



# Imprints of Increases in Evapotranspiration on Decreases in Streamflow during dry Periods, a large-sample Analysis in Germany

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**Abstract.** Decreases in streamflow ( $Q$ ) during dry periods have the potential to negatively affect river ecosystems and human societies, and understanding their causes is crucial to anticipate them. The contribution of increases in catchment actual evapotranspiration ( $E$ ) to decreases in  $Q$  during dry periods remains poorly quantified. To address this gap, we  
10 performed a data-based analysis for 363 small ( $< 1000 \text{ km}^2$ ) catchments without substantial water management influences in Germany over 1970–2019. We quantified trends in the magnitude of summer low flows, i.e., the minimum 7-day  $Q$  during summer months ( $7dQ_{\min, \text{JJA}}$ ). We attributed these trends to their main potential predictors (namely, long-term variations in  $E$ , summer precipitation,  $P$ , as well as spring and winter  $P$  as proxies for storage). Furthermore, we assessed potential changes  
15 in the annual  $P$ - $Q$  relationship during a multi-year drought in the early 1990s, and investigated whether these changes were related with trends and anomalies in  $E$  and  $P$ . Summer low flows decreased significantly in 31 % of the catchments, with a median trend of  $-3.7 \text{ \% decade}^{-1}$  across all catchments. Increases in  $E$  were a relevant driver of these decreases particularly in relatively drier Eastern catchments (contribution to long-term dynamics of  $7dQ_{\min, \text{JJA}}$  of 35 % based on multiple linear regression, and correlation coefficient between trends in  $7dQ_{\min, \text{JJA}}$  and in  $E$  of  $-0.74$ ). Changes in the  $P$ - $Q$  relationship  
20 occurred in 26 % of the catchments that experienced a multi-year drought between 1989 and 1993, with lower  $Q$  than expected from the relationship before the drought. These changes occurred in catchments with concurrent strong increases in  $E$  (median trend of  $6.1 \text{ \% decade}^{-1}$ ). Our findings point to the importance of increases in  $E$ , especially in dry catchments, when assessing potential future decreases in  $Q$  during dry periods for water management and climate adaptation strategies.

## 1 Introduction

Long-term decreases in low flows and alterations in streamflow ( $Q$ ) generation under extended dryness can alter habitat  
25 availability (Hauer et al., 2013; John et al., 2022) and human water supply (Garreaud et al., 2017; Montanari et al., 2023). As an example, Montanari et al. (2023) showed that the magnitude of summer low flows in the Po river (northern Italy) declined since the 1940s, with the minimum in 2022. This extreme event had severe ecological and socio-economic impacts, through deterioration of water quality, reductions in power production (Montanari et al., 2023 and references therein), and restrictions on public water use (Avanzi et al., 2024) for instance.

30 The magnitude of low flows (or shortly low flows in the following) has decreased in various regions worldwide over recent decades (Stahl et al., 2010; Fangmann et al., 2013; Thomas et al., 2015; Bormann & Pinter, 2017; Hammond et al., 2022;

Chagas et al., 2022). Stahl et al. (2010) analysed summer low flows for 441 small catchments (area < 1000 km<sup>2</sup>) in Europe, and they found mostly decreases in summer low flows, i.e., drying conditions, between 1962 and 2004. Specifically, Stahl et al. (2010) reported strong decreases (increases) in minimum 7-day  $Q$ , typically used to assess potential ecological impacts, for 46 (23) % of the catchments and in minimum 30-day  $Q$ , typically used to assess potential impacts on human societies, for 42 (24) % of the catchments. Bormann & Pinter (2017) later studied long-term variations in annual low flows for 79 small-to-large catchments ( $481 \leq \text{area} \leq 159300 \text{ km}^2$ ) in Germany between 1950 and 2013, and they reported significant increases, i.e., wetting conditions, in minimum 7-day  $Q$  for 26 % of the catchments and decreases for 4 % of them. Bormann & Pinter (2017) showed that increases in low flows occurred in Southern and Western Germany, while decreases occurred in the North-East of the country. They argued that changes in climate and human activities drove this spatial pattern, with increases in air temperature leading to increases in winter low flows in large, snowmelt-affected catchments in Western Germany and changes in human activities, such as mining, to the decreases in the North-East of the country. Small catchments without substantial influences of water management (or shortly small catchments henceforth) allow to study changes mainly driven by climate (Thomas et al., 2015; Hodgkins et al., 2024). Furthermore, focusing only on the summer season avoids influences of snow processes into the generation of low flows (Floriantic et al., 2020). Yet, an updated assessment of changes in summer low flows for small catchments in Germany over recent decades is lacking.

Long-term decreases in summer low flows in small catchments may originate from long-term increases in catchment actual evapotranspiration ( $E$ ), decreases in precipitation ( $P$ ) during the summer season, and decreases in water storage in the catchments ( $S$ , Montanari et al., 2023). Over recent decades,  $E$  has increased in many regions in Central Europe, including Austria (Duethmann and Blöschl, 2018), Czech Republic (Fischer et al., 2023), and Germany (Boeing et al., 2024; Bruno & Duethmann, 2024).  $P$  during summer mostly decreased in South-Central Germany and increased in the North-East of the country over 1951–2013 according to a European-wide gauge-based  $P$  dataset (Duan et al., 2019). Güntner et al. (2023) further reported long-term decreases in  $S$  across Germany and the period 2002–2021 from remote sensing observations, despite empirical uncertainties. From reanalysis data, Montanari et al. (2023) showed that increases in  $E$  and decreases in snow storage co-occurred with decreases in summer low flows in the Po river. The relative importance of long-term variations in different hydro-climatic drivers, including  $E$ , to decreases in summer low flows has not been quantified though. Dry periods with abnormally low  $P$  can extend over multiple years (multi-year droughts). In some catchments, multi-year droughts led to changes in  $Q$  generation, with lower  $Q$  than what expected from the typical annual  $P$ - $Q$  relationship reported in Australia (Saft et al., 2015), California (Avanzi et al., 2020), Chile (Alvarez-Garreton et al., 2021), and Europe (Massari et al., 2022). Massari et al., (2022) investigated potential changes in the  $P$ - $Q$  relationship during various multi-year droughts that occurred in 210 catchments in Europe over 1980–2015 (e.g., the multi-year drought in Central Europe in the early 1990s, Spinoni et al., 2015). They found that these changes occurred in 33 % of the catchments, with a median decrease in  $Q$  across the catchments of -28 % compared to the typical relationship. Potential causes of alterations in  $Q$  generation during extended dry periods are increases in  $E$  and decreases in  $S$ , with further influence of the severity of  $P$  deficits (Saft et al., 2016). Massari et al. (2022) showed positive  $E$  anomalies during the investigated multi-year droughts in Europe, which

