

Drivers of change in Peak Season Surface Ozone Concentrations and Impacts on Human Health over the Historical Period (1850-2014)

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Abstract. Elevated concentrations of ozone at the surface can lead to poor air quality and increased risks to human health. There have been large increases in surface ozone over the historical period associated with socio-economic development. Here the change in peak season ozone (OSDMA8) is estimated for the first time using hourly surface ozone output from 3 CMIP6 models over the 1850 to 2014 period. Additional results are obtained from one model to quantify the impact from different

- 5 drivers of ozone formation, including anthropogenic emissions of ozone and aerosol precursors, stratospheric ozone and climate change. The peak season ozone concentrations are used to calculate the risk to human health, in terms of the attributable fraction metric (the percentage of deaths from COPD associated with long-term exposure to elevated ozone concentrations). OSDMA8 concentrations are simulated to more than double across northern mid-latitude regions over the historical period, mainly driven by increases in anthropogenic emissions of NO_x and global CH₄ concentrations. Small contributions are made from changes
- 10 in other anthropogenic precursor emissions (CO and non-CH⁴ VOCs), aerosols, stratospheric ozone and climate change. The proportion of the global population exposed to OSDMA8 concentrations above the theoretical minimum risk exposure level (32.4 ppb), increased from <20% in 1855 to >90% in 2010. This has also increased the risk to human health mortality due to COPD from long-term ozone exposure by up to 20% across Northern Hemisphere regions in the present day. Like for OSDMA8 concentrations, the drivers of the increase in the ozone health risks are attributed mainly to changes in NO_x and
- 15 global CH₄. Fixing anthropogenic NO_x emissions at 1850 values can eliminate the risk to human health from long-term ozone exposure in the near present-day period. Understanding the historical drivers of ozone concentrations and their risk to human health can help to inform the dvelopment of future pathways that reduce this risk.

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1 Introduction

Ozone is a greenhouse gas in the troposphere and classified as a secondary air pollutant at the Earth's surface, influencing both the Earth's radiative balance and regional air quality (Szopa et al., 2021). It is formed by photochemical reactions involving 25 nitrogen oxides (NO_x) , carbon monoxide (CO) and volatile organics compounds $(VOCs)$, including CH_4) (Lelieveld and Dentener, 2000). Perturbations to the emissions associated with these precursors from both anthropogenic and natural sources can therefore alter net ozone production rates. In addition, the formation of ozone can also be impacted by geographical location and meteorological conditions such as temperature and photolysis rates (Monks et al., 2015). Furthermore, ozone concentrations at the surface are also affected by chemical destruction, surface deposition processes (Mills et al., 2018), hemispheric

- 30 transport (Liang et al., 2018) and the downward transport of ozone from the stratosphere (Stohl et al., 2003; Chen et al., 2024), stratosphere-troposphere exchange (STE). In air pollution episodes, exposure to elevated surface concentrations of ozone can lead to impacts on human health associated with respiratory diseases such as chronic obstructive pulmonary disease (COPD) (Pozzer et al., 2023). The Global Burden of Disease (GBD) study in 2019 estimated that long term exposure to concentrations of ozone resulted in 365,000 annual mortalities from COPD (Murray et al., 2020), with a recent revision by the GBD
- 35 2021 assessment to 490,000 annual mortalities in 2021 from COPD (GBD 2021 Risk Factors Collaborators, 2024). Therefore, ozone pollution represents a large current risk to the health of a population requiring an understanding of the factors controlling changes in concentrations.

Surface ozone concentrations are understood to have rapidly increased throughout the 20^{th} Century due to industrial de-40 velopment and increased anthropogenic emissions of ozone precursors. Even with the uncertainties surrounding early measurements of surface ozone at the end of the 19^{th} and beginning of the 20^{th} Century, ozone concentrations are observed to have increased across the temperate and polar regions of the Northern Hemisphere by 30 to 70% (approximately 10 to 16 ppb) since the start of the 20^{th} Century (Tarasick et al., 2019). Previous multi-model inter comparison exercises have simulated the historical change in annual or seasonal mean tropospheric and surface ozone concentrations. In both the $5th$ and $6th$ phase of

- 45 the Coupled Model Intercomparison Project (CMIP5 and CMIP6), the historical change in annual mean global surface ozone concentrations from 1850 to 2000 was simulated to be in agreement at 10 to 11 ppb (Young et al., 2013; Turnock et al., 2020). Simulated historical changes in annual mean surface ozone were larger at 15 to 20 ppb across northern hemisphere mid latitude regions in both CMIP5 and CMIP6, although some regions like south and east Asia showed large inter-model diversity. However, in both phases of CMIP the simulated historical changes in surface ozone concentrations were in part dominated by the
- 50 large uncertainty in the simulated pre-industrial values of 1850 (Young et al., 2013; Turnock et al., 2020). Evaluating historical

changes in surface ozone concentrations is limited by the availability and reliability of observations with a sufficiently long record. Sufficient observation data is available over the last five to six decades across regions within the northern mid-latitudes (North America, Europe and Japan) and shows that both CMIP5 and CMIP6 models underestimated the recent rapid increase in surface ozone concentrations that occurred over these regions (Young et al., 2018; Parrish et al., 2014, 2021). However, the 55 CMIP6 model simulated long-term change in ozone concentrations shows good agreement when compared to the observed trend at only remote observation locations (Griffiths et al., 2021). Evaluation of the present-day simulation of ozone concentrations in CMIP5 (2000) and CMIP6 (2005 to 2014 mean) shows that both generations of chemistry climate models are able to represent the spatial and seasonal distributions of ozone but overestimate the absolute concentrations at the surface compared

to ground-based observations collected as part of the Tropospheric Ozone Assessment Report (TOAR) (Young et al., 2018; 60 Turnock et al., 2020).

