



1 **Mixing state, spatial distribution, sources and**
2 **photochemical enhancement to sulfate formation of black**
3 **carbon particles in the Arctic Ocean during summer**

4

5 Longquan Wang^{1,2}, Jinpei Yan³, Afeng Chen^{1,4}, Bei Jiang^{1,5}, Fange Yue¹, Xiawei Yu¹,

6 Zhouqing Xie^{1,6*}

7

8 ¹Anhui Key Laboratory of Polar Environment and Global Change, Department of Environmental Science
9 and Engineering, University of Science and Technology of China, Hefei 230026, China.

10 ²Department of Carbon Neutral Science and Engineering, Anhui University of Science and Technology,
11 Hefei 230026, China.

12 ³Third Institute of Oceanography, Ministry of Natural Resources, Xiamen 361005, China.

13 ⁴Engineering and Technological Research Centre of National Disaster Relief Equipment, Army Logistics
14 Academy, Chongqing, 401331, China.

15 ⁵College of Ecology and Environment, Xinjiang University, Urumqi, 830017, PR China.

16 ⁶State Key Laboratory of Fire Science, University of Science and Technology of China, Hefei, 230026,
17 China.

18 *Correspondence to:* Zhouqing Xie (zqxie@ustc.edu.cn)



19 **Abstract.** Black carbon heats the atmosphere by absorbing solar radiation and regulates the radiation
20 balance of the Earth. Specifically in the Arctic region, black carbon accelerates Arctic warming by
21 simultaneously altering surface albedo. Nonetheless, assessing the climatic impacts of black carbon
22 aerosols in the Arctic is challenging due to their considerable variability in temporal and spatial
23 distribution, sources, and chemical composition. Black carbon particles (0.2-2 μ m) in the Arctic Ocean
24 were investigated using a ship-based single particle aerosol mass spectrometer from July to August 2017.
25 In the central Arctic Ocean, near the Norwegian Sea-Iceland and the North Atlantic, biomass combustion
26 is the predominant source of black carbon particles, constituting over 50%, with a particularly high
27 contribution exceeding 70% in the central Arctic Ocean. Within the Chukchi Sea region, terrestrial
28 transport from mid and low latitudes emerges as the primary source of black carbon particles,
29 representing over 50%, with biomass combustion and anthropogenic pollution sources each contributing
30 around 25%. Near Svalbard, biomass combustion sources and terrigenous transport stand out as the
31 primary sources of black carbon particles, with their contributions being comparable. Furthermore, the
32 ratio of sulfate to nitrate in black carbon particles was notably higher compared to that in sea salt particles.
33 This ratio increased with elevated black carbon content and sunlight intensity, suggesting that Arctic
34 black carbon particles substantially facilitated sulfate formation through photochemical processes. Such
35 interactions could potentially modify the mixing state of Arctic black carbon particles and their radiative
36 impacts.

37

38 **1 Introduction**

39 Black carbon, an amorphous form of carbon produced by the incomplete combustion of carbon-
40 containing fuels or biomass, plays a crucial role in atmospheric physical chemistry by absorbing solar
41 radiation and influencing Earth's radiation balance. Additionally, the warming effect of black carbon
42 modifies the vertical stability of the atmosphere and alters cloud distribution, which further impacts the
43 radiation balance (Ramanathan & Carmichael, 2008). The deposition of black carbon on ice and snow
44 surfaces significantly reduces their albedo, accelerating the melting process. This is particularly
45 pronounced in the Arctic region, where black carbon contributes to warming not only by heating the
46 atmosphere but also by changing the surface albedo and through other direct and indirect radiative effects
47 (Flanner et al., 2007; Serreze & Barry, 2011). Arctic black carbon aerosol, characterized as a transportable



48 species with various potential sources, exhibits distinct spatial and seasonal distribution patterns (Matsui
49 et al., 2022). These variations are critical for assessing its impact on radiative forcing. Furthermore, the
50 mixing state of particles containing black carbon significantly influences their radiative effects (Jacobson,
51 2001), making detailed observations and analyses essential.

52 The mass concentration of black carbon aerosols in the Arctic near-surface atmosphere exhibits
53 significant seasonal variability, typically higher in winter and spring, and lower in summer and autumn
54 (Qi et al., 2017; Sharma et al., 2006; Sharma et al., 2004; Sharma et al., 2019). Influenced by atmospheric
55 transport patterns, the sources and vertical structures of black carbon aerosols also vary seasonally.
56 During winter and spring, the Arctic near-surface atmosphere is predominantly affected by transport from
57 higher latitudes, facilitating the movement of black carbon from Siberia and Europe through low-altitude
58 pathways to the bottom of the Arctic troposphere. Conversely, black carbon from mid-latitudes reaches
59 the middle and upper troposphere of the Arctic via long-distance transport. In summer, increased
60 precipitation in middle and high latitudes, which impedes long-distance transport, shifts the primary
61 source of atmospheric black carbon to biomass burning within the Arctic and neighboring regions such
62 as Siberia and Alaska (Evangelidou et al., 2016; Matsui et al., 2022; Wang et al., 2014). Although
63 atmospheric concentrations of black carbon are lower in summer than in winter, the higher frequency of
64 summer precipitation leads to quicker removal rates and higher deposition fluxes of atmospheric black
65 carbon (Willis et al., 2018). Observational data on atmospheric black carbon aerosol concentrations in
66 the Arctic also demonstrate notable spatiotemporal variations. Long-term continuous measurements at
67 Barrow and Alert stations in the Arctic revealed that the average concentration of black carbon aerosols
68 from 1980 to 2003 was approximately 25 ng/m³ (Sharma et al., 2004). Meanwhile, the average annual
69 concentration measured at Zeppelin Station from 1998 to 2007 was 39 ng/m³ (Eleftheriadis et al., 2009),
70 significantly higher than the former. Additionally, the spatial distribution of black carbon in Arctic ice
71 and snow varies, with concentrations decreasing from the Russian Arctic, through the Canadian and
72 Alaskan Arctic, to the Arctic Ocean and Greenland, highlighting a significant decline from the Arctic
73 coastal regions to the central Arctic Ocean (Dou & Xiao, 2016). These temporal and spatial discrepancies
74 in Arctic black carbon aerosols underline the importance of long-term, multi-regional observations in the
75 Arctic.

76 Evaluating the climatic impact of black carbon aerosols in the atmosphere and on snow cover is



77 crucial for understanding Arctic climate change. Unlike greenhouse gases such as carbon dioxide and
78 methane, black carbon aerosols possess a broader absorption spectrum, capable of absorbing radiation
79 across infrared to ultraviolet wavelengths (Bond & Bergstrom, 2006). In the IPCC 5th Assessment Report,
80 the direct radiative forcing of black carbon aerosols was estimated at 0.4 W/m^2 , with a range of 0.05 to
81 0.8 W/m^2 . The effective radiative forcing, which includes direct radiative impacts along with climate
82 feedbacks and rapid adjustments, was initially estimated at $-0.45 \pm 0.5 \text{ W/m}^2$. This estimate has since
83 been revised to 0.063 W/m^2 , ranging from -0.28 to 0.42 W/m^2 , reflecting a deeper understanding of
84 black carbon's effects on clouds and water vapor. Research indicates that the direct radiative forcing from
85 black carbon aerosols originating from fossil fuels is comparable to 78% of that from carbon dioxide,
86 while those from biomass burning equate to about 58% of carbon dioxide's effect (Hansen et al., 2005).
87 Considering indirect effects like changes in ice and snow albedo and cloud properties, the black and
88 organic carbon components from fossil fuel soot exert a net positive global radiative forcing. In contrast,
89 those from biomass combustion have a net negative global radiative forcing (Hansen et al., 2005). In the
90 Arctic, black carbon aerosols starkly contrast with the bright ice and snow surfaces, enhancing their solar
91 radiation absorption and leading to higher direct radiative forcing. Studies suggest that in the Arctic, the
92 warming effect of black carbon aerosols surpasses the cooling effect of reflective aerosols, making them
93 a potent warming agent in the region (Quinn et al., 2007). Beyond atmospheric warming, black carbon
94 in snow can intensify climate change by modifying the albedo feedback of snow and ice. The deposition
95 of black carbon on snow surfaces not only increases heat absorption but also, due to near-surface
96 temperature inversions that restrict energy dispersal in the Arctic, this absorbed heat further warms the
97 near-surface atmosphere. This process accelerates ice and snow melt and alters surface albedo, thereby
98 amplifying Arctic warming through the ice-albedo feedback mechanism, a significant factor in Arctic
99 amplification (Clarke & Noone, 1985; Dou & Xiao, 2016). The estimated annual mean radiative forcing
100 of black carbon aerosols from various sources on Arctic ice and snow is approximately 0.17 W/m^2 (Dou
101 & Xiao, 2016).

