1   **The evolution of aerosols mixing state derived from a field campaign in**

2   **Beijing: implications to the particles aging time scale in urban atmosphere**

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13 **Abstract**

14   The mixing states and aging time scale of aerosol particles play a vital role in evaluating

15 their climate effects. Here, we identified four different real-time mixing patterns of size-

16 resolved particles using the field measurement by a humidity tandem differential mobility

17 analyzer (H-TDMA) in the urban Beijing. We show that the particles with external, transitional

18 and internal mixing state during the campaign account for 20-48%, 17-24% and 27-56%

19 respectively and the fraction highly depends on particles size. The diurnal variation of the

20 mixing states of particles in all sizes investigated (40, 80, 110, 150 and 200 nm) present an

21 apparent aging process from external to internal mixing state, typically spanning a duration of

22 approximately 5–8 hours from 8:00–10:00 to 15:00–17:00. Additionally, the results illustrate

23 that high ambient temperature during daytime or more humid atmosphere accelerates the aging

24 process of aerosol particles, leading to the particles from external to internal mixing on both

25 clear and cloudy days. Also, with the evolution of particulate pollution, the aerosol particles

26 become more internally-mixed. Our result implies that those fine aerosol particles experience

27 aging through both the photochemical process and aqueous growth in the polluted atmosphere

28 of urban Beijing. Furthermore, through a comprehensive review of the aging timescale of

29 particles adopted in current models and derived from observations, we show the great

30 discrepancy between observations and models, highlighting the importance to parameterize

31 their aging time scale based on more field campaigns.

32

33 **1 Introduction**

34   The mixing state of atmospheric aerosol particles can affect the hygroscopicity and the

35 ability to serve as cloud condensation nuclei (CCN), and thus the air quality and climate (Müller

36 et al., 2017; Xu et al., 2021; Yao et al., 2022; Ge et al., 2024). It has been shown that the aerosol

37 mixing state is closely related to the hygroscopicity (Chen et al., 2022; Fan et al., 2020). Ren

38 et al. (2018) predicted the concentration of CCN using five different mixing state schemes and

39 found that the influence of aerosol mixing state on its activation characteristics ranged from –

40 34% to +16%. Neglecting particle mixing structure can also lead to significant overestimation

41 of the aerosol absorption efficiency (Yao et al., 2022). Therefore, it is important to account for

42 the information of mixing state of ambient particles in climate models so as to reduce the

43 uncertainty in evaluating their environmental and climate effects.

44   The mixing states of ambient particles are complex. Particles in areas affected by primary

45 emissions are mainly with external mixing state (i.e., the chemical components of particles

46 exist independently), while aerosols in relatively clean areas are mainly transported from

47 elsewhere and have a higher degree of internal mixing (Swietlicki et al., 2008; Enroth et al.,

48 2018; Chen et al., 2022). Internal mixing typically includes uniform composition or core–shell

49 structures (Jacobson et al., 2001). The former refers to the same proportion of species in any

50 part of the aerosol component, while the latter is defined as the mixing state formed by certain

51 chemical components coating or condensing on the surface of other components during the

52 aging process. Also, the mixing state of particles is variable. Freshly emitted particles undergo

53 various processes, including photochemical and aqueous-phase processes, as well as physical

54 processes such as coagulation and condensation, leading to an increase in their degree of

55 internal mixing. This gradual transition from external to internal mixing characterizes the aging

56 process of particles. The aging timescale varies greatly between clean and polluted areas (Peng

57 et al., 2016; Chen et al., 2017; Ghosh et al., 2021). However, the timescale of aging process of

58 particles was commonly fixed in many models and did not depend on environmental conditions

59 (Chen et al., 2017; Ghosh et al., 2021), which may introduce great uncertainty in the prediction

60 of regional aerosol concentration and the evaluation of aerosol climate effects (Ghosh et al.,

61 2021). Therefore, capturing the temporal scales of the evolution of the aerosol mixing state

62 based on field campaigns is crucial for accurately parameterizing the aging timescale of aerosol

63 particles in models, thereby enhancing the precision of simulations pertaining to the

64 environmental and climatic impacts of aerosols.

65   At present, some studies have characterized the mixing state and aging process of black

66 carbon (BC) aerosols using different instruments. Transmission electron microscopy (TEM)

67 has been used to determine the mixing state of individual particles in China (Li et al., 2016;

68 Zhang J. et al., 2023). However, based on TEM technique, a large number of aerosol samples

69   are required so as to make the results with the significance of statistics, which means a high

70   cost both on labors and materials. The most important issue is that the mixing state of particles

71   may change during collection and transportation. Recently, aerosol time-of-flight mass

72   spectrometry (ATOFMAS) and soot-particle aerosol mass spectrometry (SP-AMS) have been

73   used to measure the mixing state of BC and coated aerosol species in real time (Liu et al., 2019;

74   Xie et al., 2020). Overall, previous studies focused more on the mixing state of BC of single-

75   particles (Saha et al., 2018; Xie et al., 2020; Chen et al., 2020). While, the mixing state of

