

Intended and Unintended Consequences of Atmospheric Methane Oxidation Enhancement

Hannah M. Horowitz¹

¹ Department of Civil and Environmental Engineering and Department of Climate, Meteorology, & Atmospheric Sciences, University of Illinois Urbana-Champaign, Urbana, Illinois, USA

5

Correspondence to: Hannah M. Horowitz (hmhorow@illinois.edu)

Abstract. Atmospheric oxidation enhancement (AOE) of methane via either tropospheric hydroxyl radicals (OH) or chlorine (Cl) radicals is being considered as a method to decrease greenhouse gas concentrations. The chemistry involved is coupled; is nonlinear; and affects air quality, other greenhouse gases, and ozone-depleting substances. Here I perform a suite of experiments in a three-dimensional (3D) atmospheric chemistry model representing different OH- and Cl-based atmospheric oxidation enhancement methods, to estimate the effectiveness of each at decreasing greenhouse gases and the impacts on air quality and stratospheric ozone. I find that larger emissions of iron salt aerosol are required relative to previous work to reduce methane on a global scale by at least a few percent (≥ 565 Tg/yr), which indicates uncertainty in predicting the effectiveness of this method depending on the representation of the reaction mechanism and modeling framework employed. More work is needed to understand the kinetics of chlorine release from iron salt aerosol and the potential for bromine co-release, which decreases effectiveness. Hydrogen peroxide-based approaches can decrease global methane, but the hydrogen peroxide emissions required may be too large to be feasible. I find that limiting emissions to daytime for hydrogen peroxide-based scenarios has negligible effects. All methods increase surface particulate matter (PM) pollution and in some regions lead to exceedances of annual air quality standards. Cl-based methods decrease ozone air pollution, but OH-based methods increase ozone air pollution in populated areas. While Cl-based methods can increase ozone-depleting substances, the 1-year timeframe of this study is insufficient to predict impacts on stratospheric ozone. The overall impacts of atmospheric oxidation enhancement methods on climate and human health involve not only their effectiveness at decreasing methane, but competing or complementary effects on other greenhouse gases and aerosol, as well as varying effects on surface air pollution.

15
20

1 Introduction

To achieve aggressive reductions in the greenhouse gas methane needed to reach climate goals, alternate strategies to emissions mitigation are being considered (NASEM, 2024). These include processes to decrease the atmospheric lifetime of methane by enhancing its main sinks (e.g., Abernethy et al., 2023; Gorham et al., 2024; Li, Meidan, et al., 2023; Ming et al., 2022; Wang et al., 2022), using oxidation by tropospheric OH (currently $>90\%$) and tropospheric Cl (currently 1–5%). There are interactions between Cl and OH; namely, increasing Cl decreases OH due to reductions in ozone (e.g., Horowitz et al., 2020; Li, Meidan, et al., 2023). In addition, Cl and OH can impact aerosol particles, other greenhouse gases, and ozone-depleting substances. Due to these interactions and the highly nonlinear chemistry involved, detailed atmospheric chemical

25
30

investigations are needed to understand the overall climate and pollution impacts of methods to enhance atmospheric methane oxidation. To have the greatest impact on atmospheric methane, methods appropriate for ambient concentrations (≤ 2 parts per million) are needed, but the technology is not yet available (Abernethy et al., 2023; NASEM, 2024; Pennacchio et al., 2024).

35 While commercial OH generators exist on a smaller scale (e.g., to remove volatile organic carbon pollution; Johnson et al., 2014), it is less well understood how to scale up the artificial production of OH globally (e.g., Ming et al., 2022). Potential technologies could involve the release of hydrogen peroxide, downdraft energy towers, and/or artificial ultraviolet (UV) radiation (Tao et al., 2023; Wang et al., 2022). One specific method to release Cl atom is iron salt aerosol (e.g., Oeste et al., 2017), where the presence of both iron (III) and chloride enhances photolytic Cl atom release. Photolytic Cl atom release from
40 artificial salts has been demonstrated in chamber experiments using salt pans and artificial sea salt aerosols generated from sodium chloride (NaCl) and seawater samples (Wittmer et al., 2015a,b) and in laboratory experiments (Mikkelsen et al., 2024). Recently, this process has been hypothesized to occur in the atmosphere in a study based on observations from Barbados through the mixing of natural iron-containing dust and sea salt aerosols and is thought to be due to the release of chlorine gas (Cl₂) followed by rapid photolysis (van Herpen et al., 2023). Gorham et al. (2024) present the state of the science for moving
45 forward with intentional increase of this mechanism via iron salt aerosol. Meidan et al. (2024) find that the impact of iron varies depending on the region of application. Recent work has investigated the direct release of Cl₂ in a coupled chemistry–climate model and found that at least 90 Tg/yr Cl₂ is needed for Cl increases to outweigh decreases in OH with respect to overall methane loss, and at least 1,250 Tg/yr Cl₂ is needed to decrease the methane lifetime by 50% or more (Li, Meidan, et al., 2023).

50 Here I simulate in an atmospheric model a variety of methods of atmospheric oxidation enhancement (AOE) of methane via tropospheric OH and tropospheric Cl. I assess their impacts on methane, tropospheric chemistry, other greenhouse gases and ozone-depleting substances, and surface air quality.

2 Methods

I apply the global atmospheric chemical transport model GEOS-Chem (Section 2.1) with modifications (Section 2.2)
55 to simulate atmospheric oxidation enhancement scenarios and their effects. The overall methodology of the study is outlined in Figure S1.

2.1 GEOS-Chem Model

Here I apply the 3D atmospheric chemical transport model GEOS-Chem version 13.2.1. (<https://doi.org/10.5281/zenodo.5500717>). As it is a chemical transport model, meteorological and climatic processes are not
60 simulated directly. Instead, the model is driven by assimilated offline meteorological fields of the Modern-Era Retrospective Analysis for Research and Applications (Gelaro et al., 2017) from the National Aeronautics and Space Administration/Global Modeling and Assimilation Office. Simulations are performed at 4° latitude × 5° longitude horizontal resolution, with 72

vertical layers from the surface to the mesosphere (up to 80 km) and with active chemistry through the stratosphere (up to 50 km). Simulations are performed for the arbitrary year 2019 following 1 year of initialization (year 2018).

65 Detailed coupled tropospheric halogen (chlorine, bromine, iodine) chemistry follows Wang et al. (2021) and includes sea salt debromination, acid displacement, and improvements to heterogeneous chemistry on polar stratospheric clouds (Eastham et al., 2014). Online stratospheric chemistry includes heterogeneous ozone depletion chemistry. Carbonaceous aerosol includes black carbon (Wang et al., 2014) and organic aerosol following the “simple” secondary organic aerosol (SOA) scheme (fixed-yield, direct, and irreversible formation) (Pai et al., 2020). Anthropogenic emissions follow the Community
70 Emissions Data System (CEDS) v2 inventory originally developed for the Coupled Model Intercomparison Project Phase 6 (CMIP6) (<https://data.pnnl.gov/dataset/CEDS-4-21-21>), including particulate iron emitted as a constant fraction of sulfur dioxide (SO₂). The speciation and solubility of iron in GEOS-Chem is discussed in Section 2.2.2.3. Sea salt aerosol emissions over the open ocean are wind- and sea surface temperature–dependent following Jaeglé et al. (2011), and in polar regions include blowing snow (Huang & Jaeglé, 2017). Dust emissions include natural (Fairlie et al., 2007) and anthropogenic dust
75 from the anthropogenic fugitive, combustion, and industrial dust inventory (Philip et al., 2017). Biogenic volatile organic compound emissions are from MEGAN v2.1 (Guenther et al., 2012). All meteorologically dependent emissions are calculated offline at the native model resolution (0.5° × 0.625°) (Weng et al., 2020) and appropriately scaled such that the total emissions are independent of model resolution (Lin et al., 2021). Wet deposition follows Amos et al. (2012) for gases and Liu et al. (2001) for aerosols, with snow and mixed precipitation scavenging from Wang et al. (2014). Dry deposition is a resistance-in-series
80 approach (Wang et al., 1998) with aerosol dry deposition described in Zhang et al. (2001). Ozone deposition to the ocean via reaction with sea surface iodide follows Pound et al. (2020). Planetary boundary layer mixing follows a non-local scheme (as opposed to assuming full mixing) as implemented by Lin and McElroy (2010). The first 2 kilometers contain 14 vertical levels, with gridbox heights increasing from 120m at the surface to 250m by 2km.

Methane concentrations at the surface in GEOS-Chem are a fixed boundary condition based on monthly mean
85 National Oceanic and Atmospheric Administration flask observations, after which methane advects and participates in chemistry (Murray, 2016). Hence, I use the methane feedback factor, which accounts for the feedback of methane on its own loss rate to estimate changes in steady-state methane concentrations. Impacts on steady-state methane concentrations estimated using fixed boundary conditions along with the methane feedback factor are within 10% of impacts predicted from long (>40 years) simulations reaching equilibrium with fully responsive surface methane fluxes, while significantly reducing the
90 computational cost (Khodayari et al., 2015). Fixed boundary conditions combined with the feedback factor also have been found to have a negligible impact on estimates of the global warming potential of hydrogen accounting for indirect methane feedbacks (Warwick et al., 2023). Here, I first calculate the methane lifetime following the same methods as in Horowitz et al. (2020) from Holmes et al. (2013) and Holmes (2018). Briefly, the partial lifetimes of methane against tropospheric OH and Cl are calculated from the integrated 3D reaction rates. The total methane lifetime includes these two losses, stratospheric loss
95 (assumed 120 years), and uptake to soil (assumed 150 years). Then, the change in steady-state methane concentration is estimated using the feedback factor f following Holmes (2018):

$$\Delta[\text{CH}_4] = \left(\left(\frac{\tau_{exp}}{\tau_0} \right)^f - 1 \right) [\text{CH}_4]_0 \quad (1)$$

where τ_0 and τ_{exp} are the lifetimes of methane in the standard and atmospheric oxidation enhancement experiments, respectively; $f = 1.34 \pm 0.06$ is the GEOS-Chem methane feedback factor on its loss rate (Holmes et al., 2013); and $[\text{CH}_4]_0 = 1,866.58$ parts per billion (ppb) is the global annual mean surface methane concentration in year 2019 (https://gml.noaa.gov/ccgg/trends_ch4/).

2.2 Model Experiments for Atmospheric Oxidation Enhancement

A summary of simulations is presented in Figure S2. Scenarios are grouped between the dominant intended effect on increasing methane oxidation, either through reaction with OH (OH-based; Section 2.2.1) or Cl (Cl-based; Section 2.2.2). Under these umbrellas are experiments with different compounds emitted (OH or hydrogen peroxide for OH-based, and Cl₂ or particulate iron and/or chloride for Cl-based). All emissions are constant in time except for daytime-only tests (for simplicity, double the 24-hour emissions rate released between 6am and 6pm local time) for several of the hydrogen peroxide emissions scenarios, as marked in Figure S2.

For the remainder of the paper, I will focus on scenarios detailed in Table 1, which produced nonnegligible results. Direct OH release scenarios will be discussed briefly in the context of comparisons with the hydrogen peroxide emissions scenarios that led to comparable impacts. Details on the remaining simulations are included in the supplementary information; see Tables S1 and S2.

Table 1: Details of focus scenarios in the current study.

	Emitted Species	Total Emissions (Tg/yr)	Emissions Location	Emissions Rate at Location of Emissions (kg/m ² /s)	Reaction Coefficient	Rate
H ₂ O ₂ _high	H ₂ O ₂	1.61E7	Globally at surface	1.00E-6	N/A	
H ₂ O ₂ _mid	H ₂ O ₂	7.4E4	Globally at surface	4.6E-9	N/A	
H ₂ O ₂ _low	H ₂ O ₂	1,250	Globally at surface	7.7E-11	N/A	
Cl ₂ _ocean	Cl ₂	1,250	Oceans	1.10E-10	N/A	
Cl ₂ _BrCl_Br ₂	Cl ₂	1,193		1.045E-10		
	BrCl	187	Oceans	1.137E-11	N/A	
	Br ₂	129		1.64E-11		
Iron	pFe*	565	Oceans	4.97E-11		
Iron_Max	pFe	1,250	Oceans	1.10E-10		
Chloride	Accumulation mode chloride	1,250	Oceans	1.10E-10	d[Cl ₂]/dt= α _{jNO₂} [Fe ³⁺][Cl]S	
Iron_Chloride	Accumulation mode chloride	1,250		1.10E-10		
	pFe	565	Oceans	4.97E-11		

115 *pFe is particulate iron.

2.2.1 Hydroxyl Radical

2.2.1.1 Hydrogen peroxide emission

120 Hydrogen peroxide photolysis produces two OH radicals, and the emission of hydrogen peroxide has been proposed as a possible method to enhance atmospheric methane oxidation via OH (Wang et al., 2022). For example, there is a patent to deploy hydrogen peroxide towers for this purpose (Bell, 2023). The concurrent production of hydrogen peroxide and release of OH via Fenton-like catalytic processes has also been suggested as a potential technology (Wang et al., 2022).

In this study, I tested four scenarios to investigate the impact if the current global demand of 4.1 Tg/yr H₂O₂ (Research and Markets, 2023) was released additionally (see Figure S2): globally at the surface, only at major point sources of oil and gas emissions determined from the CEDS anthropogenic emissions inventory (0.2% of Earth's area at the 4° × 5° model

125 resolution) at the surface, 600 m (stack height proposed by the Bell, 2023 patent), and 600 m emitted only during daytime to
maximize photolysis. All 4.1 Tg/yr scenarios produced negligible effects on the methane lifetime. The change to 600 m stack
height as well as from 24-hour to daytime-only did slightly increase the impact on methane (by <<1%) but still resulted in
<0.5 ppb ($\leq 0.03\%$) decrease in methane. As the spatial resolution is low ($4^\circ \times 5^\circ$ or 400–500 km at midlatitudes), this current
model study may not address fully whether this method is viable for individual, large point sources of methane. Given the
130 negligible results found here, these scenarios were not investigated further in this study. Another study using the GEOS-Chem
model investigated daytime-only 600 m H₂O₂ towers in more detail over North America at a higher spatial resolution ($0.5^\circ \times$
 0.625° or ~50 km), and found that even widespread towers at emissions rates 10x higher than what is currently proposed would
lead to negligible impact on global methane (Mayhew and Haskins, 2025). This is in part because the fraction of H₂O₂
converted to OH is ~30-60% but is driven by the small fraction of the produced OH that reacts with methane (23% at the most
135 and frequently much less) (Mayhew and Haskins, 2025).

