

**QUALITATIVE TRANSFORMATIONS OF DISSOLVED ORGANIC MATTER ALONG SUPRA-PERMAFROST FLOW:
INSIGHTS FROM ARCTIC SUBTERRANEAN ESTUARIES**

Aude Flamand^{1,2,3,4} (ORCID : 0000-0001-9983-4176)

Jean-François Lapierre^{3,4} (ORCID : 0000-0001-5862-7955)

Gwénaëlle Chaillou^{1,2} (ORCID: 0000-0002-4170-8852)

¹ Institut des Sciences de la Mer de Rimouski, Université du Québec à Rimouski, 310 Allée des Ursulines,
Rimouski, Québec, Canada, G5L 3A1.

² Groupe de recherche Québec-Océan

³ Département de sciences biologiques, Université de Montréal, 1375 Avenue Thérèse-Lavoie-Roux, Montréal,
Québec, Canada, H2V 0B3.

⁴ Groupe de Recherche Interuniversitaire en Limnologie (GRIL).

Corresponding Author: Gwénaëlle Chaillou. Email: gwenaelle_chaillou@uqar.ca

13 Abstract

14 Increasing coastal erosion and permafrost thaw along the Arctic shoreline represent major lateral sources of
15 dissolved organic matter (DOM) to the Arctic coastal ocean, ~~where its behaviour depends on composition and~~
16 ~~transport pathways~~. One key and still underrecognized ~~flow path~~~~transport mechanism~~ is non-point source supra-
17 permafrost groundwater flow discharging through sandy beaches in front of coastal bluffs. Here, we explore
18 the qualitative transformations of DOM and dissolved organic carbon (DOC) along this flow path, with a focus
19 on the beach discharge zone, where fresh supra-permafrost groundwater mixes with recirculated seawater. We
20 sampled meltwater, beach groundwater, and seawater along coastal bluff transects extending up to 0.5 - 1 km
21 offshore. Along these salinity and redox gradients, DOC and chromophoric DOM (CDOM) declined sharply,
22 partially reflecting dilution. Optical indices (a_{350} , $SUVA_{254}$, HIX) and PARAFAC-based fluorescence
23 analyses revealed a shift from humic-like, high molecular weight (HMW) DOM in meltwaters to protein-like,
24 low molecular weight (LMW) DOM in nearshore waters. The inverse relationship between humic- and protein-
25 like components, combined with reactive Fe-hydroxide coatings on sandy sediments, elevated dissolved iron
26 (Fetot) and high dissolved inorganic carbon (DIC) under low oxygen conditions in beach groundwater, suggests
27 that microbial degradation and mineral-organic interactions likely contribute to DOM transformation. ~~These~~
28 ~~processes are~~ associated with a relative decrease in humic-like and HMW fractions. Alongside these
29 transformation processes, ~~in addition to mixing processes that governs~~ bulk DOC behaviour. These findings
30 highlight the role of nearshore and intertidal zones, particularly subterranean estuaries (STEs), as zones of DOM
31 transformation along the Arctic coastline. However, whether ~~these environments~~they act primarily as transient
32 filters, long-term carbon sinks, and/or dynamic biogeochemical reactors remains unresolved,
33 ~~highlighting~~highlighting the need for further research on their role in transferring permafrost-derived carbon to
34 the ocean.

35 **Keywords:** permafrost, Arctic coastal ocean, dissolved organic matter, dissolved organic carbon, Subterranean
36 estuary, DOM transformation, Iron-organic interactions, Microbial processing.

1 Introduction

Permafrost stores around 1,300 Pg of organic carbon (OC) within its 13.9×10^6 km² surface area, which represents 60 % of the world's soil organic carbon stored in 15 % of global soil area (Schuur et al. 2015; Obu et al. 2019) . The Arctic permafrost coastline is heavily impacted by global changes, resulting in unprecedented thawing rates and the deepening of the active layer, which, in turn, increases subsurface transport (Lantuit et al. 2012; Jones et al. 2020). Unlithified and ice-bonded permafrost cliffs, such as those along the Beaufort Sea, are highly susceptible to coastal erosion. Over the past twenty years, these cliffs have experienced one of the highest coastal erosion rates recorded in the Arctic, with a recorded rate of 1.1 m yr⁻¹ between 1950 and 2000. This rate has increased by 80 – 160 % in the last two decades (Lantuit et al. 2012; Jones et al. 2020). Accelerating coastline erosion is supplying increasing quantities of terrestrial materials (Kipp et al. 2018), associated nutrients (e.g., nitrogen and phosphorus; Fritz et al. 2017) carbon (including organic carbon and methane; Bristol et al. 2021), and contaminants (such as heavy metals and organic pollutants; Kwasigroch et al. 2018) to the nearshore and coastal ocean. Recent studies indicate that land-derived nutrients from coastal erosion ~~contribute—drive~~ ~30 to 50% of primary production in the Arctic Ocean, underscoring the critical biogeochemical role of permafrost-derived nutrient ~~recycling and carbon inputs~~ on Arctic shelves (Terhaar et al. 2021; Vonk et al. 2025). Beyond erosion, subsurface pathways also contribute to these fluxes, as thaw depths and coastal hydraulic gradients peak in late summer and remobilize solutes from These diffuse, non-point sources into the coastal ocean ~~release solutes that are remobilized in late summer, when thaw depths and coastal hydraulic gradients peak~~ (Walvoord & Striegl, 2007; Lecher, 2017).

Dissolved organic matter (DOM) represents a fundamental link between terrestrial and aquatic carbon cycles, playing a significant role in the biogeochemistry of aquatic ecosystems (Hedges & Keil 1995). At the ocean-basin scale, newly mobilized organic matter from Arctic watersheds may lead, among other things, to enhanced emissions of CO₂ or other greenhouse gases (Vonk et al. 2015; Kaiser et al. 2017; Tanski et al. 2019; Lapham et al. 2020; Vonk et al. 2025) and modifications in the productivity of primary producers (Thingstad et al. 2008; McMeans et al. 2015; Nguyen et al. 2022) that sustain the marine Arctic food web, with unknown impacts on

the communities that rely on these food sources. This new input affects ocean chemistry by altering carbon and nutrient cycling (Guo et al. 2007; Stedmon et al. 2011; Vonk et al. 2014) and changes optical conditions, reducing light penetration and limiting primary production (Matsuoka et al. 2012; Fichot et al. 2013). A fraction of this terrestrial DOM can be rapidly mineralized through microbial and photochemical processes, affecting nutrient budgets, air-sea CO₂ exchanges, biological productivity, as well as acidification in coastal waters (Kaiser et al. 2017a; Kaiser et al. 2017b). For example, Kaiser et al. (2017a) showed that ~50% of the annual dissolved organic carbon (DOC) discharged by Siberian rivers was mineralized in estuaries and on Eurasian continental shelves, indicating significant removal before reaching the open ocean. Therefore, only a small fraction of terrestrially-derived compounds potentially persists in the ocean over centuries and millennia (Fichot & Benner 2014; Kaiser et al. 2017a). While the export of terrestrial dissolved organic carbon is known to significantly influence the Arctic marine ecosystem, with an estimated 25 – 38 Tg C per year discharged into the Arctic Ocean (Holmes et al. 2012), the role and significance of inputs from coastal erosion and permafrost thaw in shaping the biogeochemistry and ecology of nearshore Arctic waters remain poorly understood. This knowledge gap stems not only from the stochastic and episodic nature of erosion- and thaw-related fluxes, which complicates their quantification, but also from the complex interactions between mineral and organic phases that control the stabilization, degradation, transformation, and transport of the DOM pool along flow paths. The mechanisms of organic matter transformations occurring at mineral-organic interfaces are complex (see Li et al. 2023; and references therein), particularly in intertidal sandy sediment, where biological, geochemical, and redox conditions interplay to influence DOM concentrations and molecular compositions.

In addition to dilution and bacterial degradation, a key process in these dynamic transitional environments is also the formation of iron (Fe) curtains at the [freshwater-seawater](#) redox interface, which may serve as major zones for the storage and transformation of terrigenous organic carbon (Chen et al. 2014; Linkhorst et al. 2017; Sirois et al. 2018; Zhou et al. 2023, 2024; Amoako et al. 2025).

At the shoreline, beach subterranean estuaries (STEs), where terrestrial groundwater mixes with recirculating seawater (Moore, 1999), act as biogeochemical hotspots (Anschutz et al. 2009) with the Fe cycle playing a central role in both the remineralization and sequestration of organic carbon (Charette & Sholkovitz, 2002;

Sirois et al. 2018; Chaillou et al. 2024). Redox oscillations driven by fluctuations in water levels (e.g., tide amplitude, waves, sea-level rise, and water table variations; Santos et al. 2012) shape the zones where reactive minerals such as Fe-hydroxides precipitate, forming transient or persistent geochemical barriers that influence the mobility of compoundselements, including DOM and DOC, and thereby modulate their export to adjacent coastal waters (Riedel et al. 2013; Linkhorst et al. 2017; Sirois et al. 2018; Waska et al. 2024; Amoako et al. 2025). While most studies on beach STEs and their associated biogeochemical processes have focused on lower-latitude systems, our understanding of their occurrence and functioning in Arctic environments remains scarce. The remoteness of Arctic beaches, the logistical challenges of sampling, and the high associated costs limit our understanding of these systems. However, recent research in Alaskan coastal lagoons by Demir et al. (2024) has confirmed that supra-permafrost groundwater flows, originating from the seasonally thawed active layer, can extend beyond coastal bluffs and beaches, even in areas underlain by permafrost. The presence of unfrozen, permeable coastal sediments thus creates pathways for fresh groundwater flow that mixes with recirculating seawater (as in STEs; Moore, 1999) and discharges into coastal zones, as observed in the Laptev Sea (Charkin et al. 2017) and in the Beaufort Sea along the Alaskan and Mackenzie coasts (Bullock et al. 2024; Kipp et al. 2025).

Absorbance- and fluorescence-derived indices are widely used to characterize the origin, reactivity, and transformation of aquatic DOM (Stedmon et al. 2003; Fichot & Benner, 2014; Meilleur et al. 2023). For instance, Fouché et al. (2020) identified permafrost-derived DOM as being predominantly low molecular weight (LMW), proteinaceous, and low in aromaticity, while DOM from the seasonally thawed active layer is characterized by high molecular weight (HMW), aromatic, and condensed organic compounds. In low-latitude STE, selective retention and transport of organic molecules along the flow path lead to a fractionation of DOM, thereby shaping the export of terrestrial-derived organic compounds to adjacent coastal waters (Couturier et al. 2016; Linkhorst et al. 2017; Hébert et al. 2022). -The transformation and evolution of seasonally mobilized compounds at the land-ocean interface, particularly beyond beaches and coastal bluffs and along supra-permafrost flow paths, remains poorly constrained in Arctic coastal environments. This study aims to characterize the transformation pathways of DOM released from the thawing of coastal Arctic bluffs and to

assess the role of sandy beaches as zones of terrigenous DOM transport and transformation ~~of terrigenous DOM at the shoreline~~, with a focus on compositional changes rather than flux quantification. ~~We investigate whether Arctic beaches act as zones of transformation that may contribute to changes in the composition and reactivity of permafrost-derived DOM before its dilution and export into the coastal ocean, within a system where mixing with recirculating seawater also plays a major role in controlling bulk concentrations.~~ This study focuses on compositional changes and transformation patterns rather than on quantifying fluxes or closing carbon budgets. ~~Similar processes have been documented in temperate and tropical STEs, although their relevance in Arctic permafrost settings remains poorly constrained.~~ A fine-scale sampling strategy (<1 km) was deployed in the Kugmallit Bay region (NWT, Canada) to monitor the optical characteristics of DOM, including chromophoric dissolved organic matter (CDOM) and fluorescent dissolved organic matter (FDOM), as well as the behavior of DOC along the supra-permafrost flow, from coastal bluffs to the adjacent nearshore waters.

2 Materials and methods

2.1 Site Description

The study area is situated in the Inuvialuit Settlement Region of the Northwest Territories, near the Mackenzie Delta, the fourth-largest river system discharging into the Arctic Ocean (Macdonald et al. 1998). The Mackenzie River is a major source of freshwater, sediments, dissolved organic matter (DOM), and nutrients, exerting a strong influence on the salinity and biogeochemistry of the Beaufort Sea (Holmes et al. 2012). The tidal amplitude in Kugmallit Bay region is small (~0.3 m). Two field campaigns were conducted: the first from July 24 to August 6, 2019, and the second from June 22 to August 31, 2021, coinciding with peak thaw conditions. Approximately 60 samples were collected from four coastal sites characterized by continuous permafrost and active thaw slumps, bordered by sandy and clay-rich beaches: Tuktoyaktuk Island, Peninsula Point, Crumbling Point, and Reindeer Island (Fig. 1).

Tuktoyaktuk Island, the primary sampling site (N = 34; Fig. 2A), features a coastal bluff roughly 9 m high, 1.5 km long, and 100 m wide (Ouellette, 2021). Located southeast of Kugmallit Bay and opposite the Hamlet of

Tuktoyaktuk, the island is undergoing rapid coastal retreat, losing approximately 1.8 m of shoreline annually due to storm events and permafrost thaw. This erosion rate has increased by 22% over the past 15 years (Berry et al. 2021; Whalen et al. 2022; Tanguy et al. 2023), and the entire site is projected to vanish within the next 20–30 years (Jones et al. 2020). In front of the bluff lies a 50-m-wide beach composed of fine to medium sands (0.3–0.5 m deep) overlying a frozen clay horizon (Fig. 2A). The exact submarine extension of this permafrost layer is unknown. Peninsula Point (N=7), southwest of Tuktoyaktuk Island within the Pingo Canadian Landmark, features a complex retrogressive thaw slump system underlain by a massive ice body ranging from 5 to 20 m in thickness (Mackay, 1986). Large, muddy lobes of thawed permafrost material and meltwater flow downslope toward the coast, intermittently covering the sandy intertidal zone (Fig. 2B). The shoreline has retreated at an average rate of ~ 3.4 m yr⁻¹ between 1985 and 2018 (Hayes et al. 2022). Crumbling Point (N=5), located at the northwest edge of Kugmallit Bay, is another retrogressive thaw slump system with geomorphological characteristics similar to Peninsula Point. Reindeer Island (N=11), north of Richards Island, is part of a lagoonal complex formed by thermokarst lakes and surrounded by coastal bluffs. Although specific erosion rates for this site are not yet available, they are likely comparable to regional averages for the Canadian Beaufort Sea (~ 0.5 m yr⁻¹; Solomon, 2005). However, localized retreat rates can be substantially higher as seen in nearby Pullen Island, where erosion exceeds 12 m yr⁻¹ (Berry et al. 2021). The width of the intertidal zone along these beach systems varies from a few meters to several tens of meters, depending on oceanographic conditions (e.g., wind regime and seasonal beach morphology).