76   particles across a population in ambient atmosphere is more complex (Riemer et al., 2019). For

77   example, aerosols containing two or more components may also exhibit external mixing, as

78   evidenced by the bi- or trimodal distributions of the probability density function of hygroscopic

79   growth factor/hygroscopic parameter (κ) (Gf-PDF; κ-PDF) measured using a humidity tandem

80   differential mobility analyzer (H-TDMA) (Fan et al., 2020; Chen et al., 2022; Zhang S. et al.,

81   2023). The mixing state of particles has been retrieved according to the patterns of Gf-PDF/κ-

82   PDF (Hong et al., 2018; Shi et al., 2022; Spitieri et al., 2023). However, most studies based on

83   the H-TDMA measurements only made qualitative descriptions of the particle mixing state

84   when explaining the variations in the aerosol hygroscopicity (Wang et al., 2019; Chen et al.,

85   2022; Shi et al., 2022).

86   In this study, with the aim of obtaining insights into the mixing state and aging time scale

87   of ambient particles in urban area, we have identified four different types of size-resolved

88   particles mixing states, and characterized their real-time variations using the field measured

89   hygroscopic growth factor by the H-TDMA in urban Beijing. The dependence of mixing state

90   of particles on temperature (T), relative humidity (RH) and the pollution levels was

91   investigated. The evolution of mixing state of particles with specific sizes was also explored to

92   imply the timescale of their aging from the diurnal variations during clear and cloudy days.

93   Finally, we compared the aging timescale of aerosol particles with that adopted in current

94   models and other field observations reported in previous literatures.

95   **2 Methods**

96   The campaign was conducted to measure the Gf of particles with different dry sizes (40,

97   80, 110, 150 and 200 nm) using a H-TDMA from 19 May to 18 June 2017 in Beijing. The

98   instrument used in this study has been described previously (Zhang et al., 2017; Wang et al.,

99   2019; Fan et al., 2020; Chen et al., 2022). Here, we mainly describe the interpretation criteria

100 for the mixing state of particles. The Gf-PDF was obtained by TDMA$_{inv}$ algorithm in this study

101 (Gysel et al., 2009). Then, the κ-PDF of size-resolved particles was retrieved according to the

102 κ-Köhler theory (Petters and Kreidenweis, 2007). Subsequently, we accurately defined four

103 mixing states based on the κ-PDF patterns and the number of peaks in κ-PDF (Figure 1, Figure

104 S1). Given that the freshly emitted hydrophobic particles showed an external mixing state (Li

105 et al., 2016), the κ-PDF, that exhibited only one nearly hydrophobic mode or less hygroscopic

106 (LH; the peak of κ-PDF occurs at $\kappa < 0.1$) mode, was defined as LH external mixing state (type

107 1). Mixing type 2 indicates that the maximum peak of κ-PDF occurs at LH mode or more

108 hygroscopic (MH, the peak of κ-PDF occurs at $\kappa \geq 0.1$) mode, and the bimodal distributions of

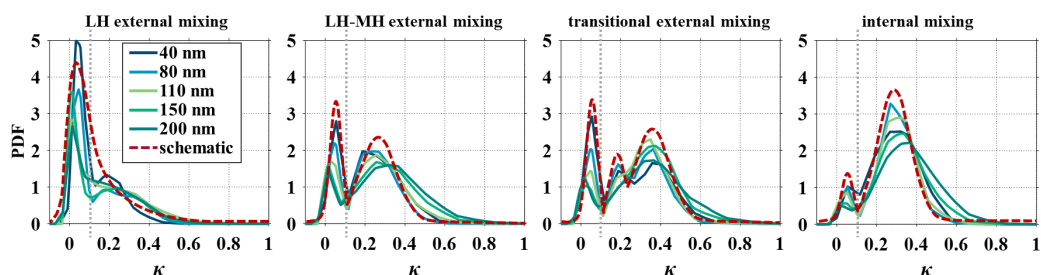
109 the LH and MH modes are present in κ-PDF. Moreover, the κ-PDF pattern with trimodal

110 distributions called transitional external mixing state in this study (type 3). The κ-PDF of type

111 4 was dominated by MH mode, displaying nearly unimodal patterns, was defined as internal

112 mixing type. In this study, we also calculated the standard deviation of κ-PDF (σ) according to

113 Spitieri et al. (2023), which indicates the degree of dispersion in the data, and is thought to

114 reflect the mixing degree of particles to a certain extent.

116 Figure 1. The mean κ-PDF of five particle sizes and the schematic diagram of κ-PDF for four

117 mixing types.

118 **3 Results and discussion**

119 **3.1 Overview of the mixing state of size-resolved particles**

120 Figure 2 shows the time series of size-resolved κ, and κ versus σ. The real-time changes

121 in the mixing states of particles are represented by different filled colors. The κ shows large

122 variations in five particle sizes, ranging from 0 to 0.5, accompanied by significant changes in

123 mixing states. In general, the mixing state of particles gradually transitions from external

124 mixing to internal mixing along with increases in κ. For example, particles with size of 40 nm

125 are dominated by LH external mixing and LH-MH external mixing states at $\kappa < 0.2$; and the

126 transitional external mixing state take up an increasingly proportion often corresponding to κ

127 values of >0.2 and the internal mixing state occurs when κ greater than 0.3 (shown by the light

128 gray dots in Figure 2).