102 In the Arctic region, the mixing state of black carbon aerosols from different sources significantly
103 influences their radiative effects (Jacobson, 2001; Matsui et al., 2022). The internal mixing of black
104 carbon aerosols with scattering aerosols such as sulfate and organic carbon can substantially alter the
105 optical properties of black carbon, increasing its positive radiative forcing (Chung & Seinfeld, 2002). It



106 has been observed that anthropogenic black carbon aerosols from Asia generally have a thicker coating
107 compared to those from Europe, Siberia, and North America. This thicker coating is attributed to the
108 rapid aging of Asian aerosols upon emission, often accompanied by condensable gases. Moreover, Asian
109 anthropogenic black carbon typically travels at higher altitudes and over longer distances, resulting in a
110 longer residence time and more thorough aging process (Matsui et al., 2022). Black carbon aerosols of
111 different origins within the same region also exhibit variations in their mixing state. For instance, black
112 carbon aerosols from biomass combustion in Siberia and North America tend to have a larger proportion
113 and thicker coatings compared to those from anthropogenic sources in these areas. This difference is
114 largely due to the transport of biomass combustion-derived black carbon to the Arctic occurring primarily
115 in summer, which facilitates a faster aging rate than in winter (Matsui et al., 2022). Anthropogenic black
116 carbon from Asia and biomass-burning-derived black carbon aerosols from Siberia and North America
117 exhibit stronger positive radiative effects (Matsui et al., 2022). Additionally, the mixing state of black
118 carbon aerosols affects nucleation and wet clearance processes, thereby indirectly influencing their
119 radiative effects (Ching et al., 2012; Ching et al., 2018). These findings highlight that the source and
120 mixing state of Arctic black carbon aerosols represent key uncertainties in assessing their radiative
121 impacts, and that field observations are crucial for accurately evaluating their climate effects.

122 The condensation of secondary species like sulfate, nitrate, and organic matter on the surfaces of
123 black carbon particles significantly alters their mixing state (Ault et al., 2010). Despite this, few studies
124 have explored the secondary processes occurring on black carbon aerosol surfaces. Black carbon is
125 known to actively participate in some of these secondary processes. Recent laboratory research indicates
126 that black carbon may catalyze the formation of sulfate, significantly contributing to the growth of
127 secondary inorganic components in urban haze in China (Zhang et al., 2020). Research has also
128 demonstrated that black carbon aerosols are photoactive, capable of releasing reactive oxygen species,
129 including excited oxygen molecules and hydroxyl radicals, into the atmosphere. These species potentially
130 facilitate the formation of sulfate and organic matter (Gehling & Dellinger, 2013; Li et al., 2019). Further
131 studies have confirmed that black carbon aerosols engage in photochemical processes that enhance
132 sulfate generation in urban settings (Zhang et al., 2021). However, such studies have not yet been
133 extended to the Arctic region. Given the Arctic's sensitivity to climate changes influenced by aerosol
134 species like black carbon and sulfate, this potential photochemical process could introduce significant



135 uncertainties in assessing the local aerosol physicochemical properties and their radiative effects.

136

137 **2 Materials and Methods**

138 **2.1 Research region and instruments**

139 Observations were conducted aboard the R/V "Xuelong" during the 8th Chinese Arctic Expedition
140 Research Cruise, which traversed a significant portion of the Arctic Ocean, spanning from 56.2° to
141 84.6°N latitude and 169.4° to 46.9°W longitude. The cruise took place from July 30 to August 27, 2017,
142 and was segmented into five distinct phases as described in Wang et al. (2022) for detailed analysis based
143 on the ship's location, as depicted in Figure 1, and the specific sampling times and locations for each
144 phase are detailed in Table S1 and summarized as follows: Chukchi Sea section, denoted as Leg I, spans
145 from 22:00 on July 30th, 2017, to 19:00 on August 1st, 2017, covering latitudes from 66.0°N to 74.8°N
146 and longitudes from 169.4°W to 159.2°W. The high Arctic section, marked as Leg II, extends from 20:00
147 on August 10th to 5:00 on August 12th, 2017, ranging from 83.7°N to 84.6°N and from 132.0°E to
148 110.4°E. Svalbard Islands section, identified as Leg III, is from 2:00 on August 17th to 11:00 on August
149 19th, 2017, with latitudes from 82.6°N to 74.3°N and longitudes from 25.4°E to 2.1°E. Norwegian Sea
150 and Iceland section, referred to as Leg IV, occurs from 14:00 on August 23rd to 14:00 on August 25th,
151 2017, spanning from 67.0°N to 61.2°N and from 2.1°W to 25.8°W. Atlantic Ocean section, labeled as
152 Leg V, ranging from 14:00 on August 25th to 17:00 on August 27th, 2017, covers from 61.1°N to 56.2°N
153 and from 25.9°W to 46.9°W. The methodology for particle detection utilized the onboard Single Particle
154 Aerosol Mass Spectrometer (SPAMS), consistent with the techniques described by Li et al. (2011). Prior
155 to analysis, sampled particles were dried using a Nafion dryer to remove moisture. A PM_{2.5} collector was
156 employed to filter out particles larger than 2.5 μm. The fine particles were then drawn into the vacuum
157 system through a critical orifice, accelerated, and focused to form a narrow particle beam. These particles
158 were subsequently exposed to two continuous Nd:YAG diode lasers (532 nm) to determine their
159 aerodynamic diameter based on their velocity. Each particle was then ionized by an Nd:YAG laser (266
160 nm) to generate positive and negative fragment ions. The ionization laser maintained a power density of
161 1.55×10^8 W/cm². The resulting fragment ions were detected using a bipolar time-of-flight mass
162 spectrometer. For calibration of the SPAMS, polystyrene latex spheres (PSL Nanosphere Size Standards,
163 Duke Scientific Corp., Palo Alto) with diameters ranging from 0.2 to 2.0 μm were used. Additionally,



164 PbNO₃ particles with a diameter of 0.35 μm, generated by an aerosol generation and monitoring system
165 (AGM-1500, MSP Corporation, USA), were utilized for mass spectrum calibration. The sampling inlet,
166 connected to the monitoring instruments, was positioned on a mast 20 meters above the sea surface at
167 the bow of the vessel to minimize the influence of ship emissions. Sampling was conducted only while
168 the ship was in motion. A specific peak threshold was set at five units, allowing the instrument to record
169 a peak signal that surpassed this threshold, thus distinguishing it from background noise (<1 unit) in the
170 mass spectra (Zhang et al., 2021). The same instrument and samples were successfully used by Wang et
171 al. (2022) to investigate the mixing state, distribution, and photochemistry of iodine-containing particles
172 through specific identification conditions. In this study, we adopt a new perspective, focusing on black
173 carbon particles from the same samples, examining their sources, distribution, and relationships to sulfate
174 formation under distinct identification conditions, as described in detail in Section 2.2.

175 2.2 Data processing of SPAMS data

176 Particle size and mass spectra were analyzed using the YAADA software toolkit
177 (<http://www.yaada.org/>) and MATLAB (<http://www.mathworks.com>). Throughout the entire cruise, over
178 2,000,000 particles were sampled and sized using two continuous Nd:YAG diode lasers (532 nm). Nearly
179 290,000 particles were ionized using an Nd:YAG laser (266 nm), generating both positive and negative
180 ion mass spectrometry data during the cruise. Black carbon particles are known to produce a series of
181 characteristic signal peaks at integer multiples of 12 for the mass-to-charge ratio ($m/z = \pm 12, \pm 24, \pm 36,$
182 $\pm 48, \pm 60,$ etc.) in the spectra (Kollner et al., 2021; Qin & Prather, 2006; Roth et al., 2016; Schmidt et al.,
183 2017; Sierau et al., 2014). Therefore, the presence of several of these signal peaks concurrently in the
184 anionic or cationic mass spectrum of a particle suggests the presence of black carbon. In this study, we
185 established specific criteria to identify black carbon particles among all ionized particles (Beddows et al.,
186 2004). A particle is determined to contain black carbon if it meets one of the following conditions:

- 187 (1) The signal peaks of $m/z = 12, 24,$ and 36 appear simultaneously in the cationic mass spectrum.
- 188 (2) The signal peaks of $m/z = -24, -36,$ and -48 appear simultaneously in the anionic mass spectrum.

189 Using the specified method, approximately 80,000 particles were identified as containing black
190 carbon. The average spectrum of black carbon particles is displayed in Figure S1. The figure reveals that,
191 in addition to the peak for carbon signals, there are prominent peaks for ³⁹K⁺ and ⁴⁰Ca⁺ in the cation
192 spectrum, as well as ²⁶CN⁻/⁴²CNO⁻ and ⁹⁷HSO₄⁻ in the anion spectrum. To categorize these particles into



193 distinct groups based on the presence and intensity of ion peaks in their mass spectra, we employed the
194 Adaptive Resonance Theory neural network algorithm (ART-2a) (Song et al., 1999). The parameters set
195 for ART-2a included a vigilance factor of 0.6, a learning rate of 0.05, and a limit of 20 iterations. This
196 clustering approach is designed to include more than 95% of the total particles when used for
197 classification purposes. For the descriptive statistical analysis of the data, SPSS version 20 (IBM Inc.,
198 USA) was utilized.

199 2.3 Satellite data acquisition

200 In this study, the intensity of solar radiation was derived using the surface incoming shortwave flux
201 (SWGDN, in watts per square meter, W/m^2) with a 1-hour time resolution. This data was sourced from
202 the NASA Goddard Earth Sciences Data and Information Services Center (<https://disc.gsfc.nasa.gov/>)
203 for the specific times and locations relevant to our research.