2.2 Water and Sediment Sampling

At each site, we carried out localized sampling along a transect, from coastal thawed bluffs, through beaches and intertidal zones, to offshore seawater as presented in Kipp et al. (2025). Meltwater and beach groundwater (here defined as porewater in sandy sediment) samples were collected from coastal permafrost slumps and the adjacent sandy shore, respectively, while seawater samples were collected 0.5 to 1.0 km offshore from the coastline at each study site. Surficial sandy sediments from beaches in front of permafrost bluffs were also collected during the 2019 field campaign at Tuktoyaktuk Island (N = 3) and Peninsula Point (N = 2). Meltwater was sampled directly in puddles formed on the thaw slumps using a submersible pump. Beach groundwater was

collected using push-point piezometers inserted ~50 cm deep into the sandy ground, above the frozen clay layer, within the intertidal zone, with continuous pumping via a Solinst® peristaltic pump. A massive ice sample was collected from a permafrost core collected at Richards Island, which was sectioned and thawed in a hermetically sealed, acid-cleaned bucket, before collecting the meltwater via a peristaltic pump. Finally, seawater samples were collected from a small vessel in front of the slump systems positioned ~0.5–1.0 km offshore ~~in front of the slump systems~~ using a submersible pump placed 0.5–1.0 m below the surface. At each site, water samples were pumped into an online flow cell where salinity (S), temperature, pH (at NBS scale), and oxygen saturation were monitored with a daily calibrated multiparametric probe (600QS, YSI Inc.). Note that these stations were likely under the influence of the Mackenzie River plume, which mainly controls the water chemistry and optical signature in the region (see Lizotte et al. 2022, and references therein).

Water samples were collected for CDOM/FDOM into acid-washed 60 mL glass amber bottles after on-line filtration through a 0.22 µm Millipore Opticap® XL4 cartridge with a Durapore® membrane. The samples were stored in the dark at 4 °C. Total dissolved Fe (Fetot) samples were collected in 60 mL metal-free Falcon® tubes after filtration through the same 0.22 µm Millipore Opticap cartridge. The samples were acidified with 3 drops of 70 % nitric acid to prevent the re-oxidation of reduced trace metals and stored at 4 °C. DOC samples were taken using 60 mL acid-cleaned polypropylene syringes and rapidly filtered with pre-combusted (450 °C for 5–6 hours) 25mm Whatman GF/F (0.7 µm) and stored in pre-combusted and acid-washed 12 mL borosilicate EPA tubes with PTFE caps. The same day, the DOC samples were acidified to pH <2 with high-purity HCl 2N and stored in the dark at 4 °C until analysis. During the 2019 campaign, samples were also collected in 30 ml scintillation vials, hermetically sealed for further water stable isotope analysis.

2.3 Chemical and Optical Analysis

Total organic carbon (TOC) content was measured on freeze-dried, ground and homogenized sediment using a CHNSO analyzer (Costech ECS 4010 CHNSO analyzer®; precision < 0.3%). Reactive Fe-hydroxide content was determined on the same sediment samples, extracted using the citrate-dithionite-bicarbonate (CDB) reduction method of Mehra and Jackson (1960), as modified by Lalonde et al. (2012), and previously applied

to sandy sediments by Sirois et al. (2018). Briefly, reactive Fe-hydroxides are reductively dissolved with dithionite at circumneutral pH (bicarbonate buffer) using citrate as a complexing agent to keep dissolved Fe in solution. Fe in solution was analyzed using a Microwave Plasma Atomic Emission Spectrophotometer (4200 MP-AES; Agilent Technologies). The detection limit of the method is $2.4 \mu\text{g L}^{-1}$ for Fe concentration at a wavelength of 391 nm and analytical precision was $<5\%$.

Stable isotopes of water ($\delta^{18}\text{O}$, $\delta^2\text{H}$) were analyzed by EA-IRMS during the following year after the collection. Uncertainties are $\pm 0.05 \text{ ‰}$ and $\pm 1 \text{ ‰}$ for $\delta^{18}\text{O}$ and $\delta^2\text{H}$, respectively. Reference materials were used throughout the isotopic water analyses, and isotopic analyses are reported compared to the international Vienna Standard Mean Ocean Water (VSMOW). DOC samples were analyzed a few weeks after data collection by Total Organic Carbon analyzer (TOC-VCPN Shimadzu) based on the method of Wurl and Tsai (2009). Trueness was verified against certified reference solutions (EMSL secondary standard and DSR Batch 21 Lot 04-21, University of Miami), with analytical uncertainty less than 4 ‰ and a detection limit of $5.8 \mu\text{mol L}^{-1}$. To ensure instrument stability, fresh acidified deionized water (blank) was regularly analyzed. The concentration of total dissolved iron (Fetot) was measured according to the ferrozine method proposed by Stookey (1970) and adapted by Viollier et al. (2000). The detection limit of the method was $0.4 \mu\text{mol L}^{-1}$ and the reproducibility was better than 0.3 ‰ .

Absorbance and fluorescence spectroscopy were used for the measurement of the chromophoric and fluorescent fractions of DOM (CDOM and FDOM) a few weeks after sampling. The CDOM absorbance was measured using a Lambda 850 UV-VIS Perkin Elmer spectrophotometer with 1 cm path-length quartz cuvettes. Measurements were taken from 220 to 800 nm at 1 nm intervals with a scanning speed of 100 nm min^{-1} and a 5 nm slit width. Blanks and references were measured using fresh Milli-Q water. The FDOM was measured concomitantly using a Varian Cary Eclipse spectrofluorometer. Fluorescence spectra were measured within the emission wavelengths of 220 to 600 nm and the excitation wavelengths of 220 to 450 nm at 5 nm intervals as described by Couturier et al. (2016). Similarly, fresh Milli-Q water was used as a blank to rinse the cuvette in between samples. Fresh deionized water was used as a blank, and absorbance measurements were used to

correct the inner-filter effect. Fluorescence spectra were further corrected for instrument-specific excitation and emission bias using manufacturer-provided correction files, and the dataset was corrected for Rayleigh and Raman scattering, according to the method described by Pucher et al. (2019).

2.4 Optical-derived indices and PARAFAC model

Absorbance and fluorescence indices were extracted using the staRdom toolbox on the R Studio Software (Pucher et al. 2019). Different indices were explored, including a₃₅₀, BIX, HIX, FI, SUVA₂₅₄ and SR. However, only three were reported here to characterize the DOM pool, as they showed distinct variations among the different categories of samples (i.e. beach groundwater, meltwater, massive ice, and seawater). The spectral absorption coefficient at 350 nm (a₃₅₀) was used as a proxy for CDOM concentration in the samples (Helms et al. 2008). It was calculated as 2.303 times the absorbance at the wavelength $\lambda=350$ nm divided by the path length of the cuvette (m). The specific UV absorbance (SUVA₂₅₄ in L mgC m⁻¹) was calculated as the absorbance at the wavelength $\lambda=254$ nm normalized to DOC concentration and is commonly used as an indicator of DOM aromaticity (Weishaar et al. 2003; Helms et al. 2008): greater SUVA₂₅₄ values correspond to a greater degree of aromaticity (Helms et al. 2008). It has also been shown to be positively correlated with the molecular weight of the DOM compounds. In the FDOM pool, the humification index (HIX) corresponds to the peak area under emission of 435–480 nm divided by the peak area under emission of 300–345 nm, at an excitation of 254 nm. HIX is an indicator of humic substances and the extent of humification of DOM compounds (Ohno, 2002; Hansen et al. 2016): higher HIX indicates a greater humification of the DOM source and HMW compounds.

In combination with absorbance and fluorescence indices, a PARAFAC model based on over 250 Arctic and subarctic coastal water samples was developed to further investigate the composition and sources of FDOM across samples (Bro 1997; Stedmon et al. 2003; Murphy et al. 2013). Five components were validated using split-half analysis, following the method adapted from Pucher et al. (2019) in R, with the model explaining more than 94% of the total variance in the dataset. The components were also matched with the literature for identification and external validation, using OpenFluor (Murphy et al. 2014). The five fluorescent components

(C1–C5) identified are presented in Fig. 3, and their characteristics based on the literature are summarized in Table 1. Briefly, components C1 and C4 are associated with terrestrial humic-like DOM, with C1 representing the dominant terrestrial humic signal, while C4 contributes a weaker humic fraction. Component C5 corresponds to a minor microbial humic-like signal. In contrast, components C2 and C3 correspond to protein-like DOM, with C2 representing the dominant tyrosine-like signal and C3 a weaker tryptophan-like component. Pearson correlation coefficients were calculated between PARAFAC components, expressed as absolute fluorescence intensities (RU). Spearman rank correlations were used when relating PARAFAC components to ratio-based, and between components and optical indices (HIX). Correlations were considered significant at $p < 0.001$. Differences in geochemical parameters and PARAFAC component intensities across sample types were assessed using Kruskal-Wallis tests, followed by pairwise Wilcoxon tests with Bonferroni correction for post-hoc comparisons.

3 Results and discussion

3.1 Physico-chemical characteristics along the permafrost-nearshore continuum

In this study, we refer to samples collected at the nearshore as “seawater”; nevertheless, we acknowledge that these samples more accurately represent a brackish environment, with salinity S values ranging from 0.5 to 20, mostly related to the Mackenzie River plume occurrence. The massive ice and meltwater samples exhibited the lowest salinities with $S < 1$, whereas the salinity of beach groundwater samples ranged between 0.2 and 5.4. The S range measured in beach groundwater samples mostly reflected the tidal and wave pumping effect and the recirculation of the infiltrated seawater within the permeable sediments (Santos et al. 2012). The higher salinities ($S > 5.4$ ~~$S < 20.0$~~) were only measured in seawater samples. The temperature varied between 8.1 and 16.7 °C (with a mean value of 13.0 ± 2.6 °C), with the higher temperatures measured in meltwater and some beach groundwater samples. Oxygen saturations ranged from 6 to 141 % of saturation. The nearshore surface seawater and the meltwater samples were all oversaturated due to mixing and photosynthetic processes. However, the low-salinity beach groundwater samples exhibited low oxygen saturation (6 - 48 %) despite the

infiltration and recirculation of well-oxygenated seawater. Redox oscillations and transient oxygen-depleted conditions are observed in micro- and meso-tidal sandy STEs (Sirois et al. 2018; Waska et al. 2021; Hébert et al. 2022), where oxygen supplied through tidal exchange is rapidly consumed by heterotrophic microbial processes (Chaillou et al. 2018, 2024; Moore et al. 2024). The total dissolved iron (Fetot) concentrations ranged from 0 $\mu\text{mol L}^{-1}$ (or below detection limits) in well-oxygenated seawater samples to more than 1400 $\mu\text{mol L}^{-1}$ in a deoxygenated meltwater sample. In beach groundwater samples, most Fetot concentrations were on the order of a few $\mu\text{mol L}^{-1}$, but one sample reached values above 600 $\mu\text{mol L}^{-1}$, corresponding to lower oxygen saturation levels. Regarding pH (NBS scale), the highest median values were observed in both meltwater and seawater (7.9), while beach groundwater exhibited a slightly lower median pH of 7.5, ranging from 7.0 to 8.0, with one sample dropping to 6.5 (also associated to the highest Fetot concentrations of $\sim 600 \mu\text{mol L}^{-1}$). Although pH data for massive ice samples were not available, their values likely fall within the lowest range of measured pH. This assumption is based on the origin of the massive ice in the region, which is thought to result from intrasedimental processes (e.g., ice intrusion and segregation) rather than from relict glacial ice (Mackay and Dallimore, 1992). Surficial beach sediments (<50 cm below the beach floor) are relatively poor in particulate organic carbon (TOC < 0.45%) and mainly composed of clastic sediment coated with reactive iron-oxide (Table 2).

3.2 *Origin of the subsurface water flow in the intertidal zone*

Little is known about the hydro(geo)logical connectivity between supra-permafrost groundwater and Arctic coastal seawater (Letcher et al. 2017). However, recent studies and conceptual models, mostly developed for lower-latitude regions, suggest that key ~~flow path transport mechanisms~~ of water and permafrost-derived solutes to the coastal zone are likely governed by three main pathways (Walvoord & Kurylyk, 2016): (i) riverine transport, which includes both active layer runoff and contributions from deeper groundwater (Dittmar & Kattner, 2003; Demir et al. 2024); (ii) coastal erosion, which directly delivers organic-rich permafrost soils and particulate matter into the marine environment (Kipp et al. 2018); and (iii) submarine groundwater discharge (SGD) through subterranean estuaries (STEs), where supra-permafrost groundwater mixes with recirculating seawater before entering nearshore waters (Demir et al. 2024; Kipp et al. 2025). The $\delta^{18}\text{O}$ and $\delta^2\text{H}$ results

support the theoretical model in which beach groundwater represents a mixture of supra-permafrost groundwater flowing seaward and recirculating seawater advected landward (Fig. 4). The $\delta^{18}\text{O}$ and $\delta^2\text{H}$ values measured in water samples collected in 2019, mostly at Tuktoyaktuk Island, Peninsula Point sites, and Crumbling Point (Fig. 1), ranged from -28 to -10 ‰ and from -215 to -82 ‰, respectively, the massive ice sample (N=1) presents the most depleted signature (Fig. 4). These depleted values are largely explained by low air temperatures and are typical of permafrost hydrology reported in the western Arctic (Fritz et al. 2011; Utting et al. 2012). The samples are well aligned along the local meteoric water line (LMWL; $\delta^2\text{H} = 7.39 \times \delta^{18}\text{O} - 6.70$; Fritz et al. 2022), regardless of their salinity values, except for the three (3) meltwater samples that are slightly below it, likely due to evaporation processes at the surface. The alignment of the different samples along the LMWL regression line suggests a common meteoric origin, probably from the permafrost watershed. The supra-permafrost groundwater flow that transits through the beach sediment does not seem to be affected by surficial processes (e.g. evaporation process), as observed in a few meltwater samples, likely limiting photochemical degradation of the flowing-concomitantly transported CDOM/FDOM. However, the high range of isotopic values in beach groundwater samples, from -25 to -12 ‰ and -192 to -112‰ for $\delta^{18}\text{O}$ and $\delta^2\text{H}$, respectively, suggested a mixing between depleted thawed waters (as surficial meltwater and supra-permafrost groundwater) and enriched seawater, as expected in a STE system. These results are consistent with the recent study by Kipp et al. (2025), which reported elevated activities of radon isotopes (^{222}Rn) in the same beach groundwater samples. As a noble gas, radon escapes rapidly upon exposure to the atmosphere, indicating a groundwater origin and minimal atmospheric interaction. These authors proposed that thaw water derived from massive ground ice is transported into the supra-permafrost groundwater system, where it is intermittently diluted by radon-poor recirculating seawater (e.g. offshore water from the Mackenzie River) percolating through the beach sediments, depending on tide conditions. Using stable isotope signatures, a first-order estimate of the relative contribution of fresh groundwater to beach groundwater samples can be derived using a simplified two-endmember mixing approach. However, this estimation remains uncertain due to the variability in the isotopic composition of the nearshore seawater endmember, which reflects a wide range of salinities ($5 < S < 20$), as well as the potential mixing of multiple freshwater sources (local and regional supra-

permafrost groundwater, and surface meltwater runoff, see Fig. 2) within the supra-permafrost system. To constrain the endmembers, we used the mean isotopic composition of the most saline sample to represent seawater ($-74.5\text{‰ } \delta^2\text{H}$; $-9.1\text{‰ } \delta^{18}\text{O}$; $S \sim 20$), and the massive ice sample to represent the fresh supra-permafrost water endmember ($-224\text{‰ } \delta^2\text{H}$; $-32\text{‰ } \delta^{18}\text{O}$). Based on these assumptions, the contribution of fresh supra-permafrost water ranges from approximately 20 to 76 %. These values should be interpreted as indicative ranges rather than precise quantifications, given the simplified assumptions and the dynamic mixing conditions within the system. Although these estimates do not constrain the total SGD flux through the beach system, they indicate that freshwater inputs can represent a substantial fraction of beach groundwater, consistent with previous SGD studies in the Beaufort Sea region (e.g., Connolly et al., 2020; Demir et al., 2024; Bullock et al., 2024). ~~These values should be interpreted as indicative ranges rather than precise quantifications, given the simplified assumptions and the dynamic mixing conditions within the system. Although these estimates do not constrain the total water flux through the beach system, they indicate that freshwater inputs can represent a substantial fraction of beach groundwater. Only a limited number of studies have quantified submarine groundwater discharge (SGD) along Arctic coastlines. Available estimates suggest that fresh supra-permafrost groundwater typically accounts for only 3–7% of total SGD, with freshwater fluxes of approximately $400\text{--}2,100\text{ m}^3\text{ km}^{-1}\text{ d}^{-1}$ reported in an Alaskan Beaufort Sea coast lagoon (Connolly et al., 2020; Demir et al., 2024). Our results indicate that supra-permafrost groundwater mixes with recirculated seawater in a nearshore zone, highlighting the dynamic nature of this system along the shoreline, where stable isotopes, salinity, oxygen saturation, and pH likely vary in response to seawater infiltration and dilution. Overall, our results support the presence of a dynamic mixing zone along the beach shoreline, where stable isotopes, salinity, oxygen saturation, and pH respond to the infiltration and dilution of offshore seawater.~~

