



Global flux-based ozone risk assessment for wheat up to 2100 under different climate scenarios

Pierluigi R. Guaita^{1,8}, Riccardo Marzuoli¹, Leiming Zhang², Steven Turnock^{3,4}, Gerbrand Koren⁵, Oliver Wild⁶, Paola Crippa⁷, Giacomo Gerosa¹

¹Dep. Mathematics and Physics, Catholic University of the Sacred Heart, Brescia, Italy
 ²Air Quality Research Division, Science and Technology Branch, Environment and Climate Change Canada, Toronto, Canada
 ³Met Office Hadley Center, Exeter, UK

⁴University of Leeds Met Office Strategic (LUMOS) Research Group, University of Leeds, UK ⁵Copernicus Institute of Sustainable Development, Utrecht University, Utrecht, The Netherlands

⁵Copernicus Institute of Sustainable Development, Utrecht University, Utrecht, The Netherlands
 ⁶Lancaster Environment Centre, Lancaster University, Lancaster, UK
 ⁷Department of Civil and Environmental Engineering and Earth Sciences, University of Notre Dame, Notre Dame, IN, USA
 ⁸Department of Applied Computational Mathematics and Statistics, University of Notre Dame, Notre Dame, IN, USA

Correspondence to: Pierluigi R. Guaita (pierluigirenan.guaita@unicatt.it), Giacomo A. Gerosa (giacomo.gerosa@unicatt.it)

- 15 Abstract. The negative effects of tropospheric ozone (O₃) on vegetation can lead to reduced photosynthesis, accelerated leaf senescence, and other negative outcomes which affect crop yields and biodiversity. This study presents a flux-based assessment of the global impact of O₃ on bread wheat (*Triticum aestivum*) for the 21st century, under various climate scenarios (Shared Socioeconomic Pathways, SSPs). A dual-sink big-leaf dry deposition model is employed to estimate the phytotoxic ozone dose (POD) absorbed by wheat through stomata, integrating data from two Earth System Models (ESMs)
- 20 from the Coupled Model Intercomparison Project 6 (CMIP6). The study explores spatial and temporal variations in O₃ concentrations and the effects of climate variables on stomatal conductance, explaining changes in POD from the present time to the century's end. The results indicate significant regional disparities in O₃ dose for wheat, particularly under weak O₃ precursor emissions control scenarios. The most vulnerable regions include Northern Europe, East China, and the Southern and Eastern edges of the Tibetan Plateau, where the POD increase by the end of the century is expected to be most
- 25 pronounced. Conversely, POD decreases worldwide under stringent pollution emission control scenarios. However, in some regions, changes in POD may be driven more by climate variables and their interaction with O₃, rather than by O₃ concentrations alone. Therefore, this study emphasizes the need for effective emission mitigation policies of both O₃ precursors and greenhouse gases to preserve global food security from O₃ damages.

1. Introduction

30 Tropospheric ozone (O_3) is a highly reactive gas that can harm vegetation and ecosystems. O_3 damages vegetation by altering plant biochemistry and physiology after entering leaf stomata, resulting in reduced photosynthesis, accelerated leaf senescence, and causing the expression of detoxification systems. This in turn reduces canopy carbon gain and export, yield,





and biodiversity, among other negative effects (Fuhrer et al., 2016; Grulke and Heath, 2020; Ramya et al., 2023; Wright et al., 2018). Vegetation impacts have consequences for food security, as highlighted by (Emberson, 2020).

- O₃ impacts on vegetation have been quantified primarily via two methods: the exposure (concentration)-based or the dose (flux)-based approach. Comparisons of the two approaches have been performed over different vegetation types (Anav et al., 2016; Hoshika et al., 2020; Karlsson et al., 2007; Mao et al., 2024; Mills et al., 2018; Paoletti et al., 2019; Pleijel et al., 2022; Simpson et al., 2007; Tai et al., 2021; Tang et al., 2014). The flux-based approach, which estimates the phytotoxic ozone dose (POD), is more biologically meaningful to estimate O₃ damage to vegetation, as it accounts for the actual amount of O₃
- 40 entering the plants through the stomata, i.e. the stomatal flux, rather than on the simple O₃ concentration surrounding the plants (Paoletti and Manning, 2007). Existing schemes for estimating O₃ stomatal flux to vegetation vary in complexity, e.g., as a simple function of temperature and solar radiation (Wesely, 1989), a single leaf Jarvis model (Baldocchi et al., 1987; Jarvis, 1976), a sunlit/shade (two-big-leaf) scheme (Emberson et al., 2000; Zhang et al., 2003), a photosynthesis approach (Ball et al., 1987; Charusombat et al., 2010), and various combinations of the above (Clifton et al., 2023; and references
- 45 therein).

65

Regional- to global-scale O_3 risk assessments for various vegetation types have been extensively conducted over the past two decades, using these schemes or exposure-based approaches, and estimating past and present O_3 damage over different canopies (Anav et al., 2011; Cheesman et al., 2023; Guaita et al., 2023; Mills et al., 2011; Savi et al., 2020; Sharps et al., 2021a). Many studies concluded that high O_3 concentrations can cause significant crop losses and economic damages

- 50 worldwide. For example, for present-day scenarios, Pleijel et al. (2018) indicated ~8% reduction in wheat production in Europe, North America and Asia, Tai et al. (2021) estimated globally aggregated yield losses of up to 7% for four staple crops, and even higher percentage losses have been reported by other studies as reviewed by Emberson (2020). Estimation of future O₃ impacts on vegetation have been mostly based on the exposure-based approach (e.g., Chuwah et al., 2015; Sicard et al., 2017). The few studies that have explicitly made regional flux-based estimates of future O₃ impacts, have
- 55 assumed either present-day meteorology with projected emissions, or present-day emissions with future climate (Klingberg et al., 2014; Simpson et al., 2007; Tang et al., 2014). To our knowledge, a comprehensive, global flux-based analysis of future O₃ impacts across the 21st century, considering both future climate and emission scenarios, is still lacking. The present study was designed within the framework of the Ozone Deposition Focus working group of the Tropospheric

Ozone Assessment Report initiative, Phase II (TOAR-II), to make a global-scale flux-based assessment of future O_3 risks for

- 60 bread wheat (*Triticum aestivum*), which is one of the world's most important staple food crops and often regarded as a reference for species with high O₃-sensitivity (Mills et al., 2011; Sitch et al., 2007). More specifically, the objectives of this research are:
 - 1. To estimate future trends of POD for bread wheat to the end of the 21st century under different climate change scenarios.
 - 2. To identify vulnerable regions to future food security threats due to the negative O_3 effects on wheat.
 - 3. To inform adaptation and mitigation policies to minimize future O_3 risk.





To address these objectives, we employ a dual-sink big-leaf dry deposition model (Guaita et al., 2023), using meteorology and O₃ concentrations inputs from Earth System Models (ESMs) participating in the Coupled Model Intercomparison Project 6 (CMIP6; Eyring et al., 2016). To generate a set of future climate simulations, these ESMs use the forcing datasets associated with the so-called "shared socioeconomic pathways" (SSP; Riahi et al., 2017), which are future pathways combining different trends in social, economic and environmental developments with different assumptions about anthropogenic emission mitigation applied on top of these to meet pre-defined climate targets (radiative forcing).

In the following sections, we describe the adopted methodology, including the analyzed CMIP6 models and the dry deposition model used to calculate the POD. We present the main results – i.e. POD trends and maps – and we quantify the interaction of O₃ risk with climate and O₃ concentrations. Finally, we discuss the confidence in our results, contextualizing them in the broader SSP framework.

2. Methodology

This section describes the CMIP6 models and experiments involved in this study, along with associated information on sowing dates, plant available water and global region definitions used in this study. Then, we summarize the main features of the O dry deposition model for wheet (Gueita et al. 2023) that is ampleued to calculate the POD metric as well as the

80 the O_3 dry deposition model for wheat (Guaita et al., 2023) that is employed to calculate the POD metric, as well as the parameterization of the stomatal conductance module. Finally, we outline the methods adopted to present the model outputs and describe the statistical tools used for the interpretation of the results.

2.1 Selection of CMIP6 models and SSPs

A subset of models from the CMIP6 experiment is selected as input to this work. Specifically, we apply the following criteria to identify the model runs suitable for our study: (i) an online/coupled-chemistry framework (AerChemMIP; Collins et al., 2017) to include the feedback of O₃ on climate and (ii) sub-daily temporal resolution of meteorological variables to enable O₃ flux calculations. According to these criteria, we identify GFDL-ESM4 (Dunne et al., 2020) and UKESM1-0-LL (Sellar et al., 2019) for this study. UKESM1-0-LL and GFDL-ESM4 are both fully-coupled global ESMs, which include a physical atmosphere-ocean model coupled with additional interactive earth system components including; ocean

- ⁹⁰ biogeochemistry, stratosphere-troposphere chemistry and aerosol scheme and terrestrial carbon cycles coupled to interactive vegetation (Table 1). They both have a horizontal grid resolution of between 100 to 140 km in the mid-latitudes, with vertical levels extending to the upper stratosphere. Comprehensive chemistry schemes are included within both ESMs simulating the reactions and transport of the major chemical species (odd-oxygen O_x , nitrogen NO_y , hydrogen $HO_x = OH +$ HO_2 , carbon monoxide CO, methane and short-chain non-methane volatile organic compounds) involved in ozone
- 95 formation. In this way the ESMs are able to simulate the interaction between changes in climate and chemistry in both the historical period and future scenarios.





Output from GFDL-ESM4 and UKESM1-0-LL simulations is publicly available from the ESGF metagrid site (Cinquini et al., 2014). The model variables required for this study are as follows: surface upward sensible heat flux (*hfss*, in ESGF Metagrid notation), near-surface specific humidity (*huss*), surface downwelling/upwelling longwave/shortwave radiation
(*rlds*, *rlus*, *rsds*, *rsus*, respectively), precipitation (*pr*), surface air pressure (*ps*), near-surface wind speed (*sfcWind*), and near-surface air temperature (*tas*). O₃ concentrations at the lowest model level (sfo3) are requested to be provided at an hourly resolution by models participating in AerChemMIP. Additionally, 3D air temperature and specific humidity (*ta* and *hus*, respectively) are used to convert model level height (Pa) of GFDL-ESM4 to geometric height (m above ground), which is a requirement of the dry deposition model used in this study (Guaita et al., 2023). On the other hand, UKESM1-0-LL uses a

105 hybrid-height coordinate system, and therefore such conversion is not required for this model. See Appendix for details.

Table 1: Key features of the GFDL-ESM4 and UKESM1-0-LL runs used in this study: horizontal nominal resolution, vertical
atmosphere resolution, atmosphere and chemistry modules with the corresponding references, variant indicating configurations
(r=realization of the ensemble, i=initialization method, p=physics version, f=forcing), output frequency (hfss=surface upward
sensible heat flux, hus= specific humidity, huss=near-surface specific humidity, rlds=surface downwelling longwave radiation, rlus
=surface upwelling longwave radiation, rsds=surface downwelling shortwave
radiation, pr=precipitation, ps=surface air pressure, sfcWind=near-surface wind speed, sfo3=O3 concentrations at the lowest
model level, ta=air temperature and tas=near-surface air temperature), and citations for documentation and datasets.

	GFDL-ESM4	UKESM1-0-LL
Horizontal resolution		
$(lon \times lat)$		
Land	1°×1.250°	$1.250^{\circ} \times 1.875^{\circ}$
Ocean	0.5° tripolar	1° tripolar
Vertical atmosphere resolution	tion 49 levels to 1Pa (~80km)	85 levels to ~85km
Atmosphere module	GFDL-AM4.1 (Horowitz et al	., HadGEM3-GC3.1 (Kuhlbrodt et
	2020)	al., 2018; Williams et al., 2018)
Chemistry module	GFDL-ATMCHEM4.1 (Horowit	z UKCA-StratTrop (Archibald et
	et al., 2020)	al., 2020; Mulcahy et al., 2018)
Variant	rlilplfl	rlilplf2
Output 1 hour	ps ^a , tas ^a , sfo3	ps, sfo3
frequency		
3 hours	psª, tasª, huss, rlds, rlus, rsds	s, rlds, rsds, rsus, pr, tas, hfss
	rsus, pr	
1 day	sfcWind	huss, rlus, sfcWind





	1 month	hfss, hus, ta	
Documentation		Dunne et al. (2020)	Sellar et al. (2019)
Dataset citation		Horowitz et al. (2018); Krasting	O'Connor (2020); Tang et al.
		et al. (2018)	(2019)

^aps and tas output frequencies are 1 hour for the historical experiment and SSP3-7.0, 3 hours for SSP1-2.6 and SSP5-8.5.

- 115 O₃ risk for wheat cultivation up to 2100 is quantified with respect to a POD baseline value, which is calculated as the average for the period 2000-2014. The POD over the baseline years is derived from the 'historical' experiment of CMIP6 (Eyring et al., 2016, p.201), which is an experiment performed by every CMIP6 model using standardized input data. Future O₃ risk is estimated based on climate and emission scenarios described by the SSPs (Riahi et al., 2017), which were developed for the CMIP6, and results were used in the IPCC Sixth Assessment Report (Intergovernmental Panel On Climate
- 120 Change (IPCC), 2023). Specifically, we focus on the scenarios SSP1-2.6 (Van Vuuren et al., 2017), SSP3-7.0 (Fujimori et al., 2017), and SSP5-8.5 (Kriegler et al., 2017), as they represent contrasting characteristics in terms of future climate change and emission policies for air quality that impact O₃ concentrations from 2015 to 2100. Herein we classify SSP1-2.6 as a low-emissions and low radiative forcing scenario, SSP3-7.0 as a high-emissions and high radiative forcing scenario, while SSP5-8.5 as a high radiative forcing and partial emission control scenario, with controls beginning in the second half of the 21st
- 125 Century (see Table 2). Thus, from an O₃ concentration perspective, SSP5-8.5 can be broadly considered as an intermediate scenario, with end-of-century pronounced climate change and O₃ concentrations akin to those observed in the historical baseline (Turnock et al., 2020). Simulations for these SSPs are available from both GFDL-ESM4 and UKESM1-0-LL include the aforementioned SSPs. Further, the SSP3-7.0pdSST experiment (present-day Sea Surface Temperature; Zanis et al., 2022) is available from UKESM1-0-LL and is included in this study to assess the effect of changing O₃ concentrations
- 130 only in a high-emission/present-day climate scenario.

