A Comprehensive Characterization of Empirical Parameterizations

for OH Exposure in the Aerodyne Potential Aerosol Mass Oxidation

Flow Reactor (PAM-OFR)

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- 19 **Abstract.** The oxidation flow reactor (OFR) has been widely used to simulate secondary organic aerosol (SOA) formation in
- 20 laboratory and field studies. OH exposure (OH_{exp}), representing the extent of hydroxyl radical (OH) oxidation and normally
- 21 expressed as the product of OH concentration and residence time in the OFR, is important in assessing the oxidation chemistry
- 22 in SOA formation. Several models have been developed to quantify the OH_{exp} in OFRs, and empirical equations have been
- 23 proposed to parameterize OHexp. Practically, the empirical equations and the associated parameters are derived under
- 24 atmospheric relevant conditions (i.e., external OH reactivity) with limited variations of calibration conditions, such as
- 25 residence time, water vapor mixing ratio, O₃ concentration, etc. Whether the equations or parameters derived under limited
- sets of calibration conditions can accurately predict the OH_{exp} under dynamically changing experimental conditions with large 26
- variations (i.e., extremely high external OH reactivity) in real applications remains uncertain. In this study, we conducted 62
- 29 the empirical equations to estimate OH_{exp}. Sensitivity tests were also conducted to obtain a minimum number of data points

sets of experiments (416 data points) under a wide range of experimental conditions to evaluate the scope of the application of

- 30 that is necessary for generating the fitting parameters. We showed that, for the OFR185 mode (185-nm lamps with internal O₃
- 31 generation), except for external OH reactivity, the parameters obtained within a narrow range of calibration conditions can be

extended to estimate the OH_{exp} when the experiments are in wider ranges of conditions. For example, parameters derived

- 33 within a narrow water vapor mixing ratio range (0.49-0.99 %, corresponding to 15.1-30.8 % of relative humidity at 101.325
- kPa and 298 K) can be extended to estimate the OH_{exp} under the entire range of water vapor mixing ratios (0.49–2.76 %,

equivalent to 15.1-85.7 % of relative humidity under identical conditions). However, the parameters obtained when the 35 external OH reactivity is below 23 s⁻¹ could not be used to reproduce the OH_{exp} under the entire range of external OH reactivity 36 37 (4–204 s⁻¹). For the OFR254 mode (254-nm lamps with external O₃ generation), all parameters obtained within a narrow range 38 of conditions can be used to estimate OH_{exp} accurately when experimental conditions are extended. Additionally, when using 39 the OFR254 mode, too-low lamp voltages should be avoided, as they will generally result in large deviations in the estimations 40 of OH_{exp} from empirical equations. Regardless of OFR185 or OFR254 mode, at least 20–30 data points from sulfur dioxide (SO₂) or (carbon monoxide) CO decay with varying conditions are required to fit a set of empirical parameters that can 41 42 accurately estimate OH_{exp}. Caution should be exercised to use fitted parameters from low external OH reactivity to high ones, 43 for instance, those from direct emissions such as vehicular exhaust and biomass burning.

As the most important oxidant in tropospheric chemistry (Ehhalt, 1999), hydroxyl (OH) radical is vital in oxidizing primary

1 Introduction

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pollutants such as volatile organic compounds (VOCs) and contributes to secondary organic aerosol (SOA) and tropospheric 46 47 ozone (O₃) formation. The OH radical has daytime concentrations of 10⁵ to 10⁷ molecules cm⁻³, exhibiting daily (Cao et al., 48 2020; Tan et al., 2017), seasonal (Friedman and Farmer, 2018), as well as spatial (Cao et al., 2020; Stone et al., 2012) variations. 49 An average daily OH radical concentration of 1.5×10^6 molecules cm⁻³ is widely used to estimate the photochemical age of an air mass (Mao et al., 2009). Typical VOCs have second-order rate constants of 10⁻¹⁵ to 10⁻¹⁰ cm³ molecule⁻¹ s⁻¹ with OH radicals 50 51 (Atkinson and Arey, 2003; Atkinson et al., 2006), which can be translated to atmospheric lifetimes of hours to approximately 52 a year (Seinfeld and Pandis, 2016). This situation poses challenges in laboratory experiments to directly simulate the OH 53 oxidation of VOCs, which is one of the most important chemical processes in the Earth's atmosphere. 54 Smog chambers (Cocker et al., 2001; Hildebrandt et al., 2009; Wang et al., 2014) and oxidation flow reactors (OFRs) (George 55 et al., 2007; Kang et al., 2007; Lambe et al., 2011) have been widely employed to simulate oxidation of VOCs and subsequent 56 SOA formation. For example, the Caltech Chamber provides oxidation conditions close to the real atmosphere, making it suitable for the study of complex multi-step reactions and low-volatility products. However, each experiment takes several 57 hours to days and long-duration experiments are prone to background interference. The Toronto Photo-Oxidation Tube (TPOT) 58 59 focuses on the study of heterogeneous oxidation reactions of aerosols. Its 0.8 L volume makes it portable, but it is prone to 60 uneven residence time distribution (RTD) and significant wall effects. The Potential Aerosol Mass Oxidation Flow Reactor (PAM-OFR) and the Gothenburg Potential Aerosol Mass Oxidation Flow Reactor (Go: PAM-OFR) are often used to study 61 62 the transformation of gaseous precursors into particles (such as the formation of SOA). The Go: PAM-OFR has a volume of 63 7.2 L, which is only half that of the PAM-OFR, making it suitable for experiments on mobile platforms. However, its small 64 volume gives it the same disadvantages as the TPOT, and it is equipped with only a single UV lamp, which does not allow for 65 as wide a range of controllable oxidation levels as the PAM-OFR. The PAM's moderate volume and central flow sampling can reduce wall effects. 66

67 These OFRs of reactors normally operate with high concentrations of oxidants (e.g., OH radicals), which lead to a significant acceleration of oxidation reactions, often by orders of magnitude. To reconcile the differences in OH concentration and 68 69 exposure time between ambient and laboratory settings, the oxidation extent, i.e., OH exposure (OH_{exp}, molecules cm⁻³ s) is 70 normally used to extrapolate laboratory findings to ambient conditions. Despite drawbacks such as possible altered reaction 71 mechanisms, this approach provides a quantitative assessment of the chemistry during OH oxidation in a reasonable time span 72 and achievable detection capability. The OH_{exp} has a significant impact on the yield and product distribution during VOC 73 oxidation (Cheng et al., 2021; Cheng et al., 2024). Accurate measurement or estimation of the OH_{exp} during laboratory 74 experiments, therefore, is the key to understanding the oxidation chemistry that can represent the ambient conditions. In this 75 study, we chose to further investigate the PAM-OFR to explore its OH_{exp}, as it offers moderate conditions in terms of 76 experiment time, deployment complexity, range of oxidation levels, and wall effects.

The Aerodyne Potential Aerosol Mass OFR (PAM-OFR) is one of the most widely used OFRs for studying SOA formation and evolution (Zhang et al., 2024). It can achieve a wide range of atmospheric OH_{exp} conditions within short residence times on the order of minutes (Kang et al., 2007; Lambe et al., 2011). The PAM-OFR can be operated in a number of modes, depending on 1) the wavelength of the ultraviolet (UV) light source, 2) the concentration of the externally generated O₃ (if any), and 3) the injection of external precursor to generate NO_x (= NO + NO₂) or other oxidants (e.g., nitrate radical or halogen atoms) upon photolysis. The most widely used methods for OH generation include combined photolysis of O₂ and H₂O at λ = 185 nm plus photolysis of O₃ at λ = 254 nm (OFR185; R1–R6) or photolysis of externally added O₃ at λ = 254 nm (OFR254;

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$$H_2O + hv_{185} \rightarrow H + OH$$
 (R1)

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R5-R6) (Rowe et al., 2020):

$$86 \quad H + O_2 \rightarrow HO_2 \tag{R2}$$

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$$O_2 + hv_{185} \rightarrow 20(^3P)$$
 (R3)

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$$O(^{3}P) + O_{2} \rightarrow O_{3}$$
 (R4)

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$$O_3 + hv_{254} \rightarrow O_2 + O(^1D)$$
 (R5)

$$90 O(^{1}D) + H_{2}O \rightarrow 2OH$$
 (R6)

To obtain the OH_{exp} under these two modes in the PAM-OFR, one can perform decay experiments on trace gases such as SO_2 and CO, and fit the OH_{exp} based on known second-order rate constants between OH radical and the trace gases, which is defined as $OH_{exp, dec}$. Based on the results of the decay experiments, Li et al. (2015) and Peng et al. (2015) developed estimation equations to parameterize OH_{exp} as a function of easily measurable quantities, which is denoted as $OH_{exp, est}$. A set of parameters (a–f and x–z, respectively) for the estimation equations of the OFR185 and OFR254 modes (see Sect. 2.3 for details) were obtained by fitting the estimation equations to $OH_{exp, dec}$ values obtained from decay experiments.

