Evaluating tropospheric nitrogen dioxide in UKCA using OMI satellite retrievals over South and East Asia

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Abstract. We compare tropospheric column nitrogen dioxide (NO_2) in the United Kingdom Chemistry and Aerosol (UKCA) model version 11.0 with satellite measurements from NASA's Earth Observing System (EOS) Aura satellite Ozone Monitoring Instrument (OMI) to investigate the seasonality and trends of tropospheric NO_2 over South and East Asia (S/E Asia). UKCA is the atmospheric composition component of the UK Earth System Model (UKESM). UKCA was run with nudged meteorology, producing hourly output over S/E Asia for 2005–2015. OMI averaging kernels have been applied to the model hourly data sampled at Aura's local overpass time of 13:45±15 to allow consistent model-data comparison. Background UKCA and OMI tropospheric column NO_2 typically ranges between $0\text{--}2 \times 10^{15}$ molecules/cm^2. Diurnal cycles and vertical profiles of the tropospheric NO_2 column in UKCA show that the daily minimum tropospheric column NO_2 occurs around the satellite overpass time. UKCA captures the seasonality but overestimates NO_2 by a factor of ~2.5, especially during winter over Eastern China and North India, at times and locations with high aerosol loadings. Heterogeneous chemistry is represented in the version of UKCA used here as uptake of N_2O_5 on internally generated sulfate aerosol. However, aerosol surface area may be underestimated in polluted locations, contributing to overestimation of NO_2. In addition, the model may underestimate emissions of volatile organic compounds and associated peroxy acetyl nitrate (PAN) formation, leading to insufficient long-range transport of oxidised nitrogen, also contributing to overestimation of NO_2 over polluted regions and underestimation over remote regions. Quantifying and understanding discrepancies in modelled NO_2 warrant further investigation as they propagate into modelling of multiple environmental issues.

1 Introduction

30 Nitrogen oxides (NO_x: the sum of nitrogen dioxide, NO_2, and nitric oxide, NO) are key gases in atmospheric chemistry, and models need to simulate them adequately in order to faithfully represent many important environmental processes. Nitrogen oxides play a central role in the atmospheric nitrogen cycle (Fowler et al., 2013) and are a precursor of nitrate aerosols (Liu et al., 2016), and the greenhouse gas (GHG) tropospheric ozone (O_3) (Bucsela et al., 2008; von Schneidemesser et al., 2015). The oxidising capacity of the atmosphere is affected by NO_x, so it also influences other GHGs such as methane (Naik et al., 2013;

35 Voulgarakis et al., 2013); hence changes in NO_x contribute to climate change (Lelieveld et al., 2015). High concentrations of NO_2 can also increase the risk of acute and chronic respiratory diseases (Brunekreef et al., 2009). Deposition of NO_2 and other species containing reactive nitrogen can lead to the eutrophication of ecosystems and loss of biodiversity (Stevens et al., 2004; Erisman et al., 2013).

Nitrogen oxides are predominantly emitted as NO, mainly originating from fossil fuel combustion (c. 58% of the total), natural

40 emissions (c. 23%), and agriculture/biofuel use (c. 19%) (Lelieveld et al., 2015). In the sunlit troposphere, NO reacts with O_3 to produce NO_2, which photolyzes to return NO and O_3, rapidly forming a photochemical equilibrium. The oxidation products of volatile organic compounds (VOCs) react with NO_2 to form peroxy acetyl nitrates (PANs), key constituents of photochemical smog (Sher, 1998; Beirle et al., 2003; Mallik and Lal, 2014). These compounds are stable at low temperatures typical of the upper troposphere, but thermally unstable in the lower troposphere, decomposing to release NO_2, thus facilitating

45 long-range transport of NO_2 from NO_x source regions to remote sites (Fiore et al., 2018). NO_2 is mainly removed by dry deposition and via oxidation to nitric acid, which readily deposits. Another sink of reactive oxidised nitrogen (NO_y), in darkness, is via heterogeneous uptake of dinitrogen pentoxide (N_2O_5) on the surface of aerosols (Dentener and Crutzen, 1993), which then deposit. These removal processes typically result in a short lifetime for NO_2 of a few hours (e.g., Beirle et al., 2011), and lead to strong spatial and temporal variations in its atmospheric abundance. Atmospheric chemistry models include

50 our best representations of these, and many other, processes that control NO_x, allowing models to simulate spatial and temporal variations in atmospheric composition in detail (e.g., Szopa et al., 2021).

Since the 1990s, various satellite-based instruments have measured tropospheric NO_2 columns. The Global Ozone Monitoring Experiment (GOME) detected NO_2 pollution hotspots around the world (Leue et al., 2001), and in 2002, the scanning imaging absorption spectrometer for atmospheric cartography (SCIAMACHY) began mapping NO_2 pollution at a spatial resolution of

55 30 x 60 km^2, with global coverage every six days, allowing the detection of temporal trends in NO_2 (van der A et al., 2008). The launch of the Ozone Monitoring Instrument (OMI) in 2004 has provided even higher spatial resolution information (13 x 24 km^2) of tropospheric NO_2 with daily global coverage (Levelt et al., 2006; Boersma et al., 2008; Liu et al., 2016). OMI NO_2 data have been validated against in-situ and surface-based observations (e.g., Irie et al., 2009; Lamsal et al., 2014), and provide

a long record of high spatial resolution daily measurements of NO_2, useful for evaluation of global atmospheric chemistry models (e.g., van Noije et al., 2006).

This study focusses on NO_2 pollution over South and East Asia (0-50°N and 55-145°E; Figure 1a) during the period 2005-2015. The region is home to nearly 50% of the Earth's population, and has some of the largest measured NO_2 columns. Rising energy demand, urbanization, traffic and industrialization have led to increases in NO_x emissions in some regions, whilst technological advances, typically introduced in response to environmental legislation, have led to reductions in NO_x emissions in other locations (Mijling et al., 2013). Regional variations in the evolution of NO_x emissions have been captured by satellite NO_2 measurements, with Eastern China showing upward trends of tropospheric NO_2 up to 2011 followed by decreases since 2012 (Shah et al., 2020; Fan et al., 2021; Cooper et al., 2022). By contrast, India shows a continuous increase of 12.5% to 29.6% from 2005 to 2019 (Krotkov et al., 2015, Singh et al., 2023).

We compare model simulations of NO_2 column with equivalent satellite measurements from OMI, to evaluate model performance in terms of simulating the magnitude and spatial distribution of NO_2 over S/E Asia, and its seasonal variations and longer-term temporal trends. We use the UK Chemistry and Aerosol (UKCA) model, the atmospheric composition component of the UK Earth System Model (UKESM). Archibald et al. (2020) compared OMI and UKCA NO_2 columns, identifying some model biases that we explore in more detail here. Because of the high reactivity and short atmospheric lifetime of NO_2, and its anthropogenic sources, its tropospheric concentration has a distinct diurnal signature, as well as a dynamically varying vertical profile. OMI takes column NO_2 measurements at a particular time each day (local time 13:45±15 at the Equator). The vertical profile of NO_2 at the time of measurement has a strong influence on the column amount measured. This is because OMI measures radiation absorption at specific ultra-violet/visible (UV/vis) wavelengths, sampling air (and hence NO_2) along the ray path from the Sun, via the atmosphere, to the satellite; this ray path depends upon the atmospheric albedo (i.e., cloud amount and height) at the time of measurement. An averaging kernel (AK) is required to translate the vertical profile to a column amount; the AK is a weighting profile that depends upon environmental conditions (e.g., cloud properties) at the time of measurement. Model evaluation therefore requires careful temporal sampling and application of the AK, so that the model is sampled in the same way as OMI samples the atmosphere. Model simulations nudged by meteorological reanalysis data are used so that the physical state of the model atmosphere resembles the real atmosphere as closely as possible. This study uses a single model (UKCA), but evaluation of NO_2 from multiple models (e.g., van Noije et al., 2006), such as in a model intercomparison project (MIP) for oxidised nitrogen, is a desirable aim to extend this work to a wider set of models.

The paper is structured as follows: Section 2 describes the version and experimental set-up of the UKCA model used for simulations and the OMI NO_2 datasets used to evaluate the model. In Section 3, we present model results for diurnal and seasonal variations of the vertical distribution of NO_2 and how they influence the NO_2 column amount, analysing the spatial distribution over S/E Asia and temporal trends over the period 2005-2015, comparing model results with satellite data. We discuss some of the reasons for model-observation discrepancies in Section 4, before drawing conclusions about model performance with respect to representation of NO_2 in Section 5.

2 Data and Methods

We focus our analysis on S/E Asia (Figure 1a), dividing this region into six different sub-regions, two politically (India and China) and four geographically (E China, W China, N India, and S India; Figure 1c). The surface NO emissions over South Asia and East Asia in 2005 and 2015 are shown in Figures 1a and 1b, respectively. We focus on the N India and E China sub-regions for detailed study as these are hotspots of high population density and high NO_2 emissions (Ramachandran et al., 2013; Sekiya et al., 2018). Apart from country-wise analysis, we also selected NO_x hotspot and cleaner regions to capture contrasting pollution profiles across S/E Asia. There are large differences in NO emission intensity (i.e. emissions per unit area) across the region: W China has NO emissions intensity approximately 30 times lower than E China, while S India's emissions intensity is relatively low, about half that of N India (Figure S1). This approach allows for a comprehensive analysis across both high-emission and cleaner regions, providing broader insights into NO_x distribution. In 2015, total surface NO emissions were around 0.09-0.10 Tg N yr^{-1} from the box over E China and 0.07-08 Tg N yr^{-1} over the N India box. Figure 1c shows percentage changes relative to 2005 in the NO surface emissions from 2005 to 2015 (from AerChemMIP, Collins et al., 2017), which show a 40-60% increase in surface NO emissions over this period for India, whereas the increase is relatively smaller (20-40%) for China. The 2005-2015 trends in surface NO emissions integrated over the whole of China and India are shown in Figure 1d. Surface NO emissions from China were 4.8 Tg N yr^{-1} in 2005, increasing to 6.5 Tg N yr^{-1} by 2011, followed by a decrease to 5.8 Tg N yr^{-1} in 2015, as also reported in other studies (Miyazaki et al., 2017, Shah et al., 2020). In contrast, India has shown a consistent upward trend, with NO emissions increasing from 1.8 Tg N yr^{-1} in 2005 to 2.5 Tg N yr^{-1} in 2015 (Krotkov et al., 2015).

