

Dear Reviewer,

We really appreciate your insightful comments and suggestions that helped a lot to improve the analysis and writing of the manuscript. The point-by-point response to your comments is listed below and the revisions/additions/edits are shown in the tracked-change file.

Review of the Manuscript „Estimating the variability of NO_x emissions from Wuhan with TROPOMI NO₂ data during 2018 to 2023“

This paper by Zhang et al. presents NO_x emission and lifetime estimates for Wuhan based on TROPOMI tropospheric NO₂ VCDs from mid 2018 to 2023 and a superposition column model introduced in previous studies. Seasonal patterns, the weekly cycle, interannual variability, and dependence on wind conditions are investigated and compared to similar studies. Estimated monthly NO_x emissions are compared to the EDGAR and ABACAS emission inventory.

General comments:

The manuscript is a follow-up study of the Zhang et al. (2023) study. Due to a longer dataset, seasonal patterns, the weekly cycle, and the interannual variability can be investigated. The methodology is based on Lorente et al. (2019) and Zhang et al. (2023), however, some adjustments have been made.

Some parts of the methodology are hard to follow, reading only this manuscript without knowing Zhang et al. (2023) and would benefit from a bit more details (see also comments in the specific comment part below). How are data rotated and how is the mean wind calculated (over which area)? Definition of emissions and lifetime? Why an area of 90km and why was it changed compared to Zhang et al. (2023)?

Response: Thank you for the comments and we have reorganized the methodology to make it clearer to follow, some details are added.

We construct a 15 x 15 grid map centered at the city center (114.6 °E, 30.6 °N) with each grid size of 0.05 ° x 0.05 ° (6km x 6km) toward the mean wind direction, with one demission of the grid map along with and the other perpendicular to the wind direction. The mean wind direction is determined by the mean meridional and zonal winds over the study domain. The original TROPOMI observation (Figure 1a) is sampled into the rotated grid (Fig. 1b). The TROPOMI NO₂ columns in the 15 grid cells perpendicular to the wind direction are integrated to form the so-called NO₂ ‘line densities’ (Beirle et al., 2011), resulting in 15 grid cells along the wind direction (Fig. 1c).

We treat each cell along with the wind direction with the simple column model proposed by Jacob (1999), finally NO_x emissions from each cell (E_i) are added up to obtain the total emission E in the study domain. The ‘chemical lifetime’ (τ , h), since

its relationship with the chemical loss rate of k (s^{-1}) in the superposition column model is $k = \frac{1}{\tau \cdot 3600}$.

We narrow down our study domain from the administrative area of Wuhan in Zhang et al. (2023) to the urban area (within the Fourth Ring Road of Wuhan, ~90 km in diameter) in this study considering that, first, most (more than 60%) of the NO_x emissions are concentrated in the urban area; second, we use regional mean wind fields and NO_x chemical loss rate, the larger study domain would induce large uncertainty in the result.

Please refer to the Section 2.4 in the revised manuscript.

I am concerned that some decisions influence the emission estimates, especially the investigated seasonal patterns (see also comments in the specific comment part below):

How does the ABACAS emission inventory filter influence the results? Is the computation of NO_x emissions, which seems to be restricted by a priori emissions, dampening the estimated emissions? How much is the method depending on the emission inventory data?

Response: Thank you for the comments. At first, we filter out the estimated NO_x emissions that beyond the $\pm 50\%$ around the ABACAS emissions, because the day to day, monthly and year to year variation and the uncertainty of the NO_x emissions were all within the $\pm 50\%$ range, the estimation that beyond this range can be seen as anomaly. The extremely high and low estimation is mainly caused by the bias of the satellite NO_2 observation.

Overall we obtained 24 days higher than 1.5 times of the a priori emission and 15 days lower than 50% of the a priori emission, the total of them made up about 10% of the total number of the estimations.

Most of the high emission days are in winter and low emission days are in summer, the filter of them dampened the seasonal variability. Therefore, in the revision of the work, we choose to keep these results for analysis.

Is the bias correction factor of 1.2 valid for all seasons?

Response: We use the factor 1.2 for all the days. The TROPOMI data version we used is the version 2.4.1–2.6.1, and even though the v2.3.1 and after versions have higher retrieval in winter and over polluted area (Van Geffen et al., 2022), they are still found to be lower in polluted area and higher in clean area, when compared to the ground measurements (Keppens and Lambert, 2023). As a consequence, we may still underestimate winter emissions and/or even overestimate summer emissions, thus leading to a higher estimated summer-to-winter emission ratio. We have added this discussion in Page 16, Line 405-412 in the revised manuscript.

How would seasonal patterns change using a daily NO_x/NO₂ ratio instead of a fixed value of 1.26? The loss rate of NO_x (k in Eq 1) is based on the fixed NO_x/NO₂ ratio and does not consider the temperature dependency of the NO₂ and OH reaction.

Response: Thank you for this comment. A fixed NO_x/NO₂ ratio is adapted in this study, while it is found to be varying by 10% from the lowest to the highest month. As a result, we overestimate the summer-to-winter emission ratio. We have clarified this in Page 11, Line 277-279 in the revised manuscript.

