emissions factors. The first set is „current legislation" (CLE), assuming efficient implementation of existing environmental legislation. The second set is the Maximum Technically Feasible Reduction (MFTR) set. The near-term achievement of MFTR is not considered feasible, but it serves as a baseline for what is ultimately possible given the current state of air pollution control technologies.

235   SSP2 assumes an implementation of CLE until 2030, while SSP1 (SSP3) assume a stronger (less strict) implementation of CLE until 2030, respectively. After 2030, the progress of the emission factors depends on the level of technological progress, i.e., the progress towards MFTR emission factors. In addition, depending on their economic developments, non-OECD countries can catch up with OECD countries with respect to the implementation of air pollution control faster (SSP1) or slower (SSP3). In all scenarios, the MFTR emission factors are stagnant in time, which means that there is no speculation about

240   possible future technologies better than MFTR. Thus, in countries and SSPs with high penetration of MFTR, this might be a conservative scenario (Rao et al., 2017).

In all SSP projections, the spatial distribution of the land transport emissions is taken from the EDGAR (Emissions Database for Global Atmospheric Research, Janssens-Maenhout et al., 2019) 4.3.2 emission inventory, aviation emissions are distributed as in CMIP5 (Coupled Model Intercomparison Project, phase 5, Taylor et al., 2012) and the shipping emissions are distributed

245   according to ECLIPSE (Evaluating the Climate and Air Quality Impacts of Short-Lived Pollutants, Stohl et al., 2015). The spatial distribution of land transport and aviation emissions are constant in time. Shipping emissions use projected changes of sulphur emissions due to changes in sulphur content and the definition of the Sulphur Emission Control Area (SECA). More details on the regridding process of the CMIP6 emissions are given by Feng et al. (2020). We apply the aviation emissions in their original form which have an error in the geographical distribution, as recently reported by Thor et al. (2023). The

250   corrected emissions would lead to an increase of the ozone RF by 7.6 %.

Figure 1 displays the NO_x emissions in the four simulations. Tables of the total emissions for NO_x, VOC and CO emissions are part of the Supplement (Sect. S1). Further analyses of the changes of the total emissions in the SSP projections are given by Righi et al. (2023).

In the SSP1-1.9 projection for 2050, anthropogenic NO_x emissions reduce by more than 70 % compared to 2015, which

255   is in agreement with the underlying narrative. Especially the emissions in the transport sector decrease strongly by around 90 % for most land transport regions, around 95 % for the shipping sector, and 64 % for the aviation sector. The reductions are caused by transformations of the transport system and by alternative fuels (e.g. electrification). Generally, SSP1 features a much larger share of public transport (e.g., buses and trains) in 2050 compared to the other SSPs (van Vuuren et al., 2017) and the aviation sector has a lower passenger demand compared to SSP2 and SSP3. In addition to this high share of public

260   transport and the quick shift towards alternative technologies, SSP1 also features the strongest air pollution control among the considered scenarios.

In the SSP2-4.5 projection for 2050, anthropogenic NO_x emissions decrease by around 25 %, but emission peaks are shifted regionally compared to 2015. As an example, land transport emissions decrease in most regions, but strongly increase over South Asia (around 60 %), and over Africa. The decrease is mainly caused by stronger air pollution control, especially in

265   the OECD regions, while the increase is partly caused by increases of the population in South Asia and parts of Africa. The
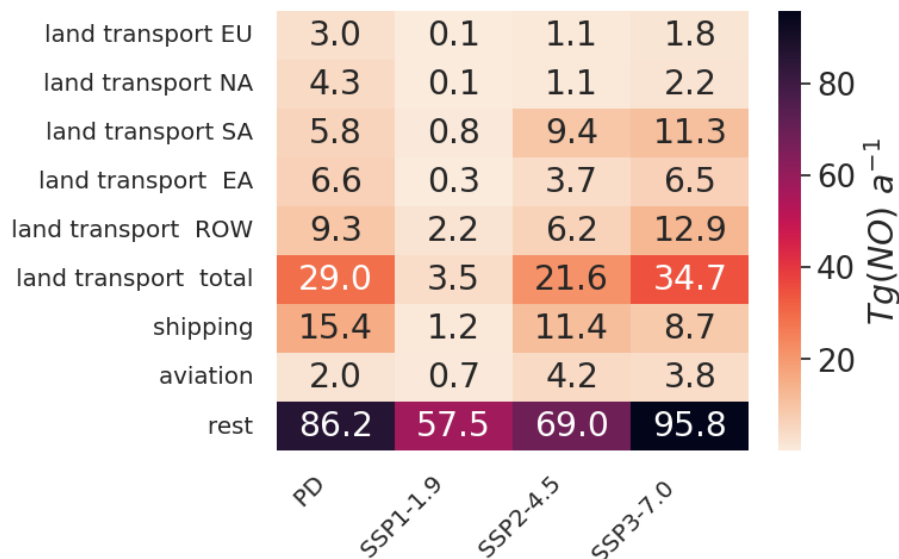
Figure 1. Comparison of the NO_x emissions (in $Tg(NO)\ a^{-1}$) between the four performed simulations for 2015 (PD) and 2050 (projections, see Sect. 2.4 for a detailed description). The land transport emissions are given for the different source regions and as global totals ("land transport total"). 'rest' indicates all other emission sources, including also soil- and lightning-NO_x emissions which are not included in the SSP emission data. Please not, that the "land transport total" emissions are only given for comparison, they are not emitted in addition to the regional land transport emissions.

aviation emissions are doubled compared to 2015, which is consistent with the increase of passenger travel demand in the aviation sector. The shipping emissions are reduced by around 30 %. These reductions mainly take place around Europe and the East coast of North America, while emissions increase, especially along the main shipping routes from China to Singapore. Most of the reductions are likely due to technological improvements (including increased air pollution control).

270   The SSP3-7.0 projection for 2050 exhibits around 12 % larger anthropogenic NO_x emissions compared to 2015. The regional changes, however, differ strongly. For example, the land transport NO_x emissions decrease in Europe and North America compared to 2015 by around 40 % and 50 % respectively, as a consequence of the stronger air pollution control (incl. technological advancements) in these regions. In South Asia, the NO_x emissions from land transport increase by almost 100 % and in the rest of the world by around 40 %. One reason for the particularly strong increase in South Asia is the strong growth of

275   population, in particular in India, in SSP2 and SSP3 in 2050 compared to 2015 (KC and Lutz, 2017). In addition, air pollution control is limited in SSP3. The shipping emissions for 2050 are lower in SSP3 compared to SSP2, even though the air pollution control is less strict in SSP3 compared to SSP2. This is likely caused by the strong decrease of international trade between the regions in SSP3 (Fujimori et al., 2017; Bauer et al., 2017).

4 Contributions to ground-level ozone

280 We first analyse the general changes in ground-level ozone for the three projections compared to 2015 (Sect. 4.1), before we analyse the individual transportation sectors in detail in Sect. 4.2 to Sect. 4.4. In these analyses the contributions to ground-level ozone (land transport and shipping) and zonally averaged contributions to ozone (aviation) are analysed. Moreover, we quantify the trends of the absolute contributions from land transport and shipping emissions to ground-level ozone for specific geographical regions. For the aviation sector we focus on relative contributions as zonal means and the trends of the contribution

285 of aviation ozone to the tropospheric ozone column for main flight regions. Moreover, we quantify the changes of the ozone burden caused by the changes of the emissions in Sect. 4.5.

4.1 Changes of ground-level ozone

Figure 2 shows the ground-level ozone mixing ratios for 2015 and the changes of the ground-level ozone mixing ratios compared to 2015 for the three projections in 2050. Figures of the ground-level ozone mixing ratios for the projections are part of

290 the Supplement (Sect. S2).

In 2050, ozone decreases in the SSP2-4.5 projection compared to 2015 in most regions by $2-6$ nmol mol^{-1}. However, ozone increases slightly over Central Africa, around India, along the east coast of China and in Central Europe. Especially in Central Europe and China, the increase is due to decreased ozone titration caused by decreasing NO_x emissions. For the SSP3-7.0 projection, ozone increases strongly in 2050 compared to 2015. The increase reaches up to 10 nmol mol^{-1} and

295 affects especially West Africa, the Arabian Peninsula, India, and South-East Asia. In contrast to this, ground-level ozone decreases strongly in 2050 for the SSP1-1.9 projection compared to 2015. In most regions the decrease is in the range of $10-15$ nmol mol^{-1}, and exceeds 20 nmol mol^{-1} on the Arabian Peninsula. The general changes of ground-level ozone for the three projections and regional features, such as the strong increase of ozone over Asia in SSP3-7.0, are in agreement with analysis of CMIP6 simulation results by Turnock et al. (2020). The magnitude of the ozone changes differ, but the CMIP6

300 models included the effects of changing meteorology and climate, while we use the same meteorology in all simulations. Therefore, we expect differences in the magnitudes of the ozone changes.

4.2 Land transport

Figure 3 shows the relative contributions of land transport emissions to ground-level ozone as simulated in the four experiments. In addition, the absolute contribution for 2015 and the changes in the absolute contribution for the various regions are shown

305 as an average over specific regions in Fig. 4. These regions are: Europe, North America, East Asia, South Asia, Southeast Asia and South America (see Appendix Sect. S9.1 for a detailed definition). Additional figures depicting absolute contributions at ground-level and zonal zonal means are part of the supplement (Sect. S3 and S4 therein).

Generally, the absolute and relative contributions over North America, Europe, and the Middle East for present-day conditions (simulation PD) are comparable to previous studies (see Mertens et al., 2018, and detailed tables therein). The contri-

310 butions for present-day conditions over India and parts of South-East Asia, however, are larger compared to previous studies.
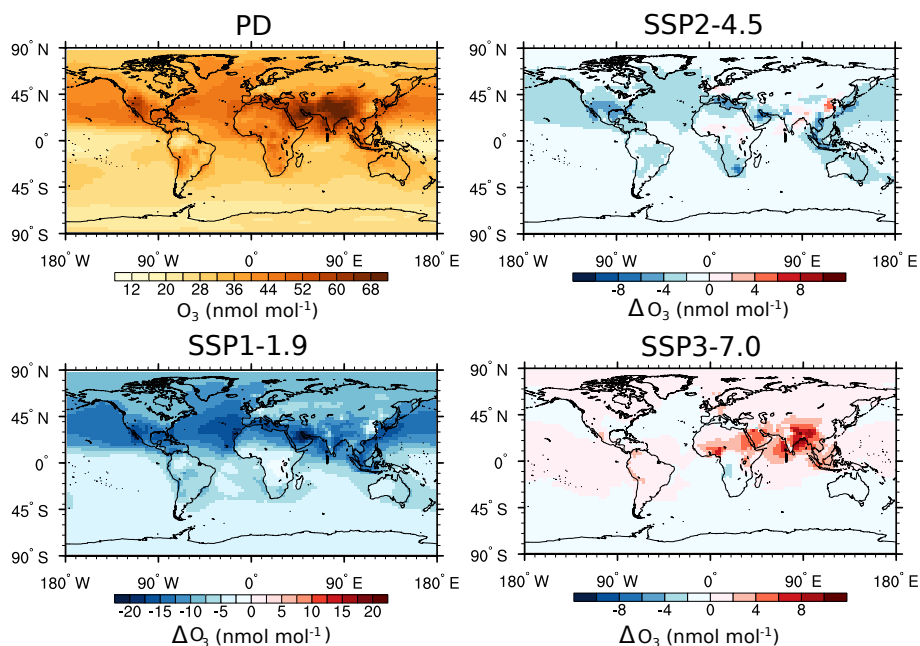
Figure 2. Ozone mixing ratios (5 year average) at ground-level for the PD simulation and differences (ΔO_3, '2050 minus 2015') for SSP1-1.9, SSP2-4.5, and SSP3-7.0. Please note that SSP1-1.9 has a different color scale than the two other simulations.

This increase is mainly caused by the larger emissions in these regions in the Community Emissions Data System (CEDS) emission inventory compared to other emission inventories (e.g. Hoesly et al., 2018; McDuffie et al., 2020). Moreover, a more recent emission year (2015) has been used here.

For the SSP2-4.5 projection in 2050, the relative contributions of emissions from land transport to ground-level ozone
315   ($\mathrm{C}^{\mathrm{tra}}(O_3)$) at the hotspots over India and Southeast Asia increase to 30 % and 25 % respectively (see Fig. 3). The absolute contributions increase over India, while they decrease over South East Asia (see Fig. 4), but due to the reduced ozone levels $\mathrm{C}^{\mathrm{tra}}(O_3)$ increases. The increase over India is consistent with the increase of emissions from land transport in the SSP2 narrative, as population grows fast and air pollution control strategies and technological development are not strong enough to counteract the increased demand of mobility. Over Europe, East Asia, and North America, absolute and relative contributions
320   of land transport emissions decrease due to stricter air pollution policies compared to present day.

The SSP1-1.90 emission pathways shows a strong reduction of land transport emissions in all regions of the world, which lead to a strong reduction of the contributions to ozone (see Fig. 3). The largest levels of $\mathrm{C}^{\mathrm{tra}}(O_3)$ are simulated in the Middle East and South Asia (8−9 %). As O_3 itself decreases in the SSP1-1.9 (see Fig. 2), absolute contributions decrease at an ever stronger pace (see Fig. 4). This strong decrease is caused by the strong change of the transport system towards public transport,
325   and electrification of the transport system as well as the very strong air pollution mitigation efforts.