97 When using the PAM-OFR in field studies, it is necessary to obtain concurrent OH_{exp} that is representative of the ambient 98 conditions. However, environmental conditions in field studies (e.g., humidity, temperature, etc.) are constantly changing, 99 making it challenging to replicate these conditions for OH_{exp} estimation. In some field studies using PAM-OFR, concurrent 100 OH_{exp} was estimated by measuring the relative decay of benzene and toluene (Liao et al., 2021; Liu et al., 2018). Additionally, 101 some studies have mentioned that OH concentrations can be indirectly measured by detecting the decay of tracers such as 3-102 pentanol, 3-pentanone, pinonaldehyde, or butanol-d9 (Barmet et al., 2012). However, the measurement of all these organic 103 tracers requires specific, sophisticated instruments such as proton-transfer-reaction time-of-flight mass spectrometers (PTR-104 MS). Additionally, switching the instrument back and forth between the front and end of the OFR during field measurements 105 can result in some loss of real-time VOCs data before entering the OFR. To obtain accurate OH_{exp}, some studies explicitly 106 modelled the radical chemistry in PAM-OFR (Li et al., 2015; Ono et al., 2014; Peng et al., 2015). The estimation equations 107 developed by Li et al. (2015) and Peng et al. (2015), although empirical, reproduced the OH_{exp} from models within 10 %, 108 making them a good choice because these equations only require the input of a few easily available parameters. Yet, it is 109 unclear whether the fitted parameters obtained under certain conditions can still accurately estimate OH_{exp} when experimental 110 conditions, such as UV light intensity, water vapor mixing ratio, residence time, and external OH reactivity (OHR_{ext}), undergo 111 significant changes. Furthermore, there is currently no consensus on the minimum number of decay experiments required to 112 obtain accurate parameterization for OH_{exp} estimation using these equations. This facet is important for field studies using 113 PAM-OFR where only limited numbers of decay experiments can be done to obtain concurrent OH_{exp} estimation. 114 In this study, we conducted a series of experiments using the decay of SO₂ and CO to estimate the OH_{exp} in the PAM-OFR under OFR185 and OFR254 modes. The applicability of previously developed OH_{exp} estimation equations to obtain accurate OHexp in the PAM-OFR has been evaluated by linear regression of OHexp, est against OHexp, dec. We have also evaluated how well estimation equations perform when using limited ranges of experimental parameters (e.g., OHR_{ext}, residence time, water

under OFR185 and OFR254 modes. The applicability of previously developed OH_{exp} estimation equations to obtain accurate OH_{exp} in the PAM-OFR has been evaluated by linear regression of $OH_{exp, est}$ against $OH_{exp, dec}$. We have also evaluated how well estimation equations perform when using limited ranges of experimental parameters (e.g., OHR_{ext} , residence time, water mixing ratio, etc.) or different trace gases (SO₂ and CO) and given recommendations. In addition, we have proposed the minimal number of trace-gas decay experiments required to obtain a set of usable parameters for the OH_{exp} estimation equations. Finally, we also compared the advantages and disadvantages of the OFR185 and the OFR254 modes from the perspective of the quantification of OH_{exp} . The methodology of this study can be applied to laboratory and field experiments

for OH_{exp} estimation using PAM-OFR or other OFRs that follow a plug-flow assumption.

2 Methods

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124 2.1 The PAM-OFR

- 125 Experiments were conducted using an Aerodyne PAM-OFR (Aerodyne Research Inc., Billerica, MA, US), which is a
- horizontal aluminium cylindrical chamber with an internal volume of 13.3 L. The PAM-OFR operates in a continuous flow
- mode. Four low-pressure Hg lamps are installed inside the reactor to produce UV light with characteristic spectral lines (e.g.,
- 128 185 and 254 nm). The OH is generated via OFR185 using two ozone-producing Hg lamps (GPH436T5VH/4P, Light Sources,

Inc.) or via OFR254 using two ozone-free Hg lamps (GPH436T5L/4P, Light Sources, Inc.) to photolyze externally added ozone. A flow of nitrogen purge gas, ranging from 0.2 to 0.3 L min⁻¹, is introduced between the lamps and sleeves. This nitrogen gas flow serves to reduce the heat generated by the lamps and prevent the formation and accumulation of ozone between the lamps and the quartz tubes that isolate them from the sample flow in the OFR. A fluorescent dimming ballast is used to control the photon flux by regulating the voltage applied to the lamps, which allows us to generate different OH concentrations. In typical measurement sequences, nine lamp voltage settings (including lights off) were cycled through every 2–3 hours. The dimming voltage ranged from 0 to 10 V direct current (DC).

OH_{exp} can be indirectly measured by detecting the decay of the tracers with known reaction rates. Inorganic trace gases SO₂

2.2 OH_{exp} estimation through decay of SO₂ and CO (OH_{exp, dec})

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138 or CO react with OH radicals at slower rates compared to most VOCs. However, considering the complex oxidation chemistry 139 of VOCs, SO₂ and CO can better capture the features of real OHR_{ext} decay and effective OHR_{ext} (Peng et al., 2015). We 140 performed systematic decay experiments with SO₂ and CO in the PAM-OFR, with conditions tabulated in Tables S1 and S2. 141 Figure S1 shows the schematics of the experimental setups in the OFR185 and OFR254 modes. In the OFR185 mode, the 142 injected gas flow at the inlet of the PAM is made up of three sub-flows: (1) The trace-gas flow, i.e. SO₂ of 0.2–8.7 ppm or CO 143 of 10.2–207.5 ppm supplied from gas cylinders (Purity: 99.9 % of SO₂, 99.95 % of CO; Shanghai Shenkai Gases Technology 144 CO., LTD.); (2) dry clean air from a zero-air generator (ZAS-100/150, Convenient) with a non-methane hydrocarbon content 145 of less than 1 ppb; (3) the humidified clean air passed through a Nafion humidifier (FC100-80-6MSS, Perma Pure). By 146 adjusting the ratio of dry air to humidified air, the water vapor mixing ratio in the PAM-OFR can be controlled. Additionally, 147 they also serve as makeup flows to maintain a constant flow rate. At the outlet of the reactor, the gas flow was sampled from 148 an internal perforated Teflon ring. The gas-phase species (O₃, SO₂, and CO) were detected using an ultraviolet ozone analyser 149 (UV-100, Eco Sensors), an SO₂ monitor (Model 43i, Thermo Scientific), and a CO monitor (G2401, Picarro), respectively. In 150 the OFR254 mode, in addition to the previously mentioned setup, externally generated O₃ (through UV photolysis) with desired 151 concentrations was injected at the inlet of the PAM-OFR. 152 Figures S2a and S2b depict examples of set and measured parameters during experiments conducted in the OFR185 and 153 OFR254 modes, respectively. In the OFR185 mode, without radical generation to oxidize the tracer species, their concentration 154 was allowed to stabilize under dark conditions. Once the concentration reached a steady state, the UV lamps were turned on. 155 Different light intensities lead to varying levels of decay of SO₂ or CO after oxidation, reflecting different OH_{exp} within the 156 PAM-OFR. In the OFR254 mode, it is necessary to obtain the initial concentration of O₃ injected into the PAM-OFR in the 157 absence of OHR_{ext}. While waiting for the SO₂ or CO concentration to stabilize, the O₃ flow was temporarily blocked outside 158 the PAM-OFR using a valve. Dry clean air was then introduced to compensate for this portion of the flow, ensuring a constant 159 total flow throughout the entire process. Once the tracer species concentration had reached a steady state, the O_3 was then 160 allowed to flow into the PAM-OFR. The total $OH_{exp, dec}$ in the reactor was varied over a wide range (approximately 10^9-10^{12} 161 molecules cm⁻³ s) by changing the UV light intensity, water mixing ratio, and residence time. The mean residence time was

- obtained from the ratio of the internal volume of and the total flow rate through the PAM-OFR. In the calculation of OH_{exp. dec}
- 163 (see the paragraph below), plug flow conditions were assumed, which has been shown to agree with the RTD approach for
- 164 OH_{exp} when using species (such as SO₂ or CO) with low reaction rate constants with OH radicals (k_{i, OH}) by Li et al. (2015)
- 165 and Peng et al. (2015).
- 166 OH_{exp, dec} in the PAM-OFR was calculated from the pseudo-first-order reaction of OH with SO₂ or CO, whose k_{i, OH} have been
- well characterized ($k_{SO2, OH} = 9.49 \times 10^{-13} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1} \text{ and } k_{CO, OH} = 2.4 \times 10^{-13} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1} \text{ at 1 atm and 298 K}$
- (Burkholder et al., 2020; Cao et al., 2020). By measuring the decay of SO₂ or CO, the corresponding OH_{exp, dec} is calculated as
- 169 follows:

$$170 \quad OH_{\text{exp, dec}} = \frac{-1}{k_{i, \text{ OH}}} \times \ln\left(\frac{C_{i, \text{ out}}}{C_{i, \text{ in}}}\right) \tag{1}$$

- 171 where $c_{i, in}$ is the concentration of reactant i injected into the PAM-OFR (ppb), $c_{i, out}$ is reactant i concentration at the PAM-
- 172 OFR outlet (ppb), and k_{i,OH} is the second-order rate constant between the trace species (SO₂ or CO) and OH radicals.