Dedicated model simulations allow for the studies focusing on the drivers of historical ozone changes. Generally, the most important drivers of historical ozone changes have been increases of anthropgenic emissions, especially emissions of NO_x (e.g. Lelieveld and Dentener, 2000; Fiore et al., 2009; Young et al., 2013; Turnock et al., 2018), with different emission sectors

- 65 having different effects on ozone due to different trends and ozone production efficiency (Dahlmann et al., 2011; Mertens et al., 2024). Besides increases in NO_x emissions, the increase of methane is also an important driver (Wild et al., 2012; Iglesias-Suarez et al., 2018; Morgenstern et al., 2018). Moreover, changes in aerosols (Xing et al., 2017), emissions of N_2O and ozone depleting substances (ODS) drive changes in tropospheric ozone (Morgenstern et al., 2018; Iglesias-Suarez et al., 2018; Zeng et al., 2022). Changes in climate via temperature, water vapor and radiation, can also influence surface ozone con-
- 70 centrations with increases and decreases generally anticipated in polluted and unpolluted regions respectively (Schnell et al., 2016; Fortems-Cheiney et al., 2017; Fu and Tian, 2019). Uncertainties of various processes, such as description of natural emissions which differ in various models, however, hinders a consistent quantification of ozone drivers among various models (Archibald et al., 2020a).
- 75 Until lately, global estimates of ozone-attributable health impacts were based on outputs from global atmospheric-chemistry models providing surface ozone concentrations in sufficient spatio-temporal distribution for ozone exposure assessments (Anenberg et al., 2010; Lim et al., 2012; Lelieveld et al., 2015; Fang et al., 2013; Silva et al., 2013; Forouzanfar et al., 2015; Silva et al., 2016; Cohen et al., 2017; Malley et al., 2017; Liang et al., 2018; Lelieveld et al., 2020). However, as part of the first phase of TOAR, Fleming et al. (2018) provided an assessment of recent changes in relevant ozone health metrics calculated
- 80 from observations, showing recent reductions since 2000 across northern mid latitudes regions (North America, Europe and East Asia). More recently, the synergistic use of in-situ ozone measurements and model outputs through data fusion have made it possible to assess ozone-related health effects (Murray et al., 2020; Malashock et al., 2022) based on observational-based gridded ozone data sets (DeLang et al., 2021; Becker et al., 2023). Over the previous years different ozone exposure metrics were applied, according to the adopted coefficients from relevant cohort studies. The exposure response function (ERF) for
- 85 ozone is actually a mathematical formula expressing the relative risk (RR) of a disease as a function of ozone abundance.

Pope et al. (2002) and Ostro et al. (2004) introduced a log-linear ERF based on epidemiological findings, while later a log-log ERF approach was proposed Cohen et al. (2005). In the GBD 2019 report (Murray et al., 2020) a log-linear ERF was used, introducing an ozone exposure metric based on the ozone season daily maximum 8-hour mixing ratio (OSDMA8). Malashock et al. (2022), using the observation-based gridded OSDMA8 data set of DeLang et al. (2021) and the GBD2019 approach 90 for mortality calculations, reported 423,000 ozone-related deaths worldwide for the year 2019, accounting for all respiratory diseases. The review by Pozzer et al. (2023) showed that global estimates of mortality from long-term exposure to ozone have a large range from 142,000 per year from COPD to 1.3 million per year from all respiratory diseases. Although there are several studies on the projected changes of ozone-related health effects under different climate change and demographic scenarios (West et al., 2007; Silva et al., 2016; Turnock et al., 2023; Pozzer et al., 2023; Akritidis et al., 2024), studies on the health

95 effects of ozone during the historical period are lacking. In particular, few studies have quantified the change in risk to human health and the drivers behind this change over this period.

Here we make use of hourly mean surface ozone concentrations from three CMIP6 models that conducted historical transient experiments; a time resolution that has not been made available in previous global multi-model inter comparison exercises. 100 This has allowed us to calculate an ozone metric relevant to human health (OSDMA8) and explore the simulated changes in

concentrations and impacts on human health over the entire CMIP6 historical period (1850 to 2014). In addition, we also use output from sensitivity experiments conducted by one of the CMIP6 models to explore the impact of changes in the different drivers of ozone formation (precursor emissions, stratospheric ozone, aerosols, climate) on the simulated concentrations and how this relates to changes in the risk to human health from long-term exposure to ozone over the same time period.

105 2 Methods

2.1 CMIP6 Model Data

We use output from the histSST experiments conducted by three different CMIP6 models as part of the Aerosols and Chemistry Model Intercomparison Project (AerChemMIP), a CMIP6 endorsed Model Intercomparison Project (Collins et al., 2017). The histSST experiments were designed to be an atmosphere-only representation of the coupled historical experiment conducted 110 by individual models using historical forcing data over the period 1850 to 2014. The three models used here, UKESM1-0- LL (Sellar et al., 2019, 2020), GFDL-ESM4 (Horowitz et al., 2020; Dunne et al., 2020) and EC-Earth3-AerChem (van Noije et al., 2021), are all global Earth system models with horizontal grid areas coarser than 100 km. All three models include an interactive representation of chemistry and aerosols, with various different couplings and feedback within the Earth system e.g. land surface, radiation etc. They are therefore suitable to use for simulating the long-term response to surface ozone from 115 perturbations to different drivers of ozone formation, along with any feedbacks from changes in climate.

Results have been obtained from UKESM1-0-LL for all the other transient historical sensitivity experiments, as this was the only model to complete all of the additional experiments requested as part of AerChemMIP (Collins et al., 2017) and thus

Table 1. Experiment configuration of the historical sensitivity experiments used in this study. The 3 CMIP6 models EC-Earth3-AerChem, GFDL-ESM4 and UKESM1-0-LL all provide data for histSST whereas, only UKESM1-0-LL provides data for the other experiments. Methane is prescribed as a global concentration value in these experiments.

provide hourly surface ozone concentrations. These additional sensitivity experiments involved running the historical transient 120 simulation but with different short lived climate forcers fixed at pre-industrial values (following an "all but one" methodology), including CH_4 , NO_x , CO , non- CH_4 VOCs and aerosols. A separate experiment was also conducted to consider the impact of ozone depleting substances (ODS - chlorofluorocarbons and hydrochlorofluorocarbons) on stratospheric ozone by fixing their concentrations at 1950 values. An additional experiment to those specified in AerChemMIP was conducted to examine the influence of historical climate change, where the underlying climate conditions (sea surface temperatures and sea ice cover) 125 were kept fixed at 1850 values throughout the whole historical time period. A full list of the experiments analysed here is

presented in Table 1.