2.1 UKCA Model

We use the United Kingdom Chemistry and Aerosols (UKCA) model version 11.0 (Archibald et al., 2020). UKCA is an aerosol-chemistry model coupled to the UK Met Office Hadley Centre HadGEM family of climate models. UKCA simulates the atmospheric composition and climate from the surface to the mesosphere (Morgenstern et al., 2009). HadGEM acts as the dynamical core and provides components for large-scale advection, convective transport and boundary layer mixing of chemical and aerosol tracers (O'Connor et al., 2014). UKCA version 11.0 comprises the GA7.1 climate model (Walters et al., 2019) with the StratTrop (CheST) chemistry scheme and the GLOMAP-mode aerosol scheme. The UKCA stratospheric and tropospheric chemistry schemes include all the well-known photochemical and night-time reactions related to NO_x, and are described in detail and evaluated by Morgenstern et al., (2009), O'Connor et al., (2014) and Archibald et al., (2020). Aerosol surface area from GLOMAP is used to drive heterogeneous chemistry.

The model's horizontal resolution (N96: 1.875° longitude × 1.25° latitude) is much coarser than the satellite data products used. The model is divided into 85 hybrid height levels with the model top at ~85 km. The vertical resolution is finest close to the surface and gradually decreases with height, i.e., layers are concentrated towards the surface, so the boundary layer (BL) is relatively well resolved in the model, with the lowest (surface) level ~18 m thick.

We used a variant of the version 11.0 'release job' (job ID u-bb210; https://www.ukca.ac.uk/wiki/index.php/Release_Job_UM11.0, dated 7th October 2019), adding a meteorological nudging scheme to allow for a more meaningful comparison of satellite data to model output. Nudging (Newtonian relaxation) is a data assimilation technique that adjusts dynamical variables of a free-running general circulation model (GCM) using meteorological reanalysis data to allow a relatively realistic representation of the atmosphere at a given time. For nudging, the European Centre for Medium-range Weather Forecasts (ECMWF) ERA-interim data were used, and the model was run from 2005 to 2015. The ERA-Interim data is at T255 (78 km) resolution on hybrid-p levels, provided at six-hourly intervals. These variables are then interpolated to the model's N96 resolution.

Monthly varying Coupled Model Intercomparison Project Phase 6 (CMIP 6) anthropogenic and biomass burning emissions of NO_x and other relevant species from AerChemMIP have been used (Collins et al., 2017). No diurnal variations in anthropogenic or biomass burning emissions are applied. Natural emissions are as described by Archibald et al. (2020); in particular, lightning NO_x emissions are interactive and follow the Price and Rind (1992) parameterization, whilst soil NO_x emissions vary monthly but are annually invariant and use the Yienger and Levy (1995) distribution.

Archibald et al. (2020) describe the dry and wet deposition schemes; with their Table 1 listing all oxidised nitrogen species deposited in the model. Dry deposition follows a resistance in series approach (Wesely, 1989), with surface resistances assigned according to the surface types specified by the JULES land surface model (Harper et al., 2018). Wet deposition is calculated using a first-order removal scheme driven by the three-dimensional distribution of convective and stratiform precipitation (Giannakopoulos et al., 1999), and scavenging coefficients related to Henry's Law coefficients for each species.

2.2 Satellite NO_2 data

The Ozone Monitoring Instrument (OMI) is a nadir-viewing sensor that measures radiation at ultraviolet-visible wavelengths, mounted on NASA's EOS Aura satellite (Boersma et al., 2011; Liu et al., 2016). Aura travels at an altitude of 705 km in a sun-synchronous polar orbit and provides daily global coverage with a daytime local equator crossing time of 13:45±15 minutes (Shah et al., 2019). OMI measures backscattered radiation from Earth's atmosphere and surface over the wavelength range 264 - 504 nm, with a spectral resolution between 0.42 - 0.63 nm and a nadir spatial resolution of 13×24 km^2 (Dobber et al., 2006; Levelt et al., 2006). The instrument consists of a telescopic system using CCD detectors which provide it a 114° field of view corresponding to a large swath of 2600 km at the Earth's surface. OMI retrieves the ozone column and profile, aerosols, SO_2, NO_2 and other trace atmospheric constituents such as HCHO, BrO, and OClO using the technique of Differential Optical Absorption Spectroscopy (DOAS). OMI tropospheric column NO_2 data utilised here come from the Tropospheric Emission Monitoring Internet Service (TEMIS) product (Boersma et al., 2011) (DOMINO v2.0). The data has been screened to only include data with a cloud fraction of below 0.2 in addition to the good data flags while excluding data with the OMI row anomaly, using the algorithm of Braak (2010). This product includes AK information which have been used for model-satellite comparison.

The AK describes the vertical structure of the atmospheric profile, accounting for the measurement sensitivity at different locations and times (Vijayaraghavan et al., 2008; Boersma et al., 2016). In other words, the AK is a linear representation of the vertical weighting of information content of retrieval parameters. The AK is specified as a vector, used to provide a measure of the vertical resolution of the estimate (Martin, 2008; Vijayaraghavan et al., 2008). The AKs have been applied to the UKCA model NO_2 as shown in Equation 1:

$$y = A.x \tag{1}$$

where A is the tropospheric AK, from the OMI product, with vertical values at specific pressures, x is the model profile (sub-columns in units of molecules/cm^2), interpolated to the OMI vertical pressure grid, and y is the modified model tropospheric column (model sub-columns with AKs applied totalled up to the satellite defined tropopause). The AKs of each day have been applied to daily model profiles, which are then averaged to produce monthly means. This modified column is then directly compared to the satellite NO_2 column.

In addition to applying the AK, the model data must be sampled at the satellite overpass time. We achieve this by producing hourly model output and matching this to the satellite data. To understand the impacts of sampling at 13:45±15 minutes local time, we compare monthly average NO_2 values (i.e., an average across all times of day) with a monthly average calculated just using values for between 1300-1400 local time. To account for the resolution difference, the OMI satellite data were spatially averaged to match the coarser resolution of the UKCA model grid (N96). This ensures consistency between the datasets and allows for a fair comparison by aligning the spatial scales of observations and simulations. For model-observation comparisons, only days and grid boxes with valid satellite retrievals (cloud fraction < 0.2) are included. This ensures consistency in spatial and temporal sampling, though the comparison may be affected by the clear-sky bias inherent in satellite observations, as NO_2 columns under cloudy conditions are excluded from the analysis.

We used linear regression to calculate trends in tropospheric NO_2 concentrations for both model (UKCA) and observation (OMI) datasets. Annual means were used, inherently removing seasonal variability. Statistical significance was tested at a 95% confidence level (α=0.05). The t-statistic for each trend was calculated as t=trend/SE, where SE is the standard error of the trend estimate. The critical t-value ($t_{critical}$) was obtained from the t-distribution using degrees of freedom (df=n-2, where n is the number of years). Trends were considered significant if $|t| > t_{critical}$, indicating that the trend is unlikely to have occurred by chance with 95% confidence. This approach provides a spatially resolved representation of NO_2 trends and their statistical reliability.

3 Results and discussion

3.1 Seasonal and diurnal variations of the vertical profile of NO_2

Figure 2 shows seasonal and diurnal variations of tropospheric column NO_2 taken directly from the UKCA model (with no vertical weighting) over N India and E China. UKCA tropospheric column NO_2 is generally lower over both regions during the late morning and early afternoon (the satellite overpass time), due to photochemical destruction which peaks around local

mid-day, before NO_2 increases in the late afternoon. While NO_2 levels remain high during the evening/night in E China, column values decrease in N India. The diurnal cycle shows the lowest daily range in June-July-August (JJA), varying from ~5-10 × 10^{15} molecules/cm^2 (N India) and ~9-11 × 10^{15} molecules/cm^2 (E China). By contrast, the December-January-February (DJF) diurnal cycle shows the largest ranges: 8-17 × 10^{15} molecules/cm^2 (N India) and 30-55 × 10^{15} molecules/cm^2 (E China). Please note that the diurnal variations depicted are solely from model simulations and cannot be directly compared with OMI data, as there is only one observation time per day. Seasonal diurnal variations of tropospheric column NO_2 for all sub-regions are shown in Figure S2. It is important to note that the UKCA does not have a diurnal cycle in emissions, so the model doesn't simulate higher NO_2 levels related to real-world processes like late afternoon rush hour. Rather, higher levels of NO_2 in the late afternoon arise solely due to dynamical and photochemical processes.

Figure 3 shows the seasonal variation in vertical profiles of NO_2 over N India and E China, as simulated by the UKCA model. The highest levels of NO_2 are found in the boundary layer, close to sources, and during winter, when the boundary layer is shallowest and when NO_2 loss chemistry proceeds more slowly. Levels of NO_2 above the boundary layer are much lower, and show seasonal maxima in summer at 10 km. NO_2 in the upper troposphere reflects a balance between sources associated with enhanced convection and lightning during the monsoon being partly offset by the higher summer photolysis rates. Equivalent average seasonal vertical profiles (2005-2015) for all regions are shown in Figure S3, whereas Figure S4 show the trends of the vertical profiles from 2005 to 2015 over all regions, which highlights that the vertical extent of NO_2 is relatively less affected by pollution in W China and S India in comparison to E China and N India. In addition to pollution levels, meteorological factors, particularly temperature, play a significant role in modulating the vertical distribution and lifetime of NO_2. Lower winter temperatures slow chemical reactions, extending the NO_2 lifetime. This effect, combined with shallow boundary layers and stable atmospheric conditions, contributes to higher NO_2 concentrations near the surface and alters its vertical distribution (Atkinson, 2000; Liu et al., 2016).