204

205 3 Results and Discussion

206 3.1 Mixing state of black carbon particles

207 Numerous studies have established a clear link between the mixing state of black carbon particles
208 and their radiative impacts in the Arctic atmosphere. Black carbon aerosols, originating from various
209 sources, each distinctly influence their mixing states and consequently their effects on radiation. (Chung
210 & Seinfeld, 2002; Matsui et al., 2022). Utilizing the ART-2a algorithm, black carbon particles were
211 classified based on the position and intensity of characteristic peaks in their mass spectra. These particles
212 were categorized into seven main types across the entire route, as detailed in Figure 2, which includes
213 the average spectra for each type:

214 (1) **Potassium-Cyanide (K-CN):** Displayed in Figure 2a, these particles show a dominant $^{39}K^+$
215 signal in the cationic mass spectrum, with a relative peak area exceeding 0.8. Although
216 carbonaceous cations ($^{12}C^+$, $^{24}C_2^+$, $^{36}C_3^+$) are less pronounced, significant cyanide peaks ($^{42}CN^-$
217 and $^{46}CNO^-$) along with inorganic and organic carbon peaks are evident in the anionic mass
218 spectrum. Accounting for 51.8% of all black carbon particles, these strong signals suggest a
219 biomass combustion origin (Silva et al., 1999).

220 (2) **Calcium-Nitrate (Ca-NO₃):** Figure 2b shows that these particles primarily feature a strong
221 $^{40}Ca^+$ signal in the cationic mass spectrum, with a significant nitrate signal ($^{46}NO_2^-$) and



222 additional carbon peaks in the anionic mass spectrum. Comprising 23.7% of the particles, this
223 type is associated with continental dust transport, possibly from mid to low-latitude dust
224 reaching the Arctic during summer (Willis et al., 2018).

225 (3) **Potassium-Nickel-Sulfate (K-Ni-SO₄)**: As seen in Figure 2c, these particles exhibit
226 prominent ³⁹K⁺ and carbonaceous cation signals in the cation mass spectrum, with notable
227 ⁵⁹Ni⁺ and intense sulfate peaks (⁹⁷HSO₄⁻) in the anionic mass spectrum. Making up 15.9% of
228 the particles, their nickel content typically points to petroleum product combustion and other
229 anthropogenic activities (Shevchenko et al., 2003; Zhan et al., 2014).

230 (4) **Sulfate (SO₄)**: Figure 2d illustrates these particles with strong sulfate signal peaks (⁹⁷HSO₄⁻)
231 in the anionic mass spectrum and a significant presence of inorganic carbon cations,
232 constituting 7.6% of the total. Correlations (Figure S2a, r=0.41, p<0.01) observed between the
233 SO₄ and K-Ni-SO₄ black carbon particles indicate a common source of anthropogenic origin.

234 (5) **Potassium-Sulfate (K-SO₄)**: Shown in Figure 2e, these particles, which account for only
235 0.6%, are characterized by strong ³⁹K⁺ signals in the cationic mass spectrum and sulfate peaks
236 (⁹⁷HSO₄⁻) alongside inorganic carbon peaks in the anionic mass spectrum. Compared to SO₄,
237 the K-SO₄ and K-Ni-SO₄ black carbon particles exhibit a stronger correlation (Figure S2b,
238 r=0.82, p<0.01), suggesting that they also originate from similar human activities.

239 (6) **Potassium-L-Glucan (K-lev)**: Figure 2f depicts these particles with ³⁹K⁺ and organic matter
240 signals in the cationic mass spectrum and characteristic L-glucan peaks (m/z = -45, -59, -71, -
241 73) in the anionic mass spectrum, indicative of biomass combustion and comprising 0.5% of
242 the particles (Simoneit et al., 1999).

243 Previous research has differentiated the sources of Arctic black carbon, identifying fossil fuels as
244 the primary winter source and biomass burning as the predominant summer source, with an annual
245 average biomass combustion contribution of 39±10% to atmospheric black carbon (Winiger et al., 2019).
246 Further studies have shown that biomass combustion contributes over 90% to Arctic ice and snow black
247 carbon (Hegg et al., 2009). This study's cluster analysis quantitatively estimated that biomass combustion
248 is the main summer source, accounting for 52.3% of black carbon aerosols, with specific mixed types
249 including K-CN and K-lev. Anthropogenic and transported sources from mid and low latitudes account
250 for 24.0% and 23.7%, respectively, with specific mixed types including K-Ni-SO₄, SO₄, K-SO₄, and Ca-



251 NO₃.

252 **3.2 Spatial distribution of black carbon particles**

253 As illustrated in Figure S3, the hourly concentration of black carbon particles along the surveyed
254 route varied substantially, ranging from 0 to 3000 particles per hour. In the initial three segments, the
255 concentration remained relatively low, with the highest value not surpassing 1500 particles per hour.
256 Conversely, in Leg IV, the concentration of black carbon particles was markedly higher, reaching a peak
257 near 3000/h, and Leg V also recorded a significant concentration of 2500/h. The proportion of black
258 carbon particles varied from 0% to 60% across the route. Unlike the number concentration, the
259 distribution of the proportion of black carbon particles was more uniform throughout the voyage. Figure
260 3 displays the hourly number and proportion of black carbon particles by box plots. The mean
261 concentrations of black carbon particles in Legs I-V were 198/h, 148/h, 255/h, 757/h, and 345/h,
262 respectively. The corresponding average proportions were 19.8 (± 10.7) %, 24.4 (± 16.2) %, 26.6
263 (± 13.6) %, 26.1 (± 10.9) %, and 26.5 (± 10.5) %. Although the average concentration of black carbon
264 particles varied significantly among the different segments, the proportion of black carbon particles
265 relative to all particles remained relatively consistent, particularly in the last three segments where the
266 average proportion showed minimal variation. This consistency suggests that black carbon particles are
267 prevalent in the Arctic summer ocean boundary layer and constitute a significant and uniformly
268 distributed component of the atmospheric aerosol. The overall average proportion of black carbon
269 particles for all segments was 24.7 (± 13.6) %.

270 While the average proportion of black carbon particles remained relatively consistent across
271 different segments, spatial variations in the proportions of black carbon particle types were observed
272 (Figure 4), suggesting diverse sources of black carbon aerosols in different regions. In Leg I, the
273 predominant black carbon types were Ca-NO₃, which accounted for about 50% of the particles, followed
274 by K-CN at approximately 25%, and K-Ni-SO₄ and SO₄ each at about 10%. K-SO₄ and K-lev were nearly
275 absent. In Leg II, K-CN was the most significant, comprising about 70% of particles, with Ca-NO₃ at
276 around 15%, K-Ni-SO₄ and K-lev each at about 5%, SO₄ less than 5%, and K-SO₄ almost nonexistent.
277 Leg III saw a dominance of K-CN and Ca-NO₃, each making up about 50%, with other types collectively
278 less than 5%. Leg IV featured K-CN as the most prevalent, at about 60%, with K-Ni-SO₄ and Ca-NO₃
279 each contributing around 20%, SO₄ at about 5%, and negligible amounts of K-SO₄ and K-lev. Leg V had



280 a similar composition to Leg IV, but with a reduced proportion of Ca-NO₃ at about 10% and an increased
281 proportion of SO₄ to about 15%. The distribution of mixed types of black carbon particles helped to
282 elucidate the source contributions in different regions. In the central Arctic Ocean, the Norwegian Sea-
283 Iceland, and the North Atlantic Ocean, biomass combustion was the primary source of black carbon,
284 accounting for more than 50% of particles, particularly in the central Arctic Ocean where it exceeded
285 70%. In the Chukchi Sea region, terrestrial transport from mid and low latitudes was the dominant source,
286 also accounting for more than 50% of particles, with biomass combustion and anthropogenic pollution
287 sources each contributing about 25%. Near Svalbard, biomass combustion sources and terrestrial
288 transport were the main contributors and were roughly equal in proportion. Previous studies have
289 indicated that black carbon aerosols from different sources possess varying thicknesses of coatings,
290 thereby affecting their radiative impacts differently (Matsui et al., 2022). In our study, the sources of
291 black carbon aerosols in different regions were quantitatively estimated, which is crucial for a more
292 accurate assessment of the climate effects of Arctic black carbon aerosols.