129 However, there exists different correlations between κ and σ for the five sizes. For

130 particles smaller than 100 nm (i.e., 40 and 80 nm), the σ increases as the κ increases from 0.1

131 to 0.4, showing a positive correlation. While, the σ for particles larger than 100 nm (i.e., 110,

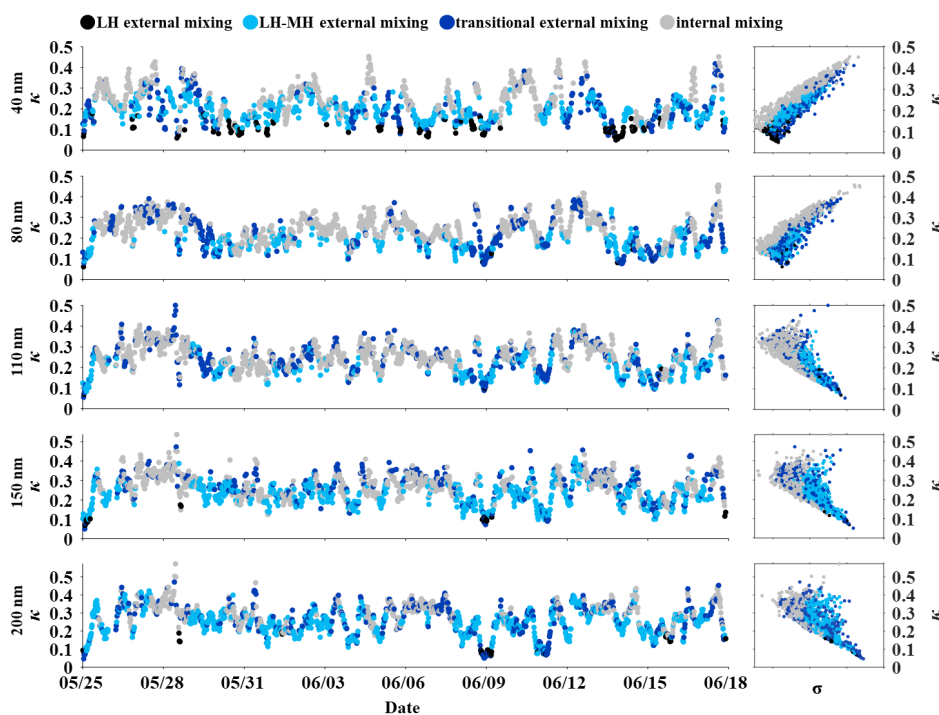
132   150 and 200 nm) exhibits a negative trend to κ variations, with the internal mixing and strong

133   particles hygroscopicity but the smaller the value of σ. Our result contrasts with that previously

134   observed in a vegetated site by Spitieri et al. (2023), which used only σ of Gf-PDF as a single

135   indicator to characterize the particles mixing state. Our result indicates that the parameter σ

136   alone cannot characterize the mixing state of particles in megacity of Beijing, where the aerosol

137   particles are usually severely affected by local anthropogenic emissions.

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139   Figure 2. Time series of hygroscopic parameter (κ), and plots of κ versus standard deviation of

140   κ-PDF (σ) for 40, 80, 110, 150 and 200 nm particles (from top to bottom). The colors denote

141   the different mixing types.

142   The size dependence of mixing type fractions and σ are shown in Figure 3. The external

143   mixing types, including LH external mixing state and LH-MH external mixing state, dominate

144 in 40, 150 and 200 nm particles, accounting for 47%, 41% and 48% respectively of all mixing

145 types. The elevated fraction in particles with external mixing state is intimately associated with

146 the local sources. For instance, previous study revealed the particle size distribution of aerosols

147 from different sources in urban Beijing, suggesting that the particles with 40 nm predominantly

148 originated from primary emissions, and particles with 150 and 200 nm displayed a strong

149 correlation with transportation sources and biomass burning sources (Ren et al., 2023).

150 Furthermore, for 40 nm particles, besides the impact from primary sources, the relative higher

151 fraction of particles with internal mixing state (37%) is likely due to that the growth and aging

152 process of newly generated particles that are with high hygroscopicity (Wu et al., 2016; Liu et

153 al., 2021a). This is similar to the result observed in Athens, Greece, in which the particles with

154 30 nm were more internally-mixed (Spitieri et al., 2023). While, for particles with sizes of 80

155 and 110 nm, the transitional and internal mixing states totally account for 50-80% of all mixing

156 types, corresponding to a significant decrease from about 10% to less than 1% of the fraction

157 of the LH external mixing type. Note that the proportion of transitional external mixing state

158 remains relative constant, with a mean value of 21% across all sizes in this study, implying a

159 continuous influence of atmospheric aging process on the particles mixing state. On average,