2.2 Sowing dates and soil hydraulic properties

The O₃ stomatal flux model requires wheat coverage and annual sowing date maps to simulate plant growth. Qiao et al. (2023) produced a global dataset of sowing dates and wheat coverage for *T. aestivum* averaged over specific multi-year periods (1990-2000, 2020-2029, 2040-2049, 2090-2099) with a resolution of $0.5^{\circ} \times 0.5^{\circ}$ for SSP1-2.6 and SSP3-7.0. In this

135 work, these dates are linearly interpolated to obtain yearly maps of sowing dates for the whole globe. The results obtained for SSP3-7.0 are also used for SSP5-8.5, as the two scenarios have comparable radiative forcing. This dataset is upscaled to the resolution of the CMIP6 models.

Contextually, O_3 stomatal flux model requires also wilting point and field capacity soil maps, to simulate plant available water during the growing season. Wilting point is defined as the minimum amount of water in the soil that plants require not

140 to wilt. On the contrary, field capacity is the maximum amount of water contained in the soil after it has been thoroughly saturated and allowed to drain. Both of them are expressed as m³ of water to m³ of soil and they vary according to soil





texture. Zhang et al. (2018) developed maps of wilting point and field capacity with a resolution of 1×1 km². The value of field capacity at each CMIP grid node is obtained by averaging all the 1×1 km² nodes provided by Zhang et al. (2018) that belong to the nodes that Qiao et al. (2023) indicated as wheat area. The same is done for the wilting point maps.

145 **2.3 Dry deposition model for wheat**

In this work we apply the O_3 dry deposition model developed by Guaita et al. (2023) to compute POD₆ values over the entire globe. This model is a dual-sink big-leaf model that simulates crop geometry, phenology, plant available water in the soil, light penetration within the canopy, and calculates stomatal conductance (g_s) and O_3 uptake by plants, following the DO3SE paradigm originally developed by Emberson et al. (2000). The O_3 stomatal flux model receives as input meteorological and

150 chemical data from the CMIP6 models under the different SSPs. The model code is developed in MATLAB (version R2023a).

 O_3 deposition from the lowest model level to the vegetated surface is calculated with a resistive network of three resistances in series: an atmospheric resistance R_a representing turbulent mixing from the lowest model level to near-surface height, a quasi-laminar sublayer resistance R_b representing diffusive transport to the surface, and a surface resistance R_c . The latter

- 155 consists of three resistances in parallel: a cuticular resistance R_{cut} , a stomatal resistance R_{stom} and a ground resistance composed of an intra-canopy resistance R_{inc} and a soil resistance R_{soil} , in series. R_a is calculated following the assumptions of the Monin-Obukhov similarity theory, R_b is calculated with the formulation of Hicks et al. (1989), R_{cut} and R_{soil} are assumed as constant values of 1500 and 200 s m⁻¹ respectively, and R_{inc} was calculated following the formulation of Erisman et al. (1994).
- 160 The R_{stom} was calculated as the inverse of g_s by applying the empirical Jarvis-Stewart approach (Jarvis, 1976; Stewart, 1988), which is based on the limitation effect of the main environmental variables on the maximum stomatal conductance to water (g_{max}) of wheat:

 $g_s = g_{max} \cdot \min\{f_{phen}, f_{O3}\} \cdot f_{light} \cdot \max\{f_{min}, f_{temp} \cdot f_{VPD} \cdot f_{soil}\}$ (1)

where the *f* functions (all ranging between 0 and 1) describe the limiting effect on g_{max} due to environmental variables such as light (f_{light}), temperature (f_{temp}), air water Vapour Pressure Deficit (f_{VPD}) and plant available water in the soil (f_{soil}), and to phenological growth (f_{phen}) and O₃ dose received by the plants (f_{O3}). The f_{min} term represents a constant value of 0.01 indicating the minimum g_s expressed relative to g_{max} during daylight hours. For this study an f_{clim} term is defined, which represents the product ($f_{temp} \cdot f_{VPD} \cdot f_{soil}$) and combines the effect of the main climatic factors on g_s .

The O_3 dry deposition model considers both sunlit and shaded leaves and varies the crop geometry according to the growth and phenology of wheat plants. Further details on the model can be found in Guaita et al. (2023).

For the simulations of this work, different parameterizations of g_s for *T. aestivum* are adopted depending on the different biogeographical regions. Contextually to sowing dates, Qiao et al., (2023) categorized wheat-growing areas into four climatic zones (temperate, cold, warm, and monsoon) based on the intensity and seasonality of temperature and precipitation.





According to this classification, we use the following wheat parameterizations of the g_s model, depending on the climatic

- 175 zone: spring wheat (Grünhage et al., 2012) in the cold region, winter wheat (Grünhage et al., 2012) in the temperate region, and mediterranean wheat (González-Fernández et al., 2013) in the warm and monsoon regions. The only difference between the spring wheat and winter wheat parameterization is in the f_{phen} limiting function, which considers the different temperature sums used to calculate the mid-anthesis date (the middle of the flowering period of a plant) and the corresponding flux accumulation period.
- 180 The gs model for the Mediterranean region was originally parameterized based on field gs measurements collected in Spain (González-Fernández et al., 2013, currently adopted in the Mapping Manual of the Convention of Long Range Transport of Air Pollution of the UN Economic Convention for Europe, LRTAP Convention, 2017). In this study, however, it is also used for warm and monsoon climatic zones. Although parameterizations exist for specific regions within these zones, these were found to be incomplete for the application in our study (Feng et al., 2012) or very similar to the one by González-Fernández 185 et al. (2013) (Yadav et al., 2021).
 - Once g_s is obtained, the seasonal O_3 stomatal dose received by wheat plants can be calculated as the phytotoxic O_3 dose above the Y threshold accounting for plants detoxification capacity (POD_Y; mmol $O_3 \text{ m}^{-2}$ of projected leaf area) by integrating the hourly O₃ stomatal flux ($F_{sO3,i}$; nmol O₃ m⁻² s⁻¹) over the flux accumulation period (i = 0, ..., t, i-th hour of the accumulation period):

190
$$POD_Y = \sum_{i=0}^{t} \max\{F_{sO3,i} - Y, 0\} \cdot 3600 \cdot 10^{-6}$$
 (2)

where the $F_{sO3,i}$ is obtained from the O₃ concentration at top canopy height (see Eq. 51, 52 in the Appendix of Guaita et al., 2023b). For *T. aestivum* a detoxifying threshold of 6 nmol $O_3 \text{ m}^{-2} \text{ s}^{-1}$ (POD₆) is recommended by the (LRTAP Convention, 2017), while the accumulation period for POD_6 calculation runs from the beginning of the anthesis to the plant's maturity, which depends on the phenology limiting function (f_{phen}). In our case, plant's maturity is set at 1775, 2325 and 2400 °C day

195 for spring wheat, winter wheat and Mediterranean wheat, respectively (González-Fernández et al., 2013; Grünhage et al., 2012).

A well-established dose-response relationship for *T. aestivum* based on POD₆ and relative grain yield (Grünhage et al., 2012; LRTAP Convention, 2017; Pleijel et al., 2007) can be used to calculate the relative yield loss (RYL; %): $RYL = 3.85 \cdot POD_6$ (3)

This relationship predicts a 5% decrease of grain yield for each increment of O₃ dose of 1.3 mmol m⁻² (PLA, plant leaf area), 200 and this is defined as the 'Critical Level' (CL).

Compared to the original model described in Guaita et al. (2023), in this exercise (i) pressure and sensible heat flux are provided as CMIP6 input, (ii) vapor pressure is calculated from specific humidity by $e [kPa] = p \cdot q/0.622$, with p air pressure [kPa], and q the specific humidity [kg H₂O (kg air)⁻¹], and (iii) net radiation is calculated by summing the radiation components provided by the CMIP6 models.

205





2.4 Assumptions and description of dry deposition model runs

The O_3 stomatal flux model uses output from the ESMs at their native spatial resolution and runs at an hourly timestep. However, the CMIP6 models a have different output frequency (Table 1) and, to make full use of the hourly O_3 concentrations provided by the CMIP6 models, other variables at coarser resolutions are interpolated in time by nearest

210 neighbor interpolation, which, compared to linear interpolation, avoids excessive underestimation of the diurnal peaks. Any output produced in this study has the same spatial resolution as the input data from the CMIP6 model used for the simulations.

Global annual POD_6 maps from 2000 to 2099 are produced for each SSP considered. The maps produced are masked with the land-use maps of Qiao et al. (2023). Every node where wheat does not reach maturity before the next prescribed sowing

215 date (according to González-Fernández et al., 2013 and Grünhage et al., 2012) is excluded from the POD₆ map for that year. The model assumes that soil water content is at field capacity at the sowing date every year. Then, to evaluate the effect of soil water on O₃ uptake, two contrasting runs are presented: (i) a run that dynamically simulates soil water availability to plants, with precipitation as the only source of water (henceforth rain-fed run), and (ii) a run that removes any limitation due to soil water content, by assuming plant available water at field capacity, i.e. with $f_{soil} = 1$ at every timestep (henceforth FC run).

2.5 ANOVA

225

A two-way ANalysis Of VAriance (ANOVA; Table 2; von Storch & Zwiers, 1999) is applied to the POD6 values computed at the end of the century in order to assess the areas of the globe where the changes of POD_6 with respect to the baseline period are statistically significant. We consider radiative forcing and control emission policies as the two factors, each one applied at two levels, i.e. low/high for radiative forcing (RF), and weak/strong for emission policy (EP). These levels are

- determined by the different SSPs, as described by Table 2. ANOVA was performed only on the POD₆ output produced through the UKESM1-0-LL, since the GFDL-ESM4 is incomplete, i.e. does not include the SSP3-7.0pdSST experiments. ANOVA is performed independently in every grid node of the map with POD₆>0 and a Bonferroni correction is applied to the p-values to account for multiple tests and for multiple
- 230 locations (#tests × #(nodes with POD₆ > 0) = $(8 \cdot 7/2)$ × 2751 = 77028). Furthermore, this analysis allows to calculate the fraction of explained total variance by each factor, defined as R^2 , according to the definition by von Storch and Zwiers, 1999. The ANOVA analysis was performed with MATLAB.

Table 2: Configuration for ANOVA by SSPs, classified by emission control policies (EP) and radiative forcing (RF).

 Emission control policies (EP)	
Strong	Weak





Radiative	Low	SSP1-2.6	SSP3-7.0pdSST
forcing	High	SSP5-8.5	SSP3-7.0
(RF)			

235 **3. Results**

3.1 Spatio-temporal patterns of O₃ and climate variables and their impacts on stomatal conductance

Figure 1 shows the O_3 mean concentration during the *baseline* years (2000-2014), and its change (ΔO_3) at 2100. O_3 concentrations are expressed at the wheat canopy height (see Eq. 51 in the Appendix of Guaita et al., 2023) and averaged over the accumulation period of the stomatal flux. During the baseline years, both GFDL-ESM4 and UKESM1-0-LL show similar global O_3 mean concentration and standard deviation (Table 3). The Northern Hemisphere presents higher O_3 concentrations than the Southern Hemisphere, with the Middle East and Asia the regions with the highest O_3 around the globe (Figure 1a,b). Nevertheless, UKESM1-0-LL shows stronger latitudinal and elevation gradients compared to GFDL-ESM4, and generally reports higher concentrations than the latter for latitudes <45°, with the only exception of the Eastern Indian peninsula. Turnock et al. (2020) showed that all CMIP6 models tend to overestimate climatological surface ozone

- 245 concentrations in the 2005 to 2014 when compared to observations. The simulated difference of ozone to observations and also between models could be due to a number of different reasons including uncertainties in the input datasets (emission inventories), output processes (deposition) or vertical mixing (Wild et al., 2020). UKESM1-0-LL simulates the strongest seasonal cycle in Northern Hemisphere surface ozone out of all of the CMIP6 models, potentially due to excessive NO_x titration of ozone in this model (Turnock et al., 2020).
- 250 Despite these differences, the two models show similar changes across different SSPs. SSP1-2.6 which assumes the implementation of strong policies to reduce air pollution indicates a decrease in O₃ concentrations by the end of the century for both models (Figure 1c,d). However, the global mean O₃ reduction is 3.1 ppb greater in GFDL-ESM4 than in UKESM1-0-LL. On the contrary, under scenario SSP3-7.0, which is characterized by weaker air pollution control policies, O₃ concentrations generally increase by 2100, even though some parts of Europe and eastern U.S. show similar or lower
- averages than the baseline (Figure 1e,f). SSP5-8.5 shows the same spatial patterns as SSP3-7.0, but with lower O_3 values at the end of the century, which is consistent with the implementation of strong air quality control policies from 2050 onward, despite the strong climate change signal (Figure 1g,h).

The limiting function f_{clim} combines the effect of the main climatic parameters (air temperature, VPD and plant available water) on stomatal conductance. $f_{clim} \approx 1$ indicates very little limitation to the g_s, while descreases in f_{clim} towards zero

260 denote a progressive stomatal closure associated with a decrease in O₃ uptake. During the baseline years, the f_{clim} over the accumulation period is greater than 0.5 over large areas of Eastern North America, Europe, and East Asia, while widely limiting conditions ($f_{clim} < 0.2$) are mostly in arid regions of the globe (Figure 2a,b). Figure 2c-h shows the changes in f_{clim}





265

stronger limitation of g_s and thus, potentially, lower POD₆ and O₃ damage. On the contrary, positive Δf_{clim} values (in red) correspond to a weaker limitation of gs and higher POD₆ and O₃ damage. SSP3-7.0 and SSP5-8.5 indicate overall similar magnitude of Δf_{clim} across the globe at the end of the century, while SSP1-2.6 shows weaker changes, as expected from the weaker RF of this scenario. However, some regional differences across models can be identified. For instance, a higher Δf_{clim} in GFDL-ESM4 is observed over the U.S. and Europe, and across the whole Asia in UKESM1-0-LL. The only consistent pattern in 2100 changes across both models and all SSPs is the large positive shift at latitudes greater than 45°, 270 which indicates higher stomatal conductance (and better growing conditions) for wheat at higher latitude. This is due to increased f_{temp} values, i.e. temperatures closer to the optimum for stomatal conductance (Figure 3a-d). The function f_{VPD} (Figure 3e-h) has non-limiting conditions ($f_{VPD} > 0.9$) over most of the globe, with the only exception in desert regions, indicating that f_{clim} is only marginally affected by this component. This function becomes generally more limiting at the end of the century, especially for UKESM1-0-LL, because of a general decrease in relative humidity, which is a feature 275 frequently observed in future climate simulations (e.g., Fang et al., 2022). The spatial patterns of f_{clim} appear to be largely affected by f_{soil} , because where f_{soil} is limiting stomatal conductance, f_{clim} becomes also limiting (Figure 3i-l). f_{soil} grows considerably at the end of the century over arid regions, especially for the SSP3-7.0 and the SSP5-8.5. Maps for SSP1-2.6 and SSP5-8.5 may be found in the appendix (Figure A1-3).

at 2100 compared to the baseline (Δf_{clim}). From an O₃ risk perspective, negative Δf_{clim} values (in blue) correspond to a

280 Table 3: Global means (±SD) for O₃ concentrations and Jarvis functions, during the baseline period and at 2100 over wheatgrowing regions, for two climate models (GFDL-ESM4 and UKESM1-0-LL) under three different SSP scenarios (SSP1-2.6, SSP3-7.0 and SSP5-8.5).