For the remainder of the paper, I will focus on three global hydrogen peroxide emissions scenarios to span a range of
impacts on methane and feasibility (H₂O₂_high, H₂O₂_mid, and H₂O₂_low; see Table 1). H₂O₂_low emits the same mass as
the Cl₂ experiment (1250 Tg/yr; see Section 2.2.2.1) for comparison purposes. Short-term tests were performed with increasing
H₂O₂ emissions to estimate the quantity needed such that methane would decline by at least 50% (leading to H₂O₂_high). All
140 scenarios were emitted at the surface (0 to 12.3 m altitude in GEOS-Chem). In the 4.1 Tg/yr point source tests, impacts
increased when the hydrogen peroxide was emitted at 600 m with respect to the surface. It is thus possible that the impacts on
methane by these global scenarios could be larger if emitted at 600 m, potentially due to a longer lifetime for H₂O₂ allowing
for greater photolysis; this would require further investigation. I also tested H₂O₂_mid and H₂O₂_low with double the 24-hour
emissions rate only released during 6am–6pm local time. Daytime-only emissions did increase the impact on methane, with a
145 larger relative effect for the smaller hydrogen peroxide emissions, but the absolute values of methane lifetime and steady-state
methane concentrations are within $\pm 1\%$ for the corresponding daytime-only and 24-hour emissions scenarios. Thus, I will
focus on results from the 24-hour emissions for the remainder of the paper.

2.2.1.2 OH chemical production

150 Not all hydrogen peroxide is immediately photolyzed to produce OH and may undergo alternate reactions. Estimates
include the previously described GEOS-Chem study which found approximately 30-60% of H₂O₂ on average is photolyzed,
but this is highly spatially variable (Mayhew and Haskins, 2025), and a conservative theoretical estimate of 10% based on all
potential chemical and physical pathways (Pennacchio et al., 2024). Thus, to examine the direct effects of OH release in the
absence of other chemical changes, I introduced a dummy reaction in GEOS-Chem to produce OH from O₂, a species with
155 fixed concentrations in GEOS-Chem (0.2095 mol/mol), from the surface up to 1 km altitude globally (see Table S2 in the
supplementary information). OH cannot be advected in GEOS-Chem due to its short lifetime and thus it cannot be emitted
directly. I tested two rates, OH_mid and OH_high; OH_high's rate of OH emission is two times higher (see Table S2) to result
in similar changes to OH as the hydrogen peroxide simulations (since H₂O₂ photolyzes to produce 2OH). In reality, OH could

be released directly through methods incorporating artificial UV radiation (e.g., Ming et al., 2022) or downdraft energy towers that generate electricity from seawater and sunlight and produce additional OH from ozone due to the water vapor introduced in dry regions (Tao et al., 2023; Wang et al., 2022); these would likely be point sources that are not feasible in the current modeling framework. Given this and the fixed methane boundary conditions, the direct OH results will only be discussed in reference to how air quality impacts change when OH versus hydrogen peroxide is emitted at comparable impacts on methane.

2.2.2 Chlorine

2.2.2.1 Direct Cl₂ emission

In the Cl₂ simulation, Cl₂ is directly emitted across the global oceans at the surface, a total of 1,250 Tg/yr, which is the midrange scenario for methane removal in Li, Meidan et al. (2023) (resulting in 45% reduction in the methane burden in their study). This is approximately 20 times higher than the current total tropospheric source of gas-phase inorganic chlorine in GEOS-Chem (54 Tg/yr; Wang et al., 2021) or the current manufacture of Cl₂ (58 Tg/yr; World Chlorine Council). Emissions of Cl₂ at the surface are more likely to be economically and technologically feasible and will limit the impact of additional chlorine on stratospheric ozone (Li, Meidan, et al., 2023). Cl₂ will then photolyze to produce two Cl atoms, of which approximately 20% will react with ozone instead of methane over the oceans (Li, Meidan, et al., 2023). Table S4 compares the modeling setup of the current study and Li, Meidan et al. (2023), including differences in resolution, year, and halogen chemistry.

2.2.2.2 Bromine contamination

It is not possible to remove 100% of bromide from pure chlorine salts (for example, Sigma-Aldrich S9888 $\geq 99.0\%$ NaCl specifications include up to 0.01% Br⁻; <https://www.sigmaaldrich.com/>). This could lead to reactive bromine release that would decrease OH in the atmosphere through the destruction of ozone, the primary HO_x (HO_x = OH + HO₂) source (e.g., Horowitz et al., 2020), while unlike Cl it does not also oxidize methane. Here I create a model experiment (Cl₂_BrCl_Br₂) including direct emission of the bromine species Br₂ and BrCl. Wittmer et al. (2015b) measured the Br versus Cl production rate from artificial sea salt created from a variety of iron-containing artificial seawater or NaCl stock solutions, with the ratio of Br/Cl produced by mass ranging from 0 to a factor of 2.5. Br atom was below the detection limit for the NaCl-based solutions (Wittmer et al., 2015b), which may better resemble artificially engineered iron salt aerosol. However, previous work with salt pans found that even when Br atom concentrations were below the detection limit, bromide impurities in “pure” NaCl ($\leq 0.01\%$ Br⁻) could lead to Cl atom release due to BrCl formation and photolysis, which would also release Br atoms in equal quantities (Wittmer et al., 2015a). Here I assume that of the total desired chlorine release (1,250 Tg/yr as in the Cl₂-ocean simulation), 20% of that by mass of bromine is released in equal parts Br₂ and BrCl (resulting in 1193 Tg/yr Cl₂, 187 Tg/yr Br₂, and 129 Tg/yr BrCl). This scenario represents a bounding case if artificial sea salt containing bromine impurities were to be continuously emitted as part of an AOE method. In prior work, increasing the flux of natural sea salt aerosol in GEOS-Chem led to relative increases in tropospheric-wide reactive bromine that were comparable to that of reactive chlorine (Horowitz et al., 2020). More laboratory studies are needed to understand the potential bromine release from engineered iron salt aerosol.

2.2.2.3 Iron salt aerosol

Here I implement a parameterization of photolytic Cl₂ release from iron-enriched salt aerosol developed by Chen et al. (2024). The production rate of Cl₂ ($d[\text{Cl}_2]/dt$: molecules cm⁻³ s⁻¹) is a function of the nitrogen dioxide (NO₂) photolysis frequency (j_{NO_2} : s⁻¹), accumulation mode aerosol iron (III) concentration ($[\text{Fe}^{3+}]$: mol l⁻¹ water, or M), accumulation mode aerosol chloride concentration ($[\text{Cl}^-]$: M), and aerosol surface area concentration (S : μm² cm⁻³), and is scaled by the factor α ($=1.4 \times 10^5$ molecules μm⁻² M⁻²) based on experimental results from Wittmer et al. (2015b) accounting for the volume of the chamber:

$$d[\text{Cl}_2]/dt = \alpha j_{\text{NO}_2} [\text{Fe}^{3+}] [\text{Cl}^-] S \quad (2)$$

This reaction ($\text{Cl}^- \rightarrow 0.5\text{Cl}_2$) occurs on chloride in accumulation mode sea salt aerosol, which in GEOS-Chem is defined as ≤ 0.5 μm in diameter. Additional details on the parameterization calculations including α and the chamber volume correction are presented in Text S1 and Table S3 in the Supplementary Information.

Spatiotemporally varying photolysis frequencies (including j_{NO_2}) are calculated online in GEOS-Chem using the Fast-JX scheme (Neu et al., 2007) as implemented by Mao et al. (2010) and account for cloud optical depth. In the standard GEOS-Chem model, Fe(III) concentrations are estimated from natural dust up to 1 μm and anthropogenic particulate Fe (pFe) for the Fe(III)-catalyzed SO₂ oxidation chemistry. This assumes that total Fe content from dust is 3.5% of total mass (Taylor & McLennan [1985], consistent with Trapp et al. [2010]), dust iron solubility is 1% (Alexander et al., 2009) while anthropogenic particulate Fe is more soluble at 10% (Shao et al., 2019), and Fe(III) is 10% of total dissolved iron in the daytime (Moffet et al., 2012) and 90% at night. These assumptions are not sufficient to match observations in polluted urban environments in China (Chen et al., 2024). Model evaluation against observations suggests four times higher solubility for both dust and anthropogenic Fe (4% and 40%, respectively) and 67% of total dissolved iron as Fe(III) (Chen et al., 2024). Observed Fe solubility in aerosols over the oceans is highly variable (0.19% to 47.8%) and is influenced by the initial source of the iron (e.g., dust vs. combustion), other chemical components in the aerosol, pH, and relative humidity (Shi et al., 2022). For the iron salt aerosol simulations, I use the Chen et al. (2024) representation of solubility and Fe(III) speciation as a maximum rate of Cl₂ production through this particular parameterization. For consistency with the version of the Chen et al. (2024) mechanism that was evaluated against observations (see Section 3.1.2), I also include this process on fine-mode aerosols only; the chamber experiment on which the mechanism is based used fine mode aerosol (Wittmer et al., 2015b) and the lifetime of smaller particles is longer relative to coarse particles such that iron and chloride have time to mix within the aerosols before being deposited (Moffet et al., 2012, Zhu et al., 2022). The dependence on aerosol surface area concentration, and the impacts of pH and potential suppression of the rate by aerosol sulfate and organics (which are not considered in this mechanism) warrant future study. van Herpen et al. (2023) also parameterized the production of Cl₂ from natural mineral dust–sea spray aerosols. Major differences between Chen et al. (2024) and van Herpen et al. (2023) include the size of the particles on which this process occurs, the percentage of photoactive iron, and that the Chen et al. (2024) study is based primarily on the Wittmer et al. (2015b) chamber experiments while van Herpen et al. (2023) follow the Fe(II)–Fe(III) cycling kinetics from Zhu et al. (1993). Differences in the model frameworks are summarized in Table S5.

First I perform GEOS-Chem simulations with the additional reaction of iron-mediated Cl₂ release from chloride aerosol from Chen et al. (2024) in the absence of any additional emissions to act as a reference point for the atmospheric oxidation enhancement experiments. Then, to release additional Cl₂ from iron salt aerosol, accumulation mode aerosol chloride and particulate iron (pFe) are released over the oceans at the surface in four scenarios to assess the driving factors and compare against the direct Cl₂ scenario. The highest iron addition tested in Wittmer et al. (2015b) was 13 mol Cl⁻/mol Fe³⁺. Given the assumed solubility of anthropogenic pFe of 40% and Fe(III) speciation fraction of 67% in my GEOS-Chem simulations, a ratio of 3.484 mol Cl⁻/mol pFe emitted would lead to Fe(III) ratios at most comparable to the highest iron addition experiments in Wittmer et al. (2015b). For accumulation mode aerosol chloride emissions of 1,250 Tg (the same mass of total chlorine emissions as the Cl₂ experiment), this is 565 Tg/yr pFe. The experiments here include particulate iron emissions alone (Iron), aerosol chloride emissions alone (Chloride), a combination of both particulate iron and chloride (Iron_Chloride), and particulate iron alone emitted at 1,250 Tg/yr (Iron_Max) as a test case (see Table 1 for details). Iron_Max does not lead to significantly more methane loss than the Iron scenario; hence, it will only be discussed in the context of methane and not with respect to tropospheric chemistry and air quality.

3 Results

3.1 Impacts on Tropospheric Chemistry

3.1.1 Major Tropospheric Oxidants and Reactive Halogens

Table 2 presents the tropospheric burdens of major oxidants, reactive halogen families, and carbon monoxide (CO) for the standard GEOS-Chem simulation and the standard simulation plus the Chen et al. (2024) parameterization, followed by relative percent changes in these burdens for the atmospheric oxidation enhancement model experiments described in Table 1.

Table 2. Percent change in simulated annual mean tropospheric burdens of selected species and chemical families.

	Br _y	Cl _y	I _y	O ₃	OH	Cl	NO _x	CO
Standard	20 Gg	240 Gg	12 Gg	338.4 Tg	214 Mg	313 kg	360 Gg	350 Tg
Standard+Chen	20 Gg	241 Gg	12 Gg	337.8 Tg	214 Mg	343 kg	359 Gg	350 Tg
H2O2_high	-12.2%	36.7%	-77.5%	-38.5%	165.9%	401.7%	8.2%	-46.4%

H2O2_mid	44.2%	12.3%	- 20.3%	-6.0%	30.3%	73.8%	-17.1%	-23.2%
H2O2_low	9.6%	0.9%	-1.2%	-0.7%	2.3%	6.3%	-4.1%	-2.9%
Cl2	-6.1%	1738%	- 42.7%	-24.4%	-27.7%	2213.6%	-18.7%	52.2%
Cl2_BrCl_Br2	2596%	1689%	- 75.2%	-67.1%	-47.1%	1869.5%	-30.6%	98.6%
Iron*	6.1%	12.0%	-5.3%	-3.4%	-2.1%	169.4%	-2.5%	3.3%
Chloride*	46.9%	1116%	-8.7%	-5.6%	-3.9%	178.6%	-3.5%	4.2%
Iron_Chloride*	43.4%	1141%	- 17.2%	-10.6%	-8.6%	658.0%	-9.2%	13.5%

Note: Percent change in annual mean tropospheric burdens for model experiments (described in Table 1) are relative to the standard version (* = relative to standard + Chen et al., 2024). Br_y, Cl_y, and I_y follow definitions in Wang et al. (2021); NO_x =NO + NO₂. See also Section 2.1.

Previous work on an intermodel comparison of atmospheric chemistry and climate models from the Atmospheric Chemistry and Climate Model Intercomparison Project found that changes in global mean OH between models and within a given model are more a function of the relative loss of reactive nitrogen versus reactive carbon than the emissions of reactive nitrogen versus carbon, as it matters how much of the loss of these species is due to OH versus other processes (Murray et al., 2021). The loss of reactive carbon is a function of a given model's chemical mechanism and structure (Murray et al., 2021). The GEOS-Chem model version used in this study (13.2.1) is several generations ahead of that used in Murray et al. (2021) (version 9-01-03), with a number of changes in reactive nitrogen chemistry including aerosol uptake and recycling of isoprene nitrates (e.g., Fisher et al., 2016, 2018; Holmes et al., 2019; McDuffie et al., 2018) that have been shown to impact the tropospheric OH burden by up to 12%. The representation of nitrogen oxides (NO_x) loss thus can lead to uncertainty in the simulated OH burden, but it is not clear if this would be a consistent bias across all model experiments in the current study and thus negligible when considering relative changes.