173 **2.3 OH**_{exp} estimation from empirical equations (OH_{exp, est})

Li et al. (2015) proposed an OH_{exp. est} estimation equation (Eq. 2) for OFR185 based on easily measurable quantities:

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$$OH_{exp, est} = 10^{\left[a + \left(b + c \times OHR_{ext}^d + e \times log\left(O_{3, out} \times \frac{180}{t}\right) \times OHR_{ext}^f\right) \times log\left(O_{3, out} \times \frac{180}{t}\right) + logH_2O + log\left(\frac{t}{180}\right)\right]}$$
 (2)

- where a-f are fitting parameters (values are reported in Table S5); O_{3, out} is ozone concentration measured at the exit of the
- 177 PAM-OFR (molecules cm⁻³), which serves as a surrogate for UV flux; H₂O is water vapor mixing ratio in PAM-OFR (%),
- 178 which is influenced by both temperature and relative humidity; t is mean residence time (s). The total external OH reactivity
- 179 is represented by $OHR_{ext}(s^{-1}) = \sum_i k_i [C_i]$, where k_i and $[C_i]$ are the rate constants with OH and the concentration of the OH-
- 180 consuming reactant i in the system (Wang et al., 2020).
- Peng et al. (2015) proposed another equation (Eq. (3)) for OH_{exp, est} in OFR254:

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$$OH_{exp, est} = 10^{\left[x + \log(-\log r_{O_3}) + y \times \left(\frac{OHR_{ext}}{O_{3, in}}\right)^z\right]}$$
 (3)

- 183 where x-z are fitting parameters (values are reported in Table S6); $\log r_{O3}$ ($\log (O_{3, out}/O_{3, in})$) is the logarithm of the ratio
- between the output and input O_3 concentrations, which serves as a surrogate for UV flux and also captures the effect of H_2O ;
- $O_{3, in}$ is the concentration of externally injected O_3 into the PAM-OFR (molecules cm⁻³).
- 186 We have performed in total of 62 sets of trace-gas decay experiments with 416 data points for the OH_{exp. dec}, with 25 sets and
- 187 175 data points in the OFR185 mode and 37 sets and 241 data points in the OFR254 mode. In OFR185 mode, the 175
- 188 experiments cover an $OH_{exp, dec}$ range of $3.57 \times 10^8 5.52 \times 10^{12}$ molecules cm⁻³ s, with an equivalent photochemical age ranging
- from 4 minutes to 43 days. In OFR254 mode, the 241 experiments cover an $OH_{exp, dec}$ range of $1.01 \times 10^9 2.18 \times 10^{12}$ molecules
- 190 cm⁻³ s, with an equivalent photochemical age ranging from 11 minutes to 17 days. The error in OH_{exp, dec} is derived from the

measurement error of the tracer gas, propagated through Eq. 1. When $OH_{exp, dec}$ ranged from $3.6 \times 10^8 - 5.5 \times 10^{12}$ molecules

192 cm⁻³ s, the resulting error values were $1.9 \times 10^8 - 2.4 \times 10^{10}$ molecules cm⁻³ s.

193 After obtaining the OH_{exp, dec} values, we used Eqs. 2 and 3 to fit the parameters a-f and x-z for OFR185 and OFR254 modes,

194 respectively, given that the experimental parameters such as OHR_{ext}, O_{3, out}, H₂O, and t (in Eq. 2), and r_{O3}, OHR_{ext}, and O_{3, in}

195 (in Eq. 3) are known. The OH_{exp, est} values were then reconstructed with the fitted parameters and the experimental parameters,

and compared with the OH_{exp. dec} values via linear regression analysis. Similarly, the error values for all OH_{exp. est} values are at

197 least one order of magnitude smaller than the respective OH_{exp, est} values. The generation of OH radicals in PAM-OFR is related

to the photon fluxes at $\lambda = 185$ nm (I_{185}) and $\lambda = 254$ nm (I_{254}). According to Rowe et al. (2020), I_{185} : I_{254} is specific to the Hg

199 lamp utilized. Since the OH_{exp} estimation equation for OFR185 uses O₃ concentration as a measurable surrogate for the UV

200 flux at 185 nm, it is also lamp-specific. Because the UV lamps used in our study are different from the BHK lamps employed

by Li et al. (2015), we anticipate that the parameters a-f fitted from our decay experiments (Table S5) should be quite different

202 from those in Li et al. (2015), which is indeed the case. Similarly, fitting parameters x-z for OFR254 mode from our decay

203 experiments (Table S6) are also different from those in Peng et al. (2015).

3 Results and Discussion

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3.1 The OFR185 mode: OHR_{ext} level relevant to ambient conditions

Field studies showed that the environmental OHR_{ext} mainly fluctuated between 10–30 s⁻¹ (Fuchs et al., 2017; Lou et al., 2010;

207 Lu et al., 2010; Tan et al., 2018; Yang et al., 2017). To investigate the factors that potentially affect the fitting parameters of

208 Eq. 2 in the estimation of OH_{exp} under ambient conditions, we first performed 16 sets of experiments with OHR_{ext} of 4–23 s⁻¹

209 using SO₂ as the OHR_{ext} source. With the measured OH_{exp, dec}, the parameters (a-f) were first derived, which were used to

210 reconstruct OH_{exp, est} using Eq. 2 with known OHR_{ext}, ozone concentration (O_{3, out}), water vapor mixing ratio (H₂O), and

211 residence time (t). The reconstructed OH_{exp, est} values were plotted against the OH_{exp, dec} values calculated from the trace-gas

decay experiments, as shown in Figure 1. The 1:2 and 2:1 lines indicate approximately half an order of magnitude difference

213 between OH_{exp, dec} and OH_{exp, est}, which is considered to be acceptable as an uncertainty in OH_{exp} estimation.

We first investigated the effect of changing residence time on the OH_{exp} estimation. With other experimental parameters (i.e.

215 H₂O, O_{3, out}, and OHR_{ext}) being similar, we set the residence time to a low value (33 s) and also a range of higher values (61–

216 200 s). The detailed ranges of each experimental condition for different datasets are listed in Table S3. With the residence time

of 33 s, the reconstructed OH_{exp, est} correlates well with the experimental OH_{exp, dec} (slope = 1.061 and $R^2 = 0.990$, Figure 1a1).

218 The set of fitted parameters a-f (FP_{st. 185}; st: short time) applied in Figure 1a1 is presented in Table S5. When the residence

219 time was increased to 61–200 s, the interpolated OH_{exp, est} utilizing FP_{st, 185} was also in good correlation with OH_{exp, dec} (slope

220 = 0.978, $R^2 = 0.959$, Figure 1a2). We also derived fitted parameters (FP_{et, 185}; et: extended t) using the data points with the

extended range of residence time (33–200 s). Not surprisingly, with the application of FPet, 185, OHexp, est also correlated well

with $OH_{exp, dec}$ (slope = 0.994, R^2 = 0.955, Figure 1a3). The results indicate that variation in residence time does not significantly

223 affect the fitting parameters of Eq. 2 for the OH_{exp} estimation. From an experimental perspective, since OH_{exp} is the product 224 of OH radical concentration ([OH]) and the residence time (t), as long as the change of t does not significantly alter the quasi-225 steady-state [OH], the fitted parameters from a narrow range of t should be applicable to situations of longer t. Mathematically, 226 two terms of 180/t and t/180 are related to t, ranging from 0.90–5.45 and 0.18 to 1.11, respectively, which do not contribute 227 significantly to the exponent in Eq. 2 after taking the logarithm of them. It is important to note that the above discussion 228 regarding residence time assumes a plug-flow condition within the PAM-OFR, which is applicable to substances with low k_i. 229 OH, such as SO₂ (or CO). For species that react rapidly with OH, such as monoterpenes or toluene, localized concentration 230 gradients can develop within the OFR, leading to a significant uneven actual RTD that affects the estimation of OH_{exp} (Palm 231 et al., 2018).

Similarly, we then investigated the impacts of H₂O on the estimation of OH_{exp}. Applying fitted parameters from experiments of low water vapor mixing ratios (0.49–0.99 %, Figure 1b1) (FP_{IH20, 185}; IH₂O: low H₂O) to data spanning a wide range of water vapor mixing ratios (0.49–2.76 %) also yielded a reasonably good correlation between OH_{exp. dest} and OH_{exp. dest} (Figure 1b2). This could be attributed to the fact that the term $log H_2O$ in Eq. 2 does not contribute significantly to the exponent.

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236 As for ozone concentration, applying fitting parameters (FP_{103, 185}; 1O₃: low O_{3, out}) from experiments of low ozone concentration (1.44 \times 10¹²–6.79 \times 10¹³ molecules cm⁻³, Figure 1c1) to reconstruct the data for a wide range (1.44 \times 10¹²–2.03 237 × 10¹⁵ molecules cm⁻³) yielded a reasonably good correlation between OH_{exp, est} and OH_{exp, dec} (Figure 1c2). It only resulted in 238 239 a mildly increased slope (from 1.063 to 1.272) and similar R² values (both are 0.970) as compared to those using the whole 240 ozone concentration range (Figure 1c3).