Variable	Model	Baseline ^a	2100ª			
			SSP1-2.6	SSP3-7.0	SSP5-8.5	
	GFDL-ESM4	$35.0\pm~9.1$	$23.7\pm~7.6$	38.1 ± 12.0	36.2 ± 10.9	
O ₃ [ppb]	UKESM1-0- LL	36.3 ± 10.4	28.1 ± 8.3	39.2 ± 10.7	37.3 ± 9.4	
<i>f_{clim}</i> [0,1]	GFDL-ESM4	0.27 ± 0.24	0.30 ± 0.24	0.35 ± 0.26	0.36 ± 0.26	
	UKESM1-0- LL	0.30 ± 0.23	0.32 ± 0.24	0.37 ± 0.23	0.34 ± 0.24	
	GFDL-ESM4	0.58 ± 0.27	0.62 ± 0.25	0.71 ± 0.19	0.71 ± 0.17	
f _{temp} [0,1]	UKESM1-0- LL	0.58 ± 0.21	0.68 ± 0.18	0.68 ± 0.17	0.71 ± 0.14	
f_{VPD} [0,1]	GFDL-ESM4	0.89 ± 0.16	0.89 ± 0.16	0.86 ± 0.18	0.85 ± 0.19	



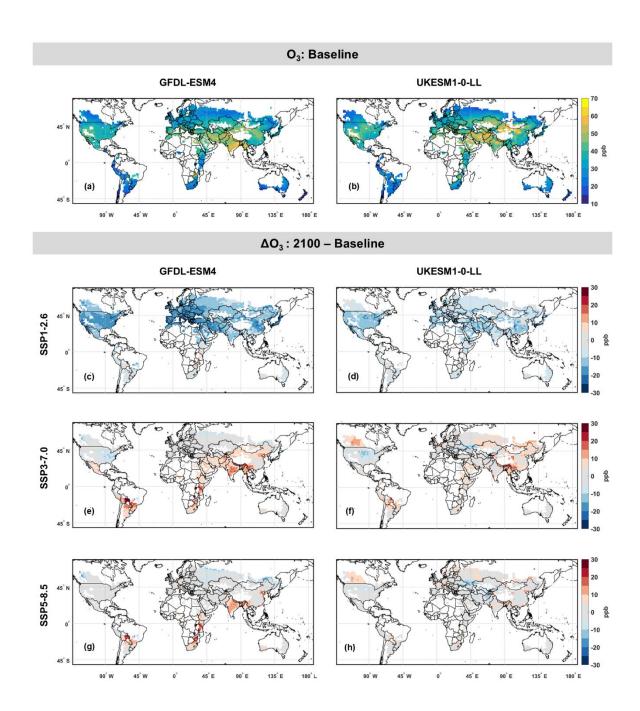


	UKESM1-0- LL	0.89 ± 0.14	0.84 ± 0.17	0.81 ± 0.20	0.76 ± 0.22
	GFDL-ESM4	0.60 ± 0.37	0.60 ± 0.36	0.59 ± 0.35	0.60 ± 0.35
<i>f_{soil}</i> [0,1]	UKESM1-0- LL	0.62 ± 0.36	0.57 ± 0.36	0.69 ± 0.34	0.64 ± 0.36

^aBaseline indicates the mean between 2000 and 2014, and 2100 indicates the mean between 2090-2099.







285 Figure 1: Ozone mean concentrations over the baseline period (a,b), and ΔO₃ at 2100 with respect to the baseline across the different SSPs for GFDL-ESM4 (c,e,g) and UKESM1-0-LL (d,f,h). O₃ values are expressed at canopy height and over the POD₆ accumulation period. Baseline indicates the mean between 2000 and 2014, and 2100 indicates the mean between 2090-2099.





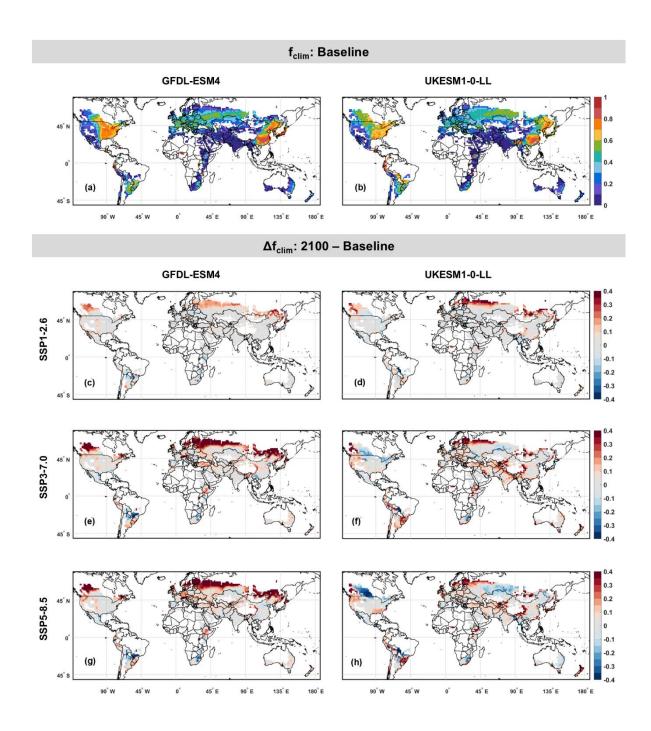


Figure 2: f_{clim} mean values over the baseline period (a,b), and Δf_{clim} at 2100 with respect to the baseline across the different SSPs for GFDL-ESM4 (c,e,g) and UKESM1-0-LL (d,f,h). $f_{clim} = f_{temp} \cdot f_{VPD} \cdot f_{soil}$ summarizes the limitations to stomatal conductance due to temperature, VPD and plant available water, with values ranging from 0 to 1 depending on whether they are limiting or not. Positive or negative Δf_{clim} correspond to higher and lower O₃ risk, respectively (red and blue, in c-h). Baseline indicates the mean between 2000 and 2014, and 2100 indicates the mean between 2090-2099.





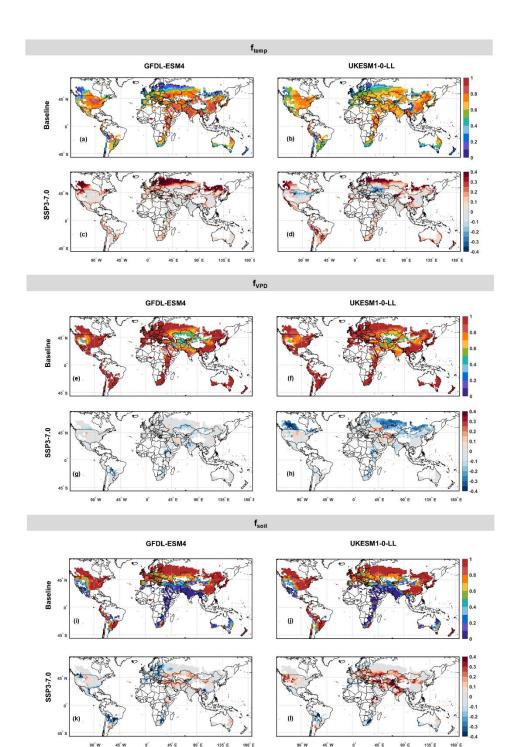


Figure 3: Jarvis limiting functions for g_s over the baseline period and the changes at 2100 with respect to the baseline under SSP3-7.0 for f_{temp} (a-d), f_{VPD} (e-h) and f_{soil} (i-l), for GFDL-ESM4 (a,c,e,g,i,k) and UKESM1-0-LL (b,d,f,h,j,l). Baseline indicates the mean between 2000 and 2014, and 2100 indicates the mean between 2090-2099.





3.2 POD₆ trends by region

- Figure 4 displays yearly POD₆ values averaged over different geographical regions under different SSPs. These values are 300 calculated individually for each model (thin lines), and for each region (different panels). In this analysis, wheat areas with potential O₃ risk are identified considering only the nodes where POD₆ was greater than 0.65 mmol m⁻² – which is half the CL for grain yield loss (LRTAP Convention, 2017) – for at least one year of the century. Then, the Δ POD₆ is calculated as the difference of each year and the baseline. Finally, the Δ POD₆ timelines of both CMIP6 models are averaged to get the regional mean trend (thick line). Table 4 lists the mean Δ POD₆ at 2050 (2045-2054) and at 2100 (2090-2099) region by
- region. ΔPOD₆ follows a decreasing trend under SSP1-2.6 in every region, which is the result of decreasing POD₆ down to near-zero values well before 2100. Trends of ΔPOD₆ generally do not decrease under SSP3-7.0, with 9 out of 12 regions displaying increasing POD₆ values at the end of the century. Under SSP5-8.5, ΔPOD₆ timelines show a characteristic downward concave shape (raising and then falling), usually reaching their maxima between 2050 and 2080, before returning closer to the baseline values (ΔPOD₆≈0). This reflects the air quality improvement policies explicitly assumed by this
 scenario starting from the mid-of the century (Kriegler et al., 2017).
- South Asia shows the greatest POD₆ increase across the three scenarios ($\pm 0.51 \text{ mmol m}^{-2}$). This area reports the maximum increase at mid-century both under SSP3-7.0 and SSP5-8.5, and the highest absolute increase at the end of the century under SSP3-7.0. Furthermore, it is one of the two regions with increased POD₆ at 2100 for SSP5-8.5, even though not statistically significant (Table 4). Sub-Saharan Africa is the region with the second highest POD₆ increase across SSPs ($\pm 0.43 \text{ mmol m}^{-2}$)
- 315 PLA). The final POD₆ at the end of the century is 2.9 times the baseline value under SSP3-7.0, the largest relative difference observed across regions and scenarios. South-East Asia and Central America also emerge as regions with distinctive positive changes both at mid-century for SSP5-8.5 (+0.58/+0.61), and at the end of the century for SSP3-7.0 (+0.66/+0.45). Furthermore, South-East Asia is the region with the highest POD₆ interannual standard deviation across SSPs (mean SD: 0.27 mmol m⁻²; data not shown), most likely reflecting the largest interannual variability of O₃ concentrations and *f_{clim}*
- among all regions (mean SD: 1.74 ppb and 0.06, respectively). Noticeably, East Asia is the region with the highest baseline POD₆ value (1.49 ± 0.84 mmol m⁻²). It experiences an increment of +0.41 and +0.52 mmol m⁻² at mid-century under SSP3-7.0 and SSP5-8.5 respectively, and a significant decrease of -0.45 mol m⁻² at 2100 under SSP5-8.5. Central Asia and Middle East also display a higher POD₆ at 2050 and 2100 with respect to the baseline, especially under SSP3-7.0, although only UKESM1-0-LL shows nodes with POD₆>0.65 mmol m⁻² for at least one year of the century. Arguably, South America and
- 325 North Africa are subject to fairly similar POD₆ changes compared to the baseline under SSP3-7.0 and SSP5-8.5 (Table 4), even though only under SSP1-2.6 South America displays statistically significant changes. North America is the only region with significantly negative POD₆ trends in all the considered scenarios, while Russia-Belarus-Ukraine has negative ΔPOD_6 under any scenario, even though the change is not significant under SSP3-7.0. On the other hand, it should be noticed that Europe has non-significant negative changes under SSP3-7.0 and SSP5-8.5, which is the result of opposing trends between
- 330 Southern and Northern Europe (see Figure 5).





335

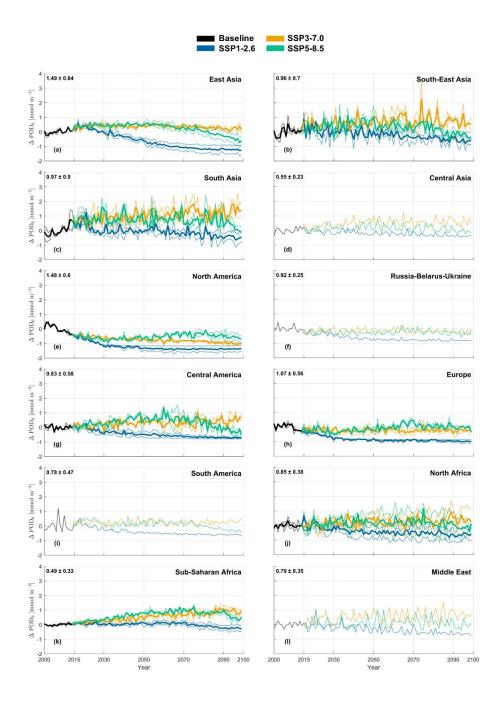


Figure 4: ΔPOD_6 across the century with respect to the baseline average over different SSPs, divided by region; only nodes with POD₆>0.65 mmol m⁻² in at least one year of the timeline are considered. Bold lines indicate the mean between the results from GFDL-ESM4 and UKESM1-0-LL, thin lines indicate the results from the individual models. In some cases (d,f,i,l), only UKESM1-0-LL had nodes with POD₆>0.65 mmol m⁻². The number in the upper-left corner of each panel is the mean POD₆ over the baseline \pm spatial SD (mmol m⁻²). These results are from the rain-fed simulation.





340

Table 4: Means ΔPOD₆ by region and SSP (mmol m⁻² PLA) at 2050 (2045-2054) and 2100 (2090-2099) with respect to the *baseline* (2000-2014). The POD₆ regional means are calculated only over the nodes with POD₆>0.65 mmol m⁻² PLA during at least one year of the century. These results are from the rain-fed simulation. Starred numbers indicate significant (p<0.05) differences.