The SSP3-7.0 projection in 2050 features globally even larger land transport emissions compared to present-day, caused by the combination of population growth, slow technological development and less strict air pollution mitigation. This increase
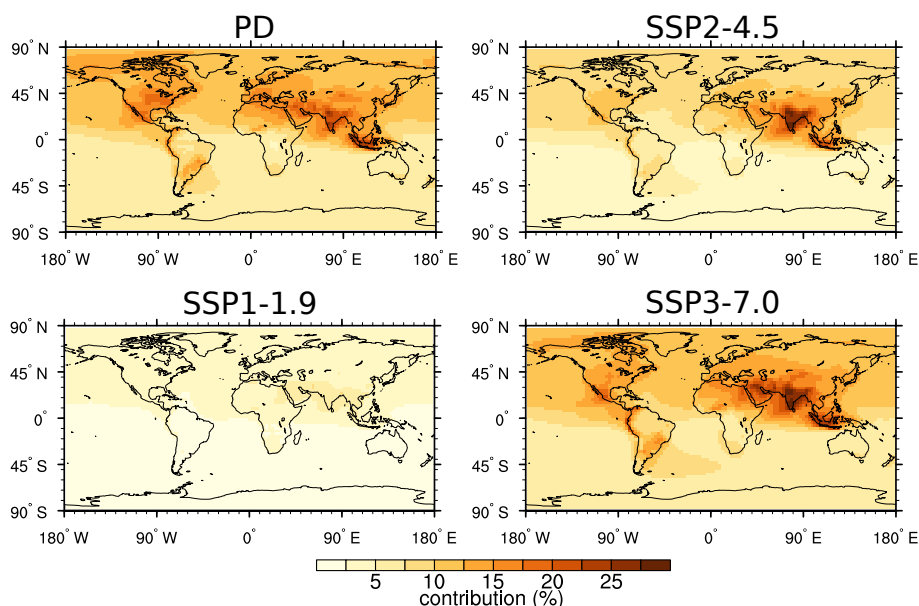
Figure 3. Relative contributions of land transport emissions to ground-level ozone (in %) for the simulations PD, SSP1-1.9, SSP2-4.5, and SSP3-7.0 (5 year average). The colour bars are identical in all panels to allow for a better comparability between the results for the different emission projections.

of the emissions globally leads to increasing contributions of the land transport emissions to ozone, but regional changes vary. Even though, land transport NO_x emissions are reduced by around 50 % compared to 2015 over North America and Europe,
330 relative $(16-18 \%)$ and absolute contributions decrease only slightly. A detailed analyses of the reason will follow in Sect. 4.5. Over the Middle East, India, and South America $C^{tra}(O_3)$ increases up to 28 %, 31 % and 22 %, respectively. In these regions also absolute contributions increase. This increase is caused by an increase of emissions due to the combination of increased mobility demand, lack of public transport, and failure of effective emission control.

To quantify whether the changes in the contribution of land transport emissions to ground-level ozone for the future pro-
335 jections are caused by changes in regional emissions or by emission changes in other regions of the world, we performed a source-receptor analysis (see Fig. 5). We define the tagged regions (see Table 1) as sources, and the regions as used in Fig. 4 as receptor regions. Please note, that the region Southeast Asia is not defined as separate source regions, but listed under rest of the world (ROW).

For all receptor regions in 2015, regional land transport emissions are the most important source for the contribution of land
340 transport emissions to ozone. This means that for example land transport emissions from the source region Europe contributes the most to the contribution of land transport emission to ozone in the receptor region Europe. This changes in the projections for 2050, particularly for Europe and North America. Here, the contribution of regional land transport emissions declines sharply due to air pollution control and the source regions of the ROW have (partly) larger contributions compared to regional emissions.
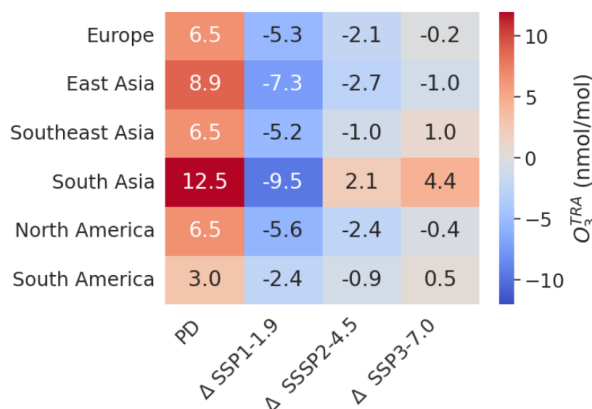
Figure 4. Absolute contributions of land transport emissions to ground-level ozone for 2015 and the differences of the absolute contributions for the three 2050 simulations ('2050 minus 2015', all in $\mathrm{nmol\ mol^{-1}}$). Values are area averaged over specific geographical areas and for 5 years (see Sect. S9.1 in the Supplement for exact definition of the regions).

Interestingly, only the emissions of the source regions North America, South Asia, and East Asia have larger influence on the contributions of land transport emissions for other receptor regions. The contributions of European land transport emissions to ground level ozone are confined around Europe (see also additional Figures in Sect. S3 in the Supplement).

The large contributions of land transport emissions for present day and in the projections SSP2-4.5 and SSP3-7.0, especially for South Asia, shows the importance of strong mitigation measures in these regions to effectively reduce O_3 levels. SSP1-1.9 shows one possible mitigation pathway, leading to a strong reduction of land transport emissions and its contribution to ozone over South Asia by strong changes of the transport system and strong pollution control. In addition, the population is less strongly increasing, leading to a lower demand for mobility in SSP1-1.9 compared to SSP2-4.5 and SSP3-7.0 in South Asia.

Overall, the importance of regional emissions in most receptor regions indicates a large benefit for ground-level ozone when reducing regional emissions. However, the increasing importance of the contribution of land transport emissions from other regions in the projections shows how important it is to reduce ozone precursors not only regionally, but also globally. Especially the increase of the ROW contribution for Europe, South Asia and North America in SSP3-7.0, in which international cooperations are strongly reduced, shows how important these international co-operations are in order to reduce air pollution by long range transport.
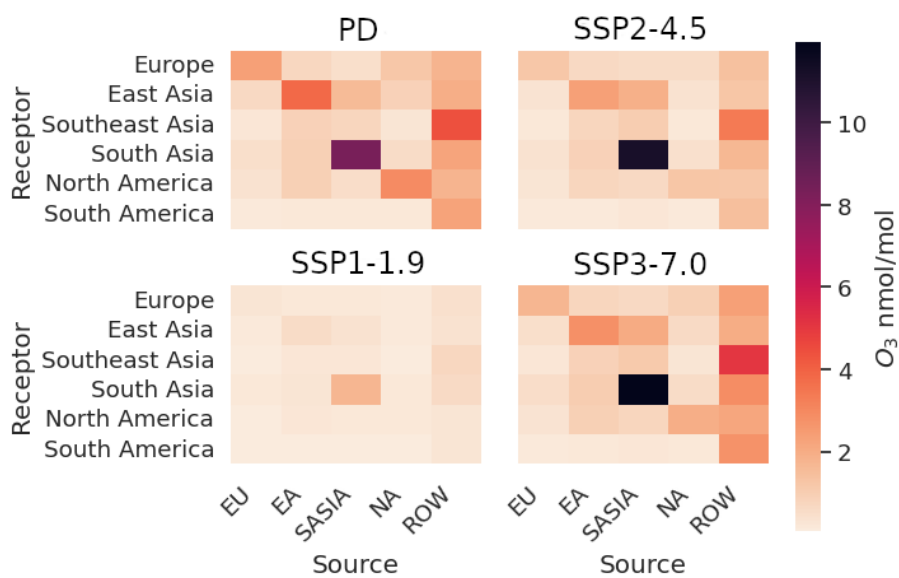
Figure 5. Source receptor analysis of the absolute contribution of land transport emissions to ground-level ozone (in $nmol\ mol^{-1}$). The values are mean values over 5 years and area weighted over the receptor regions. Exact definitions of the receptor regions are given in Sect. S9.1 in the Supplement.

4.3 Shipping

360 Figure 6 shows the relative contributions of shipping emissions to ozone ($C^{\mathrm{ship}}(O_3)$) at ground level. Quantifications of the absolute contribution of shipping emissions for specific geographical regions, which are Atlantic, Indian Ocean, Pacific and the Mediterranean Sea (see Sect. S9.2 in the Supplement for a detailed definition), are displayed in Fig. 7. Additional figures depicting absolute contributions at ground-level and zonal means are part of the Supplement (Sect. S3 and S4 therein).

Generally, $C^{\mathrm{SHP}}(O_3)$ is largest in the North Eastern Pacific and the North Eastern Atlantic. In 2015, $C^{\mathrm{SHP}}(O_3)$ reaches up
365 to 18 % in these regions, corresponding to mixing ratios of O_3^{SHP} of up to 7 nmol mol^{-1} in the North Eastern Pacific and up to 12 nmol mol^{-1} along the Western Coast of North Africa. Similarly, to the land transport emissions, the geographical distribution of shipping emissions for present day is comparable to previous studies (Mertens et al., 2018; Butler et al., 2020). Despite higher emissions from the shipping sector, the simulated contributions in the present study are slightly lower compared to Mertens et al. (2018), which is probably due to the larger overall NO_x emissions in the present study. Compared to Butler
370 et al. (2020), the simulated contributions in the present study are lower. This is due to the differences in the tagging methods used. While Butler et al. (2020) performed either NO_x or VOC tagging, we consider all precursor and their competition in ozone production an loss reactions simultaneously (see Sect. 7).

In contrast to the land transport sector, where the geographical patterns of the contributions change strongly due to the regional emission shifts resulting from the local nature of the mitigation measures in this sector, the geographical pattern of
375 the regions with the largest contributions from the shipping sector remains very similar in 2050 for SSP2-4.5 and SSP3-7.0. This can be expected as also the shipping emissions show no strong regional shifts (besides the introduction of SECAs) and mitigation measures are mostly applied globally due to technological advancements by means of improved emission factors (i.e. technological advancements). The absolute and relative contributions decrease to around 16 % (8 nmol mol^{-1}) for SSP2-4.5 and 14 % (7 nmol mol^{-1}) for SSP3-7.0. The lower contributions of shipping emissions in SSP3-7.0 compared to SSP2-4.5
380 is in accordance with lower shipping emissions caused by the decrease of global trade in the SSP3 narrative (see Sect. 3). The strong decrease of shipping emissions in SSP1-1.9 leads to strong reduction of the absolute and relative contributions from shipping emissions to ozone. Accordingly, $C^{\mathrm{ship}}(O_3)$ drops below 4 % and O_3^{ship} to 2 nmol mol^{-1} at maximum.
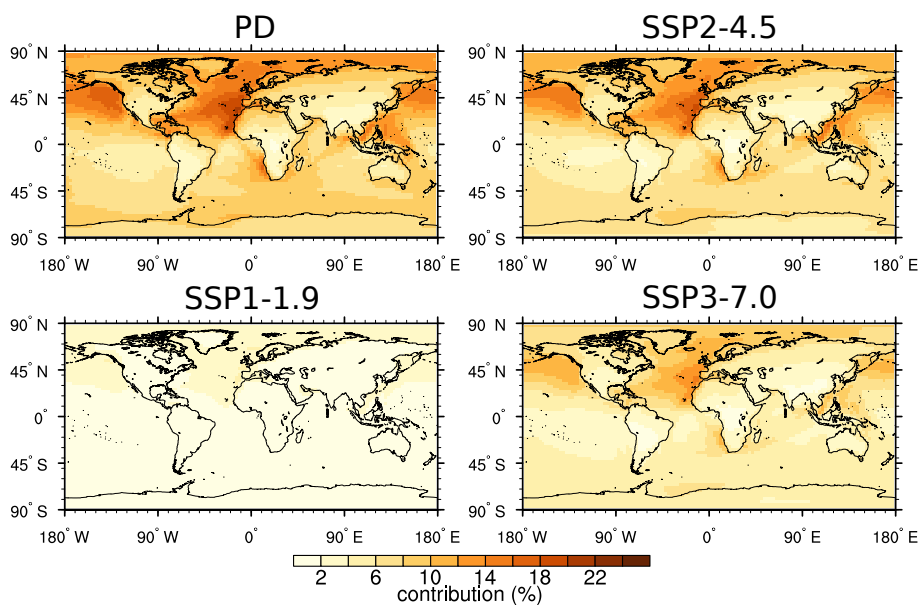
Figure 6. Relative contributions (in %) of shipping emissions to ground-level ozone as simulated by PD, SSP1-1.9, SSP2-4.5 and SSP3-7.0. All values are annual averages for 5 year average. The colour bars are identical in all panels to allow for a better comparability between the results for the different emission projections.
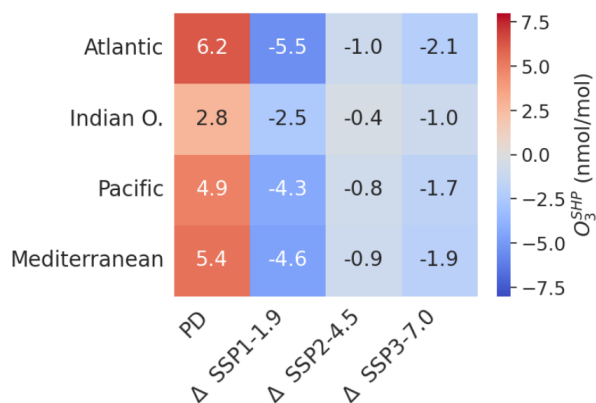
Figure 7. 5-year mean of the absolute contributions of shipping emissions to ground-level ozone for PD and the differences of the absolute contributions for the three 2050 simulations ('2050 minus 2015', all in $\mathrm{nmol\,mol^{-1}}$). Values are area-averaged for specific geographical regions (see Sect. S9.2 in the Supplement for a detailed definition).

4.4 Aviation

Figure 8 shows the zonal average of the relative contributions of aviation emissions to ozone ($C^{air}(O_3)$). Absolute contributions
385 of aviation emissions to the tropospheric column of O_3^{AIR} averaged over the main flight regions between North America and
Asia, Asia and Europe and North America and Europe (see Sect. S9.3 in the Supplement for a detailed definition) are given in
Fig. 9.

The largest values of the relative and absolute contributions in 2015 are simulated in the Northern hemisphere. They reach
up to 5 % ($4-5$ nmol mol^{-1}) and occur over a large region between $60° - 90° N$ (see Fig. S10 in the Supplement). The
390 largest relative contributions can be found between 800 and 300 hPa, while the largest absolute contributions are simulated
around 250 hPa.Generally, the geographical distribution of the regions with the largest contributions are similar to the regions
with the largest ozone changes as quantified with the perturbation approach reported by Koffi et al. (2010a). The results of
the perturbation approach, however, show a relative change o ozone which is much more limited to the upper troposphere.
This is because below the cruise altitude ozone precursors from other sources are becoming more and more important. While
395 in the perturbation approach the non-linearities in the net ozone production lead to a compensation of reduced emissions and
increased net ozone production, here such an compensation is avoided.

The increase of the aviation emissions in SSP2-4.5 and SSP3-7.0 due to the increase of passenger travel demand lead to
an increase of the ozone contributions. For the SSP2-4.5 projection, $C^{air}(O_3)$ reaches up to 10 % ($8-8.5$ nmol mol^{-1}) in
the Northern hemisphere in 2050. Like for the shipping emissions, the aviation emissions are lower in the SSP3-7.0 emission
400 pathway compared to the SSP2-4.5 pathway. Accordingly, maximum relative and absolute contributions of aviation emissions
to ozone reach up to 8 % ($7-7.5$ nmol mol^{-1}) in SSP3-7.0 for 2050. Similar as the absolute and relative contributions,
the tropospheric ozone columns of O_3^{air} increase by almost 90 % (SSP2-4.5) and 70 % (SSP3-7.0) compared to present-day
conditions (Fig. 9).