293 **3.3 Photochemical processes of Arctic black carbon aerosols**

294 In urban environments, it has been established that black carbon particles are photoactive, capable
295 of releasing reactive oxygen species such as excited oxygen molecules (¹O₂) and hydroxyl radicals. These
296 species participate in atmospheric chemical processes, including the catalytic formation of sulfate
297 (Gehling & Dellinger, 2013; Li et al., 2018; Li et al., 2019; Zhang et al., 2020; Zhang et al., 2021). In the
298 Arctic Ocean boundary layer, the release of dimethyl sulfur by marine organisms and its subsequent
299 oxidation to sulfate is known to have significant environmental and climatic impacts (Bates et al., 1987;
300 Charlson et al., 1987; Rap et al., 2013). However, the potential role of black carbon particles in enhancing
301 this sulfate formation process remains unclear, a topic we explore in this section. Although black carbon
302 particles facilitate the catalysis of sulfate formation, they do not appear to significantly promote nitrate
303 formation (Zhang et al., 2021). Consequently, this leads to a relative enrichment of sulfate compared to
304 nitrate in particles containing black carbon, increasing the sulfate to nitrate mass concentration ratio
305 (SNR). In this study, we utilized the ratio of the relative peak areas of sulfate (⁹⁷HSO₄⁻) to nitrate (⁴⁶NO₂⁻)
306 to describe SNR, specifically denote the relative enrichment of sulfate in black carbon particles. The
307 SNR for black carbon particles varied widely, ranging from 0 to 1000, with an average value of 32.25
308 and a median value of 8.14. This contrasts with measurements from a previous single-particle study of



309 urban black carbon during summer, which reported SNR values ranging from 0 to 5 and a median value
310 of about 2 (Zhang et al., 2021). The notably higher SNR in the Arctic summer suggests that emissions of
311 sulfur-containing gases by marine organisms in the ocean boundary layer may provide a substantial
312 number of precursors for sulfate generation. Additionally, we calculated the SNR for the most common
313 sea salt particles in the ocean boundary layer, which ranged from 0 to 6, with an average of 0.34 and a
314 median of 0.15. This comparison shows that the SNR of black carbon particles is significantly higher
315 than that of sea salt particles, indicating substantial sulfate enrichment in black carbon particles.
316 Moreover, we used the cumulative relative peak area values (R_{BC}) of inorganic carbonaceous anions ($^{12}C^+$,
317 $^{24}C^{\pm}$, $^{36}C^{\pm}$, $^{48}C^-$) in black carbon-containing particles to indicate the content of black carbon components
318 within the particles. As illustrated in Figure 5, the SNR value increased with the rising content of black
319 carbon, displaying a moderate correlation ($r = 0.37$, $p < 0.01$). This trend suggests that higher black
320 carbon content in particles tends to enhance sulfate enrichment over nitrate, indicating that black carbon
321 components can promote sulfate formation in the Arctic region.

322 The observed phenomenon of elevated SNR in black carbon particles can be attributed to several
323 factors: (1) Common Source for Black Carbon and Sulfate Precursors: Black carbon and sulfate, along
324 with their precursors, possibly share common sources, contributing to the increased SNR in black carbon
325 particles. Predominantly, black carbon aerosols originate from biomass combustion, which accounts for
326 over half of these emissions, supplemented by anthropogenic sources and transport from mid to low
327 latitudes in this study. In the ocean boundary layer, sulfur-containing gases emitted by marine biogenic
328 sources are primary precursors for sulfate (Becagli et al., 2016; Gondwe et al., 2003; Jarnikova et al.,
329 2018). However, anthropogenic black carbon aerosols, such as those from ship-based emissions, may
330 also emit significant amounts of SO_2 (Davis et al., 2001; Krause et al., 2021; Winther et al., 2014). After
331 excluding these anthropogenic sources, the average SNR for black carbon particles from biomass
332 combustion sources was calculated at 7.85, significantly higher than in sea salt particles. Although higher
333 SO_2 concentrations are noted in biomass combustion air masses, an increase in nitrogen oxides (Leino et
334 al., 2014) suggests a limited impact on the SNR; (2) Catalysis by Transition Metal Elements: Transition
335 metals such as iron and vanadium found in black carbon particles can catalyze the formation of sulfate
336 (Ault et al., 2010; Zhang et al., 2019). These metals typically originate from anthropogenic activities,
337 including ship-based emissions, and are consequently enriched in anthropogenic black carbon aerosols.



338 As the concentration of black carbon in the particles increases, so too does the concentration of these
339 catalytic metals, enhancing their effect on sulfate formation. Despite this, even after accounting for
340 sources that might skew data, such as those rich in transition metals, black carbon particles still exhibit
341 a higher enrichment of sulfate compared to sea salt particles. After ruling out these factors, it is evident
342 that black carbon itself contributes to the enhancement of sulfate formation, aligning with observations
343 in urban environments (Zhang et al., 2021). To further explore the potential involvement of
344 photochemical processes in this enhancement, black carbon particles were categorized based on varying
345 levels of sunlight intensity. The SWGDN served as an indicator of sunlight intensity changes. Black
346 carbon particles were segmented into five groups based on their SWGDN values, spanning from 0-100,
347 100-200, 200-300, 300-400, and over 400 W/m^2 . As depicted in Figure 6, the SNR values are lower
348 when the SWGDN value is below 200 W/m^2 . However, as the SWGDN value increases, there is a
349 corresponding rise in SNR values. The average SNR values for black carbon particles across these groups
350 were 25.2, 21.4, 44.3, 63.1, and 174.4, respectively. These findings underscore a significant impact of
351 sunlight intensity on the concentration of sulfate in black carbon particles, suggesting that light-driven
352 photochemical reactions may play a crucial role in this process.

353 The formation of sulfate and nitrate is intrinsically linked to the presence of oxidizing agents such
354 as hydroxyl radicals and active halogens, which are often involved in photochemical processes (Chen et
355 al., 2017; Liu et al., 2021; Shao et al., 2019). Additionally, conditions of strong light are thought to
356 facilitate the degradation of nitrate, potentially leading to an increased SNR in black carbon particles
357 (Zhang et al., 2021). These factors may significantly contribute to the observed light intensity-dependent
358 enrichment of sulfate in black carbon particles. To evaluate this possibility, the SNR values of sea salt
359 particles under varying lighting conditions, as shown in Figure 6, do not mirror the changes observed in
360 black carbon particles. The average SNR values for sea salt particles across the five groups—ranging
361 from low to high sunlight intensities—were 0.33, 0.28, 0.53, 0.52, and 0.33, respectively. Notably, while
362 there is a slight increase in SNR values for sea salt particles within the 200-400 W/m^2 range of SWGDN,
363 the values decrease again when SWGDN exceeds 400 W/m^2 , aligning with the lower values observed
364 under less intense light conditions. This disparity suggests that while sunlight intensity influences the
365 photoactivity of black carbon, leading to the release of reactive oxygen species and subsequent sulfate
366 formation, it does not significantly impact the SNR in sea salt particles. This indicates a unique



367 photochemical sensitivity in black carbon particles that enhances sulfate formation under higher sunlight
368 conditions—a process consistent with photochemistry observed in urban environments, rather than being
369 caused by degradation of nitrate (Zhang et al., 2020; Zhang et al., 2021). Furthermore, the interaction of
370 sulfate and organic carbon with black carbon aerosols can significantly alter the optical properties of
371 these particles, thereby increasing their positive radiative forcing (Chung & Seinfeld, 2002). This
372 photochemical facilitation not only affects the chemical composition of the aerosols but also their
373 radiative properties, which could have profound implications for the Arctic climate. Such interactions
374 and their climatic effects underscore the importance of incorporating this photochemistry into future
375 climate models to accurately predict and mitigate the impacts of black carbon in polar regions.

376

377 **4 Summary and Implications**

378 In this research, we focused on extracting and analyzing black carbon-containing particles from
379 collected atmospheric samples to understand their mixing states, sources, spatial distribution variations,
380 and their role in enhancing sulfate formation through photochemical processes. The black carbon
381 particles were categorized into six distinct groups based on their mixing states: K-CN, Ca-NO₃, K-Ni-
382 SO₄, SO₄, K-SO₄, and K-lev. Our findings indicate that K-CN and K-lev, which are primarily products
383 of biomass combustion, represent 52.3% of the sampled particles. Ca-NO₃, linked to terrestrial sources,
384 accounted for 23.7% of the particles. The remaining categories—K-Ni-SO₄, SO₄, and K-SO₄—are
385 predominantly derived from anthropogenic activities, such as emissions from ships, contributing to 24.0%
386 of the particles. Spatial analysis revealed that the distribution of black carbon-containing particles is
387 relatively stable across the Arctic regions, suggesting their pervasive presence in the Arctic summer
388 ocean boundary layer as a significant component of atmospheric aerosols. However, the source
389 contributions of these particles vary spatially. In the central Arctic Ocean and areas near the Norwegian
390 Sea-Iceland and the North Atlantic Ocean, more than 50% of the black carbon particles originate from
391 biomass combustion, with the figure rising above 70% in the central Arctic Ocean. In contrast, in the
392 Chukchi Sea region, terrestrial transport from middle and low latitudes is the dominant source,
393 accounting for more than 50% of the black carbon particles, while biomass combustion and
394 anthropogenic pollution each contribute approximately 25%. Near Svalbard, biomass combustion
395 sources and terrestrial transport are equally significant contributors to the presence of black carbon



396 particles. Moreover, our study confirms that black carbon particles in the Arctic Ocean boundary layer
397 significantly enhance sulfate formation through their involvement in photochemical reactions. This
398 interaction not only alters the mixing state of the black carbon aerosols but also affects their radiative
399 properties, potentially influencing the climate. This underscores the importance of understanding the
400 complex photochemistry of black carbon in Arctic aerosols for future climate modeling and assessment
401 strategies.