I suggest publication if the raised issues are addressed.

Specific comments:

L39: You provide Goldberg et al. (2019) and Zhang et al. (2021) as references for key information about nitrogen oxides. I think it is more appropriate to cite the references which are cited in these references, e.g., Jacob, D. J. Introduction to Atmospheric Chemistry; Princeton University Press, 1999.

Response: Thank you for the comment. We have changed the citation here to: Bassett, M. and Seinfeld, J. H.: Atmospheric equilibrium model of sulfate and nitrate aerosols, Atmospheric Environment, 17, 2237-2252, [https://doi.org/10.1016/0004-6981\(83\)90221-4](https://doi.org/10.1016/0004-6981(83)90221-4), 1983.

Penner, J. E., Atherton, C. S., Dignon, J., Chan, S. J., and Walton, J. J.: Tropospheric nitrogen: A three-dimensional study of sources, distributions, and deposition, Journal of Geophysical Research, 96, 959-990, <https://doi.org/10.1029/90JD02228>, 1991.

Jacob, D.: Introduction to Atmospheric Chemistry, Princeton Univ. Press, 1999.

L51, L53, L56: You already provide several references for the different emission estimation methods and applications; since these are, however, only some examples, I would change these parts to (e.g., reference1, reference2, ...)

Response: The 'e.g.,' has been added to each reference list in Page 2, Line 52-57 in the revised manuscript:

With the improving capabilities of later satellite sensors, more researchers started to estimate NO_x emissions on higher spatial and temporal resolutions but still depended on CTMs (e.g., Ding et al., 2017; Visser et al., 2019; Xing et al., 2022). However, there are barriers to access and employment of CTMs, and there is a substantial computational burden when our target is a single city. Therefore, CTM-independent methods have been developed and applied to estimate NO_x emissions since the early 2010s (e.g., Beirle et al., 2011; De Foy et al., 2014; Kong et al., 2019; 2019; Lorente et al., 2019; Rey-Pommier et al., 2022).

L58: Beirle et al. (2011) have not rotated the NO₂ maps, they divided the OMI data into wind sectors based on the present wind direction of the individual measurement and estimated emissions for the eight defined wind sectors. Rotation was first

introduced by Pommier et al. (2013) and Valin et al. (2013), which you mention also in L62.

Response: The reviewer's comment is taken and this sentence has been rephrased as: 'Beirle et al. (2011) reduced the 2-dimensional NO₂ map surrounding a large point source (such as a megacity or a power plant, factory) to the 1D NO₂ line density by integrating the NO₂ column density across the wind direction.' Please refer to Page 2, Line 58-59 in the revised manuscript.

L60-61: Since Beirle et al. (2011) have not only estimated emissions for cities, I would suggest deleting the word city in "city NO_x emissions" and replacing city with source in "over the city and its decay downwind of the city."

Response: Done.

L65-66: "relatively large study area", I understand that this is meant probably in comparison to the method you use with a diameter of 90 km, still the EMG method is already possible for individual cities and power plants. Maybe you can clarify this to avoid confusion.

Response: We were trying to say that the EMG model asks the cities and power plants to be isolated from other sources and requires a large fit range. We find that it is not necessary to stress it here so we decide to delete this information.

You wrote that the EMG model is limited to calculating emissions only for data averaged over a longer time period. However, this is more related to the quality of the satellite product used, e.g. Goldberg et al. (2019) showed that using the EMG method with TROPOMI NO₂ observations also single overpasses can deliver valuable results, which you are also mentioning in the next paragraph. You can change the sentence to something like "...and with OMI data it is limited to calculating mean NO_x emissions from observations over longer time periods, like some years."

Response: Thank you for the comment. The sentence is rephrased as: 'This method is more frequently used to calculate mean NO_x emissions over longer time periods (like some years) with OMI data (e.g., Lu et al., 2015; Liu et al., 2016).' Please refer to Page 2, Line 64-66 in the revised manuscript.

L69: "Lorente et al. (2019) narrowed down the study area to the domain of one city". Emission estimates for individual cities were already possible with OMI data and the EMG method.

Response: Agree. We have deleted this phrase in the revised manuscript Page 2, Line 69-70: 'Based on a single TROPOMI overpass, Lorente et al. (2019) developed a superposition column model to fit the NO₂ line density for daily NO_x emissions'.

L73: Is Valin et al. (2013) here the right reference, I think the superposition column model is not discussed in Valin et al. (2013), shouldn't it be Lorente et al. (2019) or maybe both?

Response: Valin et al. (2013) pointed out that using the average of NO₂ concentration to calculate NO_x chemical lifetimes and emissions might lead to some bias because this is a nonlinear system. The superposition column model has avoided this bias by calculating NO_x chemical lifetimes and emissions on daily basis. We have reorganized this sentence to avoid misunderstanding. Please refer to Page 3, Line 73-75 in the revised manuscript.

L87: “unprecedented nadir spatial resolution” Since the resolution of TEMPO with 2km x 4.5km is better than TROPOMI’s resolution, I suggest deleting “unprecedented”

Response: Done.