Hourly mean surface ozone concentrations were obtained from each of the experiments for a single model realization over the entire historical period (1850-2014). These values were then used to calculate the relevant peak season ozone heath metric 130 (OSDMA8) by first calculating the daily maximum of 8 hour running mean values over a 24 hour period and then a 6 month running mean is calculated from these values. An annual maximum is then calculated to represent the seasonal maximum daily exposure values (avoiding biases from different peak seasons across the world), which is consistent with the metric used in the current WHO air quality guideline values (https://apps.who.int/iris/handle/10665/345329). Regional mean OSDMA8 values have been calculated for the 21 regions used in the GBD study (region definitions shown in Figure A1).

135

Surface ozone concentrations are simulated at a relatively coarse resolution (>100km) by the CMIP6 models used in this study, which can pose problems when assessing the associated impacts on air quality and human health. Additionally, previous evaluations have shown that global chemistry climate models tend to overestimate surface ozone concentrations (Young et al., 2018; Turnock et al., 2020). To provide a more accurate representation of surface ozone concentrations, which is nec-

- 140 essary when linking concentrations to health effects from long-term exposure, the concentrations simulated by each model in the present day period (2005-2014) have been corrected for biases against the Regionalized Air Quality Model Performance (RAMP) dataset (Becker et al., 2023). The RAMP dataset provides a high-resolution $(0.1° \times 0.1°)$ global surface ozone dataset from 1990 to 2017 by using state-of-the-art geostatistical methods to integrate surface ozone observations with output from multiple atmospheric chemistry models (including contributions from the GFDL-AM4 model - the atmosphere model version
- 145 of GFDL-ESM4). First, each model is regridded to the finer spatial resolution of the RAMP dataset. We then use the last 10 year mean period (2005-2014) of the historical experiments from the CMIP6 models to bias correct against the same time period in the RAMP dataset, which maximises the availability of ground-based observations in the RAMP dataset. Figure 1 shows the 10-year mean (2005 to 2014) simulated OSDMA8 values from each CMIP6 model in the histSST experiment and also the difference versus the RAMP dataset in the same time period. EC-Earth3-AerChem shows a consistent overestimation
- 150 of OSDMA8 concentrations across all regions (up to 20 ppb), whilst GFDL-ESM4 and UKESM1-0-LL have smaller underestimations and overestimations of OSDMA8 values across different regions. Applying this correction to each model for the 2005 to 2014 10-year mean period provides a present-day climatological baseline for each model that can be used as a baseline for long term changes and also in the calculation of the human health response to exposure, as the simulated concentrations align with observations now as much as possible and the known biases in surface ozone concentrations have been reduced. Staehle
- 155 et al. (2024) discuss the merits of four different bias correction techniques including: mean bias, relative bias, delta correction and quantile mapping, finding that using delta correction is favourable due to its performance and numerical simplicity. Therefore, we use the delta correction method (Staehle et al., 2024) to calculate the change in historical ozone concentrations by applying the ratio of change between each decadal mean and the present day (2005-2014) on top of the bias corrected baseline for each model. This then generates a new historical time series for each experiment that is connected to the observed present
- 160 day values. The temporal changes in OSDMA8 values (from HistSST) are calculated by comparing each 10-year mean period to the 2005-2014 baseline period. Whereas, the impact of the historical drivers (sensitivity experiments) is assessed by comparing the 10-year mean OSDMA8 values in the histSST experiment to the same time period in each of the sensitivity simulations.
- A feature of the histSST sensitivity experiments conducted by UKESM1-0-LL is that they use prescribed values of global 165 CH₄ concentrations, which are unable to respond to any other perturbation of the inputs (e.g., fixing NO_x emissions at 1850) values), in the same way as a CH_4 emission-driven model would. This experimental set up therefore neglects the impacts of any of the sensitivity experiments on global CH_4 concentrations and also consequently ozone, as CH_4 is a precursor to ozone formation. Therefore, in each of the histSST sensitivity experiments we have adjusted the ozone concentrations to take account for the consequences of changes in the CH₄ lifetime (from the feedback on its own concentrations). Firstly, a global CH₄ 170 lifetime is calculated for each experiment, including histSST. The CH⁴ feedback factor (Prather, 1996) is then calculated by using the difference in the CH₄ lifetime and CH₄ concentrations in the histSST and histSST-piCH4 experiments over the 30year period 1930 to 1960, when large changes in CH₄ concentrations occur but there are smaller influences from other factors, e.g., halocarbons (Stevenson et al., 2020). The CH⁴ feedback factor is calculated to be 1.30 over this 30-year period, which is similar to other previous estimates (O'Connor et al., 2021; Stevenson et al., 2013). The feedback factor and CH₄ lifetime can

Figure 1. 10 year mean surface OSDMA8 values (2005 to 2014) simulated by 3 CMIP6 models a) EC-Earth3-AerChem, b) GFDL-ESM4 and c) UKESM1-0-LL and difference in the same models d) to f) when compared to the RAMP observational dataset (Becker et al., 2023) over the same time period.

- 175 then be used to calculate the equilibrium CH_4 concentrations in each experiment from the transient prescribed values used. The difference in prescribed and equilibrium CH_4 concentrations can be used as input to the relationship between CH_4 perturbations and ozone response derived by Wild et al. (2012). Using this relationship, we calculate for each sensitivity experiment the response in surface ozone concentrations that occurs due to the change in CH⁴ concentrations resulting from the adjustment of CH⁴ lifetime. The model simulated ozone concentrations are then adjusted accordingly in each of the histSST sensitivity
- 180 experiments for this change. These ozone values adjusted for the impact of CH⁴ lifetime can then be used to calculate the historical changes in surface ozone values relative to the bias-corrected baseline period.