160 the result reveals that the externally and internally mixed particles accounted for 56%±12%

161 and 44%±12% respectively. The mixing state of particles derived in this study differs from that

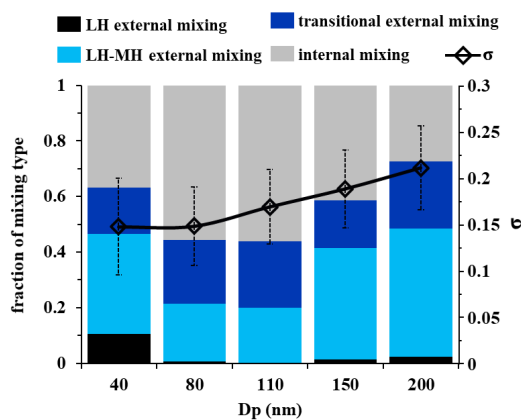
162 reported by Zhang et al., (2017), in which they conducted the measurements at a suburban site

163 of Xinzhou where the aerosols is much less affected by sources nearby, and the aerosols are

164 mainly transported from elsewhere and are thus more aged and well mixed. The result implies

165 that the influence of mixing state on hygroscopicity should be explored at specific particle sizes

166   due to the heterogeneity of chemical compositions with particle size (Fan et al., 2020, Figure

167   S2).

168

169   Figure 3. The size dependence of fraction of particles mixing types and σ.

170   Further, we compared the dependence of κ and σ on the variations of mixing states for five

171   particle sizes (Figure 4). At 40 and 80 nm sizes, it shows that both κ and σ increases markedly

172   when the particles change from external to internal mixing state. While, for particles with sizes

173   of 110, 150 and 200 nm, the σ reduced but the κ increased when the particles change from

174   external to internal mixing state. In other words, the σ of the particles larger than 80 nm shifts

175   to lower values as the particles become more internally mixed. Our results indicate that the

176   standard deviation of κ-PDF alone is insufficient to characterize mixing degree in polluted area.

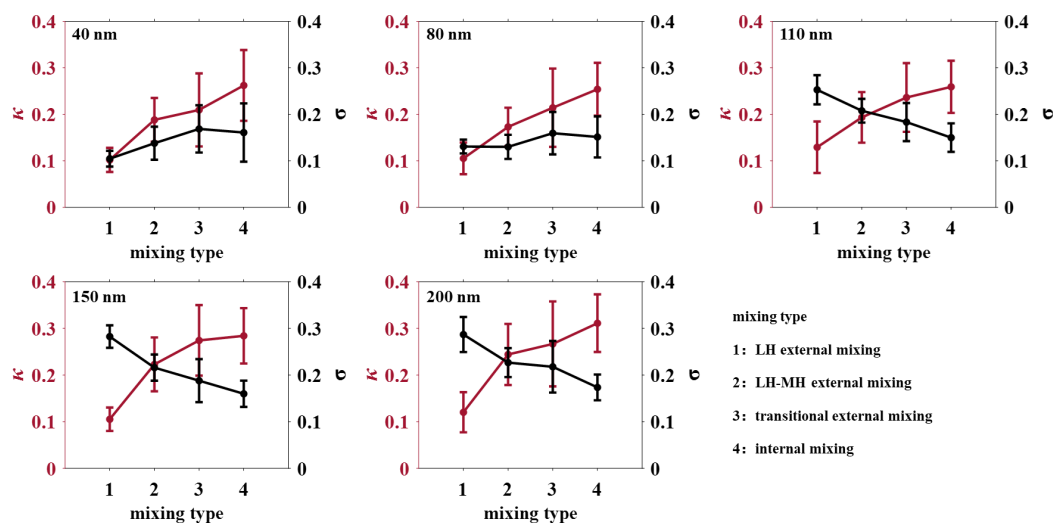



Figure 4. The mixing type dependence of hygroscopicity and σ for five particle sizes.

3.2 Evolution of mixing state of the particles

3.2.1 Diurnal variations

The average diurnal variations of the mixing state and κ of different particle sizes are shown in Figure 5. For 40 nm particles, the fraction of LH external mixing state presents three peaks at morning (9:00–12:00 local time; LT), evening rush hours (18:00–20:00 LT) and nighttime (0:00–03:00 LT). This is accompanied with the impacts from those primary cooking and traffic emissions (Xu et al., 2021; Liu et al., 2021b). Unlike the 40 nm particles, there is no apparent increase of external mixing state for the particles with sizes of 80, 110, 150, 200 nm in the rush hours or cooking times. The results indicate that the particles emitted from local primary sources are small mainly with sizes around 40 nm during the campaign. Correspondingly, the proportion of the particles with internal mixing state was smallest in the morning and nighttime, but exhibiting a rapid increase from about 9:00 until the evening rush hours (about 16:00–18:00). The particles with LH-MH external mixing and transitional

192   external mixing state represent the intermediate state of the aging process in which particles

193   transition from external mixing to internal mixing, and both showed a decrease to around 10%