	Baseline POD ₆	2050 ΔPOD ₆ [mmol m ⁻²]		2100 ΔPOD ₆ [mmol m ⁻²]			
Region ^b	[mmol m ⁻²]						
	historical	SSP1	SSP3	SSP5	SSP1	SSP3	SSP5
East Asia	1.49	-0.64*	+0.41*	+0.52*	-1.22*	+0.21	-0.45*
South-East Asia	0.96	-0.16*	+0.61*	+0.58*	-0.59*	+0.66*	-0.20*
South Asia	0.97	-0.01	+1.12*	+0.74*	-0.46*	+1.36*	+0.28
Central Asia ^a	0.55	-0.25*	+0.65*	+0.27*	-0.38*	+0.58*	-0.06
North America	1.48	-1.30*	-0.75*	-0.63*	-1.37*	-0.98*	-0.58*
Central America	0.83	-0.50*	+0.35*	+0.61*	-0.71*	+0.45*	-0.24*
South America ^a	0.78	-0.47*	+0.29	+0.22	-0.60*	+0.31	-0.33
Russia-Belarus-Ukraine ^a	0.92	-0.71*	-0.15	-0.27*	-0.81*	-0.22	-0.30*
Europe	1.07	-0.89*	-0.20	-0.13	-0.97*	-0.25	-0.05
North Africa	0.85	-0.49*	+0.35*	+0.19*	-0.64*	+0.39	-0.11*
Sub-Saharan Africa	0.49	+0.06*	+0.35*	+0.78*	-0.24*	+0.93*	+0.57*
Middle East ^a	0.79	-0.33*	+0.63*	+0.10	-0.64*	+0.68*	-0.05

^a Only UKESM1-0-LL has nodes with POD₆>0.65 mmol m⁻² PLA during at least one year of the century.

^b Region definitions are based on those established by the Hemispheric Transport of Air Pollutants (HTAP2; Huang et al., 2017).

3.3 Global POD₆ estimates and intermodel comparison

Figure 5 shows POD₆ global maps obtained with the O₃ stomatal flux model for the baseline period and at the end of the 345 century. During the baseline years, POD₆ exceeded 0.65 mmol m⁻² over the 8.6% and 34.1% of the global wheat area using GFDL-ESM4 and UKESM1-0-LL, respectively (Figure 5a,b). Despite this difference, which is most likely due to the intermodel difference in O₃ concentrations, both models identified Eastern North America, Southern Europe and East Asia as hotspots for O₃ risk at the beginning of the century. Furthermore, UKESM1-0-LL pinpoints some areas of Sub-Saharan Africa and South America potentially at risk for O₃.

Under SSP1-2.6, O_3 risk decreases across the whole globe at the end of the century (Figure 5c,d), likely due to declining O_3 concentrations (Figure 1c,d), with POD₆ exceeding the 0.65 mmol m⁻² threshold over the 1.8% and the 13.9% of the global wheat area using GFDL-ESM4 and UKESM1-0-LL. Under SSP3-7.0 and SSP5-8.5 (Figure 5e-i), changes are not as straightforward, and they need to be considered region by region, and model by model. The three regions that the baseline





analysis identified as hotspots follow different paths and dynamics in the remaining part of the century. Eastern North
America is the only region with an evident decreasing O₃ risk under any scenario, which is due to decreasing O₃ concentrations (Figure 1) in a fairly constant *f_{clim}* (Figure 2). For Eastern China, POD₆ either lowers or remains fairly similar to the baseline at the end of the century under both scenarios. The decreasing POD₆ is especially evident for SSP5-8.5, comparatively to SSP3-7.0, revealing that the policies controlling O₃ precursors emissions (Fujimori et al., 2017) might have a large impact over this region. In Europe, O₃ risk hotspots shift from Southern to Northern Europe, following less
limiting temperatures for stomatal O₃ uptake at high latitudes (Figure 3c,d). Both GFDL-ESM4 and UKESM1-0-LL recognize an increased risk in the southern and eastern edges of the Tibetan plateau (region from Kashmir to Sichuan) by the end of the century. Within this region, Nepal, Bhutan, and East India reach the POD₆ peak value of 8.7 mmol m⁻² PLA for SSP5-8.5 at 2100. Furthermore, UKESM1-0-LL identifies areas with increased O₃ risk also in Sub-Saharan Africa, Central and South America, especially under scenario SSP3-7.0.

365 Using the Eq.

375

(3, the POD_6 in

Figure 5 can be converted to relative yield losses (RYL). Based on the POD₆ values calculated using UKESM1-0-LL, the highest RYL values (95th percentile) over Europe are estimated about 8.5-9.1%, both at the beginning and at the end of the century under SSP3-7.0 and SSP5-8.5. On the other hand, under SSP1-2.6, the 95th percentile of RYL is substantially lower (1.6%).

370 Over the entire East Asia, the highest values of RYL are around 16% at the present time, with no relevant decrease for the end of the century under SSP3-7.0 (14.9%). Under SSP5-8.5 and SSP1-2.6 these peak values decrease to 9.8% and 3.9%, respectively, by the end of the century.

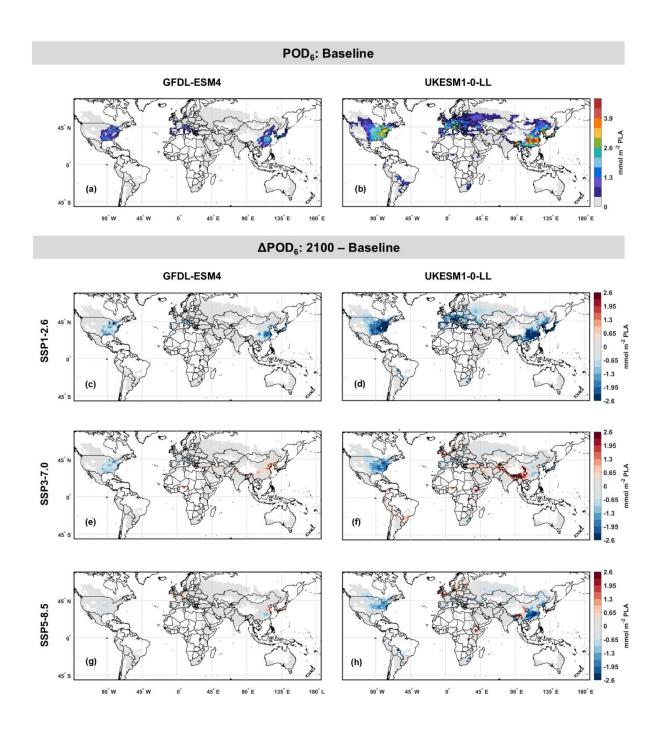
Across the southern and eastern edges of the Tibetan plateau, the present-day maximum RYL value is 25.4%, reached in the Sichuan province (China). Over this region, only under SSP1-2.6 the maximum RYL value falls to 18.7% by end of the century, while it increases to 28.6% and to 31.3% under SSP5-8.5 and SSP3-7.0, respectively.

From the present day to the end of century, the maximum RYL in South America increase from 9.8% to 10.2% under SSP3-7.0, in Sub-Saharan Africa from 8.5% to 12.3% and 15.4% under SSP3-7.0 and SSP5-8.5, and in Central America from 11.2% to 13.8% under SSP3-7.0.

Figure 6 shows the agreement between the two models GFDL-ESM4 and UKESM1-0-LL using the Pearson's correlation coefficient (ρ), calculated in every node for each SSPs. This indicator shows that the two models rarely disagree on the POD₆ changes across SSPs, as highlighted by the very few areas with negative correlation coefficients. However, while the correlation for SSP1-2.6 is relatively strong everywhere, for the other scenarios the correlation is high only in East China and in a few spot areas (e.g. Ethiopia $\rho > 0.75$). However, lower correlation coefficients ($0.15 < \rho < 0.45$) are found for Eastern US under SSP3-7.0, and for Europe and South America under SSP5-8.5.





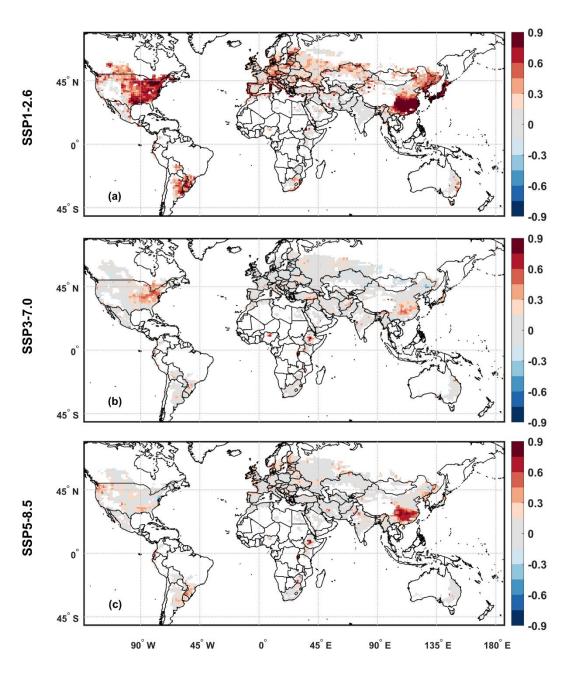


385

Figure 5: Mean POD₆ over the baseline period (a,b), and ΔPOD₆ at 2100 with respect to the baseline across the different SSPs for GFDL-ESM4 (c,e,g) and UKESM1-0-LL (d,f,h). These results are from the rain-fed simulation. Note: Baseline indicates the mean between 2000 and 2014, and 2100 indicates the mean between 2090-2099.







390 Figure 6: Pearson correlation coefficient (ρ) between ΔPOD_6 at 2100 calculated from GFDL-ESM4 and UKESM1-0-LL. These results are from the rain-fed simulation.





3.4 POD₆ under assumption of field capacity

The plant available water in the soil is commonly set to non-limiting conditions in the simplified flux calculations for integrated assessment modelling (POD_YIAM; LRTAP Convention, 2017). This approach estimates POD_Y to provide an indicative O_3 risk assessment at large scale. Due to its assumptions, POD_YIAM indicates the damage under the worst-case scenario, i.e., the potential damage to plants when soil water is not a limiting factor for g_s and O_3 uptake. Our 'field capacity' run assumes the same soil water hypothesis as this framework, and the comparison with the results of the rain-fed run offers an estimate of the O_3 risk uncertainty related to water availability for plants. In this context, the POD₆ under the FC assumption represents the maximum possible O_3 risk.

- 400 Figure 7 shows the differences between POD₆ calculated with the 'field capacity' run, and POD₆ calculated with the 'rainfed' run. Differences between the POD₆ calculated through the two runs are shown at each time window – for the baseline and for the end of the century – and for each SSP. Southern Europe (lat.<45°) and South Asia are the regions experiencing the most extensive increase of POD₆ in field capacity condition with respect to the rain-fed ones. On average, POD₆ under field capacity is 0.74 ± 0.30 mmol m⁻² greater than POD₆ under rain-fed conditions in Southern Europe, and 0.76 ± 0.21
- 405 mmol m⁻² greater in South Asia, over the whole century, and across SSPs and models. POD₆ under field capacity in South-East Asia and in North Africa increases by 1.68 ± 0.47 mmol m⁻², and 1.07 ± 0.39 mmol m⁻². However, only small fractions of these regions are affected, since most of the land-use does not include wheat. Other regions with increased POD₆ under the FC assumption are: Middle East (+0.90 ± 0.35 mmol m⁻²), Central America (+0.90 ± 0.39 mmol m⁻²), and Sub-Saharan Africa (+0.81 ± 0.19 mmol m⁻²).
- 410 It is important to underline that wheat cultivation under strong water limitations ($f_{soil} < 0.1$; Figure 3i-l) would need additional water sources besides rain. In this case, the 'field capacity' run might be more suitable to assess the potential O₃ risk, because POD₆ over widely irrigated regions would be underestimated using the rain-fed condition approach. The results in this section (Figure 7) suggest that, regardless of the considered scenario, there would be more potential hotspots worldwide than previously highlighted (Figure 5). For instance, South Asia consistently has the highest O₃ concentrations
- 415 across the globe (Figure 1), but the rain-fed simulation indicates that g_s is strongly limited by soil water availability, preventing any O_3 risk in a region with major wheat production (Monfreda et al., 2008). However, since this region requires extensive irrigation (Brauman et al., 2013; Chiarelli et al., 2020), the POD₆ calculated through the field capacity simulation might be more suitable to quantify the O_3 risk, indicating that this area might be a further hotspot for O_3 damage and food security. Similar conclusions can be drawn for any areas with large differences between field capacity POD₆ and rain-fed
- 420 POD₆, as long as irrigation is supplied for wheat production.





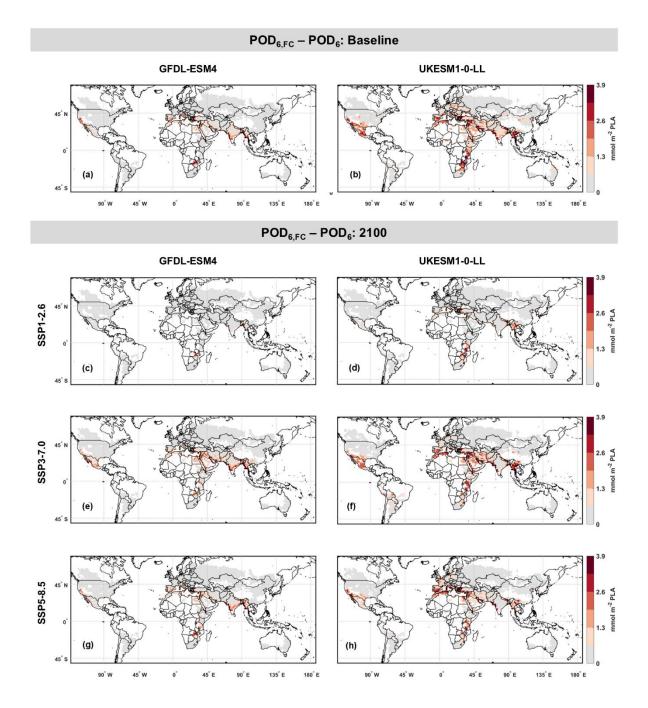


Figure 7: Mean differences between POD_{6,FC} calculated under FC conditions and the POD₆ calculated under rain-fed conditions over the baseline period (a,b), and at 2100 across the different SSPs for GFDL-ESM4 (c,e,g) and UKESM1-0-LL (d,f,h). Baseline indicates the mean between 2000 and 2014, and 2100 indicates the mean between 2090-2099.





