

# Modeling CMAQ dry deposition treatment over Western Pacific: A distinct characteristic of mineral dust and anthropogenic aerosol

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**Abstract.** Dry deposition plays a vital role in the aerosol removal process from the atmosphere. However, the chemical transport model (CTM) is sensitive to the dry deposition parameterization and yet remains to be determined due to the limited particle deposition measurement. By utilizing the CMAQv5.4 with the refined dust emission treatment, the East Asian dust (EAD) simulation during January 2023 and Spring 2021 was constructed to evaluate the performance of dry deposition parameterizations, namely S22, E20, and P22. The result showed that the dry deposition parameterization could significantly impact the CMAQ dust emission treatment. By implementing the E20 dry deposition scheme, the CMAQ simulation performance of the surface PM<sub>10</sub> has been considerably improved with the NMB of -41.9 %, as compared to the dry deposition proposed by S22 (-47.01 %) and P22 (-53.90 %). The modeled PM<sub>10</sub> pattern by E20 at the upper level (700 hPa) was mostly consistent with the observed PM<sub>10</sub> at the Lulin Atmospheric Background Station (LABS; 23.47° N, 120.87° E; 2862 m a.s.l.) where is a typical background site at Western Pacific, particularly in capturing the peak value. The correlations (R) at the high-altitude were well performed for E20 by 0.55, as compared to S22 (0.54) and P22 (0.46). Moreover, E20 improved the simulated PM<sub>10</sub> concentrations and aerosol optical depth (AOD) value over the Asian Continental during the multiple dust episodes in spring 2021, by NMB of -25.43 % and -26.19 %, respectively. The noticeable deduction of the coarse mode particle's deposition velocity ( $V_d$ ) was responsible for reducing the PM<sub>10</sub> simulation underestimation. On 22-31 January 2023, the *in-situ* measurement of the upper level observed the possibility of natural dust and anthropogenic aerosol. This is consistent with the CMAQ, which shows that both aerosol types displayed a clear "long dust-black

31 carbon belt" along the 15°N. It is revealed that the increase of wet deposition due to the surface resistivity  
32 ( $R_b$ ) leads to a significant increase in dust mass concentration but a minor increase in black carbon (BC).  
33 We proposed implementing the E20 dry deposition approach, narrowing the uncertainty of the CMAQ  
34 dust emission treatment.

## 35 **1 Introduction**

36 The chemical transport model (CTM) is a powerful tool for comprehending air pollution, encompassing  
37 emission, transport, radiative impact, and removal mechanisms at various grid scales. Among  
38 these, particle dry deposition is a crucial aerosol removal process and an important sink for particles in  
39 the model. The derivation of the dry deposition is based on the resistance framework and  
40 electrical analogue, but its implementation can vary across models (Wesley, 1989; Giardina and Buffa,  
41 2018; Gaydos et al., 2007; Khan and Perlinger, 2017; Shu et al., 2017). A key challenge in dry deposition  
42 simulation is the scarcity of measurement data for model verification, underscoring the necessity for  
43 further research to enhance the accuracy of air quality modeling.

44 An immense range of dry deposition parameterization has been implanted in the model. The  
45 deposition mechanism by Slinn (1982) includes the deposition process such as turbulent transfer,  
46 Brownian diffusion, impaction, interception, gravitational settling, and particle rebound, where the  
47 particle grows under humid conditions. Zhang et al. (2001) suggested the dry deposition scheme is  
48 sensitive to land use category and several parameters. For instance, due to the particle growth, the  
49 deposition velocity ( $V_d$ ) over the ocean is much higher than on another land surface, as the  $V_d$  increased  
50 rapidly with the increase of particle size. Some CTMs using Zhang et al. (2001) parameterization still  
51 underestimated the global  $PM_{2.5}$  concentration. The latest dry deposition scheme revision by Emerson et  
52 al. (2020) based on the flux measurement of grassland and pine forest has reduced the uncertainty,  
53 marking a significant step forward in our quest for more accurate air quality modeling.

54 An updated deposition scheme that reduces the dependence of the deposition velocity on the aerosol  
55 mode width has been proposed (Shu et al., 2022). Indeed, the approach suggested that vegetation  
56 dependence increased the  $V_d$  for submicrons and decreased for large particles by 37 % and -66 %,

57 respectively. It also reduced the functional biases by 56-97 % for vegetated land-use type and equivalence  
58 performance over the water. Moreover, adding the second inertial impaction term for microscale obstacles  
59 such as leaf hairs, microscale ridges, and needle leaf edge effects managed to increase the mass dry  
60 deposition of the accumulation mode aerosols in the model (Pleim et al., 2022). These modifications  
61 reduced the average  $PM_{2.5}$  in the atmosphere during July 2018 over the contiguous United States.