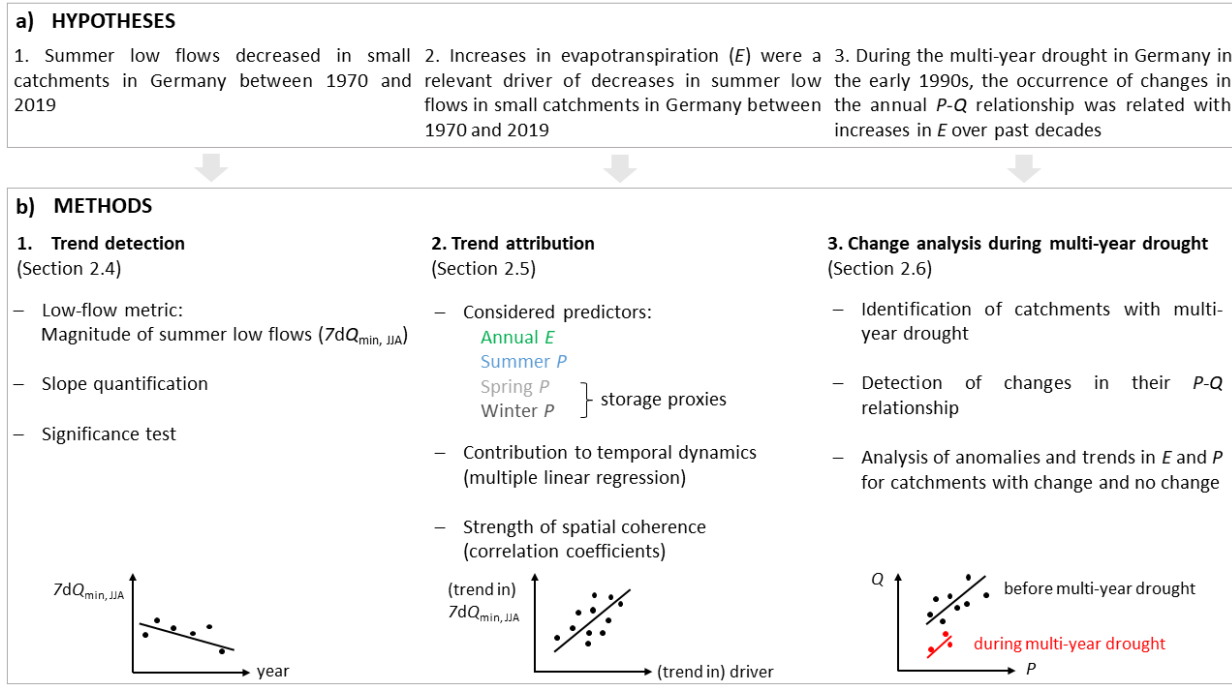


likely contributed to changes in the  $P$ - $Q$  relationship. Positive  $E$  anomalies during multi-year droughts may be caused by high atmospheric evaporative demand in warm-dry conditions and sustained by a depletion of  $S$  (Massari et al., 2022). Underlying long-term increases in  $E$  may also lead to positive  $E$  anomalies during multi-year droughts and thus, to changes in the  $P$ - $Q$  relationship. Gardiya Weligamage et al. (2023) investigated potential long-term increases in  $E$  for six catchments in South-Eastern Australia which experienced changes in the  $P$ - $Q$  relationship during the Millennium drought (ca. 1997–2010). Yet, they showed stable  $E$  before and after the drought, and pointed to decreases in  $S$  as dominant driver of changes in the  $P$ - $Q$  relationship during this event. Bruno & Duethmann (2024) reported long-term increases in  $E$  in small catchments in Germany between the 1970s and 2000s. However, whether changes in the  $P$ - $Q$  relationship during the multi-year drought in Germany in the early 1990s occurred in catchments with long-term increases in  $E$  has not been assessed yet.

To address the three research gaps delineated before, we set three working hypotheses:

1. Summer low flows decreased in small catchments in Germany between 1970 and 2019.
2. Increases in  $E$  were a relevant driver of decreases in summer low flows in small catchments in Germany between 1970 and 2019.
3. During the multi-year drought in Germany in the early 1990s, the occurrence of changes in the annual  $P$ - $Q$  relationship was related with increases in  $E$  over past decades.

To test the working hypotheses, we focused on a large-sample of catchments in Germany and we performed three main analyses (Fig. 1). Firstly, we detected trends in the magnitude of summer low flows (Sect. 2.4). Secondly, we attributed the trends in summer low flows to their main potential predictors (i.e., annual  $E$  and seasonal  $P$ , Sect. 2.5). For this trend attribution, we quantified (i) the contribution of each predictor to the long-term temporal dynamics of summer low flows, and (ii) the strength of the spatial coherence between trends in summer low flows and in the considered predictors (Sect. 2.5). Thirdly, we identified catchments which experienced a multi-year drought in the early 1990s, detected changes in their  $P$ - $Q$  relationship during this event compared to the one before the drought, and analysed the underlying trends and anomalies in  $E$  and  $P$ , for catchments with change and no change (Sect. 2.6).



90 **Fig. 1: Study design. (a) Working hypotheses and (b) overview of the methods used to test them.**

## 2 Data and methods

### 2.1 Study catchments: streamflow data, clustering, and attributes

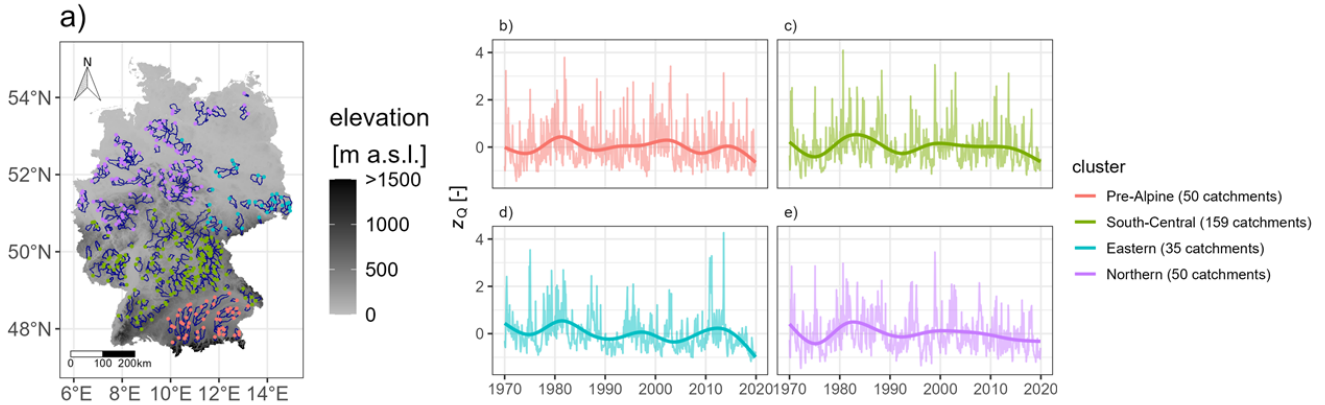
We collected daily  $Q$  observations and catchment boundaries for 363 catchments in Germany (Fig. 2a) from the Environment Agencies of the German Federal States, the Global Runoff Data Center, and the Global Streamflow Indices and Metadata Archive (Do et al., 2018a; 2018b). These catchments have no substantial influences of water management, according to information from the Environment Agencies, and have an area between 50 and 1000 km<sup>2</sup>. Furthermore, the selected catchments have  $Q$  data covering at least 48 years between 1970 and 2019, with less than 5 % of missing values in each year. Throughout the manuscript, we defined years on a hydrological basis, starting in November.

To identify regions with homogeneous long-term variations in  $Q$  over the study period and ease the trend attribution (Fig. 1b), we grouped the catchments into clusters. To this end, we focused on variations in standardized monthly  $Q$  anomalies, rather than variations in a specific low-flow metric, since various metrics can characterize low flows. For comparability across catchments with potentially different hydrological regimes, we computed standardized monthly  $Q$  anomalies (or z-scores,  $Z_Q$ ) as:

$$Z_{Q_j}(t) = \frac{Q_j(t) - \bar{Q}_j}{\sigma_{Q_j}} \quad (1)$$



where  $\overline{Q_j}$  is the mean and  $\sigma_{Q_j}$  the standard deviation in  $Q$  for month  $j$  ( $j = 1 \dots 12$ ) over the study period. We then used a  
 105 hierarchical clustering algorithm with Ward's criterion to minimize the variance of  $Z_Q$  within the clusters (Ward, 1963). We  
 set the number of clusters as four (Fig. 2), since the within-cluster variance reduced comparatively less for higher numbers  
 (Fig. S1).



110 **Fig. 2: Study catchments and streamflow ( $Q$ ) data. (a) Map of catchment boundaries (blue lines) and their outlets (dots, coloured according to the cluster they belong to), with elevation from EU-DEM (2016) as background. (b-e) Average monthly standardized  $Q$  anomalies ( $Z_Q$ ) across the catchments in the different clusters (solid lines for data smoothed by generalized additive models).**

For catchment characterization, we derived a set of attributes regarding topography, climate, land cover, and hydrology as summarized in Table 1.