As expected, the hydrogen peroxide experiments all increase tropospheric OH due to photolysis of the additional hydrogen peroxide. OH does not increase proportionally to the increase in hydrogen peroxide emissions across experiments and indicates a reduction in the effectiveness of additional emissions at higher levels of hydrogen peroxide. A >200-fold increase in hydrogen peroxide emissions from H₂O₂_mid to H₂O₂_high leads to a 5.5-fold increase in the change of OH; a 60-fold increase in hydrogen peroxide emissions from H₂O₂_low to H₂O₂_mid leads to a 13-fold increase in the change of OH (see Table 2). The increases in OH drive reductions in tropospheric CO seen in these experiments, as OH is the main oxidant of CO. The reductions in CO are nearly proportional to the OH increases in the H₂O₂_low and H₂O₂_mid scenarios; however, in the extreme H₂O₂_high scenario, the CO decrease is much lower than expected from the OH increase (see Table 2). This is

270 likely due to the much larger increase in Cl atom in the H₂O₂_high experiment, leading to additional production of CO from
Cl reactions with formaldehyde and organochlorines (see Table S6 in the SI). In all of the hydrogen peroxide experiments, Cl
atom increases. Increased HO_x (= OH + HO₂, which cycle rapidly between each other) leads to increased release of chlorine
from sea salt aerosol and organohalogens (see Table S6 in the SI and Wang et al., 2021), increasing the total Cl_y burden. Within
275 each experiment, the longer-lived reservoir species hydrogen chloride (HCl) contributes the largest mass to the total Cl_y burden.
However, relative to the standard simulation the partitioning of gas-phase Cl_y shifts away from HCl toward chlorine hydroxide
(HOCl), Cl₂, and Cl, such that the relative increase in Cl atom is greater than the relative increase in total Cl_y for each
experiment. This is due to the complex interplay of cross-reactions between fast-cycling HO_x and chlorine radicals as well as
across their longer-lived reservoir species. For example, in addition to the reactions in Table S6 which are sources of gas-phase
Cl_y rather than cycling between species, OH can react with HCl to produce Cl atom.

280 Cl-based experiments decrease tropospheric OH due to the destruction of ozone by the additional Cl, as ozone
provides the main source of tropospheric OH. Tropospheric CO increases in the chlorine-based experiments, partially due to
the decrease in OH as well as the additional production of CO from reactions of formaldehyde and organochlorines with Cl
atom. The Cl₂, Chloride, and Iron_Chloride experiments have the same amount of total chlorine emissions (see Table 1) but
vastly different effectiveness at increasing the Cl atom concentration due to the emitted species (gas-phase Cl₂ or particulate
285 chloride) having different reactivities and reactions. The Chloride and Iron_Chloride experiments, where particulate
accumulation mode chloride was emitted, have 64 to 66% as much of an increase in total gas-phase Cl_y burden relative to the
Cl₂ experiment where gas-phase Cl₂ was emitted (see Table 2). However, only 8% (Chloride only) to 30% (Iron_Chloride) of
the increase in Cl atom seen in the Cl₂ experiment is realized. This is because the vast majority of the increase in Cl_y burden
in these experiments is due to HCl, shifting Cl_y away from more reactive species. As in the hydrogen peroxide experiments,
290 HCl still remains the dominant component of Cl_y. The release of Cl₂ is iron-limited, as the addition of particulate iron emissions
with the same amount of chloride (from Chloride to Iron_Chloride) leads to a nine-fold increase in Cl₂ and a three-fold increase
in Cl (see Table S7).

In the Cl₂ and Cl₂_BrCl_Br₂ experiments, Cl_y partitioning is shifted away from HCl to Cl₂ and Cl. Thus, the relative
increase in Cl atom is even greater than the relative increase in total tropospheric Cl_y (see Table 2). In Li, Meidan et al. (2023),
295 1,250 Tg/yr Cl₂ led to a 30-fold increase in the tropospheric Cl burden; these emissions in our shorter-term modeling study led
to a 23-fold increase in the Cl₂-only case and a 20-fold increase in the Cl₂_BrCl_Br₂ case (where 1250 Tg/yr of total chlorine
was split between 1193 Tg/yr as Cl₂ and the remainder as BrCl). In GEOS-Chem, the only fate of BrCl is photolysis to produce
Cl atom. Hence, the smaller impact of the Cl₂_BrCl_Br₂ scenario on total tropospheric Cl atom is due to the increases in Br_y
in this experiment (see Table 2) and the coupled chemistry of Br_y species with Cl_y cycling.

300 All experiments decrease the tropospheric ozone burden (see Table 2), with implications for ozone radiative forcing.
All experiments have increased Cl atom concentrations, which can destroy ozone. H₂O₂_low, H₂O₂_mid, and all chlorine
experiments except Cl₂ also have increased Br_y burdens, which can lead to additional ozone loss via reactive bromine. Increases
in OH in the hydrogen peroxide experiments could lead to increased sources of Br_y from sea salt and organohalogens (see

Table S6 and Wang et al., 2021). In the H₂O₂_low, H₂O₂_mid, and chlorine experiments, part of the change in ozone is also
305 due to reductions in NO_x that lead to reduced ozone production. This is consistent with previous work which found more active
halogen chemistry leads to reductions in NO_x (e.g., due to hydrolysis of halogen nitrates) (e.g., Wang et al., 2019, 2021;
Sherwen et al., 2016b; Horowitz et al., 2020). The increases in Br_y and decreases in NO_x, which became larger as hydrogen
peroxide emissions increase from H₂O₂_low to H₂O₂_mid, flip sign once hydrogen peroxide emissions become particularly
extreme in the H₂O₂_high scenario. Thus in the H₂O₂_high case, the ozone loss is likely dominated by the large increase in Cl
310 atom (see Table 2).

As seen in Horowitz et al. (2020), the decreases in ozone drive decreases in tropospheric I_y in all experiments, as
ozone reactions with sea surface iodide are the dominant source of tropospheric I_y (Wang et al., 2021). The greatest impacts
on ozone are in the Cl₂_BrCl_Br₂ experiment, which included bromine release to represent bromide contamination (see Table
1). The 20% additional mass released as bromine (Br₂ and BrCl, which both photolyze to release Br atom) led to a much larger
315 than 20% additional reduction in ozone over the Cl₂-only experiment (factor of 2.75 higher) and hence OH (factor of 1.70
higher) (see Table 2). While 20% is likely an overestimate of the bromide content of engineered aerosol, it is well within the
range reported from natural seawater salt experiments catalyzed by iron (Wittmer et al., 2015b) and suggests that even a small
perturbation in bromine could have a much larger impact on OH.

The OH_mid and OH_high scenarios lead to 51% to 94% increases in OH, respectively (see Table S8). Despite a
320 much larger increase in OH in OH_high than in OH_mid, the decrease in CO is only 5% larger. The OH experiments' results
for CO are less reliable than the H₂O₂ experiments as the OH increase is limited to the lower 1 km where it is released. (OH is
not transported in the model; hydrogen peroxide is.) In this region, methane levels remain high due to the surface boundary
condition and long lifetime against reaction with OH; as CO is reacted away by the additional OH, its loss rate slows down
due to its shorter lifetime and fully responsive concentrations at the surface.

325 **3.1.2 Impacts on Methane**

Figure 1 (top panel) summarizes the impacts on the overall methane lifetime, the impacts on the methane lifetime
with respect to tropospheric OH, and the percent of Cl atom contributing to total chemical loss in the troposphere for each of
the experiments in Table 1 and the standard model. Figure 1 (bottom panel) shows the resulting relative change in steady-state
methane concentrations estimated with the methods described in Section 2.1.

330

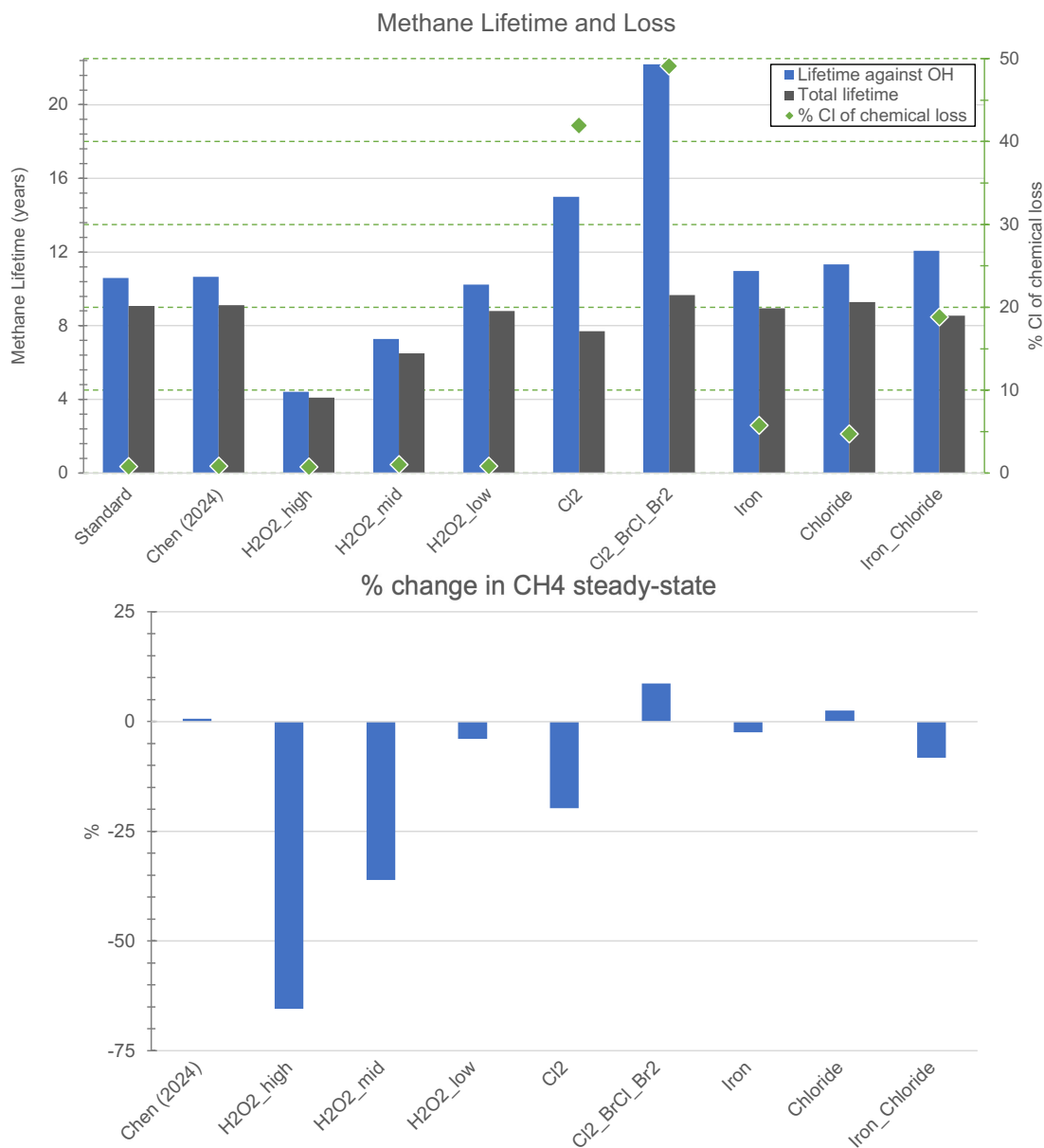


Figure 1: Top panel: total methane lifetime (gray bars) and methane lifetime against oxidation by OH (blue bars), overlaid by the percent of chemical loss contributed by Cl atom (green diamonds, right-hand y-axis). Bottom panel: estimated relative percent change in steady-state methane concentration.

335

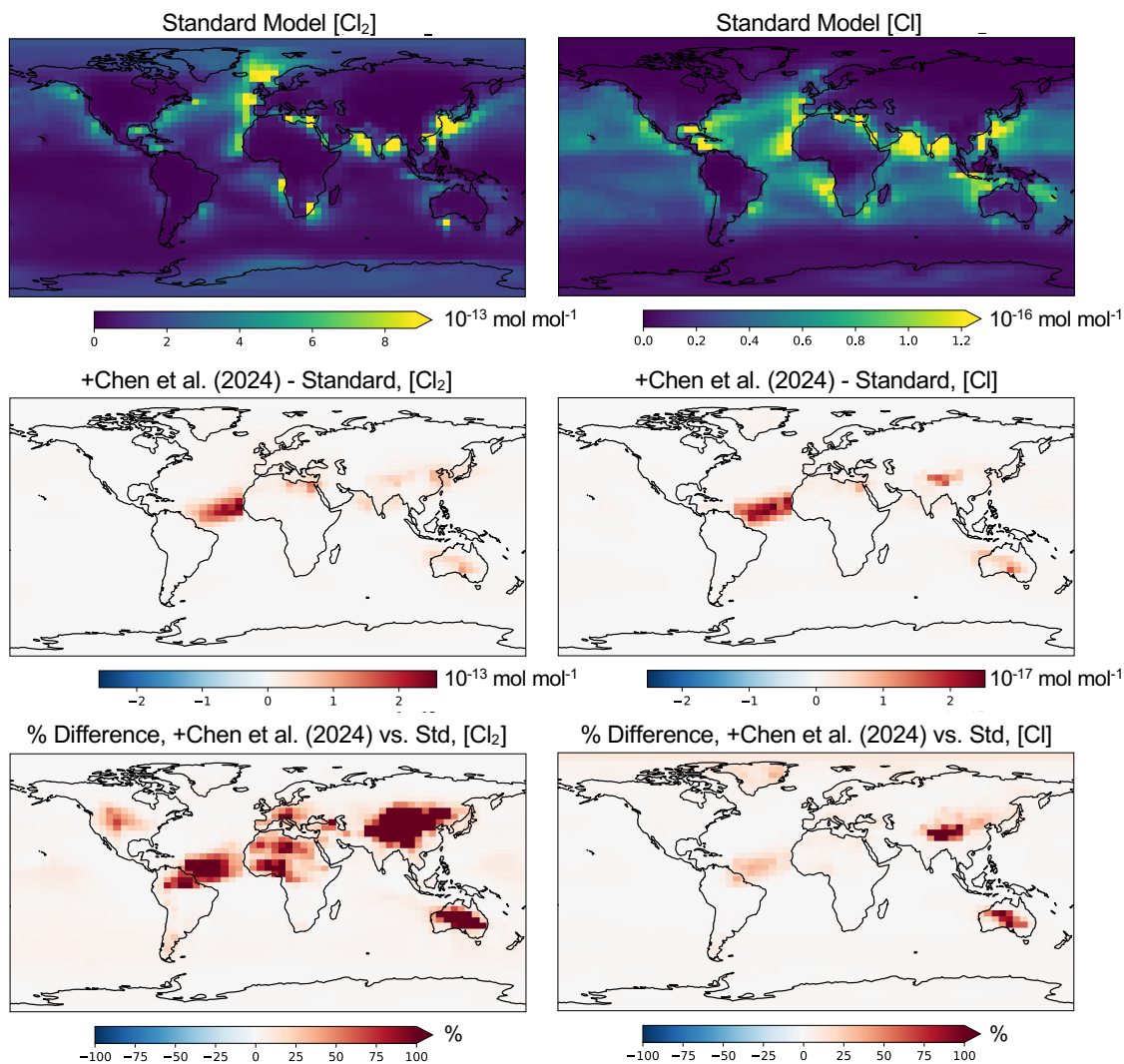
Studies suggest that the present role of chlorine in tropospheric methane oxidation is 0.23–5% based on constraints from isotopic observations (Allan et al., 2007; Gromov et al., 2018; Platt et al., 2004) and atmospheric modeling studies focused on chlorine chemistry (e.g., Hossaini et al., 2016; Wang et al., 2019, 2021). The standard version of GEOS-Chem

4used here is consistent with these constraints at 0.8%, which gives confidence to the simulation of atomic chlorine and its
340 reaction with methane.