402

403 **Acknowledgments**

404 This work was supported by the National Natural Science Foundation of China (NSFC) (41941014).

405 We thank China Arctic and Antarctic Administration for fieldwork support.

406

407 **Data availability**

408 Access to the raw data and products is available at DOI:10.5281/zenodo.13883324 or by contacting

409 either the corresponding author, Zhouqing Xie, at zqxie@ustc.edu.cn, or the first author, Longquan Wang,

410 at wq1995@mail.ustc.edu.cn. The SWGDN data featured in this publication can be accessed publicly

411 through the NASA Goddard Earth Sciences Data and Information Services Center at

412 <https://disc.gsfc.nasa.gov/>.

413

414 **Declaration of competing interests**

415 The authors declare that they have no known competing financial interests or personal relationships

416 that could have appeared to influence the work reported in this paper.



417 **Author contribution:**

418 Conceptualization: Zhouqing Xie

419 Data curation: Longquan Wang, Jinpei Yan

420 Investigation: Longquan Wang, Jinpei Yan

421 Formal analysis: Longquan Wang, Afeng Chen, Bei Jiang, Fange Yue, Xiawei Yu, Zhouqing Xie

422 Visualization: Longquan Wang

423 Funding acquisition: Zhouqing Xie

424 Supervision: Zhouqing Xie

425 Writing – original draft: Longquan Wang

426 Writing – review & editing: Longquan Wang, Zhouqing Xie



427 **Reference**

- 428 Ault, A. P., Gaston, C. J., Wang, Y., Dominguez, G., Thiemens, M. H., & Prather, K. A. (2010) .
429 Characterization of the Single Particle Mixing State of Individual Ship Plume Events Measured
430 at the Port of Los Angeles. *Environmental Science & Technology*, *44* (6) , 1954-1961.
431 doi:10.1021/es902985h
- 432 Bates, T. S., Charlson, R. J., & Gammon, R. H. (1987) . EVIDENCE FOR THE CLIMATIC ROLE
433 OF MARINE BIOGENIC SULFUR. *Nature*, *329* (6137) , 319-321. doi:10.1038/329319a0
- 434 Becagli, S., Lazzara, L., Marchese, C., Dayan, U., Ascanius, S. E., Cacciani, M., . . . Udisti, R. (2016) .
435 Relationships linking primary production, sea ice melting, and biogenic aerosol in the Arctic.
436 *Atmospheric Environment*, *136*, 1-15. doi:10.1016/j.atmosenv.2016.04.002
- 437 Beddows, D. C., Donovan, R. J., Harrison, R. M., Heal, M. R., Kinnersley, R. P., King, M. D., . . .
438 Thompson, K. C. (2004) . Correlations in the chemical composition of rural background
439 atmospheric aerosol in the UK determined in real time using time-of-flight mass spectrometry.
440 *J Environ Monit*, *6* (2) , 124-133. doi:10.1039/b311209h
- 441 Bond, T. C., & Bergstrom, R. W. (2006) . Light absorption by carbonaceous particles: An investigative
442 review. *Aerosol Science and Technology*, *40* (1) , 27-67. doi:10.1080/02786820500421521
- 443 Charlson, R. J., Lovelock, J. E., Andreae, M. O., & Warren, S. G. (1987) . OCEANIC
444 PHYTOPLANKTON, ATMOSPHERIC SULFUR, CLOUD ALBEDO AND CLIMATE.
445 *Nature*, *326* (6114) , 655-661. doi:10.1038/326655a0
- 446 Chen, Q., Schmidt, J. A., Shah, V., Jaeglé, L., Sherwen, T., & Alexander, B. (2017) . Sulfate production
447 by reactive bromine: Implications for the global sulfur and reactive bromine budgets.
448 *Geophysical Research Letters*, *44* (13) , 7069-7078. doi:10.1002/2017gl073812
- 449 Ching, J., Riemer, N., & West, M. (2012) . Impacts of black carbon mixing state on black carbon
450 nucleation scavenging: Insights from a particle-resolved model. *Journal of Geophysical*
451 *Research-Atmospheres*, *117*. doi:10.1029/2012jd018269
- 452 Ching, J., West, M., & Riemer, N. (2018) . Quantifying Impacts of Aerosol Mixing State on Nucleation-
453 Scavenging of Black Carbon Aerosol Particles. *Atmosphere*, *9* (1) . doi:10.3390/atmos9010017
- 454 Chung, S. H., & Seinfeld, J. H. (2002) . Global distribution and climate forcing of carbonaceous aerosols.
455 *Journal of Geophysical Research-Atmospheres*, *107* (D19) . doi:10.1029/2001jd001397



- 456 Clarke, A. D., & Noone, K. J. (1985) . SOOT IN THE ARCTIC SNOWPACK - A CAUSE FOR
457 PERTURBATIONS IN RADIATIVE-TRANSFER. *Atmospheric Environment*, 19 (12) , 2045-
458 2053. doi:10.1016/0004-6981 (85) 90113-1
- 459 Davis, D. D., Grodzinsky, G., Kasibhatla, P., Crawford, J., Chen, G., Liu, S., . . . Sandholm, S. (2001) .
460 Impact of ship emissions on marine boundary layer NO_x and SO₂ distributions over the Pacific
461 basin. *Geophysical Research Letters*, 28 (2) , 235-238. doi:10.1029/2000gl012013
- 462 Dou, T.-F., & Xiao, C.-D. (2016) . An overview of black carbon deposition and its radiative forcing
463 over the Arctic. *Advances in Climate Change Research*, 7 (3) , 115-122.
464 doi:10.1016/j.accre.2016.10.003
- 465 Eleftheriadis, K., Vratolis, S., & Nyeki, S. (2009) . Aerosol black carbon in the European Arctic:
466 Measurements at Zeppelin station, Ny-Alesund, Svalbard from 1998-2007. *Geophysical
467 Research Letters*, 36. doi:10.1029/2008gl035741
- 468 Evangeliou, N., Balkanski, Y., Hao, W. M., Petkov, A., Silverstein, R. P., Corley, R., . . . Skov, H. (2016) .
469 Wildfires in northern Eurasia affect the budget of black carbon in the Arctic - a 12-year
470 retrospective synopsis (2002-2013) . *Atmospheric Chemistry and Physics*, 16 (12) , 7587-
471 7604. doi:10.5194/acp-16-7587-2016
- 472 Flanner, M. G., Zender, C. S., Randerson, J. T., & Rasch, P. J. (2007) . Present-day climate forcing and
473 response from black carbon in snow. *Journal of Geophysical Research-Atmospheres*, 112(D11) .
474 doi:10.1029/2006jd008003
- 475 Gehling, W., & Dellinger, B. (2013) . Environmentally Persistent Free Radicals and Their Lifetimes
476 in PM_{2.5}. *Environmental Science & Technology*, 47 (15) , 8172-8178. doi:10.1021/es401767m
- 477 Gondwe, M., Krol, M., Gieskes, W., Klaassen, W., & de Baar, H. (2003) . The contribution of ocean-
478 leaving DMS to the global atmospheric burdens of DMS, MSA, SO₂, and NSS SO₄= . *Global
479 Biogeochemical Cycles*, 17 (2) . doi:10.1029/2002gb001937
- 480 Hansen, J., Sato, M., Ruedy, R., Nazarenko, L., Lacis, A., Schmidt, G. A., . . . Zhang, S. (2005) .
481 Efficacy of climate forcings. *Journal of Geophysical Research-Atmospheres*, 110 (D18) .
482 doi:10.1029/2005jd005776
- 483 Hegg, D. A., Warren, S. G., Grenfell, T. C., Doherty, S. J., Larson, T. V., & Clarke, A. D. (2009) .
484 Source Attribution of Black Carbon in Arctic Snow. *Environmental Science & Technology*, 43