Figure 4 shows the diurnal variation of the NO_2 vertical profiles over the same regions for the four seasons. The solid black line in Figure 4 shows the boundary layer height (BLH) of the model which is highest during afternoon (~1-2 km). Higher surface NO_2 values occur at night, and the overpass time of OMI is close to the daily minimum values of NO_2 throughout the vertical column, and the maximum BLH. Typically, an increase in the surface NO_2 concentration is observed after sunset, and the modelled BLH rapidly collapses to well below 100 m. Comparative vertical profiles for all regions are shown in Figure S5. Diurnal and seasonal variations in the boundary layer height for UKCA, ERA-5 and ERA-Interim are shown in Figure S6 and S7. These variations significantly affect NO_2 vertical profiles, particularly during the night, concentrating NO_2 near the surface.

3.2 Model-satellite data comparisons: time sampling and averaging kernel impacts

The importance of time sampling and application of the AK for model-satellite comparison is shown in Figure 5, using monthly mean OMI data averaged over the whole of S/E Asia for 2005, and comparing it with UKCA column NO_2 data generated in several ways: (i) simple monthly mean, with no AK weighting; (ii) time-matched to the satellite overpass time using hourly

model data, but with no AK weighting; and (iii) time-matched and modified by the AK weighting. Measurement uncertainty is based on the daily variation over the month.

Figure 5 shows S/E Asia regional mean OMI tropospheric column NO_2 ranges between 1.0 and 2.0 $\times10^{15}$ molecules/cm^2 over 2005, with measurement uncertainty 0.5-1.0 $\times10^{15}$ molecules/cm^2. In comparison, the UKCA simple monthly mean tropospheric column NO_2 values are larger: 2.2-2.5 $\times10^{15}$ molecules/cm^2 in summer and over 4.0 $\times10^{15}$ molecules/cm^2 in winter. Whilst the OMI tropospheric column NO_2 is measured at 13:45 local time (LT), when the NO_2 is typically relatively low (Figures 2 and 4), the modelled simple monthly mean incorporated all time periods. Therefore, the simple monthly mean modelled NO_2 is substantially larger. In contrast, once the diurnal cycle is accounted for (i.e., UKCA is sub-sampled at the satellite overpass time), modelled NO_2 is in much better agreement with OMI, with near zero biases in summer but in winter the model still overestimates (by ~80%). When the AKs are applied to the model (in addition to sub-sampling at 13:45+15 minutes LT) the summer biases remain near-zero and the winter overestimation is reduced (to ~50%) and the model seasonal cycle now sits within the satellite uncertainty range. Inclusion of the AKs has a greater impact in winter due to the shallow boundary layer which confines NO_2 near the surface (e.g. as illustrated by UKCA in Figures 4 and S5), where satellites like OMI are less sensitive because of increased aerosols, clouds, and reduced sunlight. These factors require stronger AK corrections to align models with satellite data. Boersma et al. (2008, 2016) emphasize the importance of AKs in reducing such discrepancies, while Martin (2008) highlights their role in adjusting for seasonal and vertical variability in NO_2 profiles.

Figure 5 clearly demonstrates the importance of accounting for satellite vertical sensitivities and temporal sampling when evaluating model simulations. The variation in seasonal biases following the use of the AKs is more pronounced in winter and less evident in summer. Applying the correct time sampling is much more important than including the AKs effect. In all the subsequent analysis presented here, we only show UKCA NO_2 columns sampled at the overpass time and with the AK applied.

3.3 Seasonal and spatial variations of tropospheric NO_2 column

Figure 6 shows the seasonal distribution of tropospheric NO_2 observed by OMI and simulated by UKCA, averaged between 2005 and 2015 over S/E Asia. The largest tropospheric NO_2 columns can be seen over E China in DJF from both OMI (>20 x 10^{15} molecules/cm^2) and UKCA (>30 $\times10^{15}$ molecules/cm^2). The seasonal minimum (in JJA) tropospheric column values compare well between OMI and UKCA and typically peak around 6-10 $\times10^{15}$ molecules/cm^2. Comparing UKCA and OMI indicates that the model is overestimating tropospheric column NO_2 in the major polluted regions (e.g., E China and the Indo-Gangetic Plain), especially in DJF. Over E China the differences range from +50% to +100% in March-April-May (MAM), JJA and September-October-November (SON). In DJF, the model overestimation (over +150%) is more widespread and also covers the pollution outflow regions (e.g., Pacific Ocean). The peak biases in India are also in DJF and are +100% to +150%. In the background, less polluted regions, the model tends to underestimate the observations by up to 100% in all seasons.

Scatter plots of OMI vs. UKCA tropospheric column NO_2 (Figure 7) confirm that UKCA overestimates observations in polluted regions in all seasons. The model generally performs best in JJA, and worst in DJF. UKCA captures the observed spatial variability well with R^2 values of 0.87, 0.77, 0.89 and 0.88 for MAM, JJA, SON and DJF, respectively. To understand

the biases due to the higher values another best fit, after removing the highest 10% of observed values, has been computed and plotted (green line). This shows that the model is performing better over the first 90% of the OMI data, although the fit is not improved in DJF. The main problem with the model appears to be an overestimate of NO_2 column over the most polluted regions, especially in winter. Seasonal variations in the boundary layer height (Figure S7) reveal discrepancies between UKCA and ERA datasets, which may partly explain the model's overestimation of NO_2 columns during winter.

3.4 Regional OMI and UKCA tropospheric column NO_2 variability

Figure 8 shows time series (2005-2015) of OMI and UKCA simulated tropospheric NO_2 over the whole of India and China, together with over the four regions indicated by the boxes in Figure 1b. OMI tropospheric column NO_2 typically varies between $1-2 \times 10^{15}$ molecules/cm^2 over India (Figure 8a) with a relatively small seasonal cycle. Over China tropospheric column NO_2 ranges between $2-4 \times 10^{15}$ molecules/cm^2 (Figure 8b) with more pronounced seasonality. UKCA tropospheric column NO_2 typically ranges between $2-5 \times 10^{15}$ molecules/cm^2 and $2-12 \times 10^{15}$ molecules/cm^2 over India and China, respectively. Seasonality is captured by UKCA but the amplitude is overstated by a factor of 2-3.

Figure 8c and 8d show UKCA and OMI tropospheric column NO_2 over N India and E China where values typically range between $1-2 \times 10^{15}$ and $5-20 \times 10^{15}$ molecules/cm^2, respectively. Rapid industrialization, urbanization and increased traffic activity have resulted in a significant increase in the air pollution over E China and N India in the past decades (Ghude et al., 2008; Kar et al., 2010; Mijling et al., 2013). This can be seen in the OMI data in E China between 2005 and 2011 as tropospheric column NO_2 has increased from approximately 12 to 19×10^{15} molecules/cm^2. The signal in N India is much smaller. Again, UKCA tropospheric column NO_2 captures the observed seasonality ($5-12 \times 10^{15}$ molecules/cm^2, N India; $4-45 \times 10^{15}$ molecules/cm^2, E China) but overstates the amplitude. UKCA reproduces the observed trend in E China, but not S India, and overestimates their magnitudes in both cases.

OMI and UKCA trends over the relatively clean regions of S India and W China are shown in Figure 8e and 8f, respectively. S India has an OMI lower tropospheric column NO_2 ($1-2.5 \times 10^{15}$ molecules/cm^2) and UKCA provides a good representation of the observed seasonality and magnitude. UKCA reproduces the observed marginal increase of tropospheric column NO_2 between 2005 and 2011. In W China, the observed tropospheric column NO_2 ranges between $0.5-1.5 \times 10^{15}$ molecules/cm^2 which UKCA struggles to reproduce in magnitude (~50% lower).

3.5 Trends of NO_2 over the years 2005-2011 and 2011-2015

Over China NO_x emission increased by 52% from 2005 to 2011 and thereafter decreased by 21% from 2011 to 2015 (Figure 1c) as reported elsewhere (De Foy et al., 2016; Liu et al., 2017). Therefore, we focus on OMI and UKCA trends over these time periods; the year 2011 is included in both time periods. Figures 9 and 10 show the spatial distribution of significant NO_2 trends for the periods 2005-2011 and 2011-2015, respectively. We observed the largest trends in DJF particularly in UKCA between 2005 to 2011. The seasonal variations in OMI trends are small between 2005 and 2011, with differences of approximately 0.5×10^{15} molecules/cm^2/yr. Equivalent data showing percentage changes from 2005 to 2011 are shown in

Figure S8. However, there have been increases in DJF across E China of up to 1.0×10^{15} molecules/cm^2/yr, and decreases of up to 0.5×10^{15} molecules/cm^2/yr observed over Japan, South Korea and Hong Kong. UKCA shows similar spatial distributions of changes across the majority of the domain, but overstates the magnitudes of decreases over Japan, South Korea, Hong Kong, and increases over E China. There are also substantial model decreases (approximately -1.0×10^{15} molecules/cm^2/yr) over East China in SON which are not present in the OMI observations. Between 2011 and 2015, both OMI and UKCA changes show a steady decrease of up to 2.0×10^{15} molecules/cm^2/yr over E China in almost all seasons (Figure 10 right panel). There are only small changes in the OMI trends over the India from 2011-2015, although a decrease of up to 0.5×10^{15} molecules/cm^2/yr is observed in OMI over N India in DJF (Figure 10 left panel). The corresponding percentage data from 2011 to 2015 is presented in Figure S9. Significant trends, determined at the 95% confidence level, reveal distinct spatial and seasonal patterns. For 2005-2011, some regions, especially in winter (DJF), show significant increases, indicating seasonal variations in emissions and boundary layer dynamics (Figure 9). For 2011–2015, the trends are more pronounced, with notable decreases in NO$_2$ over E China in all seasons, consistent with emission reductions during this period (Figure 10). However, in northeast China, discrepancies between observed and modelled trends suggest uncertainties in the emission inventories used in the UKCA model. The UKCA model captures these significant trends in many regions, though some discrepancies remain, particularly in the magnitude of the trends. These results highlight the ability of the UKCA model to reproduce observed NO$_2$ changes but also underscore areas requiring improvement.