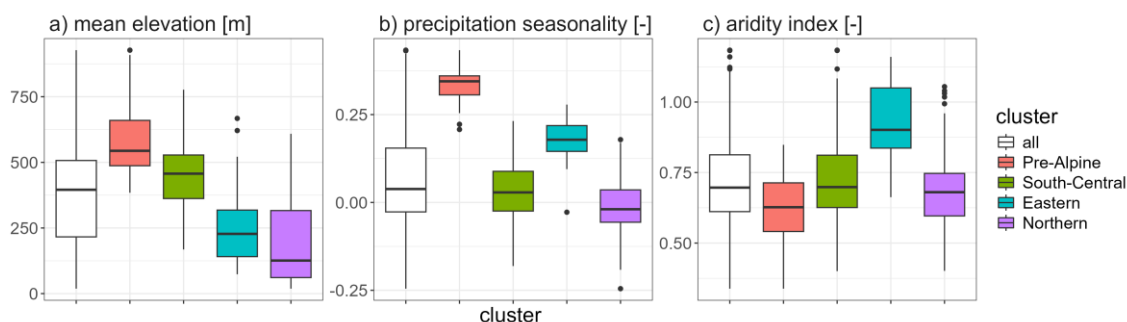
115 **Table 1: Summary of attributes for catchment characterization. Attribute, brief description with relevant references, data sources, median (first/third quartile) value across all catchments, and unit. See Fig. 3 and S2 for distributions of the attributes across the different clusters.**

Attribute	Description	Data source	Median (first/third quartile)	Unit
Area	Catchment area from catchment boundaries	Environment Agencies of the German Federal States and the Global Streamflow Indices and Metadata Archive (Do et al., 2018a; 2018b)	181 (106/331)	km <sup>2</sup>
Slope	Catchment-average slope	EU Digital Elevation Model (v1.1., EU-DEM, 2016)	5 (3/6)	°
Mean elevation	Catchment-average elevation		396 (216/507)	m a.s.l.
Snow fraction	Ratio between $P$ falling as snow (air temperature < 0°) and total $P$	E-OBS dataset (v26.0e, Cornes et al., 2018)	0.07 (0.05/0.09)	-
Precipitation	Seasonality index from Woods (2009), 0		0.04 (-)	-



seasonality	stands for uniform $P$ throughout the year, 1 for strong peak in summer season, -1 for strong peak in winter season		0.03/0.16)	
Aridity index	Ratio between mean daily potential evapotranspiration (Penman-Monteith formulation, Allen et al., 1998) and $P$		0.7 (0.61/0.81)	-
Urban areas	Percentage of urban areas, croplands, pastures, and forests within the catchment	Corine 2000 Land Cover data (v20.1, European Environment Agency, 2020)	5 (3/7)	%
Croplands			43 (26/60)	%
Pastures			7 (4/13)	%
Forests			33 (21/45)	%
Sand	Catchment-average percentage of sand, silt, and clay	SoilGrids250m dataset (Hengl et al., 2017)	40 (34/45)	%
Silt			39 (36/43)	%
Clay			20 (17/22)	%
Runoff ratio	Ratio between mean daily streamflow ( $Q$ ) and $P$	Environment Agencies of the German Federal States and the Global Runoff Data Center, E-OBS dataset (v26.0e, Cornes et al., 2018)	0.34 (0.26/0.43)	-
Baseflow index	Ratio between baseflow (Ladson et al., 2013) and $Q$		0.66 (0.59/0.74)	-

According to these attributes (Table 1), the study catchments are predominantly small (median area of 181 km<sup>2</sup> across the catchments), flat (median slope of 5 °), and lowland (median mean elevation of 396 m), especially in the Eastern and Northern clusters (Fig. 3a). They receive little snow (median snow fraction of 0.07) and  $P$  is generally distributed rather evenly throughout the year (slightly skewed to the summer season for the Pre-Alpine cluster in particular, Fig. 3b). The catchments are mostly humid, with an aridity index occasionally greater than 1, particularly in the Eastern cluster (Fig. 3c). The dominant land cover types are croplands and forests, with generally low presence of pastures and urban areas. The catchments exhibit mostly sandy and silty soils. Runoff ratios are generally low (median of 0.34) and the contribution of baseflow to  $Q$  moderate (median baseflow index of 0.66).





**Fig. 3: Selected attributes for the study catchments. Boxplots of (a) mean elevation, (b) precipitation seasonality, and (c) aridity index for all catchments and by cluster. Details on the attributes are in Table 1 and boxplots of additional attributes in Fig. S2.**

## 2.2 Precipitation data

130 Given the focus of the work on long-term variations, we selected a gridded  $P$  dataset which interpolates data only from gauges with continuous record over time to minimize inhomogeneities from a varying number of gauges. The dataset, previously used in Hoffmann et al. (2018) and Bruno & Duethmann (2024), exploits the SPHEREMAP method (Shephard, 1968; Willmott et al., 1985) for the interpolation and it provides daily  $P$  fields over Germany at a  $0.11^\circ$  resolution.

We corrected the dataset for gauge undercatch according to Richter (1995):

$$P_{corr} = P_{uncorr} + aP_{uncorr}^b \quad (2)$$

135 with  $P_{corr}$  as corrected  $P$ ,  $P_{uncorr}$  uncorrected  $P$ ,  $a$  and  $b$  coefficients which vary with wind exposure of the gauges, precipitation type (rain or snow), and season. We used the coefficients proposed by Richter (1995) for moderately sheltered locations in Germany. To discriminate between rain and snow, we relied on daily air temperature observations from the E-OBS dataset (v26.0e, Cornes et al., 2018). We then computed area-weighted catchment average time series from  $P_{corr}$  fields ( $P$  in the following) for each catchment.

## 140 2.3 Water-balance derived catchment evapotranspiration

We derived annual  $E$  from the water balance:

$$E = P - Q - \Delta S \quad (3)$$

with  $P$  as annual precipitation,  $Q$  as annual streamflow, and  $\Delta S$  as annual changes in  $S$ .

Since long-term data on  $S$  across a large number of small catchments are not available, we approximated  $S$  with the dynamic storage of the catchments ( $S_{dyn}$ , that is the portion of  $S$  connected to variations in  $Q$ ; Staudinger et al., 2017). For catchments and periods in which  $S$  is the main control of  $Q$  generation, Kirchner (2009) showed that  $S_{dyn}$  can be derived from  $Q$  time series and the recession characteristics of the catchments. We followed the details in Bruno & Duethmann (2024, Text S1) for the derivation of the recession characteristics and  $S_{dyn}$ . To constrain unrealistic values in  $S_{dyn}$ , we did not rely on the long-term water balance as in Bruno & Duethmann (2024), but we derived  $S_{dyn}$  between  $Q1$  ( $Q$  exceeded 99% of the days) to avoid numerical issues with low  $Q$  and  $Q85$  ( $Q$  exceeded 15% of the days) to minimize high  $Q$  events with relevant surface processes not controlled by  $S$  (Staudinger et al., 2017).

## 150 2.4 Trend detection for summer low flows

We characterized summer low flows in terms of their magnitude, by averaging  $Q$  over a moving window of 7 days and then selecting the minimum during the summer months June, July, and August ( $7dQ_{min, JJA}$ ). We analysed the influence of the adopted low-flow metric on the results, by using two alternative metrics which we derived extending (i) the moving window to 30 days ( $30dQ_{min, JJA}$ ) and (ii) the summer season from May to October (included,  $7dQ_{min, M-O}$ ).



We detected long-term variations through trend analysis. We used the Sen's slope estimator (Sen, 1968) as a measure of the magnitude of the trends and the Mann-Kendall test, with trend-free prewhitening to avoid lag-one autocorrelation, to verify their significance (Yue et al., 2002). For all statistical tests throughout the manuscript, we used a 5 % significance level.