The SSP1-1.9 projection shows the lowest contributions in 2050 compared to the two other pathways, caused by decreased
405 passenger travel demand compared to SSP2-4.5 and SSP3-7.0 and technological advancements compared to 2015. The max-
imum of $C^{air}(O_3)$ of around 4 % is only a slight reduction compared to the values in 2015. However, the background ozone
is much lower in SSP1-1.9 for 2050 due to the overall decreased emissions, so that the 4 % relative contribution corresponds
to maximum absolute contributions of 2.5 nmol mol^{-1}. Accordingly, also the tropospheric ozone column of O_3^{air} decrease by
around 50 % in SSP1-1.9 compared to present-day.
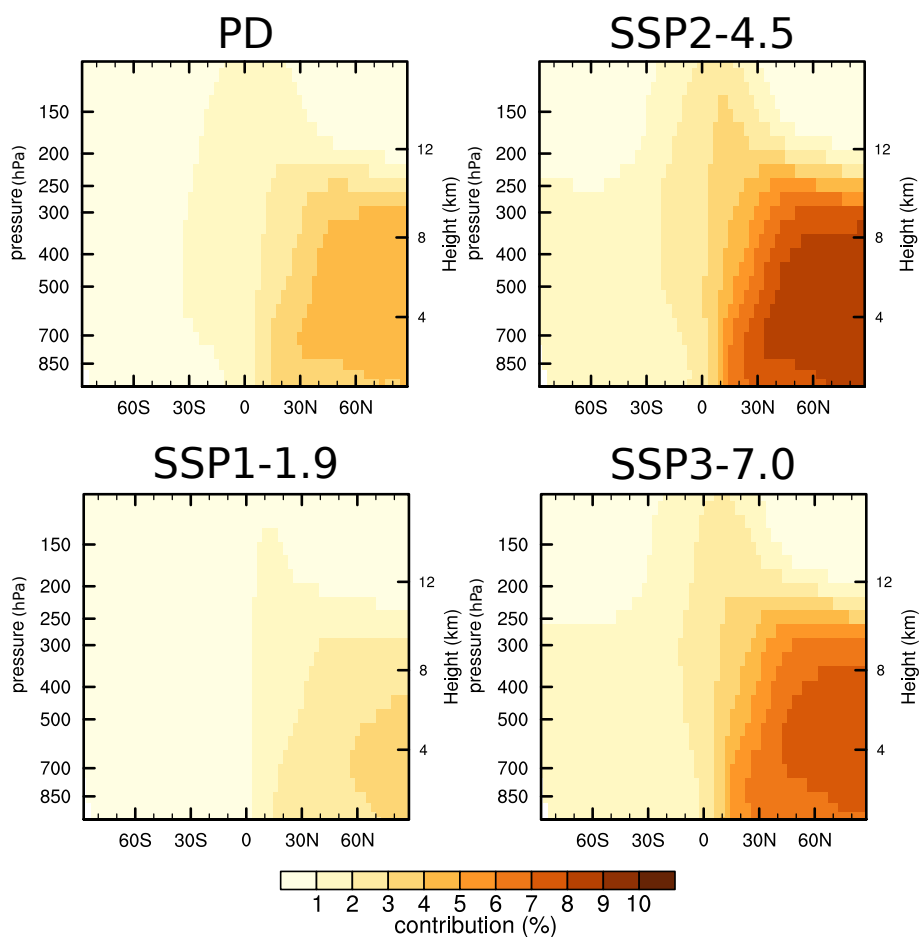
Figure 8. Zonal average of the relative contributions (in %) of aviation emissions to ozone as simulated by PD, SSP1-1.9, SSP2-4.5 and SSP3-7.0.All values are averages for 5 years. The colour bars are identical in all panels to allow for a better comparability between the results for the different emission projections.
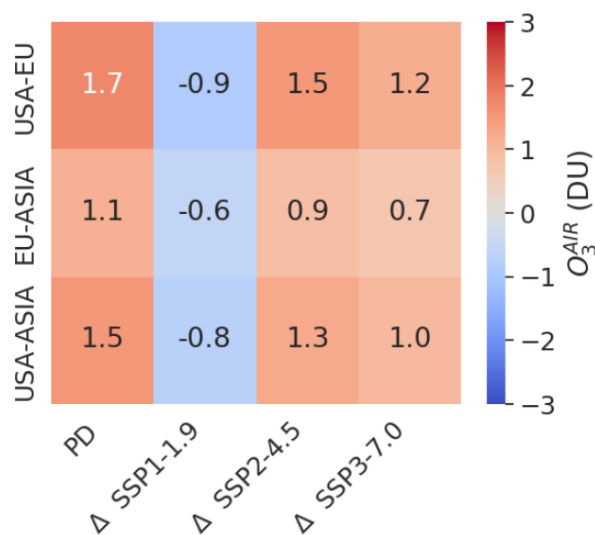
Figure 9. 5-year mean tropospheric O_3^{AIR} column for *PD* (in Dobson Units, DU) and the differences of the tropospheric O_3^{AIR} columns between the three 2050 simulations ('2050 minus 2015'. The values are area averaged over the the main flight regions (see Sect. S9.3 in the Supplement for a detailed definition).

4.5 Analysis of the non-linearity of the ozone chemistry

The production of tropospheric ozone is strongly non-linear and depends on the ratio of NO_x to VOC concentrations (e.g., Sillman, 1995; Seinfeld and Pandis, 2006; Archibald et al., 2020). In the NO_x-limited regime, the ozone production efficiency (i.e. the amount of ozone molecules formed by one NO_x molecule) increases if NO_x emissions are reduced and decreases with an increase of the NO_x emissions.

To analyse how the ozone burden in the atmosphere changes for the different projections, we introduce the following metric which we define exemplary for the shipping sector as following:

$$\chi^{SHP} = \frac{B(O_3^{SHP})}{E(NO_x^{SHP})}. \tag{3}$$

In this definition, $B(O_3^{SHP})$ is the global burden (in Tg) of O_3^{SHP} and $E(NO_x^{SHP})$ are the annual NO_x emissions of the shipping sector in Tg(NO). Increasing values of χ (called the ozone burden efficiency) indicate an increase of the ozone burden per emitted NO_x. We prefer this quantity over the commonly analysed ozone production efficiency as it includes production and loss processes of ozone and therefore is a better measure for the associated radiative forcings.

Figure 10 shows the dependency between χ and the NO_x emissions. The results clearly show the expected increase of the ozone burden efficiency with decreasing emissions. Depending on the prevailing meteorological conditions and on the chemical background conditions, χ differs among the transport sectors and in different world regions (as analysed for the land transport emissions). Generally, NO_x emissions from the aviation sector have a larger efficiency in forming ozone compared to surface emissions, because the emissions are emitted under cleaner background conditions. This is consistent with previous findings (e.g Hauglustaine et al., 1994; Derwent et al., 2001; Hoor et al., 2009; Dahlmann et al., 2011; Butler et al., 2020; Terrenoire et al., 2022). Land transport emissions in Europe or North America have rather low values of χ (at least) for present day, as they are emitted in a rather polluted background and the prevailing meteorological conditions do not favour ozone production over parts of North America and Europe.

The increase of χ with decreasing NO_x emissions underlines a major challenge for ozone mitigation. A very prominent example are the land transport emissions from Europe and North America in SSP3-7.0. Even though the NO_x emissions are reduced by around 40 % (Europe) and 50 % (North America) compared to 2015, the global burdens of ozone produced from these emissions are reduced by 5 % (Europe) and 16 % (North America), only. Accordingly, reductions of NO_x emissions from the land transport need to be large enough to overcome strong non-linear behaviour of the ozone chemistry so that the emission reductions lead to significant reductions in ozone. The SSP1-1.9 and SSP2-4.5 projections show two possible projections with such strong reductions over Europe and North America. As example, the reduction of NO_x emissions by 63 % over Europe in SSP2-4.5 compared to 2015 leads to 30 % reduction of ozone from European land transport emissions.
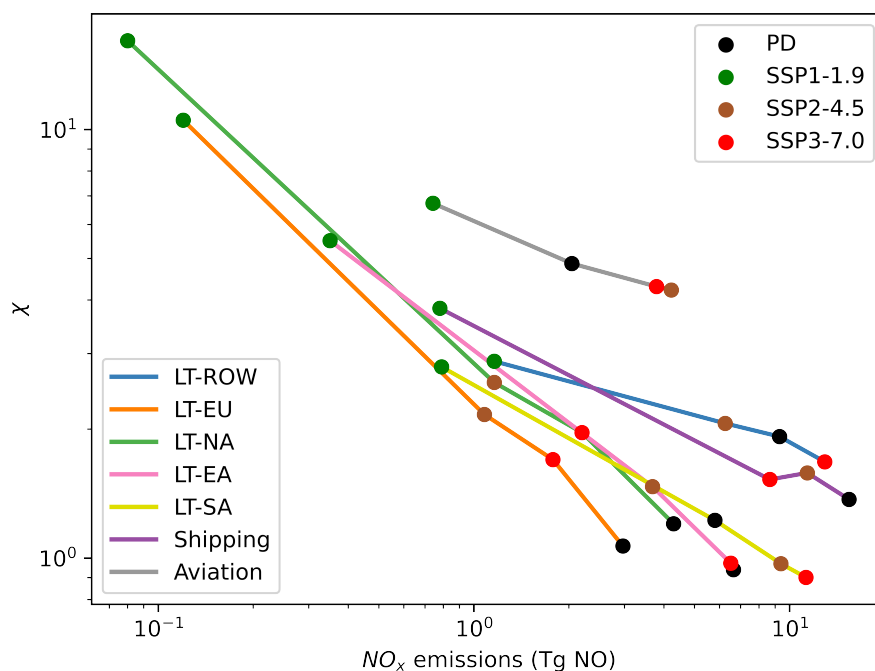
Figure 10. Dependency between χ and the annual NO_x emissions (in Tg(NO)). χ is calculated as global mean over 5 years. Shown are the results of the 7 transport categories (land transport (indicated as LT-): ROW, EU, NA, SA and EA, aviation and shipping) for each simulation (colour coded).

5   Ozone radiative forcing

440   The stratospherically adjusted radiative forcings of ozone (for simplicity from now on called RF) of the different transport
sectors are depicted in Figure 11. The RF is given for the three main transport sectors (Fig. 11a) and for the land transport
emissions from specific regions in Fig. 11b (see Table 1 for naming convention). The total RF of the land transport sector in
2015 is 121 mW m^{-2}. This value is 29 mW m^{-2} (≈ 32 %) larger compared to Mertens et al. (2018), which is partly due to
the ~ 35 % larger NO$_x$ emissions from land transport in the present study. Moreover, especially the land transport emissions

445   in Asia are larger in the present study, which have a rather high efficiency for RF (see Sect. 5.1).

The RF of the shipping sector in 2015 is 54 mW m^{-2}, which is 8 mW m^{-2} (≈ 13 %) lower compared to Mertens et al.
(2018), despite the about 20 % larger NO$_x$ emissions in the present study. This is most likely caused by the overall larger
NO$_x$ emissions from anthropogenic sources in the present study compared to Mertens et al. (2018). This larger overall NO$_x$
emissions lead to a more polluted background and a lower ratio of the shipping emissions compared to all anthropogenic

450   emissions (around 7 % in the present study compared to 12 %).

The RF of the aviation sector in 2015 is 29 mW m^{-2}, which lies between the best estimates by Lee et al. (2021) for 2011
(27 mW m^{-2}) and 2018 (36 mW m^{-2}). Note that the values from Lee et al. (2021) are based on a literature review from
various models using the perturbation approach. Only a limited number of model results from EMAC are included in the multi-
model assessment. The included results of EMAC tend to be at the lower end of the ozone response in Lee et al. (2021) and

455   also in other multi-model assessments (Søvde et al., 2014, and Cohen et al., pers. communication). A recent assessment with
EMAC, comparing different methodologies to asses the effects of aviation emissions on ozone, diagnoses a difference of ozone
between the perturbation and tagging method of a factor of 1.16 to 2.55, with a mean value of 1.83 (Maruhashi et al., 2024).

As mentioned in Sect. 3, an inconsistency of the geographical distributions in the CMIP6 aviation emissions applied here,
lead to an underestimation of the aviation O$_3$ RF of about 8 % (see Thor et al., 2023, for a detailed analysis).

460   The RF from land transport emissions is reduced to 103 mW m^{-2} in 2050 with the SSP2-4.5 projection. The strong increase
of the emissions from the land transport sector in SA leads to an increase of the RF from O$_3^{LT-SA}$ compared to 2015 of around
7 mW m^{-2}. The RF of the land transport emissions of all other regions decrease slightly. The RF of the shipping emissions
decreases slightly by 2 mW m^{-2} in 2050 compared to 2015, while the RF of the aviation sector almost doubles with SSP2-4.5.
With this large increase, the RF of the aviation sector is slightly larger than the RF of the shipping sector in 2050 for SSP2-4.5.

465   Moreover, the RF of the total transport sector (land, aviation and shipping) in 2050 with the SSP2-4.5 projection is almost as
large as for present day conditions, as the increase of the RF in the aviation is almost of same magnitude as the reductions in
shipping and land transport.

The strong increase of the land transport emissions in SSP3-7.0 compared to SSP2-4.5 leads to an increase of the RF from
land transport emissions to 140 mW m^{-2} in 2050. In comparison to 2015 and to SSP2-4.5 (in 2050) especially the RF of the

470   land transport emissions from the rest of the world (RoW) increases. Also the contributions of land transport emissions from
South Asia increase mainly due to the strong increase of the emissions in India. The RF of the land transport emissions from
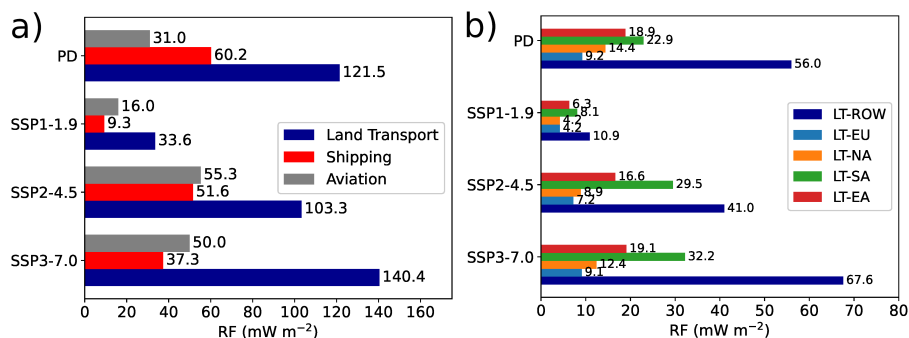
Figure 11. Stratospheric adjusted radiative forcing of ozone (in mW m^{-2}) from each sector (a) and for land transport emissions from specific regions (b).
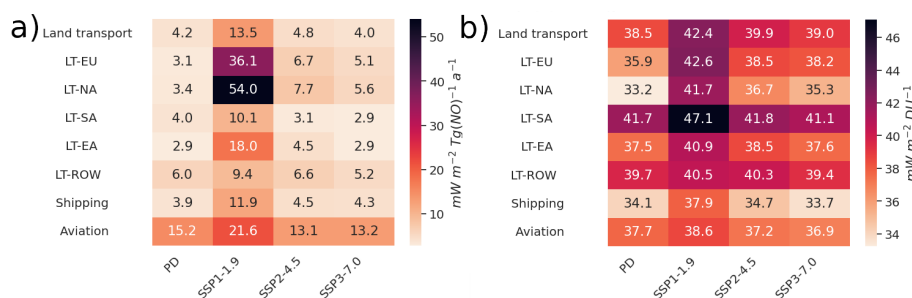
Figure 12. Global averages of RF_e (a, in $\text{mW m}^{-2}((\text{Tg(NO) a}^{-1})^{-1})$) and RF_{DU} (b, in $\text{mW m}^{-2}\text{DU}^{-1}$) for the emission sectors land transport, shipping and aviation as well as for the land transport emissions from specific regions only.