Ideally, trace-gas decay experiments covering the entire ranges of the t, H₂O, and O_{3, out} variations under real experimental conditions should be conducted, which is labor-intensive. Practically, due to the atmospherically relevant variations that occur in t, H₂O, and O_{3, out} during the real experiments, the ranges of t, H₂O, and O_{3, out} covered by trace-gas decay experiments are usually narrower compared to the real experiments. Our results suggest that the fitting parameters (a-f) obtained from calibration experiments with relatively narrow ranges of t, H₂O, and O_{3, out} can still provide a reliable estimation of OH radical levels during the real experiments, which would cover wider ranges of these conditions.

It is noteworthy that reliable estimations can be achieved regardless of whether the narrow range is situated within the lower or higher interval of the full condition range. Figure 1 demonstrated the case where the narrow range was situated within the lower interval, while Figure S3 presented the case where the narrow range was situated within the higher interval. The detailed ranges of each experimental condition for different datasets are listed in Table S3. As shown in Figure S3, the data points in panel al had residence times of 100–296 s, the data points in panel bl had water vapor mixing ratios of 1.04–2.76 %, and the data points in panel c1 had $O_{3, \text{out}}$ of $8.45 \times 10^{13} - 2.03 \times 10^{15}$ molecules cm⁻³. Panels a2, b2, and c2 built on panels a1, b1, and c1 by incorporating data points with shorter t (33–61 s), lower H_2O (0.49–0.97 %), and lower $O_{3, \text{ out}}$ (1.44 × 10¹²–6.79 × 10¹³ molecules cm $^{-3}$), respectively, but still used fitting parameters a-f obtained from the higher range of conditions to estimate $OH_{exp. est.}$ In panels a3, b3, and c3, the parameters a-f were refitted using all the data points included in the expanded t, H_2O , example, the slope and R^2 values in a2 and a3 were very close to 1, reflecting the good consistency between $OH_{exp, est}$ and $OH_{exp, dec}$. In the OFR254 mode discussed later (Figure 4, panels c1–c3), this narrower range can also be situated within the middle interval of the full condition range. This applicability of fitting parameters obtained from narrow ranges of experimental conditions is beneficial for quickly obtaining concurrent OH_{exp} during the experiments in field measurements.

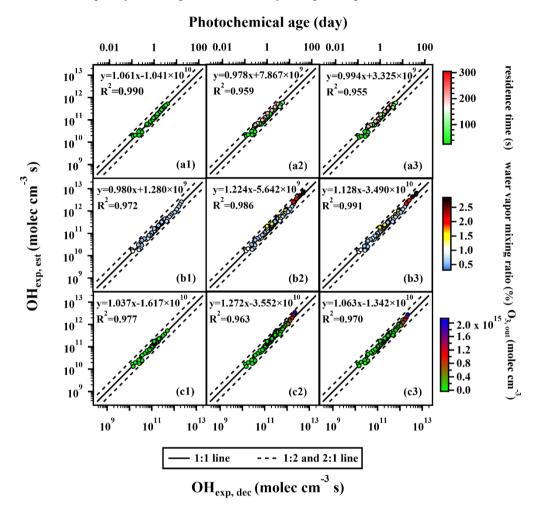


Figure 1: The regression results of $OH_{exp, est}$ and $OH_{exp, dec}$ when variations occurred in (a1–a3) residence time, (b1–b3) water vapor mixing ratio, and (c1–c3) output O_3 concentration under atmospheric relevant OHR_{ext} level (4–23 s⁻¹). Compared to panels a1, b1, and c1, panels a2, b2, and c2 respectively incorporated additional data points with higher t, H₂O, and $O_{3, out}$ values, but still utilized the fitting parameters $FP_{st, 185}$, $FP_{IH2O, 185}$, and $FP_{IO3, 185}$ obtained from the lower condition range to estimate $OH_{exp, est}$. In panels a3, b3, and c3, all data points within the extended condition range were used to re-fit the parameters a–f, and the resulting $FP_{et, 185}$, $FP_{eH2O, 185}$, and $FP_{eO3, 185}$ were employed to estimate $OH_{exp, est}$ (s: short, l: low, e: extended).

3.2 The OFR185 mode: OHRext level relevant to emission sources

The experimental conditions in the PAM-OFR often involve not only general atmospheric conditions (OHR_{ext} $< 30 \text{ s}^{-1}$) but also high-concentration conditions, e.g., those directly from emission sources. For instance, the OHR_{ext} of direct vehicle

271 emission can be as high as 1000 s⁻¹ with plenty of reducing gases such as CO and VOCs (Nakashima et al., 2010). To evaluate 272 the applicability of Eq. 2 under situations of high OHR_{ext}, we performed high OHR_{ext} (up to 204 s⁻¹) experiments using high 273 concentrations of SO₂ as the OHR_{ext} source. Compared to the data points shown in Figure 2a (4–23 s⁻¹), Figure 2b and Figure 274 2c included additional data points with higher OHR_{ext} values (198–204 s⁻¹), while the other conditions remained similar. In 275 Figure 2b, the parameters a-f (FP_{10HR, 185}; IOHR: low OHR_{ext}) obtained from the low-OHR_{ext} data points were used to estimate 276 $OH_{exp. est}$, yet those used in Figure 2c were refitted from the data points with extended OHR_{ext} range (4–204 s⁻¹). It could be 277 observed from Figure 2b that when estimating OH_{exp} using FP_{IOHR, 185}, OH_{exp, est} of the high-OHR_{ext} data points were 278 significantly overestimated, with a difference of more than two orders of magnitudes compared to OH_{exp. dec}. This observation 279 suggests that, different from cases for residence time, water vapor mixing ratio, and ozone concentration shown in the section 280 above, FP_{IOHR, 185} were not applicable to high-OHR_{ext} conditions.

We then investigated the possible causes of the discrepancy for OH_{exp} estimation between FP_{IOHR}, 185 and FP_{eOHR}, 185 (eOHR: 281 extended OHR_{ext}). From a mathematical perspective, according to Eq. 2, the third term $c \times \text{OHR}_{\text{ext}}^d \times \log(O_{3, \text{ out}} \times 180/t)$ and 282 the fourth term $e \times OHR_{ext}^f \times [log(O_{3, out} \times 180/t)]^2$ are associated with OHR_{ext} , which involve fitted parameters of c-f. To 283 investigate their relationships with OHR_{ext}, we performed a sensitivity test with a fixed ozone concentration $(1.77 \times 10^{14}$ 284 285 molecules cm⁻³) and residence time (89 s), which were mean values during our experiments. When using the c-f values of 286 FP_{IOHR, 185} (-0.13922, 0.26786, 0.0026332, and 0.4917), the variations of the third term, the fourth term, and their sum with respect to OHR_{ext} were shown in Figure S4a1–a3, respectively. The third term (Figure S4a1) was negative and decreased as 287 288 OHR_{ext} increased, while the fourth term (Figure S4a2) was positive and increased as OHR_{ext} increased. The sum of them (Figure S4a3), however, first decreased and then started to increase at approximately OHR_{ext} = 21 s⁻¹, owing possibly to a 289 290 slower decrease in the third term or a faster increase in the fourth. If contributions from other terms in Eq. 2 were constant, this led to an increase of OH_{exp} as OHR_{ext} increased beyond 21 s⁻¹. Our results showed that the expectation that OH_{exp} should 291 292 decrease with increasing OHR_{ext} (Li et al., 2015) was applicable to the lower ranges of OHR_{ext}, i.e., under atmospheric relevant 293 conditions. With further increase of OHR_{ext}, i.e., above atmospheric relevant condition, the fitted parameters obtained from the 294 dataset with FP_{IOHR, 185} were not applicable.

When using the *c*–*f* values of FP_{eOHR, 185} (-0.079114, 0.36805, 0.0041654, and 0.38722), the trends of the third and the fourth terms (Figure S4b1 and S4b2, respectively) were similar to those with low OHR_{ext} (Figure S4a1 and S4a2, respectively); their sum, however, gave a monotonical decreasing trend as OHR_{ext} increased (Figure S4b3), consistent with the expectation that OH_{exp} should decrease with increasing OHR_{ext} (Li et al., 2015). The curve in Figure S4b3 can continue to decrease monotonically at higher OHR_{ext} values, at least until 2000 s⁻¹.

From the perspective of oxidation chemistry, high concentrations of gas phase SO₂ could lead to more SO₂ entering the particle phase. The H₂O₂ in the liquid water of nucleated sulfuric acid aerosols would further oxidize SO₂ (Liu et al., 2020), which could lead to the discrepancy for OH estimation between low OHR and extended high OHR.