3.3 *Behaviour of the DOC and DOM pool*

The DOC concentrations dropped from $2,360\text{ }\mu\text{mol L}^{-1}$ in meltwater samples to $236\text{ }\mu\text{mol L}^{-1}$ in the saltiest seawater sample ($S = 20$; Fig. 5A), with concentrations declining sharply along the supra-permafrost flow path,

reaching values below $400 \mu\text{mol L}^{-1}$ in beach groundwater samples. These differences across sample types were statistically significant (Kruskal-Wallis, $p < 0.001$). DOC concentrations in seawater samples, and to a lesser extent in beach groundwater, closely followed the theoretical mixing line connecting the freshest and saltiest endmembers. This mixing line, based on the most enriched and depleted stable isotope signatures in the dataset, corresponds to salinity values of 20.08 (N = 1 seawater sample) and 0.18-0.30 (N = 8 beach groundwater samples), respectively. This pattern suggests that dilution becomes the dominant process controlling DOC behaviour once supra-permafrost flow enters the STE and is exported to adjacent coastal seawaters. ~~This is consistent with the small volumetric contribution of fresh supra-permafrost groundwater relative to recirculating seawater in along the Arctic coastline (Section 3.2).~~ Dilution of the high-concentration, low-volume fresh groundwater endmember therefore contributes to the observed decline in bulk DOC concentrations, ~~but does not fully account for the magnitude of the changes of the other parameters, suggesting that additional processes are involved.~~ Absorption coefficients at 350 nm (a_{350}) ranged from 2.0 to a maximum of 134.0 m^{-1} , the latter observed in a meltwater sample (Fig. 5B). Like As DOC concentrations, the a_{350} values declined sharply along the flow path, with a median value of 24.0 m^{-1} in meltwater samples and median values of 8.4 and 7.8 m^{-1} in beach groundwater and seawater, respectively. However, unlike DOC, a_{350} values in beach groundwaters and seawater deviated from the theoretical mixing line, indicating that they did not follow simple dilution between the freshest and saltiest endmembers. The variability of the endmembers, particularly in seawater, limits the robustness of a single mixing model and contributes to the observed dispersion, pointing to processes in addition to dilution. This interpretation is further supported by the decoupling between DOC and a_{350} (Fig. 6), as discussed below. ~~However, unlike DOC, a_{350} values in beach groundwaters and seawater deviated from the theoretical mixing line, indicating that they did not follow a simple dilution between the freshest and saltiest endmembers.~~ The HIX indices showed a wide range of values across all sample types, with median values decreasing along the continuum from 3.5 in meltwater samples to 1.4 in beach groundwater and 0.8 in seawater (Fig. 5C). Linear mixing models do not apply to ratio-based indices, and non-linear mixing curves could not be reconstructed, as the underlying fluorescence components used to calculate HIX were not retained during data processing. The decrease in HIX along the salinity gradient is consistent with a shift toward less humified DOM.

However, as for a350, the variability of HIX values, particularly in seawater samples, limits the strength of this interpretation. ~~The consistent decline in HIX along the salinity gradient indicates a progressive shift toward less humified, lower molecular weight DOM from meltwater to seawater. Similar to a350, HIX values in beach groundwater and seawater did not follow conservative mixing behaviour.~~ In contrast to other parameters, SUVA₂₅₄ increased slightly along the flow path, from 2.4 mg C L⁻¹ in the meltwater to 2.9 in beach groundwater and 3.4 in seawater (Fig. 5D), and generally followed a dilution pattern along the salinity gradient. However, the variability observed in seawater samples limits the strength of this interpretation.

DOC concentration and a350 are commonly used as proxies to assess both the quantity and optical properties of DOM. The relationship between these two parameters provides insights into the biogeochemical sources and transformation processes shaping the DOM pool composition along environmental gradients (Stedmon et al. 2003; Fichot and Benner, 2011; Spencer et al. 2013; Massicotte et al. 2017). A strong correlation between DOC and CDOM typically reflects a stable DOM pool, a linear relation frequently observed in freshwater systems (Frenette et al. 2012; Massicotte et al. 2017 and references therein). However, deviations from this relationship, referred to as DOC-CDOM decoupling, suggest shifts in DOM composition and point to selective processing of specific DOM fractions. Microbial degradation, photo-oxidation and organo-mineral interactions mechanisms selectively degrade specific organic compounds, and they lead to shifts in CDOM composition or a loss of CDOM along the land-ocean continuum (Nelson et al. 1998; Del Vecchio & Blough, 2004; Weyhenmeyer et al. 2012; Nelson & Siegel, 2013). In subarctic siliciclastic STE, Couturier et al. (2016), for example, reported a strong decoupling between CDOM-DOC, where HMW CDOM compounds were largely retained within the STE sediments, limiting their export to nearshore adjacent seawater. In such environments, the Fe cycling at the oxic-anoxic interface acts as a geochemical barrier, promoting the sequestration of terrestrial organic compounds (Sirois et al. 2018). As in soils and cohesive marine sediments, molecular fractionation of DOM within STEs tends to promote the *in situ* degradation and/or preferential export of non-Fe-stabilized, ~~marine~~-protein-like derived DOM to coastal waters (Linkhorst et al. 2017; Sirois et al. 2018; Waska et al. 2022; Zhou et al. 2023).

The strong decoupling between DOC and a350, particularly in seawater, massive ice, and beach groundwater samples (Fig. 6), indicates that these ~~variables-compounds~~ follow different trajectories along the supra-permafrost flow path (~~Fig. 5A and 5B~~), leading to a decrease in humification and a shift toward lower molecular weight material as revealed by HIX indices (~~Fig. 5A to 5C~~). Given the light-limited conditions in beach groundwater environments, microbial degradation and mineral-organic interactions likely control the fate and composition of the DOM pool. Despite this general decline in humification and molecular weight of the DOM pool, the impact on DOM aromaticity (as indicated by SUVA₂₅₄) remains relatively weak. SUVA₂₅₄ values tend to increase slightly along the flow path, likely reflecting the dilution effect with the offshore seawater (Fig. 5D). Together, these patterns indicate that DOM composition evolves along the supra-permafrost flow path, resulting in a more decomposed and refractory DOM pool in nearshore waters. This interpretation should be considered in light of the high variability observed in seawater samples, which span a broad range of DOM compositions and reflect the complexity of the system, likely influenced by extensive terrestrial input and degradation during the transport in the Mackenzie River watershed and plume (McKnight et al. 2001; Huguët et al. 2009).

3.3.3.4 ~~the decoupling between Drivers of DOC and~~ DOM decoupling along the supra-permafrost flow path

In addition to the transformations inferred from optical properties, other sources and processes may contribute to DOM composition along the flow path. The *in situ* production of a new DOM pool along the supra-permafrost flow path, ~~characterized by more degraded components with lower molecular weights and higher aromaticity, along the supra-permafrost flow path~~ cannot be ruled out despite the low TOC content in the surficial sandy sediment (TOC < 0.50% dry sediment; Table 2). Similar patterns have been observed in transgressive subarctic STEs (Hebert et al. 2022) and in areas with buried peat lenses (Waska et al. 2021), where ancient sedimentary organic matter preserved in paleosols acts as a hotspot for terrestrial DOM production along the groundwater flow path. Such local inputs-production can significantly influence the chemical and optical properties of the beach groundwater DOM pool. Moreover, the presence of permafrost extending below the STE suggests that additional particulate organic matter reservoirs may be mobilized under thawing conditions along the supra-permafrost flow path, ~~further contributing to newly the mobilization, ed DOM transformation and export of~~

DOM. These deeper *in situ* inputs may explain the very depleted stable water isotopic signature of a few beach groundwater samples (to 76% of supra-permafrost groundwater contribution) and the high radon activities (Kipp et al., 2025). Further efforts are needed to improve our understanding of Arctic beach hydrogeology by using geochemical and molecular proxies.

Despite their low TOC content, sandy sediments can act as hotspots of redox activity (Anschutz et al. 2009). In beach groundwater, the absence of light combined with fluctuations in pH, salinity, and redox conditions creates a dynamic environment conducive to a complex network of redox reactions, including the remineralization of organic matter and the cycling of iron between its oxidized (Fe(III)) and reduced (Fe(II)) states. Field observations support the occurrence of these reactions. The presence of reactive Fe-hydroxide coatings on sandy sediment surfaces (Table 2), along with intermittently high total dissolved iron concentrations (median F_{TOT} of $1.6 \mu\text{mol L}^{-1}$, reaching $671.0 \mu\text{mol L}^{-1}$; Fig. 7A) and high Fe:DOC ratio observed in beach groundwater (median of 0.00844, ranging from 0.00353 to 0.92404 in more acidic samples; Fig. 7B), together suggest the potential for the formation of Fe-OM associations through oxidative precipitation of reactive Fe-hydroxide. However, at the Fe:DOC molar ratios observed here, Fe-DOM interactions are expected to drive selective molecular fractionation of the DOM pool rather than substantial bulk DOC removal, preferentially scavenging aromatic, high-molecular-weight humic compounds from solution while leaving bulk DOC concentrations largely unaffected (Linkhorst et al. 2017; Riedel et al. 2013). This is consistent with the ~~deviation~~ sharper decline of a350 relative to DOC along the salinity gradient and HIX below the theoretical mixing line, while bulk DOC follows more closely ~~follows~~ conservative mixing behaviour. The concurrent decrease in HIX further supports a progressive loss of humic, high molecular weight compounds along the flow path, independent of simple dilution. The chemical nature of the DOM pool plays a critical role in these interactions: higher molecular weight, aromatic, oxygen-rich terrigenous DOM tends to bind more strongly to Fe(III) than aliphatic, marine-derived DOM (Linkhorst et al. 2017), and slightly acidic conditions further promote Fe(III)-DOM coagulation (Nierop et al. 2002; Amoako et al. 2025). Further evidence consistent with active microbial remineralization within the STE is provided by DIC concentrations measured in beach groundwater in 2019 at the same sites, which exceeded $3,000 \mu\text{mol L}^{-1}$ (Lizotte et al., 2022), together with a low O_2 saturation. Further evidence of

~~active microbial remineralization within the STE comes from~~ Concurrently, DIC concentrations measured in
~~samples collected in 2019 at the same sites showed that beach groundwater in 2019 at the same sites, which~~
~~exceeded 3,000 $\mu\text{mol L}^{-1}$ (Lizotte et al., 2022), alongside a mean and a mean O_2 saturation of 34%, both~~
~~substantially elevated and depleted, respectively, relative to above seawater values at comparable salinities,~~
~~consistent with active microbial remineralization of organic carbon within the STE.~~ The relative influence of
the different mechanisms controlling the molecular fractionation of the DOM pool in supra-permafrost
groundwater remains to be determined, and further studies are required to explore the role of mineral-organic
interactions and Fe-OM associations in regulating the export of thaw-derived material.

While the present dataset does not allow a quantitative separation between mixing-driven dilution and *in-situ*
transformation processes, the convergence of multiple independent lines of evidence discussed above, suboxic
conditions, elevated DIC, Fe-OM associations, and the deviation of a_{350} and HIX from the theoretical mixing
line, collectively points to significant *in-situ* biogeochemical transformation of the DOM pool within the beach
STE, operating alongside the mixing dynamics that govern bulk DOC behaviour.

3.43.5 **PARAFAC components in the DOM pool**

Among the five fluorescent components identified by the PARAFAC model, the protein-like (tyrosine-like)
component C2 dominated the FDOM pool across all sample types and salinities (Table 1; Fig. 8B). While no
~~strictly linear~~ trend was observed along the salinity gradient, clear shifts in DOM composition were evident
along the continuum. Component C2 was strongly negatively correlated with both humic-like components C1
and C4 ($r = -0.87$ and $r = -0.77$, respectively; $p < 0.001$), as well as with HIX ($\rho_F = -0.9082$; $p < 0.001$). In
contrast, humic-like components were highly positively correlated with each other (C1 vs C4: $r = 0.90$, $p <$
 0.001) and with HIX ($\rho_F = 0.9682$ and $\rho_F = 0.9384$ for C1 and C4, respectively; $p < 0.001$).

Consistent with optical indices (HIX, a_{350}) and DOC patterns previously described, humic-like components
(C1 and C4) ~~declined-decreased~~ along the continuum, with the lowest median values ~~found-observed~~ in seawater
(Fig. 8A, D), while the absolute intensity of protein-like DOM (C2) increased significantly across sample types
(Kruskal-Wallis, $p = 0.017$), with values significantly higher in seawater than in meltwater (Wilcoxon, $p =$

0.013). Massive ice and meltwater samples were predominantly composed of humic-like, HMW, and terrestrially derived DOM (C1 and C4), consistent with active layer-derived FDOM described by Fouché et al. (2020). The tryptophan-like component C3 and microbial humic-like component C5 showed comparatively low and stable intensities across sample types and salinity gradients, consistent with their minor contributions to the overall DOM fluorescence signal ~~as described above~~.

As DOM is transported along the supra-permafrost flow path, the ~~terrestrials~~ humic-like signal (C1 and C4) decreases, coinciding with an increase in protein-like, LMW DOM that is consistent with microbial processing. This shift is coherent with the trends observed in optical indices (decreasing HIX and variable a350), and supports the decoupling between DOC and CDOM, indicating that bulk DOC concentrations and DOM composition are influenced by distinct processes. Together, these results observations suggest that meltwater and massive ice act as precursors of supra-permafrost solutes exported toward the coastal ocean through Arctic sandy beaches. Along this flow path, the decline in humic-like DOM pool is accompanied by an increase in biologically derived, tyrosine-like FDOM (component C2), typically associated with microbial production (Table 1 and references therein). This ~~shift mirrors patterns is consistent with~~ transformations reported by Fouché et al. (2020), where microbial processing converts DOM into LMW, protein-like material.