Europe, North America and East Asia stay relatively similar to 2015. Even though the NO_x emissions from land transport are reduced strongly, ozone is only reduced slightly due to non-linear effects (see Sect. 4.5).

Consistent with the changes of NO_x emissions and the O_3 contributions, RF from the aviation and shipping sector are
475 smaller for SSP3-7.0 compared to SSP2-4.5. The RF from shipping emissions is $37\ \text{mW m}^{-2}$ in 2050 with SSP3-7.0, which is a decrease of 17 and $15\ \text{mW m}^{-2}$ compared to SSP2-4.5 in 2015 and 2050, respectively. The RF from aviation emissions decreases for SSP3-7.0 by $5\ \text{mW m}^{-2}$ compared to SSP2-4.5 in 2050, but increases by $21\ \text{mW m}^{-2}$ compared to 2015.

In the SSP1-1.9 projection, RF from all transport sectors decreases strongly. The total land transport RF is reduced by a factor of four, the shipping RF by a factor of five and the RF from aviation emissions by a factor of three compared to 2015.
480 Also the RF of the land transport emissions from various regions are reduced strongly compared to 2015 and also compared to the other projections.

5.1 Radiative Forcing efficiency

The RFs discussed in Sect. 5 are difficult to compare among the different sectors due to the different amount of emissions for each sector and region. To set the discussed RFs into perspective with the emissions of the individual emission sectors, we calculate the RF efficiency with respect to the NO_x emissions for each sector (defined as RF_e, e.g., Dahlmann et al., 2011) as:

$$RF_e^i = \frac{RF^i}{E(NO_x^i)},$$
(4)

with $E(NO_x^i)$ denoting the NO_x emissions in $Tg(NO)\ a^{-1}$ and RF^i the stratospheric adjusted RF of the considered sector (denoted as i). In addition, we calculate the RF with respect to the tropospheric ozone columns, denoted as $O3_{tc}$, in Dobson Units (DU), for each emission sector (denoted with i) . We call this quantity RF_{DU} and define it as:

$$RF_{DU}^i = \frac{RF^i}{O3_{tc}^i}.$$
(5)

Figure 12 shows RF_e and RF_{DU} for the different projections and the different emission sectors. For present day, the aviation sector has the largest RF_e value of $15.2\ mW\ m^{-2}\ Tg(NO)^{-1}\ a^{-1}$ ($32.5\ mW\ m^{-2}\ Tg(N)^{-1}\ a^{-1}$). This is in general in-line with previous studies (e.g Hauglustaine et al., 1994; Derwent et al., 2001; Hoor et al., 2009; Butler et al., 2020; Terrenoire et al., 2022) and the high ozone burden efficiency from aviation NO_x emissions discussed in Sect. 4.5. Moreover, the value is 30 % ($7.5\ mW\ m^{-2}\ Tg(N)^{-1}\ a^{-1}$) larger as the multi-model mean of $25.1 \pm 7.3\ mW\ m^{-2}\ Tg(N)^{-1}\ a^{-1}$ by Lee et al. (2021) (see also discussion in Sect. 5). The inclusion of the correction of the CMIP6 aviation emissions as reported by Thor et al. (2023) increases the efficiency to $16.3\ mW\ m^{-2}\ Tg(NO)^{-1}\ a^{-1}$ ($34.9\ mW\ m^{-2}\ Tg(N)^{-1}\ a^{-1}$).

RF_e for shipping is slightly smaller than for total land transport emissions, caused by the much more effective upward transport of land transport emissions by convection in the upper troposphere compared to shipping emission as already discussed by (Hoor et al., 2009, see also zonal means in Sect. S4 in the Supplement). Land transport emissions from East Asia, Europe and North America show the smallest RF_e values, because ozone is formed rather inefficiently from emissions of these regions (see Sect. 4.5). Moreover, especially for Europe and North America, the produced ozone is not transported into the upper troposphere very efficiently where it would be most radiatively active (e.g. Riese et al., 2012; Rap et al., 2015). The latter is analysed using RF_{DU}, which is largest for land transport emissions from South Asia. This is in agreement with results from Dahlmann et al. (2011) reporting a strong latitude dependence of RF_{DU} with the highest values near the equator. Even though aviation has the largest RF_e values, RF_{DU} is lower as most emissions take place in mid-high latitudes (Dahlmann et al., 2011).

In the future projections, RF_e increases for those emission sectors and regions whose emissions decrease, and RF_e decreases where the emissions increase. The change of RF_e agrees with the change of χ. The large values of RF_e for some sectors with SSP1-1.9 emissions are partly also dominated by the small amount of emissions, leading to numerical difficulties calculating the quotients.

The strong decrease of the emissions in the SSP1-1.9 projections lead to an increase of RF_{DU}, caused by reduced saturation effects of the radiative efficiency as analysed by Dahlmann et al. (2011). This increase of the radiative efficiency causes an additional penalty on ozone mitigation, as ozone gets more radiativly efficient if tropospheric ozone levels are strongly reduced.

6 Contributions of transport emissions to OH and to the methane lifetime

515 Emissions of the transport sector affect not only tropospheric ozone levels, they also have a strong influence on OH mixing ratios and with this on the tropospheric oxidation capacity. Increases in OH lead to a reduction of the methane lifetime. Previous studies analysed the influence of transport emissions on OH and methane lifetime with the perturbation approach (e.g., Hoor et al., 2009). Here, we determine the contribution of the emissions from the transport sector to OH and the associated contributions to the methane lifetime for the first time with a source attribution method. Due to the methodological differences,

520 the results obtained in the present study are not directly comparable to previous studies. Therefore, we also performed three perturbation simulations, in which we applied a 5 % emission reduction to land transport, shipping and aviation, respectively. The changes of the methane lifetime by emissions from land transport, aviation and shipping with this method are in agreement with Hoor et al. (2009), who applied the same 5 % perturbation (see Sect. S5 in the Supplement).

6.1 Contributions to OH

525 Figure 13 depicts the zonal average of the contribution of land transport emissions to OH (OH^{TRA}). The zonal mean pattern of the contribution for the present-day simulation is similar as the one presented by Rieger et al. (2018). Two maxima of OH^{tra} are simulated, one in the lower troposphere between the surface and 600 hPa at 0-30° N peaking at 20 fmol mol^{-1}, and one in the upper tropical troposphere between 100 and 200 hPa peaking at 26 fmol mol^{-1} The maximum in the upper troposphere, however, is mainly due to the overall very large OH mixing ratio in these regions and the relative contributions

530 in these regions are rather low (up to 5 % in 2015). Near the surface, the relative contributions of land transport emissions to OH reach up to 40 % in the NH in 2015. Hoor et al. (2009) reported only one maximum in the lower troposphere by applying the perturbation approach. This corroborates the fact that results of the attribution method and of the perturbation methods are not directly comparable. This second maximum of OH^{TRA} is caused by photolysis of O_3^{tra} in the upper troposphere and subsequent formation of OH.

535 Due to the strong interconnection between NO_x, O_3 and OH in atmospheric chemistry, the overall changes of OH^{TRA} in the future projections are similar as discussed for the O_3 contributions in Sect. 4.2. Accordingly, increases of the NO_x emissions from land transport lead to an increase of OH^{TRA} and vice versa.

As for O_3^{SHP}, OH^{SHP} is mainly limited to the lower troposphere (see Fig. 14). The maxima are simulated near the surface, with enhanced contributions up to 500 hPa. In the tropical upper troposphere only small contributions are simulated, because

540 also O_3^{SHP} is limited to the lower troposphere (see Sect. S3 in the Supplement). In 2015, the zonal average of OH^{SHP} reaches up to 16 fmol mol^{-1} near the surface and decreases to $8-12$ fmol mol^{-1} in the middle and upper troposphere.

In accordance with the changes of the emissions in the three SSP projections, the contributions of OH^{SHP} is lowest in SSP1-1.9, followed by SSP3-7.0 and SSP2-4.5, which has the largest contributions in 2050. The contributions in 2050 with SSP2-4.5 emissions, however, are lower compared to 2015.

545 The zonal average of OH^{AIR} (see Fig. 15) shows a clear peak in the Northern hemisphere between $30-60°$N and $200-300$ hPa which is in accordance with Rieger et al. (2018). The region with the maxima of OH^{air} is similar with the region of the largest
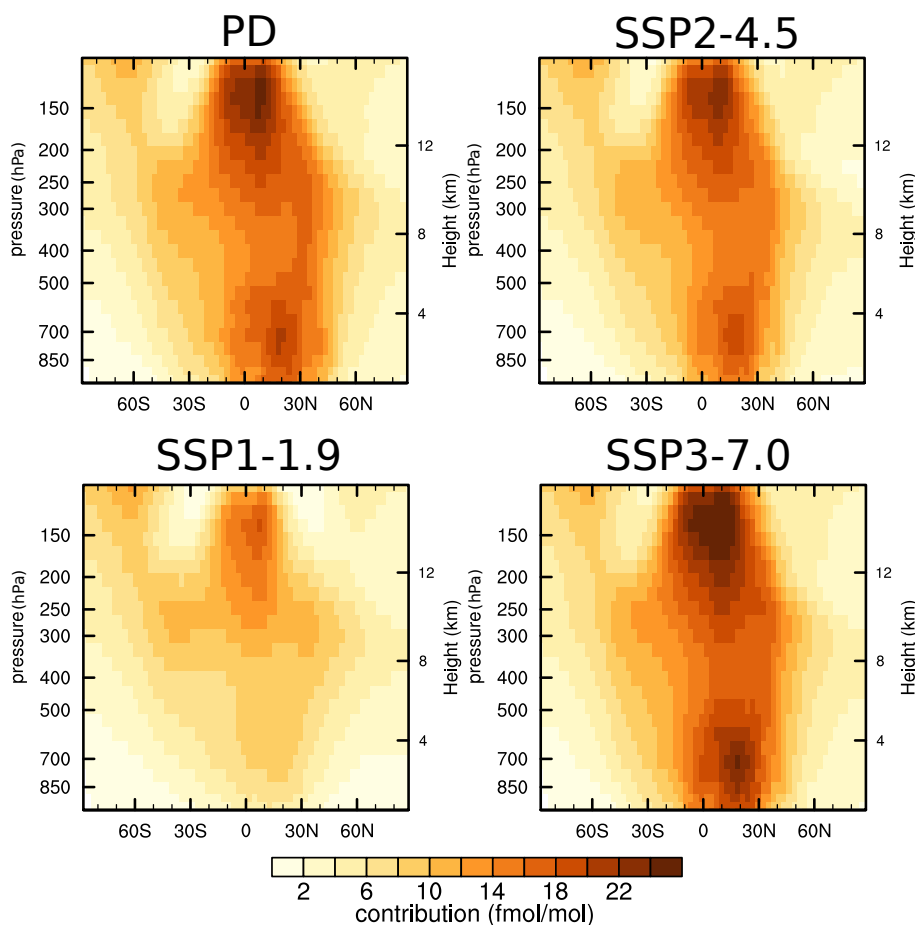
Figure 13. Zonal average of the absolute contribution of land transport emissions to OH (in $\mathrm{fmol\,mol^{-1}}$ for PD, SSP1-1.9, SSP2-4.5 and SSP3-7.0. The shown values are 5 year averages.

impact of aviation emissions on OH at around $40°$ N and between $200-300\,\mathrm{hPa}$ reported by Hoor et al. (2009). In 2015 the zonal average of OH^{air} peaks up to $22\,\mathrm{fmol\,mol^{-1}}$. In agreement with the changes of the aviation emissions and O_3^{AIR}, OH^{air} increases in 2050 for SSP2-4.5 (up to $36\,\mathrm{fmol\,mol^{-1}}$) and SSP3-7.0 (up to $33\,\mathrm{fmol\,mol^{-1}}$) compared to 2015. Due to the large
550 decrease of emissions in SSP1-1.9, peak values of the zonal average of OH^{AIR} decrease to $13\,\mathrm{fmol\,mol^{-1}}$.