2.2 Emissions in Historical Scenarios

- There have been large changes in emissions of ozone precursors (CO, NO_x , non-CH₄ VOCs and CH₄) since 1850 due to 185 increasing human industrial activity and economic development (Hoesly et al., 2018). Figure 2 shows the relative change (compared to 1850 values) in these ozone precursors that are used as input to the CMIP6 historical experiments and sensitivity experiments considered in this study (Table 1). Global anthropogenic emissions of N_{α} have seen the largest increase of the ozone precursor emissions, a factor of >11 increase globally since 1850. Global emissions of non-CH⁴ VOCs and CO, along with global CH₄ concentrations have all increased globally by a factor of >2 since 1850. These large changes will all have 190 a substantial effect on surface ozone concentrations throughout the historical period considered in the CMIP6 experiments (1850 to 2014). Additionally, changes in climate and stratospheric ozone concentrations over the historical period will also have an influence on surface ozone concentrations; the changes in these are considered in the sensitivity experiments in Table 1. The transient change in surface air temperature simulated by UKESM1-0-LL over the historical period (Fig. 2 right panel)
- showed particularly large cold biases compared to observations throughout the latter half of the 20th Century (peak of ≈0.5k), 195 which was similar to a number of other CMIP6 models (Flynn and Mauritsen, 2020). However, in 2014 the global annual mean surface temperature in HadCRUT5 observations (Morice et al., 2021) and UKESM1-0-LL simulations (Fig. 2) both increased by a similar amount of \approx 1K (compared to a 1850 to 1900 mean period), mainly due to increasing concentrations of long-lived greenhouse gases (Chen et al., 2021). Total column ozone burdens have shown large reductions since the 1960s in UKESM1-0- LL historical simulations due to the reductions in stratospheric ozone from increases in ozone depleting substances, e.g. CFCs.
- 200 However, total column ozone simulated by UKESM1-0-LL in CMIP6 simulations tends to be higher than other CMIP6 models and observations, and also simulates a stronger depletion of stratospheric ozone from 1960 to 2000 (Keeble et al., 2021). All of these drivers of surface ozone are used as transient changes to the inputs for the histSST experiment and are also fixed at 1850 values (or 1950 for ODS) for the individual sensitivity experiments.

205 2.3 Health Calculations

To assess the health impacts from long-term exposure to ozone concentrations during the historical period, we use the attributable fraction (AF) metric, which is defined here as the percentage of deaths from chronic obstructive pulmonary disease (COPD) attributable to ozone pollution. A full health assessment is not undertaken due to the absence of long term datasets for changes in population (on a latitude longitude grid) and country specific baseline mortality rates over the 1850 to 2014 period.

210 In more detail, $AF(x,y,t)$ is calculated at a given location with x and y coordinates and for a specific time period t, as shown in equation (1):

$$
AF(x, y, t) = \frac{RR(x, y, t) - 1}{RR(x, y, t)}
$$
\n
$$
(1)
$$

where $RR(x,y,t)$ is the relative risk (also known as hazard ratio) for ozone-related excess mortality from COPD. Following the 215 GBD 2019 methodology (Murray et al., 2020), RR is estimated using a log-linear function as shown in equation (2):

Figure 2. Relative change in total historical emissions of ozone precursors (NO_x , CO and NM (non-CH₄)VOCs) and global CH₄ concentrations compared to 1850 values (left panel). Global annual mean surface air temperature anomaly, relative to an 1850-1900 mean period, from UKESM1-0-LL historical simulations and HadCRUT5 (Morice et al., 2021) (blue lines, right panel). Global annual mean total column ozone values from 1850 to 2014 simulated by UKESM1-0-LL (green line, right panel).

$$
RR(x, y, t) = e^{\beta(X(x, y, t) - TMREL)}
$$
\n(2)

where $X(x,y,t)$ is the health-relevant ozone metric (OSDMA8) at the given location and time period and TMREL is the theoretical minimum risk exposure level, below which ozone is considered as not being harmful for human health. For TMREL we 220 adopted the value of 32.4 (29.1–35.7) ppb, while the β parameter value(s) result from the RR per 10 ppb of ozone (OSDMA8) value of 1.063 (1.029–1.098), estimated from a meta-regression of five cohorts studies (Murray et al., 2020).

Since ozone concentrations (OSDMA8) are the only varying input in the AF calculations, the AF temporal changes over the historical period reflect the changes of the percentage of ozone-related COPD excess mortality from ozone variations only. The 225 ozone exposure values were calculated as 10-year mean OSDMA8 values across the historical period from the bias-corrected model simulation data. An assessment of population exposure to ozone is also provided by calculating the number of people exposed to OSDMA8 concentrations concentrations above the TMREL in 3 different time periods (1850, 1980 and 2010).

Population data for these periods is obtained from Hyde (History database of the Global Environment) (Klein Goldewijk et al., 2017).