In this context, the presence of non-Fe-stabilized DOM and redox ~~active conditions oscillations~~ likely ~~favour~~ promotes microbial reprocessing and transformation. ~~These processes are~~ consistent with the production formation of LMW and protein-like compounds and contribute to the observed compositional changes along the flow path that are subsequently exported and diluted in adjacent coastal waters.

4 Conclusion

This study reveals a rapid decline in DOC and CDOM concentrations along the diffuse and non-point source supra-permafrost groundwater flow from coastal permafrost bluffs through sandy beach sediments and to adjacent nearshore seawater. This decline is likely largely influenced by the dilution of a high-concentration, low-volume supra-permafrost groundwater (~~fresh endmember~~) with volumetrically dominant recirculating

seawater. At the same time, our results provide evidence consistent with Fe-mineral interactions and microbial reprocessing contributing to a concurrent compositional transformation of the DOM pool within the STE. Rather than acting as a quantitative DOC sink, the Arctic beach STE can be described as a DOM quality filter, where these processes are associated with a relative decrease in humic-like, high-molecular-weight compounds and a shift toward more protein-like, microbially-derived DOM prior to export to coastal waters. Our results indicate-suggest that nearshore and intertidal zones, particularly subterranean estuaries where supra-permafrost groundwater mixes with recirculated seawater, ~~play an important role in~~ contribute to shaping the DOM composition of DOM prior to its export to coastal waters ~~DOM composition and persistence~~. Similar transformation processes have been reported in temperate and tropical STEs, suggesting that Arctic beach systems share common biogeochemical features despite the influence of permafrost. However, whether Arctic beaches primarily act as transient filters, ~~permanent-long-term~~ carbon sinks, or dynamic biogeochemical reactors remains ~~an open question~~unresolved, underscoring the need for further research into their contribution to the transfer of carbon from permafrost to the ocean. Given ongoing permafrost thaw and increasing coastal erosion, improving our understanding of these nearshore processes is important for constraining Arctic carbon cycling and predicting their potential role in climate feedbacks.

Data availability: Along with this submission, the dataset used in this research was submitted and accepted for publication to Pangaea Data Publisher (www.pangaea.de). Once this article accepted for publication, the moratorium in place will be lifted and the dataset generated during the study will be freely available in the Pangaea repository. Here is the hyperlink and DOI toward the dataset: <https://doi.pangaea.de/10.1594/PANGAEA.960986>

Author contribution: AF: Investigation, Conceptualization, Methodology, Formal analysis, Validation, Data curation, Visualization, Writing - Original Draft Preparation, Writing - Review & Editing; J-FL: Supervision, Methodology, Validation, Data curation, Writing - Review & Editing; GC: Conceptualization, Methodology, Validation, Data curation, Supervision, Project administration, Funding acquisition, Writing - Review & Editing.

517 **Competing interests:** The authors declare that there are no competing interests.

518 **Acknowledgements:** This project was made possible through the tremendous support of the hunters and
519 trappers committee, the hamlet and town council, in addition to several members of the Inuvialuit Nunangit
520 Sannaiqtuaq community of Tuktoyaktuk. The authors would like to express their gratitude to Dustin Whalen
521 (NRCan) for his invaluable contributions that made the fieldwork possible, as well as for his assistance in
522 providing resources and coordinating the fieldwork. We also want to thank Charlotte Irish, James Pokiak, Angus
523 Robertson, Bay Berry, and Brian Mayhew for their precious help during fieldwork in the summers of 2019 and
524 2021. We thank Claude Belzile for analyzing the DOC sample at ISMER-UQAR, Frédéric Bélanger for
525 conducting the CDOM and FDOM analysis of the 2019 samples and Antoine Biehler for his help with database
526 management. We also want to thank Simon Bélanger (UQAR) and Celine Guéguen (U. Sherbrooke) for
527 providing early feedback on this manuscript. We thank the different anonymous reviewers and the associate
528 editor for their constructive and insightful feedback. This work represents a contribution to the scientific
529 programs of Nunataryuk, ArcticNet, and Québec-Océan.

530 **Financial support:** Financial support for this project has been provided by the Network of Centers of
531 Excellence of Canada ArcticNet (grant no. P-66), by Québec-Océan, funded through the Fonds de Recherche
532 du Québec – Nature et Technologies, and by the Aurora Research Institute – Aurora College. The authors
533 received additional funding from NSERC (RGPIN-2021-04332 to GC) and in-kind support from Natural
534 Resources Canada (Climate Change Geoscience Program and Polar Continental Shelf Program; grant no. 007-
535 19) for logistics support (equipment and helicopter time), and Crown–Indigenous Relations and Northern
536 Affairs Canada (Climate Change Preparedness in the North program, CCPN grant no. CCPN PN-NT-077-2018;
537 Beaufort Sea Regional Strategic Environmental Assessment Program, BRSEA agreement no. 239) in the form
538 of consultation tour of the ISR (March 2019). AF also received grants from the Northern Scientific Training
539 Program (NSTP) and from Québec-Océan.

5 References

- Amoako, K., Reckhardt, A., Roberts, M., Meyer, R., Brick, S., Dittmar, T., and Waska, H.: Organo-mineral interactions modulate organic carbon retention and mobility in a deep subterranean estuary of a high-energy beach. *Frontiers in Water* 7:1507564. 2025
- Anschutz, P., T. Smith, A. Mouret, J. Deborde, S. Bujan, D. Poirier, and P. Lacroix: Tidal sands as biogeochemical reactors. *Estuarine, Coastal and Shelf Science* 84 (1): 84–90. <https://doi.org/10.1016/j.ecss.2009.06.015>, 2009.
- Benner, R., and Amon, R. M.: The size-reactivity continuum of major bioelements in the ocean, *Annu. Rev. Mar. Sci.*, 7, 185–205, <https://doi.org/10.1146/annurev-marine-010213-135126>, 2015.
- Berggren, M., Gudas, C., Guillemette, F., Hensgens, G., Ye, L. and Karlsson, J.: Systematic microbial production of optically active dissolved organic matter in subarctic lake water. *Limnology and Oceanography*, 65(5), 1071–1085. <https://doi.org/10.1002/lno.11362>, 2020.
- Berry, H. B., Whalen, D., and Lim, M.: Long-term ice-rich permafrost coast sensitivity to air temperatures and storm influence: lessons from Pullen Island, Northwest Territories, Canada, *Arct. Sci.*, 7, 723–745, <https://doi.org/10.1139/as-2020-0003>, 2021.
- Bristol, E. M., Connolly, C. T., Lorenson, T. D., Richmond, B. M., Ilgen, A. G., Choens, R. C., Bull, D. L., Kanevskiy, M., Iwahana, G., and Jones, B. M.: Geochemistry of coastal permafrost and erosion-driven organic matter fluxes to the Beaufort Sea near Drew Point, Alaska, *Front. Earth Sci.*, 8, <https://doi.org/10.3389/FEART.2020.598933>, 2021.

559 Bro, R.: PARAFAC. Tutorial and applications, *Chemometr. Intell. Lab.*, 38, 149–171,
560 [https://doi.org/10.1016/S0169-7439\(97\)00032-4](https://doi.org/10.1016/S0169-7439(97)00032-4), 1997.

561 Bullock, E. J., Schaal, I. V., Cardenas, M. B., McClelland, J. W., Henderson, P. B., & Charette, M. A. :
562 Seasonality of submarine groundwater discharge to an Arctic coastal lagoon. *Limnology and Oceanography*.
563 <https://doi.org/10.1002/lno.12585>, 2024.

564 Chaillou, G., Lemay-Borduas, F., Larocque, M., Couturier, M., Biehler, A., and Tommi-Morin, G.: Flow and
565 discharge of groundwater from a snowmelt-affected sandy beach, *J. Hydrol.*, 557, 4–15,
566 <https://doi.org/10.1016/J.JHYDROL.2017.12.010>, 2018.

567 Chaillou, G., Tommi-Morin, G., and Mucci, A. : Production and fluxes of inorganic carbon and alkalinity in a
568 subarctic subterranean estuary. *Front. Mar. Sci.* 11:1323463. <https://doi.org/10.3389/fmars.2024.1323463>,
569 2024.

570 Charkin, A. N., Van Der Loeff, M. R., Shakhova, N. E., Gustafsson, Ö., Dudarev, O. V., Cherepnev, M. S., et
571 al.: Discovery and characterization of submarine groundwater discharge in the Siberian Arctic seas: A case
572 study in the Buor-Khaya Gulf, Laptev Sea. *The Cryosphere*, 11, 2305–2327. [https://doi.org/10.5194/tc-11-](https://doi.org/10.5194/tc-11-2305-2017)
573 2305-2017, 2017.

574 Charette, M. A., and Sholkovitz, E. R.: Oxidative precipitation of groundwater-derived ferrous iron in the
575 subterranean estuary of a coastal bay, *Geophys. Res. Lett.*, 29, 85-81, <https://doi.org/10.1029/2001GL014512>,
576 2002.

577 Charrette, M. A., and Sholkovitz, E. R.: Trace element cycling in a subterranean estuary: Part 2. Geochemistry
578 of the pore water, *Geochim. Cosmochim. Ac.*, 70, 811–826, <https://doi.org/10.1016/j.gca.2005.10.019>, 2006.

579 Chen, C. M., Dynes, J. J., Wang, J., Sparks, D. L. : Properties of Fe-Organic Matter Associations via
 580 Coprecipitation versus Adsorption. *Nature Reviews Earth & Environment* 48 (23), 13751– 13759,
 581 <https://doi.org/10.1021/es503669u>, 2014.

582 Connolly, C. T., Cardenas, M. B., Burkart, G. A., Spencer, R. G. M., and McClelland, J. W.: Groundwater as a
 583 major source of dissolved organic matter to Arctic coastal waters, *Nat. Commun.*, 11, 1479,
 584 <https://doi.org/10.1038/s41467-020-15250-8>, 2020.

585 Couturier, M., Nozais, C., and Chaillou, G.: Microtidal subterranean estuaries as a source of fresh terrestrial
 586 dissolved organic matter to the coastal ocean, *Mar. Chem.*, 186, 46–57,
 587 <https://doi.org/10.1016/j.marchem.2016.08.001>, 2016.

588 DeFrancesco, C., & Guéguen, C.: Long-term trends in dissolved organic matter composition and its relation to
 589 sea ice in the Canada Basin, Arctic Ocean (2007–2017). *Journal of Geophysical Research: Oceans*, 126,
 590 e2020JC016578. <https://doi.org/10.1029/2020JC016578>, 2021.

591 Del Vecchio, R., and Blough, N. V.: Spatial and seasonal distribution of chromophoric dissolved organic
 592 matter and dissolved organic carbon in the Middle Atlantic Bight, *Mar. Chem.*, 89, 169–187,
 593 <https://doi.org/10.1016/j.marchem.2004.02.027>, 2004.

594 Demir, C., McClelland, J. W., Bristol, E., Charette, M. A. & Cardenas, M. B.: Coastal Supra-Permafrost
 595 Aquifers of the Arctic and Their Significant Groundwater, Carbon, and Nitrogen Fluxes. *Geophysical Research*
 596 *Letter* 51, e2024GL109142, 2024.

597 Dittmar, T. & Kattner, G.: The biogeochemistry of the river and shelf ecosystem of the Arctic Ocean: a review.
 598 *Marine Chemistry* 83, 103–120, 2003.

599 Frenette, J-J., Massicotte, P., and Lapierre, J-F.: Colorful Niches of Phytoplankton Shaped by the Spatial
600 Connectivity in a Large River Ecosystem: A Riverscape Perspective. PLoS ONE, 7(4), e35891,
601 <https://doi.org/10.1371/journal.pone.0035891>, 2012.

602 Fichot, C. G., and Benner, R.: A novel method to estimate DOC concentrations from CDOM absorption
603 coefficients in coastal waters, *Geophys. Res. Lett.*, 38, <https://doi.org/10.1029/2010GL046152>, 2011.

604 Fichot, C. G., Kaiser, K., Hooker, S. B., Amon, R. M. W., Babin, M., Bélanger, S., Walker, S. A., and
605 Benner, R.: Pan-Arctic distributions of continental runoff in the Arctic Ocean, *Sci. Rep.-UK*, 3, 1053,
606 <https://doi.org/10.1038/srep01053>, 2013.

607 Fichot, C. G., Kaiser, K., Hooker, S. B., Amon, R. M. W., Babin, M., Bélanger, S., Walker, S. A., and Benner,
608 R.: Pan-Arctic distributions of continental runoff in the Arctic Ocean, *Scientific Reports* 3, 1053. [https://doi.org/](https://doi.org/10.1038/srep01053)
609 [10.1038/srep01053](https://doi.org/10.1038/srep01053), 2013.

610 Fichot, C. G., and Benner, R.: The fate of terrigenous dissolved organic carbon in a river-influenced ocean
611 margin, *Global Biogeochem. Cy.*, 28, 300–318, <https://doi.org/10.1002/2013GB004670>, 2014.

612 Fouché, J., Christiansen, C. T., Lafrenière, M. J., Grogan, P., and Lamoureux, S. F.: Canadian permafrost stores
613 large pools of ammonium and optically distinct dissolved organic matter, *Nat. Commun.*, 11, 4500,
614 <https://doi.org/10.1038/s41467-020-18331-w>, 2020.

615 Fritz, M., Wetterich, S., Meyer, H., Schirrmeister, L., Lantuit, H., and Pollard, W. H.: Origin and characteristics
616 of massive ground ice on Herschel Island (western Canadian Arctic) as revealed by stable water isotope and
617 hydrochemical signatures, *Permafrost Periglac.*, 22, 26–38, <https://doi.org/10.1002/ppp.714>, 2011.

618 Fritz, M., Vonk, J. E., and Lantuit, H.: Collapsing Arctic coastlines, *Nat. Clim. Change*, 7, 6–7,
619 <https://doi.org/10.1038/nclimate3188>, 2017.

620 Fritz, M., Wetterich, S., McAlister, J., and Meyer, H.: A new local meteoric water line for Inuvik (NT, Canada),
 621 Earth Syst. Sci. Data, 14, 57–63, <https://doi.org/10.5194/essd-14-57-2022>, 2022.

622 Guo, L., Ping, C. L., and Macdonald, R. W.: Mobilization pathways of organic carbon from permafrost to
 623 arctic rivers in a changing climate, Geophys. Res. Lett., 34, <https://doi.org/10.1029/2007GL030689>, 2007.

624 Hansen, A.M., Kraus, T.E.C., Pellerin, B.A., Fleck, J.A., Downing, B.D. and Bergamaschi, B.A.: Optical
 625 properties of dissolved organic matter (DOM): Effects of biological and photolytic degradation. Limnol.
 626 Oceanogr., 61: 1015-1032. <https://doi.org/10.1002/lno.10270>, 2016.

627 Hayes, S., Lim, M., Whalen, D., Mann, P. J., Fraser, P., Penlington, R., and Martin, J.: The role of massive ice
 628 and exposed headwall properties on retrogressive thaw slump activity. J. Geophys. Res. Earth Surf., 127,
 629 e2022JF006602. <https://doi.org/10.1029/2022JF006602>, 2022.

630 Hayes, S.: Massive ice and topographic controls on retrogressive thaw slump dynamics: Peninsula point,
 631 western canadian arctic (Order No. 28187262). Doctoral thesis, Northumbria University. from Northumbria
 632 Research Link: <http://nrl.northumbria.ac.uk/id/eprint/44081>, 2020.