Figures 11 and 12 compare seasonal trends between UKCA and OMI over 2005-2011 and 2011-2015 respectively. UKCA overestimates the magnitudes of trends in NO$_2$ at most locations, with the gradients of best fits (OMI trend over the UKCA trend) in the range 0.15-0.39 for the 2005-2011 (Figure 11), but showing a closer correspondence (0.39-0.67) for 2011-2015 (Figure 12), when the NO$_2$ tropospheric column starts decreasing over China. The overestimation of trends by the model is consistent with the overestimation of NO$_2$ columns in polluted regions, again with the worst agreement in DJF, and better performance in JJA.

4 Discussion

It is well understood that to usefully compare satellite measurements of column NO$_2$ with model simulations, the model atmosphere needs to be sampled in the same way that the satellite samples the real atmosphere (e.g., Boermsa et al, 2008, 2011, 2016). Sampling UKCA at the OMI overpass time and application of a satellite-derived vertical weighting function (averaging kernel) significantly influence the modelled NO$_2$ column and make it more comparable to the OMI values (Figure 5), although differences remain, particularly during winter, when UKCA over-estimates NO$_2$ columns.

The results presented in Figures 2 and 4 illustrate some of the challenges faced by models in accurately simulating column NO$_2$ values measured by satellite instruments such as OMI, particularly during winter (DJF) and at higher latitudes. Diurnal variations in simulated column NO$_2$ for N India and E China (Figure 2) show that at the OMI overpass time the column is changing least (it is approximately flat) in JJA, whilst in DJF it is rising towards a late afternoon peak, particularly further

north. This means that any errors in the shape of the simulated diurnal cycle of NO_2 will translate into larger errors in column NO_2 in winter and at higher latitudes.

320 One source of error in the simulated diurnal cycle of NO_2 arises due to the use of diurnally invariant anthropogenic and biomass burning emissions in these UKCA simulations. Boersma et al. (2008) show that using diurnally varying NO_x emissions has significant effects on the diurnal cycle of the simulated NO_2 column, tending to increase it during daylight hours, as this is when more emissions occur. Hence inclusion of diurnally varying emissions would likely exacerbate the model-observation differences seen in this study.

325 Figures 6 and 7 show that the model is overestimating NO_2 column over the more polluted regions but underestimating it over the cleaner regions. This may reflect a lack of PAN formation, or equivalent sequestration of NO_x in other reservoir species. PAN is a compound that locks up NO_2 in a reaction with the PA (peroxy acetyl) radical (Fiore et al., 2018). The PA comes from oxidation of certain volatile organic compounds (VOCs). PAN is stable at cold temperatures, but unstable at high temperatures, decomposing back to NO_2 and PA. If PAN formation is too low (e.g., because VOCs are too low), this may
330 cause more NO_2 in source regions, and less transport of NO_2 to remote regions.

Another potential contributing factor to the overestimation of NO_2 in source regions may be underestimation of heterogeneous conversion of N_2O_5 to nitrate aerosol (e.g., Dentener and Crutzen, 1993; Riemer et al., 2003; Chen et al., 2018) in UKCA. These modelling studies have shown that this heterogeneous chemistry tends to reduce NOx, especially during winter and in polluted regions with high aerosol loads, and it seems likely that the aerosol surface areas simulated by UKCA in these regions
335 are underestimated.

Modelled trends in column NO_2 over S/E Asia are larger than trends seen in the OMI data, particularly during DJF. This is partly explained by the general overestimation of NO_2 columns, especially in polluted areas. Upwards trends in aerosols would tend to enhance heterogeneous loss of oxidised N, so the underestimation of this process in UKCA would lead to an overestimate of NO_2 trends. Some of the model-observation discrepancies may also reflect uncertainties in emissions
340 magnitudes, spatial distributions, and trends.

Many of these reasons for model-observation differences are likely to be present in other models. Future research, such as a NO_x-focussed model intercomparison and evaluation (cf. van Noije et al., 2006) would help identify and quantify how widespread such problems may be amongst models. Given the central importance of NO_x for multiple environmental issues investigated by models, such future research should be a high priority.

345 **5 Conclusions**

In this work we evaluated tropospheric column NO_2 from the UKCA model using OMI satellite retrievals over S/E Asia. This required sampling the model at the satellite overpass time and application of vertical weighting profiles (averaging kernels) that account for how the satellite retrieval is influenced by the presence of clouds. UKCA can capture the NO_2 seasonality over S/E Asia but generally overestimates NO_2 column, especially in polluted regions during winter. UKCA overestimates column

NO$_2$ near source regions, but underestimates column NO$_2$ in remote regions, suggesting it is not converting enough NO$_2$ into longer-lived reservoir species such as PAN. Overestimations in polluted regions may be due to the UKCA model underestimating heterogeneous conversion of N$_2$O$_5$ to nitrate aerosol, which has been shown to quite strongly reduce NO$_x$ levels in the presence of aerosol, which is present at high levels across much of the region. UKCA also overestimates trends in NO$_2$ column over the region. Underestimation of heterogeneous chemistry may be further contributing to the trend overestimations, as the influence of increases in aerosols over time will be missed.

Given the importance of accurate simulation of oxidised N for many processes important to climate, air quality and the wider environment, further investigation of these discrepancies in simulated NO$_2$ in UKCA are required. In particular, we recommend inclusion of schemes to more comprehensively represent heterogeneous chemistry and diurnal variation of emissions, together with exploration of the VOC emissions and PAN formation mechanisms in the model, to see if their improved representation can lead to improvements in the simulation of column NO$_2$. We also recommend similar studies with other models to understand if these issues are common across models.

Code and Data Availability: This work used the United Kingdom Chemistry and Aerosol model. The model outputs were pre-processed using netCDF Operator (NCO) and Climate Data Operator (CDO). The analysis was carried out using Python.

Author contributions: AKP and DSS conceptualised and planned the research study. AKP performed the UKCA model simulations with support from DSS. AKP performed the model and satellite data analysis with help of AZ. RJP and MPC helped in the satellite and model data comparison. KK and RH commented on the manuscript. AKP and DSS wrote most of the first draft. All authors helped to shape the paper content by editing prior versions of the paper.

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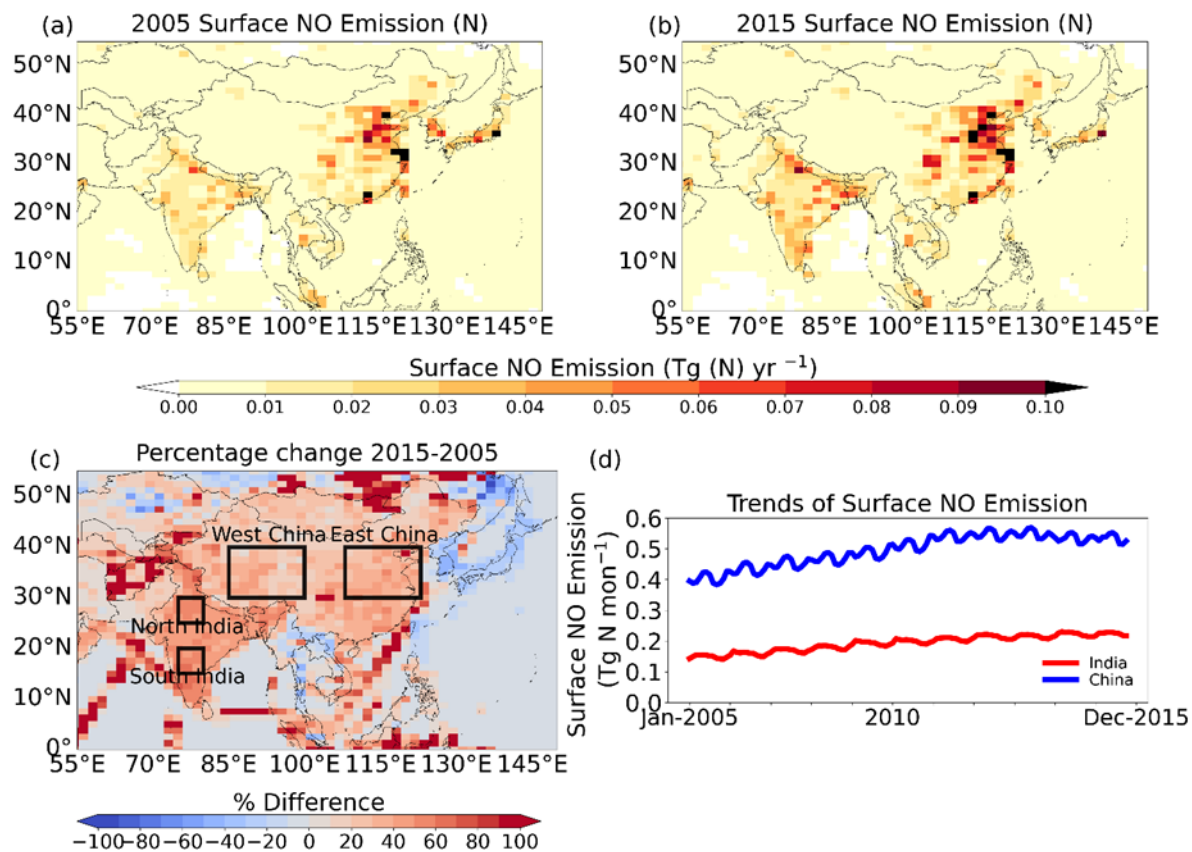
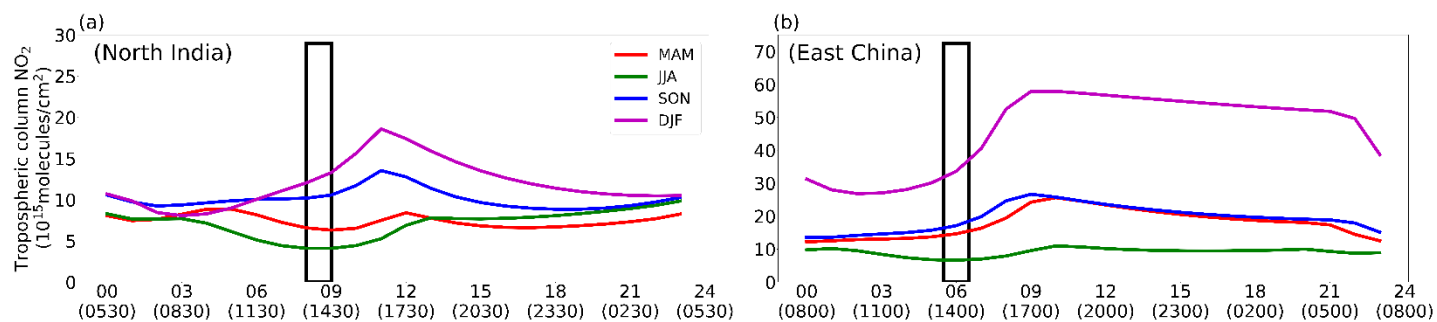