Although the H₂O₂_high experiment had more than 5x greater increase in tropospheric OH than that of H₂O₂_mid
(see Table 2), the methane lifetime against OH decreases by less than a factor of 2 (Figure 1). The efficacy of H₂O₂ with respect
to oxidizing methane via tropospheric OH and decreasing steady-state methane concentrations decreases with increasing
emissions (see Figure S3). There is a larger difference between the scenarios for the change in steady-state methane, which
345 also accounts for changes in the lifetime against tropospheric Cl. While all H₂O₂ experiments also lead to increased Cl atom
(see Table 2), H₂O₂_low is most efficient at also decreasing the methane lifetime against oxidation by tropospheric Cl for its
smaller level of emissions (Figure S3). Significant methane reductions for the hydrogen peroxide scenarios (>5%) require
more than 1,250 Tg/yr emissions (H₂O₂_low).

For the same annual emissions of Cl₂ (1,250 Tg/yr) as in Li, Meidan et al. (2023), I find slightly smaller decreases in
350 the methane lifetime (here -15% vs. 50%) and methane concentrations (here -20% at steady-state vs. -45% in year 2050)
despite vastly different modeling frameworks (Table S4). Li, Meidan et al. (2023) finds a larger increase in the tropospheric
burden of Cl atom, as described in the previous section (factor of 30 vs. factor of 23), and a larger percent of methane chemical
loss due to tropospheric chlorine (60% vs. 41%). Differences in our results may be due to differences in the vertical distribution
of Cl as well as the meteorology between the two models, as the Cl + CH₄ reaction is temperature-dependent (see Table S6)
355 and the impacts of local changes in the reaction rate on the global methane lifetime are weighted by air density. The vertical
distribution of Cl in the two models is not only impacted by the halogen chemical mechanism (see Table S4); in GEOS-Chem,
an improved representation of entrainment-limited uptake in clouds and ice cloud particle properties for cloud heterogeneous
chemistry leads to significant increases in reactive chlorine, particularly in the upper troposphere where reaction rates are
decreased due to low temperatures (Table S6), due to changes in HCl-ClO_x (chlorine oxides) cycling (Holmes et al., 2019).
360 Other differences that could impact the results are the increase in methane emissions in Li, Meidan et al. (2023) following the
representative concentration pathway 8.5 scenario, and the shorter simulation period in our study (see Table S4). Bromine
contamination of 20% more than reverses the gains seen in the Cl₂ case, as the Cl₂_BrCl_Br₂ case has an increase in steady-
state methane concentrations of 8.7% (see Figure 1).

Figures 2 and S4 show the change in annual mean Cl₂ and Cl atom concentrations at the surface and zonally averaged
365 through 20km, respectively, from adding the Chen et al. (2024) mechanism for Cl₂ production from aerosol iron
photochemistry. The absolute increase in surface [Cl] (Figure 2) is largest over the North Atlantic ocean, with the largest
relative increases over China reaching a factor of 3 (204%). In Chen et al. (2024), simulations were performed for December
9 - 31, 2017 over North China at 16 times higher spatial resolution (0.25° latitude × 0.3125° longitude) than this study (4° ×
5°). They also added high-resolution anthropogenic fine-mode aerosol Cl⁻ emissions in China from Fu et al. (2018) which are
370 not included here and would further increase Cl₂ production through their mechanism (see equation 2). They found a maximum
increase in [Cl] in an individual model gridbox of a factor of 20 to 40 which is consistent with the increased model horizontal
resolution relative to my study.



375 **Figure 2: Impacts of adding natural Cl_2 emissions from the Chen et al. (2024) parameterization on simulated year 2019 annual mean surface-level concentrations of Cl_2 (left panels) and Cl (right panels). The top row is the standard model concentrations (mol mol^{-1}), with the second and third rows containing the absolute difference in mol mol^{-1} and relative difference in percent, respectively, from adding the Chen et al. (2024) mechanism.**

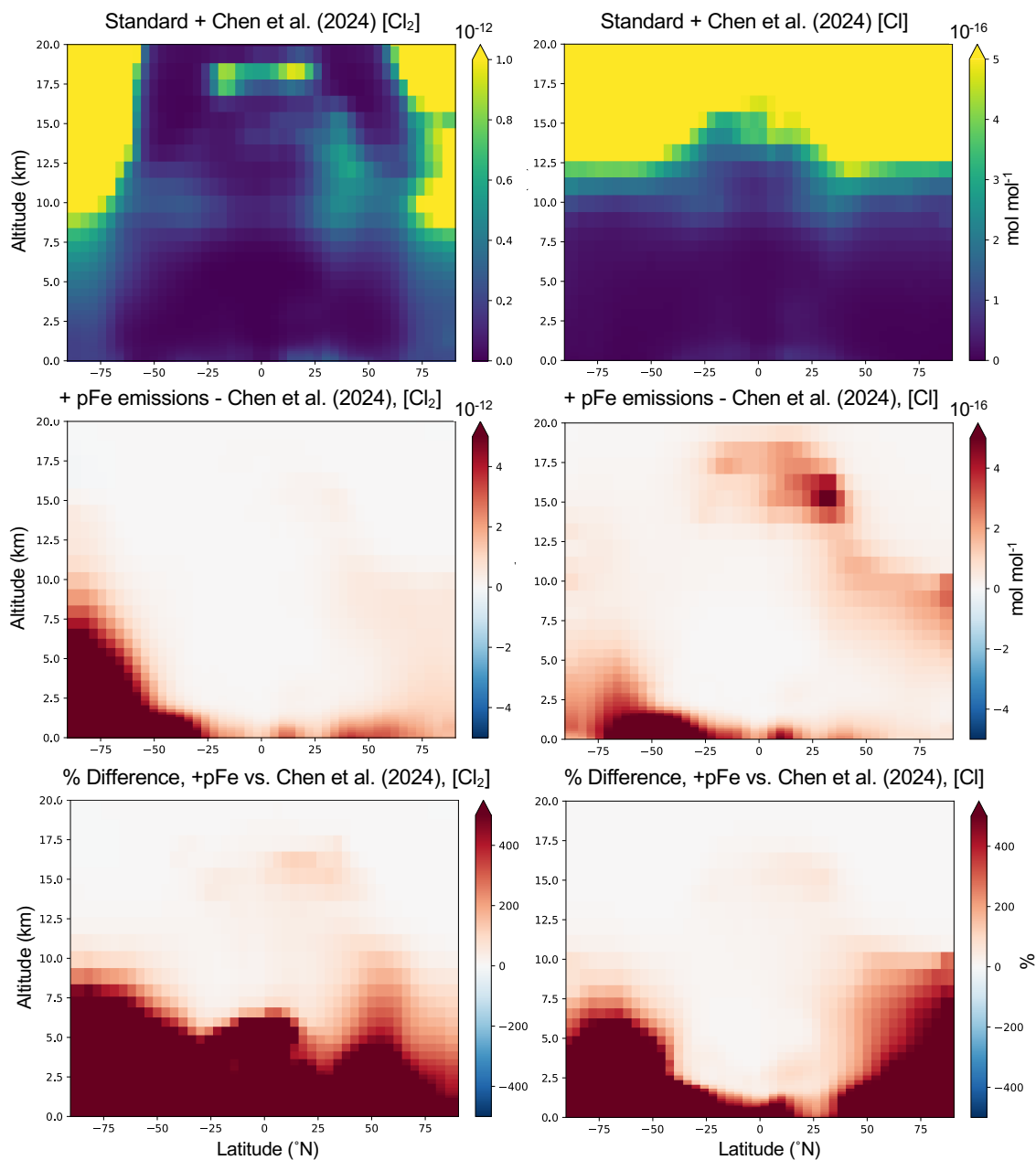
380 van Herpen et al. (2023) found that including their parameterization of Cl_2 release from mineral dust–sea spray aerosol (without any additional emissions) led to an increase in methane loss via the Cl atom of $\sim 20\%$, although the overall global methane loss was decreased slightly due to a reduction in methane loss via reaction with OH . Here I find an increase in the tropospheric burden of Cl atom of 9.4% (see Table 2) when the Chen et al. (2024) parameterization is included (without any additional emissions), which translates to an increase of 5.4% in the methane loss via Cl atom. While van Herpen et al. (2023)

385 do not report global changes in chlorine burdens, they find that total inorganic chlorine production is increased by 41%. This led to a change in -0.7% in the tropospheric ozone burden after including Cl_2 release from natural iron salt aerosol (van Herpen et al., 2023), while I find a change of -0.18% in the tropospheric ozone burden when I include the Chen et al. (2024) parameterization. Thus, it is likely that the two parameterizations (see Table S5) result in Cl_2 production rates that are up to 4 times higher in van Herpen et al. (2023). In the zonal mean, increases in $[\text{Cl}_2]$ and $[\text{Cl}]$ from including the Chen et al. (2024) production mechanism are largest aloft (Figure S4). van Herpen et al. (2023) limited their mechanism to the lower troposphere. Hence, differences in the vertical distributions of changes in chlorine vs. concentrations of ozone and methane in the two parameterizations may further contribute to the differences in impacts. The Chen et al. (2024) mechanism was evaluated directly against hourly observations of gas-phase Cl_2 in Wangdu, China, as well as key parameters including aerosol $[\text{Cl}^-]$, aerosol total iron, j_{NO_2} , and aerosol surface area (see equation 2). GEOS-Chem also captured the variability and magnitude of daily total Fe and Cl^- concentrations in fine-mode aerosol collected from 29 sites across North China, and slightly underestimated the observed solubility of aerosol Fe which was sampled at two sites, during the intensive study period at Wangdu (Chen et al., 2024). While adding the mechanism greatly improved the model performance of $[\text{Cl}_2]$ at Wangdu, increasing simulated concentrations by a factor of 28 to 48, model concentrations remained underestimated. Chen et al. (2024) hypothesized this may be due to overestimated aerosol water in the model's thermodynamic module and thus underestimated aqueous-phase $[\text{Cl}^-]$ and $[\text{Fe}^{3+}]$. Low aerosol water content at high altitudes and hence high aqueous concentrations may explain the increases aloft seen in Figure S4. It is outside of the scope of this paper to address uncertainties in the parameterization at higher altitudes. Figure 3 shows the annual mean change in Cl_2 and Cl atom concentrations in the Iron experiment (see Table 1) relative to the model with the Chen et al. (2024) mechanism alone, averaged zonally through 20km altitude. Here, zonal mean increases in $[\text{Cl}_2]$ and $[\text{Cl}]$ due to iron emissions are largest near the surface (much greater than 400%, up to 60,000%) where emissions are added, unlike when the Chen et al. (2024) parameterization is included in the absence of additional surface emissions (Figure S4). Figures 3 and S4 show that the additional release of Cl_2 due to natural or emitted iron leads to changes in Cl_2 and Cl that have differing spatial patterns and magnitudes in the annual mean due to physical and chemical processing. Tropospheric-wide differences between changes in Cl_2 and changes in Cl (see Table S7) are minor from adding the Chen et al. (2024) parameterization alone or in the Chloride experiment, but diverge most in experiments where iron emissions are added (Iron and Iron_Chloride). There, the relative increase in Cl_2 is 4 to 6.5 times greater than what is realized for Cl (Table S7).

With the parametrization used in this study, iron salt aerosol can lead to changes in steady-state methane of -8.3% to $+2.5\%$ depending on the emissions employed and their relative effects on Cl vs. OH (see Figure 1). The Iron experiment led to a small decrease in steady-state methane (-2.5%) from a tropospheric-wide factor of 2.7 increase in Cl atom (see Table S7). The Iron_Chloride experiment led to a larger decrease in steady-state methane (-8.3%) due to a larger 7.6-fold increase in Cl atom burden, despite a much larger reduction in OH (-8.6% vs. -2.1% ; see Table 2 and Table S7). For a similar increase in Cl atom burden, emitting chloride (Chloride, 2.8-fold increase in Cl) instead of iron aerosol (Iron, 2.7-fold increase in Cl) led to a larger decrease in OH (-3.9% vs. -2.1% ; see Table 2) and hence net increase in methane ($+2.5\%$). Li, Meidan et al. (2023)

found that a 2.8-fold increase in Cl burden (from 88 Tg/yr gas-phase Cl₂ emission) was insufficient to decrease methane, while
420 a 7.9-fold increase in Cl burden (from 313 Tg/yr gas-phase Cl₂ emissions) overcame the OH competition and led to a decrease
in methane concentrations by about 6% after 10 years (Li, Meidan, et al., 2023). This suggests that the threshold of additional
chlorine needed to overcome the OH limitation depends on what is emitted and the background chemistry in the model
employed.

Here I find that emitting sea salt chloride along with particulate iron increases methane loss. This is a function of the
425 formulation of the Chen et al. (2024) mechanism, which occurs on sea salt chloride aerosol and increases with increasing [Cl⁻]
] concentrations (see Section 2.2.2.3 and Equation 2). Artificial chloride aerosol emissions in addition to the particulate iron
emissions can replenish the sea salt chloride that was converted to Cl₂ and increase aerosol [Cl⁻] concentrations, leading to
greater overall production of Cl₂ (see Table S7).



430

Figure 3: Impacts of adding surface pFe emissions (Iron experiment in Table 1) on simulated year 2019 zonal, annual mean concentrations of Cl_2 (left panels) and Cl (right panels) up to 20km altitude. The top row is modeled concentrations in mol mol^{-1} with the Chen et al. (2024) parameterization alone, with the second and third rows containing the absolute difference in mol mol^{-1} and relative difference in percent, respectively, after pFe emissions are added to the surface layer in this simulation.