## 2.5 Trend attribution

160 As predictors of trends in summer low flows ( $7dQ_{\min, JJA}$  in the following), we considered variations in  $E$  and in the cumulative  $P$  over summer ( $P_{JJA}$ ), spring ( $P_{MAM}$ ), and winter ( $P_{DJF}$ ). We used  $P_{MAM}$  and  $P_{DJF}$  as proxies of storage processes (Duethmann et al., 2015). We did not use estimates of  $S_{\text{dyn}}$  (Sect. 2.3) or baseflow from  $Q$  data to have predictors as much independent as possible from the target variable. To disentangle the relative contribution of these predictors, we performed two analyses.

165 Firstly, we modelled the temporal dynamics of summer low flows ( $7dQ_{\min, JJA}$  in Eq.(4) from the dynamics of the predictors through multiple linear regression:

$$7dQ_{\min, JJA} = \alpha_1 E + \alpha_2 P_{JJA} + \alpha_3 P_{MAM} + \alpha_4 P_{DJF} + \varepsilon \quad (4)$$

with  $\alpha_i$  ( $i = 1 \dots 4$ ) the regression coefficient for each predictor and  $\varepsilon$  the model residuals. We used cluster-average time series to minimize potential data issues in specific catchments, while analysing the main signal at a regional scale. Moreover, we adopted 5-year averages to focus on long-term dynamics and further reduce storage effects in water balance-derived  $E$ . We verified that the considered predictors did not have a high level of multicollinearity (variance inflation factors  $< 10$ , Draper and Smith, 1998). We assessed the performance of the multiple linear regressions in terms of coefficient of determination ( $R^2$ ) between predicted and observed summer low flow. We determined significant predictors according to a two-sided t-test, and expressed the contribution of each predictor  $j$  ( $j = 1 \dots 4$ ) as:

$$C_j = \frac{|\tilde{\alpha}_j|}{\sum_{i=1}^4 |\tilde{\alpha}_i|} * 100 \quad (5)$$

175 where  $\tilde{\alpha}_i$  ( $i = 1 \dots 4$ ) are the standardized regression coefficients (i.e., centered around their mean and scaled by their standard deviation) of the considered predictors. We quantified the uncertainty in the regression coefficients as their standard errors.

As a second analysis, we assessed the strength of the spatial coherence between catchment-scale trends in summer low flows and trends in predictors through a correlation analysis (Pearson's correlation coefficient,  $r$ ) for each cluster. We quantified trends in predictors as done for summer low flows (Sect. 2.4).

We repeated both analyses for the alternative low flow metrics (Sect. 2.4).

## 180 2.6 Change analysis during multi-year drought

We identified catchments that experienced the multi-year drought based on annual  $P$  standardized anomalies ( $Z_P$ ):

$$Z_P(t) = \frac{P(t) - \bar{P}}{\sigma_P} * 100 \quad (6)$$





where  $\bar{P}$  is the mean and  $\sigma_P$  the standard deviation in annual  $P$  over the study period. We relied on the criteria of Saft et al. (2015) and Massari et al. (2022) to determine start and end years of the drought for each catchment. Firstly, we computed 3-year moving average of  $Z_P$  ( $3yZ_P$ ) to minimize the potential effect of single wet years in the identification of the drought (Saft et al., 2015). We then considered multi-year droughts as periods with consecutive negative values in  $3yZ_P$ , by discarding end years with (i)  $Z_P > 0.15$ , and (ii)  $Z_P > 0$  for that year and the previous one (Saft et al., 2015). Finally, we retained only catchments that experienced a multi-year drought longer than 3 years with mean  $Z_P < -0.8$  and concomitant  $Q$  data available (Sect. 2.1), to have relevant drought severity (Massari et al., 2022) and data consistency.

For each catchment with a multi-year drought, we relied on the approach of Saft et al. (2015) to identify changes in the  $P$ - $Q$  relationship during this event compared to the conditions before the drought. To ensure that annual  $Q$  data follow approximately a normal distribution for subsequent parametric analyses, we applied a Box-Cox transformation (Box and Cox, 1964):

$$Q_{BC}(t) = \frac{Q(t)^\lambda - 1}{\lambda} \quad (7)$$

where  $\lambda$  is a transformation parameter that we determined through maximum likelihood estimation (Massari et al., 2022). We then identified changes in the  $P$ - $Q$  relationship during the multi-year drought through multiple linear regression (see Sect. 2.5 for methodological details on the regression):

$$Q_{BC} = \beta_1 + \beta_2 I + \beta_3 P + \varepsilon_1 \quad (8)$$

with  $Q_{BC}$  as the Box-Cox transformed  $Q$  ((7) before and during the multi-year drought),  $I$  a binary indicator (1 for years in the multi-year drought and 0 for years before),  $P$  annual precipitation before and during the multi-year drought,  $\beta_i$  ( $i = 1..3$ ) regression coefficients, and  $\varepsilon_1$  model residuals. A significant  $\beta_2$  indicates a change in the  $P$ - $Q$  relationship, as compared to the one before the drought. For catchments where these changes occurred, we quantified their magnitude as:

$$C_{P-Q \text{ rel}} = \frac{(\beta_1 + \beta_2 I + \beta_3 P^* + \varepsilon_1) - (\beta_1 + \beta_3 P^* + \varepsilon_1)}{(\beta_1 + \beta_2 I + \beta_3 P^* + \varepsilon_1)} * 100 \quad (9)$$

with  $P^*$  as a representative annual  $P$ , given by the average between the minimum and average  $P$  over the years before and during the multi-year drought (Saft et al., 2015) to account for non-linearities in non-transformed  $P$  data.

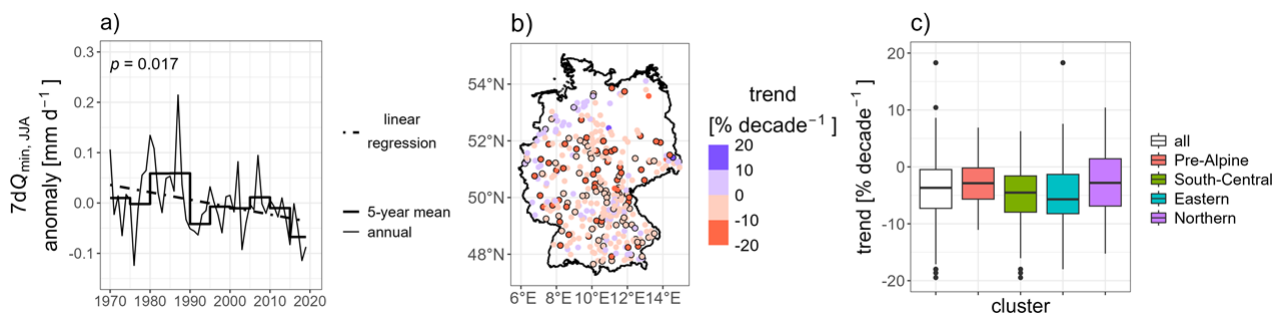
Finally, for each catchment where a multi-year drought occurred, we quantified trends (Sect. 2.4) in  $E$  and  $P$  over the period from 1970 to the end of the multi-year drought, as well as mean annual anomalies (Eq. (6) in  $E$  and  $P$  during the drought. We furthermore computed mean annual anomalies in detrended  $E$  and  $P$ . For the detrending, we removed from the original data the linear trend over the years before and during the drought. We then verified whether catchments with change and no change had statistically different distributions in trends and anomalies according to the Kolmogorov-Smirnov two-sample test.