Figure 1 Surface nitrogen oxide (NO) emissions over S/E Asia (Tg N yr⁻¹) in (a) 2005 and (b) 2015; (c) Percentage change in the NO surface emissions from 2005 to 2015; (d) trends of NO surface emissions (Tg N month⁻¹) from 2005 to 2015 over India and China. Boxes shown in (b) indicate regions referred to in the text.

570 **Figure 2 Diurnal cycles of tropospheric column NO₂ (10¹⁵ molecules/cm²) simulated by UKCA over (a) North India and (b) East China for the four seasons (averaged over 2005-2015). The time axis displays UTC, with local time shown in parentheses. The box highlights the OMI overpass time, representing the period used for UKCA-OMI comparisons**

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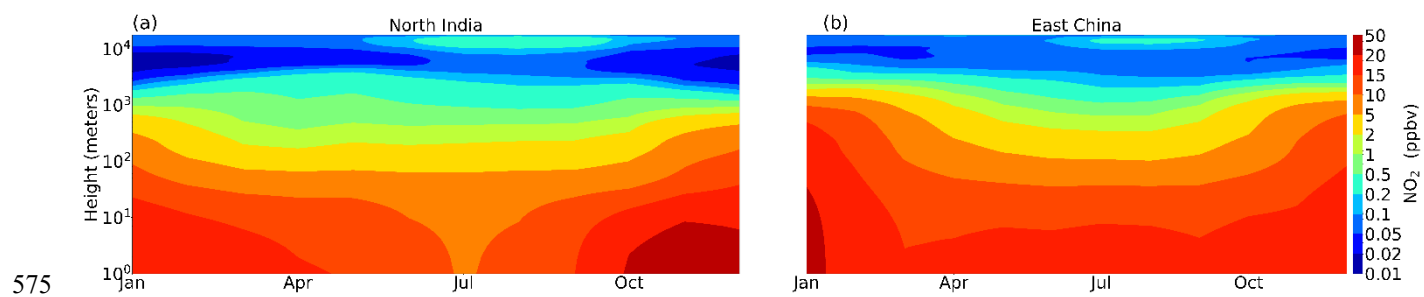
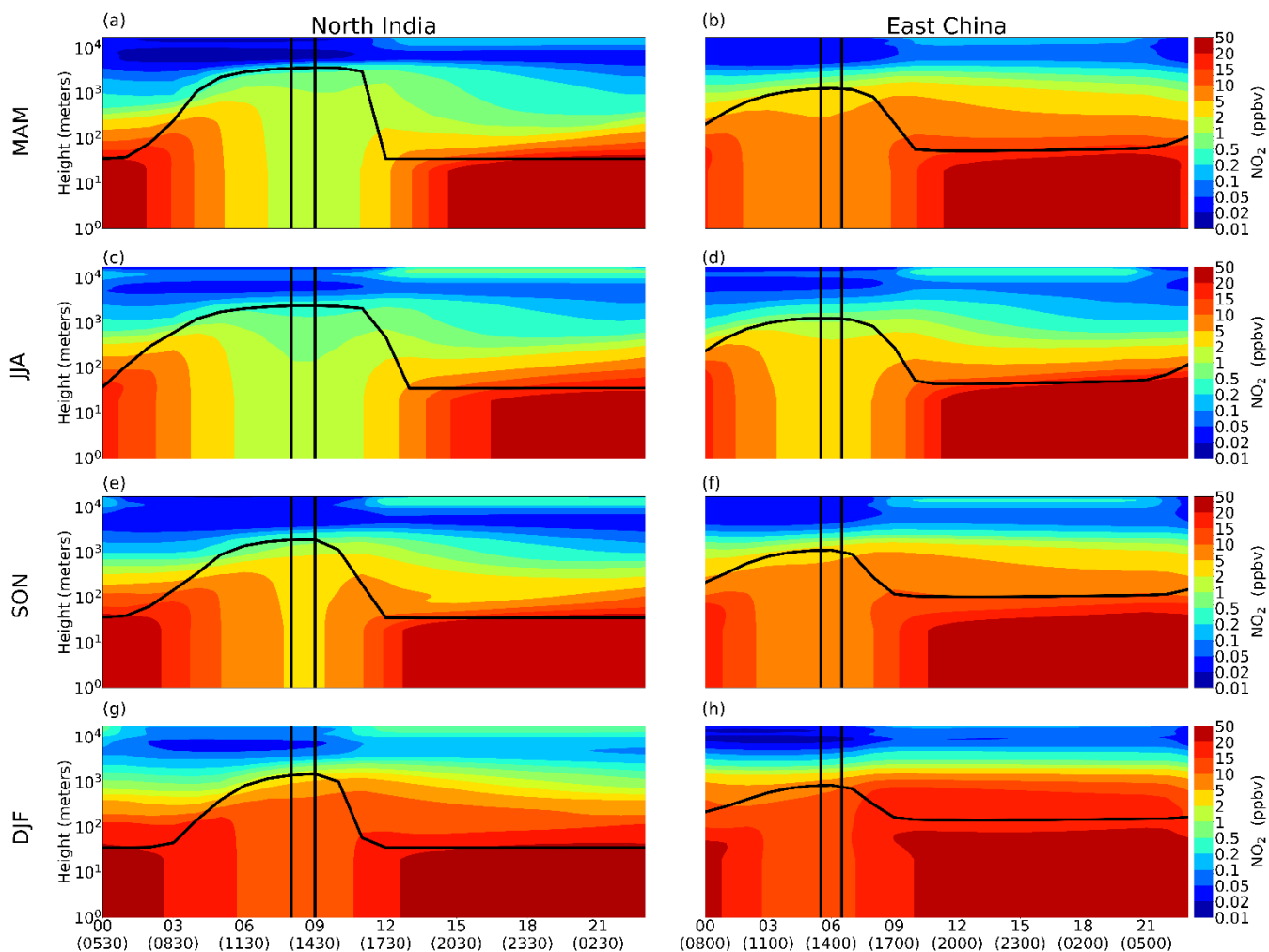


Figure 3 Average seasonal vertical profiles (2005-2015) of NO$_2$ (ppbv) in UKCA over (a) North India and (b) East China.

580   **Figure 4 Diurnal vertical profile of NO₂ (ppbv) simulated by UKCA over N India (left) and E China (right) for the four seasons (averaged over 2005-2015). The time axis shows the time in UTC and, in brackets, the local time. The box is the OMI overpass time. The solid black line shows the boundary layer height in the UKCA model.**