633 Hébert, A.-J., Flamand, A., and Chaillou, G.: Origins and transformations of terrigenous dissolved organic
 634 matter in a transgressive coastal system, Estuar. Coast. Shelf S., 279, 108137,
 635 <https://doi.org/10.1016/J.ECSS.2022.108137>, 2022.

636 Hedges, J. I., and Keil, R. G.: Sedimentary organic matter preservation: an assessment and speculative
 637 synthesis, Mar. Chem., 49, 81–115, [https://doi.org/10.1016/0304-4203\(95\)00008-F](https://doi.org/10.1016/0304-4203(95)00008-F), 1995.

638 Helms, J. R., Stubbins, A., Ritchie, J. D., Minor, E. C., Kieber, D. J., and Mopper, K.: Absorption spectral
 639 slopes and slope ratios as indicators of molecular weight, source, and photobleaching of chromophoric dissolved
 640 organic matter, Limnol. Oceanogr., 53, 955–969, <https://doi.org/10.4319/LO.2008.53.3.0955>, 2008.

641 Holmes, R.M., McClelland, J.W., Peterson, B.J. *et al.* Seasonal and Annual Fluxes of Nutrients and Organic
642 Matter from Large Rivers to the Arctic Ocean and Surrounding Seas. *Estuaries and Coasts*, 35, 369–382,
643 <https://doi.org/10.1007/s12237-011-9386-6>, 2012.

644 Huguet, A., Vacher, L., Relexans, S., Saubusse, S., Froidefond, J.-M., and Parlanti, E.: Properties of fluorescent
645 dissolved organic matter in the Gironde Estuary, *Org. Geochem.*, 40, 706–719,
646 <https://doi.org/10.1016/j.orggeochem.2009.03.002>, 2009.

647 Jones, B. M., Irrgang, A. M., Farquharson, L. M., Lantuit, H., Whalen, D., Ogorodov, S., Grigoriev, M.,
648 Tweedie, C., Gibbs, A. E., Strzelecki, M. C., Baranskaya, A., Belova, N., Sinitsyn, A., Kroon, A., Maslakov,
649 A., Vieira, G., Grosse, G., Overduin, P., Nitze, I., Maio, C., Overbeck, J., Bendixen, M., Zagórski, P., and
650 Romanovsky, V. E.: Arctic Report Card 2020: Coastal Permafrost Erosion, [https://doi.org/10.25923/e47w-](https://doi.org/10.25923/e47w-dw52)
651 [dw52](https://doi.org/10.25923/e47w-dw52), 2020.

652 Kaiser, K., Benner, R., and Amon, R. M. W.: The fate of terrigenous dissolved organic carbon on the Eurasian
653 shelves and export to the North Atlantic, *J. Geophys. Res.-Oceans*, 122, 4–22,
654 <https://doi.org/10.1002/2016jc012380>, 2017a.

655 Kaiser, K., Canedo-Oropeza, M., McMahon, R., and Amon, R. M. W.: Origins and transformations of dissolved
656 organic matter in large Arctic rivers, *Sci. Rep.*, 7, <https://doi.org/10.1038/s41598-017-12729-1>, 2017b.

657 Kipp, L. E., Charette, M. A., Moore, W. S., Henderson, P. B., and Rigor, I. G.: Increased fluxes of shelf-derived
658 materials to the central arctic ocean, *Sci. Adv.*, 4, <https://doi.org/10.1126/SCIADV.AAO1302>, 2018.

659 Kipp, L. E., Chaillou, G., Kienast, M., Tamborski, J., and Whalen, D. : Radon and radium isotope signatures
660 on a massive ice- and permafrost-rich coastline, *Geochimica et Cosmochimica Acta*,
661 <https://doi.org/10.1016/j.gca.2025.06.014>, 2025.

662 Kwasigroch, U., Beldowska, M., Jędruch, A., and Saniewska, D.: Coastal erosion—a “new” land-based source
 663 of labile mercury to the marine environment, *Environ. Sci. Pollut. Res.*, 25, 28682–28694,
 664 <https://doi.org/10.1007/S11356-018-2856-7>, 2018.

665 Lalonde, K., Mucci, A., Ouellet, A., Gélinas, Y. : Preservation of organic matter in sediments promoted by iron.
 666 *Nature* 483, 198–200. <http://dx.doi.org/10.1038/nature10855>, 2012.

667 Lantuit, H., Overduin, P. P., Couture, N., Wetterich, S., Aré, F., Atkinson, D., Brown, J., Cherkashov, G.,
 668 Drozdov, D., Donald Forbes, L., Graves-Gaylord, A., Grigoriev, M., Hubberten, H. W., Jordan, J., Jorgenson,
 669 T., Ødegård, R. S., Ogorodov, S., Pollard, W. H., Rachold, V., Sedenko, S., Solomon, S., Steenhuisen, F.,
 670 Streletskaya, I., and Vasiliev, A.: The Arctic Coastal Dynamics Database: A New Classification Scheme and
 671 Statistics on Arctic Permafrost Coastlines, *Estuaries Coasts*, 35, 383–400, [https://doi.org/10.1007/S12237-010-](https://doi.org/10.1007/S12237-010-9362-6)
 672 9362-6, 2012.

673 Lapham, L. L., Dallimore, S. R., Magen, C., Henderson, L. C., Powers, L. C., Gonsior, M., Clark, B., Côté, M.,
 674 Fraser, P., and Orcutt, B. N: Microbial Greenhouse Gas Dynamics Associated With Warming
 675 Coastal Permafrost, Western Canadian Arctic. *Front. Earth Sci.* 8:582103.
 676 <https://doi.org/10.3389/feart.2020.582103>, 2020.

677 Lecher, A. L.: Groundwater Discharge in the Arctic: A Review of Studies and Implications for
 678 Biogeochemistry, *Hydrol.* 4(3), 41, <https://doi.org/10.3390/hydrology4030041>, 2017.

679 Li, Q., Hu, W., Li, L., and Li, Y.: Interactions between organic matter and Fe oxides at soil micro-interfaces:
 680 Quantification, associations, and influencing factors, *Sci. Total Environ.*, 855,
 681 <https://doi.org/10.1016/j.scitotenv.2023.158710>, 2023.

682 Linkhorst, A., Dittmar, T., and Waska, H.: Molecular Fractionation of Dissolved Organic Matter in a Shallow
 683 Subterranean Estuary: The Role of the Iron Curtain, *Environ. Sci. Technol.*, 51, 1312–1320,
 684 <https://doi.org/10.1021/acs.est.6b03608>, 2017.

685 Lizotte, M., Juhls, B., Matsuoka, A., Massicotte, P., Mével, G., Anikina, D. O. J., Antonova, S., Bécu, G.,
 686 Béguin, M., Bélanger, S., Bossé-Demers, T., Bröder, L., Bruyant, F., Chaillou, G., Comte, J., Couture, R.-M.,
 687 Devred, E., Deslongchamps, G., Dezutter, T., Dillon, M., Doxaran, D., Flamand, A., Fell, F., Ferland, J., Forget,
 688 M.-H., Fritz, M., Gordon, T. J., Guilmette, C., Hilborn, A., Hussherr, R., Irish, C., Joux, F., Kipp, L., Laberge-
 689 Carignan, A., Lantuit, H., Leymarie, E., Mannino, A., Maury, J., Overduin, P., Oziel, L., Stedmon, C., Thomas,
 690 C., Tisserand, L., Tremblay, J.-É., Vonk, J., Whalen, D., and Babin, M.: Nunataryuk field campaigns:
 691 Understanding the origin and fate of terrestrial organic matter in the coastal waters of the Mackenzie Delta
 692 region, *Earth Syst. Sci. Data*, 15, 1617–1653, <https://doi.org/10.5194/essd-2022-163>, 2023.

693 Macdonald, R. W., Solomon, S. M., Cranston, R. E., Welch, H. E., Yunker, M. B., and Gobeil, C.: A sediment
 694 and organic carbon budget for the Canadian Beaufort Shelf, *Mar. Geol.*, 144, 255–273,
 695 [https://doi.org/10.1016/S0025-3227\(97\)00106-0](https://doi.org/10.1016/S0025-3227(97)00106-0), 1998.

696 Mackay, J.R.: Fifty years (1935-1985) of coastal retreat west of Tuktoyaktuk, District of Mackenzie, Geological
 697 Survey of Canada, 86-1A, 727–735, 1986.

698 Massicotte, P., Asmala, E., Stedmon, C., and Markager, S.: Global distribution of dissolved organic matter
 699 along the aquatic continuum: Across rivers, lakes and oceans, *Sci. Total Environ.*, 609, 180–191,
 700 <https://doi.org/10.1016/j.scitotenv.2017.07.076>, 2017.

701 Matsuoka, A., Bricaud, A., Benner, R., Para, J., Sempéré, R., Prieur, L., Bélanger, S., and Babin, M.: Tracing
 702 the transport of colored dissolved organic matter in water masses of the Southern Beaufort Sea: relationship
 703 with hydrographic characteristics, *Biogeosciences*, 9, 925–940, <https://doi.org/10.5194/bg-9-925-2012>, 2012.

704 Mehra, O.P., Jackson, M.L., : Iron oxide removal from soils and clays by a dithionitecitrate system buffered
705 with sodium bicarbonate. *Clay Clay Miner.* 7, 317–327, 1960.

706 McKnight, D. M., Boyer, E. W., Westerhoff, P. K., Doran, P. T., Kulbe, T., and Andersen, D. T.:
707 Spectrofluorometric characterization of dissolved organic matter for indication of precursor organic material
708 and aromaticity, *Limnol. Oceanogr.*, 46, 38–48, <https://doi.org/10.4319/lo.2001.46.1.0038>, 2001.

709 McMeans, B. C., McCann, K. S., Humphries, M., Rooney, N., and Fisk, A. T.: Food Web Structure in
710 Temporally-Forced Ecosystems, *Trends Ecol. Evol.*, 30, 662–672,
711 <https://doi.org/10.1016/J.TREE.2015.09.001>, 2015.

712 Meilleur, C., Kamula, M., Kuzyk, Z. A., and Guéguen, C.: Insights into surface circulation and mixing in James
713 Bay and Hudson Bay from dissolved organic matter optical properties, *J. Mar. Syst.*, 238, 103841,
714 <https://doi.org/10.1016/J.JMARSYS.2022.103841>, 2023.

715 Moore, W. S.: The subterranean estuary: a reaction zone of ground water and sea water, *Mar. Chem.*, 65(1–2),
716 111–125, doi:10.1016/S0304-4203(99)00014-6, 1999.

717 Moore, W. S., Benitez-Nelson, C., Schutte, C., Moody, A., Shiller, A., Sibert, R. J., and Joye, S.: SGD-OD:
718 investigating the potential oxygen demand of submarine groundwater discharge in coastal systems, *Sci. Rep.*,
719 14, 9249, <https://doi.org/10.1038/s41598-024-59229-7>, 2024.

720 Murphy, K. R., Stedmon, C. A., Graeber, D., and Bro, R.: Fluorescence spectroscopy and multi-way techniques.
721 PARAFAC, *Anal. Methods*, 5, <https://doi.org/10.1039/c3ay41160e>, 2013.

722 Murphy, K.R., Stedmon, C.A., Wenig, P. and Bro, R.: OpenFluor – an online spectral library of auto-
723 fluorescence by organic compounds in the environment. *Analytical Methods*, 6, 658–661.
724 <https://doi.org/10.1039/C3AY41935E>, 2014.

725 Nelson, N. B., Siegel, D. A., and Michaels, A. F.: Seasonal dynamics of colored dissolved material in the
 726 Sargasso Sea, *Deep Sea Res. Part I Oceanogr. Res. Pap.*, 45, 931–957, [https://doi.org/10.1016/S0967-](https://doi.org/10.1016/S0967-0637(97)00106-4)
 727 0637(97)00106-4, 1998.

728 Nelson, N. B., and Siegel, D. A.: The global distribution and dynamics of chromophoric dissolved organic
 729 matter, *Annu. Rev. Mar. Sci.*, 5, 447–476, <https://doi.org/10.1146/annurev-marine-120710-100751>, 2013.

730 Nguyen, H. T., Lee, Y. M., Hong, J. K., Chen, M., and Hur, J. : Climate warming-driven changes in the flux of
 731 dissolved organic matter and its effects on bacterial communities in the Arctic Ocean: A review. *Frontiers*
 732 *Marine Sciences*. 9, 968583. <https://doi.org/10.3389/fmars.2022.968583>, 2022.

733 Obu, J., Westermann, S., Bartsch, A., Berdnikov, N., Christiansen, H. H., Dashtseren, A., Delaloye, R.,
 734 Elberling, B., Etzelmüller, B., Kholodov, A., Khomutov, A., Kääb, A., Leibman, M. O., Lewkowicz, A. G.,
 735 Panda, S. K., Romanovsky, V., Way, R. G., Westergaard-Nielsen, A., Wu, T., Yamkhin, J., and Zou, D.:
 736 Northern Hemisphere permafrost map based on TTOP modelling for 2000–2016 at 1 km² scale, *Earth-Sci.*
 737 *Rev.*, 193, 299–316, <https://doi.org/10.1016/j.earscirev.2019.04.023>, 2019.

738 Ohno, T.: Fluorescence Inner-Filtering Correction for Determining the Humification Index of Dissolved
 739 Organic Matter, *Environ. Sci. Technol.*, 36 (4), 742–746, <https://doi.org/10.1021/es0155276>, 2002.

740 Olefeldt, D., Persson, A. and Turetsky, M.R.: Influence of the permafrost boundary on dissolved organic
 741 matter characteristics in rivers within the Boreal and Taiga plains of western Canada. *Environmental Research*
 742 *Letters*, 9(3), 035005. <https://doi.org/10.1088/1748-9326/9/3/035005>, 2014.

743 Ouellette, D. S.: Thermal and Mechanical Modeling of Coastal Erosion Processes on Tuktoyaktuk Island,
 744 Northwest Territories, Master thesis, University of Calgary, Calgary, Canada, 2021.

745 Pucher, M., Wünsch, U., Weigelhofer, G., Murphy, K., Hein, T., and Graeber, D.: staRdom: Versatile Software
 746 for Analyzing Spectroscopic Data of Dissolved Organic Matter in R, *Water*, 11, 2366,
 747 <https://doi.org/10.3390/w11112366>, 2019.

748 Retelletti Brogi, S., Jung, J., Ha, S.-Y., Hur, J.: Seasonal differences in dissolved organic matter properties and
 749 sources in an Arctic fjord: Implications for future conditions, *Science of The Total Environment*, 694,
 750 <https://doi.org/10.1016/j.scitotenv.2019.133740>, 2019.

751 Riedel, T., Zak, D., Biester, H., and Dittmar, T.: Iron traps terrestrially derived dissolved organic matter at redox
 752 interfaces, *PNAS*, 110, 10101–10105, <https://doi.org/10.1073/pnas.1221487110>, 2013.

753 Santos, I. R., Eyre, B. R., Huettel, M.: The driving forces of porewater and groundwater flow in permeable
 754 coastal sediments: A review, *Estuarine, Coastal and Shelf Science* : 98, 1-15,
 755 <https://doi.org/10.1016/j.ecss.2011.10.024>, 2012.