425 **3.5** The contribution of radiative forcing and pollutant emissions to POD₆

Here we perform an ANOVA (see details in Section 2.5) to estimate which factor among EP, RF and their interaction (I), has the most significant impact on POD₆ changes at the end of the century around the globe. The different levels of EP and RF are described in Table 2 for each SSP, and are responsible for the different O_3 concentrations and f_{clim} values reported in Figure 1, Figure 2, and Figure A6. Figure 8a shows for each node the factor with the most significant effect (i.e. the lowest 430 p-value among the significant ones) among EP, RF, or their interaction (I), obtained from the two-way ANOVA, while Figure 8b-d display the total variance explained by each factor (\mathbb{R}^2), allowing further quantification of their contributions to the POD₆ changes. The ANOVA highlights different regions where the POD₆ changes are mostly driven by the same factor. For example, in East China the decrease of POD_6 under SSP1-2.6 and SSP5-8.5 (Figure 4; Figure 5; Table 4) significantly depends on EP. In South America POD₆ changes are controlled by EP too. In Eastern U.S. and in Japan the large decrease in 435 the POD_6 under any SSP (Figure 5) is controlled by the combined effect of EP and RF (interaction), i.e. the effect of a change in emission policies on the POD_6 depends on the given specific climate change signal and viceversa. In Russia-Belarus-Ukraine and in Northern Europe (lat.>45°), changes in POD₆ are driven by RF, which largely determines f_{clim} differences across scenarios, with decreases over Russia-Belarus-Ukraine, and increases over Northern Europe. This finding suggests that regions where RF controls POD₆ changes could experience an increase in POD₆ even under strong EP. This 440 may happen, for instance, not only because of increased f_{clim} , but also because O₃ increases more as a result of the RF

- enhancement rather than due to precursors emissions, which would be a climate penalty effect on O_3 concentrations (for instance, see Northern Europe under SSP5-8.5, Figure 5h). In regions where POD₆ decreases the ANOVA identifies the factors that are mainly responsible for this decrease, such as in East China, Eastern U.S. and Japan. In contrast, where POD₆ increases the ANOVA suggests the policies that should be adopted to mitigate O_3 risk for crops. For instance, in Northern
- Europe, where RF is responsible for the POD₆ increase, the adoption of stricter reduction policies on greenhouse gases emissions, rather than only the adoption of local EP, could lead to reduction of O₃ risk for crops as a climate co-benefit. However, this ANOVA is performed using the data from UKESM1-0-LL only, because GFDL doesn't include the scenario SSP3-7.0pdSST. Including more models could lead to different results, especially for the climate change component (RF), since it is known that UKESM1-0-LL simulates higher warming compared to many other CMIP6 models (McBride et al.,
- 450 2021; Scafetta, 2023).





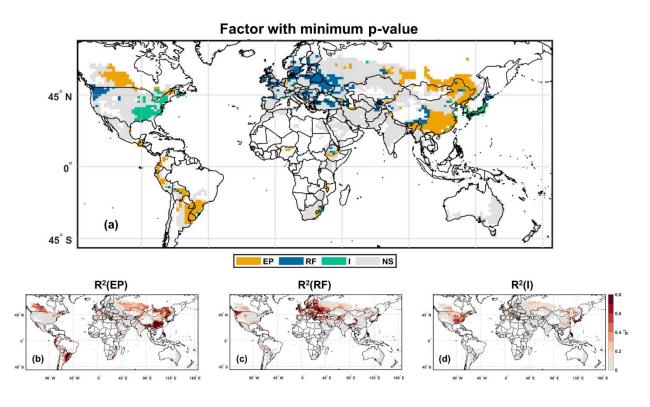


Figure 8: Map of the factor with the minimum significant p-value (p<0.05, corrected with Bonferroni), determined by the two-way ANOVA (a). Explained variance R^2 associated with each factor in each node, with black dots indicating significant p-value (b-d). EP=Emission Policy, RF=Radiative Forcing, I=Interaction, NS=Not Significant.

455 **4. Discussion**

This study highlights how wheat in different areas of the globe may be affected by O_3 damage in the future. Our results may be associated with different degrees of confidence depending on the agreement between the two available CMIP6 models, and whether an O_3 risk is predicted under rain-fed conditions, or just under the FC assumption. Under the rain-fed simulation (and therefore in the FC simulation as well), both the GFDL-ESM4 and UKESM1-0-LL models, consistently identify

- 460 Europe, East Asia, and the southern and eastern edges of the Tibetan plateau as the main areas at risk throughout the 21^{st} century, despite discrepancies in the estimated POD₆ values (Figure 5). The Eastern U.S. is currently among the most critical regions for O₃ damage to wheat, but both models project lower POD₆ values under all scenarios at the end of the century, suggesting that this area should not be a future concern. Additionally, the FC run indicates that South Asia might be extensively affected by O₃ damage in the future. Given the region's widespread need for irrigation (Brauman et al., 2013;
- 465 Chiarelli et al., 2020), the FC simulation might give a better estimate of POD_6 changes over this area, although the POD_6 calculated under the FC assumption represents the maximum possible O_3 risk and is therefore associated with higher uncertainty. The Indian peninsula and the southern and eastern edges of the Tibetan plateau, being major wheat producers





(Monfreda et al., 2008), should be considered among the areas at highest O_3 risk globally across the 21st century. South America, Central America and Sub-Saharan Africa are identified only by UKESM1-0-LL as regions subject to O₃ damage to 470 wheat. The increases in POD₆, particularly in Central America and Sub-Saharan Africa, are relatively large compared to the baseline and could grow larger under the FC assumption. This fact is particularly concerning for Sub-Saharan Africa, where intense population growth and development is expected during the 21st century (Kc and Lutz, 2017; United Nations Department of Economic and Social Affairs, Population Division, 2022).

- Some external sources of uncertainty in this study should also be discussed and are mainly related to the spatiotemporal 475 resolution and the bias of the input variables. The coarser temporal resolution from the CMIP6 models can lead to underestimation of O_3 flux peaks and POD₆ values (Guaita et al., 2023). Additionally, the coarse spatial resolution of ESMs fails to capture local meteorological features influenced by topology and land cover, and it can misrepresent chemical processes (Brands, 2022). More generally, the coarse resolution is related to biases for any variable, which have been partially evaluated in past literature (e.g. temperature Sellar et al., 2019, O3 in Turnock et al., 2020). Bias correction is
- 480 beyond the scope of this research; however, we performed a sensitivity analysis by perturbing input variables to evaluate the effect of potential biases on the final POD₆ estimates. For the sake of this, we altered separately temperature ($\pm 2^{\circ}$ C), vapor pressure (± 0.2 kPa), and O₃ concentrations (± 5 ppb) in the rain-fed simulation over the baseline for the UKESM1-0-LL model. The results of these simulations reveal that the standard deviation of the POD_6 differences between the perturbed and the original POD₆ simulations is smaller than the critical level (1.3 mmol m^{-2} PLA), with temperature the factor contributing
- 485 the most to the perturbed POD_6 response (Figure A5). A more detailed analysis of the sensitivity of POD to ozone and meteorological variables under present-day conditions is being undertaken in a companion study comparing different stomatal ozone flux models [Emmerichs et al., in prep].

Past studies, although using different modelling approaches, generally show good agreement with our results in spatial patterns and identified hotspots of O₃ damage (Lombardozzi et al., 2015; Schauberger et al., 2019; Sitch et al., 2007; Tai et

- 490 al., 2021). More specifically, they identified the eastern U.S., Europe, and East Asia as hotspots, with substantial overlaps. However, there are differences with the available present-day regional assessments that evaluated O₃ effects on wheat. For example, POD₆ was estimated to reach values up to 8.5 mmol m^{-2} in the Iberian Peninsula (De Andrés et al., 2012), while our results indicate potential threats without reaching such extreme POD_6 values. Mills et al. (2011) found that central Europe and Mediterranean coasts are the areas most affected by O₃, which is a feature that is only partially represented in our
- results. On the other hand, our findings are consistent with past estimates of POD₃IAM in Sub-Saharan Africa under the field 495 capacity assumption (Sharps et al., 2021b), and with the identified areas most affected by O_3 in China (Cao et al., 2024; Qinyi et al., 2023; Wang et al., 2022). The observed differences are likely due to the different spatial resolutions adopted in these regional assessments. Therefore, it is important to emphasize the need of bias-corrected high-resolution spatiotemporal projections of O₃ concentrations and meteorology for future regional-scale assessments (Liu et al., 2022).
- 500 The most recent literature (Zhou et al., 2024) used ModelE2-YIBs to estimate the O₃ effect on GPP, both for the present (2010) and future times (2060) under SSP1-2.6 and SSP5-8.5 scenarios. Their identified spatial patterns are quite similar to





ours, with only minor differences in Western Africa and South America which are likely due to our focus on wheat crops rather than vegetation in general. However, their coarser spatial resolution $(2^{\circ} \times 2.5^{\circ})$ might mask some of the spatial patterns identified in our results.

- 505 Our results and their consequences for food security can be interpreted in the broader context of the SSPs. Under SSP1 there is a clear decline in O₃ risk everywhere, implying a relatively unaffected crop production at the end of the century (globally, 95th percentile of RYL: 2.2%), which would favor food sovereignty, and thus easing adaptation challenges. On the other hand, under SSP3, POD₆ would increase throughout the century. This will imply higher crop losses at 2100 for regions with high population growth, particularly South Asia and Sub-Saharan Africa (95th perc. of RYL: 20.0% and 9.3%, respectively).
- 510 Since this scenario also includes barriers to trade and to international cooperation, these high food security threats would exacerbate adaptation challenges. With respect to O_3 threats, SSP5 is classified as an intermediate scenario, since O_3 risk rises to about 2050 in many regions (globally, 95th perc. of RYL: 10.9%) and then reverts to levels similar to or lower than the baseline by the end of the century (7.7%), depending on the area. However, this scenario describes a world that emphasizes technological development, which might also include effective adaptations to food security threats. Ultimately,
- 515 SSP3 raises an environmental injustice issue, because it implies that the most affected nations facing food security threats would be the ones with lower resources to address adaptation challenges. On the contrary, a large mitigation effort, such as the one implied within SSP1, would guarantee environmental justice by avoiding food security issues worldwide.

5. Conclusions

This study quantified the global and regional evolution of the risk to wheat from O₃, expressed as POD₆, under different 520 Shared Socioeconomical Pathways (SSPs) up to 2100. The results show that for SSP1-2.6, the O₃ risk will decrease worldwide by the end of the century. Under SSP5-8.5 the POD₆ will increase by mid-century before reverting to present-day levels by 2100 in most of the considered regions. On the other hand, under SSP3-7.0, there will be an increase in POD₆ over many regions by the end of the century.

In this latter scenario, both CMIP6 models (UKESM1-0-LL and GFDL-ESM4) agree that Northern Europe, East China, and 525 the southern and eastern edges of the Tibetan Plateau will be the most extensively at risk for future food security, both under rain-fed and field capacity conditions, with POD₆ values that will be about six times greater than the current critical level for grain yield loss over the latter region. However, for other regions, like South Asia and Sub-Saharan Africa, the predicted adverse O₃ effects on wheat are more uncertain, as they are shown only by one of the CMIP6 models or only occurring under field capacity conditions. Analysis of these regions would benefit from employing finer spatio-temporal resolutions, as

530 revealed by the comparison made with other regional-scale studies. Finally, although the dry deposition model employed in this research allows explicit evaluation of the O_3 risk for wheat, it does not account for any vegetation feedback to O_3 concentrations and climate beyond what is already embedded in the CMIP6 models.





535

More stringent policies to reduce radiative forcing and O₃ precursor emissions, as expressed in SSP1-2.6, would lead to almost complete removal of the O₃ risk for wheat, thus avoiding O₃-induced food security issues worldwide. Conversely, the control policies directed to limit solely O₃ precursors but not greenhouse gases (SSP5-8.5) does not necessarily lead to benefits, as observed for instance in Northern Europe. Finally, the SSP3-7.0 simulations highlight that enhanced radiative forcing and uncontrolled O₃ precursor emissions may raise environmental justice issues, as food security threats would exacerbate adaptation challenges for low-income nations.

6. Author contributions

540 PRG and GG conceptualized the study, the methodology, and prepared the software; PRG curated and analysed the data, and created the visualization; PRG and ST performed the investigation and the data collection of UKESM1-0-LL and GFDL-ESM4 outputs; GG administered the project; GG and RM provided the computing resources; GG, RM and PC supervised the research activity; PRG, GG, RM, PC and LZ contributed to the original draft preparation; all authors reviewed and edited the final version of the paper.

545 7. Competing interests

Leiming Zhang is member of the editorial board of journal Atmospheric Chemistry and Physics.

8. Code and Data availability

The individual available data for years used produce the figures in this study is at to https://doi.org/10.5281/zenodo.13485000. The Matlab code for the model can be made available upon request to the authors.

550 9. Acknowledgements

We thank dr. Shengchao Qiao and colleagues for their valuable contribution in providing the sowing dates dataset of wheat and for their useful comments to adapt their results to our case-study. This manuscript benefits from discussions during the meeting of the Ozone Deposition Focus working group of the Tropospheric Ozone Assessment Report, phase II. We would especially like to thank prof. Lisa Emberson for her fruitful insights on this work. We are grateful to Cazimir Kowalski for

555 offering helpful feedback on the writing style and structure of this manuscript. This work was also supported by the "5x1000 funding to research" of the Italian Government and by Catholic University of the Sacred Heart in the frame of its Programs of promotion and dissemination of the scientific research (Funding line D3.1).





References

Anav, A., Menut, L., Khvorostyanov, D., and Viovy, N.: Impact of tropospheric ozone on the Euro-Mediterranean vegetation, Global Change Biology, 17, 2342–2359, https://doi.org/10.1111/j.1365-2486.2010.02387.x, 2011.

Anav, A., De Marco, A., Proietti, C., Alessandri, A., Dell'Aquila, A., Cionni, I., Friedlingstein, P., Khvorostyanov, D., Menut, L., Paoletti, E., Sicard, P., Sitch, S., and Vitale, M.: Comparing concentration-based (AOT40) and stomatal uptake (PODY) metrics for ozone risk assessment to European forests, Global Change Biology, 22, 1608–1627, https://doi.org/10.1111/gcb.13138, 2016.

Archibald, A. T., O'Connor, F. M., Abraham, N. L., Archer-Nicholls, S., Chipperfield, M. P., Dalvi, M., Folberth, G. A., Dennison, F., Dhomse, S. S., Griffiths, P. T., Hardacre, C., Hewitt, A. J., Hill, R. S., Johnson, C. E., Keeble, J., Köhler, M. O., Morgenstern, O., Mulcahy, J. P., Ordóñez, C., Pope, R. J., Rumbold, S. T., Russo, M. R., Savage, N. H., Sellar, A., Stringer, M., Turnock, S. T., Wild, O., and Zeng, G.: Description and evaluation of the UKCA stratosphere–troposphere chemistry scheme (StratTrop vn 1.0) implemented in UKESM1, Geosci. Model
570 Dev., 13, 1223–1266, https://doi.org/10.5194/gmd-13-1223-2020, 2020.

Baldocchi, D. D., Hicks, B. B., and Camara, P.: A canopy stomatal resistance model for gaseous deposition to vegetated surfaces, Atmospheric Environment (1967), 21, 91–101, https://doi.org/10.1016/0004-6981(87)90274-5, 1987.