435

The iron emissions needed to decrease methane are also uncertain and sensitive to the mechanism and rate of chlorine release from iron salt aerosol. The Iron and Iron_Chloride scenarios presented here have 565 Tg/yr iron emissions (see Table 1) leading to a decrease in steady-state methane of 2.5% to 8.3%. In a Community Earth System Model (CESM) study using the van Herpen et al. (2023) parameterization for chlorine release, additional iron emissions of 200 Tg/yr (which increased Cl₂ production by ~850 to 930 Tg/yr depending on the region of emission) lead to a ~20% decrease in global methane concentrations after 10 years (Meidan et al., 2024). This study found that a threshold of at least 6 Tg/yr of iron was needed in the most idealized setup to avoid increasing methane, though larger emissions of >60 Tg/yr of iron may be needed for significant methane reductions.

Here, the methane loss in the Iron_Chloride scenario decreases in the tropical and southern hemisphere's free troposphere, with increases in methane loss rates limited mostly to the surface and some regions of the upper troposphere. There may also be a reduction in Cl₂ production from chloride via other reactions with OH and chlorine nitrate (ClONO₂) due to reductions in tropospheric OH and NO_x; thus, the representation of competing heterogeneous halogen chemical reactions may play a role in the predicted effectiveness of iron salt aerosol. Additional laboratory experiments and field observations are needed to constrain this process in natural and engineered aerosol mixtures.

Overall, OH-based scenarios and gas-phase emission of Cl₂ lead to significant decreases in steady-state methane but require extremely large emissions. Cl₂ is much more effective at reducing methane (-20%) than hydrogen peroxide (-4%) for the same level of emission (1,250 Tg/yr). There are uncertainties in the exact impact on methane of a given amount of Cl₂ emission (e.g., in this study vs. Li, Meidan, et al., 2023) due to a variety of factors including the representation of complex reactive halogen chemistry. Despite this, the methane lifetime and methane concentration reductions from 1,250 Tg/yr of gas-phase Cl₂ emissions presented here are qualitatively similar to those of Li, Meidan et al. (2023). In contrast, by using a different chlorine release mechanism from iron salt aerosol from Chen et al. (2024) and a different atmospheric model, I find that larger iron emissions (565 Tg/yr) than those used in Meidan et al. (2024) (200 Tg/yr) lead to a smaller decrease in methane (-2.5% at steady-state vs. -20% after 10 years). The mechanism of Cl₂ emission, whether this would also release bromine, and how fast the release from iron salt aerosol is will affect whether this method would decrease or increase steady-state methane.

3.1.3 Impacts on Other Climate Forcers and Ozone-Depleting Substances

GEOS-Chem simulates additional greenhouse gases and ozone-depleting substances including nitrous oxide (N₂O), halons (three species), hydrochlorofluorocarbons (HCFCs; four species), chlorofluorocarbons (CFCs; five species), and other halomethanes (12 species, e.g., chloromethane, methyl bromide, chloroform). I examined the changes in tropospheric burdens of each. Due to the short simulation length of 1 year following 1 year of initialization, I focus here only on short-lived species (lifetimes <6 months).

Details of changes in tropospheric burdens (accounting for the temporally and spatially varying tropopause) for selected short-lived species with significant impacts from the AOE experiments are shown in Table S6. Of the short-lived halomethanes, bromoform (CHBr₃) and dibromomethane (CH₂Br₂) have the largest tropospheric-wide, with decreases in the

hydrogen peroxide-based experiments and increases in the Cl-based experiments. Dichloromethane and chloroform follow
470 the same pattern, but changes are smaller overall. For dibromomethane, dichloromethane, and chloroform, the addition of
bromine in the Cl₂_BrCl_Br₂ experiment leads to larger increases relative to Cl₂ only. These short-lived gases could contribute
to ozone depletion in the stratosphere and are not regulated by the Montreal Protocol (e.g., Hossaini et al., 2017). Iodine-
containing halomethane tropospheric burdens decrease across all experiments due to the reduction in overall I_y (Section 3.1.1),
with the largest decreases seen for methyl iodide (CH₃I), which is not thought to contribute to stratospheric ozone loss (Zhang
475 et al., 2020). The shortest-lived HCFC simulated by GEOS-Chem is HCFC-123 (atmospheric lifetime 1.3 years). One year of
AOE leads to large changes in its tropospheric burden, with decreases in the hydrogen peroxide-based scenarios (up to -74%)
and increases in the Cl-based scenarios (up to 63%). As HCFC-123 has a 100-year global warming potential of 90.4 (Smith et
al., 2021) and is a class-II ozone-depleting substance, longer-term simulations are needed to fully quantify the climate impact
of AOE

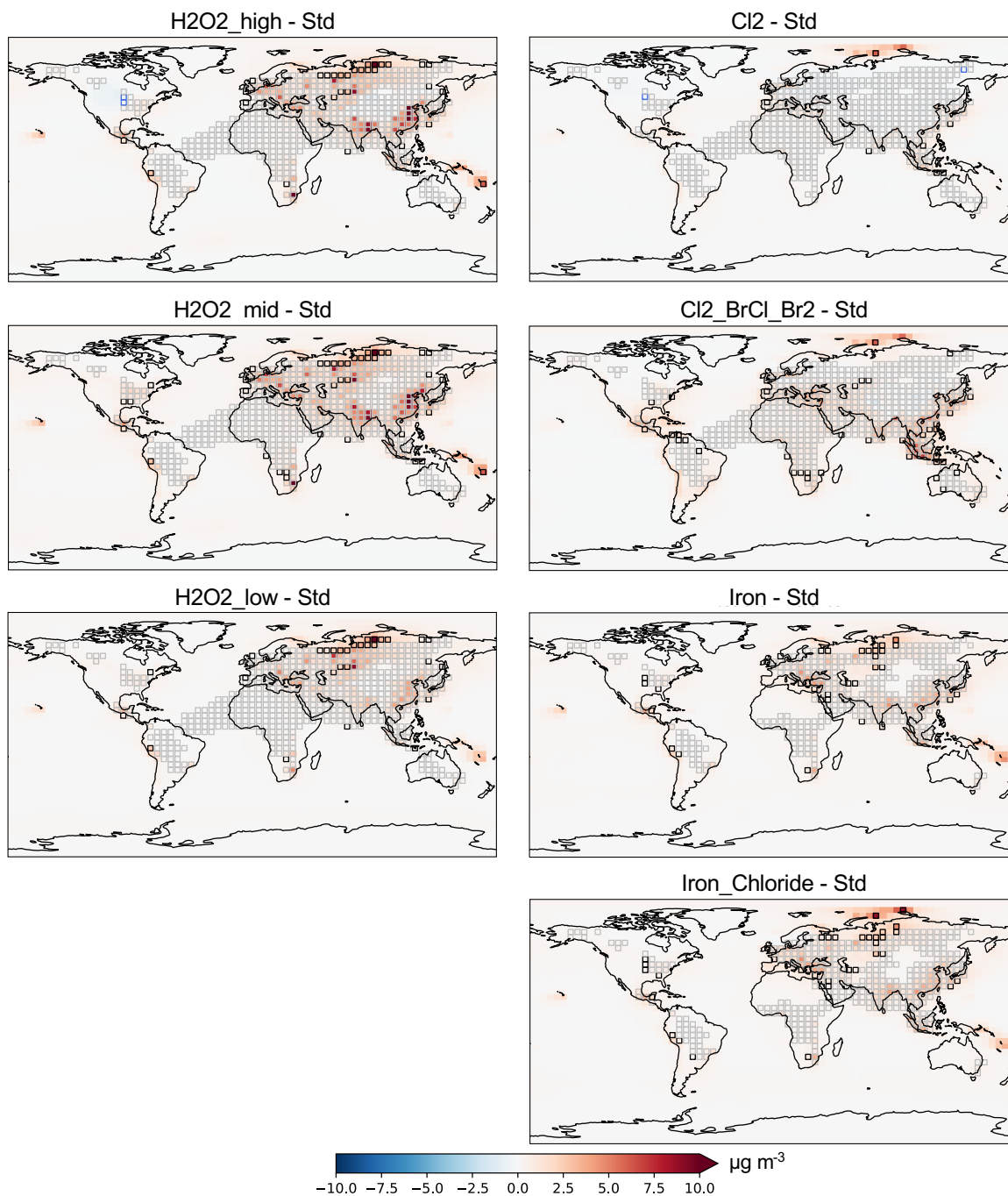
480 Across all experiments except Chloride, tropospheric inorganic aerosol increases (see Table S9) due to the increase
in tropospheric sulfate and ammonium, partially offset by decreases in tropospheric nitrate burdens. This is likely due to the
HOCl/HOBr+S(IV) pathways that produce sulfate (Wang et al., 2021), as Br_y and Cl_y both increase in nearly all experiments.
This is a departure from Li, Meidan et al. (2023), who found decreases in sulfate of ~10% from 1,250 Tg/yr Cl₂ emissions.
This difference may reflect the coupled halogen-sulfate chemistry in GEOS-Chem (Chen et al., 2017; Wang et al., 2021) that
485 does not appear to be included in the CESM chemical mechanism (see Table S4). For the hydrogen peroxide experiments,
increased hydrogen peroxide and OH will also increase the oxidation of SO₂ to sulfate in the aqueous and gas phases,
respectively. For the Iron and Iron_Chloride experiments, sulfate may also increase due to the Fe(III)-catalyzed pathway
(Alexander et al., 2009). Partitioning of ammonium and nitrate aerosol is determined by the thermodynamic model
ISORROPIA-2 in GEOS-Chem (see Section 2.1). Thus, ammonium likely concurrently increases to neutralize the excess
490 sulfate in all experiments. For the Chloride experiment, decreases in nitrate and sulfate are nearly exactly offset by increases
in ammonium. Nitrate is likely reduced due to reductions in NO_x in most experiments (see Table 2) and is influenced by the
change in gas-particle partitioning due to increased sulfate production. Increases in the tropospheric inorganic aerosol burden
could lead to an additional negative radiative forcing. Changes in these aerosols at the surface and the associated air quality
implications are examined in Section 3.1.4.

495 3.1.4 Impacts on Surface Air Quality

Surface PM_{2.5}, CO, ozone, and NO₂ are air pollutants impacted by the atmospheric oxidation enhancement
experiments. These largely follow the changes in tropospheric burdens (see Tables 2 and S6). Surface NO₂, however, was
negligibly impacted by all experiments; most of the changes in tropospheric NO_x occurred in the upper troposphere (see Figure
S5). The change in global annual mean surface PM_{2.5}, CO, and ozone is shown in Figure S6 across the different scenarios,
500 with spatial variations in annual mean surface PM_{2.5} and ozone shown in Figures 5 and 6, respectively.

In the mean across all scenarios, there are declines in surface ozone air pollution (Figure S6). Surface PM_{2.5} increases in all scenarios except in the Chloride case, where in the absence of iron to mediate the Cl₂ release, the additional chloride leads to decreased sulfate and nitrate. Interestingly, H₂O₂_mid has worse air quality impacts than H₂O₂_high despite having less emissions; PM_{2.5} increases more, and CO and ozone decrease less. There are larger impacts on PM_{2.5} from the hydrogen peroxide and OH methods (up to 14–19 μg/m³) than the Cl₂ and Cl₂_BrCl_Br₂ methods (up to 6–7 μg/m³). These results are consistent with prior work suggesting that aerosol production can be oxidant-limited (e.g., Mayhew and Haskins, 2025; Shah et al., 2018). Figure 4 shows the absolute change in PM_{2.5} spatially in year 2019. Gray boxes highlight areas already in exceedance of the U.S. Environmental Protection Agency’s (EPA’s) annual mean PM_{2.5} primary standard of 9 μg m⁻³, which remain in exceedance when the experiment is applied. Black boxes highlight areas where the experiment to increase methane oxidation brought a region from <9 μg m⁻³ to an exceedance of 9 μg m⁻³. Blue boxes show where the experiment led to improved PM_{2.5} from an exceedance to < 9 μg m⁻³.

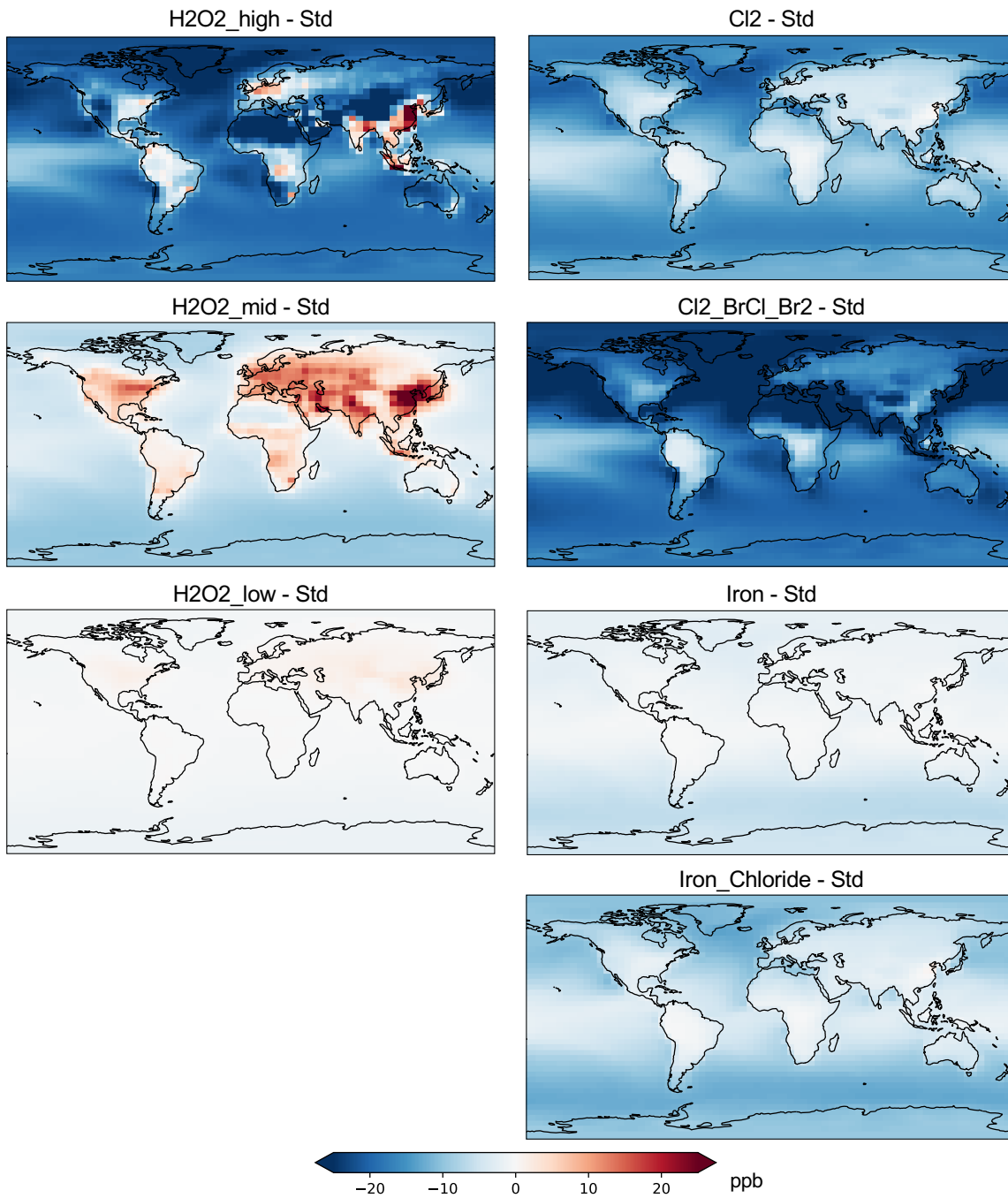
These results represent one single year of AOE application (2019). As such, the exact magnitude and spatial patterns of changes in annual mean PM_{2.5} due to AOE may vary interannually due to variability in meteorology and its effects on natural emissions, pollutant transport, and PM_{2.5} removal, and the variability in background anthropogenic emissions. Moreover, long-term simulations using the CESM2 model suggest that relative changes in tropospheric sulfate aerosol and ozone due to Cl₂ gas-based AOE increase during the first 15 years of continuous application and then stabilize (Li, Meidan et al., 2023). Here, Figure 4 highlights the potential risks in surface PM_{2.5} air quality in different regions in the initial years of deployment. Although the emissions for Cl₂ and Cl₂_BrCl_Br₂ are focused over the oceans, there are larger changes seen on land in both cases than over the oceans—with the exception of the Arctic ocean, where ammonium aerosol drives a larger increase (Figure 4). The largest absolute increases in surface PM_{2.5} in the hydrogen peroxide experiments occur in Europe, India, China, Russia, Equatorial Asia, and southern Africa (Figure 4). The impact of daytime-only emissions (not shown) varied depending on the level of hydrogen peroxide emissions; for the H₂O₂_mid scenario, daytime-only emissions had higher PM_{2.5} concentrations, but for the H₂O₂_low scenario the daytime-only emissions reduced PM_{2.5} concentrations. Similarly to the tropospheric-wide changes, these changes are driven by increases in sulfate with decreases in nitrate partially compensating, particularly in the Cl₂ and Cl₂_BrCl_Br₂ methods. Decreases in nitrate lead to small decreases in total PM_{2.5} in some regions in the Cl₂ scenario. Overall, AOE exacerbates existing PM_{2.5} air quality issues in populated regions of Europe and Asia in year 2019 (Figure 4).