As OH is strongly coupled to O_3 (and vice-versa) also OH shows a similar non-linearity as O_3. Accordingly, also the OH contributions respond non-linearly to the changes of the emissions from the traffic sectors (see Sect. S7 in the Supplement).
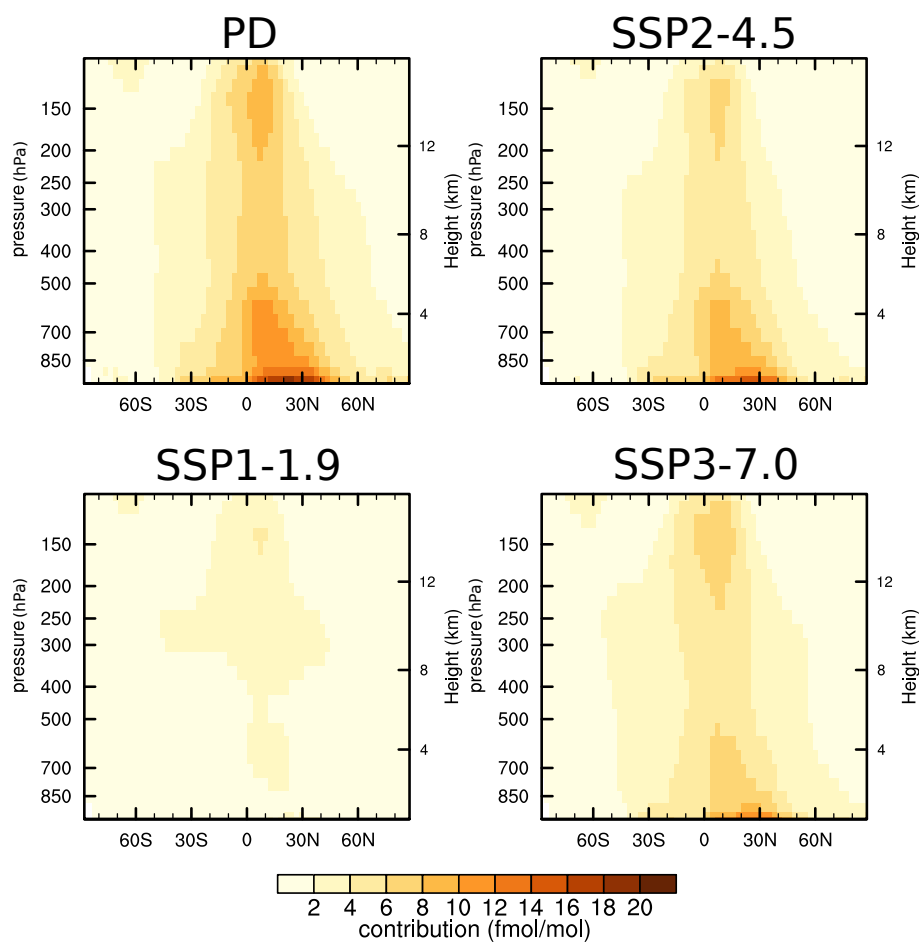
Figure 14. Zonal average of the absolute contribution of shipping emissions to OH (in $\mathrm{fmolmol^{-1}}$ for PD, SSP1-1.9, SSP2-4.5 and SSP3-7.0. The shown values are 5 year averages.
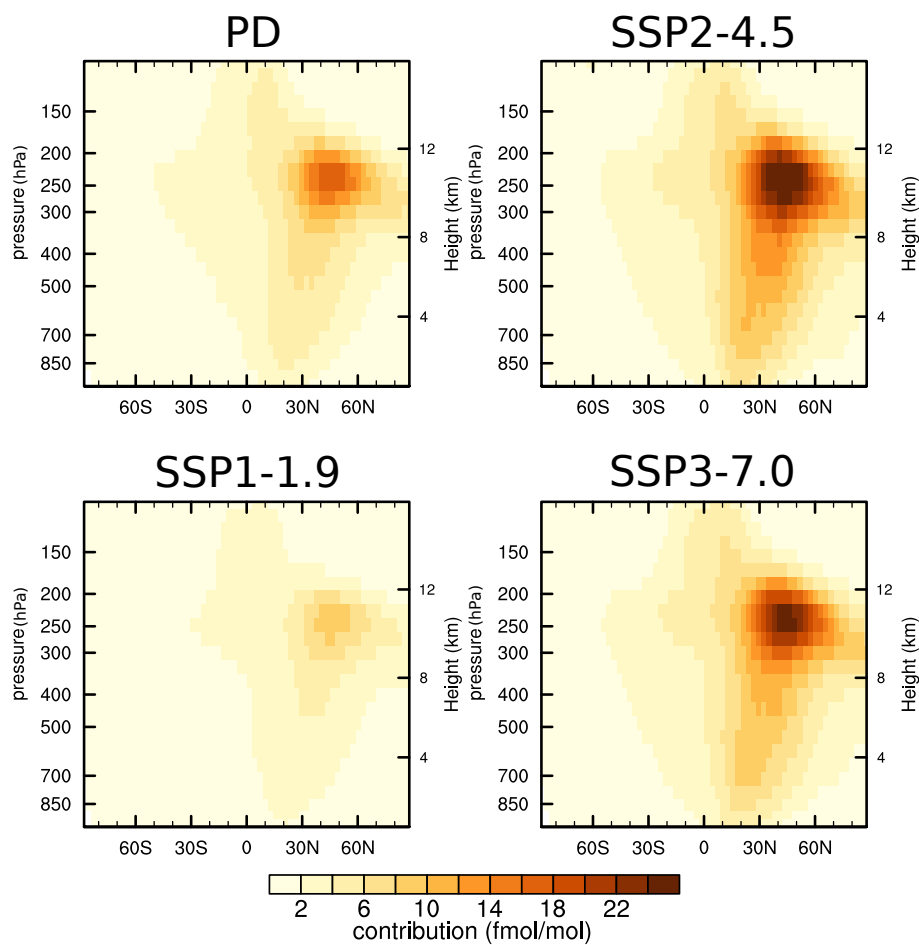
Figure 15. Zonal average of the absolute contribution of aviation emissions to OH (in fmolmol^{-1} for *PD*, *SSP1-1.9*, *SSP2-4.5* and *SSP3-7.0*. The shown values are 5 year averages.

6.2 Reductions of the methane lifetime

Methane is an important greenhouse gas and therefore changes in lifetime, e.g. due to changes in OH, caused by emissions
555   of the transport sector might contribute significantly to the climate effect of transport emissions. Based on the contributions of
the different emissions sectors to OH, the effect of the different emission source onto the methane lifetime is analysed. This
analyses helps us to understand how the transport emissions affect the methane lifetime. In the following we define methane
lifetime as the yearly global mean of CH_4 divided by the yearly global mean (expressed as overbars) of the loss of CH_4 by OH
as diagnosed by TAGGING ($loss(CH_4)_{total}$, see Rieger et al., 2018);

560
$$\tau(CH_4) = \frac{\overline{CH_4}}{\overline{loss(CH_4)_{total}}} \tag{6}$$

The lifetime is calculated for the troposphere, i.e. below the monthly mean tropopause (pressure) height. In the present study
the tropopause is defined according to the WMO definition based on the temperature lapse rate for latitudes equatorward of
$30°$, and as the potential vorticity iso-surface of 3.5 PVU at latitudes poleward of $30°$ (see Grewe and Dameris, 1996, for an
anlyssis on the effects of different tropopause definitions).

565   The loss rates of CH_4 for 18 different sectors, including the transportation sectors land, ship, and aviation are diagnosed
based on the OH contribution discussed in Sect. 6.1. Based on these OH contributions, the relative reduction of the methane
lifetime of a specific emission source (for example shipping emissions) is then calculated as:

$$\Delta_{rel}\tau(CH_4)_{SHP} = \frac{\overline{\tau(CH_4)_{total}} - \overline{\tau(CH_4)_{no-SHP}}}{\overline{\tau(CH_4)_{no-SHP}}} = -\frac{\overline{loss(CH_4)_{SHP}}}{\overline{loss(CH_4)_{total}}}. \tag{7}$$

'SHP' is the CH_4 loss attributed to the emissions of the shipping sector based on the OH^{SHP}, 'no-SHP' is the loss of all
570   other categories, except the shipping sector. Please note, the reciprocals of the methane lifetime calculated for each tagged
OH are additive and sum up to the total methane lifetime as calculated by Equation 6. The relative reduction as calculated
according to Equation 7 always refer to the total methane loss for one specific simulation. Thus, the relative reduction of the
methane lifetime for one simulation sum up to 100 %. In the different simulations, the methane lifetime differs slightly. This
needs to be considered when comparing the relative reductions of the methane lifetime analysed for the different simulations.
575   The methane lifetimes analysed for the different simulations by Equation 6 are 7.46 a (2015), 7.45 a (SSP2-4.5 in 2050), 7.43 a
(SSP3-7.0 in 2050) and 8.01 a (SSP1-1.9 in 2050).

Figure 16a shows the relative reduction of the methane lifetime (calculated with Equation 7)) due to the emissions of
the different transport sectors for 2015 and the three emissions projections in 2050. Generally, transport emissions lead to a
reduction of the methane lifetime and hence to a reduction of the climate effects of methane. From all transport modes. the
580   land transport emissions have the largest effect on the methane lifetime. The relative reductions of the methane lifetime by
land transport emissions ($\Delta_{rel}\tau(CH_4)_{TRA}$) ranges between -7 % (SSP1-1.9) and -16 % (SSP3-7.0). The relative reductions
of the methane lifetime from shipping emissions (($\Delta_{rel}\tau(CH_4)_{SHP}$) range from -9 % (*PD*) to -2 % (SSP1-1.9). The relative

reduction of the methane lifetime of aviation emissions (($\Delta_{rel}\tau(CH_4)_{AIR}$) is lowest with SSP1-1.9 (-3 %) and largest with SSP2-4.5 (-7 %) and SSP3-7.0 (-6 %) emissions. Despite for 2015 conditions, shipping and aviation have similar relative

585   reductions of the methane lifetime, even though the shipping sector has much larger emissions (see Fig. 1). This is due to the larger mixing rations of OH^{AIR} compared to OH^{SHP} (see Sect. 6.1). Moreover, due to the strong non-linearity of the OH chemistry, even in the clean SSP1-1.9 projection the transport sectors, especially the land transport sector, have a rather large effect on the methane lifetime. This non-linearity is larger for the reduction of the methane lifetime than for the ozone RF. As example, the ozone RF of land transport decreased by around 70 % between 2015 and SSP1-1.9, while the relative reduction of

590   the methane lifetime is 50 %. The reverse effect can be observed when increasing transport emissions, for example the ozone RF of land transport emissions between 2015 and SSP3-7.0 increase by 15 %, while the reduction of the methane lifetime changes by 8 %.

    Figure 16b breaks down the reduction o the methane lifetime by land transport emissions to the different geographical regions. In agreement with the contributions to the ozone RF (see Sect. 5) the largest relative reduction of the methane lifetime

595   are caused by land transport emissions from the rest of the world (LT-ROW). Relative reductions of the other regions are smaller by a factor of $2-3$, with the largest relative reductions from South Asia (LT-SA).

    Generally, the relative reductions of the methane lifetime as analysed by the source attribution approach are much larger than relative reductions of the methane lifetime calculated by a 5 % perturbation (see Sect. S5 in the Supplement for a detailed analysis for 2015 conditions). The reduction of the methane lifetime as analysed by the perturbation approach for the land

600   transport emissions is -1.9 %, -4.5 %, for the shipping emissions and -1.1 % for the aviation emissions, respectively. Accordingly, the reductions of the methane lifetime diagnosed by the source attribution approach are larger by a factor of 7.7 (land transport), 1.9 (shipping) and 3.5 (aviation) compared to the reductions as analysed by a 5 % reduction with the perturbation approach. This large difference of the relative reductions of the methane lifetime as analysed by the two methods also leads to shifts in the ranking of the different transport sectors w.r.t. the influence of the methane lifetime. While with the perturbation

605   method the shipping emissions are a factor two larger than for land transport emissions, with the source attribution method the contribution of land transport emissions is a factor 2 larger than for shipping emissions. The main reason for this different behaviour is that most shipping emissions take place in cleaner environments with less emissions compared to the land transport emissions. In these cleaner environments the non-linearity of the ozone chemistry is not that pronounced and the linearisation of the perturbation approach does not affect the results strongly (see also Grewe et al., 2010).

7   Limitations and uncertainties

In the following we discuss the limitations and uncertainties of this study. The main points we discuss are the usage of the same meteorological conditions and methane emissions in all simulations, the different available tagging method for source attribution and the influence of the errors in the geographical distribution of the aviation emissions in CMIP6.

    The signals of emissions from specific regions or specific emission sources (such as e.g. aviation) are small. To quantify these

615   signals, we apply the QCTM mode (see Sect. 2) in which chemistry and dynamics are decoupled. Accordingly, the dynamics
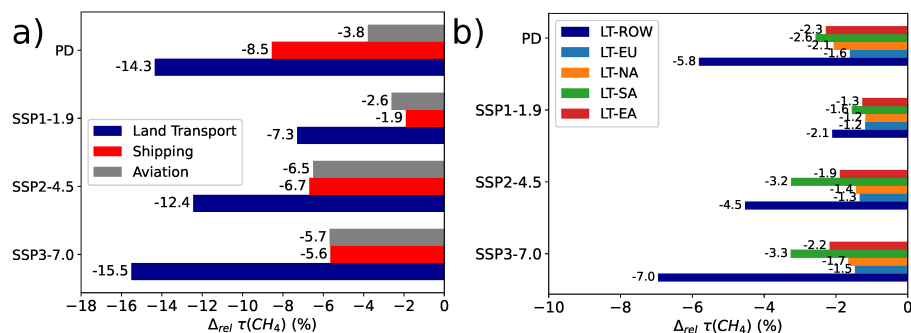
Figure 16. Methane lifetime reduction by different emission sectors based on their contributions to OH. a) Sectors land, aviation, ship, and for different SSPs, b) total transportation (all transport sectors) from different regions and for different SSPs. Note: Lifetime reductions due to individual sectors are additive. See the text for a description of the calculation. All values are 5 year averages.

(and therefore the climatic state) is identical in every model simulation. This approach of applying present day dynamics for future emission scenarios is commonly used when the effects of certain emission changes or sources on the atmospheric chemistry are investigated on the global and regional scale (e.g., Eyring et al., 2007; Hoor et al., 2009; Hodnebrog et al., 2012; Righi et al., 2015; Matthias et al., 2016). Due to this approach, however, emissions which are based on meteorological conditions (e.g. biogenic emissions, lightning-NO_x) are identical in all simulations. With climate change, these emissions are likely to increase (von Schneidemesser et al., 2015). This increase could alter the contributions of the anthropogenic emissions, for instance increased biogenic VOC emissions may affect the ozone production efficiency, while increased lightning-NO_x in the upper troposphere may compete with NO_x emissions from the aviation sector. The atmospheric kinetics and atmospheric transport patterns are also altered by a changing climate. All of these effects are not considered with the approach used in this study. Koffi et al. (2010a) considered the effects of climate change on the ozone effects of transport emissions applying a 5 % emissions reduction. Globally, they report a small decrease of the ozone changes caused by transport emissions due to climate change, but with strongly varying regional patterns.

To be able to estimate only the effect of the changing anthropogenic emissions of short lived species, we also kept emissions of methane constant in all model simulations. We chose this approach to be able to attribute and understand the effects of the emission changes in more detail. Increasing methane levels strongly affect atmospheric chemistry. In addition, increasing methane levels can have very complex effects on ozone produced form a specific emission sources. As an example, Butler et al. (2020) reported a strong increase of ozone attributable to shipping emissions with increasing methane levels, while the contribution of other emission sources showed a smaller increase. To suppress such complex effects, we kept the CH_4 emissions constant in all simulations. However, to investigate how increased methane levels would affect the results, we performed an additional sensitivity study for 2050 taking short-lived emissions and prescribed methane mixing ratios at the surface from the *SSP2-4.5* projection for 2050. In this simulation, globally averaged prescribed methane mixing ratios at the surface are around 9 % larger compared to the *SSP2-4.5* simulation. This increase leads to ≈ 2.5 % (or 0.9 DU) increased globally averaged tropospheric ozone columns. Similar as reported by Butler et al. (2020), the contributions show a very heterogeneous response.