230 3 Results

3.1 Multi-model Historical Changes in Surface Ozone and Risk to Human Health

3.1.1 Multi-model Historical Changes in OSDMA8

The regional mean change in OSDMA8 concentrations over the 1850 to 2014 period from the 3 bias-corrected CMIP6 models is shown in Figure 3. Large changes in OSDMA8 concentrations have been simulated across all regions over the historical 235 period. Globally, the 10-year multi-model mean (+/- 1 S.D. of 3 model values) OSDMA8 concentrations are simulated to have increased by 12 +/- 2.6 ppb (50% increase) from 1855 to 2010, which is of similar magnitude to annual mean changes simulated by 6 CMIP6 models (Turnock et al., 2020). The largest regional mean increases over the same time period have been simulated to occur over South Asia (29.8 +/- 5.3 ppb, 115%), East Asia (24.8 +/- 2.2 ppb, 88%), High-income Asia Pacific (24.8 +/- 0.4 ppb, 110%), North Africa and Middle East (21.4 +/- 4.8 ppb, 78%), Central Europe (21.3 +/- 3.9 ppb, 90%) and High-income 240 North America (18.6 +/- 1.9, 81%). Previously, CMIP6 models were shown to be able to represent the changes in surface ozone concentrations since 1960 at five long-term monitoring remote locations around the world (Griffiths et al., 2021), providing confidence in the ability of models to simulate long-term changes. Across most regions OSDMA8 concentrations only show small increases up until about 1950. After this time period concentrations rapidly increase (more than doubling) to 2005-2014 values, driven by the large rapid changes in all anthropogenic precursors of ozone $(CH_4, NO_x, CO, Non-CH_4 VOCs)$, with the 245 largest relative changes occuring in NO_x emissions (Fig. 2). The increase in concentrations continued across certain regions, most notably South Asia, East Asia and other tropical and sub-tropical regions, due to the continued increases in anthropogenic ozone precursors from socio-economic development across these regions (Figure A3). However, across some regions in the northern mid-latitudes (high-income North America, Europe and high-income Asia Pacific) OSDMA8 concentrations stopped increasing from about the 1980s, remaining at and near these values until the end of the CMIP6 historical period when a slight 250 reduction in concentrations occurred, in agreement with the observations in the RAMP dataset (Fig. 3). The recent change in OSDMA8 concentrations is consistent with the mitigation measures adopted over these regions to improve regional air quality

by reducing primary air pollutant emissions (United Nations Economic Commission for Europe, 2004; EMEP Steering Body and Working Group on Effects of the Convention on Long-Range Transboundary Air Pollution, 2016), shown by a reduction in NO_x emissions from the peak in the 1980-1990s (Figure A3).

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Fig. 3 shows that bias correcting the model data to the RAMP dataset has resulted in the simulated OSDMA8 concentrations from all 3 models being similar in the 2005 to 2014 10-year mean period. However, there is still a diversity of 10 to 15% in the model simulated OSDMA8 concentrations in the pre-industrial period (1850 to 1860), with some of the largest model differences (20%) occurring over South Asia. UKESM1-0-LL consistently simulates larger bias-corrected OSDMA8 concentrations

- 260 across most regions in the first 10 year mean period of 1855, with GFDL-ESM4 tending to have the lowest concentrations of the 3 models. This diversity, along with differences in the chemical sensitivity of ozone formation in each model, results in the different simulated historical changes in OSDMA8 across the models (Turnock et al., 2020). Given its higher 1855 concentrations, UKESM1-0-LL is the model that tends to simulate the smallest historical change in OSDMA8 values over the historical period (9.2 ppb globally), whilst GFDL-ESM4 simulates the largest (14.3 ppb globally). If the models are not corrected against
- 265 the observation based dataset then the simulated concentrations show on average a larger 20% variability between models (up to 30% across South Asia), with the largest diversity in concentrations in the 1855 period. The comparison to the RAMP dataset in the 2005 to 2014 period shows that simulated OSDMA8 values tend to be consistently overestimated by all 3 models (Fig. 1, Figure A2). Overestimating OSDMA8 concentrations has important consequences for the assessment of the long-term impact on human health due to the use of theoretical minimum risk exposure level, below which no risk to health is considered
- 270 (Section 2.3). Using the uncorrected simulated OSDMA8 concentrations from the models could result in an overestimation of the impacts to human health from long-term exposure, and thus the bias-corrected model simulated concentrations are used for any assessment of health impacts.

3.1.2 Historical Changes in Ozone exposure and risk to Human Health

- 275 The increase in OSDMA8 concentrations over the historical period across all regions has increased the risk to human health via enhanced long-term exposure to ozone. Figure 3 shows how the simulated regional OSDMA8 concentrations compare with the TMREL of 32.4 ppb. All three models simulate concentrations in first 10-year mean period of 1855 that are below this threshold across most regions of the world, resulting in <20% of the world's population in 1850 being exposed to concentrations above this value (Fig. 4a). However, the simulated increase in OSDMA8 concentrations over the historical period has resulted in con-280 centrations being above the TMREL value in almost every region of the world in the near present-day period. This is shown by >90% of the world's population in 2010 being exposed to ozone concentrations above this value, meaning that over the historical period there has been a large increase in the risk to human health from long-term exposure to elevated ozone concentrations.
- Figure 5 shows the spatial distribution of the multi-model mean attributable fraction over the pre-industrial period 1850-1859 285 and the near present-day period 2005-2014. High AF values up to \approx 20% are found over regions of North India, East China, Middle East, and Western Unites States, indicating that across these regions approximately one out of five COPD deaths in the near present-day period are attributed to long-term ozone exposure. As the AF calculation does not consider changes in population and baseline mortality rates, a distinct increase of the AF is depicted almost globally compared to the 1850-1859 period due entirely to increases in ozone concentrations. As ozone concentrations in the pre-industrial period are below the 290 TMREL value almost all over the world, the respective AF values are near zero. This suggests that there is no risk for human
- health from ozone exposure in the pre-industrial period, although the increases in ozone concentrations up to the near presentday became a major threat for human health, being responsible for more than 10% of COPD deaths in many regions of the Northern Hemisphere.

Figure 3. Regional mean surface OSDMA8 values from 3 CMIP6 models simulated over the historical period (1850-2014) bias-corrected to a 10-year mean of observations from Becker et al. (2023). Observations from Becker et al. (2023) are shown as circles on each relevant region with the TMREL value of 32.4 ppb as the dashed line.