530 **Figure 4: Absolute change in surface PM_{2.5} relative to the standard model in year 2019. Gray boxes: areas already in exceedance of USEPA's annual mean PM_{2.5} primary standard of $9 \mu\text{g m}^{-3}$, which remain in exceedance when the experiment is applied. Black boxes: the experiment brought a region from $< 9 \mu\text{g m}^{-3}$ to an exceedance of $\geq 9 \mu\text{g m}^{-3}$. Blue boxes: experiment led to improved PM_{2.5} from an exceedance to $< 9 \mu\text{g m}^{-3}$.**

Figure 5 shows the absolute change in annual mean surface ozone across model experiments. In the global mean, surface ozone decreases on average in all methods (see Figure S6), with largest mean decreases in the Cl₂_BrCl_Br₂ and H₂O₂_high experiments. At the same time, individual grid boxes in the hydrogen peroxide experiments see increases in annual mean ozone across populated areas (see Figure 5). This has implications for the location of additional OH release, balancing the targeting of methane point sources versus populated areas where ozone may increase. Figure 3 again highlights that H₂O₂_mid has worse air quality impacts for ozone than H₂O₂_high.

Surface CO follows the tropospheric changes presented in Section 3.1.1 with decreases in hydrogen peroxide-based scenarios and increases in chlorine-based scenarios. Increases in surface CO concentrations in the chlorine-based scenarios are small relative to health guidelines and largely occur over the oceans. Surface CO decreases everywhere in the hydrogen peroxide-based experiments. These changes represent the short-term impact over 1 year of applied AOE.



545 **Figure 5: Absolute change in surface ozone (ppb) between model experiments detailed in Table 1 and the standard model in year 2019.**

4 Uncertainties

Model resolution. Here I use a coarse-resolution global model simulation. This is not appropriate to examine point source applications near high methane emitters. Model resolution can lead to biases due to nonlinear atmospheric chemistry. However, these effects vary spatially and temporally and for tropospheric NO₂ in GEOS-Chem are within ±8% of high-resolution simulations (C. Li et al., 2023). Resolution effects are most important in polluted regions and are becoming less important as anthropogenic sources are reduced (C. Li et al., 2023). At the same time, Pennacchio et al. (2025) demonstrated the challenges in representing high-chlorine conditions in global lower-resolution models that result from point source emissions of iron for atmospheric methane oxidation enhancement, including how dilution of iron emission plumes and their interactions with the surrounding NO_x and ozone gradients can change the direction of the change in methane predicted.

Uncertainties in iron salt aerosol mechanism. Representation of chlorine release from iron salt aerosols in models remains highly parameterized. Complexities in iron solubility and speciation are not well represented. Depending on the reaction rate and representation, a given mass of iron salt aerosol release may lead to increases or decreases in the methane lifetime. Additional laboratory studies of natural and engineered iron salt aerosol, including mixtures with ambient species, are needed to improve the understanding of the kinetics and driving factors. Field studies and additional observations that could help evaluate this mechanism in models (e.g., isotopic composition, chloride, chlorine, iron and its speciation) will also aid in constraining this process.

Methods for OH release. The method and location of OH release will have different impacts. Here I only examined hydrogen peroxide as a mechanism for OH release, versus OH release without a specific mechanism. Effects of water vapor or artificial radiation for producing OH could have additional effects on climate forcers and air pollution.

Chemistry and time horizon. Here I use the “simple” SOA mechanism (fixed-yield, direct, and irreversible formation); additional simulations would be needed to better understand the change in SOA production under varying chemical regimes of OH and chlorine addition for methane oxidation. In addition, emissions are kept constant in year 2019. Most of the PM_{2.5} effects predicted here are mediated by sulfate; future reductions in SO₂ emissions would likely limit this effect. We do not include interactive hydrogen (H₂) chemistry in our model simulation. Increased methane oxidation by OH in the hydrogen peroxide-based scenarios could lead to increased atmospheric H₂, as would potential future increases in hydrogen applications and their associated H₂ emissions from leakage. This would lead to a positive feedback on methane, as H₂ reacts with OH and reduces the amount of OH available to oxidize methane (e.g., Bertagni et al., 2022; Ocko & Hamburg, 2022; Warwick et al., 2023). As the simulations here are performed for only 1 year following 1 year of initialization, they cannot assess the stratospheric impact of AOE including on ozone depletion due to the long lifetime of air in the stratosphere. Over 30-year-long CESM2 simulations, Li, Meidan et al. (2023) find that chlorine-based AOE leads to stratospheric ozone depletion. Finally, uncertainties in the representation of halogen chemistry may impact the prediction of unintended consequences. For example, whether the halogen chemical mechanism includes interactions with sulfur seems to lead to different results for how increased chlorine affects sulfate aerosol between this study (increases) and Li, Meidan et al. (2023) (decreases).

5 Conclusions

Here we simulate multiple scenarios of enhancing atmospheric methane oxidation via OH or Cl including hydrogen peroxide release, Cl₂ emissions, and iron salt aerosol in a global chemical transport model to assess the potential for decreasing the methane lifetime and resulting impacts on other climate forcers, stratospheric ozone, and surface air quality. The overall
585 impacts of atmospheric oxidation enhancement methods on climate and human health involve multiple competing factors (see Figure 6).

Based on the present work, current global production of hydrogen peroxide is not sufficient to affect methane on a global scale, and this presents a challenge. Release of gas-phase Cl₂ is promising, but the exact mechanism to accomplish this and ensure it is large enough to reduce rather than increase methane remains a major challenge. Two different model
590 approaches (this work, and Li, Meidan, et al., 2023) show that a large quantity of Cl₂ (>100–300 Tg/yr) must be added to the atmosphere in order to have an impact. Smaller amounts, or increasing particulate chloride as shown in this study, will increase methane. Here we find that emitting particulate iron alone to catalyze chlorine release from sea salt aerosol following the reaction rate parameterization from Chen et al. (2024) can release sufficient Cl₂ to decrease global methane by 2.5% but requires larger emissions than in previous studies (565 Tg/yr) to accomplish a smaller change.

In addition to the impacts on methane, increased atmospheric oxidation via hydrogen peroxide- and chlorine-based
595 methods largely has a climate co-benefit, via increased tropospheric aerosols and decreased tropospheric ozone. Chlorine-based methods increase other greenhouse gases and thus have reduced climate co-benefits.

Overall, ozone-depleting substances are increased in chlorine-based methods and decreased in OH-based methods. Most of the species impacted are short-lived. Longer simulations are needed to understand the full impacts on stratospheric
600 chemistry.

Chlorine-based methods reduce surface ozone air pollution. OH-based methods largely result in ozone reductions as well but lead to increases in ozone in already polluted areas. Surface PM_{2.5} pollution increases over land and in highly populated regions across most chlorine scenarios and all OH-based methods, regardless of whether emissions are limited to the oceans. These increases are on the order of present EPA air quality standards for annual mean PM_{2.5}. These results from a one-year
605 simulation highlight potential risks to air quality during the initial phases of AOE deployment.

Co-emission of bromine with chlorine appears to remove any benefit from chlorine-based approaches. In the case of 20% bromine release by mass with respect to chlorine, the methane lifetime increases by 6.7%. The addition of bromine also generally results in worse outcomes with respect to surface air quality, halogenated greenhouse gases, and ozone-depleting substances. Additional experimental measurements of bromine species released from natural and engineered iron salt aerosol
610 are needed to constrain these effects.

Overall, additional research in higher-resolution, longer-term modeling frameworks as well as laboratory experiments is needed to constrain whether the AOE methods have the desired intended consequences of sufficiently decreasing

atmospheric methane, and the risk level for the unintended consequences of potential increases in halogenated greenhouse gases, ozone-depleting substances, and particulate matter air pollution.

615

		OH-based methods	CI-based methods
Methane		↓	↓
Other climate forcers	aerosol	↑	↑
	GHGs*	↓	↓↑
Ozone-depleting substances		↓	↑
Stratospheric ozone		----	----
Surface air quality	PM2.5	↑	↑
	ozone*	↓↑	↓

Figure 6: The overall impacts of atmospheric oxidation enhancement methods on climate and human health involve multiple competing factors. Note: * = sign of change depends on species (greenhouse gases [GHGs]) or location (ozone).

620

Competing interests. The contact author has declared that there are no competing interests.

Acknowledgments. I acknowledge helpful discussions with Jessica Haskins, Katherine Travis, and Qianjie Chen, and funding from the National Academy of Sciences contract PO-10000845. Portions are reproduced with permission from the National Academy of Sciences, Courtesy of the National Academies Press, Washington, D.C.

Data Availability. The model data used in this study will be made available with a permanent DOI when the review process is complete through the Illinois Data Bank (<https://datbank.illinois.edu/>) and will be subject to the terms of the Creative Commons Attribution 4.0 International (CC BY 4.0). The standard model code is available at <https://doi.org/10.5281/zenodo.5500717>. For the review process, model data is temporarily available at the following link: <https://www.dropbox.com/scl/fo/rtktdj7cru74p200x5q5/A191DLtDyZJaXJlJEPY4j8Y?rlkey=x0yt2wjkb9agv3gf4hwhv57q&st=0l0vx6az&dl=0>

References

635 Abernethy, S., Kessler, M. I., & Jackson, R. B. (2023). Assessing the potential benefits of methane oxidation technologies using a concentration-based framework. *Environmental Research Letters*, 18(9), Article 094064. <https://doi.org/10.1088/1748-9326/acf603>

- Alexander, B., Park, R. J., Jacob, D. J., & Gong, S. (2009). Transition metal-catalyzed oxidation of atmospheric sulfur: Global implications for the sulfur budget. *Journal of Geophysical Research*, *114*(D2). <https://doi.org/10.1029/2008JD010486>
- 640 Allan, W., Struthers, H., & Lowe, D. C. (2007). Methane carbon isotope effects caused by atomic chlorine in the marine boundary layer: Global model results compared with Southern Hemisphere measurements. *Journal of Geophysical Research: Atmospheres*, *112*(D4). <https://doi.org/10.1029/2006JD007369>
- Amos, H. M., Jacob, D. J., Holmes, C. D., Fisher, J. A., Wang, Q., Yantosca, R. M., Corbitt, E. S., Galarneau, E., Rutter, A. P., Gustin, M. S., Steffen, A., Schauer, J. J., Graydon, J. A., St. Louis, V. L., Talbot, R. W., Edgerton, E. S., Zhang, Y., & Sunderland, E. M. (2012). Gas-particle partitioning of atmospheric Hg(II) and its effect on global mercury deposition. *Atmospheric Chemistry and Physics*, *12*(1), 591–603. <https://doi.org/10.5194/acp-12-591-2012>
- 645 Bell, S. C. (2023). WO2023108278 - Systems and Methods for Atmospheric Dispersion of Oxidant for Net Conversion of Atmospheric Methane to Carbon Dioxide.
- Bertagni, M. B., Pacala, S. W., Paulot, F., & Porporato, A. (2022). Risk of the hydrogen economy for atmospheric methane. *Nature Communications*, *13*, Article 7706. <https://doi.org/10.1038/s41467-022-35419-7>
- 650 Chen, Q., Schmidt, J. A., Shah, V., Jaeglé, L., Sherwen, T., & Alexander, B. (2017). Sulfate production by reactive bromine: Implications for the global sulfur and reactive bromine budgets. *Geophysical Research Letters*, *44*(13), 7069–7078. <https://doi.org/10.1002/2017GL073812>
- Chen, Q., Wang, X., Fu, X., Li, X., Alexander, B., Peng, X., Wang, W., Xia, M., Tan, Y., Gao, J., Chen, J., Mu, Y., Liu, P., & Wang, T. (2024). Impact of molecular chlorine production from aerosol iron photochemistry on atmospheric oxidative capacity in North China. *Environmental Science & Technology*, *58*(28), 12585–12597. <https://doi.org/10.1021/acs.est.4c02534>
- 655 Eastham, S. D., Fritz, T., Sanz-Morère, I., Prashanth, P., Allroggen, F., Prinn, R. G., Speth, R. L., & Barrett, S. R. H. (2022). Impacts of a near-future supersonic aircraft fleet on atmospheric composition and climate. *Environmental Science: Atmospheres*, *2*(3), 388–403. <https://doi.org/10.1039/D1EA00081K>
- 660 Eastham, S. D., Weisenstein, D. K., & Barrett, S. R. H. (2014). Development and evaluation of the unified tropospheric–stratospheric chemistry extension (UCX) for the global chemistry-transport model GEOS-Chem. *Atmospheric Environment*, *89*, 52–63. <https://doi.org/10.1016/j.atmosenv.2014.02.001>
- Fairlie, T. D., Jacob, D. J., & Park, R. J. (2007). The impact of transpacific transport of mineral dust in the United States. *Atmospheric Environment*, *41*(6), 1251–1266. <https://doi.org/10.1016/j.atmosenv.2006.09.048>
- 665 Fisher, J. A., Atlas, E. L., Barletta, B., Meinardi, S., Blake, D. R., Thompson, C. R., Ryerson, T. B., Peischl, J., Tzompa-Sosa, Z. A., & Murray, L. T. (2018). Methyl, ethyl, and propyl nitrates: Global distribution and impacts on reactive nitrogen in remote marine environments. *Journal of Geophysical Research: Atmospheres*, *123*(21), 12429–12451. <https://doi.org/10.1029/2018JD029046>
- 670 Fisher, J. A., Jacob, D. J., Travis, K. R., Kim, P. S., Marais, E. A., Chan Miller, C., Yu, K., Zhu, L., Yantosca, R. M., Sulprizio, M. P., Mao, J., Wennberg, P. O., Crounse, J. D., Teng, A. P., Nguyen, T. B., St. Clair, J. M., Cohen, R. C., Romer, P.,