194   during the corresponding period of the day as the proportion of particles with internal mixing

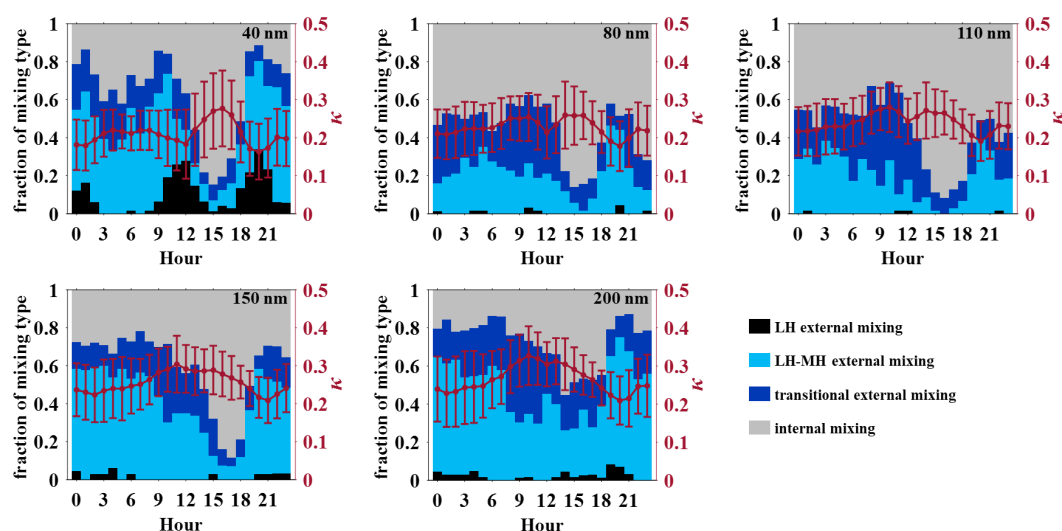
195   state increased up to 80%. The diurnal variations imply an apparent aging process that leads

196   the particles changed from external mixing in the early morning to internal mixing in the

197   afternoon. The aging process also changes particles physiochemical properties. As a result, an

198   obvious enhancement in particles hygroscopicity was observed during the daytime, indicating

199   the impact of particles mixing and aging on their hygroscopicity (Hersey et al., 2013; Müller

200   et al., 2017).

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202   Figure 5. Diurnal variations of the mixing state fraction and κ for five particle sizes.

203   **3.2.2 Dependence of the mixing state on T, RH and PM$_1$**

204   To elucidate the effect of meteorological condition and particulate pollution level on

205   mixing state of particles, the dependences of particles mixing types on ambient T, RH and PM$_1$

206   mass concentrations are further examined (Figure 6). It shows that there is a clear dependence

207   of mixing state of the particles across all observed sizes on T, for example, for the 40 nm

208   particles, the fraction with external mixing state reduces from 60% to 20%, while that of the

209   internal mixing state increases from 30% to 70% with the increase of T from <20℃ to >35℃.

210   This can be easily explained that the higher ambient T (usually corresponding to the periods in

211   daytime) will generally promote the photochemical aging and mixing of the particles (Riemer

212   et al., 2004; Zhang et al., 2015). But note that, the fraction of particles with internal mixing

213   state decreases evidently when the T increases from 15 ℃ to about 25℃. With the increase of

214   RH from 20% to 80%, the fraction of particles with internal mixing state at 40 nm reduces from

215   60% to 20%. However, the fraction of the internal mixing state for particles larger than 80 nm

216   is reduced with the increase of RH when its value <60%, while it increases with the RH increase

217   as it larger than 60%, indicating that the very humid atmosphere with high RH is more

218   favorable for particle aging that was reported promoted by the aerosols aqueous process (Zhang

219   et al., 2021).

220   With the evolution of PM_1 pollution, the fraction of internal mixing state for 40 nm

221   particles significantly decreases to less than 10% due to accumulation of small particles from

222   local primary emissions under heavy pollution conditions, which thus leads to enhanced

223   heterogeneity of particles chemical composition and externally-mixed degree (Fan et al., 2020).