3.2 Drivers of Historical Change in Surface Ozone and Risk to Human Health

295 3.2.1 Drivers of Historical Change in OSDMA8

Figure 6 shows the change in OSDMA8 concentrations in the present day time period (2005 to 2014) due to the historical changes in the individual drivers of ozone formation i.e. histSST minus sensitivity experiment. Considering only historical changes in emissions of ozone precursors (i.e., histSST minus histSST-piO3 = NO_x , CO and non-CH₄ VOCs) has resulted in a large increase in global mean OSDMA8 concentrations of 9.6 ppb (37%) in the 2005 to 2014 mean period. Of these three

300 precursors, the majority of this change in global mean OSDMA8 concentrations in 2005-14 can be attributed to historical changes in anthropogenic NO_x emissions (8.6 ppb, 32%), compared to CO and non-CH₄ VOCs (1.5 ppb, 4%). Regionally, the

Figure 4. Percentage of total global population in exceedance of the theoretical minimum risk exposure level (TMREL - 32.4 ppb) to ozone across 3 historical time periods from 3 different CMIP6 models a) and in the different present day (2010) for sensitivity scenarios with different drivers of ozone fixed b). Total global population for each time period is shown in parenthesis on a).

largest changes in OSDMA8 concentrations occur mainly over the polluted continental regions of the northern mid-latitudes. OSDMA8 concentrations more than doubled (\approx 25 ppb) over the High-income Asia Pacific region, due to this region having the largest relative increase in anthropogenic NO_x emissions over the historical period (Figure A3). Similarly large increases 305 in OSDMA8 concentrations occur over Europe (16 ppb), High-income North America (16 ppb), East Asia (23 ppb) and South Asia (24 ppb), all mainly due to the large historical changes in anthropogenic NO_x emissions over these regions (Figure A3).

Historically, global CH₄ concentrations, the other major precursor gas to ozone formation, have increased by a factor of \approx 2 since 1850 (Fig. 2), which has increased global mean OSDMA8 concentrations by 5.9 ppb (20%) in the 2005 to 2014 period.

310 The impact of changes in CH₄ on surface ozone concentrations tend to be more globally uniform due to the longer chemical lifetime of $CH₄$ and that it is input to these model experiments as a global annual mean concentration, rather than gridded emissions. The largest increase in OSDMA8 concentrations occurred over parts of Asia and North Africa and Middle East (up to 8.5 ppb, 21%), whilst the smallest increase (2 to 3 ppb, 10 to 13%) occurred over tropical regions of the southern hemisphere.

Figure 5. Attributable fraction as the percentage of deaths from COPD attributable to ozone pollution calculated from the multi-model mean OSDMA8 values in the 10 year mean periods of 1850-1859 (top) and 2005-2014 (bottom).

Over some of these more remote (low NO_x) regions like sub-Saharan Africa and southern Latin America, historical changes in 315 CH₄ had a larger proportional impact on OSDMA8 concentrations, contributing almost as much as changes in NO_x emissions. This shows that historical changes in CH_4 concentrations have been important in altering the background ozone concentrations in most regions of the world.

The other drivers of surface ozone formation considered here in the sensitivity experiments are historical changes in cli-320 mate, aerosols and stratospheric ozone via ODS, which have all tended to have smaller impacts than anthropogenic ozone precursor emissions (Fig. 6 and Figure A4 for more detail). The increasing emissions of ODS since the 1950s reduced stratospheric ozone concentrations (shown as total ozone column on Fig. 2) significantly by the start of the 21^{st} Century, which has also led to reductions in OSDMA8 concentrations, potentially via less downward transport to the surface or from changes to photolysis rates. Global mean OSDMA8 concentrations reduced by 2.8 ppb (7%) in the 2005 to 2014 period due to smaller 325 amounts of stratospheric ozone. Some of the largest impacts (>2 ppb) occurred over remote southern hemisphere regions, including southern Latin America and Australasia (Figure A4). In addition, changes in stratospheric ozone also reduced OS-DMA8 concentrations across the well known hot spots regions of stratosphere-to-troposphere transport (STT) (Škerlak et al., 2014; Akritidis et al., 2021) of high altitude Central Asia, the Eastern Mediterranean and Middle East, and the Western United

States.

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Increasing emissions of aerosols and aerosol precursors over the historical period (Figure A5) has resulted in a small reduction of global mean OSDMA8 concentrations of 0.8 ppb (2%) in the 2005 to 2014 period (see Figure A4 for more detail). Increasing aerosols over the historical period implies that there is an increase in the heterogeneous loss of nitrogen oxides to aerosol surfaces (the only active heterogeneous tropospheric mechanism currently in UKESM1-0-LL (Archibald et al., 2020b)), 335 leading to a reduction in ozone formation. This effect is probably underestimated due to the absence of the loss mechanism involving HO² uptake on aerosols in UKESM1-0-LL (Ivatt et al., 2022). Regionally, increasing aerosols causes the largest reduction in OSDMA8 concentrations of up to 2.1 ppb (4%) across Asia (central, south and east), North Africa and Middle East regions in the 2005 to 2014 period when aerosol concentrations were higher. However, the largest reduction on OSDMA8 concentrations from aerosols (2.7 ppb, 6%) occurred in the 1980s across Europe. Aerosol concentrations peaked in Europe at 340 this time due to the subsequent enacting of air pollutant controls, meaning aerosol concentrations and their impact on surface ozone were reduced thereafter (Turnock et al., 2015).

The climate change signal over the historical period has resulted in an approximate 1K increase in global mean surface temperatures simulated by UKESM1-0-LL by the 2005 to 2014 period (Fig. 2), which has resulted in a small reduction in global 345 mean OSDMA8 concentrations of 0.8 ppb (2%). The global mean reduction in ozone has mainly been driven by decreases over the ocean, which can be attributed to more ozone destruction here due to increased hydroxyl formation from enhanced amounts of water vapour occurring in a warmer world (Johnson et al., 1999; Doherty et al., 2013; Zanis et al., 2022). Historical climate change has resulted in enhanced ozone formation rates over some polluted continental regions including including Southeast Asia, Central Europe, east Asia, high-income Asia Pacific and parts of North America (see Figure A4 for more detail), which 350 is shown by a small increase of OSDMA8 concentrations (<1 ppb, 2%). There has also been a similar small increase in OS-DMA8 concentrations across some of major biogenic emission regions (Central Sub-Saharan Africa and parts South America), suggesting a small climate impact on natural sources of ozone precursor emissions. The magnitude of the changes in ozone due to 1K of warming over the historical period are consistent with studies analysing the change of ozone in response to future warming (Archibald et al., 2020c; Zanis et al., 2022), with the impact of climate change on surface ozone being less in this

355 study due to the smaller level of historical warming compared to those projected for the future.