L87: Since you also use data before 6 August 2019, briefly mention the spatial resolution before the change.

Response: Added.

L90-92: You wrote that “The version 2.3.1 includes a different treatment of the surface albedo compared to earlier versions, which led to a 10~15% increase of tropospheric NO₂ columns over polluted scenes.” This is misleading, the surface albedo in the NO₂ window (OMI LER) and for the cloud product (GOME-2 LER) is replaced with the TROPOMI DLER in v2.4, which you also describe in the following sentences. The main changes in v2.3.1 compared to v1.x are the switch to the FRESCO-wide cloud product and a correction of the surface albedo for cloud-free scenes (only for specific scenes with cloud fractions < 0 and > 1). All changes together result in tropospheric NO₂ VCDs that are 10-40% larger than in v1.x.

Response: The reviewer’s comment is taken and we have rewritten the description about TROPOMI version 2.3.1: ‘Compared to the previous versions v1.x, the version 2.3.1 includes a different treatment of the surface albedo to avoid negative and > 1 cloud fractions, and updates the FRESCO-wide cloud retrieval that leads to a lowering cloud pressure. These result in 10% – 40% increase of tropospheric NO₂ columns, depending on the level of pollution and season (Van Geffen et al., 2022).’ Please refer to Page 3-4, Line 99-101 in the revised manuscript.

L105: You use a scale factor of 1.2 to correct the low bias of the TROPOMI data based on the ground-based validation with the Xianghe station. Can you comment on possible differences (bias variations) between different seasons, might this influence your seasonal investigations?

Response: The reviewer's comment is well taken and we have discussed the fixed scale factor of 1.2 on the seasonal pattern investigation of this study. We use the scale factor 1.2 for all seasons and pollution levels to partly correct the potential low bias of the TROPOMI data. However, even though the v2.3.1 and after versions have higher retrieval in winter and over polluted area (Van Geffen et al., 2022), they are still found to be lower in polluted area and higher in clean area, when compared to the ground measurements (Keppens and Lambert, 2023). As a consequence, we may still underestimate winter emissions and/or even overestimate summer emissions, thus leading to a higher estimated summer-to-winter emission ratio. Please refer to Page 16, Line 408-411 in the revised manuscript.

L112: Why have you decided on the 950hPa level?

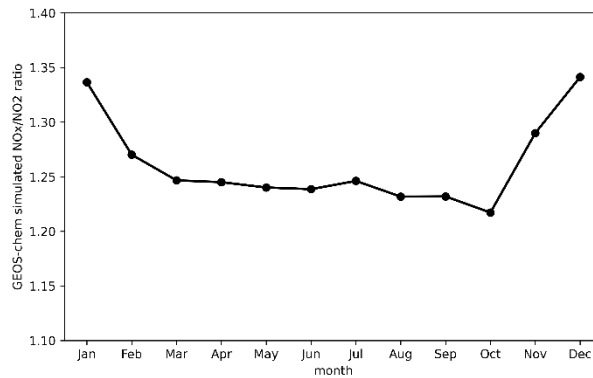
Response: Considering the vertical consistency of wind speeds and directions, we use the 3 levels mean meridional and zonal wind below 950hPa. We have clarified this in Page 4, Line 122-123 in the revised manuscript.

L122: Do you use daily OH concentrations, please clarify.

Response: We did not use the daily OH concentration but the monthly mean value in the initial work. During the revision of the work, we discard the OH concentration and chemical loss rate constant k' from the GEOS-chem model in the fitting. Instead, we use 4 hours for the cold months (October to March) and 2 hours for the warm months (April to September) as the initial guess of the chemical lifetime of NO_2 , which avoids the uncertainty caused by the model output OH concentration and NO_x chemical loss rate and make this method computational efficient.

L124: You decided to use a fixed value of 1.26 for the NO_x/NO_2 ratio. You mention that it varies less than 10% in season. Can you show the seasonal variation of the NO_x/NO_2 ratio over the year? I think this is especially relevant as you are investigating seasonal patterns.

Response: Thank you for the comment. The figure below shows the monthly NO_x/NO_2 ratio simulated by GEOS-Chem, it is highest in winter and lowest in summer, with the difference of 10.2% (highest/lowest -1). We keep the NO_x/NO_2 ratio fixed for all the days, and it leads to high estimation in summer and lower estimation in winter, and thus we overestimate the summer-to-winter emission ratio. We have added this information in the comparison between this work and Lange et al. (2022). Please refer to Page 11, Line 276-280 in the revised manuscript.



L128: Information missing about the ABACAS emission inventory. Is it annual or monthly data, based on which year?

Response: Thank you for the comment. The ABACAS NO_x emission inventory is monthly data for the year 2019. We have added this information in Page 4, Line 126-130 in the revised manuscript.

L130 & Fig. 2: EDGAR provides not only annual but also monthly time series for 2018 NO_x, did you have a look at these? Relevant for L212.