Nevertheless, the good agreement between $OH_{exp, est}$ and $OH_{exp, dec}$ in Figure 2c (using re-fitted parameters from the dataset of extended OHR_{ext}) indicate that Eq. 2 can still be used to estimate OH_{exp} under high- OHR_{ext} conditions. This conclusion is

further supported by the results of OH_{exp} obtained using CO as the OHR_{ext} source (see Figure 3 and the section below) under extremely high- OHR_{ext} conditions (up to 1200 s^{-1}). This is advantageous for the use of PAM-OFR in simulations of SOA formation from direct emission sources (e.g., vehicular exhaust and biomass burning) where OHR_{ext} is extremely high. It is, however, desirable to have OH_{exp} estimated under similarly high OHR_{ext} for those experiments to accurately represent the extent of oxidation.

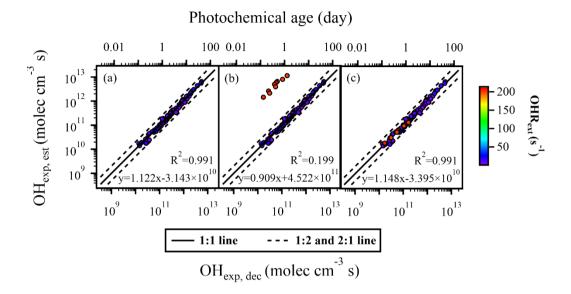


Figure 2: The regression results of $OH_{exp, est}$ and $OH_{exp, dec}$ with different OHR_{ext} levels. In panel a, data points with atmospheric relevant OHR_{ext} level (4–22 s⁻¹) were applied. In addition to the data points contained within panel a, panel b included additional data points with emission sources related OHR_{ext} level (198–204 s⁻¹), but $FP_{IOHR, 185}$ were still used to estimate $OH_{exp, est}$. In panel b, data points in red showed that the $OH_{exp, est}$ of these high- OHR_{ext} data points were significantly overestimated. $FP_{IOHR, 185}$ were not applicable to high- OHR_{ext} conditions. The data points in panel c were identical to those in panel b, but the estimation of $OH_{exp, est}$ utilized the $FP_{eOHR, 185}$ obtained by fitting all data points across the full range of OHR_{ext} levels.

3.3 The OFR185 mode: SO₂ and CO as OHR_{ext} sources

Peng et al. (2015) suggested that SO_2 can better capture the features of real OHR_{ext} decay and effective OHR_{ext}. The reaction between SO_2 and OH is relatively straightforward and is not expected to undergo too many side reactions. CO is a typical gaseous inorganic compound emitted during combustion process. Using CO as an OHR_{ext} source to explore the estimation of OH_{exp} in the simulation of oxidation chemistry for emission sources (i.e., high OHR_{ext} level) is representative. Therefore, we compared the results with SO_2 (Figure 3a) and CO (Figure 3b) as the OHR_{ext} source. When using SO_2 as the OHR_{ext} source, all data points agreed within a factor of 2 (Figure 3a). while only approximately 83 % of the data points agreed within a factor of 2 when CO was used as the OHR_{ext} source (Figure 3b). The deviating data points were mostly concentrated in areas with high OHR_{ext} (> 600 s⁻¹) and low $O_{3, \text{ out}}$ concentration (10^{12} – 10^{13} molecules cm⁻³), where the removal of CO was relatively low. Li et al. (2015) have observed increased deviations between OH_{exp, est} and OH_{exp, dec}, which was attributed, at least in part, to the increased measurement uncertainties for CO when the decrease of its concentration was marginal. We believe that

328 measurement uncertainty might not be the main reason in our case, because most of the decreases in CO concentration during 329 our experiments were larger than the precision of the Picarro G2401 Analyzer (~1.5 ppb at 5 min time resolution). Another 330 possible reason is that in addition to the reaction with OH radicals, CO may react with some other oxidants, leading to its 331 consumption, while SO₂ was less affected, thereby resulting in more scattered data points for CO. The reaction rate of CO with HO_2 is very slow, and is unlikely to play a significant role ($K_{CO, HO2} = 5.55 \times 10^{-27}$ cm³ molecule⁻¹ s⁻¹ at 300 K) (You et al., 332 2007). Cohen and Heicklen (1972) suggested that CO could also react with atomic oxygen (O(¹D)). Clerc and Barat (1967) 333 have reported some appreciable rate coefficients (10⁻¹¹ to 10⁻¹² cm³ molecule⁻¹ s⁻¹) for the reaction between CO and O(¹D). 334 which are higher than those for the reactions of CO with OH ($k_{CO,OH} = 2.4 \times 10^{-13}$ cm³ molecule⁻¹ s⁻¹ at 298 K) (Burkholder et 335 336 al., 2020). It is therefore possible that reaction between CO and O(1D) might have complicated the decay of CO in the PAM-337 OFR. To further investigate this aspect, we used the KinSim, a kinetic simulator, to calculate the average mixing ratios of OH, 338 O(1D), and HO₂ under the specific conditions in the PAM-OFR, and then assessed the relative importance of the reactions CO + OH \rightarrow CO₂ + H, CO + O(1 D) \rightarrow CO₂, and CO + HO₂ \rightarrow CO₂ + OH (Li et al., 2015; Peng and Jimenez, 2019, 2020). The 339 340 results show that although the reaction rate constant of CO and O(1D) is 1-2 orders of magnitude higher than that of CO and 341 OH, the concentration of OH is about 6–7 orders of magnitude higher than the concentration of O(¹D), indicating that the 342 reaction of CO with O(1D) will not have a significant impact on the consumption of CO. The real reason for the scattered data 343 points when using CO in the trace-gas decay experiment is still unknown. 344 Figure 3c includes the results of trace-gas decay experiments using both SO₂ and CO as the OHR_{ext} source. Despite having 345 different reaction rates with OH radicals, the data points could be collectively utilized to fit the parameters for the estimation 346 equation. With approximately 95 % of the results agreeing within a factor of 2, OH_{exp. est} obtained using the fitted parameters exhibited good agreements (slope = 1.101, $R^2 = 0.991$) with $OH_{exp, dec}$. Our results thus suggest that although using CO as the 347 348 OHR_{ext} might result in some scattered data points, it was still feasible to use Eq. 2 to estimate OH_{exp} given that experiments 349 were not done solely in conditions with high OHR_{ext} (i.e., high CO concentrations) and low O₃ concentrations. Another benefit 350 of using CO as OHRext source for the estimation of OHexp is that it introduces complexity in the precursor, which resembled 351 those in real applications. Although not tested in this study, we also note that further trace-gas decay experiments in the 352 presence of N_2O/NO_x (typical urban environment) should be conducted when oxidation chemistry in the presence of NO_x is 353 studied (Cheng et al., 2021).

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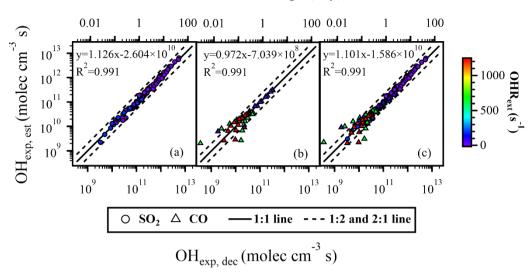


Figure 3: The regression results of $OH_{exp, dec}$ and $OH_{exp, est}$ in the OFR185 mode with (a) SO_2 and (b) CO as OHR_{ext} sources. (c) Results from all experiments (using SO_2 and CO) in the OFR185 mode.

3.4 The OFR254 mode

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The equation for OH_{exp} estimation in OFR254 mode is simpler compared to that of OFR185 mode. According to Eq. 3, under OFR254 mode, the three parameters potentially affecting the OH_{exp} are OHR_{ext}, input O₃ concentration, and r_{O3}. The detailed ranges of each experimental condition for different datasets are listed in Table S4. We found that compared to Figure 1, the data points in Figure 4 were more scattered. Most of the R² values in Figure 4 were below 0.9, indicating that using SO₂ as the OHR_{ext} source, the estimation of OH_{exp} (using Eq. 3) under the OFR254 mode performed not as well as those under the OFR185 mode (using Eq. 2). Firstly, we investigated the impacts of OHR_{ext}. Figure 4a1 showed the regression results of OH_{exp, est} and $OH_{exp, dec}$ when OHR_{ext} ranged from 5 to 14 s⁻¹. The parameters x-z (FP_{IOHR, 254}; IOHR: low external OHR) (Table S6) were obtained by fitting Eq. 3 to OH_{exp, dec}. In Figure 4a2, the same set of fitted parameters FP_{IOHR, 254} from Figure 4a1 were used for a wider range of OHR_{ext} (5-21 s⁻¹). From the regression results (slopes of 1.050 and 1.024, R² of 0.890 and 0.894), the same set of parameters yielded similar estimation performance for OH_{exp} despite a wider range of OHR_{ext} in Figure 4a2 compared to that of Figure 4a1. At the same time, these results were not much different from those (slope = 1.071, $R^2 = 0.891$) using a re-fitted set of parameters (FP_{eOHR}, 254; eOHR: extended external OHR) for the wider range of OHR_{ext} (Figure 4a3). Even though the correlation was not as good as those in the OFR185 mode, approximately 85 % of the data points agreed within a factor of 2. We did not further extend the OHR_{ext} to values as high as those in the OFR185 mode as discussed above, since the OFR254 mode was much less oxidative and might not be suitable for simulating the oxidation chemistry of extremely high OHR_{ext} as those from direct emissions.