In summary, the model experiments isolating the impact of the individual drivers shows that historical changes in anthropogenic NO_x emissions contribute the most, approximately 42%, to the summed total change in historical global mean OS-DMA8 concentrations from all drivers combined. The contribution to historical changes in global mean OSDMA8 concen-360 trations from other drivers was 29% from global CH⁴ concentrations, 14% from stratospheric ozone, 7% from anthropogenic emissions of CO and non-CH₄ VOCs, and 4% from both aerosols and climate change. However, the linear combination of the historical change in global mean OSDMA8 concentrations, calculated from the individual driver experiments, is found to be about 20% larger than that calculated when all drivers are varied simultaneously (histSST). This indicates that there are

Figure 6. 10-year mean in surface OSDMA8 values simulated by UKESM1-0-LL in the HistSST experiment for 2005-14 (bias-corrected) and the change in the same time period in the HistSST experiment relative to each of the sensitivity experiments with different drivers of ozone fixed

important non-linearities in the formation of ozone from the interactions between the different drivers.

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3.2.2 Drivers of change in Ozone Exposure and risk to Human Health

The change in the individual drivers of ozone formation show different impacts of OSDMA8 concentrations over the historical period, which will also have an impact on the risk to human health via changes to long-term exposure. Figure 4b) shows a comparison of the different fraction of the global population in 2010 that are exposed to OSDMA8 concentrations above the 370 TMREL of 32.4 ppb in each of the sensitivity scenarios that isolate the different drivers of ozone formation. If emissions of

ozone precursors (NO_x , CO and non-CH₄ VOCs) are kept fixed at 1850 values then the fraction of the near present-day global population exposed to OSDMA8 concentrations above 32.4 ppb is reduced substantially to 27%, from >90% when emissions

are allowed to increase in histSST. Fixing solely anthropogenic NO_x emissions at 1850 is shown to be the dominant cause of this reduction by decreasing the global population exposure above the TMREL to 34%. Fixing the emissions of the other 375 precursors (CO and non-CH⁴ VOCs) at 1850 values results in a very small (2%) change in global population exposure. A relatively small impact on global population exposure is shown when fixing global CH₄ concentrations at 1850, which reduces the proportion of the global population exposed to OSDMA8 concs above TMREL to 83% (from >90%). Fixing the other historical drivers of ozone concentrations (climate change, stratospheric ozone and aerosols) slightly increases the global population exposure to OSDMA concentrations above the TMREL by 1 to 2%. These results show that historical changes in 380 anthropogenic NO_x emissions are the leading cause of increased ozone exposure that is considered harmful to human health.

Figure 7 presents the regional mean attributable fraction for the histSST and the sensitivity experiments isolating the different drivers of ozone formation for the near present-day period 2015-2014. Ozone pollution in histSST is responsible for more than 10% of COPD deaths in South and East Asia; while percentages in other regions are: North Africa and Middle East (9%), 385 High-income Asia Pacific (9%), Central Europe (7%), Western Europe (6%) and High-income North America (5%). As for the drivers of ozone health effects, NO_x emissions in the historical period are found to be the main contributor to ozone-related mortality for the 2005-2014 period, as fixing them to pre-industrial levels (histSST-piNOx) eliminates the risk for human health from ozone exposure in the majority of regions. Historical increases in methane concentrations also appear to be a significant contributor to near present-day AF both globally and regionally. Setting methane concentrations to pre-industrial 390 levels reduces ozone-related mortality in the highly populated regions of South Asia, East Asia and Western Europe by $\approx 4\%$ and that of High-income North America by $\approx 3\%$. Historical changes in climate, aerosol precursor emissions and stratospheric ozone have relatively small impacts on regional mean near present-day AF. Isolating these drivers at 1850 levels (or 1950 for stratospheric ozone) shows a small ($\approx 1\%$) increase in the risk to ozone-related mortality.

4 Conclusions

- 395 Surface ozone is an important greenhouse gas and secondary pollutant in the lower atmosphere, influencing both the Earth's climate and air quality. Exposure to elevated concentrations of ozone at the surface can have detrimental impacts on human health. In this work we use numerical results of hourly surface ozone, obtained from the latest generation of global chemistry climate models that participated in CMIP6, as part of the Aerosol and Chemistry Model Intercomparison Project (AerChem-MIP). We use output from three different CMIP6 models, over the historical period (1850 to 2014), to calculate changes in
- 400 the peak season ozone metric (OSDMA8), that can relate risks to human health from long-term exposure. All three CMIP6 models tended to overestimate the recent observed OSDMA8 concentrations, in agreement with other recent multi-model evaluations of surface ozone concentrations (Young et al., 2018; Turnock et al., 2020). The most recent 10-year mean period (2005 to 2014) of OSDMA8 concentrations simulated by the three CMIP6 models were bias-corrected to the observations to avoid overestimating the present-day risk to human health from long-term ozone exposure. However, there is still a 10 to 15%
- 405 difference in the simulated OSDMA8 concentrations in the pre-industrial period (1850 to 1859) due to physical and chemi-

Figure 7. Regional mean attributable fraction as the percentage of deaths from COPD calculated from the 10 year (2005-14) mean OSDMA8 values simulated by UKESM1-0-LL in the HistSST and sensitivity experiments with different drivers of ozone fixed

cal differences across the three models, which also influences the magnitude of the simulated historical change in OSDMA8. Global increases of more than 50% in OSDMA8 concentrations were simulated over the historical period by all three CMIP6 models, with the largest changes over the Northern Hemisphere regions (North America, Europe, Asia). This change has increased the proportion of the global population being exposed to OSDMA8 concentrations above the theoretical minimum risk

- 410 exposure level (TMREL), below which ozone is considered as not being harmful for human health, from less than 20% in 1855 to more than 90% in 2010. The risk to human health from long-term exposure to ozone concentrations has been assessed in terms of changes to the attributable fraction (AF), defined here as the percentage of deaths from chronic obstructive pulmonary disease (COPD) attributable to ozone pollution. The AF has increased by more than 10% over many regions of the Northern Hemisphere, showing that increased ozone concentrations over the historical period have contributed to a significant increase
- 415 in the long-term risk to human health across many populated regions of the world. In addition, the historical increase in surface ozone concentrations has made the risk of human mortality from long-term exposure to ozone similar to other environmental

risk factors, but still much less than from long-term exposure to fine particulate matter (GBD 2021 Risk Factors Collaborators, 2024).