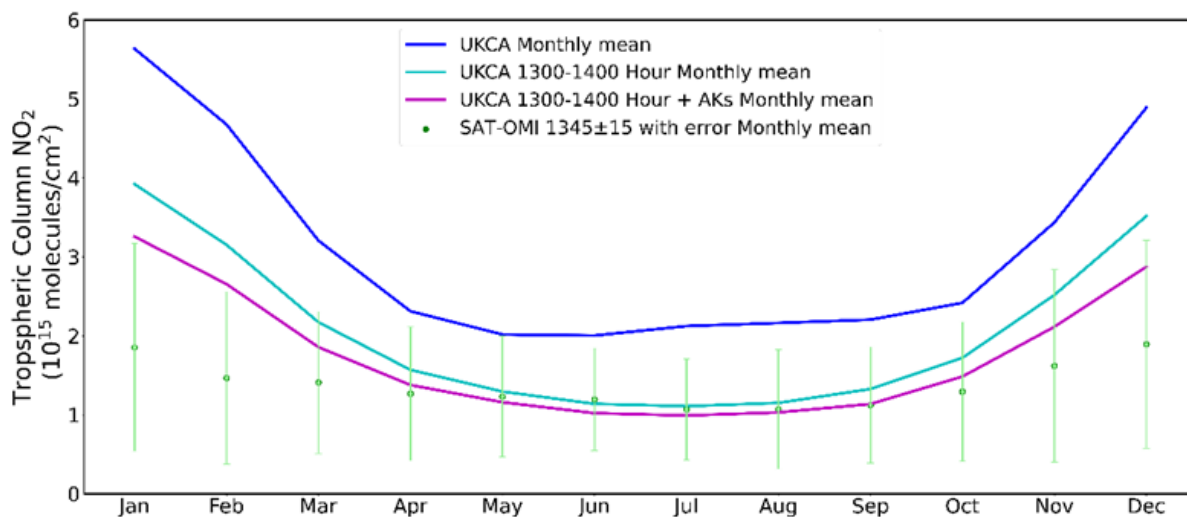
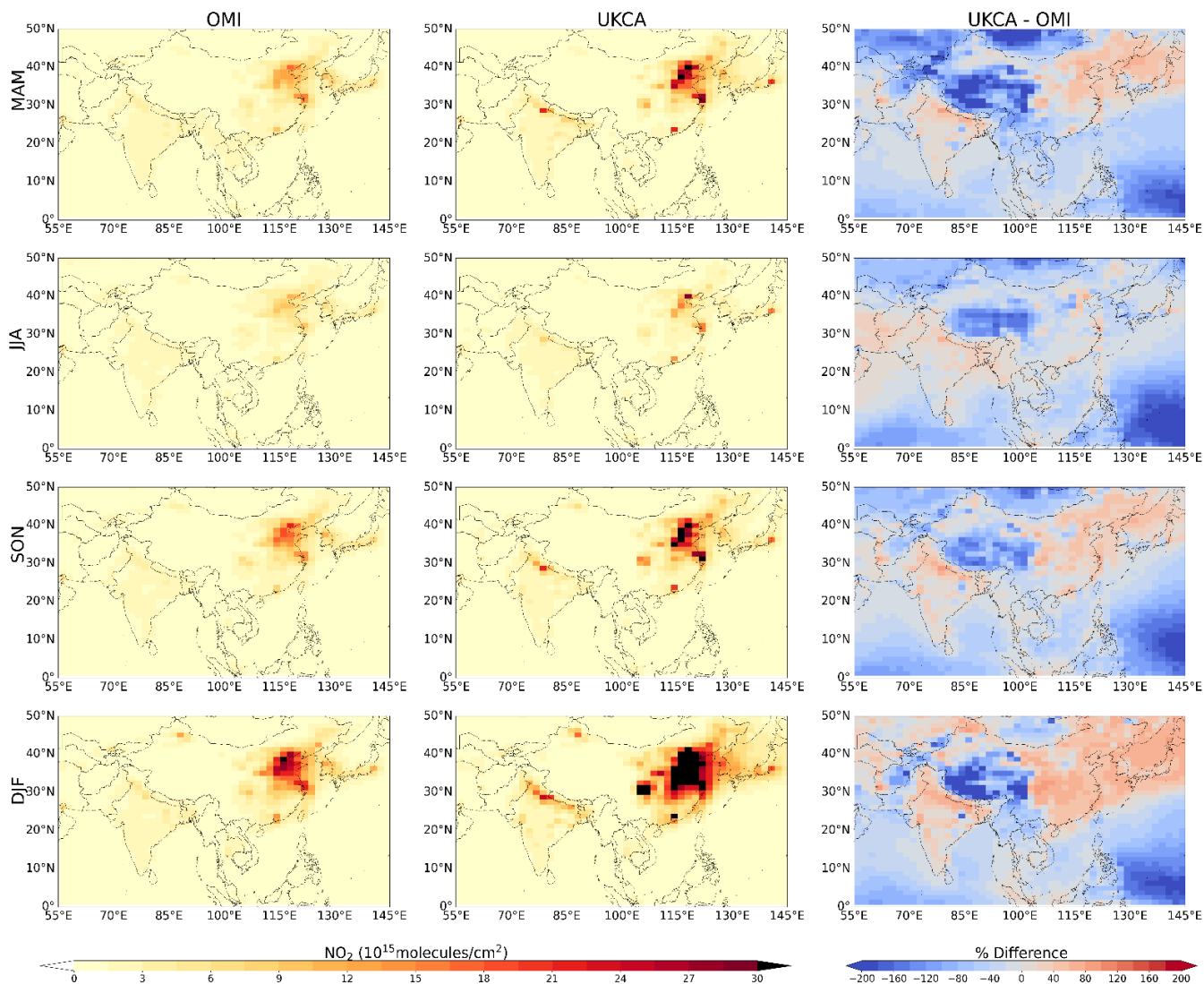



Figure 5 Comparison of monthly mean tropospheric column NO₂, averaged over 2005-2015 for the whole S/E Asia region (Figure 1) from OMI (green, with uncertainty indicated by shading), and from UKCA sampled in three different ways: (i) simple monthly mean (blue); (ii) sampled at the OMI overpass time (cyan); and (iii) sampled at the overpass time and with satellite averaging kernels applied (magenta).

590   **Figure 6 Seasonal tropospheric column NO₂ (10¹⁵ molecules/cm²) distributions from OMI (left), simulated by UKCA (middle), and the percentage difference (100% x (UKCA-OMI)/UKCA)) between UKCA and OMI (right).**

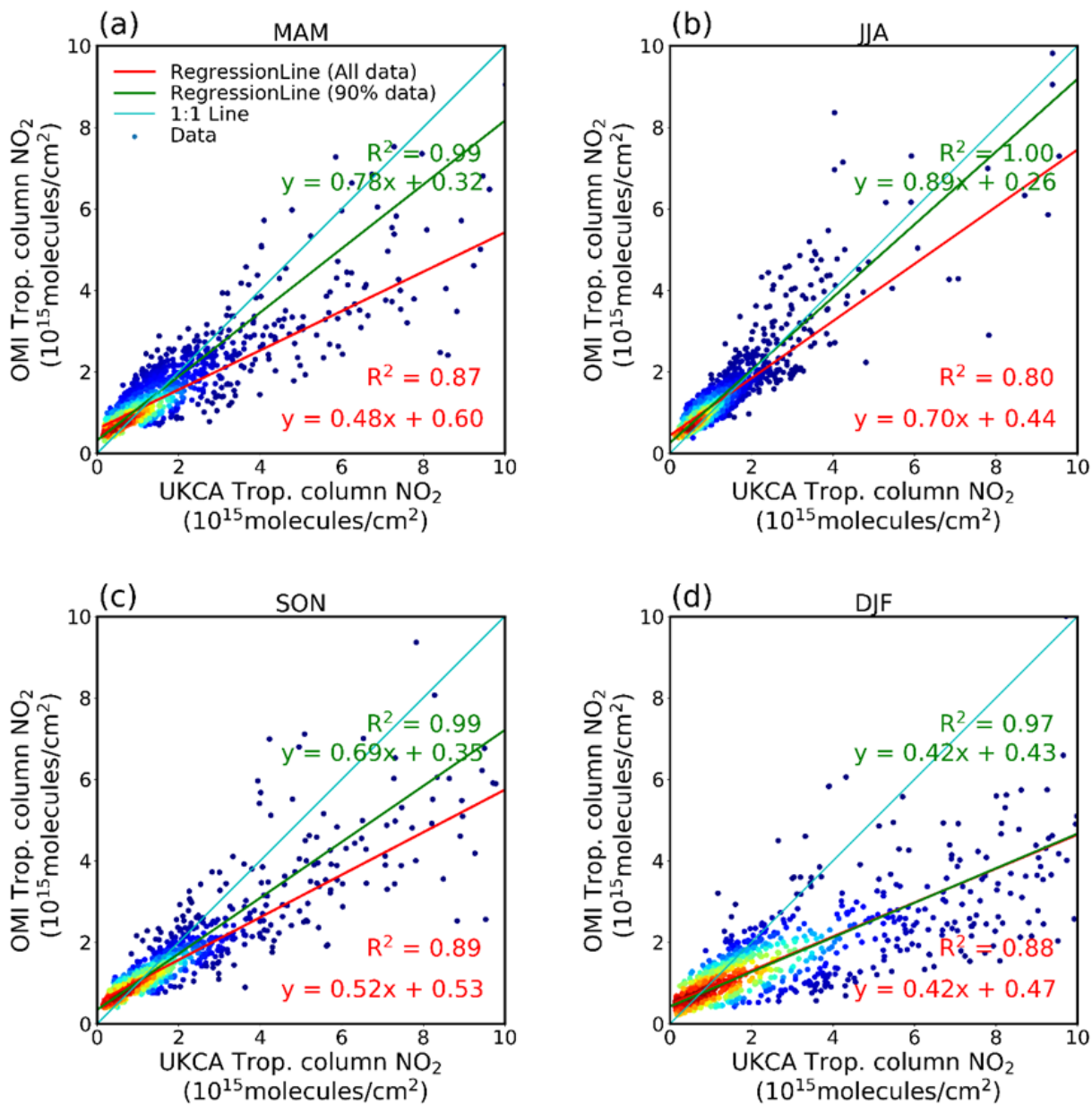



Figure 7 Scatter plots of OMI and UKCA Tropospheric column NO₂ for the four seasons averaged over 2005-2015. Scatter data points are plotted as a heat map where red corresponds to more data. The 1:1 line is shown in cyan colour, best fit in red line (all data) and green line (lowest 90% of data). The equations of best fit and the coefficients of determination (R²) are also shown in the respective colours.