The contribution to tropospheric ozone of emission categories with strong NO_x emissions increases, with the largest increases by the categories lightning (3.7 %), N_2O degradation (3.1 %), aviation (2.9 %), and shipping (2.4 %). In addition, ozone attributed to methane degradation also increases by 9.1 % (see also Sect. S6 in the Supplement). Moreover, the response of the methane increase in our study differs from the results reported by Butler et al. (2020), which is caused by differences in the tagging approach (see next paragraph) and by different emissions inventories and different magnitude of the methane increases.

The tagging method which we applied for source attribution is a diagnostic method. The method itself is mathematically correct, but it has some simplifications which influence the results, as discussed in detail by Mertens et al. (2018) and Butler et al. (2018). In addition, different ozone source attribution methods exist. Our approach attributes NO_x and VOC emissions at the same time, while the approach of Butler et al. (2018) either tags NO_x or VOC. Especially for the shipping sector, but most likely also for the aviation and the land transport sector, which all have rather large NO_x emissions and low VOC emissions, our approach leads to lower contributions compared to the NO_x tagging approach (see also Butler et al., 2020).

Finally, we would like to point out that we used the original aviation emissions from CMIP6. Thor et al. (2023) reported a discrepancy in the geographical distribution of the aviation emissions in CMIP6. This discrepancy leads to overestimated emissions in Arctic and Antarctic regions and underestimated emissions in the tropics. Correcting for this discrepancy in the 2015 emissions leads to an around 8% larger ozone radiative forcing and 6 % larger relative reduction of the methane lifetime of the aviation emissions compared to the values discussed in the present manuscript (see Thor et al., 2023, for detailed analyses).

Future studies should aim to consider the effects of changes methane emissions and the effects due to climate change on the contribution of transport emissions on ozone and OH and the associated radiative forcings and reductions of the methane lifetime. Due to the effects discussed above, such studies need a much more complicated design of the numerical experiments. Moreover, much longer integration times needs to be taken into account to yield signals of the transport emissions larger than the internal variability (see for example Koffi et al., 2010b). As the response of atmospheric chemistry, and more importantly also natural emissions, of climate change vary (e.g. Turnock et al., 2020) a multi-model approach would be of interest here, however only few models feature ozone tagging approaches as used in the present study.

8 Conclusions and Outlook

This manuscript presents an analysis of the effects of transport emissions from the sectors land transport, aviation, and shipping on ozone and methane lifetime. We investigated present day conditions for 2015 and three different projections of the emissions for 2050 according to the Shared Socioeconomic Pathway (SSPs) SSP1-1.9, SSP2-4.5 and SSP3-7.0. To do so, we performed simulations with the chemistry-climate model EMAC for 5 years each. Atmospheric dynamics are kept identical in order to quantify the small signals of specific transport sectors. For the source attribution of the emission effects on ozone, OH, HO_2 and ozone related precursors, we used a tagging method. Based on the contributions of the specific emissions to ozone and OH, we are able to calculate the radiative forcings of ozone and the relative reduction of the methane lifetime for all transport sectors. The calculation of the reduction of the methane lifetime based on the OH contribution is a novelty. Previous studies used the so-called perturbation approach.

Our analyses of the model results can be summarised in the following most important findings:

– Globally, transport emissions are reduced strongly in the SSP1 projection, following the high share of public transport, the strong technological advancements including adaption of newer technologies, increased use of high speed trains and overall reduced demand for transportation assumed in this scenario. In the SSP2 projection the emissions of the land transport sector decrease compared to 2015 in most regions except South Asia. Here, the large increase in demand for mobility caused by increasing population and insufficient pollution control and limited technological advancements lead to an increase of land transport emissions. Also global NO_x emissions from aviation are doubled compared to 2015 due to increasing demands, while the shipping emissions are reduced slightly. In the SSP3 projection only Europe and North America show a larger decrease of land transport emissions in 2050 compared to 2015, while emissions increase in most other regions. The aviation emissions are similar as in the SSP2 projection and shipping emissions are reduced compared to SSP2 and to 2015, most likely due to international trade barriers in the narrative of the scenario.

– Due to the non-linearity of atmospheric chemistry, the response of atmospheric composition to the changed emissions is rather complex. In accordance with previous publications, we find a large efficiency of the ozone formation from aviation emissions, while ozone from land transport emissions in Europe and North America is formed rather inefficient. A similar behaviour is also observed for OH, which is strongly bound to the NO_x-O_3 chemistry.

– Due to the strong decrease of emissions in SSP1 the non-linearity of the NO_x-O_3 chemistry compensates for some of the emission reductions. However, the total decrease of the emissions is so large that the contributions of the transport emissions to ozone and the associated radiative forcing decrease strongly. The non-linearity of the OH-response and the associated relative reduction of the methane lifetimes is stronger as for the O_3 response. Accordingly, the (cooling) effect of the transport emissions on the climate by the decrease of the methane lifetime can get more important compared to the warming effect of O_3, because the reduction of the methane lifetime is larger as the increase of ozone.

– In the SSP2 projection the land transport emissions are reduced strongly in many regions, leading to a large reduction of the ozone contributions and associated radiative forcings. However, ozone formed from land transport emissions over South Asia have rather large radiative forcings as this ozone is transported in the upper troposphere very efficiently. Therefore, the radiative forcing from global land transport emissions is reduced by only 15 %. Although the total NO_x emissions of the aviation sector are low compared to the land transport sector, the doubling of the aviation NO_x emissions in SSP2 compared to 2015 almost compensates for the reductions in the land transport and the shipping sector in terms of ozone RF. Accordingly, the total radiative forcing of the transport sector is similar in 2050 for SSP2 compared to 2015. The contributions to the methane life-time show a similar behaviour.

– In the SSP3 projection land transport emissions in Europe and North America are reduced, but the non-linearity of the ozone chemistry strongly compensates for this reduction and the ozone contributions are not reduced likewise. In addition, due to the lack of international co-operations and increasing emission in other parts of the world, the contribution of ozone from land transport transported via long-range transport increases over Europe and North America with SSP3 in

2050 compared to 2015. The global increase of land transport emissions, especially in South Asia, leads to an increase of the ozone radiative forcing from land transport emissions. Moreover, also the doubling of aviation emissions compared to 2015 leads to an overall larger ozone radiative forcing and larger contributions on the methane lifetime of the transport sector in 2050 compared to 2015.

– Overall, the land transport sector is the transport sector with the largest emissions and the largest effects on ozone and OH. In 2015, this is followed by shipping and aviation. Due to the increase of aviation emissions in SSP2 and SSP3 compared to 2015, caused by increased demand, and the reduction of shipping emissions in SSP2 and SSP3 compared to 2015, the effects of the aviation sector to ozone radiative forcing and methane lifetime are larger compared to the shipping sector.

– Already for present day land transport emissions from South Asia strongly contribute to tropospheric ozone. This contribution is projected to increase in SSP2 and SSP3. Given the large radiative efficiency of ozone formed from this emissions and their negative effects on air quality, very urgent mitigation policies aiming at a reduction of land transport emissions in this region are needed.

– The complex, non-linear atmospheric responses of O_3 and OH highlight the necessity to investigate effects of possible mitigation options for the transport sector using state of the art atmospheric chemistry models. These effects are overlooked if only emission reductions are analysed.

– The application of the source attribution method for OH and the diagnosed reductions of the methane lifetime yields to results which deviate strongly from the results of the perturbation approach. With the 5 % perturbation the shipping sector leads to a larger reduction of the methane lifetime than the land transport and aviation sector. With the source attribution method emissions from the land transport sector lead to the largest reduction of the methane lifetime (around 7 times larger compared to a 5 % perturbation), followed by the shipping and the aviation sector. Such larger reductions of the methane lifetime also influence the total radiative forcing caused by the NO_x emissions which will be analysed in a follow up study.

Due to the very complex response of atmospheric chemistry on emission changes, we did not consider the effects of climate change in our simulations. Also the effects of increased methane emissions was investigated in a short sensitivity simulation only. Future studies should incorporate both effects. Especially, for example changes in lightning-NO_x emissions and biogenic emissions could also strongly influence contributions of land transport emissions to the atmospheric composition. Moreover, in very strong mitigation cases such as the SSP1, the role of these natural emissions will be more and more important due to the non-linearitiy of the ozone chemistry, as the effects of the transport emissions strongly depend on the background conditions. Such a study, however, would need much longer model integration times and due to different responses of atmospheric composition by climate change by different models a multi-model approach would be very valuable here. The lack of publicly available information on the detailed assumptions and trends in the transport sectors from the SSP narratives complicated

the detailed interpretation of some of the results. For future studies it would therefore be very valuable if such important information on the developments in the transport sector would also be available publicly.

Code and data availability. MESSy is continuously developed and applied by a consortium of institutions. MESSy and the source code
740 are licensed to all affiliates of institutions which are members of the MESSy Consortium. Institutions can become members of the MESSy Consortium by signing the MESSy Memorandum of Understanding. More information can be found on the MESSy Consortium website (http://www.messy-interface.org, last access: 27 November 2023). The model configuration discussed in this paper is based on EMAC version 2.55. The exact set-up used to produce the results of this paper is archived at the German Climate Computing Center (DKRZ) and can be made available to members of the MESSy community upon request. The model results analysed in this study are available on request
745 and will be made available during the peer-review process.

Author contributions. M.M. performed the model simulations and led the evaluation of the model data, S.B. assisted with the emissions processing and performed the methane lifetime analyses, V.G. and V.R. developed the improved HO_x TAGGING needed for this study, J.H. designed and coordinated the experimental setup, P.J. supported the necessary model development and methane lifetime analyses, P.G. supported the model setup and emission processing, A.L. performed analyses of the ozone non-linearity and emission, S.M., V.G. and R.N.T.
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References

760 Archibald, A. T., Neu, J. L., Elshorbany, Y. F., Cooper, O. R., Young, P. J., Akiyoshi, H., Cox, R. A., Coyle, M., Derwent, R. G., Deushi, M., Finco, A., Frost, G. J., Galbally, I. E., Gerosa, G., Granier, C., Griffiths, P. T., Hossaini, R., Hu, L., Jöckel, P., Josse, B., Lin, M. Y., Mertens, M., Morgenstern, O., Naja, M., Naik, V., Oltmans, S., Plummer, D. A., Revell, L. E., Saiz-Lopez, A., Saxena, P., Shin, Y. M., Shahid, I., Shallcross, D., Tilmes, S., Trickl, T., Wallington, T. J., Wang, T., Worden, H. M., and Zeng, G.: Tropospheric Ozone Assessment Report: A critical review of changes in the tropospheric ozone burden and budget from 1850 to 2100, Elementa: Science of the Anthropocene, 8,

765 034, https://doi.org/10.1525/elementa.2020.034, 2020.

Bauer, N., Calvin, K., Emmerling, J., Fricko, O., Fujimori, S., Hilaire, J., Eom, J., Krey, V., Kriegler, E., Mouratiadou, I., Sytze de Boer, H., van den Berg, M., Carrara, S., Daioglou, V., Drouet, L., Edmonds, J. E., Gernaat, D., Havlik, P., Johnson, N., Klein, D., Kyle, P., Marangoni, G., Masui, T., Pietzcker, R. C., Strubegger, M., Wise, M., Riahi, K., and van Vuuren, D. P.: Shared Socio-Economic Pathways of the Energy Sector – Quantifying the Narratives, Global Environmental Change, 42, 316–330,

770 https://doi.org/https://doi.org/10.1016/j.gloenvcha.2016.07.006, 2017.

Bellouin, N., Quaas, J., Gryspeerdt, E., Kinne, S., Stier, P., Watson-Parris, D., Boucher, O., Carslaw, K. S., Christensen, M., Daniau, A.-L., Dufresne, J.-L., Feingold, G., Fiedler, S., Forster, P., Gettelman, A., Haywood, J. M., Lohmann, U., Malavelle, F., Mauritsen, T., McCoy, D. T., Myhre, G., Mülmenstädt, J., Neubauer, D., Possner, A., Rugenstein, M., Sato, Y., Schulz, M., Schwartz, S. E., Sourdeval, O., Storelvmo, T., Toll, V., Winker, D., and Stevens, B.: Bounding Global Aerosol Radiative Forcing of Climate Change, Reviews

775 of Geophysics, 58, e2019RG000 660, https://doi.org/https://doi.org/10.1029/2019RG000660, e2019RG000660 10.1029/2019RG000660, 2020.

Butler, T., Lupascu, A., Coates, J., and Zhu, S.: TOAST 1.0: Tropospheric Ozone Attribution of Sources with Tagging for CESM 1.2.2, Geosci. Model Dev., 11, 2825–2840, https://doi.org/10.5194/gmd-11-2825-2018, 2018.

Butler, T., Lupascu, A., and Nalam, A.: Attribution of ground-level ozone to anthropogenic and natural sources of nitrogen oxides and

780 reactive carbon in a global chemical transport model, 20, 10 707–10 731, https://doi.org/10.5194/acp-20-10707-2020, 2020.

Clappier, A., Belis, C. A., Pernigotti, D., and Thunis, P.: Source apportionment and sensitivity analysis: two methodologies with two different purposes, Geosci. Model Dev., 10, 4245–4256, https://doi.org/10.5194/gmd-10-4245-2017, 2017.

Dahlmann, K., Grewe, V., Ponater, M., and Matthes, S.: Quantifying the contributions of individual NOx sources to the trend in ozone radiative forcing, Atmos. Environ., 45, 2860–2868, https://doi.org/http://dx.doi.org/10.1016/j.atmosenv.2011.02.071, 2011.

785 Deckert, R., Jöckel, P., Grewe, V., Gottschaldt, K.-D., and Hoor, P.: A quasi chemistry-transport model mode for EMAC, Geosci. Model Dev., 4, 195–206, https://doi.org/10.5194/gmd-4-195-2011, 2011.

Dee, D. P., Uppala, S. M., Simmons, A. J., Berrisford, P., Poli, P., Kobayashi, S., Andrae, U., Balmaseda, M. A., Balsamo, G., Bauer, P., Bechtold, P., Beljaars, A. C. M., van de Berg, L., Bidlot, J., Bormann, N., Delsol, C., Dragani, R., Fuentes, M., Geer, A. J., Haimberger, L., Healy, S. B., Hersbach, H., Hólm, E. V., Isaksen, L., Kållberg, P., Köhler, M., Matricardi, M., McNally, A. P., Monge-Sanz, B. M., Mor-

790 crette, J.-J., Park, B.-K., Peubey, C., de Rosnay, P., Tavolato, C., Thépaut, J.-N., and Vitart, F.: The ERA-Interim reanalysis: configuration and performance of the data assimilation system, Quart. J. Roy. Meteor. Soc., 137, 553–597, https://doi.org/10.1002/qj.828, 2011.