62 With a plethora of deposition approaches in use, it becomes paramount to comprehend their impact  
63 on model performance in predicting aerosol behaviour. The surface fine particle concentrations can vary  
64 up to 5-15 %, and the particle dry deposition has more than 200 % discrepancy due to the different dry  
65 deposition schemes. (Saylor et al., 2019). A comprehensive evaluation of five different parameterizations  
66 has been conducted, with the simplest and most effective deposition mechanism suggested for the CTM  
67 (Khan and Perlinger, 2017). However, the model's reliance on meteorological factors such as frictional  
68 velocity, relative humidity, rainfall, or wind speed, which can significantly influence  
69 the model's accuracy, remains a challenge (Kong et al., 2021).

70 Besides the model bias on  $PM_{2.5}$ , the simulation of  $PM_{10}$  has been underestimated due to the  
71 uncertainty of the deposition mechanism, particularly over the western Pacific (Kong et al., 2021). The  
72  $V_d$  is overestimated for coarse particles, where the dry deposition velocity is too high for coarse particles  
73 when the frictional velocity is large, which is why the surface  $PM_{10}$  concentration is underestimated (Ryu  
74 and Min, 2022). The model performance of  $PM_{10}$  simulation that is widely influenced by the dust  
75 treatment embedded within CMAQ has been revised (Dong et al., 2016; Liu et al., 2021; Kong et al.,  
76 2021, 2024) and are found to effectively simulate the  $PM_{10}$  over the western Pacific region such as  
77 Taiwan. However, the issue regarding the deposition algorithm's impact on the model performance at the  
78 corresponding region needs to be discussed. The present research intends to evaluate the CMAQ model  
79 performance due to the different deposition schemes on aerosols in the Taiwan region.

80 The model performance in Taiwan is paramount in our study, as the area is equipped with a substantial  
81 number of well-maintained surface observation sites, providing comprehensive coverage. The LABS  
82 station in the high-altitude subtropical western North Pacific region serves as the sole background station

83 for monitoring transboundary pollutants. This station is crucial in our research as it provides unique data  
84 on the long-range transport of pollutants, further underscoring the relevance of our study.

85 The transboundary pollutants mechanisms have been widely discussed through LABS measurements,  
86 cooperating with the backward trajectory, reanalysis dataset, and modeling approach. Previous research  
87 reveals that LABS pollutants could be associated with severe fire emissions from northern Peninsular  
88 Southeast Asia (Huang et al., 2020; Ooi et al., 2021) and Indonesia (Ravindra Babu et al., 2023).  
89 Moreover, the intense wind speed in northwest China could transport the mineral dust through the surface  
90 and high-altitude layer detected at LABS (Kong et al., 2021; Kong et al., 2022). Additionally, the transport  
91 process of East Asian haze due to the cold surge from the Asian Continental industrial region towards  
92 Taiwan has been widely discussed (Chuang et al., 2020). Instead of pure aerosol, the coexistence of dust  
93 and biomass burning over Taiwan, a condition discovered in previous research, has significant  
94 implications for the regional climate (Dong et al., 2018; Dong et al., 2019). However, the high-altitude  
95 synoptic pattern associated with the coexistence between natural dust and anthropogenic pollutants  
96 remains unknown due to a lack of observations at the upper layers.

97 This study used the chemical transport model to investigate the long-range transport of East Asian  
98 dust (EAD) that occurred on 22-31 January 2023 and 12 March-20 April 2021. Due to the limitation of  
99 the dust model, the CMAQ version 5.4, embedded with three types of dry deposition schemes, was  
100 implemented to justify the effectiveness of improving our latest refined dust model (Kong et al., 2024).  
101 The dry deposition scheme proposed by Shu et al. (2022) has reduced certain model bias as compared to  
102 the base scheme. However, the revised scheme response to the natural phenomenon such as wind-blown  
103 dust has not being tested. In the other way, the number of concentrations of the large size particle has  
104 been decreased over land, and increased over ocean area globally by the adjusted collective efficiency  
105 proposed by Emerson et al. (2020). Pleim et al. (2022) has included the consideration of white cap effect  
106 which dependent on wind speed and sea surface temperature into the dry deposition scheme. Hence, the  
107 response of the CMAQ dust model under the newly developed dry deposition schemes are worth  
108 investigating in reducing the model uncertainty.

109 LABS detected the recent transboundary episode in January 2023 as a mixing aerosol type (see  
110 Section 3.1), which has not been widely discussed, and the multiple dust storm episodes mentioned by  
111 Kong et al. (2024) provide an opportunity to model the EAD over the downwind region. Recognizing the  
112 significant transboundary events detected through Taiwan’s observations, the improvement of the CMAQ  
113 dust model by the dry deposition schemes, and its application in characterizing the transport mechanism  
114 can be vital. The paper is organized as follows. The model setup and ancillary datasets are discussed in  
115 Sect. 2. The results and discussion are presented in Sect. 3, followed by the conclusions in Sect. 4.