Ball, J. T., Woodrow, I. E., and Berry, J. A.: A Model Predicting Stomatal Conductance and its Contribution to the
 Control of Photosynthesis under Different Environmental Conditions, in: Progress in Photosynthesis Research,
 vol. 4, edited by: Biggens, J., Martinus Nijhoff Publishers, Netherlands, 221–224, https://doi.org/10.1007/978-94-017-0519-6_48, 1987.

Brands, S.: A circulation-based performance atlas of the CMIP5 and 6 models for regional climate studies in the Northern Hemisphere mid-to-high latitudes, Geosci. Model Dev., 15, 1375–1411, https://doi.org/10.5194/gmd-15-1375-2022, 2022.

- Brauman, K. A., Siebert, S., and Foley, J. A.: Improvements in crop water productivity increase water sustainability and food security—a global analysis, Environ. Res. Lett., 8, 024030, https://doi.org/10.1088/1748-9326/8/2/024030, 2013.
- Cao, J., Yue, X., and Ma, M.: Simulation of ozone-vegetation coupling and feedback in China using multiple ozone damage schemes, Atmos. Chem. Phys., 24, 3973–3987, https://doi.org/10.5194/acp-24-3973-2024, 2024.

Charusombat, U., Niyogi, D., Kumar, A., Wang, X., Chen, F., Guenther, A., Turnipseed, A., and Alapaty, K.: Evaluating a New Deposition Velocity Module in the Noah Land-Surface Model, Boundary-Layer Meteorol, 137, 271–290, https://doi.org/10.1007/s10546-010-9531-y, 2010.

Cheesman, A. W., Brown, F., Farha, M. N., Rosan, T. M., Folberth, G. A., Hayes, F., Moura, B. B., Paoletti, E.,
Hoshika, Y., Osborne, C. P., Cernusak, L. A., Ribeiro, R. V., and Sitch, S.: Impacts of ground-level ozone on sugarcane production, Science of The Total Environment, 904, 166817, https://doi.org/10.1016/j.scitotenv.2023.166817, 2023.





Chiarelli, D. D., Passera, C., Rosa, L., Davis, K. F., D'Odorico, P., and Rulli, M. C.: The green and blue crop water requirement WATNEEDS model and its global gridded outputs, Sci Data, 7, 273. 595 https://doi.org/10.1038/s41597-020-00612-0, 2020.

Chuwah, C., Van Noije, T., Van Vuuren, D. P., Stehfest, E., and Hazeleger, W.: Global impacts of surface ozone changes crop vields and land use. Atmospheric Environment, 106, 11-23, on https://doi.org/10.1016/j.atmosenv.2015.01.062, 2015.

- Cinquini, L., Crichton, D., Mattmann, C., Harney, J., Shipman, G., Wang, F., Ananthakrishnan, R., Miller, N., Denvil, S., Morgan, M., Pobre, Z., Bell, G. M., Doutriaux, C., Drach, R., Williams, D., Kershaw, P., Pascoe, S., 600 Gonzalez, E., Fiore, S., and Schweitzer, R.: The Earth System Grid Federation: An open infrastructure for access to distributed geospatial data, Future Generation Computer Systems, 36. 400-417, https://doi.org/10.1016/j.future.2013.07.002, 2014.
- Clifton, O. E., Schwede, D., Hogrefe, C., Bash, J. O., Bland, S., Cheung, P., Coyle, M., Emberson, L., Flemming, 605 J., Fredj, E., Galmarini, S., Ganzeveld, L., Gazetas, O., Goded, I., Holmes, C. D., Horváth, L., Huijnen, V., Li, Q., Makar, P. A., Mammarella, I., Manca, G., Munger, J. W., Pérez-Camanyo, J. L., Pleim, J., Ran, L., San Jose, R., Silva, S. J., Staebler, R., Sun, S., Tai, A. P. K., Tas, E., Vesala, T., Weidinger, T., Wu, Z., and Zhang, L.: A singlepoint modeling approach for the intercomparison and evaluation of ozone dry deposition across chemical transport models (Activity 2 of AQMEII4), Atmos. Chem. Phys., 23, 9911–9961, https://doi.org/10.5194/acp-23-9911-2023, 2023. 610
 - Collins, W. J., Lamarque, J.-F., Schulz, M., Boucher, O., Eyring, V., Hegglin, M. I., Maycock, A., Myhre, G., Prather, M., Shindell, D., and Smith, S. J.: AerChemMIP: quantifying the effects of chemistry and aerosols in CMIP6, Geosci. Model Dev., 10, 585–607, https://doi.org/10.5194/gmd-10-585-2017, 2017.
- De Andrés, J. M., Borge, R., De La Paz, D., Lumbreras, J., and Rodríguez, E.: Implementation of a module for risk of ozone impacts assessment to vegetation in the Integrated Assessment Modelling system for the Iberian 615 Peninsula. Evaluation for wheat and Holm oak, Environmental Pollution, 165. 25 - 37, https://doi.org/10.1016/j.envpol.2012.01.048, 2012.

Dunne, J. P., Horowitz, L. W., Adcroft, A. J., Ginoux, P., Held, I. M., John, J. G., Krasting, J. P., Malyshev, S., Naik, V., Paulot, F., Shevliakova, E., Stock, C. A., Zadeh, N., Balaji, V., Blanton, C., Dunne, K. A., Dupuis, C., Durachta, J., Dussin, R., Gauthier, P. P. G., Griffies, S. M., Guo, H., Hallberg, R. W., Harrison, M., He, J., Hurlin,

- 620 W., McHugh, C., Menzel, R., Milly, P. C. D., Nikonov, S., Paynter, D. J., Ploshay, J., Radhakrishnan, A., Rand, K., Reichl, B. G., Robinson, T., Schwarzkopf, D. M., Sentman, L. T., Underwood, S., Vahlenkamp, H., Winton, M., Wittenberg, A. T., Wyman, B., Zeng, Y., and Zhao, M.: The GFDL Earth System Model Version 4.1 (GFDL-ESM 4.1): Overall Coupled Model Description and Simulation Characteristics, J Adv Model Earth Syst, 12,
- e2019MS002015, https://doi.org/10.1029/2019MS002015, 2020. 625

Emberson, L.: Effects of ozone on agriculture, forests and grasslands, Phil. Trans. R. Soc. A., 378, 20190327, https://doi.org/10.1098/rsta.2019.0327, 2020.

Emberson, L. D., Ashmore, M. R., Cambridge, H. M., and Simpson, D.: Modelling stomatal ozone flux across Europe, Environmental Pollution, 109, 403-413, 2000.



650



630 Erisman, J. W., Van Pul, A., and Wyers, P.: Parametrization of surface resistance for the quantification of atmospheric deposition of acidifying pollutants and ozone, Atmospheric Environment, 28, 2595–2607, https://doi.org/10.1016/1352-2310(94)90433-2, 1994.

Eyring, V., Bony, S., Meehl, G. A., Senior, C. A., Stevens, B., Stouffer, R. J., and Taylor, K. E.: Overview of the Coupled Model Intercomparison Project Phase 6 (CMIP6) experimental design and organization, Geosci. Model
Dev., 9, 1937–1958, https://doi.org/10.5194/gmd-9-1937-2016, 2016.

Fang, Z., Zhang, W., Brandt, M., Abdi, A. M., and Fensholt, R.: Globally Increasing Atmospheric Aridity Over the 21st Century, Earth's Future, 10, e2022EF003019, https://doi.org/10.1029/2022EF003019, 2022.

Feng, Z., Tang, H., Uddling, J., Pleijel, H., Kobayashi, K., Zhu, J., Oue, H., and Guo, W.: A stomatal ozone flux–response relationship to assess ozone-induced yield loss of winter wheat in subtropical China, Environmental
Pollution, 164, 16–23, https://doi.org/10.1016/j.envpol.2012.01.014, 2012.

Fuhrer, J., Val Martin, M., Mills, G., Heald, C. L., Harmens, H., Hayes, F., Sharps, K., Bender, J., and Ashmore, M. R.: Current and future ozone risks to global terrestrial biodiversity and ecosystem processes, Ecology and Evolution, 6, 8785–8799, https://doi.org/10.1002/ece3.2568, 2016.

Fujimori, S., Hasegawa, T., Masui, T., Takahashi, K., Herran, D. S., Dai, H., Hijioka, Y., and Kainuma, M.: SSP3:
AIM implementation of Shared Socioeconomic Pathways, Global Environmental Change, 42, 268–283, https://doi.org/10.1016/j.gloenvcha.2016.06.009, 2017.

González-Fernández, I., Bermejo, V., Elvira, S., De La Torre, D., González, A., Navarrete, L., Sanz, J., Calvete, H., García-Gómez, H., López, A., Serra, J., Lafarga, A., Armesto, A. P., Calvo, A., and Alonso, R.: Modelling ozone stomatal flux of wheat under mediterranean conditions, Atmospheric Environment, 67, 149–160, https://doi.org/10.1016/j.atmosenv.2012.10.043, 2013.

Grulke, N. E. and Heath, R. L.: Ozone effects on plants in natural ecosystems, Plant Biol J, 22, 12–37, https://doi.org/10.1111/plb.12971, 2020.

Grünhage, L., Pleijel, H., Mills, G., Bender, J., Danielsson, H., Lehmann, Y., Castell, J.-F., and Bethenod, O.: Updated stomatal flux and flux-effect models for wheat for quantifying effects of ozone on grain yield, grain mass and protein yield, Environmental Pollution, 165, 147–157, https://doi.org/10.1016/j.envpol.2012.02.026,

mass and protein yield, Environmental Pollution, 165, 147–157, https://doi.org/10.1016/j.envpol.2012.02.026, 2012.

Guaita, P. R., Marzuoli, R., and Gerosa, G. A.: A regional scale flux-based O3 risk assessment for winter wheat in northern Italy, and effects of different spatio-temporal resolutions, Environmental Pollution, 333, 121860, https://doi.org/10.1016/j.envpol.2023.121860, 2023.

660 Hicks, B. B., Matt, D. R., McMillen, R. T., and Box, P. O.: A micrometeorological investigation of surface exchange of O3, SO2 and NO2: a case study., Boundary-Layer Meteorology, 47, 321–336, 1989.

Horowitz, L. W., Naik, V., Sentman, L., Paulot, F., Blanton, C., McHugh, C., Radhakrishnan, A., Rand, K., Vahlenkamp, H., Zadeh, N. T., Wilson, C., Ginoux, P., He, J., John, J. G., Lin, M., Paynter, D. J., Ploshay, J.,





- Zhang, A., and Zeng, Y.: NOAA-GFDL GFDL-ESM4 model output prepared for CMIP6 AerChemMIP (20230720), https://doi.org/10.22033/ESGF/CMIP6.1404, 2018. 665
 - Horowitz, L. W., Naik, V., Paulot, F., Ginoux, P. A., Dunne, J. P., Mao, J., Schnell, J., Chen, X., He, J., John, J. G., Lin, M., Lin, P., Malyshev, S., Paynter, D., Shevliakova, E., and Zhao, M.: The GFDL Global Atmospheric Chemistry-Climate Model AM4.1: Model Description and Simulation Characteristics, J Adv Model Earth Syst, 12, e2019MS002032, https://doi.org/10.1029/2019MS002032, 2020.
- 670 Hoshika, Y., Paoletti, E., Agathokleous, E., Sugai, T., and Koike, T.: Developing Ozone Risk Assessment for Larch Species, Front. For. Glob. Change, 3, 45, https://doi.org/10.3389/ffgc.2020.00045, 2020.

Intergovernmental Panel On Climate Change (IPCC): Technical Summary, in: Climate Change 2021 - The Physical Science Basis: Working Group I Contribution to the Sixth Assessment Report of the Intergovernmental Panel on Climate Change, Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA, 33-144, https://doi.org/10.1017/9781009157896, 2023.

Jarvis, P. G.: The interpretation of the variations in leaf water potential and stomatal conductance found in canopies in the field, Phil. Trans. R. Soc. Lond. B, 273, 593-610, https://doi.org/10.1098/rstb.1976.0035, 1976.

Karlsson, P. E., Braun, S., Broadmeadow, M., Elvira, S., Emberson, L., Gimeno, B. S., Le Thiec, D., Novak, K., Oksanen, E., Schaub, M., Uddling, J., and Wilkinson, M.: Risk assessments for forest trees: The performance of 608-616, the ozone flux versus the concepts, Environmental Pollution. 146. AOT https://doi.org/10.1016/j.envpol.2006.06.012, 2007.

Kc, S. and Lutz, W.: The human core of the shared socioeconomic pathways: Population scenarios by age, sex and level of education for all countries to 2100, Global Environmental Change, 42, 181-192, https://doi.org/10.1016/j.gloenvcha.2014.06.004, 2017.

Klingberg, J., Engardt, M., Karlsson, P. E., Langner, J., and Pleijel, H.: Declining ozone exposure of European 685 vegetation under climate change and reduced precursor emissions, Biogeosciences, 11, 5269-5283, https://doi.org/10.5194/bg-11-5269-2014, 2014.

Krasting, J. P., John, J. G., Blanton, C., McHugh, C., Nikonov, S., Radhakrishnan, A., Rand, K., Zadeh, N. T., Balaji, V., Durachta, J., Dupuis, C., Menzel, R., Robinson, T., Underwood, S., Vahlenkamp, H., Dunne, K. A., Gauthier, P. P., Ginoux, P., Griffies, S. M., Hallberg, R., Harrison, M., Hurlin, W., Malyshev, S., Naik, V., Paulot, 690 F., Paynter, D. J., Ploshay, J., Reichl, B. G., Schwarzkopf, D. M., Seman, C. J., Silvers, L., Wyman, B., Zeng, Y., Adcroft, A., Dunne, J. P., Dussin, R., Guo, H., He, J., Held, I. M., Horowitz, L. W., Lin, P., Milly, P. C. D., Shevliakova, E., Stock, C., Winton, M., Wittenberg, A. T., Xie, Y., and Zhao, M.: NOAA-GFDL GFDL-ESM4 model output prepared for CMIP6 CMIP historical (20230220), https://doi.org/10.22033/ESGF/CMIP6.8597, 2018.

695

675

680

Kriegler, E., Bauer, N., Popp, A., Humpenöder, F., Leimbach, M., Strefler, J., Baumstark, L., Bodirsky, B. L., Hilaire, J., Klein, D., Mouratiadou, I., Weindl, I., Bertram, C., Dietrich, J.-P., Luderer, G., Pehl, M., Pietzcker, R., Piontek, F., Lotze-Campen, H., Biewald, A., Bonsch, M., Giannousakis, A., Kreidenweis, U., Müller, C., Rolinski, S., Schultes, A., Schwanitz, J., Stevanovic, M., Calvin, K., Emmerling, J., Fujimori, S., and Edenhofer,



705



700 O.: Fossil-fueled development (SSP5): An energy and resource intensive scenario for the 21st century, Global Environmental Change, 42, 297–315, https://doi.org/10.1016/j.gloenvcha.2016.05.015, 2017.