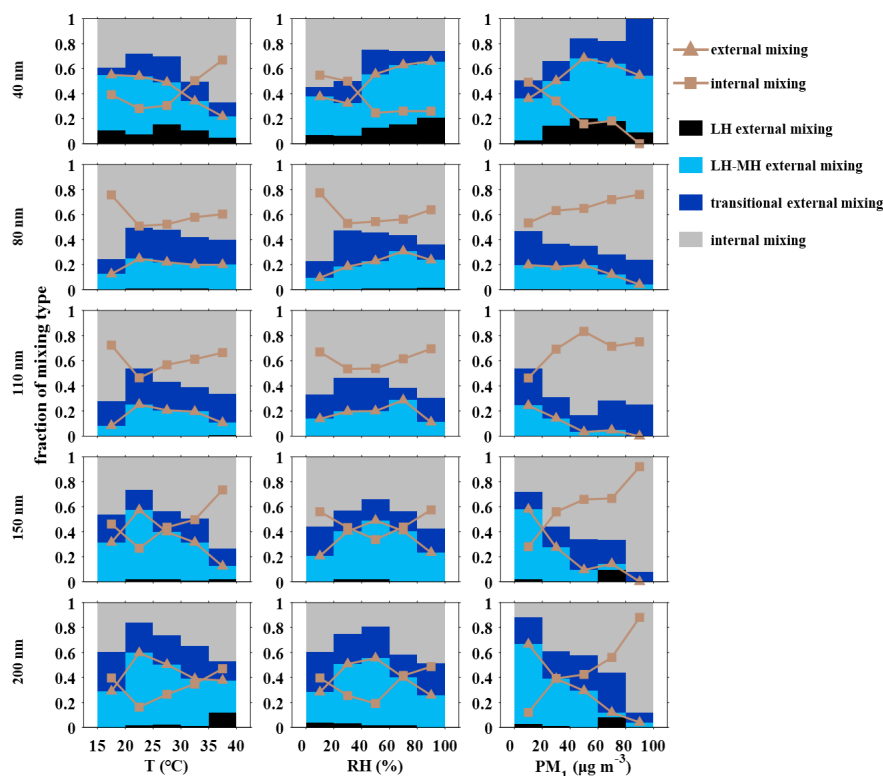
224   While for particles with sizes of 80, 110, 150 and 200 nm, the proportion of internally-mixed

225   state increase to nearly 90% under heavy PM_1 pollution, which is consistent with previous

226   observations in winter of urban Beijing (Chen et al., 2020) and in Xi'an (Wang et al., 2014).

227   Our result implies that those fine aerosol particles experience aging through both the

228   photochemical processes and aqueous growth in the polluted atmosphere of urban Beijing,

229   thereby affecting their mixing state.

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231   Figure 6. Mixing type fraction as a function of T, RH and PM_1 for five particle sizes. Solid

232   triangles and solid squares represent the fraction of internal and external (LH external mixing

233   + LH-MH external mixing state) mixing states, respectively.

3.3 Particles aging timescale: on clear and cloudy days

235   Figure 7 shows the average diurnal variations of the fraction of particles with internal

236   mixing state and the mean fractions of all the four mixing types on clear and cloudy days. On

237   clear days, the fraction of internal mixing state increases significantly from ~10-40% before

238   9:00 to nearly 100% during 12:00–15:00, especially when particle size smaller than 150 nm.

239   While, there is no such notable enhancement on cloudy days. The difference in the proportion

240   of particles with internal mixing state between the period before 9:00 and during 12:00–17:00

241   usually does not exceed 50% on cloudy days, and the maximum fractions of particles with

242   internal mixing state on cloudy days were about 50%-90%. Compared with the cloudy days,

243   the fraction of particles with internal mixing for 40 nm is noticeably higher on clear days, which

244   accounts for about 40% of all mixing states, indicating that the 40 nm particles are likely from

245   nucleation processes. On clear days, the particles undergo a gradual shift from external to

246   internal mixing states during 8:00–16:00 accompanied by a growth in particle size from 20 nm

247   to about 100 nm (Figure S3), which is generally 1–2 hours shorter than that observed on cloudy

248   days.

249   Overall, the aging process is confirmed by the transition of particle from external to

250   internal mixing states, the growth of particle size, and the increase of secondary organic aerosol

251   (Figure S3), which typically spans a duration of approximately 5 to 8 hours, as is similar to the

252   aging timescale of aerosol particles in the polluted Indo-Gangetic Plain (< 10 hours) (Ghosh et

253   al., 2021). Actually, the aerosol particles aging would be largely affected by local atmospheric

254   conditions, and thus would vary both spatially and temporally (Pöschl et al., 2001; Huang et

255   al., 2013). For example, using an environmental chamber approach, Peng et al. (2016) revealed

256   that the timescale of BC aerosols aging is 2.3 and 4.6 hours, 9 and 18 hours over two cities—

257   Beijing and Houston respectively. Note that the faster aging time in Beijing derived in the

258   chamber experiment is probably due to the different levels in the concentration of gaseous

259   precursors. In addition, there is only photochemical aging occurred in the chamber experiment,