## 116 **2 Data and Methodology**

### 117 **2.1 Dust emission treatment**

118 Before delving into the details, it's important to understand the process of dust transport. Dust is primarily  
119 transported by wind through a process known as sandblasting (Kok et al., 2012). For dust to be uplifted,  
120 the horizontal wind speed must exceed a certain threshold frictional velocity ( $u_{*,t}$ ), which is estimated by  
121 the model as follows:

$$122 \quad u_{*,t} = u_{*,to} f_m f_r \quad (1)$$

123 Where  $u_{*,to}$  is the ideal threshold friction velocity, while  $f_m$  and  $f_r$  are the correction factors of soil  
124 moisture and surface roughness, respectively.

125 Through a collaborative effort, the windspeed, soil texture, soil moisture, and surface roughness  
126 length derived from field and laboratory studies have been integrated into the windblown dust treatment,  
127 which is now a part of the Community Multiscale Air Quality (CMAQ) modeling system (Foroutan et  
128 al., 2017). This model, developed and evaluated over the continental United States, has also been extended  
129 to the East Asia region (Dong et al., 2016; Liu et al., 2021; Kong et al., 2021, 2024). Kong et al. (2024)  
130 have proposed further improvements, including the integration of the revised soil moisture fraction, dust  
131 emission speciation profile, and bulk soil density, to enhance the representation of the Asian dust  
132 simulation. This ongoing collaboration is crucial for the continuous improvement of our understanding  
133 and management of dust emissions.

## 134 **2.2 Particle dry deposition schemes**

135 Particle dry deposition is a complex process relating to the deposition velocity, particle size, source and  
136 composition, land use surface, and meteorological condition. Generally, the flux of the particle mass  
137 through the surface boundary layer is estimated as:

$$138 \quad F = C \times V_d \quad (2)$$

139 where  $F$  is the deposition flux,  $C$  is the particle concentration at the surface layer, and  $V_d$  is the deposition  
140 velocity.

141 The difference in the particle concentration and deposition prediction among the various  
142 atmospheric chemistry models was probably due to the algorithm of the dry deposition particle. The  
143 algorithm describing particle deposition velocity as a function of particle size in almost all current air  
144 quality model systems is descended from Slinn (1982). The particle deposition according to vegetative  
145 canopies formulated the deposition velocity as:

$$146 \quad V_d = V_g + \frac{1}{R_a + R_b} \quad (3)$$

147 where  $V_g$  is the gravitation settling velocity,  $R_a$  is the resistivity aerodynamic,  $R_b$  is the surface resistivity,  
148 also known as quasi-laminar sub-layer resistivity in STAGE. The  $V_s$  is calculated according to Stokes's  
149 Law as:

$$150 \quad V_g = \frac{\rho_p D_p^2 g C_c}{18\eta} \quad (4)$$

151 where,  $\rho_p$  is the density of the particle;  $D_p$  is the diameter of the particle;  $g$  is gravitational acceleration;  
152  $C_c$  is the Cunningham correction factor for small particles; and,  $\eta$  is the dynamic viscosity of air.

153 CMAQ is embedded with M3Dry dry deposition calculation that implements the scheme of Pleim  
154 and Ran (2011), which is based on Slinn (1982). As noted by Pleim and Ran (2011), chemical surface  
155 flux modeling has become an essential process in the air quality model. For instance, the linkages of

156 ambient concentration levels to the deposition of SO<sub>x</sub> and NO<sub>x</sub>. Moreover, Surface Tiled Aerosol and  
157 Gaseous Exchange (STAGE) deposition has been implemented within the CMAQv5.3, where estimated  
158 fluxes from sub-grid cell fractional land-use values, aggregate the fluxes to the model grid cell and unifies  
159 the bidirectional and unidirectional deposition schemes using the resistance framework (Massad et al.,  
160 2010; Nemitz et al., 2001). The updated STAGE version in CMAQv5.4 could aggregate the grid-scale  
161 values that match the grid-scale values from most kinds of Land Surface Model of WRF (Hogrefe et al.,  
162 2023). Since the present study is primary focused on the impact of dry deposition scheme on CMAQ dust  
163 model, the simulations with the STAGE module are the mandatory concern.

### 164 **2.3 CMAQ model design**

165 This study applied WRF v4.0 for the meteorological field parameters and CMAQv5.4 to simulate the  
166 transboundary East Asian dust episodes on 22-31 January 2023, and the multiple dust storm episodes  
167 during 12 March-20 April 2021. The modeling domain was set up to cover the Taklamakan and Gobi  
168 Desert, with a resolution of 45 km, and nested towards Taiwan at a resolution of 15 km (d02) and 5 km  
169 (d03) (Fig .1, Table 2). Also, as Taiwan is influenced by biomass burning, the domain covers up to the  
170 peninsular Southeast Asia (PSEA), which will be carried out in the future (Ooi et al., 2021). The model  
171 consisted of 40 vertical layers, with eight layers below ~1 KM altitude, 13 layers below ~3 KM altitude,  
172 and 27 layers covering the upper layer to ~21 KM. The model's initial and lateral boundary  
173 conditions were constructed using the National Centers for Environmental Prediction (NCEP) Final  
174 Analyses (FNL) reanalysis dataset on a 0.5° × 0.5° grid. The data assimilation was conducted by grid  
175 nudging in all the domains. The CB06 gas-phase chemical mechanism and the AERO7 aerosol module  
176 model were implemented in CMAQ for the present study.