### 3. Results

#### 3.1 Trends in summer low flows

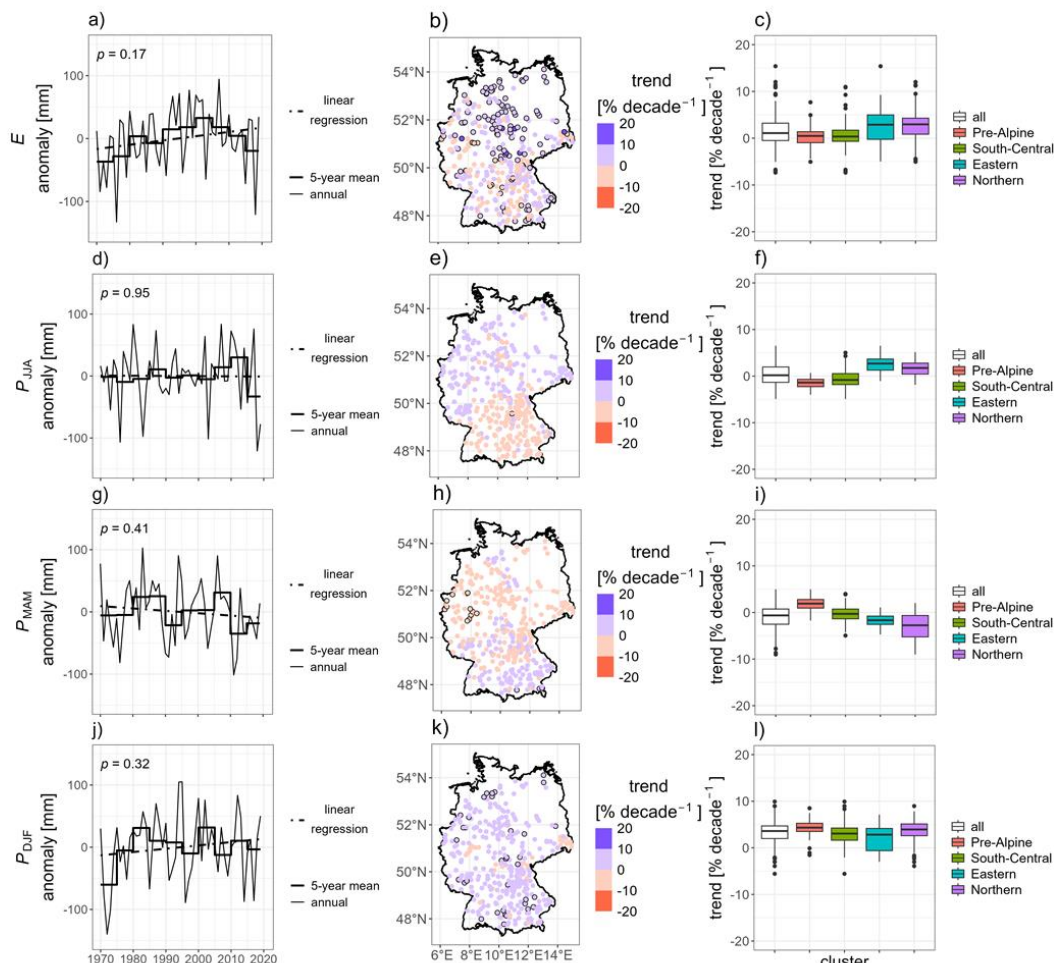
210 Summer low flows largely decreased in small catchments in Germany between 1970 and 2019 (Fig. 4, S6a-d). Mean  $7dQ_{\min, JJA}$  across all catchments showed generally positive anomalies before the 1990s and negative anomalies afterwards (Fig. 4a). Trends in  $7dQ_{\min, JJA}$  were significantly negative in 31 % of the catchments (positive in 2 %, Fig. 4b), with median (interquartile range, IQR) of  $-3.7$  ( $-7.5/-0.6$ ) % decade<sup>-1</sup> across all catchments (Fig. 4c). Each cluster showed negative median trends (the strongest in the Eastern one, equal to  $-5.8$  % decade<sup>-1</sup>, Fig. 4c). Alternative metrics for summer low flows had  
 215 comparable behaviour, with a median trend across the catchments of  $-4.2$  % decade<sup>-1</sup> for  $30dQ_{\min, JJA}$  and of  $-3.1$  % decade<sup>-1</sup> for  $7dQ_{\min, M-O}$ , and significant decreases for 28 % of the catchments for both metrics (Fig. S3 and S4).



220 **Fig. 4: Long-term variations in summer low flows ( $7dQ_{\min, JJA}$ ) over 1970–2019. (a) Average anomalies across all catchments (Average anomalies across the catchments in each cluster in Fig. S6a–d). (b) Map of catchment-scale trends (dots with black edge if significant). (c) Boxplots of trends for all catchments and by cluster.**

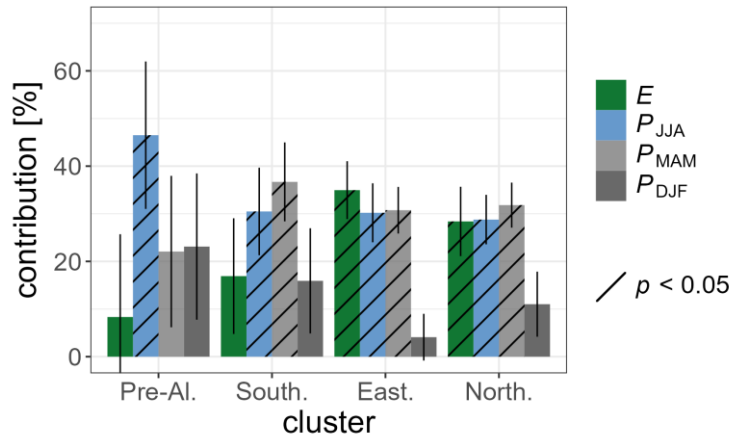
#### 3.2 Attribution of trends in summer low flows

Annual  $E$  increased between 1970 and 2019, with significant increases in 27 % of the catchments and a median trend (IQR) of  $1.1$  ( $-0.5/3.2$ ) % decade<sup>-1</sup> across all of them (Fig. 5a-c). All clusters displayed positive median trends in  $E$ , particularly in the Northern ( $3.0$  % decade<sup>-1</sup>) and Eastern one ( $2.9$  % decade<sup>-1</sup>, Fig. 5c). Increases in  $E$  occurred especially between 1970  
 225 and the early 2000s (Fig. 5a), with a median trend (IQR) of  $4.2$  ( $1.8/6.9$ ) % decade<sup>-1</sup> across the catchments over 1970–1999 (Fig. S5). Other predictors of trends in summer low flows (i.e., long-term variations in  $P_{JJA}$ ,  $P_{MAM}$ , and  $P_{DJF}$ ) showed contrasting directions of change over time (Fig. 5d, g, j) and only small trends (significant in  $< 10$  % of the catchments, Fig. 5e, h, k). Trends in  $P_{JJA}$  and  $P_{MAM}$  further exhibited contrasting signs among the clusters (Fig. 5e, h).  $P_{JJA}$  generally decreased in the Pre-Alpine cluster (median trend of  $-1.4$  % decade<sup>-1</sup>) and the Central one (median trend of  $-0.8$  % decade<sup>-1</sup>),  
 230 while it generally increased in Eastern (median trend of  $2.7$  % decade<sup>-1</sup>) and Northern catchments (median trend of  $1.7$  % decade<sup>-1</sup>, Fig. 5f).  $P_{MAM}$  underwent general increases in the Pre-Alpine cluster (median trend of  $1.9$  % decade<sup>-1</sup>) and decreases elsewhere (median trends of  $-0.3$  % decade<sup>-1</sup> in the Central cluster,  $-1.7$  % decade<sup>-1</sup> in the Eastern one, and  $-2.7$  % decade<sup>-1</sup> in Northern catchments, Fig. 5i). Median trends in  $P_{DJF}$  were slightly positive for all clusters (Fig. 5j).