Response: The reviewer's suggestion is well taken and we have used EDGAR v8.1 monthly NO_x emissions from 2018 to 2022 and MEIC v1.4 monthly emissions for 2018-2020 to compare with the result from this study. We found that the TROPOMI estimation is close to the bottom-up emission inventories during cold months, while much lower during warm months. For the three years (2018 to 2020) when MEIC v1.4 data is available, the difference between TROPOMI and MEIC v1.4 is within 35%, and both of them capture the NO_x emission reduction in early 2020 due to COVID-19 lockdown. TROPOMI and EDGAR v8.1 are close to each other (within 30% difference) in 2018 and 2019, but the discrepancy is larger since 2020. EDGAR v8.1 is > 50% higher than TROPOMI from 2020 to 2022. Please refer to Page 8, Line 213-222 and the new Figure 2 in the revised manuscript.

Section 2.4: You write you rotate with the mean wind direction. How is the mean wind calculated, over which area? What is rotated? Is there a difference between the wind errors in Fig 1(a) and (b)? How do you determine the diameter of your circle?

Response: Thank you for the comment. We construct a 15 × 15 grid map centered at the city center with each grid size of 0.05 ° × 0.05 ° (6km × 6km) toward the mean wind direction, with one demission of the grid map along with and the other perpendicular to the wind direction. The mean wind direction is determined by the mean meridional and zonal winds over the study domain. The only difference between Figure 1a and b is that Figure 1a is plotted directly from the TROPOMI observation, and Figure 1b is the rotated result of the study domain. The circle covers the region within the Fourth Ring

Road of Wuhan, and it is the urban area, with a diameter of about 90 km. Please refer to the revised Figure 1 and Line 149-152 in Page 5 of the revised manuscript.

Equation 1: Use points for multiplication signs instead of crosses.

Response: Corrected.

L154-157: k represents the loss rate of NO_x . You use a fixed value for the rate constant k' between NO_2 and OH. How large is the seasonality due to the temperature dependency of this reaction? How large is the seasonal variation of the NO_x/NO_2 ratio over the year (see comment above)? Relevant for section 3.2.2 about seasonal patterns:

Response: We thank the reviewer for this comment.

According to Burkholder et al. (2020), the NO_2 chemical loss rate constant increases with temperature: $k'(T) = 2.8 \times 10^{-11} \times \frac{T}{300}$. In Wuhan's case, the seasonal variation of k' would be ~8% (the temperature record shows that in 2023 the lowest monthly mean noon temperature is in February of 11°C, and the highest is 34°C in July); Based on the GEOS-Chem model simulation, the seasonal variation of NO_x/NO_2 ratio is ~10%.

In our initial work, both the fixed k' and NO_x/NO_2 ratio lead to overestimation of the summer-to-winter emission ratio. In the revision of the work, we discard the OH concentration from GEO-Chem model, instead we give an initial guess for NO_2 lifetime τ (h) of 2 h in warm months (April to September) and 4 h in cold months (October to March) and the NO_x chemical loss rate k (s^{-1}) is derived by $k = \frac{1}{\tau \cdot 3600} \cdot$

$$\frac{[\text{NO}_2]}{[\text{NO}_x]}$$

The fixed NO_x/NO_2 value will dampen the seasonality of NO_x emission estimation, and we have added this discussion in Page 11, Line 276-280 in the revised manuscript.

Equation 3: for $x > x_i$ (?) missing

Response: It is not for $x > x_i$, it is for all the x . We have placed equations 1-3 together to make it clearer. Please refer to Page 6, Line 159-161 in the revised manuscript.

L163: What do you mean with “upend point”, at the upwind end point of the grid/city/circle?

Response: Yes, it is the upwind end point of the city. It is corrected in the revised manuscript.

L166: You write you use the monthly instead of the daily noon time mean OH concentration, can this create an issue in the weekly cycle investigation as the monthly mean is dominated by weekdays?

Response: Thank you for this comment. We don't think that the monthly OH concentration will affect the weekly cycle of NO_x emissions. First, we did not set the OH concentration fixed, it was allowed to vary within a certain range. Second, the observation of NO₂ and O₃ concentrations also reveal no significant weekdays and weekends difference, thus we could say that the OH concentration does not change much from weekdays to weekends. We have added this information in Page 16, Line 403-405 in the revised manuscript.

In the revision of the work, we discard the chemical transport model simulated OH concentration, instead we use an initial guess for the NO₂ chemical lifetime of 2 h for warm months (April to September) and 4 h for cold months (October to March), and we let it vary between 1/4 to 4 times of the initial value during the fitting. Also, we don't think the initial guess of the NO₂ chemical lifetime would affect the weekly cycle of estimated NO_x emissions for the same reasons we stated above.

L178: You write clear skies here and also at other parts of the manuscript, but you mean for cloud radiance fractions < 0.5, which means there is not always clear sky.

Response: Thank you for this comment and we have replaced the 'clear sky' with 'full-NO₂-coverage' in the revised manuscript.