177 The anthropogenic emission inventories in East Asia, crucial for our research, were obtained from  
178 the MICS-Asia (Model Inter-Comparison Study for Asia) Phase III emission inventory (Li et al., 2017).  
179 The emissions of SO<sub>2</sub>, NO<sub>x</sub>, NMVOC, NH<sub>3</sub>, CO, PM<sub>10</sub>, PM<sub>2.5</sub>, BC, OC and CO<sub>2</sub> has been meticulously  
180 modified, taking into account of the relative changes in China's anthropogenic emissions between 2010  
181 and 2017 (Zheng et al., 2018). Additionally, the modified emission of NO<sub>2</sub> was adjusted further by the  
182 satellite imagery OMI-NO<sub>2</sub> in January 2023 (Huang et al., 2021). Biogenic emissions for Taiwan were

183 prepared by the Biogenic Emission Inventory System version 3.09 (BEIS3, Vukovich and Pierce, 2002)  
184 and, for regions outside Taiwan, by the Model of Emissions of Gases and Aerosols from Nature v2.1  
185 (MEGAN, Guenther et al., 2012). TEDS 10.0 (Taiwan Emission Database System, TWEPA, 2011;  
186 <https://erdb.epa.gov.tw/>, last access: 18 January 2024) was used for domain 3 (d03). To ensure the  
187 precision of the multiple dry deposition parameterizations, the present research conducted five simulation  
188 scenarios, namely CMAQ\_Off\_S22, CMAQ\_Dust\_S22, CMAQ\_Dust\_E20 and CMAQ\_Dust\_P22. The  
189 CMAQ\_Off\_S22 scenario did not include the inline dust calculation (Table 2). Meanwhile, the latest  
190 refined integrated dust treatment was implemented in the CMAQ\_Dust\_S22 scenario (Kong et al., 2024).  
191 Indeed, both CMAQ\_Off\_S22 and CMAQ\_Dust\_S22 used the dry deposition mechanism by Shu et al.  
192 (2011). The dry deposition mechanism of Emerson et al. (2020) and Pleim et al. (2022) were implemented  
193 in CMAQ\_Dust\_E20 and CMAQ\_Dust\_P22 scenarios, respectively.

194  $V_d$  over the ocean surface has been shown to influence the CTMs in simulating aerosol,  
195 particularly  $PM_{10}$ . The modeled  $PM_{10}$  can be increased by reducing  $V_d$  by a factor of 10 based on the bare  
196 soil measurement (Tav et al., 2018; Ryu et al., 2022). However, the adjusted  $V_d$  in estimating the aerosol  
197 was too coarse. In Eq. (3), the parameterizations of  $R_a$  and  $R_b$  determined the magnitude of  $V_d$ . P22 dry  
198 deposition scheme in CMAQv5.4 includes the white-cap effect over the ocean surface, which is related  
199 to the particle collection efficiency by impaction, as a function of  $R_b$  (Pleim et al., 2022). The impact of  
200 the white cap can increase as the wind speed increases, which can be an essential parameterization in  
201 simulating transboundary events (Albert et al., 2020). Our findings on dust transport, which was highly  
202 related to turbulence, have significant implications for future research and modeling, inspiring further  
203 exploration and innovation (Zhang et al., 2022). In CMAQv5.4,  $R_b$  is estimated separately by vegetation  
204 and non-vegetation type.  $R_b$  at the smooth surface (non-vegetation) is related to the surface resistivity of  
205 the bare soil and the water layer. Since EAD aerosol particles are mostly uplifted from the bare soil surface  
206 layer and the aerosol deposition at the marine boundary layer (Kong et al., 2021), the sensitivity of  $R_b$  at  
207 the smooth surface impact on the CMAQ dust model simulation can be vital. By assuming  $R_b = 1/V_d$ ,  $V_d$   
208 is inversely dependent on  $R_b$ , we increased  $R_b$  to a factor of 10 as P22E01. To further carry out the  
209 sensitivity test, we scale  $R_s$  by 50 and 100, for P22E02 and P22E03, respectively (Table 2).