Derwent, R. G., Collins, W. J., Johnson, C. E., and Stevenson, D. S.: 49, 463–487, https://doi.org/https://doi.org/10.1023/A:1010648913655, 2001.

Dietmüller, S., Jöckel, P., Tost, H., Kunze, M., Gellhorn, C., Brinkop, S., Frömming, C., Ponater, M., Steil, B., Lauer, A., and Hendricks, J.: A
new radiation infrastructure for the Modular Earth Submodel System (MESSy, based on version 2.51), Geosci. Model Dev., 9, 2209–2222,
https://doi.org/10.5194/gmd-9-2209-2016, 2016.

Eyring, V., Stevenson, D. S., Lauer, A., Dentener, F. J., Butler, T., Collins, W. J., Ellingsen, K., Gauss, M., Hauglustaine, D. A., Isaksen, I.
S. A., Lawrence, M. G., Richter, A., Rodriguez, J. M., Sanderson, M., Strahan, S. E., Sudo, K., Szopa, S., van Noije, T. P. C., and Wild,
O.: Multi-model simulations of the impact of international shipping on Atmospheric Chemistry and Climate in 2000 and 2030, Atmos.
Chem. Phys., 7, 757–780, https://doi.org/10.5194/acp-7-757-2007, 2007.

Feng, L., Smith, S. J., Braun, C., Crippa, M., Gidden, M. J., Hoesly, R., Klimont, Z., van Marle, M., van den Berg, M., and van der Werf,
G. R.: The generation of gridded emissions data for CMIP6, Geoscientific Model Development, 13, 461–482, https://doi.org/10.5194/gmd-
13-461-2020, 2020.

Fleming, Z. L., Doherty, R. M., von Schneidemesser, E., Malley, C. S., Cooper, O. R., Pinto, J. P., Colette, A., Xu, X., Simpson, D., Schultz,
M. G., Lefohn, A. S., Hamad, S., Moolla, R., Solberg, S., and Feng, Z.: Tropospheric Ozone Assessment Report: Present-day ozone
distribution and trends relevant to human health, Elementa: Science of the Anthropocene, 6, 12, https://doi.org/10.1525/elementa.273,
2018.

Forster, P., Storelvmo, T., Armour, K., Collins, W., Dufresne, J.-L., Frame, D., Lunt, D., Mauritsen, T., Palmer, M., Watanabe, M., Wild, M.,
and Zhang, H.: The Earth's Energy Budget, Climate Feedbacks, and Climate Sensitivity, in: Climate Change 2021: The Physical Science
Basis. Contribution of Working Group I to the Sixth Assessment Report of the Intergovernmental Panel on Climate Change, edited by
Masson-Delmotte, V., Zhai, P., Pirani, A., Connors, S. L., Péan, C., Berger, S., Caud, N., Chen, Y., Goldfarb, L., Gomis, M. I., Huang, M.,
Leitzell, K., Lonnoy, E., Matthews, J. B. R., Maycock, T. K., Waterfield, T., Yelekçi, O., Yu, R., and Zhou, B., book section 7, Cambridge
University Press, Cambridge, UK and New York, NY, USA, https://doi.org/10.1017/9781009157896.009, 2021.

Fricko, O., Havlik, P., Rogelj, J., Klimont, Z., Gusti, M., Johnson, N., Kolp, P., Strubegger, M., Valin, H., Amann, M., Ermolieva, T., Forsell,
N., Herrero, M., Heyes, C., Kindermann, G., Krey, V., McCollum, D. L., Obersteiner, M., Pachauri, S., Rao, S., Schmid, E., Schoepp, W.,
and Riahi, K.: The marker quantification of the Shared Socioeconomic Pathway 2: A middle-of-the-road scenario for the 21st century,
Global Environmental Change, 42, 251–267, https://doi.org/https://doi.org/10.1016/j.gloenvcha.2016.06.004, 2017.

Fujimori, S., Hasegawa, T., Masui, T., Takahashi, K., Herran, D. S., Dai, H., Hijioka, Y., and Kainuma, M.:
SSP3: AIM implementation of Shared Socioeconomic Pathways, Global Environmental Change, 42, 268–283,
https://doi.org/https://doi.org/10.1016/j.gloenvcha.2016.06.009, 2017.

Gidden, M. J., Riahi, K., Smith, S. J., Fujimori, S., Luderer, G., Kriegler, E., van Vuuren, D. P., van den Berg, M., Feng, L., Klein, D., Calvin,
K., Doelman, J. C., Frank, S., Fricko, O., Harmsen, M., Hasegawa, T., Havlik, P., Hilaire, J., Hoesly, R., Horing, J., Popp, A., Stehfest,
E., and Takahashi, K.: Global emissions pathways under different socioeconomic scenarios for use in CMIP6: a dataset of harmonized
emissions trajectories through the end of the century, Geoscientific Model Development, 12, 1443–1475, https://doi.org/10.5194/gmd-12-
1443-2019, 2019.

Grange, S. K., Lee, J. D., Drysdale, W. S., Lewis, A. C., Hueglin, C., Emmenegger, L., and Carslaw, D. C.: COVID-19 lock-
downs highlight a risk of increasing ozone pollution in European urban areas, Atmospheric Chemistry and Physics, 21, 4169–4185,
https://doi.org/10.5194/acp-21-4169-2021, 2021.

Grewe, V. and Dameris, M.: Calculating the global mass exchange between stratosphere and troposphere, Annales Geophysicae, 14, 431–442,
https://doi.org/10.1007/s00585-996-0431-x, 1996.

Grewe, V., Brunner, D., Dameris, M., Grenfell, J., Hein, R., Shindell, D., and Staehelin, J.: Origin and variability of upper tropospheric nitrogen oxides and ozone at northern mid-latitudes, Atmos. Environ., 35, 3421–3433, https://doi.org/http://dx.doi.org/10.1016/S1352-2310(01)00134-0, 2001.

Grewe, V., Tsati, E., and Hoor, P.: On the attribution of contributions of atmospheric trace gases to emissions in atmospheric model applications, Geosci. Model Dev., 3, 487–499, https://doi.org/10.5194/gmd-3-487-2010, 2010.

Grewe, V., Tsati, E., Mertens, M., Frömming, C., and Jöckel, P.: Contribution of emissions to concentrations: the TAGGING 1.0 submodel based on the Modular Earth Submodel System (MESSy 2.52), Geosci. Model Dev., 10, 2615–2633, https://doi.org/10.5194/gmd-10-2615-2017, 2017.

Guenther, A., Hewitt, C. N., Erickson, D., Fall, R., Geron, C., Graedel, T., Harley, P., Klinger, L., Lerdau, M., Mckay, W. A., Pierce, T., Scholes, B., Steinbrecher, R., Tallamraju, R., Taylor, J., and Zimmerman, P.: A global model of natural volatile organic compound emissions, J. Geophys. Res., 100, 8873–8892, 1995.

Hauglustaine, D. A., Granier, C., Brasseur, G. P., and Mégie, G.: Impact of present aircraft emissions of nitrogen oxides on tropospheric ozone and climate forcing, 21, 2031–2034, https://doi.org/https://doi.org/10.1029/94GL01729, 1994.

Hodnebrog, Ø., Berntsen, T. K., Dessens, O., Gauss, M., Grewe, V., Isaksen, I. S. A., Koffi, B., Myhre, G., Olivié, D., Prather, M. J., Stordal, F., Szopa, S., Tang, Q., van Velthoven, P., and Williams, J. E.: Future impact of traffic emissions on atmospheric ozone and OH based on two scenarios, Atmos. Chem. Phys., 12, 12 211–12 225, https://doi.org/10.5194/acp-12-12211-2012, 2012.

Hoesly, R. M., Smith, S. J., Feng, L., Klimont, Z., Janssens-Maenhout, G., Pitkanen, T., Seibert, J. J., Vu, L., Andres, R. J., Bolt, R. M., Bond, T. C., Dawidowski, L., Kholod, N., Kurokawa, J.-I., Li, M., Liu, L., Lu, Z., Moura, M. C. P., O'Rourke, P. R., and Zhang, Q.: Historical (1750–2014) anthropogenic emissions of reactive gases and aerosols from the Community Emissions Data System (CEDS), 11, 369–408, https://doi.org/10.5194/gmd-11-369-2018, 2018.

Hoor, P., Borken-Kleefeld, J., Caro, D., Dessens, O., Endresen, O., Gauss, M., Grewe, V., Hauglustaine, D., Isaksen, I. S. A., Jöckel, P., Lelieveld, J., Myhre, G., Meijer, E., Olivie, D., Prather, M., Schnadt Poberaj, C., Shine, K. P., Staehelin, J., Tang, Q., van Aardenne, J., van Velthoven, P., and Sausen, R.: The impact of traffic emissions on atmospheric ozone and OH: results from QUANTIFY, Atmos. Chem. Phys., 9, 3113–3136, https://doi.org/10.5194/acp-9-3113-2009, 2009.

Janssens-Maenhout, G., Crippa, M., Guizzardi, D., Muntean, M., Schaaf, E., Dentener, F., Bergamaschi, P., Pagliari, V., Olivier, J. G. J., Peters, J. A. H. W., van Aardenne, J. A., Monni, S., Doering, U., Petrescu, A. M. R., Solazzo, E., and Oreggioni, G. D.: EDGAR v4.3.2 Global Atlas of the three major greenhouse gas emissions for the period 1970–2012, Earth System Science Data, 11, 959–1002, https://doi.org/10.5194/essd-11-959-2019, 2019.

Jöckel, P., Tost, H., Pozzer, A., Brühl, C., Buchholz, J., Ganzeveld, L., Hoor, P., Kerkweg, A., Lawrence, M., Sander, R., Steil, B., Stiller, G., Tanarhte, M., Taraborrelli, D., van Aardenne, J., and Lelieveld, J.: The atmospheric chemistry general circulation model ECHAM5/MESSy1: consistent simulation of ozone from the surface to the mesosphere, Atmos. Chem. Phys., 6, 5067–5104, https://doi.org/10.5194/acp-6-5067-2006, 2006.

Jöckel, P., Kerkweg, A., Pozzer, A., Sander, R., Tost, H., Riede, H., Baumgaertner, A., Gromov, S., and Kern, B.: Development cycle 2 of the Modular Earth Submodel System (MESSy2), Geosci. Model Dev., 3, 717–752, https://doi.org/10.5194/gmd-3-717-2010, 2010.

Jöckel, P., Tost, H., Pozzer, A., Kunze, M., Kirner, O., Brenninkmeijer, C. A. M., Brinkop, S., Cai, D. S., Dyroff, C., Eckstein, J., Frank, F., Garny, H., Gottschaldt, K.-D., Graf, P., Grewe, V., Kerkweg, A., Kern, B., Matthes, S., Mertens, M., Meul, S., Neumaier, M., Nützel, M., Oberländer-Hayn, S., Ruhnke, R., Runde, T., Sander, R., Scharffe, D., and Zahn, A.: Earth System Chemistry integrated Modelling (ES-

CiMo) with the Modular Earth Submodel System (MESSy) version 2.51, Geosci. Model Dev., 9, 1153–1200, https://doi.org/10.5194/gmd-9-1153-2016, 2016.

870 KC, S. and Lutz, W.: The human core of the shared socioeconomic pathways: Population scenarios by age, sex and level of education for all countries to 2100, Global Environmental Change, 42, 181–192, https://doi.org/https://doi.org/10.1016/j.gloenvcha.2014.06.004, 2017.

Kerkweg, A., Sander, R., Tost, H., and Jöckel, P.: Technical note: Implementation of prescribed (OFFLEM), calculated (ONLEM), and pseudo-emissions (TNUDGE) of chemical species in the Modular Earth Submodel System (MESSy), Atmos. Chem. Phys., 6, 3603–3609, https://doi.org/10.5194/acp-6-3603-2006, 2006.

875 Koffi, B., Szopa, S., Cozic, A., Hauglustaine, D., and van Velthoven, P.: Present and future impact of aircraft, road traffic and shipping emissions on global tropospheric ozone, Atmospheric Chemistry and Physics, 10, 11 681–11 705, https://doi.org/10.5194/acp-10-11681-2010, 2010a.

Koffi, B., Szopa, S., Cozic, A., Hauglustaine, D., and van Velthoven, P.: Present and future impact of aircraft, road traffic and shipping emissions on global tropospheric ozone, Atmos. Chem. Phys., 10, 11 681–11 705, https://doi.org/10.5194/acp-10-11681-2010, 2010b.

880 Lee, D., Fahey, D., Skowron, A., Allen, M., Burkhardt, U., Chen, Q., Doherty, S., Freeman, S., Forster, P., Fuglestvedt, J., Gettelman, A., León, R. D., Lim, L., Lund, M., Millar, R., Owen, B., Penner, J., Pitari, G., Prather, M., Sausen, R., and Wilcox, L.: The contribution of global aviation to anthropogenic climate forcing for 2000 to 2018, Atmospheric Environment, 244, 117 834, https://doi.org/https://doi.org/10.1016/j.atmosenv.2020.117834, 2021.

Maruhashi, J., Mertens, M., Grewe, V., and Dedoussi, I., C.: A multi-method assessment of the regional sensitivities between flight altitude
885 and short-term climate warming from aircraft NOx emissions, submitted to Environmental Research Letters, xx, xx, 2024.

Matthias, V., Aulinger, A., Backes, A., Bieser, J., Geyer, B., Quante, M., and Zeretzke, M.: The impact of shipping emissions on air pollution in the greater North Sea region – Part 2: Scenarios for 2030, Atmos. Chem. Phys., 16, 759–776, https://doi.org/10.5194/acp-16-759-2016, 2016.

Matthias, V., Quante, M., Arndt, J. A., Badeke, R., Fink, L., Petrik, R., Feldner, J., Schwarzkopf, D., Link, E.-M., Ramacher, M. O. P.,
890 and Wedemann, R.: The role of emission reductions and the meteorological situation for air quality improvements during the COVID-19 lockdown period in central Europe, Atmospheric Chemistry and Physics, 21, 13 931–13 971, https://doi.org/10.5194/acp-21-13931-2021, 2021.

McDuffie, E. E., Smith, S. J., O'Rourke, P., Tibrewal, K., Venkataraman, C., Marais, E. A., Zheng, B., Crippa, M., Brauer, M., and Martin, R. V.: A global anthropogenic emission inventory of atmospheric pollutants from sector- and fuel-specific sources (1970–2017): an
895 application of the Community Emissions Data System (CEDS), 12, 3413–3442, https://doi.org/10.5194/essd-12-3413-2020, 2020.

Mertens, M., Grewe, V., Rieger, V. S., and Jöckel, P.: Revisiting the contribution of land transport and shipping emissions to tropospheric ozone, Atmos. Chem. Phys., 18, 5567–5588, https://doi.org/10.5194/acp-18-5567-2018, 2018.