L179: You remove overpasses with inhomogeneous wind fields and days with estimated NO_x emissions beyond 0.5–1.5 times the ABACAS bottom-up emissions. First, I don't understand the reason for the ABACAS filter, and second, can you mention how many days or overpasses are filtered for each of the filters? You only mention how much it is in total. Do you maybe filter high emission days, especially in winter, which results in dampened seasonal patterns?

Response: Thank you very much for the insightful comment. The day to day, monthly and year to year variation and the uncertainty of the NO_x emissions were all within the ±50% range, so the estimated data beyond this range were excluded. The extremely high and low estimation is mainly caused by the bias of the satellite NO₂ observation. Overall we obtained 24 days higher than 1.5 times of the a priori emission and 15 days lower than 50% of the a priori emission, the total of them made up about 10% of the total number of the estimations.

Most of the high emission days are in winter and low emission days are in summer, so we agree with the reviewer that the filter of them will dampen the seasonal variability. Thereby, in the revised work, we choose to keep these results for analysis, and we obtain a lower summer-to-winter emission ratio of 0.77 compared to the 0.87 in the initial work.

L197: Why was the area changed between Zhang et al. (2023) and this study?

Response: We narrow down our study domain from the administrative area of Wuhan in Zhang et al. (2023) to the urban area (within the Fourth Ring Road of Wuhan, ~90 km in diameter) in this study considering that, first, most (~60%) of the NO_x emissions are concentrated in the urban area; second, we use regional mean wind fields and NO_x chemical loss rate, the larger study domain would induce large uncertainty in the result. We have added this information in Section 2.4 in Page 5, line 147-150 in the revised manuscript.

L201: Change November 2020 to January 2020.

Response: Changed.

Figure 2 and text: y-labels say “noontime NO_x emissions”, which is true for TROPOMI, but I think not for EDGAR and ABACAS. Please correct this and also clarify in the text that the emission inventory emissions are not around noon time.

Response: Thank you for the comment. Yes, the bottom-up emission inventories give monthly total emissions, and when we compared them with our estimation, we converted the monthly total to hourly mean through: hourly mean emission =

$$\frac{\text{monthly total}}{(\text{number of days in the month}) * 24}$$

Then, a scale factor of 1.4 (provided by the ABACAS inventory) is applied to the hourly mean emissions to obtain the ‘noontime NO_x emissions’. We have made it clear in Page 8, Line 213-215 in the revised manuscript.

L240: Possible explanations for deviations with Lange et al. (2022): Larger area in Lange et al., this study limited to city center, maybe different behaviors and sources in the different urban/suburban areas? See comment L166, monthly OH concentrations dampening weekly cycle? Fixed NO_x/NO₂ conversion factor in this study compared to daily conversion factors in Lange et al.

Response: Thank you for the comments. We have added the discussion about the influence of the area difference on the deviations between this study and Lange et al. (2022). Please refer to Page 10, Line 257-260 in the revised manuscript.

We find from literature that the observation of NO₂ and O₃ concentrations also reveal no significant weekdays and weekends difference (Wei et al., 2022; Yang et al., 2020), thus we could say that the OH concentration and the fixed NO_x/NO₂ ratio do not change much from weekdays to weekends.

Section 3.2.2 Seasonal pattern: You see a much more dampened seasonal emission pattern than Lange et al. Possible explanations for deviations:

Response: Two possible factors may contribute to the large difference in the summer-to-winter emission ratio between this study and Lange et al. (2022). First is the different treatment to the NO_x-to-NO₂ ratio. We use a fixed NO_x-to-NO₂ ratio of 1.26, while Lange et al. (2022) calculated the ratio from day to day, and it was lower in summer than in winter, leading to a lower NO_x emission estimation in summer. Second is that we use the bottom-up emission inventory to constrain our estimation, the flat seasonality of the bottom-up emissions leads to a higher summer-to-winter ratio of this study.

We have added this information in Page 11, Line 276-280 in the revised manuscript.

See comment above: You use a fixed value for the rate constant k' between NO₂ and OH. How large is the seasonality due to the temperature dependency of this reaction? How large is the seasonal variation of the NO_x/NO₂ ratio over the year? Different areas, sampling issues due to a shorter period in Lange et al., different TROPOMI NO₂ product versions have seasonal bias differences (van Geffen et al. 2022), bias correction used in this study.

Response: Thank you for the comment. According to Burkholder et al. (2020), the NO₂ chemical loss rate constant increases with temperature: $k'(T) = 2.8 \times 10^{-11} \times \frac{T}{300}$. In Wuhan's case, the seasonal variation of k' would be ~8% (the temperature record shows that in 2023 the lowest monthly mean noon temperature is in February of 11°C, and the highest is 34°C in July); Based on the GEOS-Chem model simulation, the seasonal variation of NO_x/NO₂ ratio is ~10%.

The above two factors would lead to higher summer-to-winter emission ratio of this study compared to Lange et al. (2022). We have added this comparison in Page 11, line 276-280 in the revised manuscript.

The different study area and different study period lengths may also affect the different finding in the weekly cycle between the two studies, and we have added this discussion in Page 10, Line 257-260 in the revised manuscript.