Similarly good correlations were observed when we only used the fitted parameters (FP_{103, 254} and FP_{mrO3, 254}, respectively; IO₃: low O_{3, in}, mrO₃: medium rO₃) from narrow ranges of input O₃ concentration and r_{O3} (Figure 4b1 and Figure 4c1, respectively) to reconstruct the OH_{exp, est} values with extended ranges of these experimental conditions (Figure 4b2 and Figure 4c2, respectively). Such correlations were as good as those with re-fitted parameters (FP_{eO3, 254} and FP_{erO3, 254}, respectively; eO₃: extended O_{3, in}, erO₃: extended rO₃) from data points in the extended ranges of O₃ concentration and r_{O3} (Figure 4b3 and Figure 4c3, respectively). These observations thus indicate that under the OFR254 mode, when OHR_{ext}, O_{3, in}, and r_{O3} vary within certain ranges (5–21 s⁻¹, 6.46×10^{13} – 4.8×10^{14} molecules cm⁻³, and 0.61–0.99, respectively), Eq. 3 can be used to estimate OH radical levels reasonably well using the fitting parameters (*x*–*z*) obtained from a narrower range of data points.

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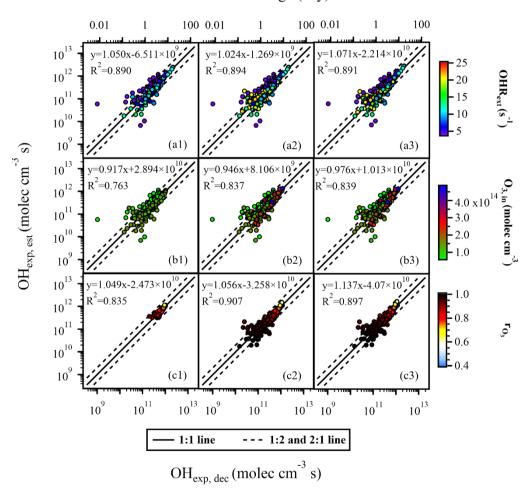


Figure 4: The regression results of $OH_{exp,\ est}$ and $OH_{exp,\ dec}$ when variations occurred in (a1–a3) OHR_{ext} , (b1–b3) input O_3 concentration, and (c1–c3) r_{O3} . Compared to panels a1, b1, and c1, panels a2, b2, and c2 respectively incorporated additional data points with extended OHR_{ext} , $O_{3,\ in}$, and r_{O3} values, but still utilized the fitting parameters $FP_{IOHR,\ 254}$, $FP_{IO3,\ 254}$, and $FP_{mrO3,\ 254}$ obtained from the lower or medium condition range to estimate $OH_{exp,\ est}$. In panels a3, b3, and c3, all data points within the extended

condition range were used to re-fit the parameters x–z, and the resulting FP_{eOHR}, 254, FP_{eO3}, 254, and FP_{erO3}, 254 were employed to estimate OH_{exp, est}.

Figure 5a and Figure 5b depicted the correlation between $OH_{exp, est}$ estimated from Eq. 3 and $OH_{exp, dec}$ calculated from Eq. 1 with SO_2 and CO as OHR_{ext} sources, respectively. When using SO_2 as the OHR_{ext} source, approximately 86 % of the data points agreed within a factor of 2 (Figure 5a). Similar to the case of OFR185, when CO was used as the OHR_{ext} source, the data points were more scattered, with the percentage of data points within a factor of 2 dropping to only about 64 % (Figure 5b). Figure 5c included data points using both SO_2 and CO as the OHR_{ext} sources. Overall, regardless of the OHR_{ext} source, when r_{O3} was higher than 0.93, which meant a low UV intensity, the majority of data points for $OH_{exp, est}$ and $OH_{exp, dec}$ differed by a factor of two or more. It is therefore recommended that when using the OFR254 mode, too low lamp power settings, for example, UV lamp voltage below 1.5V should be avoided in the case of our study.

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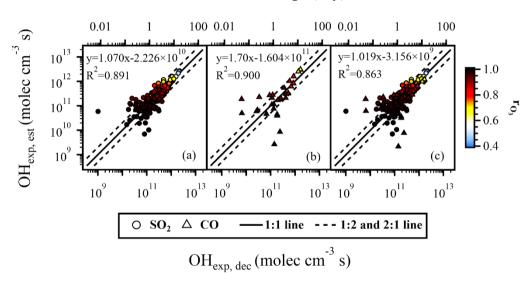


Figure 5: The regression results of $OH_{exp, dec}$ and $OH_{exp, est}$ in the OFR254 mode with (a) SO_2 and (b) CO as OHR_{ext} sources. (c) Results from all experiments (SO_2 and CO) in the OFR254 mode.

4 Conclusions

A series of OH_{exp} estimation experiments using the PAM-OFR were conducted in OFR185 and OFR254 modes to explore the applicability of the empirical equations under a wide range of conditions. The results indicate that for OFR185 mode, when varying the residence time, water vapor mixing ratio, and output O₃ concentration (as a surrogate for UV intensity) within certain ranges, the empirical equation (Eq. 2) for OH_{exp} proves to be effective in estimating OH_{exp}. Unless there is a significant change in OHR_{ext}, such as transitioning from ambient conditions to emission source conditions, there is no need to re-fit the parameters *a*–*f* in the estimation equation to estimate OH_{exp}. Compared with OFR254 mode, the consistency between OH_{exp}, est and OH_{exp}, dec in OFR185 mode is better. For the OFR254 mode, when OHR_{ext}, input O₃ concentration, and r_{O3} vary within

408 certain ranges, the empirical equation (Eq. 3) can be used to estimate OH_{exp} reasonably well using the parameters x-z obtained 409 from a narrower range of data points. It is important to note that for the OFR185 mode, the above conclusions are valid only 410 if one already has a set of a-f values that are appropriate for the specific UV lamps being used, as the I_{185} : I_{254} that affects the OH_{exp} is lamp-specific. For a PAM-OFR that employs a different Hg lamp, a series of calibration experiments should be 411 412 conducted in any case. Alternatively, based on the research by Rowe et al. (2020), the exponential relationship between the a-413 f values and the I_{185} : I_{254} could be used to first obtain a set of a-f values suitable for the UV lamps being used. 414 To obtain reliable estimates of OH_{exp} using Eqs. 2 and 3 for the OFR185 mode or OFR254 mode, respectively, it is desirable to have sufficient data points (that is, OH_{exp, dec} from trace-gas decay experiments) to fit the parameters for the calculation of 415 OH_{exp. est}. There is currently no consensus on how many data points in trace-gas decay experiments are enough for reliable 416 417 fitted parameters, which could be important for in-situ OH_{exp} estimation in field studies where a limited number of experiments 418 are done to reduce downtime. We aim to address this by random sampling from the data points in our experiments and 419 determine the minimum number of experiments that are needed to obtain reliable OH_{exp}. 420 For OFR185 mode, we first used randomly selected N data points from the 175 data points presented previously to fit the 421 parameters (a-f) using Eq. 2. The fitted parameters were then used to reconstruct $OH_{exp, ext}$ for all the 175 data points. The 422 OH_{exp. est} values were then compared with the corresponding 175 OH_{exp. dec} values. This procedure was repeated 10 times for 423 each N, with N starting from 7 till approximately 50 (Figure 6a). The average R², slope, and intercept from the 10 attempts 424 were then shown as a function of N for experiments with SO₂ only (Figure 6a) and those with SO₂ and CO (Figure 6b). It can 425 be observed that around 30 data points are needed for experiments with SO₂ only while around 20 data points are needed to 426 have stable R² values and slopes when using both SO₂ and CO. For OFR254 mode, the same procedure was applied to the 241 427 data points. It was not surprising that the results were a lot more scattered (Figure 6c and Figure 6d) compared to those for 428 OFR185 mode given their performance shown in the previous section. Nevertheless, our analysis suggests that around 25 data 429 points are needed to obtain reliable OH_{exp, est} for OFR254 mode, whether SO₂ alone (Figure 6c) or SO₂ and CO (Figure 6d) are 430 used for the trace-gas decay experiments. Therefore, despite the limitation that this practice only randomly samples the data points without considering the range of any experimental conditions, our analysis suggests that 20–30 data points are normally 431

needed to obtain reliable OH_{exp} for both OFR185 and OFR254 modes.