756 Schuur, E. A. G., McGuire, A. D., Schädel, C., Grosse, G., Harden, J. W., Hayes, D. J., Hugelius, G., Koven,
 757 C. D., Kuhry, P., Lawrence, D. M., Natali, S. M., Olefeldt, D., Romanovsky, V. E., Schaefer, K., Turetsky, M.
 758 R., Treat, C. C., and Vonk, J. E.: Climate change and the permafrost carbon feedback, *Nature*, 520, 171–179,
 759 <https://doi.org/10.1038/nature14338>, 2015.

760 Sirois, M., Couturier, M., Barber, A., Gélinas, Y., and Chaillou, G.: Interactions between iron and organic
 761 carbon in a sandy beach subterranean estuary, *Mar. Chem.*, 202, 86–96,
 762 <https://doi.org/10.1016/j.marchem.2018.02.004>, 2018.

763 Solomon, S. M.: Spatial and temporal variability of shoreline change in the Beaufort-Mackenzie region,
 764 northwest territories, Canada, *Geo-Mar. Lett.*, 25, 127–137, <https://doi.org/10.1007/S00367-004-0194-X>, 2005.

765 Söndergaard, M., Stedmon, C. A., and Borch, N. H.: Fate of terrigenous dissolved organic matter (DOM) in
 766 estuaries: Aggregation and bioavailability. *Ophelia*, 57(3), 161–176.
 767 <https://doi.org/10.1080/00785236.2003.10409512>, 2003.

768 Spencer, R. G., Aiken, G. R., Dornblaser, M. M., Butler, K. D., Holmes, R. M., Fiske, G., and Stubbins, A.:
 769 Chromophoric dissolved organic matter export from US rivers, *Geophys. Res. Lett.*, 40, 1575–1579,
 770 <https://doi.org/10.1002/grl.50357>, 2013.

771 Stedmon, C. A., Amon, R. M. W., Rinehart, A. J., and Walker, S. A.: The supply and characteristics of colored
 772 dissolved organic matter (CDOM) in the Arctic Ocean: Pan Arctic trends and differences, *Mar. Chem.*, 124,
 773 108–118, <https://doi.org/10.1016/j.marchem.2010.12.007>, 2011.

774 Stedmon, C. A., Markager, S., and Bro, R.: Tracing dissolved organic matter in aquatic environments using a
 775 new approach to fluorescence spectroscopy, *Mar. Chem.*, 82, 239–254, [https://doi.org/10.1016/s0304-](https://doi.org/10.1016/s0304-4203(03)00072-0)
 776 [4203\(03\)00072-0](https://doi.org/10.1016/s0304-4203(03)00072-0), 2003.

777 Stookey, L. L.: Ferrozine—A New Spectrophotometric Reagent for Iron, *Anal. Chem.*, 42, 779–781,
 778 <https://doi.org/10.1021/AC60289A016>, 1970.

779 Tanguy, R., Whalen, D., Prates, G., and Vieira, G.: Shoreline change rates and land to sea sediment and soil
 780 organic carbon transfer in eastern Parry Peninsula from 1965 to 2020 (Amundsen Gulf, Canada), *Arctic Sci.*,
 781 1–20, <https://doi.org/10.1139/as-2022-0028>, 2023.

782 Tanski, G., Wagner, D., Knoblauch, C., Fritz, M., Sachs, T., and Lantuit, H.: Rapid CO₂ release from eroding
 783 permafrost in seawater. *Geophysical Research Letters*, 46, 11244–11252.
 784 <https://doi.org/10.1029/2019GL084303>, 2019.

785 Terhaar, J., Lauerwald, R., Regnier, P. *et al.* Around one third of current Arctic Ocean primary production
786 sustained by rivers and coastal erosion. *Nature Communications*, 12, 169. [https://doi.org/10.1038/s41467-020-](https://doi.org/10.1038/s41467-020-20470-z)
787 20470-z, 2021

788 Thingstad, T. F., Bellerby, R. G. J., Bratbak, G., Børsheim, K. Y., Egge, J. K., Heldal, M., Larsen, A., Neill,
789 C., Nejstgaard, J., Norland, S., Sandaa, R. A., Skjoldal, E. F., Tanaka, T., Thyrhaug, R., and Töpper, B.:
790 Counterintuitive carbon-to-nutrient coupling in an Arctic pelagic ecosystem, *Nature*, 455, 387–390,
791 <https://doi.org/10.1038/nature07235>, 2008.

792 Utting, N., Clark, I., Lauriol, B., Wieser, M., and Aeschbach-Hertig, W.: Origin and Flow Dynamics of
793 Perennial Groundwater in Continuous Permafrost Terrain using Isotopes and Noble Gases: Case Study of the
794 Fishing Branch River, Northern Yukon, Canada, *Permafrost Periglac.*, 23, 91–106,
795 <https://doi.org/10.1002/ppp.1732>, 2012.

796 Viollier, E., Inglett, P. W., Hunter, K., Roychoudhury, A. N., and Van Cappellen, P.: The ferrozine method
797 revisited: Fe(II)/Fe(III) determination in natural waters, *Appl. Geochem.*, 15, 785–790,
798 [https://doi.org/10.1016/S0883-2927\(99\)00097-9](https://doi.org/10.1016/S0883-2927(99)00097-9), 2000.

799 Vonk, J. E., Semiletov, I. P., Dudarev, O. V., Eglinton, T. I., Andersson, A., Shakhova, N., Charkin, A., Heim,
800 B., and Gustafsson, Ö.: Preferential burial of permafrost-derived organic carbon in Siberian-Arctic shelf waters,
801 *J. Geophys. Res.: Oceans*, 119, 8410–8421, <https://doi.org/10.1002/2014JC010261>, 2014.

802 Vonk, J. E., Tank, S. E., Bowden, W. B., Laurion, I., Vincent, W. F., Alekseychik, P., Amyot, M., Billet, M.
803 F., Canário, J., Cory, R. M., Deshpande, B. N., Helbig, M., Jammet, M., Karlsson, J., Larouche, J.,
804 Macmillan, G., Rautio, M., Walter Anthony, K. M., and Wickland, K. P.: Reviews and syntheses: Effects of
805 permafrost thaw on Arctic aquatic ecosystems, *Biogeosciences*, 12, 7129–7167, [https://doi.org/10.5194/bg-](https://doi.org/10.5194/bg-12-7129-2015)
806 12-7129-2015, 2015.

807 Vonk, J. E., Fritz, M., Speetjens, N. J. et al. : The land–ocean Arctic carbon cycle. *Nature Reviews Earth*
808 *Environment* 6, 86–105, <https://doi.org/10.1038/s43017-024-00627-w>, 2025.

809 Walker, S. A., Amon, R. M. W., Stedmon, C., Duan, S., and Louchouart, P.: The use of PARAFAC modeling
810 to trace terrestrial dissolved organic matter and fingerprint water masses in coastal Canadian Arctic surface
811 waters, *J. Geophys. Res.*, 114, G00F06, <https://doi.org/10.1029/2009JG000990>, 2009.

812 Walvoord, M. A., and Striegl, R. G.: Increased groundwater to stream discharge from permafrost thawing in
813 the Yukon River basin: Potential impacts on lateral export of carbon and nitrogen, *Geophys. Res. Lett.*, 34,
814 <https://doi.org/10.1029/2007gl030216>, 2007.

815 Walvoord, M. A. & Kurylyk, B. L. : Hydrologic Impacts of Thawing Permafrost—A Review. *Vadose Zone J.*
816 15, 1–20, 2016.

817 Warren, C., Martin, J., Coote, G., Lim, M., Lee, R., Whalen, D.: 3D imaging of ground ice in the Canadian
818 Arctic. 13th International Workshop on Advanced Ground Penetrating Radar – IWAGPR2025. Thessaloniki,
819 Greece (2025)

820 Waska, H., Simon, H., Ahmerkamp, S., Greskowiak, J., Ahrens, J., Seibert, S. L., Schwalfenberg, K., Zielinski,
821 O., and Dittmar, T.: Molecular traits of dissolved organic matter in the subterranean estuary of a high-energy
822 beach: Indications of sources and sinks, *Front. Mar. Sci.*, 8, 54,
823 <https://doi.org/10.3389/FMARS.2021.607083/BIBTEX>, 2021.

824 Weishaar, J. L., Aiken, G. R., Bergamaschi, B. A., Fram, M. S., Fujii, R., and Mopper, K.: Evaluation of
825 specific ultraviolet absorbance as an indicator of the chemical composition and reactivity of dissolved organic
826 carbon, *Environ. Sci. Technol.*, 37, 4702–4708, <https://doi.org/10.1021/es030360x>, 2003.

827 Weyhenmeyer, G.A., Fröberg, M., Karlun, E., Khalili, M., Kothawala, D., Temnerud, J.,Tranvik, L.J., 2012.
828 Selective decay of terrestrial organic carbon during transportfrom land to sea. *Glob. Chang. Biol.* 18 (1), 349–
829 355, <https://doi.org/10.1111/j.1365-2486.2011.02544.x>, 2012.

830 Whalen, D., Forbes, D. L., Kostylev, V., Lim, M., Fraser, P., Nedimović, M. R., and Stuckey, S.: Mechanisms,
831 volumetric assessment, and prognosis for rapid coastal erosion of Tuktoyaktuk Island, an important natural
832 barrier for the harbour and community, *Can. J. Earth Sci.*, 59, 945–960, <https://doi.org/10.1139/cjes-2021-0101>,
833 2022.

834 Wurl, O., and Tsai, M. S.: Analysis of Dissolved and Particulate Organic Carbon with the HTCO Technique,
835 in: *Practical Guidelines for the Analysis of Seawater*, CRC Press, pp. 45–60, 2009.

836 Zhou, Z., Henkel, S., Kasten, S., and Holtappels, M.: The iron “redox battery” in sandy sediments: Its impact
837 on organic matter remineralization and phosphorus cycling. *Sci. Total. Environ.*, 865, 161168,
838 <https://doi.org/10.1016/j.scitotenv.2022.161168>., 2023.

839 Zhou, Z., Waska, H., Henkel, S., Dittmar, T., Kasten, S., and Holtappels, M.: Iron promotes the retention of
840 terrigenous Dissolved organic matter in subtidal Permeable Sediments. *Environ. Sci. Technol.*, 58,
841 6204–6214, <https://doi.org/10.1021/acs.est.3c09531>, 2024.

842

843

844

845

846 **TABLES**

847 Table 1 : EEM-PARAFAC components and their characteristics, based on literature assignments.

848

Comp.	Peak max Ex/Em (nm)	Coble peak	Description	# of matches (TCC>0.95)	Literature
C1	250(330)/428	A, C	Terrestrial humic-like, aromatic, HMW	16	Stedmon and Markager (2005); Murphy et al. (2018); Fouché et al. (2020)
C2	260/296	B	Protein-like (tyrosine), microbial/autochthonous, LMW	4	Stedmon and Markager (2005); Murphy et al. (2008); Olefeldt et al. (2014)
C3	280/324	T	Protein-like (tryptophan), associated with primary production	54	Berggren et al. (2020); Meilleur et al. (2023)
C4	285(375)/478	C	Terrestrial humic-like, visible range	47	Søndergaard et al. (2003); Walker et al. (2009); DeFrancesco and Gueguen (2021)
C5	295/412	M	Microbial humic-like associated with recent production	12	Retelletti Brogi et al. (2019); DeFrancesco and Gueguen (2021)

849

850 Table 2 : Particulate fraction characteristics of sandy beach sediment collected at Tuktoyaktuk Island and
851 Peninsula Point.

852

	Tuktoyaktuk Island N=3	Peninsula Point N=2
Sediment depth layer (cm)	0–50	0–30
Total organic carbon (% dry sediment)	0.43 ± 0.12	0.21 ± 0.07
Reactive sediment Fe (µg g ⁻¹ dry sediment)	2.246 ± 0.850	1.320 ± 0.250

853

854

855 **FIGURE CAPTIONS**

856 **Fig. 1** Map of the four sampling sites (red dots) located in the Northwest Territories, Canada.

857 **Fig. 2** Schematic cross-sections and photographs illustrating the coastal permafrost landscape at Tuktoyaktuk
858 Island (A) and Peninsula Point (B; not to scale) as examples of a retrogressive thaw slump system. Black arrows
859 show theoretical surface (solid lines) and subsurface (dashed lines) flow paths, while blue arrows represent
860 recirculating seawater driven by tidal and wave inputs into the intertidal zone of sandy beaches. Blue dots in
861 panel B mimic melt ponds. The locations of the different sample types (“massive ice,” “meltwater,” “beach
862 groundwater,” and “seawater”) are indicated. These conceptual models are adapted from Kipp et al. (2025)
863 based on Hayes et al. (2022) for Peninsula Point and Whalen et al. (2022) and Warren et al. (2025) for
864 Tuktoyaktuk Island.

865 **Fig. 3** Excitation-emission matrix (EEM) fluorescence spectra of the 5-component PARAFAC model.
866 Fluorescence is expressed in Raman Unit (R.U.).

867 **Fig. 4** Isotopic composition of massive ice, meltwater, beach groundwater and seawater samples collected in
868 2019, ~~plotted against showing the mixing line between samples across the salinity gradient, from the meltwater~~
869 ~~to the seawater.~~ The global meteoric water line (GMWL; Craig, 1961) and the local meteoric water line for
870 Inuvik (LMWL, Fritz et al. 2022) are also reported. Samples are colour-coded by salinity (S). Note that only
871 one massive ice sample (N=1) was collected in 2019.

Fig. 5 Distribution of (A) DOC, (B) a_{350} , (C) HIX, and (D) $SUVA_{254}$ across the salinity gradient (S) and among the different sample types (i.e., beach groundwater, massive ice, meltwater, and seawater). Note that only one massive ice sample is included. The zoomed-in panels display only beach groundwater and seawater samples, along with a theoretical mixing line (dashed black line) connecting the “fresh” and “salty” endmembers for DOC and a_{350} . These endmembers were defined based on the most enriched and depleted stable isotope signatures in the dataset, corresponding to salinity values of 20.08 (N = 1) and 0.18-0.30 (N = 8), respectively. As the underlying fluorescence and absorbance components used to calculate HIX and $SUVA_{254}$ indices were not retained, non-linear mixing curves could not be reconstructed. In the boxplots, black lines indicate the medians, whiskers represent the full range of data, and individual points denote outliers.

Fig. 6 Relationship between absorption coefficients at 350 nm (a_{350} , m^{-1}) and DOC concentrations ($\mu mol L^{-1}$) across all samples along the permafrost-to-nearshore aquatic continuum. ~~Data are shown on a logarithmic scale.~~

Fig. 7 Distribution of total dissolved Fe (F_{tot} , $\mu mol L^{-1}$) (A), and Fe-DOC ratio (B) across the salinity gradient (S), and by sample type. The y-axis in each panel is shown on a logarithmic scale to accommodate the broad range of Fe concentrations. Note that there is no F_{tot} data from the massive ice samples and Fe : DOC ratios were calculated only for F_{tot} concentration higher than the detection limit.

Fig. 8 Distribution of PARAFAC components (A) C1, (B) C2, (C) C3, (D) C4, and (E) C5 along the salinity gradient (S) and for the different types of collected samples (i.e., beach groundwater, massive ice, meltwater and nearshore seawater samples). Note that only one massive ice sample is reported here. For the boxplots, the black lines represent the median values, the whiskers represent the extent of the data, and the dot points represent the outlier values. Note that there is only one massive ice sample.