This study uses the TROPOMI v2.4.0-2.6.0 and Lange et al. (2022) used earlier v1.x. According to Van Geffen et al. (2022), the v2.3 and later versions produce higher NO₂ columns over polluted area and in winter. However, this does not support the higher summer-to-winter emission ratio in this study.

We use a fixed scale factor 1.2 to correct the potential underestimation of TROPOMI NO₂ in this study, but the v2.4.0-2.6.0 data is still found to be lower in polluted area and higher in clean area, when compared to the ground measurements (Keppens and Lambert, 2023). Thereby our work may have overestimated the summer-to-winter emission ratio (0.77), though it is even higher in the bottom-up emission inventory. This discussion is added to Page 16, Line 408-412 in the revised manuscript.

L263/264: Any ideas for the differences in summer lifetime between Zhang et al. (2023) and this study?

Response: The possible reason for the shorter summer NO_x lifetime estimated in this study than Zhang et al. (2023) can be that we use the bottom-up emission inventory to constrain our fitting, thus resulting in higher summer NO_x emissions and lower chemical lifetime.

In the revision of the work, we use the result from Zhang et al. (2023) as the initial guess of the chemical loss rate of NO₂ during cold and warm months to replace the model simulated OH concentration.

L272: You write that the computation of NO_x emissions is restricted by a priori emissions. Is this restriction dampening your estimated emissions and a possible explanation for differences with Lange et al.? See also L179 (comment above) and L388-390 in your manuscript.

Response: Yes, like the fixed NO_x/NO₂ ratio we use in the superposition column model, the fact that we use the bottom-up emissions to restrict our emission estimation also leads to higher summer-to-winter emission ratio estimated by our study than Lange et al. (2022). We have clarified this in the revised manuscript in Page 11, Line 276-280.

In addition, we use a fixed scale factor 1.2 to correct the potential underestimation of TROPOMI NO₂ in this study, but the v2.4.0-2.6.0 data is still found to be lower in polluted area and higher in clean area, when compared to the ground measurements (Keppens and Lambert, 2023). Thereby our work may have overestimated the summer-to-winter emission ratio (0.77), though it is even higher in the bottom-up emission inventory. This discussion is added to Page 16, Line 405-412 in the revised manuscript.

L298: Any ideas for large deviations between Lonsdale and Sun (2023) and this study, which kind of method is used in Lonsdale and Sun (2023)?

Response: Yes, we find that when we compared NO_x emissions from two different years, we simply computed the mean from all valid days within each year. However, considering the large seasonal difference of estimated NO_x emissions with the satellite data, the estimated annual mean emission will bias low when the valid summer days are more than winter days, and bias high when otherwise. Thereby, in the revision of the work, following what Lonsdale and Sun (2023) did, when we make the comparison between two years, the annual mean emissions are calculated based on the months available for both years. For example, January and February estimations are absent for 2019, June and July are not available for 2020, therefore March to May and August to December monthly values are used to determine the annual mean emissions for the comparison of the two years. As a result, the estimated year-to-year

variations in these two studies are close to each other. Please refer to Page 12, Line 311-320 in the revised manuscript.

Section 3.3 Wind field dependence: You start using the term chemical lifetime in this section, I think until now you only used the term lifetime. How do you calculate lifetimes? Usually, lifetimes estimated by these kinds of models are defined as effective lifetimes (see, e.g., Beirle et al.) as they include effects of deposition, chemical conversion, and wind advection, which you are illustrating here also for your model. I think you should clarify the differences between chemical and effective lifetime in your text.

Response: Thank you for the comment. The ‘lifetime’ mentioned in this study should be the ‘chemical lifetime’ (τ , h). In the superposition model, we assume a first-order chemical loss of NO₂ with the loss rate of k (s⁻¹). the relationship between k and NO₂ chemical lifetime $\tau_{[NO_2]}$ (h) is $k = \frac{1}{\tau_{[NO_2]} * 3600}$. We have clarified this in the revised manuscript, please refer to Page 6, Line 164-165 in the revised manuscript.

Figure 7 and Lines following L 329: I have issues understanding this figure and the text. Explanations for what is visible in the Figure are missing in the caption and also in the text below. Is this one day, averaged for several days? How are the red, yellow, green, pink, and purple curves determined? The text in L332 says that the fitted emissions are basically in line with those from the bottom-up emissions, but in the figure differences are quite large, especially close to the city center.

Response: Thank you for the comment. As we mentioned in Section 2.4, in our fitting approach, we assume that the NO_x emissions within the city is distributed as a Gaussian shape along with the wind direction: $E_i = amp * \exp[-(x_i - cen)^2 / wid]$, the parameter cen is initially set as the location of the city center, and we allow it to shift along the wind.

The original Figure 7 displays the estimated distribution of NO_x emissions along with the wind direction. All the results are divided into four groups of the westerly, easterly, northerly and southerly wind directions. The lines in Figure 7a and b represent the mean emissions of all the days under each wind direction. We find mirrored distribution of the estimated NO_x emissions between the westerly and easterly wind directions and between the southerly and northerly wind directions. The final fitted distribution is much shallower than that of the bottom-up emission inventory.