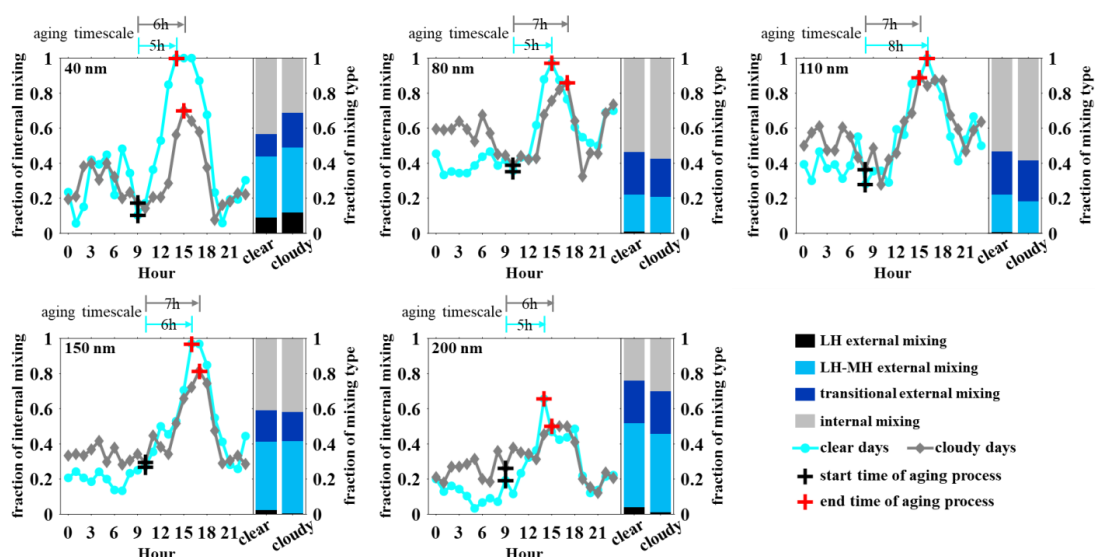
260   however, in the atmosphere, the particles aging is also through coagulation process which

261   usually occurs slower than the photochemical reaction and condensation processes (Chen et al.,

262   2017). This can also explain our observed faster aging time on clear days when the

263   photochemical process is more significant considering that the difference of the pollutant

264   concentrations in daytime between clear and cloudy days is not obvious (i.e., SO_2, NO_2, O_3,

265   PM_1; Figure S4).

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267   Figure 7. Diurnal variations of internal mixing proportion and statistical results of the fraction

268   of mixing state for five particle sizes on clear and cloudy days. The line segments and time

269   above the graph represent the aging timescale of particles. The start time and end time of

270   particle aging process was selected when the proportion of particles with internal mixing state

271   is closest after sunrise, and when the fraction of internally-mixed particles reaches its maximum,

272   represented by black and red cross respectively.

273   **3.4 Implications to parameterization of the current models**

274   The aging timescale of aerosol significantly impacts their physiochemical properties,

275   which in turn affects their atmospheric lifetime, transport characteristics (Zhang J. et al., 2023)

276   and the direct radiative forcing of aerosol particles (Moffet and Prather, 2009; Wang et al.,

277   2018). We further compared our results with the results derived from five field sites and the

278   values adopted in current models (Table S1). As shown in Figure 8, the aging timescale of

279   particles in Beijing achieved in this study is 5–8 hours, which is comparable with the

280   observations of that reported in other urban areas like Mexico City (3 hours; Moffet and Prather,

281   2009) and Los Angeles (3 hours; Krasowsky et al., 2016). The result is also close to the aging

282   timescale of particles in source area of biomass burning in California (4 hours; Akagi et al.,

283   2012). However, the aging time of particles displays large spatial variations at different sites.

284   For example, the aging timescale of particles observed in Beijing (4.6 hours) was four times

285   faster than that in Houston (18 hours) where the precursors concentrations are extremely low

286   (Peng et al., 2016). Overall, the aging time of particles obtained in ambient atmosphere is much

287   shorter than the default values adopted in most models, which is commonly with a duration of

288   1.15–2.5 days. In addition, the values among different models range greatly from 1 to 20 days

289   (Figure 8). For example, a timescale of 20 days was used to represent a slow aging process (i.e.,

290   coagulation) by Liu et al. (2011), as may be not properly applied in regions with high particle

291   number concentration where the particles coagulation is also efficient (Chen et al., 2017).

292   Although using dynamic parameterization scheme of particles aging in models could

293   achieve the application of different particles aging time to different regions, the simulated

294   values show large variations and uncertainties among different models. For example, in

295   RegCM4 model, the conversion time from fresh to aged BC range from about 5 hours to 7 days

296   (Ghosh et al., 2021). The range in KAMM/DRAIS, however, is only 2 hours to about 1.6 days

297   (Riemer et al., 2004). The aging timescales has been reported ranging from 12 hours to 7 days

298   over central-eastern China based on a regional chemical transport model by Chen et al. (2017),

299   which is much longer than that derived in urban Beijing by this study.

300   The large uncertainties in the aging timescale can significantly affect the accuracy of

301   simulation and assessment for atmospheric lifetime, loading and radiation forcing of aerosols.

302   For example, to better simulate the intercontinental transport of aerosols, Huang et al. (2013)

303   implemented a variable aging scheme in the GEOS-Chem model, which showed that the total

304   atmospheric burdens and global average lifetimes of BC (OC, organic carbon) were increase

305   by 8% (2%) compared to the default value (1.15 days). Similarly, due to the implementation of

306   the dynamic aging scheme in models, the column burden and surface mass concentration of

307   carbonaceous aerosols increased during the dry season in the polluted Indo-Gangetic Plain, and

308   the atmospheric heating increased by at least 1.2 W m^{-2} (Ghosh et al., 2021). Therefore, given

309   the large spatiotemporal variations in aging timescale of particles, the study emphasizes the

310   urgency of conducting investigations at more field sites. In addition, the other factors such as

311   the meteorology, aerosol mass loading and particle sizes that affect the aerosols aging should

312   be accounted for so as to improve the dynamic aging schemes in climate models.