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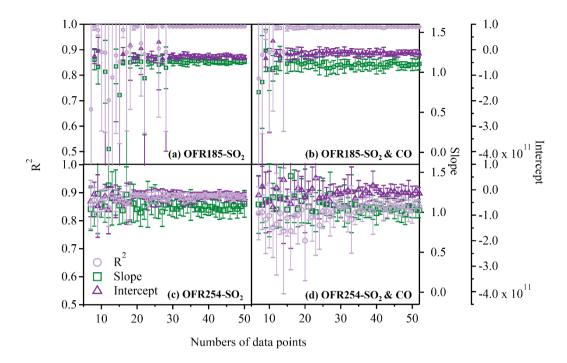


Figure 6: The regression results of $OH_{exp, dec}$ and $OH_{exp, est}$ (characterized by the R^2 , slope, and intercept) when different numbers of data points were chosen. (a) SO_2 as OHR_{ext} source in OFR185 mode, (b) SO_2 or CO as OHR_{ext} source in OFR185 mode, (c) SO_2 as OHR_{ext} source in the OFR254 mode, and (d) SO_2 or CO as OHR_{ext} source in the OFR254 mode.

Our study suggests that the OH_{exp, est} estimated from the empirical equations agrees better with OH_{exp, dec} for the OFR185 (Figure 3) than for the OFR254 mode (Figure 5). This can be understood from the perspective of OH generation and its consumption by OHR_{ext} (Li et al., 2015). For the OFR185 mode, there are two pathways to generate OH radicals: the photolysis of H₂O and the photolysis of O₃. For the OFR254 mode, the main pathway for OH radical generation is solely the photolysis of O₃. Consequently, when OHR_{ext} changes, the disruption to OH_{exp} in the system is more significant in the case of the OFR254 mode, while the OH_{exp} in the OFR185 mode remains more stable. In addition, pseudo-first-order kinetics between OH radicals and SO₂ or CO is assumed, with [OH] being at a pseudo-steady state. Yet, the relatively low OH radical generation capacity in the OFR254 mode might not necessarily always fulfil such an assumption, leading to higher uncertainties for estimating OH_{exp}. Therefore, the OFR185 mode offers certain advantages such as relatively high OH_{exp}, more accurate OH_{exp} estimation, as well as no external input of O₃ needed. However, for substances that exhibit strong absorption at the wavelength of 185 nm and are prone to photolysis, such as aromatic species (Peng et al., 2016), using the OFR254 mode is a better choice. For users of other OFRs (non-PAM-OFR) who would like to apply the conclusions above, at least two conditions must be met: (1) the concentration of [OH] within the OFR should remain stable, and (2) the assumption of plug flow conditions is acceptable, allowing for the neglect of differences in the actual RTD, heterogeneity in the UV light intensity and the concentration of

- 451 radicals/oxidants at different points within the reactor, which are caused by different designs of reactors (such as wall materials,
- 452 shapes, or volumes).

453 **Data availability**

- 454 The data shown in the paper are available on request from the corresponding authors (huangdd@saes.sh.cn and
- 455 yongjieli@um.edu.mo).

456 Supplement link

457 The supplement related to this article is available online at:

458 Author contribution

- 459 QL, DDH, and YJL conceived and planned the experiments. QL and YW carried out the experiments. QL, DDH, and YJL
- 460 analysed the data and took the lead in writing the paper. QL, DDH, YJL, ATL, and XC contributed to the interpretation of the
- 461 results. SL, LZ, CYH, ST, QC, KIH, HW, KMM, and CH provided significant input during the revision of the manuscript. All
- authors provided feedback on the paper.

463 Competing interests

464 The authors declare no competing interests.

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471 References

- 472 Atkinson, R. and Arey, J.: Atmospheric degradation of volatile organic compounds, Chem. Rev., 103, 4605-4638, doi:
- 473 10.1021/cr0206420, 2003.

- 474 Atkinson, R., Baulch, D. L., Cox, R. A., Crowley, J. N., Hampson, R. F., Hynes, R. G., Jenkin, M. E., Rossi, M. J., and Troe,
- 475 J.: Evaluated kinetic and photochemical data for atmospheric chemistry: Volume II-gas phase reactions of organic species,
- 476 Atmos. Chem. Phys., 6, 3625-4055, doi: 10.5194/acp-6-3625-2006, 2006.
- 477 Barmet, P., Dommen, J., DeCarlo, P., Tritscher, T., Praplan, A., Platt, S., Prévôt, A., Donahue, N., and Baltensperger, U.: OH
- 478 clock determination by proton transfer reaction mass spectrometry at an environmental chamber, Atmospheric Measurement
- 479 Techniques, 5, 647-656, 2012.
- 480 Burkholder, J. B., Sander, S. P., Abbatt, J. P. D., Barker, J. R., Cappa, C., Crounse, J. D., Dibble, T. S., Huie, R. E., Kolb, C.
- 481 E., Kurylo, M. J., Orkin, V. L., Percival, C. J., Wilmouth, D. M., and Wine, P. H.: Chemical kinetics and photochemical data
- 482 for use in atmospheric studies, Evaluation Number 19, Pasadena, CA: Jet Propulsion Laboratory, National Aeronautics and
- 483 Space Administration, http://jpldataeval.jpl.nasa.gov, 2020.
- 484 Cao, J., Wang, Q., Li, L., Zhang, Y., Tian, J., Chen, L. A., Ho, S. S. H., Wang, X., Chow, J. C., and Watson, J. G.: Evaluation
- 485 of the oxidation flow reactor for particulate matter emission limit certification, Atmos. Environ., 224, 117086, doi:
- 486 10.1016/j.atmosenv.2019.117086, 2020.
- 487 Cheng, X., Chen, Q., Jie Li, Y., Zheng, Y., Liao, K., and Huang, G.: Highly oxygenated organic molecules produced by the
- 488 oxidation of benzene and toluene in a wide range of OH exposure and NOx conditions, Atmos. Chem. Phys., 21, 12005-12019,
- 489 doi: 10.5194/acp-21-12005-2021, 2021.
- 490 Cheng, X., Li, Y. J., Zheng, Y., Liao, K., Koenig, T. K., Ge, Y., Zhu, T., Ye, C., Qiu, X., and Chen, Q.: Oxygenated organic
- 491 molecules produced by low-NOx photooxidation of aromatic compounds: contributions to secondary organic aerosol and steric
- 492 hindrance, Atmos. Chem. Phys., 24, 2099-2112, doi: 10.5194/acp-24-2099-2024, 2024.
- 493 Clerc, M. and Barat, F.: Kinetics of CO formation studied by far-UV flash photolysis of CO2, J. Chem. Phys., 46, 107-110,
- 494 doi: 10.1063/1.1840358, 1967.
- 495 Cocker, D. R., Flagan, R. C., and Seinfeld, J. H.: State-of-the-art chamber facility for studying atmospheric aerosol chemistry,
- 496 Environ. Sci. Technol., 35, 2594-2601, doi: 10.1021/es0019169, 2001.
- 497 Cohen, N. and Heicklen, J.: The Oxidation of Inorganic Non-metallic Compounds, in: Reactions of Non-Metallic Inorganic
- 498 Compounds, Compr. Chem. Kinet., 1-137, doi: 10.1016/s0069-8040(08)70303-0, 1972.
- 499 Ehhalt, D. H.: Photooxidation of trace gases in the troposphere Plenary Lecture, Phys. Chem. Chem. Phys., 1, 5401-5408, doi:
- 500 10.1039/A905097C, 1999.
- 501 Friedman, B. and Farmer, D. K.: SOA and gas phase organic acid yields from the sequential photooxidation of seven
- 502 monoterpenes, Atmos. Environ., 187, 335-345, doi: 10.1016/j.atmosenv.2018.06.003, 2018.
- 503 Fuchs, H., Tan, Z., Lu, K., Bohn, B., Broch, S., Brown, S. S., Dong, H., Gomm, S., Häseler, R., and He, L.: OH reactivity at a
- 504 rural site (Wangdu) in the North China Plain: contributions from OH reactants and experimental OH budget, Atmos. Chem.
- 505 Phys., 17, 645-661, doi: 10.5194/acp-17-645-2017, 2017.
- 506 George, I. J., Vlasenko, A., Slowik, J. G., Broekhuizen, K., and Abbatt, J. P. D.: Heterogeneous oxidation of saturated organic
- 507 aerosols by hydroxyl radicals: uptake kinetics, condensed-phase products, and particle size change, Atmos. Chem. Phys., 7,
- 508 4187-4201, doi: 10.5194/acp-7-4187-2007, 2007.
- 509 Hildebrandt, L., Donahue, N. M., and Pandis, S. N.: High formation of secondary organic aerosol from the photo-oxidation of
- 510 toluene, Atmos. Chem. Phys., 9, 2973-2986, doi: 10.5194/acp-9-2973-2009, 2009.
- 511 Kang, E., Root, M. J., Toohey, D. W., and Brune, W. H.: Introducing the concept of potential aerosol mass (PAM), Atmos.
- 512 Chem. Phys., 7, 5727-5744, doi: 10.5194/acp-7-5727-2007, 2007.
- Lambe, A. T., Ahern, A. T., Williams, L. R., Slowik, J. G., Wong, J. P. S., Abbatt, J. P. D., Brune, W. H., Ng, N. L., Wright,
- 514 J. P., and Croasdale, D. R.: Characterization of aerosol photooxidation flow reactors: heterogeneous oxidation, secondary
- 515 organic aerosol formation and cloud condensation nuclei activity measurements, Atmos. Meas. Tech., 4, 445-461, doi:
- 516 10.5194/amt-4-445-2011, 2011.