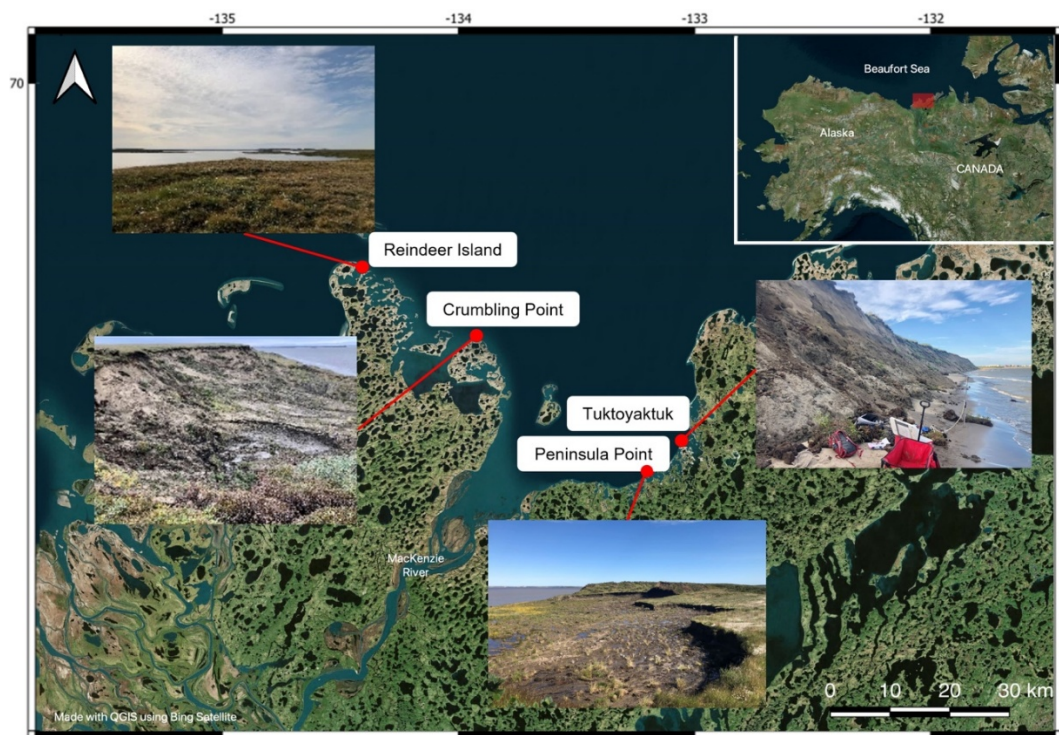


Fig. 1 Map of the four sampling sites (red dots) located in the Northwest Territories, Canada.

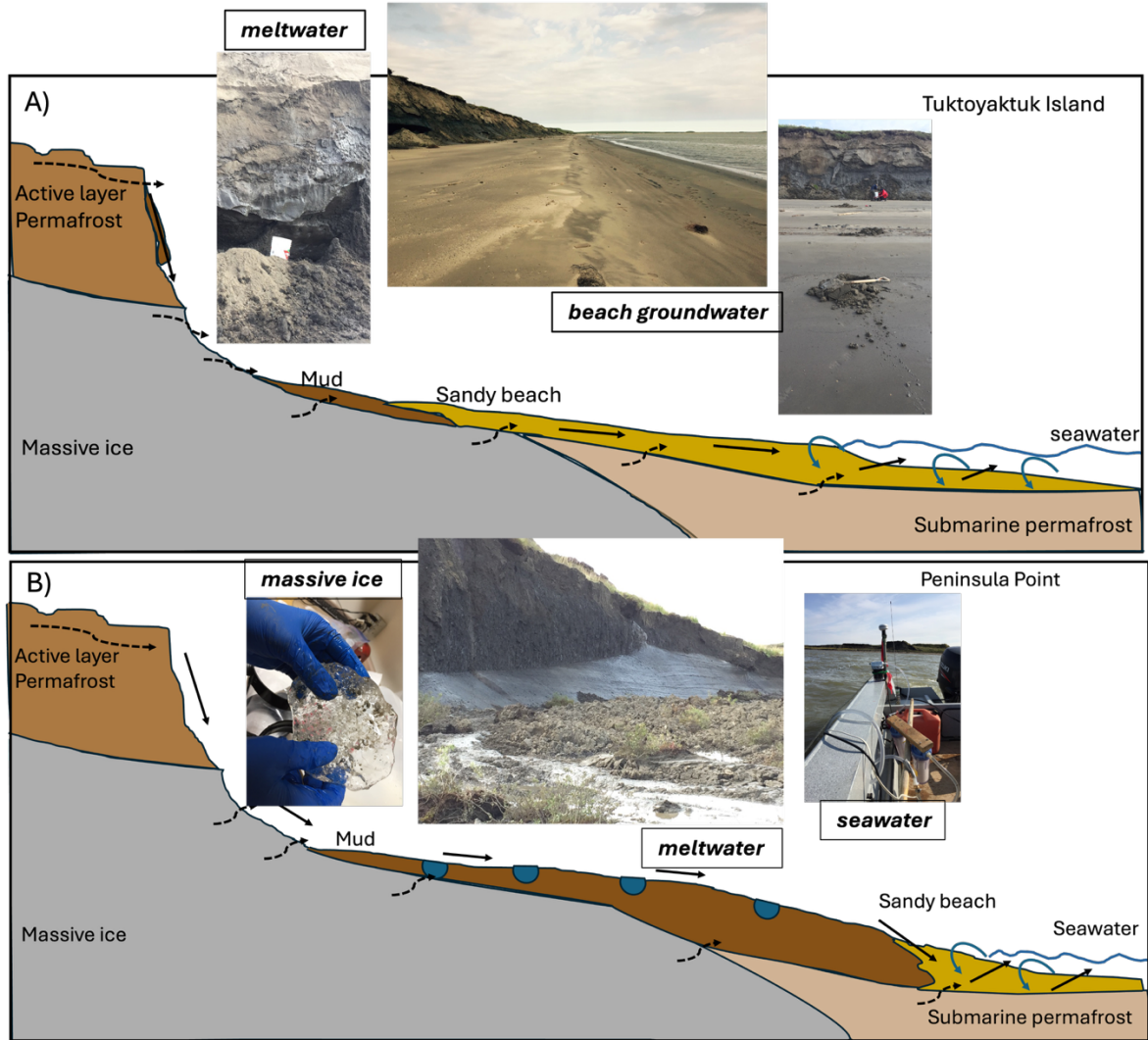
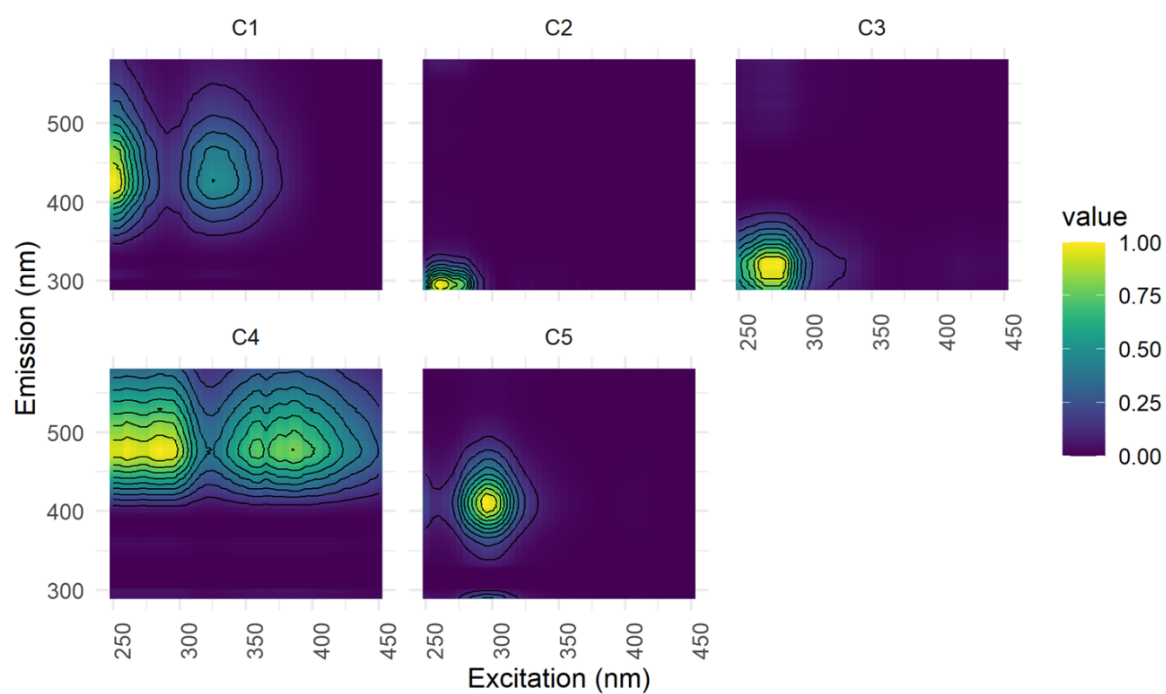


Fig. 2 Schematic cross-sections and photographs illustrating the coastal permafrost landscape at Tuktoyaktuk Island (A) and Peninsula Point (B; not to scale) as examples of a retrogressive thaw slump system. Black arrows show theoretical surface (solid lines) and subsurface (dashed lines) flow paths, while blue arrows represent recirculating seawater driven by tidal and wave inputs into the intertidal zone of sandy beaches. Blue dots in panel B mimic melt ponds. The locations of the different sample types (“massive ice,” “meltwater,” “beach groundwater,” and “seawater”) are indicated. These conceptual models are adapted from Kipp et al. (2025) based on Hayes et al. (2022) for Peninsula Point and Whalen et al. (2022) and Warren et al. (2025) for Tuktoyaktuk Island.

906



907

908 **Fig. 3** Excitation-emission matrix (EEM) fluorescence spectra of the 5-component PARAFAC model.

909

Fluorescence is expressed in Raman Unit (R.U.).

910

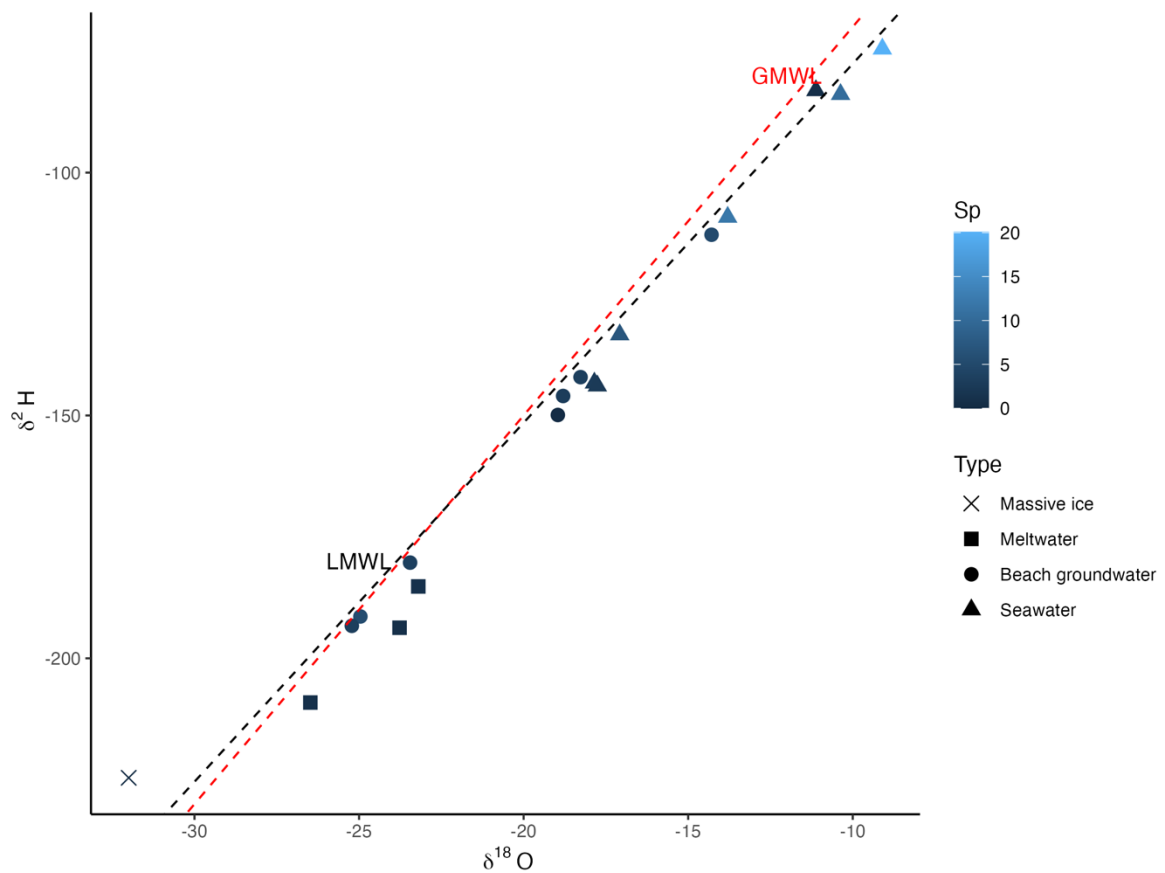


Fig. 4 Isotopic composition of massive ice, meltwater, beach groundwater and seawater samples collected in 2019, plotted against $\delta^{18}\text{O}$ showing the mixing line between samples across the salinity gradient, from the meltwater to the seawater. The global meteoric water line (GMWL; Craig, 1961) and the local meteoric water line for Inuvik (LMWL, Fritz et al. 2022) are also reported. Samples are colour-coded by salinity (S). Note that only one massive ice sample (N=1) was collected in 2019.