In the revision of the work, we find that this part of analysis is nothing new or different compared to the work of Lorente et al. (2019), so we chose to remove this part in the revised manuscript.

L357: Lange et al. are not using OH concentrations for their estimates and are not providing an uncertainty for OH.

Response: Thank you for the comment, since we have discarded the model simulate OH in our method, this part has been removed in the revised manuscript.

Technical corrections:

L47: Divide the sentence into two sentences. "...ultraviolet/visible spectrum. Various satellite instruments..."

Response: Done

L50: Misleading, change to something like "Limited by the coarse spatial resolution of the early instruments, researchers ..."

Response: Done.

L55: when the target is an individual city.

Response: Corrected.

L61: Change model to method.

Response: Changed.

L63: To avoid confusion, I would change it to "The EMG model has been first applied to OMI NO₂ data..."

Response: Changed.

L70: Change to: "for daily NO_x emission estimates"

Response: Changed.

L72: Change to: "avoids the bias caused by using the averaged NO₂ columns in the nonlinear system"

Response: Changed.

L81: Change conclusion to concluding

Response: Done.

L86: I suggest replacing "TROPOMI observes NO₂ at 405-465nm of the UV-visible spectral band..." maybe with something like "TROPOMI NO₂ columns are retrieved in the spectral range from 405-465nm..."

Response: Done.

L95/96: Change DLER to “the DLER”.

Response: Done.

L96/97: Check the sentence starting with “The v2.4.0 version.” The version is twice, and in general, it is not good to read.

Response: We rephrased this sentence to “The version 2.4.0 made a complete mission reprocessing from 1 May 2018 to 22 July 2022 and then switched to the offline mode.”

L125: Change to: An initial guess...

Response: Done.

L161: Missing word: and added/combined/... with the contribution from the background

Response: we use ‘combined’ here.

L173: I suggest changing “dilute” with “reduce.”

Response: changed.

L202: an ad hoc bias correction factor

Response: corrected.

L210/211: I would first name EDGAR, then ABACAS, following the logic of Fig. 2.

Response: Corrected. We now use EDGAR v8.1 and MEIC v1.14 for the validation.

L230: Change to “NO_x emission estimates.”

Response: Changed.

L237: Replace “minimum” with “reductions” and add “for Wuhan”

Response: Replaced and added.

Fig 3 caption: “The number of valid measurement days for each day of the week is listed in the plot”

Response: Corrected.

L260: Do you mean dominated instead of determined?

Response: Yes, it should be 'dominated'.

L279: Split into two sentences: "...2024a). We find a similar..."

Response: Done.

L280: Change "dramatic changes" to "strong reductions"

Response: Changed.

L283: You write "We have also found" and give a reference to Zhang et al. (2023) at the end of the sentence. Do you mean "Zhang et al. (2023) have also found"?

Response: It is a mistake, and the reference is deleted here.

Caption Figure 5: Two times NO_x in the first sentence. Change dash to dashed. Split the second sentence into two sentences.

Response: Done.

Beirle, S., Boersma, K. F., Platt, U., Lawrence, M. G., and Wagner, T.: Megacity emissions and lifetimes of nitrogen oxides probed from space, *Science*, 333, 1737-1739, 10.1126/science.1207824, 2011.

Beirle, S., Borger, C., Dorner, S., Li, A., Hu, Z., Liu, F., Wang, Y., and Wagner, T.: Pinpointing nitrogen oxide emissions from space, *Science Advances*, 5, eaax9800, doi: 10.1126/sciadv.aax9800, 2019.

Burkholder, J. B., Sander, S. P., Abbatt, J. P. D., Barker, J. R., Cappa, C., Crouse, J. D., Dibble, T. S., Huie, R. E., Kolb, C. E., Kuryo, M. J., Orkin, V. L., Percival, C. J., Wilmouth, D. M., and Wine, P. H.: Chemical kinetics and photochemical data for use in atmospheric studies Evaluation Number 19, JPL Publication 19-5, 2020.

de Foy, B., Wilkins, J. L., Lu, Z., Streets, D. G., and Duncan, B. N.: Model evaluation of methods for estimating surface emissions and chemical lifetimes from satellite data, *Atmospheric Environment*, 98, 66-77, 10.1016/j.atmosenv.2014.08.051, 2014.

Ding, J., van der A, R. J., Mijling, B., and Levelt, P. F.: Space-based NO_x emission estimates over remote regions improved in DECSO, *Atmospheric Measurement Techniques*, 10, 925-938, 10.5194/amt-10-925-2017, 2017.

Jacob, D.: *Introduction to Atmospheric Chemistry*, Princeton Univ. Press, 1999.

Keppens, A. and Lambert, J.-C.: Quarterly Validation Report of the Copernicus Sentinel-5 Precursor Operational Data Products #21: April 2018 - November 2023, <https://mpc-vdaf.tropomi.eu>, 2023.