## 210 **2.4 Ancillary dataset**



211 PM<sub>10</sub> (particulate matter  $\leq 10 \mu\text{m}$  in aerodynamic diameter) and PM<sub>2.5</sub> (particulate matter  $\leq 2.5 \mu\text{m}$  in  
212 aerodynamic diameter) concentrations during the dust events in January 2023 were obtained from Lulin  
213 Atmospheric Background Station (LABS; 23.47° N, 120.87° E, 2862 m MSL) and Cape Fuguei (25.30°  
214 N, 121.54° E, 10 m MSL). In addition, the hourly PM<sub>10</sub> and PM<sub>2.5</sub> of nearly 100 sites distributed over  
215 mainland China (Fig. S1), covering the period of 12 March-20 April 2021, obtained from the Chinese air  
216 quality online monitoring analysis platform's website ([www.aqistudy.cn/](http://www.aqistudy.cn/)). The Modern Era  
217 Retrospective-analysis for Research and Application version 2 (MERRA-2) reanalysis data was used to  
218 demonstrate the spatiotemporal distribution of dust, compare with the air quality model, irrespective of  
219 the influence of clouds. MERRA-2 (Gelaro et al., 2017) is a NASA reanalysis utilizing Goddard Earth  
220 Observing System Data Assimilation System Version 5 (GEOS-5) and covering the data assimilated  
221 system at a native spatial resolution of  $0.5^\circ \times 0.625^\circ$ . Also, Moderate Resolution Imaging  
222 Spectroradiometer (MODIS) Terra satellite images and the level-3 MODIS AOD at 550 nm  
223 (MYD08) were obtained from the U.S. National Aeronautics and Space Administration  
224 (<https://worldview.earthdata.nasa.gov/>).

## 225 **3 Results and Discussion**

### 226 **3.1 Observed air quality and weather conditions**

227 Figure 2 shows the dust outbreak over East Asia, displayed by the MODIS Terra sensor and MODIS  
228 AOD at 550 nm from 22-31 January 2023. The satellite image showed dust induced by a high-pressure  
229 system on 24-25 January (Fig. 2a3, 2a4). The next day, the same region was covered by a thick cloud,  
230 and dust was again widely distributed from 27-30 January 2023. Using MODIS AOD to verify the dust  
231 plume (Han et al., 2012; Kong et al., 2021), the dust plume was distributed in Central China and northern  
232 Taiwan on 24 January 2023. Moreover, the most intense dust plume in the eastern China and East China  
233 Sea region was observed on 27 January. Fig. S2 shows the synoptic weather map across the study domain.  
234 On 22-23 January, the southward high-pressure system was responsible for pushing the pollutant across  
235 the Asian Continent, which is consistent with Chuang et al. (2018) and Kong et al. (2021, 2022, 2024)  
236 (Fig. S1a-b). The high-pressure system that moved southward will then move eastward toward the western

237 Pacific Ocean (Fig S1c-d). Meanwhile, the high-pressure system on the northwest side again expands in  
238 the southeast direction. The second high-pressure system again pushed the pollutant for the second time  
239 and caused the high pollutant problem on 27 January.

240 The impact of East Asian dust on the air quality over the high-altitude western Pacific region was  
241 widely discussed (Kong et al., 2022). Two interesting high pollution events at Mt. Lulin (2,862 m above  
242 sea level) during 24-26 Jan and 27-30 January, respectively, are shown in Fig. 3. The latter event was  
243 more intense compared to the earlier one, where the maximum PM<sub>10</sub> concentration can reach up to 35 µg  
244 m<sup>-3</sup>. Moreover, it was observed that the BC concentrations could reach up to a maximum of 400 ng m<sup>-3</sup>.  
245 Based on the *in-situ* measurement, it was interesting to find the mixing state between dust, BC, and brown  
246 carbon (Fig. 3c). Different from what has been discussed by Kong et al. (2022), the long-range transport  
247 air pollution at the high-altitude not just merely EAD, but also included the anthropogenic pollutant from  
248 mainland China.

### 249 **3.2 Evaluation of CMAQ dust emission and dry deposition parameterizations**

250 Table 3 shows the statistical analysis of PM<sub>10</sub> and PM<sub>2.5</sub> concentrations over Cape Fuguei (northern  
251 Taiwan) from 22-31 January under the multiple deposition mechanisms. The threshold of the statistical  
252 index is based on Emery (2001). CMAQ\_Off\_S22, the PM<sub>10</sub> simulation presented without the inline dust  
253 calculation, recorded the normalized mean bias (NMB) of -52.81 %. CMAQ\_Dust\_S22 improved the  
254 simulation over Cape Fuguei (northern Taiwan) by -47.01 % as we included the refined dust treatment  
255 (Kong et al., 2024). However, the improvement is insignificant due to the weak intensity dust episodes  
256 and the limitation due to the excessive deposition mechanism within the model (Kong et al., 2021). Hence,  
257 we expanded the sensitivity simulation to examine the impact of the deposition algorithm on the aerosol  
258 prediction. CMAQ\_Dust\_E20 simulations utilizing the Emerson et al. (2020) approach increased the  
259 modeled PM<sub>10</sub> simulation by NMB of -41.9 %.

260 Instead of PM<sub>10</sub> simulation, the present study found that the inline dust treatment and deposition  
261 algorithms could influence PM<sub>2.5</sub> simulation performances. For instance, the modeled PM<sub>2.5</sub> improved  
262 from -12.63 % (CMAQ\_Off\_S22) to -8.84 % (CMAQ\_Dust\_S22). Meanwhile, the deposition algorithm

263 embedded in CMAQv5.4 has recorded modeled  $PM_{2.5}$  by -10.65 % and -15.22 % under  
264 CMAQ\_Dust\_E20 and CMAQ\_Dust\_P22, respectively. This incident suggested that the East Asian dust  
265 from northwest China transported to the Western Pacific Ocean could also carry the anthropogenic  
266 emission of East China.