235 **Fig. 5: Long-term variations in predictors of variations in summer low flows over 1970–2019 (annual evapotranspiration,  $E$ , panels a–c, precipitation over summer,  $P_{JJA}$ , panels d–f, spring,  $P_{MAM}$ , panels g–i, and winter  $P_{DJF}$ , panels j–l). (a, d, g, and j) Average anomalies across the catchments (average anomalies across the catchments in each cluster in Fig. S6e–t). (b, e, h, and k) Maps of catchment-scale trends (black edges if significant). (c, f, i, and l) Boxplots of trends for all catchments and by cluster.**

Multiple linear regression for predicting long-term dynamics in summer low flows achieved satisfactory performances (for  
 240  $7dQ_{\min, JJA}$ ,  $R^2 > 0.7$  for each cluster, Table S1). Significant predictors of  $7dQ_{\min, JJA}$  differed between clusters (Fig. 6, Table S1). In the Pre-Alpine cluster,  $P_{JJA}$  was the only significant predictor (contribution to the simulation of  $7dQ_{\min, JJA}$  equal to 46 %). In the South-Central cluster,  $P_{JJA}$  and  $P_{MAM}$  were significant predictors, with comparable contribution to the simulations (< 40 % for each predictor). In the Eastern and Northern clusters, significant predictors were  $E$ ,  $P_{JJA}$ , and  $P_{MAM}$ , with similar relative contributions (slightly higher for  $E$  than for  $P_{JJA}$  and  $P_{MAM}$  in the Eastern cluster, and equal to 35 %).



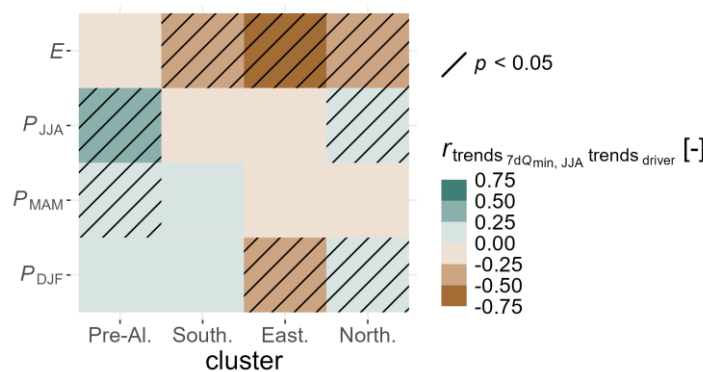
245

**Fig. 6: Attribution of long-term variations in summer low flows to their predictors (contribution to temporal dynamics): relative contribution of annual evapotranspiration ( $E$ ), summer ( $P_{JJA}$ ), spring ( $P_{MAM}$ ), and winter precipitation ( $P_{DJF}$ ) to the predicted long-term dynamics of summer low flows ( $7dQ_{min, JJA}$ ) from multiple linear regression (Sect. 2.5) for the different clusters. Pre-al. refers to Pre-Alpine, South. to South-Central, East. to Eastern, and North. to Northern cluster. Vertical lines indicate the uncertainty of the regression coefficients.**

250

The strength of the correlation between catchment-scale trends in summer low flows and in their predictors varied according to the considered predictor and cluster (Fig. 7). By focusing on significant correlations, trends in  $7dQ_{min, JJA}$  negatively correlated with trends in  $E$  in all clusters but the Pre-Alpine one ( $r$  equal to  $-0.26$  in the South-Central,  $-0.74$  in the Eastern, and  $-0.36$  in the Northern cluster), indicating that decreases in summer low flows corresponded to increases in  $E$  and vice versa. Moreover, trends in  $7dQ_{min, JJA}$  positively correlated with trends in  $P_{JJA}$  for the Pre-Alpine ( $r = 0.3$ ) and Northern ( $r = 0.25$ ) clusters, illustrating that decreases in summer low flows paralleled decreases in  $P_{JJA}$  and vice versa. Trends in  $7dQ_{min, JJA}$  furthermore showed a positive correlation with trends in  $P_{MAM}$  for the Pre-Alpine cluster ( $r = 0.23$ ), a negative correlation with trends in  $P_{DJF}$  for the Eastern cluster ( $r = -0.49$ ), and a positive correlation with trends in  $P_{DJF}$  for the Northern cluster ( $r = 0.16$ ).

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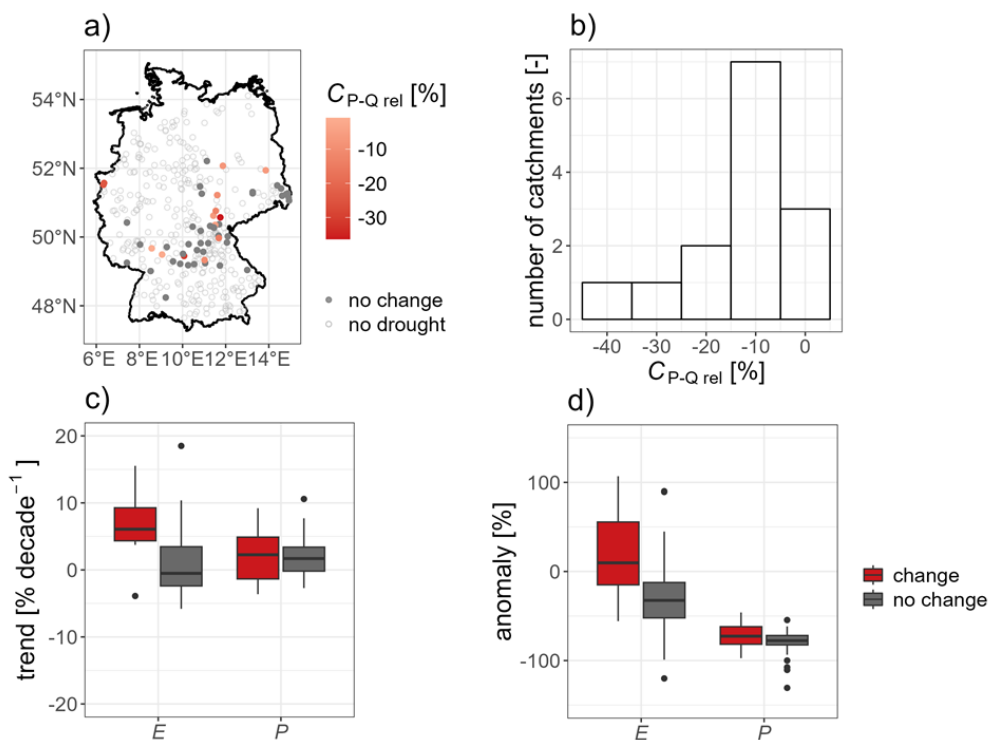
265 **Fig. 7: Attribution of long-term variations in summer low flows to their predictors (strength of spatial coherence): Pearson's correlation coefficients ( $r$ ) between catchment-scale trends in summer low flows ( $7dQ_{\min, JJA}$ ) and in potential predictors (annual evapotranspiration,  $E$ , summer precipitation  $P_{JJA}$ , spring precipitation  $P_{MAM}$ , and winter precipitation  $P_{DJF}$ ) over 1970–2019, for the catchments in the different clusters. Pre-al. refers to Pre-Alpine, South. to South-Central, East. to Eastern and North. to Northern cluster.**

Correlation between trends in  $7dQ_{\min, JJA}$  and in their predictors differed also according to the considered decades. Over 1970–1999, trends in  $7dQ_{\min, JJA}$  for instance negatively correlated with trends in  $E$  in all clusters, including the Pre-Alpine one ( $r = -0.28$ , Fig. S7).

Results for  $30dQ_{\min, JJA}$  and  $7dQ_{\min, M-O}$  (not shown) were comparable to those for  $7dQ_{\min, JJA}$  for both attribution analyses.

### 270 3.3 Changes in the $P$ - $Q$ relationship during a multi-year drought and their potential predictors

Fifteen percent of the study catchments experienced a multi-year drought within 1989–1993 (or drought in the following, Fig. 8a). Of these, 26 % exhibited a change in the annual  $P$ - $Q$  relationship during the drought (Fig. 8a). The median magnitude of change across the catchments was of -10 % of  $Q$  for given  $P$ , as compared to the relationship before the drought (Fig. 8b). The catchments with change showed largely positive trends in  $E$  over the period before and during the drought (median of 6.1, IQR of 4.4/9.2 % decade<sup>-1</sup> across the catchments, Fig. 8c). The catchments with no change had mostly small trends in  $E$  before and during the drought (median of -0.5, IQR of -0.5/3.4 % decade<sup>-1</sup>, distribution statistically different from the one for catchments with change, Fig. 8c). Catchments with and without change similarly experienced generally positive trends in annual  $P$  before and during the drought (Fig. 8c). Mean  $E$  anomalies during the drought were mostly positive across the catchments with change (median of 10 %) and negative for catchments with no change (median of -32 %, Fig. 8d). Mean anomalies in annual  $P$  during the drought were largely negative, and relatively similar for catchments with and without change (Fig. 8d). Mean anomalies in detrended variables ( $E_d$  and  $P_d$ ) during the drought were mostly negative both for catchments with and without change (Fig. S8).