- 420 An analysis of the drivers of historical changes in OSDMA8 concentrations and risks to human health is provided from sensitivity experiments conducted using a single CMIP6 model. Historical increases in anthropogenic emissions of NO_x is identified as the leading contributor to increases in OSDMA8 concentrations ($\approx 40\%$ globally) over the period 1850 to 2014, particularly over Northern Hemisphere regions. Increases in global CH⁴ concentrations is also shown to have an important contribution to the historical changes in OSDMA8 concentrations ($\approx 30\%$ globally), particularly over more southern hemisphere 425 regions where its contribution can be as large of that from NO_x . The increase in anthropogenic emissions of other ozone pre-
- cursors (CO and non-CH₄ VOCs) has had a relatively small impact on increasing historical OSDMA8 concentrations ($\approx 7\%$) globally) compared to NO_x and $CH₄$. The other drivers considered here (stratospheric ozone, aerosols and climate change) all have relatively small contributions to historical changes in OSDMA8 concentrations. The depletion of stratospheric ozone since the 1950s has resulted in less downward transport of ozone to the surface and a small reduction in historical surface
- 430 concentrations, particularly over more remote Southern Hemisphere regions. Increasing anthropogenic emissions of aerosols also caused a small reduction in OSDMA8 concentrations, particularly over Asia, although this effect might be underestimated due an under representation of heterogeneous chemistry mechanisms within the model (Ivatt et al., 2022). Historical climate change contributed to a small reductions in historical OSDMA8 concentrations over remote regions and a small increase over polluted continental regions due to changes in ozone production and destruction pathways. The response of ozone to climate
- 435 change was small due to the small 1K of global warming considered over the historical period, although the results were consistent with projected changes due to future warming (Zanis et al., 2022).

Increasing anthropogenic NO_x emissions over the 1850 to 2014 period is shown to be the main contributor to the long-term health effects from ozone exposure in the near present day period (2010). Fixing anthropogenic NO_x emissions at 1850 reduces 440 the proportion of the near present day global population exposed to OSDMA8 concentrations above the TMREL to 34%, from more than 90% when they are allowed to increase over the historical period along with other precursors. Fixing NO_x emissions at 1850 also reduces the AF to near zero across most regions, thereby eliminating the risk to human health from long-term ozone exposure. Historical global CH_4 concentrations are shown to be the next most important contributor to the health effects from long-term ozone exposure. Fixing global CH⁴ concentrations at 1850 values reduces the proportion of the global popula-445 tion exposed to OSDMA8 concentrations above the TMREL to 83%, and reduces the AF, and risk to human health, by about half in populated regions. However, the influence of the other drivers is small, having little impact on the health risk of the near present day population from exposure to elevated surface ozone concentrations.

The results from these experiments provide a unique opportunity to quantify the long term changes and drivers behind 450 the changes of an ozone exposure metric relevant to health impacts. Substantial increases have occurred in surface ozone concentrations and the risk to human health from long-term exposure to these elevated concentrations over the historical

period. These changes can be mainly attributed to changes in anthropogenic sources of ozone precursors such as NO_x and CH4, with smaller contributions from CO and non-CH⁴ VOCs. Sensitivity studies show that drastic changes to these main drivers can have large impacts on the risk to human health. However, there are also non-negligible changes due to other drivers 455 of ozone (stratospheric ozone, aerosols and climate change), which could become more significant in the future depending on the particular emissions and climate pathway the world takes. Therefore, understanding the historical changes and drivers of ozone and human health could help to inform future policy measures.

Data availability. The hourly surface ozone data used in this study has been obtained from the CMIP6 data archive which is hosted at the Earth System Grid Federation and is freely available to download from https://esgf-node.llnl.gov/search/cmip6/. The references for 460 AerChemMIP data used from GFDL-ESM4, EC-Earth3-AerChem and UKESM1-0-ll are Horowitz et al. (2018); EC-Earth Consortium (2020); O'Connor (2019). The CMIP6 variable name sfo3 has been used from the experiments listed in Table 1. Processed model data and calculations related to human health that are used in this study can be found on zenodo at https://doi.org/10.5281/zenodo.13385648. The RAMP surface ozone observational data is from Becker et al. (2023).

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Appendix A: Supplementary Figures

Figure A1. 21 regions used in this study based on those from the Global Burden of Disease study

Figure A2. Uncorrected 10-year mean regional mean surface OSDMA8 values from 3 CMIP6 models simulated over the historical period (1850-2014) with 10-year mean values from RAMP observations (Becker et al., 2023) shown in grey diamonds.

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Figure A3. Relative change in historical regional emissions of ozone precursors (NOx, CO and non-CH4 VOCs- NMVOCs) compared to 1850 values.

Figure A4. 10-year mean in surface OSDMA8 values simulated by UKESM1-0-LL in the HistSST experiment for 2005-14 (bias-corrected) and the change in the same time period in the HistSST experiment relative to the histpiSST, histSST-1950HC and histSST-piAer sensitivity experiments with different drivers of ozone (climate change, stratospheric ozone and aerosols) fixed. Same information as Figure 6 but different colour scale used

Figure A5. Relative change in historical regional emissions of ozone precursors (SO2, OC and BC) compared to 1850 values

465 *Author contributions.* Steven T. Turnock set out the conceptual idea of the research project with inputs from Dimitris Akritidis and Andrea Pozzer. The main formal analysis, including the creation of figures, was conducted by Steven T. Turnock and Dimitris Akritidis. Observational data and additional analysis on this data was provided by Hantao Wang. Additional model data was provided by Larry Horowitz and Putian Zhou. Steven T. Turnock prepared the manuscript with contributions of writing and editing from all co-authors.

Competing interests. At least one of the (co-)authors is a member of the editorial board of Atmospheric Chemistry and Physics.

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