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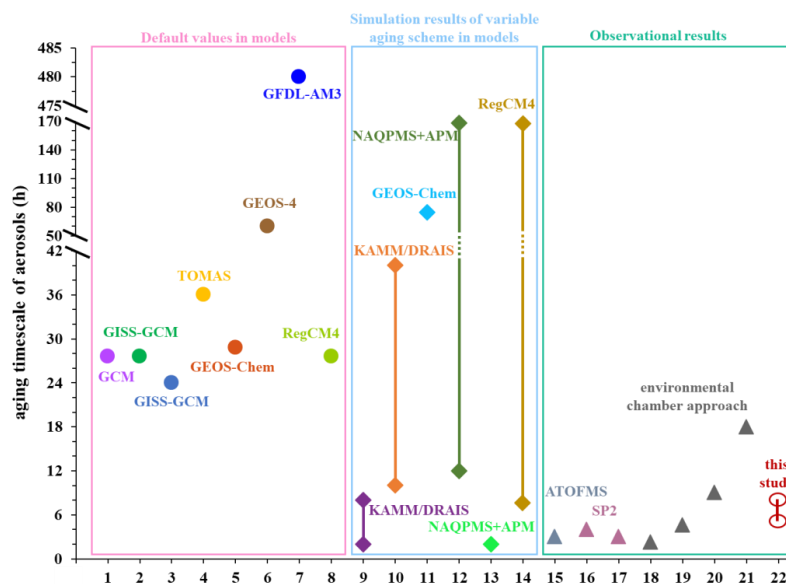
Figure 8. The aging timescale of particles reported in literatures (1. Cooke et al., 2002; 2. Chung

and Seinfeld, 2002; 3. Koch and Hansen, 2005; 4. Pierce et al., 2007; 5. Yu and Luo, 2009; 6.

Colarco et al., 2010; 7. Liu et al., 2011; 8, 14. Ghosh et al., 2021; 9, 10. Riemer et al., 2004; 11.

Huang et al., 2013; 12, 13. Chen et al., 2017; 15. Moffet and Prather, 2009; 16. Akagi et al.,

2012; 17. Krasowsky et al., 2016; 18-21. Peng et al., 2016). The solid circle, diamond and the

triangle denote the default aging time of particles used in models and stimulation results of

variable aging scheme in models, as well as the observational results, respectively.

4 Conclusions

The real-time mixing state of ambient aerosol particles with dry particle sizes of 40, 80,

110, 150 and 200 nm was investigated in urban Beijing, according to PDF of hygroscopic

growth factor measured using the H-TDMA system. Four mixing states of ambient size-

resolved particles were captured in this study. In general, particles with LH external, LH-MH

external, transitional external and internal mixing state account for 0-10%, 20-46%, 17-24%

327   and 27-56% respectively, which depends on particles size greatly. The diurnal variation of

328   mixing state of particles in all sizes considered present a visible aging process, showing that

329   the fraction of particles with internal mixing state increases significantly from ~10-40% before

330   9:00 to about 100% during 12:00–15:00 on clear days, accompanied by a growth in particle

331   size from 20 nm to 100 nm, which is more notable than that on cloudy days. Specifically, the

332   aging process of particles within approximately 5–8 hours. In addition, the mixing state of

333   particles was observed to be dependent both on T and RH, showing that the particles with

334   internal mixing state elevates obviously when T or RH exceeds 25°C or 60% respectively.

335   Moreover, with the intensification of particulate pollution, the particles become more

336   internally-mixed. The results indicate that both high ambient temperature and very humid

337   atmosphere are more favorable for particle aging in polluted atmosphere. The large difference

338   of aging timescale of particles between values in models and the timescale achieved by

339   observations, emphasizing the vital role of exploring the aging timescale through more field

340   measurements to improve the accuracy of aging schemes in climate models. The results

341   revealed in our study highlight the considerable impact of atmospheric aging on mixing state

342   of fine aerosol particles in polluted megacities.

343

344   **Data availability**

345   All data used in the study are available from the corresponding author upon request

346   (zhangfang2021@hit.edu.cn).

Author contributions

FZ and JL conceived the conceptual development of the paper. FZ directed and performed of the experiments with JR, LC and JL. JL conducted the data analysis and wrote the draft. All authors edited and commented on the various sections of the paper.

Competing interests

The authors declare that they have no conflict of interest.

Acknowledgments

This work was funded by the National Natural Science Foundation of China (NSFC) research project (Grant No. 42475112, 42405191), Shenzhen Science and Technology Plan Project (Grant No. GXWD20220811174022002; KCXST20221021111404011), Guangdong Natural Science Foundation (Grant No. 2024A1515011005), Hebei Natural Science Foundation (Grant No. D2024205006). We thank all participants in the field campaigns for their tireless work and cooperation.

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