Kuhlbrodt, T., Jones, C. G., Sellar, A., Storkey, D., Blockley, E., Stringer, M., Hill, R., Graham, T., Ridley, J., Blaker, A., Calvert, D., Copsey, D., Ellis, R., Hewitt, H., Hyder, P., Ineson, S., Mulcahy, J., Siahaan, A., and Walton, J.: The Low-Resolution Version of HadGEM3 GC3.1: Development and Evaluation for Global Climate, J Adv Model Earth Syst, 10, 2865–2888, https://doi.org/10.1029/2018MS001370, 2018.

Liu, Z., Doherty, R. M., Wild, O., O'Connor, F. M., and Turnock, S. T.: Correcting ozone biases in a global chemistry–climate model: implications for future ozone, Atmos. Chem. Phys., 22, 12543–12557, https://doi.org/10.5194/acp-22-12543-2022, 2022.

Lombardozzi, D., Levis, S., Bonan, G., Hess, P. G., and Sparks, J. P.: The Influence of Chronic Ozone Exposure
on Global Carbon and Water Cycles, Journal of Climate, 28, 292–305, https://doi.org/10.1175/JCLI-D-14-00223.1, 2015.

LRTAP Convention: Chapter III: mapping critical level for vegetation, in: Modelling and Mapping Manual, 2017.

Mao, H., Felker-Quinn, E., Sive, B., Zhang, L., Ye, Z., and Fang, H.: Examining indicators and methods for quantifying ozone exposure to vegetation, Atmospheric Environment, 316, 120195, https://doi.org/10.1016/j.atmosenv.2023.120195, 2024.

McBride, L. A., Hope, A. P., Canty, T. P., Bennett, B. F., Tribett, W. R., and Salawitch, R. J.: Comparison of CMIP6 historical climate simulations and future projected warming to an empirical model of global climate, Earth Syst. Dynam., 12, 545–579, https://doi.org/10.5194/esd-12-545-2021, 2021.

 Mills, G., Hayes, F., Simpson, D., Emberson, L., Norris, D., Harmens, H., and Büker, P.: Evidence of widespread
 effects of ozone on crops and (semi-)natural vegetation in Europe (1990-2006) in relation to AOT40- and fluxbased risk maps: OZONE EFFECTS ON VEGETATION IN EUROPE, Global Change Biology, 17, 592–613, https://doi.org/10.1111/j.1365-2486.2010.02217.x, 2011.

Mills, G., Sharps, K., Simpson, D., Pleijel, H., Broberg, M., Uddling, J., Jaramillo, F., Davies, W. J., Dentener, F., Van Den Berg, M., Agrawal, M., Agrawal, S. B., Ainsworth, E. A., Büker, P., Emberson, L., Feng, Z., Harmens,

H., Hayes, F., Kobayashi, K., Paoletti, E., and Van Dingenen, R.: Ozone pollution will compromise efforts to increase global wheat production, Global Change Biology, 24, 3560–3574, https://doi.org/10.1111/gcb.14157, 2018.

Monfreda, C., Ramankutty, N., and Foley, J. A.: Farming the planet: 2. Geographic distribution of crop areas, yields, physiological types, and net primary production in the year 2000, Global Biogeochemical Cycles, 22, 2007GB002947, https://doi.org/10.1029/2007GB002947, 2008.

Mulcahy, J. P., Jones, C., Sellar, A., Johnson, B., Boutle, I. A., Jones, A., Andrews, T., Rumbold, S. T., Mollard, J., Bellouin, N., Johnson, C. E., Williams, K. D., Grosvenor, D. P., and McCoy, D. T.: Improved Aerosol Processes and Effective Radiative Forcing in HadGEM3 and UKESM1, J Adv Model Earth Syst, 10, 2786–2805, https://doi.org/10.1029/2018MS001464, 2018.





735 O'Connor, F.: NERC UKESM1.0-LL model output prepared for CMIP6 AerChemMIP (20230720), https://doi.org/10.22033/ESGF/CMIP6.405, 2020.

Paoletti, E. and Manning, W. J.: Toward a biologically significant and usable standard for ozone that will also protect plants, Environmental Pollution, 150, 85–95, https://doi.org/10.1016/j.envpol.2007.06.037, 2007.

Paoletti, E., Alivernini, A., Anav, A., Badea, O., Carrari, E., Chivulescu, S., Conte, A., Ciriani, M. L., DalsteinRichier, L., De Marco, A., Fares, S., Fasano, G., Giovannelli, A., Lazzara, M., Leca, S., Materassi, A., Moretti, V.,
Pitar, D., Popa, I., Sabatini, F., Salvati, L., Sicard, P., Sorgi, T., and Hoshika, Y.: Toward stomatal-flux based
forest protection against ozone: The MOTTLES approach, Science of The Total Environment, 691, 516–527,
https://doi.org/10.1016/j.scitotenv.2019.06.525, 2019.

Pleijel, H., Danielsson, H., Emberson, L., Ashmore, M. R., and Mills, G.: Ozone risk assessment for agricultural
crops in Europe: Further development of stomatal flux and flux–response relationships for European wheat and
potato, Atmospheric Environment, 41, 3022–3040, https://doi.org/10.1016/j.atmosenv.2006.12.002, 2007.

Pleijel, H., Broberg, M. C., Uddling, J., and Mills, G.: Current surface ozone concentrations significantly decrease wheat growth, yield and quality, Science of The Total Environment, 613–614, 687–692, https://doi.org/10.1016/j.scitotenv.2017.09.111, 2018.

750 Pleijel, H., Danielsson, H., and Broberg, M. C.: Benefits of the Phytotoxic Ozone Dose (POD) index in doseresponse functions for wheat yield loss, Atmospheric Environment, 268, 118797, https://doi.org/10.1016/j.atmosenv.2021.118797, 2022.

Qiao, S., Harrison, S. P., Prentice, I. C., and Wang, H.: Optimality-based modelling of wheat sowing dates globally, Agricultural Systems, 206, 103608, https://doi.org/10.1016/j.agsy.2023.103608, 2023.

- 755 Qinyi, W., Mei, H., Shaoqiang, W., Bin, C., Zhenhai, L., Zhaosheng, W., Shiliang, C., Hui, L., Tongtong, Z., Donghui, L., Yuelin, L., Hu, L., and Leigang, S.: Evaluation of the impacts of ozone on the vegetation productivity of woodland and grassland ecosystems in China, Ecological Modelling, 483, 110426, https://doi.org/10.1016/j.ecolmodel.2023.110426, 2023.
- Ramya, A., Dhevagi, P., Poornima, R., Avudainayagam, S., Watanabe, M., and Agathokleous, E.: Effect of ozone
 stress on crop productivity: A threat to food security, Environmental Research, 236, 116816, https://doi.org/10.1016/j.envres.2023.116816, 2023.

Riahi, K., Van Vuuren, D. P., Kriegler, E., Edmonds, J., O'Neill, B. C., Fujimori, S., Bauer, N., Calvin, K., Dellink, R., Fricko, O., Lutz, W., Popp, A., Cuaresma, J. C., Kc, S., Leimbach, M., Jiang, L., Kram, T., Rao, S., Emmerling, J., Ebi, K., Hasegawa, T., Havlik, P., Humpenöder, F., Da Silva, L. A., Smith, S., Stehfest, E., Bosetti,

765 V., Eom, J., Gernaat, D., Masui, T., Rogelj, J., Strefler, J., Drouet, L., Krey, V., Luderer, G., Harmsen, M., Takahashi, K., Baumstark, L., Doelman, J. C., Kainuma, M., Klimont, Z., Marangoni, G., Lotze-Campen, H., Obersteiner, M., Tabeau, A., and Tavoni, M.: The Shared Socioeconomic Pathways and their energy, land use, and greenhouse gas emissions implications: An overview, Global Environmental Change, 42, 153–168, https://doi.org/10.1016/j.gloenvcha.2016.05.009, 2017.





770 Savi, F., Nemitz, E., Coyle, M., Aitkenhead, M., Frumau, K., Gerosa, G., Finco, A., Gruening, C., Goded, I., Loubet, B., Stella, P., Ruuskanen, T., Weidinger, T., Horvath, L., Zenone, T., and Fares, S.: Neural Network Analysis to Evaluate Ozone Damage to Vegetation Under Different Climatic Conditions, Front. For. Glob. Change, 3, 42, https://doi.org/10.3389/ffgc.2020.00042, 2020.

Scafetta, N.: CMIP6 GCM ensemble members versus global surface temperatures, Clim Dyn, 60, 3091–3120, https://doi.org/10.1007/s00382-022-06493-w, 2023.

Schauberger, B., Rolinski, S., Schaphoff, S., and Müller, C.: Global historical soybean and wheat yield loss estimates from ozone pollution considering water and temperature as modifying effects, Agricultural and Forest Meteorology, 265, 1–15, https://doi.org/10.1016/j.agrformet.2018.11.004, 2019.

Sellar, A. A., Jones, C. G., Mulcahy, J. P., Tang, Y., Yool, A., Wiltshire, A., O'Connor, F. M., Stringer, M., Hill,
R., Palmieri, J., Woodward, S., De Mora, L., Kuhlbrodt, T., Rumbold, S. T., Kelley, D. I., Ellis, R., Johnson, C. E., Walton, J., Abraham, N. L., Andrews, M. B., Andrews, T., Archibald, A. T., Berthou, S., Burke, E., Blockley, E., Carslaw, K., Dalvi, M., Edwards, J., Folberth, G. A., Gedney, N., Griffiths, P. T., Harper, A. B., Hendry, M. A., Hewitt, A. J., Johnson, B., Jones, A., Jones, C. D., Keeble, J., Liddicoat, S., Morgenstern, O., Parker, R. J., Predoi, V., Robertson, E., Siahaan, A., Smith, R. S., Swaminathan, R., Woodhouse, M. T., Zeng, G., and
Zerroukat, M.: UKESM1: Description and Evaluation of the U.K. Earth System Model, J Adv Model Earth Syst, 11, 4513–4558, https://doi.org/10.1029/2019MS001739, 2019.

Sharps, K., Hayes, F., Harmens, H., and Mills, G.: Ozone-induced effects on leaves in African crop species, Environmental Pollution, 268, 115789, https://doi.org/10.1016/j.envpol.2020.115789, 2021a.

Sharps, K., Vieno, M., Beck, R., Hayes, F., and Harmens, H.: Quantifying the impact of ozone on crops in SubSaharan Africa demonstrates regional and local hotspots of production loss, Environ Sci Pollut Res, 28, 62338–
62352, https://doi.org/10.1007/s11356-021-14967-3, 2021b.

Sicard, P., Anav, A., De Marco, A., and Paoletti, E.: Projected global ground-level ozone impacts on vegetation under different emission and climate scenarios, Atmos. Chem. Phys., 17, 12177–12196, https://doi.org/10.5194/acp-17-12177-2017, 2017.

795 Simpson, D., Ashmore, M. R., Emberson, L., and Tuovinen, J.-P.: A comparison of two different approaches for mapping potential ozone damage to vegetation. A model study, Environmental Pollution, 146, 715–725, https://doi.org/10.1016/j.envpol.2006.04.013, 2007.

Sitch, S., Cox, P. M., Collins, W. J., and Huntingford, C.: Indirect radiative forcing of climate change through ozone effects on the land-carbon sink, Nature, 448, 791–794, https://doi.org/10.1038/nature06059, 2007.

800 Stewart, J. B.: Modelling surface conductance of pine forest, Agricultural and Forest Meteorology, 43, 19–35, https://doi.org/10.1016/0168-1923(88)90003-2, 1988.

von Storch, H. and Zwiers, F.: Analysis of Variance, in: Statistical Analysis in Climate Research, Cambridge University Press, 171–192, 1999.





Tai, A. P. K., Sadiq, M., Pang, J. Y. S., Yung, D. H. Y., and Feng, Z.: Impacts of Surface Ozone Pollution on
 Global Crop Yields: Comparing Different Ozone Exposure Metrics and Incorporating Co-effects of CO2, Front.
 Sustain. Food Syst., 5, 534616, https://doi.org/10.3389/fsufs.2021.534616, 2021.

Tang, H., Pang, J., Zhang, G., Takigawa, M., Liu, G., Zhu, J., and Kobayashi, K.: Mapping ozone risks for rice in China for years 2000 and 2020 with flux-based and exposure-based doses, Atmospheric Environment, 86, 74–83, https://doi.org/10.1016/j.atmosenv.2013.11.078, 2014.

810 Tang, Y., Rumbold, S., Ellis, R., Kelley, D., Mulcahy, J., Sellar, A., Walton, J., and Jones, C.: MOHC UKESM1.0-LL model output prepared for CMIP6 CMIP historical (20230220), https://doi.org/10.22033/ESGF/CMIP6.6113, 2019.

Turnock, S. T., Allen, R. J., Andrews, M., Bauer, S. E., Deushi, M., Emmons, L., Good, P., Horowitz, L., John, J. G., Michou, M., Nabat, P., Naik, V., Neubauer, D., O'Connor, F. M., Olivié, D., Oshima, N., Schulz, M., Sellar,
A., Shim, S., Takemura, T., Tilmes, S., Tsigaridis, K., Wu, T., and Zhang, J.: Historical and future changes in air pollutants from CMIP6 models, Atmos. Chem. Phys., 20, 14547–14579, https://doi.org/10.5194/acp-20-14547-2020, 2020.

United Nations Department of Economic and Social Affairs, Population Division: World population prospects 2022: summary of results, UN DESA/POP/2022/TR/NO, 2022.

- 820 Van Vuuren, D. P., Stehfest, E., Gernaat, D. E. H. J., Doelman, J. C., Van Den Berg, M., Harmsen, M., De Boer, H. S., Bouwman, L. F., Daioglou, V., Edelenbosch, O. Y., Girod, B., Kram, T., Lassaletta, L., Lucas, P. L., Van Meijl, H., Müller, C., Van Ruijven, B. J., Van Der Sluis, S., and Tabeau, A.: Energy, land-use and greenhouse gas emissions trajectories under a green growth paradigm, Global Environmental Change, 42, 237–250, https://doi.org/10.1016/j.gloenvcha.2016.05.008, 2017.
- 825 Wang, Y., Wild, O., Ashworth, K., Chen, X., Wu, Q., Qi, Y., and Wang, Z.: Reductions in crop yields across China from elevated ozone, Environmental Pollution, 292, 118218, https://doi.org/10.1016/j.envpol.2021.118218, 2022.