- 485 (11) , 4016-4021. doi:10.1021/es803623f
- 486 Jacobson, M. Z. (2001) . Strong radiative heating due to the mixing state of black carbon in atmospheric
487 aerosols. *Nature*, *409* (6821) , 695-697. doi:10.1038/35055518
- 488 Jarnikova, T., Dacey, J., Lizotte, M., Levasseur, M., & Tortell, P. (2018) . The distribution of methylated
489 sulfur compounds, DMS and DMSP, in Canadian subarctic and Arctic marine waters during
490 summer 2015. *Biogeosciences*, *15* (8) , 2449-2465. doi:10.5194/bg-15-2449-2018
- 491 Kollner, F., Schneider, J., Willis, M. D., Schulz, H., Kunkel, D., Bozem, H., . . . Borrmann, S. (2021) .
492 Chemical composition and source attribution of sub-micrometre aerosol particles in the
493 summertime Arctic lower troposphere. *Atmospheric Chemistry and Physics*, *21* (8) , 6509-6539.
494 doi:10.5194/acp-21-6509-2021
- 495 Krause, K., Wittrock, F., Richter, A., Schmitt, S., Pohler, D., Weigelt, A., & Burrows, J. P. (2021) .
496 Estimation of ship emission rates at a major shipping lane by long-path DOAS measurements.
497 *Atmospheric Measurement Techniques*, *14* (8) , 5791-5807. doi:10.5194/amt-14-5791-2021
- 498 Leino, K., Riuttanen, L., Nieminen, T., Dal Maso, M., Vaananen, R., Pohja, T., . . . Kulmala, M. (2014) .
499 Biomass-burning smoke episodes in Finland from eastern European wildfires. *Boreal
500 Environment Research*, *19*, 275-292.
- 501 Li, L., Huang, Z. X., Dong, J. G., Li, M., Gao, W., Nian, H. Q., . . . Zhou, Z. (2011) . Real time bipolar
502 time-of-flight mass spectrometer for analyzing single aerosol particles. *International Journal of
503 Mass Spectrometry*, *303* (2-3) , 118-124. doi:10.1016/j.ijms.2011.01.017
- 504 Li, M., Bao, F. X., Zhang, Y., Sheng, H., Chen, C. C., & Zhao, J. C. (2019) . Photochemical Aging of
505 Soot in the Aqueous Phase: Release of Dissolved Black Carbon and the Formation of O-1 (2) .
506 *Environmental Science & Technology*, *53* (21) , 12311-12319. doi:10.1021/acs.est.9b02773
- 507 Li, M., Bao, F. X., Zhang, Y., Song, W. J., Chen, C. C., & Zhao, J. C. (2018) . Role of elemental carbon
508 in the photochemical aging of soot. *Proceedings of the National Academy of Sciences of the
509 United States of America*, *115* (30) , 7717-7722. doi:10.1073/pnas.1804481115
- 510 Liu, T. Y., Chan, A. W. H., & Abbatt, J. P. D. (2021) . Multiphase Oxidation of Sulfur Dioxide in
511 Aerosol Particles: Implications for Sulfate Formation in Polluted Environments. *Environmental
512 Science & Technology*, *55* (8) , 4227-4242. doi:10.1021/acs.est.0c06496
- 513 Matsui, H., Mori, T., Ohata, S., Moteki, N., Oshima, N., Goto-Azuma, K., . . . Kondo, Y. (2022) .



- 514 Contrasting source contributions of Arctic black carbon to atmospheric concentrations,
515 deposition flux, and atmospheric and snow radiative effects. *Atmospheric Chemistry and*
516 *Physics*, 22 (13) , 8989-9009. doi:10.5194/acp-22-8989-2022
- 517 Qi, L., Li, Q. B., Li, Y. R., & He, C. L. (2017) . Factors controlling black carbon distribution in the
518 Arctic. *Atmospheric Chemistry and Physics*, 17 (2) , 1037-1059. doi:10.5194/acp-17-1037-
519 2017
- 520 Qin, X. Y., & Prather, K. A. (2006) . Impact of biomass emissions on particle chemistry during the
521 California Regional Particulate Air Quality Study. *International Journal of Mass Spectrometry*,
522 258 (1-3) , 142-150. doi:10.1016/j.ijms.2006.09.004
- 523 Quinn, P. K., Shaw, G., Andrews, E., Dutton, E. G., Ruoho-Airola, T., & Gong, S. L. (2007) . Arctic
524 haze: current trends and knowledge gaps. *Tellus Series B-Chemical and Physical Meteorology*,
525 59 (1) , 99-114. doi:10.1111/j.1600-0889.2006.00236.x
- 526 Ramanathan, V., & Carmichael, G. (2008) . Global and regional climate changes due to black carbon.
527 *Nature Geoscience*, 1 (4) , 221-227. doi:10.1038/ngeo156
- 528 Rap, A., Scott, C. E., Spracklen, D. V., Bellouin, N., Forster, P. M., Carslaw, K. S., . . . Mann, G. (2013) .
529 Natural aerosol direct and indirect radiative effects. *Geophysical Research Letters*, 40 (12) ,
530 3297-3301. doi:10.1002/grl.50441
- 531 Roth, A., Schneider, J., Klimach, T., Mertes, S., van Pinxteren, D., Herrmann, H., & Borrmann, S.
532 (2016) . Aerosol properties, source identification, and cloud processing in orographic clouds
533 measured by single particle mass spectrometry on a central European mountain site during
534 HCCT-2010. *Atmospheric Chemistry and Physics*, 16 (2) , 505-524. doi:10.5194/acp-16-505-
535 2016
- 536 Schmidt, S., Schneider, J., Klimach, T., Mertes, S., Schenk, L. P., Kupiszewski, P., . . . Borrmann, S.
537 (2017) . Online single particle analysis of ice particle residuals from mountain-top mixed-
538 phase clouds using laboratory derived particle type assignment. *Atmospheric Chemistry and*
539 *Physics*, 17 (1) , 575-594. doi:10.5194/acp-17-575-2017
- 540 Serreze, M. C., & Barry, R. G. (2011) . Processes and impacts of Arctic amplification: A research
541 synthesis. *Global and Planetary Change*, 77 (1-2) , 85-96. doi:10.1016/j.gloplacha.2011.03.004
- 542 Shao, J., Chen, Q., Wang, Y., Lu, X., He, P., Sun, Y., . . . Alexander, B. (2019) . Heterogeneous sulfate



- 543 aerosol formation mechanisms during wintertime Chinese haze events: air quality model
544 assessment using observations of sulfate oxygen isotopes in Beijing. *Atmospheric Chemistry
545 and Physics*, 19 (9) , 6107-6123. doi:10.5194/acp-19-6107-2019
- 546 Sharma, S., Andrews, E., Barrie, L. A., Ogren, J. A., & Lavoue, D. (2006) . Variations and sources of
547 the equivalent black carbon in the high Arctic revealed by long-term observations at Alert and
548 Barrow: 1989-2003. *Journal of Geophysical Research-Atmospheres*, 111 (D14) .
549 doi:10.1029/2005jd006581
- 550 Sharma, S., Barrie, L. A., Magnusson, E., Brattstrom, G., Leaitch, W. R., Steffen, A., & Landsberger, S.
551 (2019) . A Factor and Trends Analysis of Multidecadal Lower Tropospheric Observations of
552 Arctic Aerosol Composition, Black Carbon, Ozone, and Mercury at Alert, Canada. *Journal of
553 Geophysical Research-Atmospheres*, 124 (24) , 14133-14161. doi:10.1029/2019jd030844
- 554 Sharma, S., Lavoue, D., Cachier, H., Barrie, L. A., & Gong, S. L. (2004) . Long-term trends of the
555 black carbon concentrations in the Canadian Arctic. *Journal of Geophysical Research-
556 Atmospheres*, 109 (D15) . doi:10.1029/2003jd004331
- 557 Shevchenko, V., Lisitzin, A., Vinogradova, A., & Stein, R. (2003) . Heavy metals in aerosols over the
558 seas of the Russian Arctic. *Science of the Total Environment*, 306 (1-3) , 11-25.
559 doi:10.1016/s0048-9697 (02) 00481-3
- 560 Sierau, B., Chang, R. Y. W., Leck, C., Paatero, J., & Lohmann, U. (2014) . Single-particle
561 characterization of the high-Arctic summertime aerosol. *Atmospheric Chemistry and Physics*,
562 14 (14) , 7409-7430. doi:10.5194/acp-14-7409-2014
- 563 Silva, P. J., Liu, D. Y., Noble, C. A., & Prather, K. A. (1999) . Size and chemical characterization of
564 individual particles resulting from biomass burning of local Southern California species.
565 *Environmental Science & Technology*, 33 (18) , 3068-3076. doi:10.1021/es980544p
- 566 Simoneit, B. R. T., Schauer, J. J., Nolte, C. G., Oros, D. R., Elias, V. O., Fraser, M. P., . . . Cass, G. R.
567 (1999) . Levoglucosan, a tracer for cellulose in biomass burning and atmospheric particles.
568 *Atmospheric Environment*, 33 (2) , 173-182. doi:10.1016/s1352-2310 (98) 00145-9
- 569 Song, X. H., Hopke, P. K., Fergenson, D. P., & Prather, K. A. (1999) . Classification of single particles
570 analyzed by ATOFMS using an artificial neural network, ART-2A. *Analytical Chemistry*, 71 (4) ,
571 860-865. doi:10.1021/ac9809682