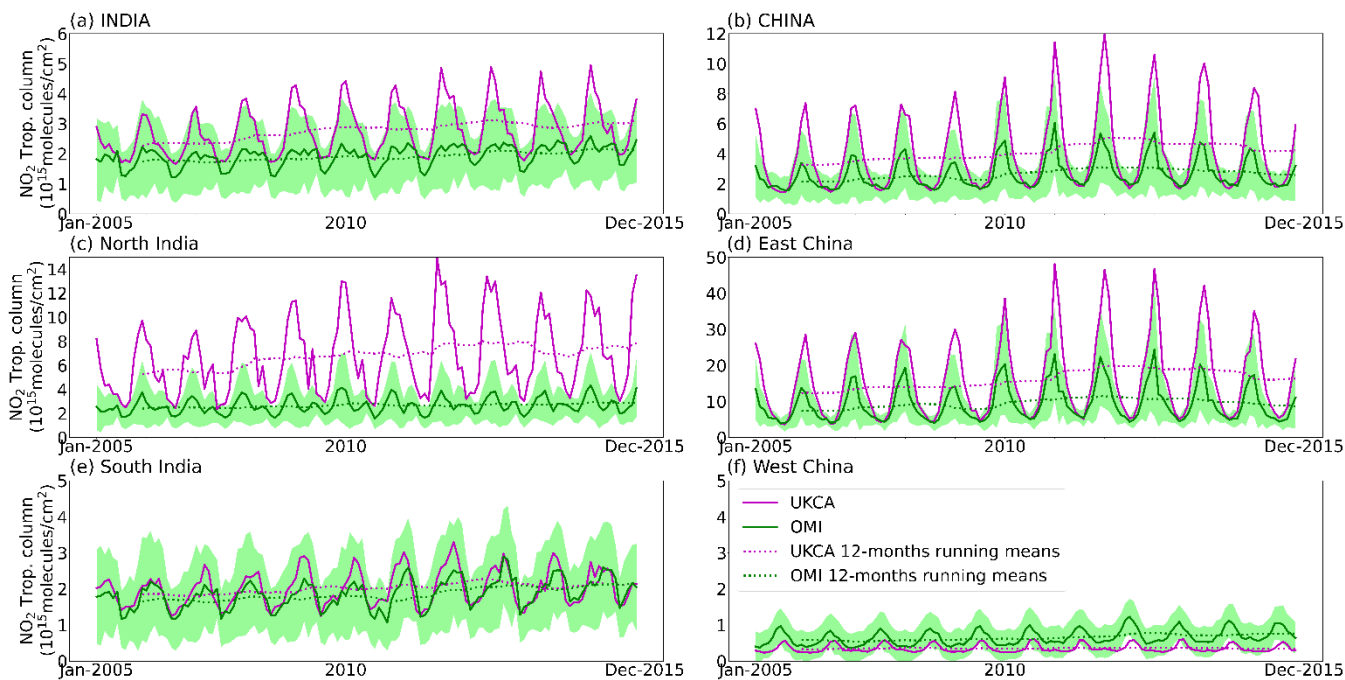



Figure 8 OMI and UKCA tropospheric column NO₂ (10¹⁵ molecules/cm²) time series over (a) India, (b) China, (c) North India, (d) East China, (e) South India and (f) West China. Twelve month running means are shown in the dotted lines. Regions are indicated by the boxes in Figure 1b. Green shading represents the uncertainty in the OMI data.

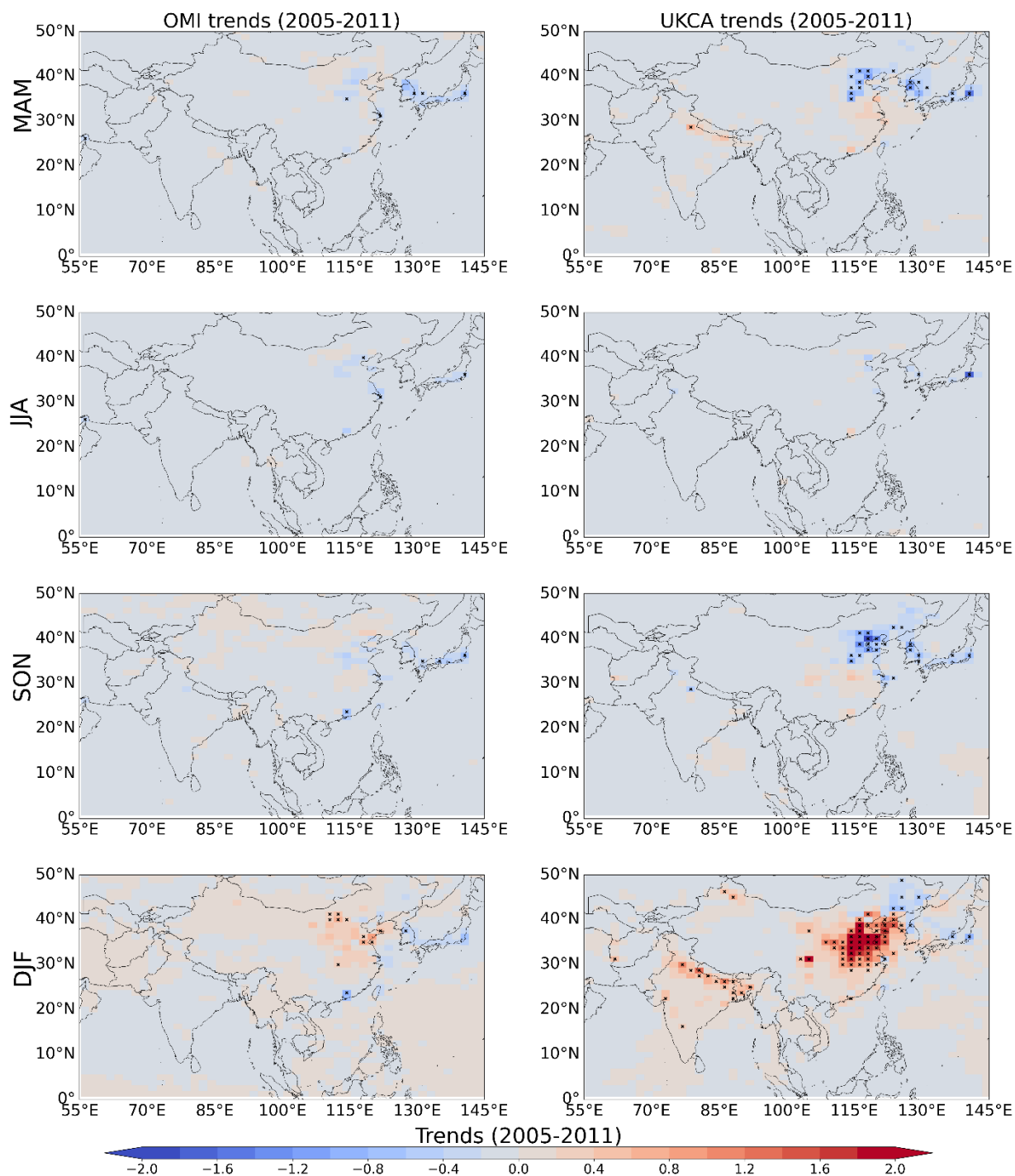



Figure 9 Trends of tropospheric column NO$_2$ (10^{15} molecules/cm^2/yr) from 2005 to 2011 from OMI (left) and UKCA (right) for the four seasons. Scatter plots of these data are shown in Figure 11. Equivalent data in percentage is shown in supplementary Figure S8. Crosses indicate grid squares with significant trends.

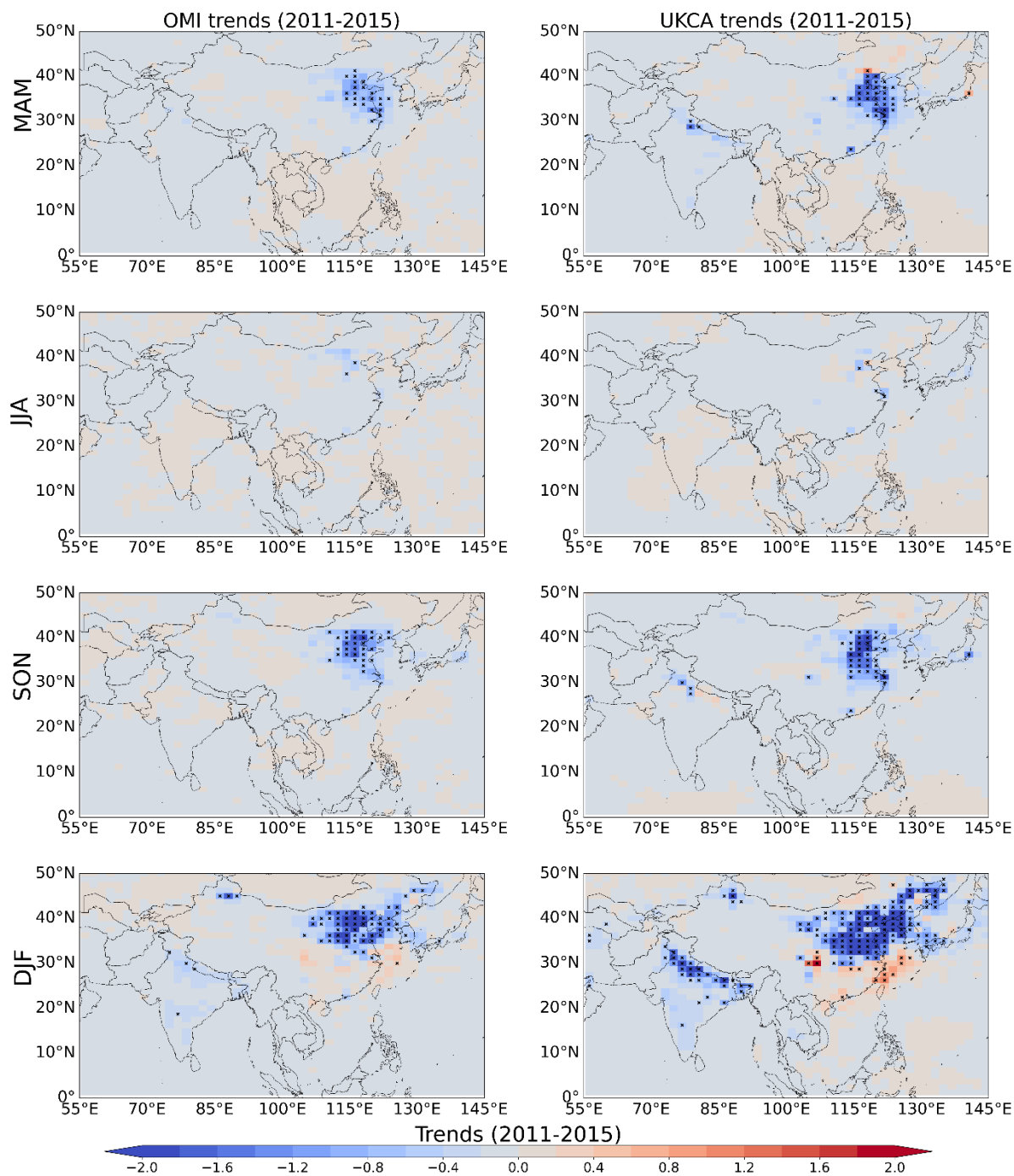



Figure 10 Trends of tropospheric column NO$_2$ (10^{15} molecules/cm^2/yr) from 2011 to 2015 from OMI (left) and UKCA (right) for the four seasons. Scatter plots of these data are shown in Figure 12. Equivalent data in percentage is shown in supplementary Figure S9. Crosses indicate grid squares with significant trends.