- 517 Li, R., Palm, B. B., Ortega, A. M., Hlywiak, J., Hu, W., Peng, Z., Day, D. A., Knote, C., Brune, W. H., and De Gouw, J. A.:
- 518 Modeling the radical chemistry in an oxidation flow reactor: Radical formation and recycling, sensitivities, and the OH
- 519 exposure estimation equation, J. Phys. Chem. A, 119, 4418-4432, doi: 10.1021/jp509534k, 2015.
- 520 Liao, K., Chen, Q., Liu, Y., Li, Y. J., Lambe, A. T., Zhu, T., Huang, R.-J., Zheng, Y., Cheng, X., and Miao, R.: Secondary
- 521 organic aerosol formation of fleet vehicle emissions in China: Potential seasonality of spatial distributions, Environ. Sci.
- 522 Technol., 55, 7276-7286, doi: 10.1021/acs.est.0c08591, 2021.
- 523 Liu, J., Chu, B., Chen, T., Liu, C., Wang, L., Bao, X., and He, H.: Secondary organic aerosol formation from ambient air at an
- 524 urban site in Beijing: effects of OH exposure and precursor concentrations, Environ. Sci. Technol., 52, 6834-6841, doi:
- 525 10.1021/acs.est.7b05701, 2018.
- 526 Liu, T., Clegg, S. L., and Abbatt, J. P.: Fast oxidation of sulfur dioxide by hydrogen peroxide in deliquesced aerosol particles,
- 527 Proceedings of the National Academy of Sciences, 117, 1354-1359, 2020.
- 528 Lou, S., Holland, F., Rohrer, F., Lu, K., Bohn, B., Brauers, T., Chang, C. C., Fuchs, H., Häseler, R., and Kita, K.: Atmospheric
- 529 OH reactivities in the Pearl River Delta-China in summer 2006: measurement and model results, Atmos. Chem. Phys., 10,
- 530 11243-11260, doi: 10.5194/acp-10-11243-2010, 2010.
- 531 Lu, K., Zhang, Y., Su, H., Brauers, T., Chou, C. C., Hofzumahaus, A., Liu, S. C., Kita, K., Kondo, Y., and Shao, M.: Oxidant
- 532 (O3 + NO2) production processes and formation regimes in Beijing, J. Geohys. Res.: Atmospheres, 115, doi:
- 533 10.1029/2009JD012714, 2010.
- Mao, J., Ren, X., Brune, W. H., Olson, J. R., Crawford, J. H., Fried, A., Huey, L. G., Cohen, R. C., Heikes, B., and Singh, H.
- 535 B.: Airborne measurement of OH reactivity during INTEX-B, Atmos. Chem. Phys., 9, 163-173, doi: org/10.5194/acp-9-163-
- 536 2009, 2009.
- 537 Nakashima, Y., Kamei, N., Kobayashi, S., and Kajii, Y.: Total OH reactivity and VOC analyses for gasoline vehicular exhaust
- 538 with a chassis dynamometer, Atmos. Environ., 44, 468-475, doi: 10.1016/j.atmosenv.2009.11.006, 2010.
- 539 Ono, R., Nakagawa, Y., Tokumitsu, Y., Matsumoto, H., and Oda, T.: Effect of humidity on the production of ozone and other
- radicals by low-pressure mercury lamps, J Photochem Photobiol, A, 274, 13-19, doi: 10.1016/j.jphotochem.2013.09.012, 2014.
- Palm, B. B., de Sá, S. S., Day, D. A., Campuzano-Jost, P., Hu, W., Seco, R., Sjostedt, S. J., Park, J.-H., Guenther, A. B., and
- 542 Kim, S.: Secondary organic aerosol formation from ambient air in an oxidation flow reactor in central Amazonia, Atmospheric
- 543 Chemistry and Physics, 18, 467-493, 2018.
- 544 Peng, Z. and Jimenez, J. L.: KinSim: a research-grade, user-friendly, visual kinetics simulator for chemical-kinetics and
- 545 environmental-chemistry teaching, J. Chem. Educ, 96, 806–811, doi: 10.1021/acs.jchemed.9b00033, 2019.
- Peng, Z. and Jimenez, J. L.: Radical chemistry in oxidation flow reactors for atmospheric chemistry research, Chem. Soc. Rev.,
- 547 49, 2570-2616, doi: 10.1039/C9CS00766K, 2020.
- 548 Peng, Z., Day, D. A., Stark, H., Li, R., Lee-Taylor, J., Palm, B. B., Brune, W. H., and Jimenez, J. L.: HOx radical chemistry
- 549 in oxidation flow reactors with low-pressure mercury lamps systematically examined by modeling, Atmos. Meas. Tech., 8,
- 550 4863-4890, doi: 10.5194/amt-8-4863-2015, 2015.
- Peng, Z., Day, D. A., Ortega, A. M., Palm, B. B., Hu, W., Stark, H., Li, R., Tsigaridis, K., Brune, W. H., and Jimenez, J. L.:
- 552 Non-OH chemistry in oxidation flow reactors for the study of atmospheric chemistry systematically examined by modeling,
- 553 Atmos. Chem. Phys., 16, 4283-4305, doi: 10.5194/acp-16-4283-2016, 2016.
- 554 Rowe, J. P., Lambe, A. T., and Brune, W. H.: Effect of varying the λ=185 and 254 nm photon flux ratio on radical generation
- 555 in oxidation flow reactors, Atmos. Chem. Phys., 20, 13417-13424, doi: 10.5194/acp-20-13417-2020, 2020.
- 556 Seinfeld, J. H. and Pandis, S. N.: Atmospheric chemistry and physics: from air pollution to climate change, John Wiley &
- 557 Sons, Inc., ISBN 9781119221166, 2016.

- 558 Stone, D., Whalley, L. K., and Heard, D. E.: Tropospheric OH and HO2 radicals: field measurements and model comparisons,
- 559 Chem. Soc. Rev., 41, 6348-6404, doi: 10.1039/C2CS35140D, 2012.
- 560 Tan, Z., Fuchs, H., Lu, K., Hofzumahaus, A., Bohn, B., Broch, S., Dong, H., Gomm, S., Häseler, R., and He, L.: Radical
- 561 chemistry at a rural site (Wangdu) in the North China Plain: observation and model calculations of OH, HO2 and RO2 radicals,
- 562 Atmos. Chem. Phys., 17, 663-690, doi: 10.5194/acp-17-663-2017, 2017.
- 563 Tan, Z., Rohrer, F., Lu, K., Ma, X., Bohn, B., Broch, S., Dong, H., Fuchs, H., Gkatzelis, G. I., and Hofzumahaus, A.:
- 564 Wintertime photochemistry in Beijing: observations of RO x radical concentrations in the North China Plain during the BEST-
- 565 ONE campaign, Atmos. Chem. Phys., 18, 12391-12411, doi: 10.5194/acp-18-12391-2018, 2018.
- 566 Wang, N., Zannoni, N., Ernle, L., Bekö, G., Wargocki, P., Li, M., Weschler, C. J., and Williams, J.: Total OH reactivity of
- 567 emissions from humans: in situ measurement and budget analysis, Environ. Sci. Technol., 55, 149-159, doi:
- 568 10.1021/acs.est.0c04206, 2020.

579

- 569 Wang, X., Liu, T., Bernard, F., Ding, X., Wen, S., Zhang, Y., Zhang, Z., He, Q., Lü, S., and Chen, J.: Design and
- 570 characterization of a smog chamber for studying gas-phase chemical mechanisms and aerosol formation, Atmos. Meas. Tech.,
- 571 7, 301-313, doi: 10.5194/amt-7-301-2014, 2014.
- 572 Yang, Y., Shao, M., Keßel, S., Li, Y., Lu, K., Lu, S., Williams, J., Zhang, Y., Zeng, L., and Nölscher, A. C.: How the OH
- 573 reactivity affects the ozone production efficiency: case studies in Beijing and Heshan, China, Atmos. Chem. Phys., 17, 7127-
- 574 7142, doi: 10.5194/acp-17-7127-2017, 2017.
- 575 You, X., Wang, H., Goos, E., Sung, C.-J., and Klippenstein, S. J.: Reaction kinetics of CO + HO2→ products: ab initio
- transition state theory study with master equation modeling, J. Phys. Chem. A, 111, 4031-4042, doi: 10.1021/jp067597a, 2007.
- 577 Zhang, Z., Xu, W., Lambe, A. T., Hu, W., Liu, T., and Sun, Y.: Insights Into Formation and Aging of Secondary Organic
- 578 Aerosol From Oxidation Flow Reactors: A Review, Curr. Pollut. Rep., 1-14, doi: 10.1007/s40726-024-00309-7, 2024.