267 Figure 4 shows the time series of hourly  $PM_{10}$  and  $PM_{2.5}$  concentrations over Cape Fuguei  
268 (northern Taiwan) and LABS (high altitude region) from 22-31 January 2023 under the multiple  
269 deposition mechanisms. Generally, all the patterns of  $PM_{10}$  simulations were consistent with the observed  
270  $PM_{10}$ , especially in capturing the peak value. For instance, the maximum observed (CMAQ\_Dust\_E20)  
271  $PM_{10}$  concentrations at the surface during Jan 24 and 27 were 141 (102.6)  $\mu g m^{-3}$  and 114 (163.2)  $\mu g m^{-3}$ ,  
272 respectively. A similar time-series pattern was found for the  $PM_{2.5}$  simulation (Fig. 4b).

273 More importantly, the CMAQ model performance over the high-altitude region needed to be  
274 carried out and discussed. The biomass-burning episode of the northern PSEA over Mt. Lulin has been  
275 finely correlated by plume rise injection (Chuang et al., 2016; Ooi et al., 2021). From Fig. 4c, the modeled  
276  $PM_{10}$  pattern for CMAQ\_Off\_S22 could not correlate well with observed  $PM_{10}$  over Mt. Lulin, with a  
277 poor correlation of 0.30. The correlation was increased for CMAQ\_Dust\_S22 (0.54), CMAQ\_Dust\_P22  
278 (0.46), and primarily well performed for CMAQ\_Dust\_E20 (0.55). The modeled result  
279 was somehow consistent with the surface  $PM_{10}$  simulation at Cape Fuguei. The high observed  $PM_{10}$   
280 episodes during 27-28 January with a maximum of 34.5  $\mu g m^{-3}$  was only 53.3 % higher than  
281 CMAQ\_Dust\_E20 of 22.5  $\mu g m^{-3}$ . For the CMAQ  $PM_{2.5}$ , the simulation generally underestimated the  
282 observed  $PM_{2.5}$ .

283 During the spring of 2021, a series of dust storms (15 March, 27 March, and 18 April) occurred  
284 over the Gobi area, with one of the most significant dust storms in the past decade (15 March, the “3.15”  
285 dust storm hereafter) causing environmental impact over the continental (Jin et al., 2022; Gui et al., 2022;  
286 He et al., 2022; Liang et al., 2022; Tang et al., 2022). More interestingly, one of the multiple dust storm  
287 episodes reached western Pacific Ocean due to the extreme typhoon episode (Kong et al., 2024). Hence,  
288 we intend to re-emphasize the precision of various deposition schemes on the CMAQ for the recent dust

289 storm episode over the Asian Continental highlighted by Kong et al. (2024). We evaluated the CMAQ  
290 simulations with the different dry deposition schemes for the 40-day sensitivity test on 12 March-20 April  
291 2021 against measured  $PM_{10}$  and  $PM_{2.5}$  concentrations across the observation sites in mainland China  
292 (Table 4). The observation sites used for the model comparison are marked in Fig. S1. Generally, the  
293 evaluation results for Taiwan and mainland China were consistent. During the 40 days of Spring 2021,  
294 the CMAQ  $PM_{10}$  of NMB was the highest for Off\_S22 (NMB = -75.00 %), followed by Dust\_S22 (-45.97  
295 %). The latest inline dust emission scheme embedded with E20 dry deposition scheme for  $PM_{10}$  was well  
296 performed by NMB of -25.43 %, compared to the Dust\_P22 (-59.82 %). For the  $PM_{2.5}$  simulation,  
297 Dust\_S22 has been improved from Off\_S22, and Dust\_S22 was slightly better than Dust\_E20 and  
298 Dust\_P22.

299 Figure 5 shows the scatter plot of simulated and observed PM across mainland China. The  
300 correlation coefficient (R), a factor of two (FAC2), and the mean observed and simulated PM are marked  
301 in Figure 5. The modeled  $PM_{10}$  without the dust scheme had the lowest correlation. Among all of these  
302 simulations, Dust\_E20 performed the best correlation ( $R > 0.3$ ) compared to Dust\_S22 and Dust\_P22.  
303 However, for  $PM_{2.5}$ , the correlation between the model and measured values was similar for all the dry  
304 deposition schemes. The statistical index of FAC2 was used in the present work since either low or high  
305 outliers less influence it (Chan and Hanna, 2004). The dataset is reliable for FAC2 values between 0.5  
306 and 2.0, with the ideal model of 1.0. The simulated  $PM_{10}$  by Dust\_E20 performed well, with a nearly  
307 perfect value of 1.1. Meanwhile, the  $PM_{2.5}$  by Dust\_S22 simulation was slightly better than Dust\_E20 but  
308 much better than the other experiments.