- Nault, B. A., ... Mikoviny, T. (2016). Organic nitrate chemistry and its implications for nitrogen budgets in an isoprene- and monoterpene-rich atmosphere: Constraints from aircraft (SEAC⁴RS) and ground-based (SOAS) observations in the Southeast US. *Atmospheric Chemistry and Physics*, 16(9), 5969–5991. <https://doi.org/10.5194/acp-16-5969-2016>
- 675 Gelaro, R., McCarty, W., Suárez, M. J., Todling, R., Molod, A., Takacs, L., Randles, C. A., Darmenov, A., Bosilovich, M. G., Reichle, R., Wargan, K., Coy, L., Cullather, R., Draper, C., Akella, S., Buchard, V., Conaty, A., Da Silva, A. M., Gu, W., ... Zhao, B. (2017). The Modern-Era Retrospective Analysis for Research and Applications, version 2 (MERRA-2). *Journal of Climate*, 30(14), 5419–5454. <https://doi.org/10.1175/JCLI-D-16-0758.1>
- 680 Gorham, K. A., Abernethy, S., Jones, T. R., Hess, P., Mahowald, N. M., Meidan, D., Johnson, M. S., van Herpen, M. M. J. W., Xu, Y., Saiz-Lopez, A., Röckmann, T., Brashear, C. A., Reinhardt, E., and Mann, D.: Opinion: A research roadmap for exploring atmospheric methane removal via iron salt aerosol, *Atmos. Chem. Phys.*, 24, 5659–5670, <https://doi.org/10.5194/acp-24-5659-2024>, 2024. Gromov, S., Brenninkmeijer, C. A. M., & Jöckel, P. (2018). A very limited role of tropospheric chlorine as a sink of the greenhouse gas methane. *Atmospheric Chemistry and Physics*, 18(13), 9831–9843. <https://doi.org/10.5194/acp-18-9831-2018>
- 685 Guenther, A. B., Jiang, X., Heald, C. L., Sakulyanontvittaya, T., Duhl, T., Emmons, L. K., & Wang, X. (2012). The Model of Emissions of Gases and Aerosols from Nature version 2.1 (MEGAN2.1): An extended and updated framework for modeling biogenic emissions. *Geoscientific Model Development*, 5(6), 1471–1492. <https://doi.org/10.5194/gmd-5-1471-2012>
- 690 Holmes, C. D. (2018). Methane feedback on atmospheric chemistry: Methods, models, and mechanisms. *Journal of Advances in Modeling Earth Systems*, 10(4), 1087–1099. <https://doi.org/10.1002/2017MS001196>
- Holmes, C. D., Bertram, T. H., Confer, K. L., Graham, K. A., Ronan, A. C., Wirks, C. K., & Shah, V. (2019). The role of clouds in the tropospheric NO_x cycle: A new modeling approach for cloud chemistry and its global implications. *Geophysical Research Letters*, 46(9), 4980–4990. <https://doi.org/10.1029/2019GL081990>
- 695 Holmes, C. D., Prather, M. J., Søvde, O. A., & Myhre, G. (2013). Future methane, hydroxyl, and their uncertainties: Key climate and emission parameters for future predictions. *Atmospheric Chemistry and Physics*, 13(1), 285–302. <https://doi.org/10.5194/acp-13-285-2013>
- 700 Horowitz, H. M., Holmes, C., Wright, A., Sherwen, T., Wang, X., Evans, M., Huang, J., Jaeglé, L., Chen, Q., Zhai, S., & Alexander, B. (2020). Effects of sea salt aerosol emissions for marine cloud brightening on atmospheric chemistry: Implications for radiative forcing. *Geophysical Research Letters*, 47(4), Article e2019GL085838. <https://doi.org/10.1029/2019GL085838>
- Hossaini, R., Chipperfield, M. P., Montzka, S. A., Leeson, A. A., Dhomse, S. S., & Pyle, J. A. (2017). The increasing threat to stratospheric ozone from dichloromethane. *Nature Communications*, 8(1), Article 15962. <https://doi.org/10.1038/ncomms15962>

- Hossaini, R., Chipperfield, M. P., Saiz-Lopez, A., Fernandez, R., Monks, S., Feng, W., Brauer, P., & Von Glasow, R. (2016).
705 A global model of tropospheric chlorine chemistry: Organic versus inorganic sources and impact on methane oxidation.
Journal of Geophysical Research: Atmospheres, *121*(23), 14271–14297. <https://doi.org/10.1002/2016JD025756>
- Huang, J., & Jaeglé, L. (2017). Wintertime enhancements of sea salt aerosol in polar regions consistent with a sea ice source
from blowing snow. *Atmospheric Chemistry and Physics*, *17*(5), 3699–3712. <https://doi.org/10.5194/acp-17-3699-2017>
- Jaeglé, L., Quinn, P. K., Bates, T. S., Alexander, B., & Lin, J.-T. (2011). Global distribution of sea salt aerosols: New
710 constraints from in situ and remote sensing observations. *Atmospheric Chemistry and Physics*, *11*(7), 3137–3157.
<https://doi.org/10.5194/acp-11-3137-2011>
- Johnson, M. S., E. J. K. Nilsson, E. A. Svensson, S. Langer, Gas Phase Advanced Oxidation for Effective, Efficient In Situ
Control of Pollution, *Environmental Science & Technology* *48*(15), 8768–8776, 2014.
- Khodayari, A., Olsen, S. C., Wuebbles, D. J., & Phoenix, D. B. (2015). Aviation NO_x-induced CH₄ effect: Fixed mixing ratio
715 boundary conditions versus flux boundary conditions. *Atmospheric Environment*, *113*, 135–139.
<https://doi.org/10.1016/j.atmosenv.2015.04.070>
- Li, C., Martin, R. V., Cohen, R. C., Bindle, L., Zhang, D., Chatterjee, D., Weng, H., & Lin, J. (2023). Variable effects of spatial
resolution on modeling of nitrogen oxides. *Atmospheric Chemistry and Physics*, *23*(5), 3031–3049.
<https://doi.org/10.5194/acp-23-3031-2023>
- 720 Li, Q., Fernandez, R. P., Hossaini, R., Iglesias-Suarez, F., Cuevas, C. A., Apel, E. C., Kinnison, D. E., Lamarque, J.-F., &
Saiz-Lopez, A. (2022). Reactive halogens increase the global methane lifetime and radiative forcing in the 21st century.
Nature Communications, *13*(1), Article 2768. <https://doi.org/10.1038/s41467-022-30456-8>
- Li, Q., Meidan, D., Hess, P., Añel, J. A., Cuevas, C. A., Doney, S., Fernandez, R. P., Van Herpen, M., Höglund-Isaksson, L.,
Johnson, M. S., Kinnison, D. E., Lamarque, J.-F., Röckmann, T., Mahowald, N. M., & Saiz-Lopez, A. (2023). Global
725 environmental implications of atmospheric methane removal through chlorine-mediated chemistry-climate interactions.
Nature Communications, *14*(1), Article 4045. <https://doi.org/10.1038/s41467-023-39794-7>
- Lin, H., Jacob, D. J., Lundgren, E. W., Sulprizio, M. P., Keller, C. A., Fritz, T. M., Eastham, S. D., Emmons, L. K., Campbell,
P. C., Baker, B., Saylor, R. D., & Montuoro, R. (2021). Harmonized Emissions Component (HEMCO) 3.0 as a versatile
emissions component for atmospheric models: Application in the GEOS-Chem, NASA GEOS, WRF-GC, CESM2,
730 NOAA GEFS-Aerosol, and NOAA UFS models. *Geoscientific Model Development*, *14*(9), 5487–5506.
<https://doi.org/10.5194/gmd-14-5487-2021>
- Lin, J.-T., and M. McElroy, *Impacts of boundary layer mixing on pollutant vertical profiles in the lower troposphere:
Implications to satellite remote sensing*, *Atmospheric Environment*, **44**(14), 1726-1739,
doi:10.1016/j.atmosenv.2010.02.009, 2010.
- 735 Liu, H., Jacob, D. J., Bey, I., & Yantosca, R. M. (2001). Constraints from ²¹⁰Pb and ⁷Be on wet deposition and transport in a
global three-dimensional chemical tracer model driven by assimilated meteorological fields. *Journal of Geophysical
Research: Atmospheres*, *106*(D11), 12109–12128. <https://doi.org/10.1029/2000JD900839>

- Mao, J., Jacob, D. J., Evans, M. J., Olson, J. R., Ren, X., Brune, W. H., Clair, J. M. St., Crouse, J. D., Spencer, K. M., Beaver, M. R., Wennberg, P. O., Cubison, M. J., Jimenez, J. L., Fried, A., Weibring, P., Walega, J. G., Hall, S. R., Weinheimer, A. J., Cohen, R. C., Chen, G., Crawford, J. H., McNaughton, C., Clarke, A. D., Jaeglé, L., Fisher, J. A., Yantosca, R. M., Le Sager, P., and Carouge, C.: Chemistry of hydrogen oxide radicals (HO_x) in the Arctic troposphere in spring, *Atmos. Chem. Phys.*, 10, 5823–5838, <https://doi.org/10.5194/acp-10-5823-2010>, 2010.
- 740 Mayhew, A. W., & Haskins, J. D. (2025). Potential air quality side-effects of emitting H₂O₂ to enhance methane oxidation as a climate solution. *Environmental Science & Technology*, 59(1), 679–688.
- 745 McDuffie, E. E., Fibiger, D. L., Dubé, W. P., Lopez-Hilfiker, F., Lee, B. H., Thornton, J. A., Shah, V., Jaeglé, L., Guo, H., Weber, R. J., Reeves, J. M., Weinheimer, A. J., Schroder, J. C., Campuzano-Jost, P., Jimenez, J. L., Dibb, J. E., Veres, P., Ebben, C., Sparks, T. L., ... Brown, S. S. (2018). Heterogeneous N₂O₅ uptake during winter: Aircraft measurements during the 2015 WINTER campaign and critical evaluation of current parameterizations. *Journal of Geophysical Research: Atmospheres*, 123(8), 4345–4372. <https://doi.org/10.1002/2018JD028336>
- 750 Meidan, D., Li, Q., Cuevas, C. A., Doney, S. C., Fernandez, R. P., Van Herpen, M. M. J. W., Johnson, M. S., Kinnison, D. E., Li, L., Hamilton, D. S., Saiz-Lopez, A., Hess, P., & Mahowald, N. M. (2024). Evaluating the potential of iron-based interventions in methane reduction and climate mitigation. *Environmental Research Letters*, 19(5), Article 054023. <https://doi.org/10.1088/1748-9326/ad3d72>
- Moffet, R. C.; Furutani, H.; Rödel, T. C.; Henn, T. R.; Sprau, P. O.; Laskin, A.; Uematsu, M.; Gilles, M. K. Iron Speciation and Mixing in Single Aerosol Particles from the Asian Continental Outflow. *J. Geophys. Res.* 2012, 117, No. D07204.
- 755 Ming, T., Li, W., Yuan, Q., Davies, P., de Richter, R., Peng, C., Deng, Q., Yuan, Y., Caillol, S., & Zhou, N. (2022). Perspectives on removal of atmospheric methane. *Advances in Applied Energy*, 5, Article 100085. <https://doi.org/10.1016/j.adapen.2022.100085>
- Moffet, R. C., Furutani, H., Rödel, T. C., Henn, T. R., Sprau, P. O., Laskin, A., Uematsu, M., & Gilles, M. K. (2012). Iron speciation and mixing in single aerosol particles from the Asian continental outflow: Aerosol iron speciation in Asian outflow. *Journal of Geophysical Research: Atmospheres*, 117(D7). <https://doi.org/10.1029/2011JD016746>
- 760 Murray, L. T. (2016). Lightning NO_x and impacts on air quality. *Current Pollution Reports*, 2(2), 115–133. <https://doi.org/10.1007/s40726-016-0031-7>
- Murray, L. T., Fiore, A. M., Shindell, D. T., Naik, V., & Horowitz, L. W. (2021). Large uncertainties in global hydroxyl projections tied to fate of reactive nitrogen and carbon. *Proceedings of the National Academy of Sciences*, 118(43), Article e2115204118. <https://doi.org/10.1073/pnas.2115204118>
- 765 National Academies of Sciences, Engineering and Medicine (NASEM) report [2024] 'A Research Agenda Toward Atmospheric Methane Removal'
- Neu, J., Prather, M., and Penner, J.: Global atmospheric chemistry: integrating over fractional cloud cover, *J. Geophys. Res.*, 112, D11306, 12 pp., doi:10.1029/2006JD008007, 2007.
- 770