- 572 Wang, H. L., Rasch, P. J., Easter, R. C., Singh, B., Zhang, R. D., Ma, P. L., . . . Beagley, N. (2014) .
573 Using an explicit emission tagging method in global modeling of source-receptor relationships
574 for black carbon in the Arctic: Variations, sources, and transport pathways. *Journal of*
575 *Geophysical Research-Atmospheres*, *119* (22) , 12888-12909. doi:10.1002/2014jd022297
- 576 Wang, L., Yan, J., Saiz-Lopez, A., Jiang, B., Yue, F., Yu, X., & Xie, Z. (2022) . Mixing state and
577 distribution of iodine-containing particles in Arctic Ocean during summertime. *Science of the*
578 *Total Environment*, *834*. doi:10.1016/j.scitotenv.2022.155030
- 579 Willis, M. D., Leaitch, W. R., & Abbatt, J. P. D. (2018) . Processes Controlling the Composition and
580 Abundance of Arctic Aerosol. *Reviews of Geophysics*, *56* (4) , 621-671.
581 doi:10.1029/2018rg000602
- 582 Winiger, P., Barrett, T. E., Sheesley, R. J., Huang, L., Sharma, S., Barrie, L. A., . . . Gustafsson, O. (2019).
583 Source apportionment of circum-Arctic atmospheric black carbon from isotopes and modeling.
584 *Science Advances*, *5* (2) . doi:10.1126/sciadv.aau8052
- 585 Winther, M., Christensen, J. H., Plejdrup, M. S., Ravn, E. S., Eriksson, O. F., & Kristensen, H. O. (2014).
586 Emission inventories for ships in the arctic based on satellite sampled AIS data. *Atmospheric*
587 *Environment*, *91*, 1-14. doi:10.1016/j.atmosenv.2014.03.006
- 588 Zhang, F., Wang, Y., Peng, J. F., Chen, L., Sun, Y. L., Duan, L., . . . Zhang, R. Y. (2020) . An unexpected
589 catalyst dominates formation and radiative forcing of regional haze. *Proceedings of the National*
590 *Academy of Sciences of the United States of America*, *117* (8) , 3960-3966.
591 doi:10.1073/pnas.1919343117
- 592 Zhang, G. H., Fu, Y. Z., Peng, X. C., Sun, W., Shi, Z. B., Song, W., . . . Bi, X. H. (2021) . Black Carbon
593 Involved Photochemistry Enhances the Formation of Sulfate in the Ambient Atmosphere:
594 Evidence From In Situ Individual Particle Investigation. *Journal of Geophysical Research-*
595 *Atmospheres*, *126* (19) . doi:10.1029/2021jd035226
- 596 Zhan, J. Q., Gao, Y., Li, W., Chen, L. Q., Lin, H. M., & Lin, Q. (2014) . Effects of ship emissions on
597 summertime aerosols at Ny-Alesund in the Arctic. *Atmospheric Pollution Research*, *5* (3) ,
598 500-510. doi:10.5094/apr.2014.059



599 **Figure Captions**

600 **Figure 1.** Sampling locations from open water to the high Arctic are indicated by a blue line. The
601 expedition is divided into five segments, each outlined by dashed red boxes: Chukchi Sea section,
602 denoted as Leg I, spans from 22:00 on July 30th, 2017, to 19:00 on August 1st, 2017, covering latitudes
603 from 66.0°N to 74.8°N and longitudes from 169.4°W to 159.2°W. The high Arctic section, marked as
604 Leg II, extends from 20:00 on August 10th to 5:00 on August 12th, 2017, ranging from 83.7°N to 84.6°N
605 and from 132.0°E to 110.4°E. Svalbard Islands section, identified as Leg III, is from 2:00 on August 17th
606 to 11:00 on August 19th, 2017, with latitudes from 82.6°N to 74.3°N and longitudes from 25.4°E to 2.1°E.
607 Norwegian Sea and Iceland section, referred to as Leg IV, occurs from 14:00 on August 23rd to 14:00 on
608 August 25th, 2017, spanning from 67.0°N to 61.2°N and from 2.1°W to 25.8°W. Atlantic Ocean section,
609 labeled as Leg V, spanning from 14:00 on August 25th to 17:00 on August 27th, 2017, covers from
610 61.1°N to 56.2°N and from 25.9°W to 46.9°W.

611 **Figure 2.** Spectra of various types of black carbon particles with peaks displaying strong signals
612 highlighted in blue. The X-axis represents the mass-to-charge ratio (m/z), and the Y-axis shows the
613 relative peak area. The cation spectrum is presented in the upper half of each map, while the anion
614 spectrum is depicted in the lower half. (a) Spectra of K-CN black carbon particles; (b) Spectra of Ca-
615 NO₃ black carbon particles; (c) Spectra of K-Ni-SO₄ black carbon particles; (d) Spectra of SO₄ black
616 carbon particles; (e) Spectra of K-SO₄ black carbon particles; (f) Spectra of K-lev black carbon particles.

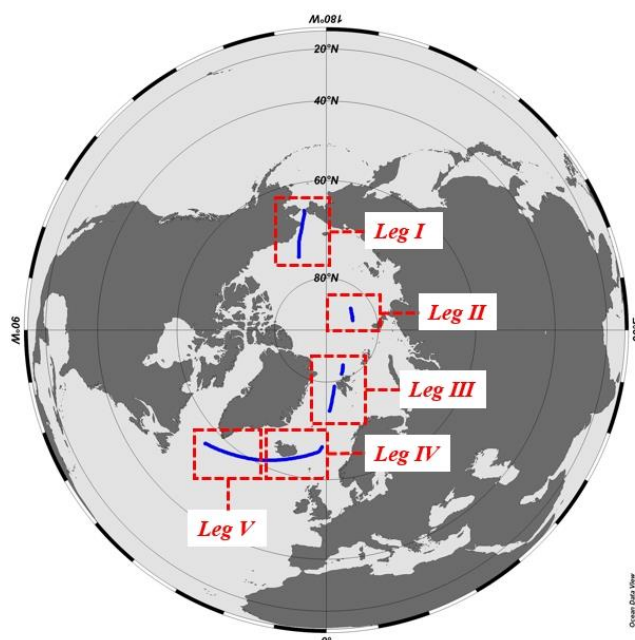
617 **Figure 3.** Box plots of hourly counts (in red) and fractions (in blue) of black carbon particles from Leg
618 I to Leg V. The tops and bottoms of the boxes indicate the upper and lower quartile values, respectively.
619 Horizontal lines within the boxes denote median values. Hollow squares and dotted black lines represent
620 mean values. Vertical whiskers extend to scattered values above and below the boxes, with caps on the
621 whiskers indicating maximum and minimum values. Cross symbols mark the 99th and 1st percentile
622 values, respectively.

623 **Figure 4.** Fraction of each type of black carbon particles from Leg I to Leg V. Different colors represent
624 various types of black carbon particles: blue for K-CN, green for Ca-NO₃, orange for K-Ni-SO₄, magenta
625 for SO₄, cyan for K-SO₄, and red for K-lev.

626 **Figure 5.** Correlation of SNR and R_{BC} in black carbon particles. A red dashed line indicates the linear fit
627 curve, with the correlation coefficient (r) and significance level (p) provided.

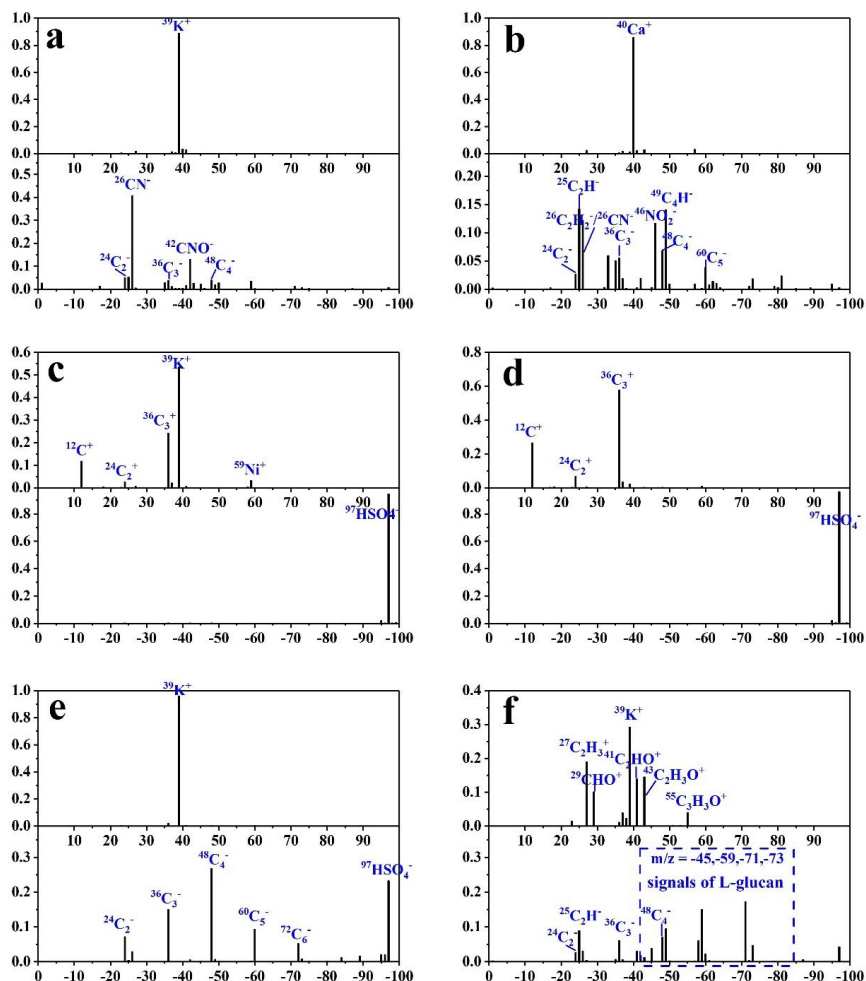


628 **Figure 6.** Box plots of SNR for sea salt particles (in red) and black carbon particles (in blue), grouped
629 by SWGDN. The upper and lower parts of the boxes indicate the quartile values, while the horizontal
630 lines within the boxes denote the median values. Hollow squares and black dotted lines show the mean
631 values. Vertical whiskers extend to capture scattered values above and below the boxes, with caps on the
632 whiskers indicating the maximum and minimum values. Cross symbols mark the 99th and 1st percentile
633 values, respectively.