Wesely, M. L.: Parameterization of surface resistances to gaseous dry deposition in regional-scale numerical models, Atmospheric Environment, 23, 1293–1304, https://doi.org/10.1016/0004-6981(89)90153-4, 1989.

Wild, O., Voulgarakis, A., O'Connor, F., Lamarque, J.-F., Ryan, E. M., and Lee, L.: Global sensitivity analysis of
 chemistry–climate model budgets of tropospheric ozone and OH: exploring model diversity, Atmos. Chem. Phys.,
 20, 4047–4058, https://doi.org/10.5194/acp-20-4047-2020, 2020.

Williams, K. D., Copsey, D., Blockley, E. W., Bodas-Salcedo, A., Calvert, D., Comer, R., Davis, P., Graham, T., Hewitt, H. T., Hill, R., Hyder, P., Ineson, S., Johns, T. C., Keen, A. B., Lee, R. W., Megann, A., Milton, S. F., Rae, J. G. L., Roberts, M. J., Scaife, A. A., Schiemann, R., Storkey, D., Thorpe, L., Watterson, I. G., Walters, D. N., West, A., Wood, R. A., Woollings, T., and Xavier, P. K.: The Met Office Global Coupled Model 3.0 and 3.1 835 (GC3.0 and GC3.1) Configurations, J Adv Model Earth Syst, 10. 357-380. https://doi.org/10.1002/2017MS001115, 2018.





Wright, L. P., Zhang, L., Cheng, I., Aherne, J., and Wentworth, G. R.: Impacts and Effects Indicators of Atmospheric Deposition of Major Pollutants to Various Ecosystems - A Review, Aerosol Air Qual. Res., 18, 1953–1992, https://doi.org/10.4209/aaqr.2018.03.0107, 2018.

Zanis, P., Akritidis, D., Turnock, S., Naik, V., Szopa, S., Georgoulias, A. K., Bauer, S. E., Deushi, M., Horowitz, L. W., Keeble, J., Le Sager, P., O'Connor, F. M., Oshima, N., Tsigaridis, K., and Van Noije, T.: Climate change penalty and benefit on surface ozone: a global perspective based on CMIP6 earth system models, Environ. Res. Lett., 17, 024014, https://doi.org/10.1088/1748-9326/ac4a34, 2022.

845 Zhang, L., Brook, J. R., and Vet, R.: A revised parameterization for gaseous dry deposition in air-quality models, Atmos. Chem. Phys., 2003.

Zhang, Y., Schaap, M. G., and Zha, Y.: A High-Resolution Global Map of Soil Hydraulic Properties Produced by a Hierarchical Parameterization of a Physically Based Water Retention Model, Water Resources Research, 54, 9774–9790, https://doi.org/10.1029/2018WR023539, 2018.

850 Zhou, X., Yue, X., and Tian, C.: Responses of Ecosystem Productivity to Anthropogenic Ozone and Aerosols at the 2060, Earth's Future, 12, e2023EF003781, https://doi.org/10.1029/2023EF003781, 2024.





Appendix A

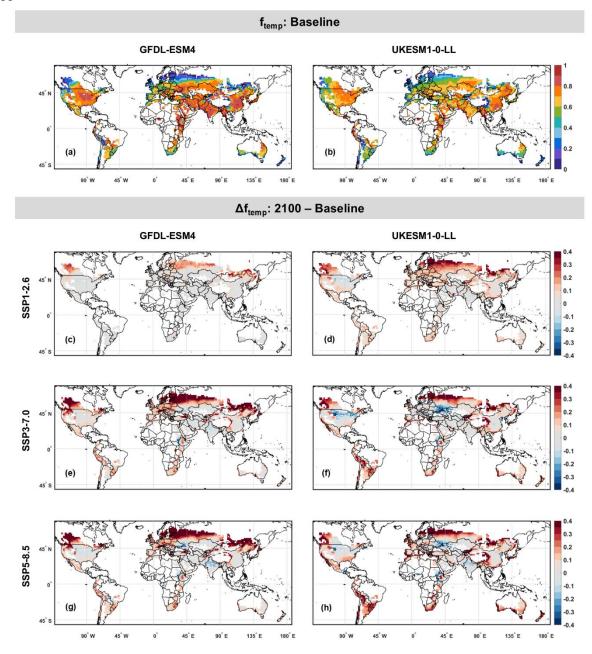
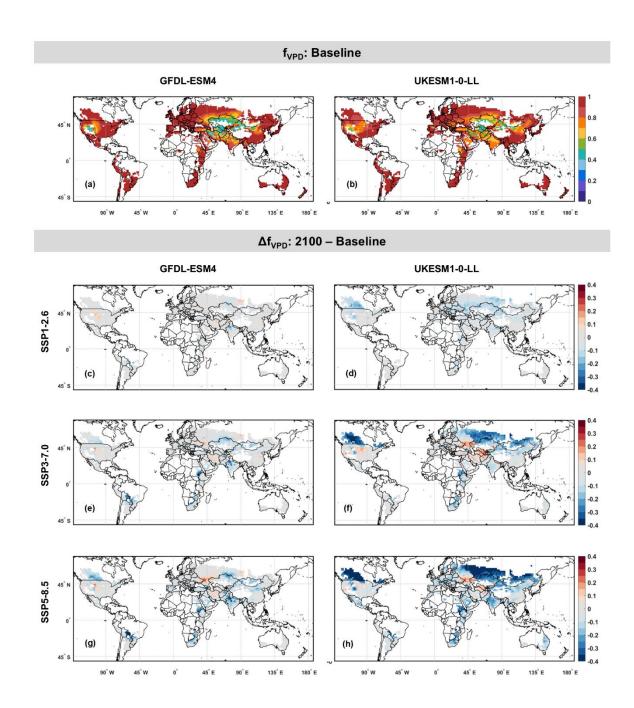


Figure A1: f_{temp} mean values over the baseline period (a,b), and Δf_{temp} at 2100 with respect to the baseline across the different SSPs for GFDL-ESM4 (c,e,g) and UKESM1-0-LL (d,f,h). f_{temp} ranges from 0 to 1 depending on whether it is limiting or not. From this study's perspective, positive or negative Δf_{temp} correspond to higher and lower O3 risk respectively (red and blue, in c-h).

Note: Baseline indicates the mean between 2000 and 2014, and 2100 indicates the mean between 2090-2099.







860 Figure A2: f_{VPD} mean values over the baseline period (a,b), and Δf_{VPD} at 2100 with respect to the baseline across the different SSPs for GFDL-ESM4 (c,e,g) and UKESM1-0-LL (d,f,h). f_{VPD} ranges from 0 to 1 depending on whether it is limiting or not. From this study's perspective, positive or negative Δf_{VPD} correspond to higher and lower O3 risk respectively (red and blue, in c-h). Note: Baseline indicates the mean between 2000 and 2014, and 2100 indicates the mean between 2090-2099.





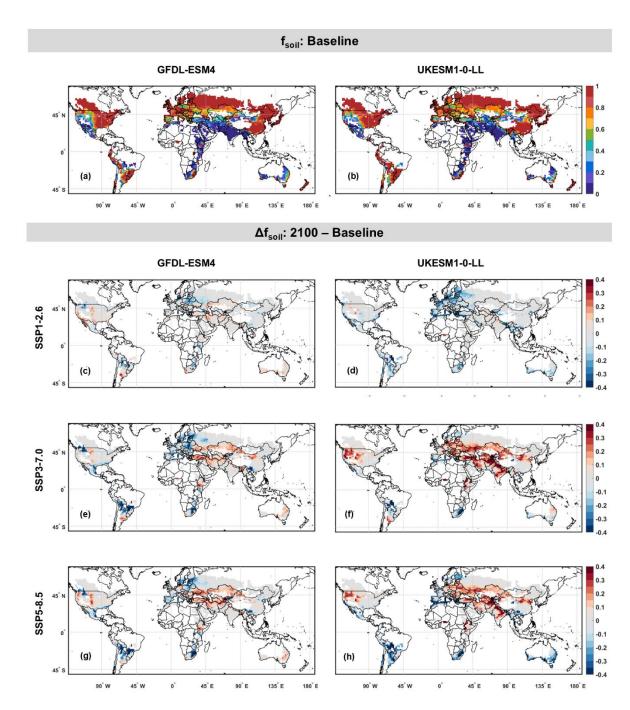
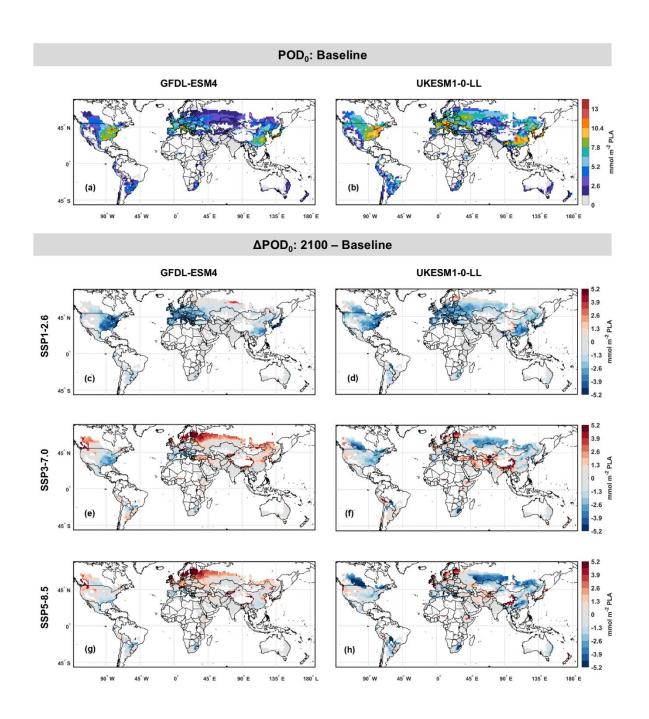


Figure A3: f_{soil} mean values over the baseline period (a,b), and Δf_{soil} at 2100 with respect to the baseline across the different SSPs for GFDL-ESM4 (c,e,g) and UKESM1-0-LL (d,f,h). f_{soil} ranges from 0 to 1 depending on whether it is limiting or not. From this study's perspective, positive or negative Δf_{soil} correspond to higher and lower O3 risk respectively (red and blue, in c-h). Note: Baseline indicates the mean between 2000 and 2014, and 2100 indicates the mean between 2090-2099.







870 Figure A4: Mean POD₀ over the baseline period (a,b), and ΔPOD₀ at 2100 with respect to the baseline across the different SSPs for GFDL-ESM4 (c,e,g) and UKESM1-0-LL (d,f,h). These results are from the rain-fed simulation.
 Note: Baseline indicates the mean between 2000 and 2014, and 2100 indicates the mean between 2090-2099.





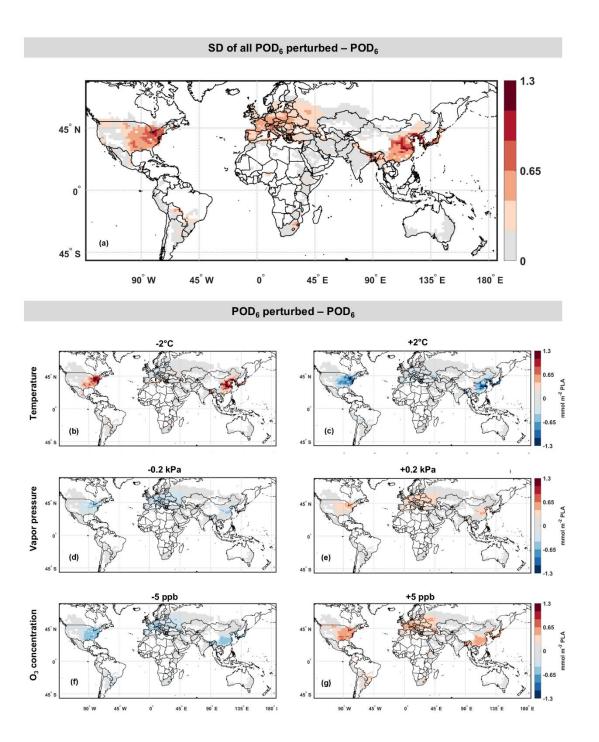


Figure A5: Standard Deviation (SD) of yearly differences of POD₆ over the baseline between all the perturbed simulations and the rain-fed simulations (a). Mean differences of POD₆ over the baseline between the perturbed simulations and the rain-fed simulations for each variable (b-g). Note: Baseline indicates the mean between 2000 and 2014.

41





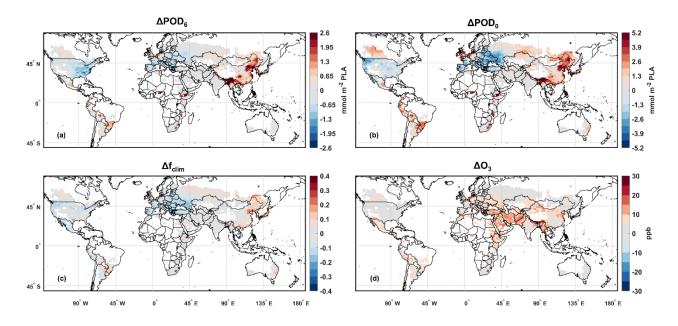


Figure A6: values at 2100 for the experiment SSP3-7.0pdSST with respect to the baseline of (a) ΔPOD_6 , (b) ΔPOD_0 , (c) Δf_{clim} , (d) ΔO_3 . These results are from the rain-fed simulation.

Appendix B

The conversion from model levels to geometric height above the ground is achieved using the hydrostatic assumption and integrating the following equation from the surface pressure to the actual pressure at the model level:

$$dz = \frac{R_d T_v(p)}{gp} dp \tag{B1}$$

- Here, *p* denotes pressure, $R_d = 287 \text{ J kg}^{-1}\text{K}^{-1}$ is the specific gas constant of dry air, $g = 9.81 \text{ m s}^{-2}$ is the acceleration due to gravity and it is assumed constant since calculations are performed for the model levels closest to the ground, and thus there is no need to incorporate geopotential height. $T_v(p)$ [K] denotes the virtual temperature profile, obtained by interpolating between model levels after converting the values from air temperature T [K] through $T_v = T(1 + 0.609133q)$, with *q* being the specific humidity [kg H₂O kg⁻¹ air].
- 890 UKESM1-0-LL explicitly indicates that the near-surface model level for O_3 concentration is at 20 m above the ground, and therefore no calculation for geometric height is required.