- Ocko, I. B., & Hamburg, S. P. (2022). Climate consequences of hydrogen emissions. *Atmospheric Chemistry and Physics*, 22(14), 9349–9368. <https://doi.org/10.5194/acp-22-9349-2022>
- 775 O'Connor, F. M., Johnson, B. T., Jamil, O., Andrews, T., Mulcahy, J. P., & Manners, J. (2022). Apportionment of the pre-industrial to present-day climate forcing by methane using UKESM1: The role of the cloud radiative effect. *Journal of Advances in Modeling Earth Systems*, 14(10), Article e2022MS002991. <https://doi.org/10.1029/2022MS002991>
- Oeste, F. D., de Richter, R., Ming, T., & Caillol, S. (2017). Climate engineering by mimicking natural dust climate control: The iron salt aerosol method. *Earth System Dynamics*, 8(1), 1–54. <https://doi.org/10.5194/esd-8-1-2017>
- 780 Ordóñez, C., Lamarque, J.-F., Tilmes, S., Kinnison, D. E., Atlas, E. L., Blake, D. R., Sousa Santos, G., Brasseur, G., & Saiz-Lopez, A. (2012). Bromine and iodine chemistry in a global chemistry-climate model: Description and evaluation of very short-lived oceanic sources. *Atmospheric Chemistry and Physics*, 12(3), 1423–1447. <https://doi.org/10.5194/acp-12-1423-2012>
- 785 Pai, S. J., Heald, C. L., Pierce, J. R., Farina, S. C., Marais, E. A., Jimenez, J. L., Campuzano-Jost, P., Nault, B. A., Middlebrook, A. M., Coe, H., Shilling, J. E., Bahreini, R., Dingle, J. H., & Vu, K. (2020). An evaluation of global organic aerosol schemes using airborne observations. *Atmospheric Chemistry and Physics*, 20(5), 2637–2665. <https://doi.org/10.5194/acp-20-2637-2020>
- Pennacchio, L., Mikkelsen, M. K., Krogsbøll, M., van Herpen, M., and Johnson, M. S. (2024). Physical and practical constraints on atmospheric methane removal technologies. *Environmental Research Letters*, 19(10), 104058.
- 790 Pennacchio, L.; van Herpen, M.; Meidan, D.; Saiz-Lopez, A.; Johnson, M. S. Catalytic Efficiencies for Methane Removal: Impact of HO_x, NO_x, and Chemistry in the High-Chlorine Regime. *ACS Earth Space Chem.* **2025**, 9, 504–512, DOI: 10.1021/acsearthspacechem.4c00283
- Philip, S., Martin, R. V., Snider, G., Weagle, C. L., Van Donkelaar, A., Brauer, M., Henze, D. K., Klimont, Z., Venkataraman, C., Guttikunda, S. K., & Zhang, Q. (2017). Anthropogenic fugitive, combustion and industrial dust is a significant, underrepresented fine particulate matter source in global atmospheric models. *Environmental Research Letters*, 12(4), Article 044018. <https://doi.org/10.1088/1748-9326/aa65a4>
- 795 Platt, U., Allan, W., & Lowe, D. (2004). Hemispheric average Cl atom concentration from ¹³C/¹²C ratios in atmospheric methane. *Atmospheric Chemistry and Physics*, 4(9/10), 2393–2399. <https://doi.org/10.5194/acp-4-2393-2004>
- Pound, R. J., Sherwen, T., Helmig, D., Carpenter, L. J., & Evans, M. J. (2020). Influences of oceanic ozone deposition on tropospheric photochemistry. *Atmospheric Chemistry and Physics*, 20(7), 4227–4239. <https://doi.org/10.5194/acp-20-4227-2020>
- 800 Research and Markets. (2023). *Global hydrogen peroxide market analysis: Plant capacity, production, operating efficiency, technology, demand & supply, end user industries, distribution channel, regional demand, 2015-2030*. <https://www.researchandmarkets.com/reports/5174562/global-hydrogen-peroxide-market-analysis-plant>

- Saiz-Lopez, A., Fernandez, R. P., Ordóñez, C., Kinnison, D. E., Gómez Martín, J. C., Lamarque, J.-F., & Tilmes, S. (2014).
805 Iodine chemistry in the troposphere and its effect on ozone. *Atmospheric Chemistry and Physics*, *14*(23), 13119–13143.
<https://doi.org/10.5194/acp-14-13119-2014>
- V. Shah, L. Jaeglé, J.A. Thornton, F.D. Lopez-Hilfiker, B.H. Lee, J.C. Schroder, P. Campuzano-Jost, J.L. Jimenez, H. Guo,
A.P. Sullivan, R.J. Weber, J.R. Green, M.N. Fiddler, S. Bililign, T.L. Campos, M. Stell, A.J. Weinheimer, D.D. Montzka,
& S.S. Brown, Chemical feedbacks weaken the wintertime response of particulate sulfate and nitrate to emissions
810 reductions over the eastern United States, *Proc. Natl. Acad. Sci. U.S.A.* *115* (32) 8110-8115,
<https://doi.org/10.1073/pnas.1803295115> (2018).
- Shao, J., Chen, Q., Wang, Y., Lu, X., He, P., Sun, Y., Shah, V., Martin, R. V., Philip, S., Song, S., Zhao, Y., Xie, Z., Zhang,
L., & Alexander, B. (2019). Heterogeneous sulfate aerosol formation mechanisms during wintertime Chinese haze
events: Air quality model assessment using observations of sulfate oxygen isotopes in Beijing. *Atmospheric Chemistry*
815 *and Physics*, *19*(9), 6107–6123. <https://doi.org/10.5194/acp-19-6107-2019>
- Sherwen, T., Evans, M. J., Carpenter, L. J., Andrews, S. J., Lidster, R. T., Dix, B., Koenig, T. K., Sinreich, R., Ortega, I.,
Volkamer, R., Saiz-Lopez, A., Prados-Roman, C., Mahajan, A. S., & Ordóñez, C. (2016a). Iodine’s impact on
tropospheric oxidants: A global model study in GEOS-Chem. *Atmospheric Chemistry and Physics*, *16*(2), 1161–1186.
<https://doi.org/10.5194/acp-16-1161-2016>
- 820 Sherwen, T., Schmidt, J. A., Evans, M. J., Carpenter, L. J., Großmann, K., Eastham, S. D., Jacob, D. J., Dix, B., Koenig, T.
K., Sinreich, R., Ortega, I., Volkamer, R., Saiz-Lopez, A., Prados-Roman, C., Mahajan, A. S., & Ordóñez, C. (2016b).
Global impacts of tropospheric halogens (Cl, Br, I) on oxidants and composition in GEOS-Chem. *Atmospheric Chemistry*
and Physics, *16*(18), 12239–12271. <https://doi.org/10.5194/acp-16-12239-2016>
- Shi, J., Guan, Y., Gao, H., Yao, X., Wang, R., & Zhang, D. (2022). Aerosol iron solubility specification in the global marine
825 atmosphere with machine learning. *Environmental Science & Technology*, *56*(22), 16453–16461.
<https://doi.org/10.1021/acs.est.2c05266>
- Smith, C., Nicholls, Z. R. J., Armour, K., Collins, W., Forster, P., Meinshausen, M., Palmer, M. D., & Watanabe, M. (2021).
The Earth’s energy budget, climate feedbacks and climate sensitivity supplementary material. In V. Masson-Delmotte,
P. Zhai, A. Pirani, S. L. Connors, C. Péan, S. Berger, N. Caud, Y. Chen, L. Goldfarb, M. I. Gomis, M. Huang, K. Leitzell,
830 E. Lonnoy, J. B. R. Matthews, T. K. Maycock, T. Waterfield, Ö. Yelekçi, R. Yu, & B. Zhou (Eds.), *Climate change*
2021: The physical science basis. Contribution of working group I to the Sixth Assessment Report of the
Intergovernmental Panel on Climate Change.
https://www.ipcc.ch/report/ar6/wg1/downloads/report/IPCC_AR6_WGI_Chapter07_SM.pdf
- Tao, T., Wang, Y., Ming, T., Mu, L., de Richter, R., & Li, W. (2023). Downdraft energy tower for negative emissions: Analysis
835 on methane removal and other co-benefits. *Greenhouse Gases: Science and Technology*, *13*(5), 713–720.
<https://doi.org/10.1002/ghg.2233>

- Taylor, S. R., & McLennan, S. M. (1985). *The continental crust: Its composition and evolution*.
<https://www.osti.gov/biblio/6582885>
- 840 Trapp, J. M., Millero, F. J., & Prospero, J. M. (2010). Trends in the solubility of iron in dust-dominated aerosols in the equatorial Atlantic trade winds: Importance of iron speciation and sources: Fe solubility in North Atlantic trade wind aerosols. *Geochemistry, Geophysics, Geosystems*, *11*(3). <https://doi.org/10.1029/2009GC002651>
- 845 van Herpen, M. M. J. W., Li, Q., Saiz-Lopez, A., Liisberg, J. B., Röckmann, T., Cuevas, C. A., Fernandez, R. P., Mak, J. E., Mahowald, N. M., Hess, P., Meidan, D., Stuut, J.-B. W., & Johnson, M. S. (2023). Photocatalytic chlorine atom production on mineral dust–sea spray aerosols over the North Atlantic. *Proceedings of the National Academy of Sciences*, *120*(31), Article e2303974120. <https://doi.org/10.1073/pnas.2303974120>
- Wang, Q., Jacob, D. J., Spackman, J. R., Perring, A. E., Schwarz, J. P., Moteki, N., Marais, E. A., Ge, C., Wang, J., & Barrett, S. R. H. (2014). Global budget and radiative forcing of black carbon aerosol: Constraints from pole-to-pole (HIPPO) observations across the Pacific. *Journal of Geophysical Research: Atmospheres*, *119*(1), 195–206. <https://doi.org/10.1002/2013JD020824>
- 850 Wang, X., Jacob, D. J., Downs, W., Zhai, S., Zhu, L., Shah, V., Holmes, C. D., Sherwen, T., Alexander, B., Evans, M. J., Eastham, S. D., Neuman, J. A., Veres, P. R., Koenig, T. K., Volkamer, R., Huey, L. G., Bannan, T. J., Percival, C. J., Lee, B. H., & Thornton, J. A. (2021). Global tropospheric halogen (Cl, Br, I) chemistry and its impact on oxidants. *Atmospheric Chemistry and Physics*, *21*(18), 13973–13996. <https://doi.org/10.5194/acp-21-13973-2021>
- 855 Wang, X., Jacob, D. J., Eastham, S. D., Sulprizio, M. P., Zhu, L., Chen, Q., Alexander, B., Sherwen, T., Evans, M. J., Lee, B. H., Haskins, J. D., Lopez-Hilfiker, F. D., Thornton, J. A., Huey, G. L., & Liao, H. (2019). The role of chlorine in global tropospheric chemistry. *Atmospheric Chemistry and Physics*, *19*(6), 3981–4003. <https://doi.org/10.5194/acp-19-3981-2019>
- 860 Wang, X., Jacob, D. J., Fu, X., Wang, T., Breton, M. L., Hallquist, M., Liu, Z., McDuffie, E. E., & Liao, H. (2020). Effects of anthropogenic chlorine on PM_{2.5} and ozone air quality in China. *Environmental Science & Technology*, *54*(16), 9908–9916. <https://doi.org/10.1021/acs.est.0c02296>
- Wang, Y., Jacob, D. J., & Logan, J. A. (1998). Global simulation of tropospheric O₃-NO_x-hydrocarbon chemistry: 1. Model formulation. *Journal of Geophysical Research: Atmospheres*, *103*(D9), 10713–10725. <https://doi.org/10.1029/98JD00158>
- 865 Wang, Y., Ming, T., Li, W., Yuan, Q., de Richter, R., Davies, P., & Caillol, S. (2022). Atmospheric removal of methane by enhancing the natural hydroxyl radical sink. *Greenhouse Gases: Science and Technology*, *12*(6), 784–795. <https://doi.org/10.1002/ghg.2191>
- Warwick, N. J., Archibald, A. T., Griffiths, P. T., Keeble, J., O'Connor, F. M., Pyle, J. A., & Shine, K. P. (2023). Atmospheric composition and climate impacts of a future hydrogen economy. *Atmospheric Chemistry and Physics*, *23*(20), 13451–13467. <https://doi.org/10.5194/acp-23-13451-2023>

- 870 Weng, H., Lin, J., Martin, R., Millet, D. B., Jaeglé, L., Ridley, D., Keller, C., Li, C., Du, M., & Meng, J. (2020). Global high-resolution emissions of soil NO_x, sea salt aerosols, and biogenic volatile organic compounds. *Scientific Data*, 7(1), Article 148. <https://doi.org/10.1038/s41597-020-0488-5>
- Wittmer, J., Bleicher, S., Ofner, J., & Zetzsch, C. (2015b). Iron(III)-induced activation of chloride from artificial sea-salt aerosol. *Environmental Chemistry*, 12(4), 461–475. <https://doi.org/10.1071/EN14279>
- 875 Wittmer, J., Bleicher, S., & Zetzsch, C. (2015a). Iron(III)-induced activation of chloride and bromide from modeled salt pans. *The Journal of Physical Chemistry A*, 119(19), 4373–4385. <https://doi.org/10.1021/jp508006s>
- World Chlorine Council <https://worldchlorine.org/wp-content/themes/brickthemewp/pdfs/sustainablefuture.pdf>
- Zhang, J., Wuebbles, D. J., Kinnison, D. E., & Saiz-Lopez, A. (2020). Revising the ozone depletion potentials metric for short-lived chemicals such as CF₃I and CH₃I. *Journal of Geophysical Research: Atmospheres*, 125(9), Article e2020JD032414. <https://doi.org/10.1029/2020JD032414>
- 880 Zhang, L., Gong, S., Padro, J., & Barrie, L. (2001). A size-segregated particle dry deposition scheme for an atmospheric aerosol module. *Atmospheric Environment*, 35(3), 549–560. [https://doi.org/10.1016/S1352-2310\(00\)00326-5](https://doi.org/10.1016/S1352-2310(00)00326-5)
- Zhu, L., Jacob, D. J., Eastham, S. D., Sulprizio, M. P., Wang, X., Sherwen, T., Evans, M. J., Chen, Q., Alexander, B., Koenig, T. K., Volkamer, R., Huey, L. G., Le Breton, M., Bannan, T. J., & Percival, C. J. (2019). Effect of sea salt aerosol on tropospheric bromine chemistry. *Atmospheric Chemistry and Physics*, 19(9), 6497–6507. <https://doi.org/10.5194/acp-19-6497-2019>
- 885 Zhu, X., Prospero, J. M., Savoie, D. L., Millero, F. J., Zika, R. G., & Saltzman, E. S. (1993). Photoreduction of iron(III) in marine mineral aerosol solutions. *Journal of Geophysical Research: Atmospheres*, 98(D5), 9039–9046. <https://doi.org/10.1029/93JD00202>
- 890 Zhu, Y.; Li, W.; Wang, Y.; Zhang, J.; Liu, L.; Xu, L.; Xu, J.; Shi, J.; Shao, L.; Fu, P.; Zhang, D.; Shi, Z. Sources and Processes of Iron Aerosols in a Megacity in Eastern China. *Atmos. Chem. Phys.* 2022, 22 (4), 2191–2202.