285 **Fig. 8:** Changes in the annual relationship between precipitation ( $P$ ) and streamflow ( $Q$ ,  $P-Q$  relationship) during the multi-year  
drought between 1989 and 1993, and their potential predictors. (a) Map of the magnitude of changes in the  $P-Q$  relationship ( $C_{P-Q}$   
rel), across the study catchments. (b) Histogram of  $C_{P-Q \text{ rel}}$ . (c) Boxplots of trends in annual catchment actual evapotranspiration ( $E$ )  
and  $P$  over 1970–1993, for catchments with change and no change in the  $P-Q$  relationship. (d) Boxplots of mean anomalies in  $E$   
and  $P$  over the drought, for catchments with change and no change in the  $P-Q$  relationship.

## 290 4. Discussion

### 4.1 Summer low flows decreased in small catchments in Germany between 1970 and 2019

Trend analysis confirmed widespread decreases in summer low flows in all clusters (Fig. 4). Decreases in  $7dQ_{\min, \text{JJA}}$  were significant in 31 % of the catchments with a median trend of  $-3.7 \text{ \% decade}^{-1}$  across all catchments (Fig. 4b and c). These findings extend for two more decades the decreases in summer low flows which were previously reported for parts of  
295 Germany up to the mid-2000s (Stahl et al., 2010; Fangmann et al., 2013). Moreover, our results complement observations on increases in annual low flows found by Bormann & Pinter (2017) for catchments with potentially significant human impacts in Germany between 1950 and 2013. Bormann & Pinter (2017) argued that water management was the main driver of these increases, possibly in combination with climatic changes for catchments with their annual  $Q$  minimum in winter. Here we analysed only small catchments with no substantial influences of water management, despite the presence of human  
300 activities in land cover (Table 1). For this set of catchments, we observed consistent decreases in summer low flows between 1970 and 2019. Our findings indicate potential differences in low-flow trends between small catchments without substantial



influences of water management and regulated catchments, or between summer and winter low flows, due to different generating processes. The agreement among trends in  $7dQ_{\min, JJA}$  and alternative metrics (i.e.,  $30dQ_{\min, JJA}$  and  $7dQ_{\min, M-O}$ , Fig. S3 and S4) suggests that our main conclusions remain valid irrespective of the specific choice of the low-flow metric. However, we found a slightly lower percentage of significant decreases for  $30dQ_{\min, JJA}$  (Fig. S3b) than for  $7dQ_{\min, JJA}$  (Fig. 305 4b), which may point to stronger potential impacts for river ecosystems than for human societies.

Here we focused on the magnitude of summer low flows. Alternative metrics of low flows also showed changes over past decades. For instance, the timing of low flows displayed mostly negative trends (i.e., low flows occurring earlier in the year) in small catchments in Europe up to 2004 (Stahl et al., 2010), and the spatial extent of low flows both increased and 310 decreased across catchments with potentially significant human impacts in Europe between 1969 and 2011 (Brunner and Gilleland, 2021). For a similar set of small catchments as the one used here, future work could investigate recent long-term variations in these alternative metrics of low flows. Similarly, future studies could focus on streamflow droughts, that are abnormal low flows as compared to the climatology (Van Loon, 2015). Peña-Angulo et al. (2022) found increases in the duration, frequency, and severity of streamflow droughts (i.e., drying conditions over time) in catchments with potentially 315 significant human impacts in Germany between 1962 and 2017. The relevance of water management on these changes might be assessed by selecting catchments with and without substantial influences of water management.

#### **4.2 Increases in $E$ were a relevant driver of decreases in summer low flows in small catchments in Germany between 1970 and 2019**

The analysis of predictors of decreases in summer low flows revealed concomitant increases in  $E$  in all clusters (significant 320 in 27 % of the catchments, Fig. 5b, c, S6), but contrasting changes in seasonal  $P$  across the clusters (Fig. 5, S6). Specifically, we found that (i)  $P_{JJA}$  decreased in the Pre-Alpine and South-Central Germany, and increased elsewhere, (ii)  $P_{MAM}$  generally decreased, and (iii)  $P_{DJF}$  generally increased over 1970–2019 (Fig. 5b, e, h), which is coherent with Duan et al. (2019). These changes indicate that widespread increases in evaporative losses ( $E$ ), local decreases in water input during summer ( $P_{JJA}$ ), and local decreases in storage recharge during spring (approximated by  $P_{MAM}$ ) superimposed to generate decreases in 325 summer low flows, and they overcompensated local increases in storage recharge during winter (approximated by  $P_{DJF}$ ). We further quantified the relative importance of these predictors and showed that the main drivers differed by cluster, i.e., by region with distinct variations in  $Q$  over time. Increases in  $E$  were a significant driver of decreases in  $7dQ_{\min, JJA}$  in the Eastern and Northern clusters, together with changes in seasonal  $P$ , whereas decreases in  $P_{JJA}$  dominated the decreases in  $7dQ_{\min, JJA}$  in the Pre-Alpine cluster (Fig. 6 and 7). Both multiple linear regressions, focusing on drivers of temporal 330 variations, and the correlation analysis, focusing on drivers of spatial patterns, agreed on these results. For the South-Central cluster the two analyses led to mixed results, which may be due to the generally small trends in this cluster (Fig. 5). These results suggest giving particular attention to catchments with strong increases in  $E$  (e.g., in the Eastern and Northern clusters, Fig. 5c) to anticipate potential decreases in summer low flows. Moreover, the relevance of increases in  $E$  for decreases in summer low flows was strongest for catchments with the highest aridity indices in our sample (i.e., the Eastern cluster),



335 although still mostly humid (aridity index  $< 1$ , Fig. 3c). Therefore, the aridity index appears as a key attribute to identify  
catchments where summer low flows are particularly vulnerable to increases in  $E$ . These findings expand results by  
Montanari et al. (2023) on concomitant decreases in summer low flows and increases in  $E$  in northern Italy over recent  
decades. Furthermore, they align with the results of Thomas et al. (2015), Fangmann & Haberlandt (2019), and Floriancic et  
al. (2021) on the importance of potential evapotranspiration for the interannual variability of low flows in German  
340 catchments, and those of Hammond et al. (2022) on decreases in low flows in catchments in the USA where aridity  
increased. Increases in  $E$  showed a halt in Germany after the 2000s (Fig. 4a and Bruno & Duethmann, 2024), leading to  
milder trends in  $E$  over 1970–2019 than over 1970–1999 (Fig. S5). Increases in  $E$  over 1970–1999 correlated with decreases  
in summer low flows in the Pre-Alpine cluster as well (Fig. S7), suggesting that the contribution of increases in  $E$  to  
decreases in summer low flows may be even larger than the one that we illustrated here, under continuing  $E$  increases.

345 High  $E$  contributed to streamflow droughts in the Alps during the warm and dry summer 2003 (Mastrotheodoros et al.,  
2020), and delayed the recovery of  $S$  deficits during the 2018–2021 drought in Germany (Boeing et al., 2024). However, the  
effect of increases in  $E$  on potential increases in streamflow drought conditions remains unclear. Future research could  
quantify it, similarly to the analyses by Lan et al. (2024) to disentangle the role of vegetation greening on increases in  
streamflow drought conditions in China between 1982 and 2015.