309 The comparison of AOD between CMAQ and MODIS for the three dust storm episodes: 14-16  
310 March 2021 (“3.15” dust storm), 26-28 March 2021 (“3.27” dust storm), and 17-19 April 2021 was shown  
311 (“4.18” dust storm) (Table 4). Overall, CMAQ Dust\_E20 above 30°N has evaluated well the MODIS  
312 AOD by NMB of -26.2 %, as compared to S22 (-32.0 %) and P22 (-35.8 %). The CMAQ AOD by E20  
313 during the most intense Super Dust Storm in 3.15 has significantly improved over northern China, the  
314 dust source region, as shown in the red dash rectangular box (Fig. S3). Additionally, the modeled AOD  
315 by E20 over the western Pacific Ocean (shown in red dash rectangular box) increased in episode 4.18,

316 reporting a value of 0.7 compared to 0.5 by S22. Significantly, the E20 deposition scheme has primarily  
317 enhanced the PM<sub>10</sub> prediction over the marine boundary layer, addressing the model uncertainty due to  
318 the typhoon mentioned by Kong et al. (2024) and demonstrating the practical implications of our research.

319 The present work is consistent with the dust scheme in the WRF-Chem, where the dust loading is  
320 very sensitive to the dry deposition schemes and dust emission schemes, especially over the downwind  
321 region (Zeng et al., 2020). Fig. 6 shows the CMAQ estimated averaged mean PM<sub>10</sub> and PM<sub>2.5</sub> in January  
322 2023 and Spring 2021 for the Off\_S22 and its corresponding change by Dust\_S22, Dust\_E20, and  
323 Dust\_P22, respectively. Generally, the spatial distribution of the high PM<sub>10</sub> concentrations by > 50 µg m<sup>-3</sup>  
324 was distributed over northwest China, which is the dust source region's location, consistent with the  
325 simulation suggested by Kong et al. (2021, 2022, 2024). Such high particulate matter dissipated to east  
326 China, indicating the transport pathway in the southeastern direction towards the western Pacific (Fig. 6a,  
327 h). The larger PM<sub>10</sub> distribution by E20 than S22 and P22 over northwest China, meaning E20  
328 successfully increased the PM<sub>10</sub> concentrations. Another fascinating fact about E20 was that the PM<sub>10</sub>  
329 increased over the southern South China Sea (Fig. 6b). For the modeled PM<sub>2.5</sub> concentrations, the high  
330 concentration was distributed over the Asian Continental under all dry deposition mechanism.

### 331 **3.3 Impact on the CMAQ ambient particle concentrations**

332 Figure 7 shows the boxplot of the averaged simulated V<sub>d</sub> for the Aitken, accumulation, and coarse  
333 particles modes under multiple deposition schemes in January 2023 (S22\_2023, E20\_2023, and  
334 P22\_2023) and in Spring 2021 (S22\_2021, E20\_2021, and P22\_2021). These different dry deposition  
335 treatments have a substantial impact on the aerosol profile, altering the ambient total dry deposition  
336 regionally. For instance, the median deposition velocity of S22\_2023, E20\_2023, and P22\_2023 of the  
337 Aitken (accumulation) modes particle were 0.069 (0.020) cm s<sup>-1</sup>, 0.039 (0.014) cm s<sup>-1</sup> and 0.034 (0.029)  
338 cm s<sup>-1</sup>, respectively. The E20 simulation median V<sub>d</sub> decreased by -12.65 % for coarse-mode particles  
339 compared to S22. Also, the 75<sup>th</sup> percentile V<sub>d</sub> of the coarse mode has been significantly reduced by -32.13  
340 %. On the other hand, P22 showed a different simulation by the median V<sub>d</sub> increment of 71.38 %. These  
341 findings suggest that the choice of dry deposition treatment can significantly influence the distribution  
342 and concentration of aerosols in the atmosphere, with potential implications for air quality and climate.

343 As shown in Figure 7, the results during the spring of 2021 are similar to those for the 2023 results  
344 for the median deposition velocity. For instance, the  $V_d$  Aitken, accumulation and coarse mode for  
345 E20\_2023 (E20\_2021) were 0.039 (0.039)  $\text{cm s}^{-1}$ , 0.014 (0.012)  $\text{cm s}^{-1}$  and 0.20 (0.19)  $\text{cm s}^{-1}$ ,  
346 respectively. The result was consistent with the best simulated  $\text{PM}_{10}$  by E20 in 2023 and 2021 displayed  
347 in Table 3 and 4, respectively. The lowest  $V_d$  of the coarse mode particle was responsible for reducing  
348 the  $\text{PM}_{10}$  simulation underestimation of S22 and P22, consistent with the simulation by Ryu and Min  
349 (2022). The slow  $V_d$  means the total loss of aerosol to the surface has been minimized, leading to  
350 increased aerosol concentration. These implications are crucial for understanding the behaviour of  
351 aerosols in the atmosphere and their impact on air quality.