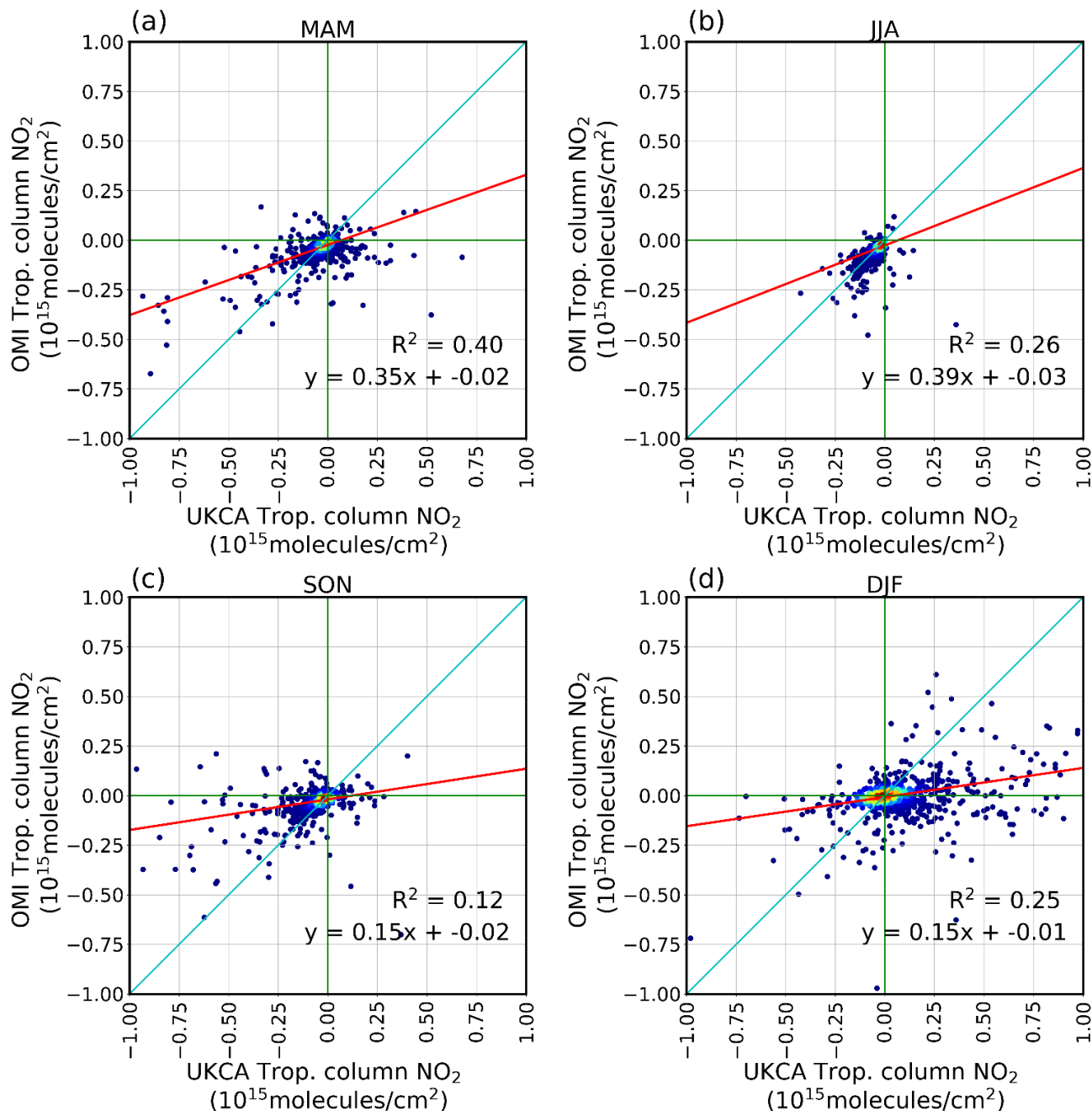



Figure 11 Scatter plot of UKCA and OMI tropospheric column NO₂ trends (10¹⁵ molecules/cm²/yr) from 2005 to 2011 by season.
The 1:1 line is shown in cyan colour, best fit line in red colour. Data points shown as a heat map where red corresponds to more
data. The equation of best fit and the coefficient of determination (R²) are also shown. Note that N = 2050 data points are used for
the fit, though some points fall outside the scales shown.

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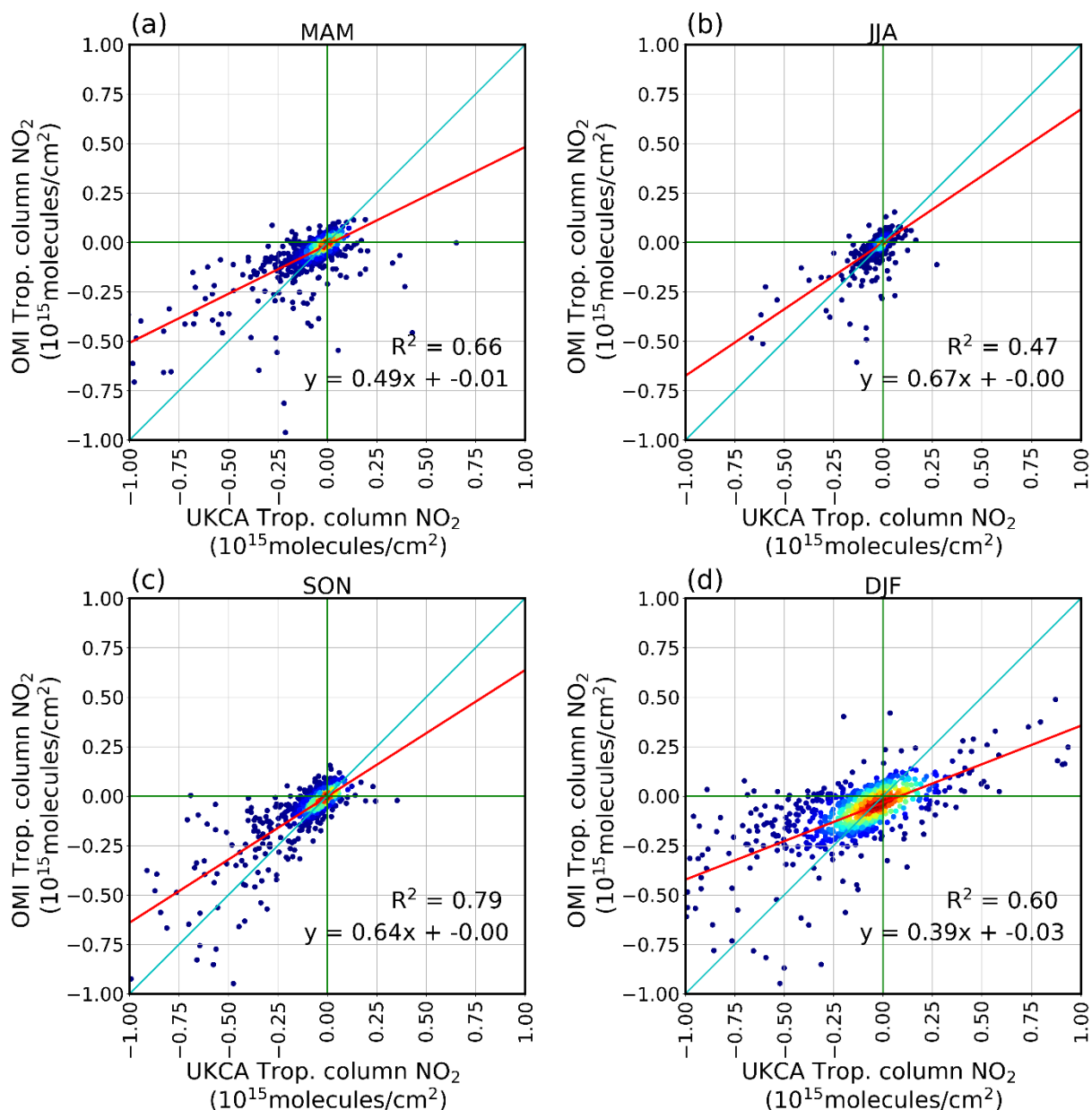
Figure 12 Scatter plots of UKCA and OMI tropospheric column NO₂ trends (10^{15} molecules/cm²/yr) from 2011 to 2015 by season. The 1:1 line is shown in cyan colour, best fit line in red colour. Data points shown as a heat map where red corresponds to more data. The equation of best fit and the coefficient of determination (R^2) are also shown. Note that N = 2050 data points are used for the fit, though some points fall outside the scales shown.