Mertens, M., Kerkweg, A., Grewe, V., Jöckel, P., and Sausen, R.: Are contributions of emissions to ozone a matter of scale?, Geosci. Model Dev., 13, 363–383, https://doi.org/10.5194/gmd-13-363-2020, 2020a.

900 Mertens, M., Kerkweg, A., Grewe, V., Jöckel, P., and Sausen, R.: Attributing ozone and its precursors to land transport emissions in Europe and Germany, 20, 7843–7873, https://doi.org/10.5194/acp-20-7843-2020, 2020b.

Mertens, M., Jöckel, P., Matthes, S., Nützel, M., Grewe, V., and Sausen, R.: COVID-19 induced lower-tropospheric ozone changes, 16, 064 005, https://doi.org/10.1088/1748-9326/abf191, 2021.

Mertens, M. B.: Contribution of road traffic emissions to tropospheric ozone in Europe and Germany, http://nbn-resolving.de/urn:nbn:de:
905 bvb:19-207288, 2017.

Mills, G., Pleijel, H., Malley, C. S., Sinha, B., Cooper, O. R., Schultz, M. G., Neufeld, H. S., Simpson, D., Sharps, K., Feng, Z., Gerosa, G., Harmens, H., Kobayashi, K., Saxena, P., Paoletti, E., Sinha, V., and Xu, X.: Tropospheric Ozone Assessment Report: Present-day tropospheric ozone distribution and trends relevant to vegetation, Elementa: Science of the Anthropocene, 6, 47, https://doi.org/10.1525/elementa.302, 2018.

910 Monks, P. S., Archibald, A. T., Colette, A., Cooper, O., Coyle, M., Derwent, R., Fowler, D., Granier, C., Law, K. S., Mills, G. E., Stevenson, D. S., Tarasova, O., Thouret, V., von Schneidemesser, E., Sommariva, R., Wild, O., and Williams, M. L.: Tropospheric ozone and its precursors from the urban to the global scale from air quality to short-lived climate forcer, Atmos. Chem. Phys., 15, 8889–8973, https://doi.org/10.5194/acp-15-8889-2015, 2015.

O'Neill, B. C., Kriegler, E., Ebi, K. L., Kemp-Benedict, E., Riahi, K., Rothman, D. S., van Ruijven, B. J., van Vuuren, D. P., Birkmann, J.,
915 Kok, K., Levy, M., and Solecki, W.: The roads ahead: Narratives for shared socioeconomic pathways describing world futures in the 21st century, Global Environmental Change, 42, 169–180, https://doi.org/https://doi.org/10.1016/j.gloenvcha.2015.01.004, 2017.

Putero, D., Cristofanelli, P., Chang, K.-L., Dufour, G., Beachley, G., Couret, C., Effertz, P., Jaffe, D. A., Kubistin, D., Lynch, J., Petropavlovskikh, I., Puchalski, M., Sharac, T., Sive, B. C., Steinbacher, M., Torres, C., and Cooper, O. R.: Fingerprints of the COVID-19 economic downturn and recovery on ozone anomalies at high-elevation sites in North America and Western Europe, EGUsphere, 2023,
920 1–25, https://doi.org/10.5194/egusphere-2023-1737, 2023.

Rao, S., Klimont, Z., Smith, S. J., Van Dingenen, R., Dentener, F., Bouwman, L., Riahi, K., Amann, M., Bodirsky, B. L., van Vuuren, D. P., Aleluia Reis, L., Calvin, K., Drouet, L., Fricko, O., Fujimori, S., Gernaat, D., Havlik, P., Harmsen, M., Hasegawa, T., Heyes, C., Hilaire, J., Luderer, G., Masui, T., Stehfest, E., Strefler, J., van der Sluis, S., and Tavoni, M.: Future air pollution in the Shared Socio-economic Pathways, Global Environmental Change, 42, 346–358, https://doi.org/https://doi.org/10.1016/j.gloenvcha.2016.05.012, 2017.

925 Rap, A., Richards, N. A. D., Forster, P. M., Monks, S. A., Arnold, S. R., and Chipperfield, M. P.: Satellite constraint on the tropospheric ozone radiative effect, Geophys. Res. Lett., 42, 5074–5081, https://doi.org/10.1002/2015GL064037, 2015GL064037, 2015.

Riahi, K., van Vuuren, D. P., Kriegler, E., Edmonds, J., O'Neill, B. C., Fujimori, S., Bauer, N., Calvin, K., Dellink, R., Fricko, O., Lutz, W., Popp, A., Cuaresma, J. C., KC, S., Leimbach, M., Jiang, L., Kram, T., Rao, S., Emmerling, J., Ebi, K., Hasegawa, T., Havlik, P., Humpenöder, F., Da Silva, L. A., Smith, S., Stehfest, E., Bosetti, V., Eom, J., Gernaat, D., Masui, T., Rogelj, J., Stre-
930 fler, J., Drouet, L., Krey, V., Luderer, G., Harmsen, M., Takahashi, K., Baumstark, L., Doelman, J. C., Kainuma, M., Klimont, Z., Marangoni, G., Lotze-Campen, H., Obersteiner, M., Tabeau, A., and Tavoni, M.: The Shared Socioeconomic Pathways and their energy, land use, and greenhouse gas emissions implications: An overview, Global Environmental Change, 42, 153–168, https://doi.org/https://doi.org/10.1016/j.gloenvcha.2016.05.009, 2017.

Rieger, V. S., Mertens, M., and Grewe, V.: An advanced method of contributing emissions to short-lived chemical species (OH and HO_2):
935 the TAGGING 1.1 submodel
based on the Modular Earth Submodel System (MESSy 2.53), Geosci. Model Dev., 11, 2049–2066, https://doi.org/10.5194/gmd-11-2049-2018, 2018.

Riese, M., Ploeger, F., Rap, A., Vogel, B., Konopka, P., Dameris, M., and Forster, P.: Impact of uncertainties in atmospheric mixing on simulated UTLS composition and related radiative effects, 117, https://doi.org/https://doi.org/10.1029/2012JD017751, 2012.

940 Righi, M., Hendricks, J., and Sausen, R.: The global impact of the transport sectors on atmospheric aerosol in 2030 – Part 1: Land transport and shipping, Atmos. Chem. Phys., 15, 633–651, https://doi.org/10.5194/acp-15-633-2015, 2015.

Righi, M., Hendricks, J., and Brinkop, S.: The global impact of the transport sectors on aerosol and climate under the Shared Socioeconomic Pathways (SSPs), Earth System Dynamics Discussions, 2023, 1–34, https://doi.org/10.5194/esd-2022-52, 2023.

Roeckner, E., Brokopf, R., Esch, M., Giorgetta, M., Hagemann, S., Kornblueh, L., Manzini, E., Schlese, U., and Schulzweida, U.: Sensitivity of Simulated Climate to Horizontal and Vertical Resolution in the ECHAM5 Atmosphere Model, J. Climate, 19, 3771–3791, https://doi.org/10.1175/jcli3824.1, 2006.

Sander, R., Baumgaertner, A., Gromov, S., Harder, H., Jöckel, P., Kerkweg, A., Kubistin, D., Regelin, E., Riede, H., Sandu, A., Taraborrelli, D., Tost, H., and Xie, Z.-Q.: The atmospheric chemistry box model CAABA/MECCA-3.0, Geosci. Model Dev., 4, 373–380, https://doi.org/10.5194/gmd-4-373-2011, 2011.

Schumann, U. and Huntrieser, H.: The global lightning-induced nitrogen oxides source, Atmos. Chem. Phys., 7, 3823–3907, https://doi.org/10.5194/acp-7-3823-2007, 2007.

Seinfeld, J. H. and Pandis, S. N.: Atmospheric Chemistry and Physics: From Air Pollution to Climate Change, Wiley, 2006.

Sillman, S.: The use of NOy , H2O2, and HNO3 as indicators for ozone-NOx -hydrocarbon sensitivity in urban locations, J. Geophys. Res. Atmos., 100, 14 175–14 188, https://doi.org/10.1029/94JD02953, 1995.

Stevenson, D. S., Young, P. J., Naik, V., Lamarque, J.-F., Shindell, D. T., Voulgarakis, A., Skeie, R. B., Dalsoren, S. B., Myhre, G., Berntsen, T. K., Folberth, G. A., Rumbold, S. T., Collins, W. J., MacKenzie, I. A., Doherty, R. M., Zeng, G., van Noije, T. P. C., Strunk, A., Bergmann, D., Cameron-Smith, P., Plummer, D. A., Strode, S. A., Horowitz, L., Lee, Y. H., Szopa, S., Sudo, K., Nagashima, T., Josse, B., Cionni, I., Righi, M., Eyring, V., Conley, A., Bowman, K. W., Wild, O., and Archibald, A.: Tropospheric ozone changes, radiative forcing and attribution to emissions in the Atmospheric Chemistry and Climate Model Intercomparison Project (ACCMIP), Atmospheric Chemistry and Physics, 13, 3063–3085, https://doi.org/10.5194/acp-13-3063-2013, 2013.

Stohl, A., Aamaas, B., Amann, M., Baker, L. H., Bellouin, N., Berntsen, T. K., Boucher, O., Cherian, R., Collins, W., Daskalakis, N., Dusinska, M., Eckhardt, S., Fuglestvedt, J. S., Harju, M., Heyes, C., Hodnebrog, Ø., Hao, J., Im, U., Kanakidou, M., Klimont, Z., Kupiainen, K., Law, K. S., Lund, M. T., Maas, R., MacIntosh, C. R., Myhre, G., Myriokefalitakis, S., Olivié, D., Quaas, J., Quennehen, B., Raut, J.-C., Rumbold, S. T., Samset, B. H., Schulz, M., Seland, Ø., Shine, K. P., Skeie, R. B., Wang, S., Yttri, K. E., and Zhu, T.: Evaluating the climate and air quality impacts of short-lived pollutants. Atmospheric Chemistry and Physics, Atmospheric Chemistry and Physics, 15(18), pp.10 529–10 566, https://doi.org/https://doi.org/10.5194/acp-15-10529-2015, 2015.

Szopa, S., Naik, V., Adhikary, B., Artaxo, P., Berntsen, T., Collins, W., Fuzzi, S., Gallardo, L., Kiendler-Scharr, A., Klimont, Z., Liao, H., Unger, N., and Zanis, P.: The Physical Science Basis. Contribution of Working Group I to the Sixth Assessment Report of the Intergovernmental Panel on Climate Change, chap. Short-Lived Climate Forcers, 2021.

Søvde, O. A., Matthes, S., Skowron, A., Iachetti, D., Lim, L., Owen, B., Øivind Hodnebrog, Di Genova, G., Pitari, G., Lee, D. S., Myhre, G., and Isaksen, I. S.: Aircraft emission mitigation by changing route altitude: A multi-model estimate of aircraft NOx emission impact on O3 photochemistry, Atmospheric Environment, 95, 468–479, https://doi.org/https://doi.org/10.1016/j.atmosenv.2014.06.049, 2014.

Taylor, K. E., Stouffer, R. J., and Meehl, G. A.: An Overview of CMIP5 and the Experiment Design, Bulletin of the American Meteorological Society, 93, 485 – 498, https://doi.org/https://doi.org/10.1175/BAMS-D-11-00094.1, 2012.

Terrenoire, E., Hauglustaine, D. A., Cohen, Y., Cozic, A., Valorso, R., Lefèvre, F., and Matthes, S.: Impact of present and future aircraft NO$_x$ and aerosol emissions on atmospheric composition and associated direct radiative forcing of climate, 22, 11 987–12 023, https://doi.org/10.5194/acp-22-11987-2022, 2022.

Thor, R. N., Mertens, M., Matthes, S., Righi, M., Hendricks, J., Brinkop, S., Graf, P., Grewe, V., Jöckel, P., and Smith, S.: An inconsistency in aviation emissions between CMIP5 and CMIP6 and the implications for short-lived species and their radiative forcing, Geoscientific Model Development, 16, 1459–1466, https://doi.org/10.5194/gmd-16-1459-2023, 2023.

Tost, H., Jöckel, P., Kerkweg, A., Sander, R., and Lelieveld, J.: Technical note: A new comprehensive SCAVenging submodel for global atmospheric chemistry modelling, Atmos. Chem. Phys., 6, 565–574, https://doi.org/10.5194/acp-6-565-2006, 2006.

Turnock, S. T., Allen, R. J., Andrews, M., Bauer, S. E., Deushi, M., Emmons, L., Good, P., Horowitz, L., John, J. G., Michou, M., Nabat, P., Naik, V., Neubauer, D., O'Connor, F. M., Olivié, D., Oshima, N., Schulz, M., Sellar, A., Shim, S., Takemura, T., Tilmes, S., Tsigaridis, K., Wu, T., and Zhang, J.: Historical and future changes in air pollutants from CMIP6 models, 20, 14547–14579, https://doi.org/10.5194/acp-20-14547-2020, 2020.

van Vuuren, D. P., Stehfest, E., Gernaat, D. E., Doelman, J. C., van den Berg, M., Harmsen, M., de Boer, H. S., Bouwman, L. F., Daioglou, V., Edelenbosch, O. Y., Girod, B., Kram, T., Lassaletta, L., Lucas, P. L., van Meijl, H., Müller, C., van Ruijven, B. J., van der Sluis, S., and Tabeau, A.: Energy, land-use and greenhouse gas emissions trajectories under a green growth paradigm, Global Environmental Change, 42, 237–250, https://doi.org/https://doi.org/10.1016/j.gloenvcha.2016.05.008, 2017.

von Kuhlmann, R., Lawrence, M. G., Pöschl, U., and Crutzen, P. J.: Sensitivities in global scale modeling of isoprene, Atmos. Chem. Phys., 4, 1–17, https://doi.org/10.5194/acp-4-1-2004, 2004.

von Schneidemesser, E., Monks, P. S., Allan, J. D., Bruhwiler, L., Forster, P., Fowler, D., Lauer, A., Morgan, W. T., Paasonen, P., Righi, M., Sindelarova, K., and Sutton, M. A.: Chemistry and the Linkages between Air Quality and Climate Change, Chemical Reviews, 115, 3856–3897, https://doi.org/10.1021/acs.chemrev.5b00089, pMID: 25926133, 2015.

Wang, Z. S., Chien, C.-J., and Tonnesen, G. S.: Development of a tagged species source apportionment algorithm to characterize three-dimensional transport and transformation of precursors and secondary pollutants, J. Geophys. Res. Atmos., 114, n/a–n/a, https://doi.org/10.1029/2008JD010846, d21206, 2009.

Yienger, J. J. and Levy, H.: Empirical model of global soil-biogenic NOx emissions, J. Geophys. Res. Atmos., 100, 11447–11464, https://doi.org/10.1029/95JD00370, 1995.