350 **4.3 During the multi-year drought in Germany in the early 1990s, the occurrence of changes in the annual  $P$ - $Q$   
relationship was related with increases in  $E$  over past decades**

Changes in the annual  $P$ - $Q$  relationship during the multi-year drought in Germany in the early 1990s generally occurred in  
catchments with increases in  $E$  (Fig. 8). Following Saft et al. (2015) and Massari et al. (2022), we found that 15 % of the  
study catchments underwent a multi-year drought between 1989 and 1993, and 26 % of these catchments showed a change in  
355 the  $P$ - $Q$  relationship, with respect to the one before the drought (Fig. 8a). For catchments with change, we found a median  
decrease in  $Q$  of -10 % compared to the  $P$ - $Q$  relationship before the drought. Thus, these catchments generated less  $Q$  during  
the multi-year drought than what predicted by their  $P$ - $Q$  relationship before the drought and given  $P$ . These changes should  
not be seen as long-term shifts in the hydrological regime of the catchments, since we did not include in the analysis years  
after the drought (to exclude potential non-recovery from the drought conditions for possibly different processes, Peterson et  
al., 2021). Furthermore, the multi-year drought in Germany in the early 1990s had less severe impact on  $Q$  generation than  
360 the Millennium drought in Australia (changes in 56 % of the catchments, with median decrease of approximately -50 %, Saft  
et al. 2015), the 2012–2016 event in California (mean decreases of -28 % across three catchments, Avanzi et al. 2020), and  
the 2010–2020 drought in Chile (changes in 61 % of the catchments and mean decrease of -19 %, Alvarez-Garreton et al.,  
2021). These differences may be related to the characteristics of the  $P$  anomalies, like their severity and duration, as well as  
365 to hydro-climatic characteristics of the catchments. For catchments in Europe and droughts of similar duration to the one we  
analysed here, Massari et al. (2022) reported changes in 33 % of the catchments, and a median decrease of -28 % across



catchments and droughts. Our findings confirm less widespread and severe changes in the  $P$ - $Q$  relationship for specific droughts in Europe than for other events in different hydro-climatic settings (e.g., the Millennium drought in Australia).

Massari et al. (2022) suggested positive  $E$  anomalies during the multi-year droughts as causes of these changes in Europe. We revealed that catchments with changes during the drought had predominantly strong increases in  $E$  before and during the drought, while catchments with no change mostly had negligible changes in  $E$  over the same period (Fig. 8c). Furthermore, catchments with change in the  $P$ - $Q$  relationship had mostly positive mean anomalies in  $E$  (Fig. 8d), but not in detrended  $E$  (Fig. S8). Our results point out that long-term increases in  $E$  between the 1970s and 1990s were related with positive  $E$  anomalies during the multi-year drought in Germany in the early 1990s and thus, led to decreases in  $Q$  compared to the  $P$ - $Q$  relationship before the drought. Monitoring long-term variations in  $E$ , therefore, can help identifying catchments potentially prone to alterations in  $Q$  generation during extended dry periods. We also found that catchments with and without change in the  $P$ - $Q$  relationship had comparable  $P$  trends (mild long-term increases in  $P$  before and during the drought; Fig. 8c) and comparable  $P$  anomalies (negative  $P$  anomalies of similar magnitude; Fig. 8d) during the analysed multi-year drought. In other words, the catchments with changes in the  $P$ - $Q$  relationship were not characterized by more severe  $P$  deficits or drying trends than those with no change. This reinforces the importance of increases in  $E$  as a predictor of changes in the  $P$ - $Q$  relationship during this event.

Additional processes, such as decreases in the contribution of  $S$  to  $Q$  under prolonged dryness, may have further contributed to changes in the  $P$ - $Q$  relationship during the multi-year drought that we analysed and other events, as reported for the Millennium drought (Trotter et al., 2024; Gardiya Weligamage et al., 2023). Future research might investigate for instance to what extent changes in the  $P$ - $Q$  relationship occurred during other multi-year droughts in Germany over recent decades, such as the 2018–2022 event (Boeing et al., 2024; Sodoge et al., 2024), which our dataset covers only partly, and their potential causes.

#### 4.4 Implications

Summer is the season when  $Q$  is most critical for the health of river ecosystems and specific human uses, such as irrigation. Multi-year droughts can have different impacts on ecosystems and societies than single dry years, as shown for the 2018–2022 drought in Germany which had stronger effects on forestry, recreational activities, aquaculture, and waterborne transportation than the 2003 and 2015 events, with impacts mostly on agriculture (Sodoge et al., 2024). Long-term decreases in summer low flows and stronger reductions in  $Q$  than expected during multi-year droughts can have direct ecological and socio-economic implications. A proper understanding of the causes of these phenomena is required to better predict them. Under further projected future increases in  $E$  (Yang et al., 2023) and multi-year droughts in many regions, as shown by Van Der Wiel et al. (2023) for the Rhine basin for instance, we underline the importance of monitoring changes in  $E$  for the prediction of potential decreases in  $Q$  during dry periods. Furthermore, given the frequent challenges of hydrological models under changing conditions (Saft et al., 2016; Duethmann et al., 2020) and in representing long-term variations in  $E$  (Duethmann et al., 2020), we recommend evaluating models with specific respect to long-term variations in  $E$ , and possibly



400 improving them, when used for climate impact assessments on river ecosystems (Barbarossa et al., 2021) and human societies (Naumann et al., 2021).

## 5. Conclusions

The magnitude of summer low flows widely decreased across 363 small catchments without substantial influences of water management in Germany over 1970–2019, with significant negative trends in 31 % of the catchments (Fig. 4). Long-term  
405 increases in  $E$  were a relevant driver of decreases in summer low flows, particularly for humid-to-dry Eastern catchments, both in terms of contribution to temporal dynamics (35 %, Fig. 6) and coherence of spatial patterns ( $r$  equal to -0.74, Fig. 7). A change in the  $P$ - $Q$  relationship occurred in 26 % of the catchments with a multi-year drought in the early 1990s, with lower  $Q$  than expected from the relationship before the drought. Catchments with a change were characterized by strong underlying increases in  $E$  (median trend of 6.1 % decade<sup>-1</sup>) while catchments without change had on average negligible  
410 changes in  $E$  (median trend -0.5 % decade<sup>-1</sup>; Fig. 8). We illustrated the imprint of long-term increases in  $E$  on decreases in  $Q$  during dry conditions, especially for specific regions and periods. Therefore, the consideration of long-term variations in  $E$  is required to understand and model changes in  $Q$  during dry periods, for the adaptation of water management to global changes.

## Code Availability

415 Codes will be made available upon request.

## Data Availability

Streamflow data and catchment boundaries are available through the Environment Agencies of German Federal States (Landesamt für Umwelt Brandenburg; Landesanstalt für Umwelt Baden-Württemberg; Bayerisches Landesamt für Umwelt; Landesamt für Umwelt, Naturschutz und Geologie Mecklenburg-Vorpommern; Niedersächsischer Landesbetrieb für  
420 Wasserwirtschaft, Küsten- und Naturschutz; Landesamt für Natur, Umwelt und Verbraucherschutz Nordrhein-Westfalen; Landesamt für Umwelt Rheinland-Pfalz; Landesamt für Landwirtschaft, Umwelt und ländliche Räume Schleswig-Holstein; Landesamt für Umwelt- und Arbeitsschutz Saarland; Sächsisches Landesamt für Umwelt, Landwirtschaft und Geologie; Landesbetrieb für Hochwasserschutz und Wasserwirtschaft Sachsen-Anhalt; and Thüringer Landesamt für Umwelt, Bergbau und Naturschutz).



#### 425 **Author contribution**

GB: conceptualization; methodology; software; validation; formal analysis; investigation; data curation; writing – original draft; writing – review and editing; visualization. LS: software; writing – review and editing. DD: methodology; software; formal analysis; investigation; writing – review and editing; supervision; project administration; funding acquisition.

#### **Competing interests**

430 The authors declare that they have no conflict of interest.

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