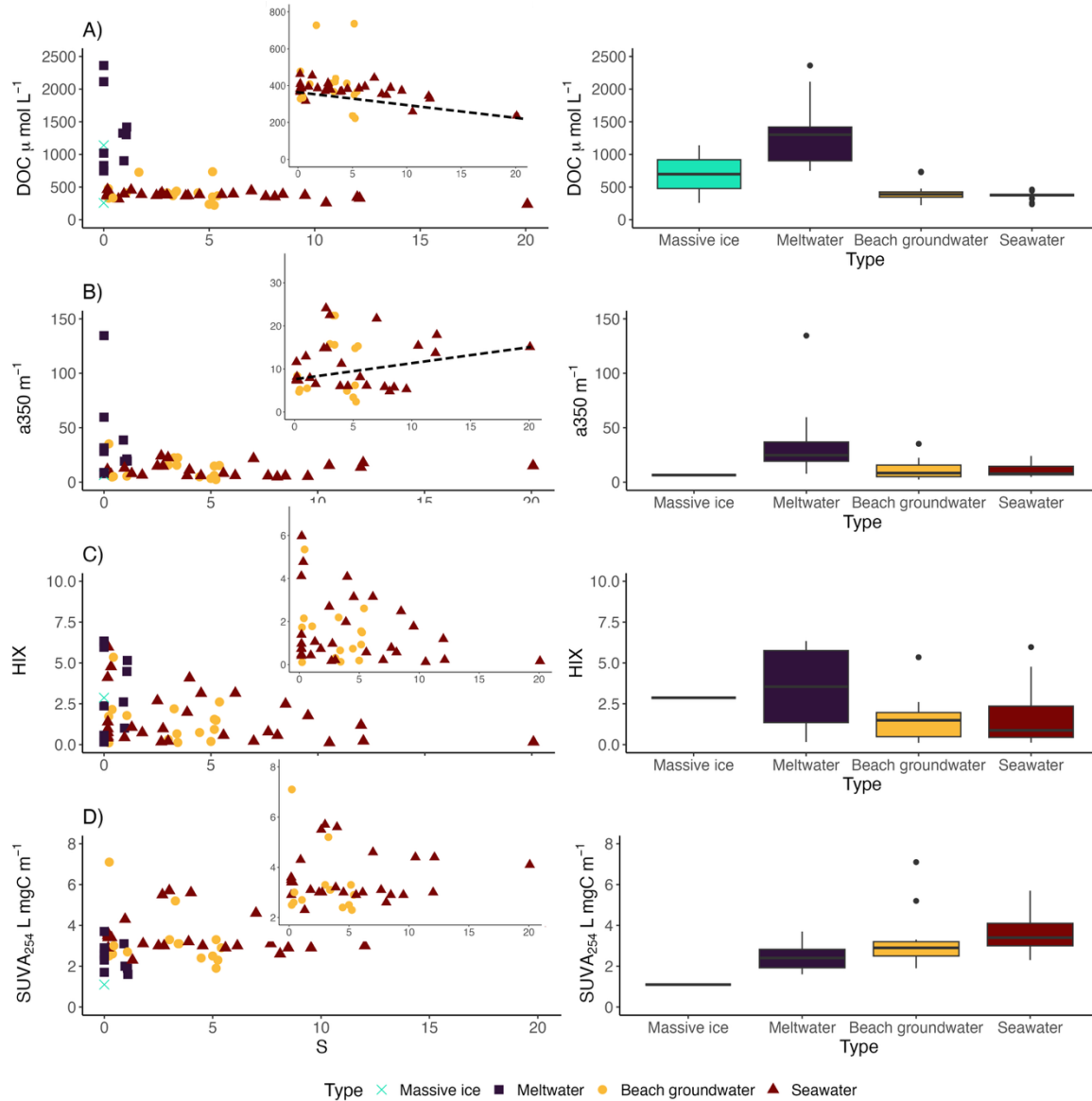


Fig. 5 Distribution of (A) DOC, (B) a_{350} , (C) HIX, and (D) SUVA_{254} across the salinity gradient (S) and among the different sample types (i.e., beach groundwater, massive ice, meltwater, and seawater). Note that only one massive ice sample is included. The zoomed-in panels display only beach groundwater and seawater samples, along with a theoretical mixing line (dashed black line) connecting the “fresh” and “salty” endmembers for DOC and a_{350} .—These endmembers were defined based on the most enriched and depleted stable isotope signatures in the dataset, corresponding to salinity values of 20.08 (N = 1) and 0.18–0.30 (N = 8), respectively. As the underlying fluorescence and absorbance components used to calculate HIX and SUVA_{254} indices were not retained, non-linear mixing curves could not be reconstructed. In the boxplots, black lines indicate the medians, whiskers represent the full range of data, and individual points denote outliers.

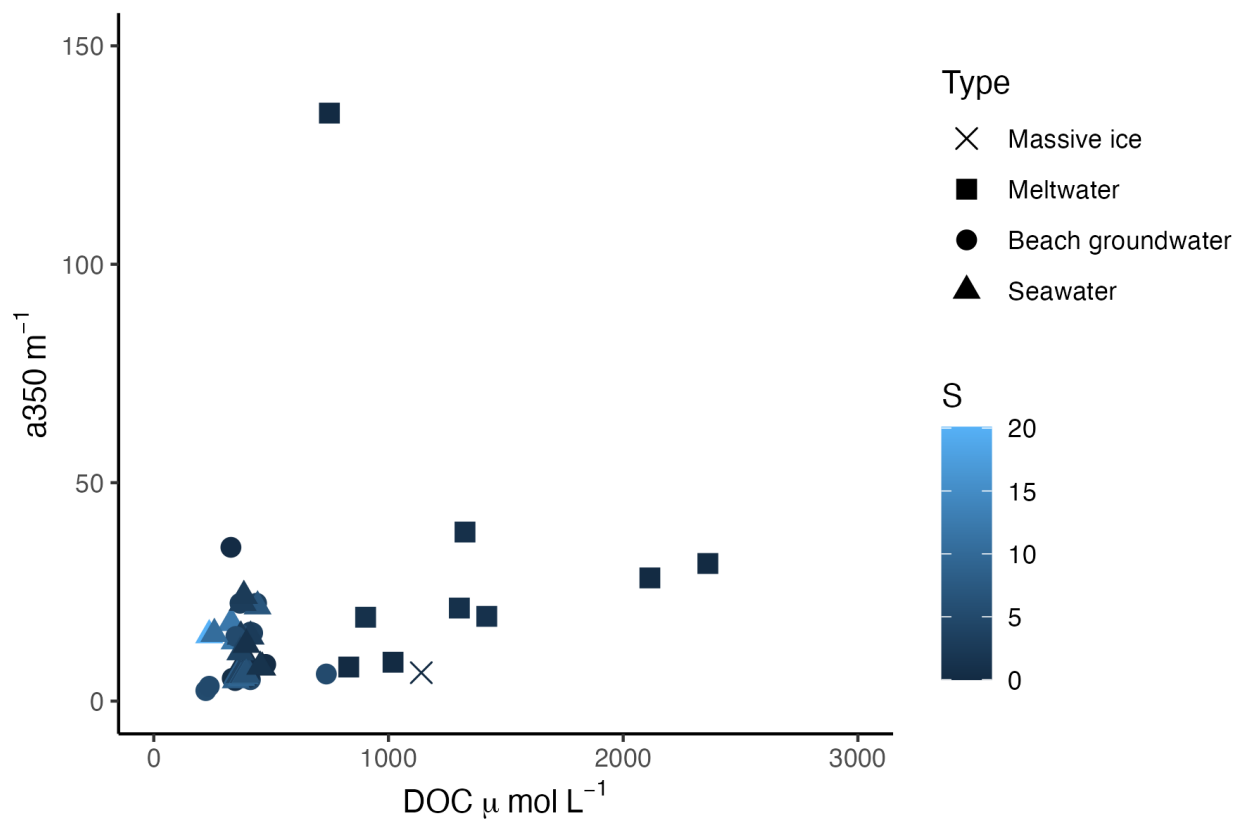


Fig. 6 Relationship between absorption coefficients at 350 nm (a_{350}, m^{-1}) and DOC concentrations ($\mu\text{mol L}^{-1}$) across all samples along the permafrost-to-nearshore aquatic continuum. Global relationship between absorption coefficients at 350 nm (a_{350}, m^{-1}) and DOC concentrations (in $\mu\text{mol L}^{-1}$) along the permafrost to nearshore aquatic continuum. Note that the data is reported in Log units.

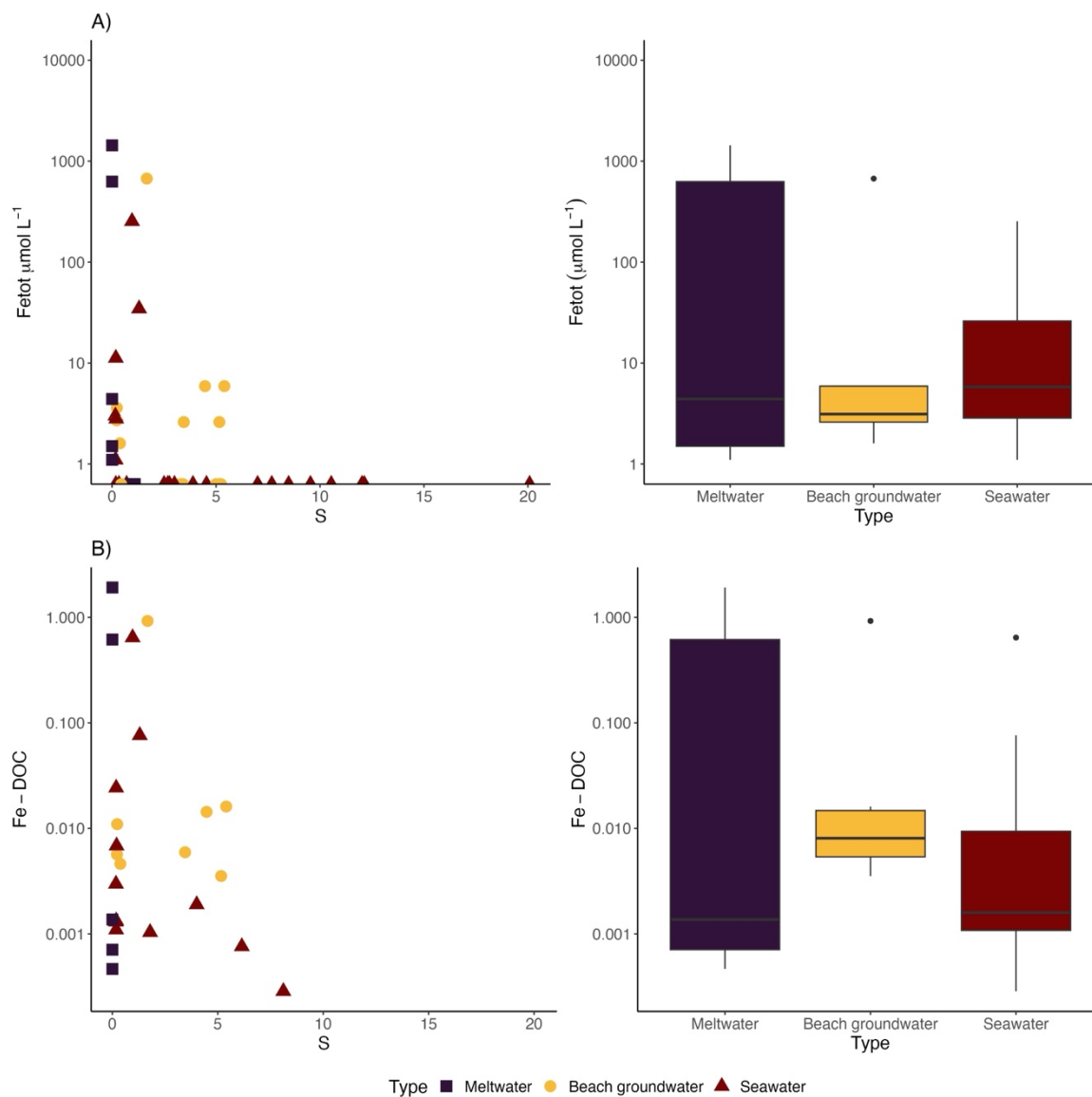


Fig. 7 Distribution of total dissolved Fe (Fetot, $\mu\text{mol L}^{-1}$) (A), and Fe-DOC ratio (B) across the salinity gradient (S), and by sample type. The y-axis in each panel is shown on a logarithmic scale to accommodate the broad range of Fe concentrations. Note that there is no Fetot data from the massive ice samples and Fe : DOC ratios were calculated only for Fetot concentration higher than the detection limit.

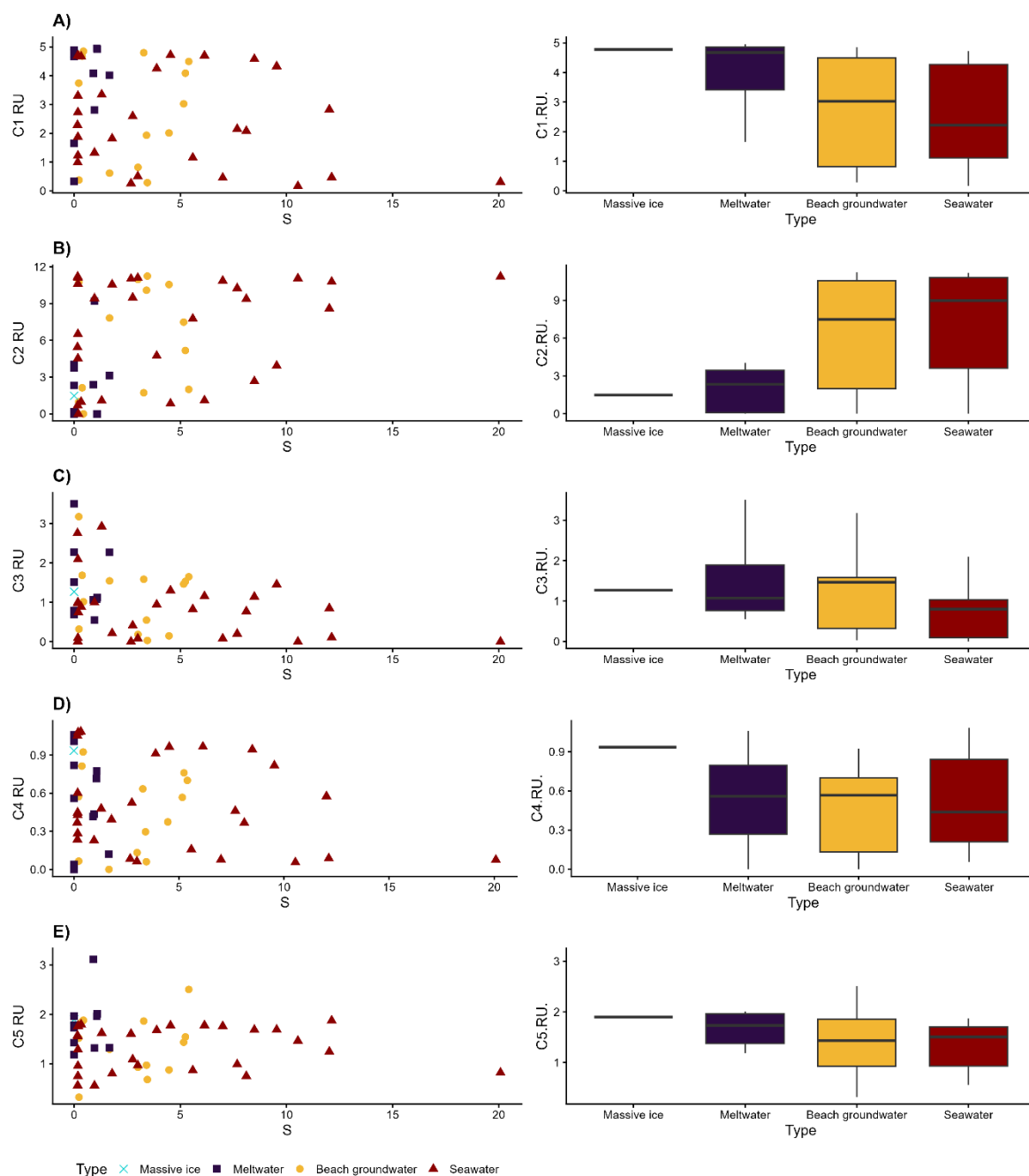


Fig. 8 Distribution of PARAFAC components (A) C1, (B) C2, (C) C3, (D) C4, and (E) C5 along the salinity gradient (S) and for the different types of collected samples (*i.e.*, beach groundwater, massive ice, meltwater and nearshore seawater samples). Note that only one massive ice sample is reported here. For the boxplots,

947 the black lines represent the median values, the whiskers represent the extent of the data, and the dot points
948 represent the outlier values. Note that there is only one massive ice data.

949