Kong, H., Lin, J., Zhang, R., Liu, M., Weng, H., Ni, R., Chen, L., Wang, J., Yan, Y., and Zhang, Q.: High-resolution ($0.05^\circ \times 0.05^\circ$) NO_x emissions in the Yangtze River Delta inferred from OMI, *Atmospheric Chemistry and Physics*, 19, 12835-12856, 10.5194/acp-19-12835-2019, 2019.

Lange, K., Richter, A., and Burrows, J. P.: Variability of nitrogen oxide emission fluxes and lifetimes estimated from Sentinel-5P TROPOMI observations, *Atmospheric Chemistry and Physics*, 22, 2745-2767, doi: 10.5194/acp-22-2745-2022, 2022.

Liu, F., Beirle, S., Zhang, Q., Dörner, S., He, K., and Wagner, T.: NO_x lifetimes and emissions of cities and power plants in polluted background estimated by satellite observations, *Atmospheric Chemistry and Physics*, 16, 5283-5298, 10.5194/acp-16-5283-2016, 2016.

Lonsdale, C. R. and Sun, K.: Nitrogen oxides emissions from selected cities in North America, Europe, and East Asia observed by the TROPospheric Monitoring Instrument (TROPOMI) before and after the COVID-19 pandemic, *Atmospheric Chemistry and Physics*, 23, 8727-8748, 10.5194/acp-23-8727-2023, 2023.

Lorente, A., Boersma, K. F., Eskes, H. J., Veefkind, J. P., van Geffen, J., de Zeeuw, M. B., Denier van der Gon, H. A. C., Beirle, S., and Krol, M. C.: Quantification of nitrogen oxides emissions from build-up of pollution over Paris with TROPOMI, *Scientific Report*, 9, 20033, doi: 10.1038/s41598-019-56428-5, 2019.

Lu, Z., Streets, D. G., de Foy, B., Lamsal, L. N., Duncan, B. N., and Xing, J.: Emissions of nitrogen oxides from US urban areas: estimation from Ozone Monitoring Instrument retrievals for 2005-2014, *Atmospheric Chemistry and Physics*, 15, 10367-10383, doi: 10.5194/acp-15-10367-2015, 2015.

Rey-Pommier, A., Chevallier, F., Ciais, P., Broquet, G., Christoudias, T., Kushta, J., Hauglustaine, D., and Sciare, J.: Quantifying NO_x emissions in Egypt using TROPOMI observations, *Atmospheric Chemistry and Physics*, 22, 11505-11527, 10.5194/acp-22-11505-2022, 2022.

Valin, L. C., Russell, A. R., and Cohen, R. C.: Variations of OH radical in an urban plume inferred from NO_2 column measurements, *Geophysical Research Letters*, 40, 1856-1860, 10.1002/grl.50267, 2013.

van Geffen, J., Eskes, H., Compernelle, S., Pinardi, G., Verhoelst, T., Lambert, J.-C., Sneep, M., ter Linden, M., Ludewig, A., Boersma, K. F., and Veefkind, J. P.: Sentinel-5P TROPOMI NO_2 retrieval: impact of version v2.2 improvements and comparisons with OMI and ground-based data, *Atmospheric Measurement Techniques*, 15, 2037-2060, 10.5194/amt-15-2037-2022, 2022.

Visser, A. J., Boersma, K. F., Ganzeveld, L. N., and Krol, M. C.: European NO_x emissions in WRF-Chem derived from OMI: impacts on summertime surface ozone, *Atmospheric Chemistry and Physics*, 19, 11821-11841, 10.5194/acp-19-11821-2019, 2019.

Wei, J., Liu, S., Li, Z., Liu, C., Qin, K., Liu, X., Pinker, R. T., Dickerson, R. R., Lin, J., Boersma, K. F., Sun, L., Li, R., Xue, W., Cui, Y., Zhang, C., and Wang, J.: Ground-Level NO_2 Surveillance from Space Across China for High Resolution Using Interpretable Spatiotemporally Weighted Artificial Intelligence, *Environmental Science & Technology*, 56, 9988-9998, doi: 10.1021/acs.est.2c03834, 2022.

Xing, J., Li, S. M., Zheng, S., Liu, C., Wang, X., Huang, L., Song, G., He, Y., Wang, S., Sahu, S. K., Zhang, J., Bian, J., Zhu, Y., Liu, T.-Y., and Hao, J.: Rapid Inference of nitrogen oxide emissions based on a top-down method with a physically informed variational autoencoder, *Environmental Science & Technology*, 56, 12, 10.1021/acs.est.1c08337, 2022.

Yang, G., Liu, Y., and Li, X.: Spatiotemporal distribution of ground-level ozone in China at a city level, *Scientific Reports*, 10, 10.1038/s41598-020-64111-3, 2020.

Zhang, Q., Boersma, K. F., Zhao, B., Eskes, H., Chen, C., Zheng, H., and Zhang, X.: Quantifying daily NO_x and CO_2 emissions from Wuhan using satellite observations from TROPOMI and OCO-2, *Atmospheric Chemistry and Physics*, 23, 551-563, doi: 10.5194/acp-23-551-2023, 2023.