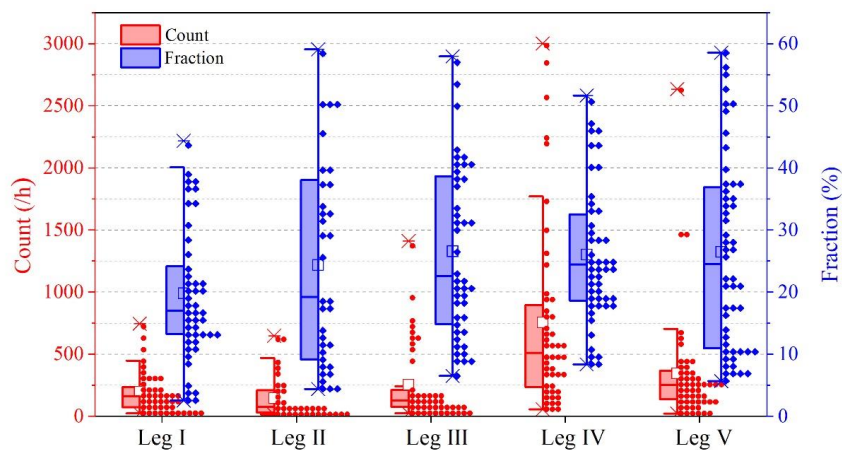
634

635 **Figure 1.** Sampling locations from open water to the high Arctic are indicated by a blue line. The
636 expedition is divided into five segments, each outlined by dashed red boxes: Chukchi Sea section,
637 denoted as Leg I, spans from 22:00 on July 30th, 2017, to 19:00 on August 1st, 2017, covering latitudes
638 from 66.0°N to 74.8°N and longitudes from 169.4°W to 159.2°W. The high Arctic section, marked as
639 Leg II, extends from 20:00 on August 10th to 5:00 on August 12th, 2017, ranging from 83.7°N to 84.6°N
640 and from 132.0°E to 110.4°E. Svalbard Islands section, identified as Leg III, is from 2:00 on August 17th
641 to 11:00 on August 19th, 2017, with latitudes from 82.6°N to 74.3°N and longitudes from 25.4°E to 2.1°E.
642 Norwegian Sea and Iceland section, referred to as Leg IV, occurs from 14:00 on August 23rd to 14:00 on
643 August 25th, 2017, spanning from 67.0°N to 61.2°N and from 2.1°W to 25.8°W. Atlantic Ocean section,
644 labeled as Leg V, spanning from 14:00 on August 25th to 17:00 on August 27th, 2017, covers from
645 61.1°N to 56.2°N and from 25.9°W to 46.9°W.



646

647 **Figure 2.** Spectra of various types of black carbon particles with peaks displaying strong signals
 648 highlighted in blue. The X-axis represents the mass-to-charge ratio (m/z), and the Y-axis shows the
 649 relative peak area. The cation spectrum is presented in the upper half of each map, while the anion
 650 spectrum is depicted in the lower half. (a) Spectra of K-CN black carbon particles; (b) Spectra of Ca-
 651 NO₃ black carbon particles; (c) Spectra of K-Ni-SO₄ black carbon particles; (d) Spectra of SO₄ black
 652 carbon particles; (e) Spectra of K-SO₄ black carbon particles; (f) Spectra of K-lev black carbon particles.



653

654 **Figure 3.** Box plots of hourly counts (in red) and fractions (in blue) of black carbon particles from Leg

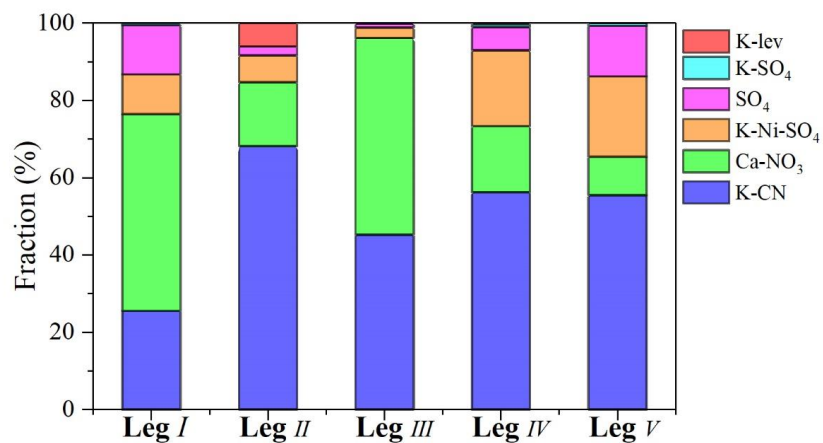
655 I to Leg V. The tops and bottoms of the boxes indicate the upper and lower quartile values, respectively.

656 Horizontal lines within the boxes denote median values. Hollow squares and dotted black lines represent

657 mean values. Vertical whiskers extend to scattered values above and below the boxes, with caps on the

658 whiskers indicating maximum and minimum values. Cross symbols mark the 99th and 1st percentile

659 values, respectively.

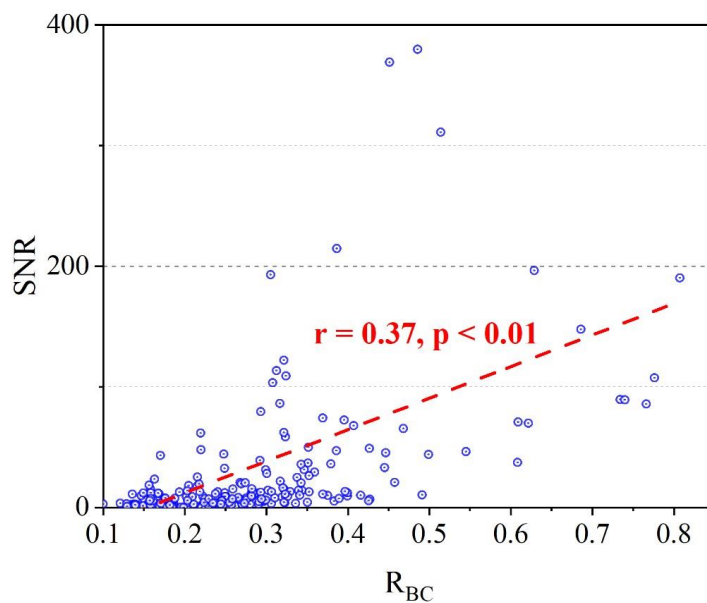


660

661 **Figure 4.** Fraction of each type of black carbon particles from Leg I to Leg V. Different colors represent

662 various types of black carbon particles: blue for K-CN, green for Ca-NO₃, orange for K-Ni-SO₄, magenta

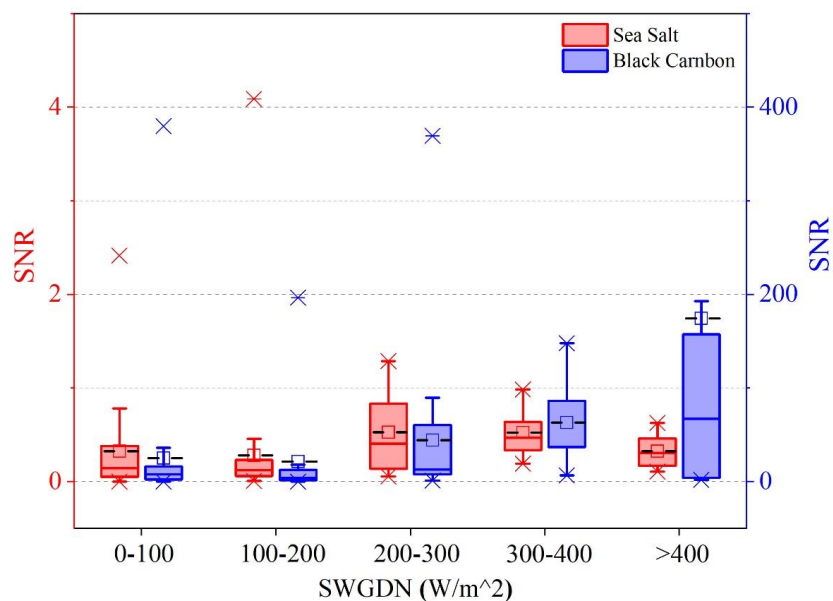
663 for SO₄, cyan for K-SO₄, and red for K-lev.



664

665 **Figure 5.** Correlation of SNR and R_{BC} in black carbon particles. A red dashed line indicates the linear fit

666 curve, with the correlation coefficient (r) and significance level (p) provided.



667

668 **Figure 6.** Box plots of SNR for sea salt particles (in red) and black carbon particles (in blue), grouped
669 by SWGDN. The upper and lower parts of the boxes indicate the quartile values, while the horizontal
670 lines within the boxes denote the median values. Hollow squares and black dotted lines show the mean
671 values. Vertical whiskers extend to capture scattered values above and below the boxes, with caps on the
672 whiskers indicating the maximum and minimum values. Cross symbols mark the 99th and 1st percentile
673 values, respectively.