352 We estimated the CMAQ averaged particle modes for the S22\_2023, E20\_2023, and P22\_2023  
353 dry deposition scheme (Fig. 8). For S22\_2023, we found that high  $V_d$  corresponding to the Aitken and  
354 accumulation modes distributed mainly over most of the CMAQ domain, which was most evident over  
355 Asian continent (Fig 8a, 8d). Meanwhile, the magnitude of  $V_d$  distribution was the most significant over  
356 the western Pacific Ocean by S22\_2023 and the least for E20\_2023 (dash rectangular box in Fig. 8d, e,  
357 f). For the coarse mode particles, the  $V_d$  was the lowest for E20\_2023 compared to S22\_2023 and  
358 P22\_2023, particularly over the ocean area near northeast China, Japan, and Korea (white-dash  
359 rectangular box in Fig. 8d, e, f). This leads to a significant deposition over the downwind region, causing  
360 less  $\text{PM}_{10}$  simulated by P22\_2023 and S22\_2023 than E20\_2023. A previous study proposed the  $V_d$  for  
361 the aerosol at the water surface was associated with the CTM uncertainty at the downwind region (Kong  
362 et al., 2021, 2024; Ryu and Min, 2022). The  $V_d$  at land surface was generally higher than at water surfaces.  
363 Interestingly, the coarse mode  $V_d$  at the water surface for E20\_2023 (0.060  $\text{cm s}^{-1}$ ) was lower than  
364 S22\_2023 (0.085  $\text{cm s}^{-1}$ ) and P22\_2023 (0.116  $\text{cm s}^{-1}$ ), respectively, suggesting that E20\_2023 deposition  
365 schemes could minimize the excessive deposition over the marine boundary layer (Table 5). Such minimal  
366 deposition velocity distributing over a large part of the western Pacific Ocean, including the Sea of Japan,  
367 Yellow Sea, East China Sea, and South China Sea, might be responsible for reducing the modeled  $\text{PM}_{10}$   
368 underestimation over Taiwan (Fig. 8h), as mentioned by Kong et al. (2021).

369 To better understand the behavior of the  $V_d$  during the 40-day simulation of Spring 2021  
370 corresponding to the aerosol simulation, we visualized the CMAQ averaged particle modes for the  
371 S22\_2023, E20\_2023, and P22\_2023 dry deposition scheme (Fig. 9). The  $V_d$  of the coarse mode particles  
372 for E20\_2021 was the lowest among the others over the ocean area, which shows similarity as E20\_2023  
373 (Fig. 9g, h, i). As mentioned by Kong et al. (2024), one of the continuous EAD episodes was related to  
374 the typhoon. The strong wind speed and extreme precipitation due to the intense anticyclonic system  
375 caused nearly zero dust simulation. In S22\_2021 ( $0.060 \text{ cm s}^{-1}$ ) and P22\_2021 ( $0.070 \text{ cm s}^{-1}$ ), the model  
376 suggested high coarse mode  $V_d$  at the western Pacific Ocean. In E20\_2021, the  $V_d$  ( $0.053 \text{ cm s}^{-1}$ ) is lower  
377 than the rest of the dry deposition mechanism, particularly the area affected by typhoon (black-dash  
378 rectangular box). This means that the E20 dry deposition has reduced the uncertainty of the excessive  
379 dust loss at the marine boundary layer. Figure 6 (g, h, i) shows more simulated mineral dust at the western  
380 Pacific by E20 than S22 and P22 during the spring of 2021.

### 381 **3.4 CMAQ of dust and black carbon synoptic pattern at the upper level**

382 Black carbon (BC), often known as elemental carbon, released from the biofuels, fossil fuels and biomass  
383 burning, has been proven to impact the radiative budget and regional climate (Ramanathan, V and  
384 Carmicheal, 2008; Pani et al., 2016, 2020). In the meantime, China has been a significant contributor to  
385 global anthropogenic BC emission, particularly in the cities of the northern part (Xiao et al., 2023; Wang  
386 et al., 2024). During the severe dust episodes in the spring of 2023, the contribution of black carbon  
387 brought by EAD was captured in North China (Wang et al., 2024). As depicted in Fig. 2, the  
388 transboundary episode observed in the upper level of Taiwan during this event could be the mixing of the  
389 natural dust and anthropogenic haze episodes, which demonstrates the consistency. Additionally,  
390 blending mineral dust with anthropogenic transport due to the north easterly wind, a wind that blows from  
391 the northeast, has been a subject of extensive discussion (Lin et al., 2007, 2012; Li et al., 2012). During  
392 the EAD, the dust from the Gobi Desert that was transported towards the western Pacific region could  
393 also carry anthropogenic aerosol, contributing to different levels of pollutant concentration. However, the  
394 distinct transport pathway at the high altitude between both aerosol types is a topic that has received less  
395 attention but is of significant importance to our understanding of atmospheric dynamics.

396 Figure 10 illustrates mineral dust and BC concentration's spatial and temporal distribution under  
397 the CMAQ\_Dust\_E20 scenario at 700 hPa from 24-31 January. The model reveals a high proportion of  
398 modeled dust aerosol (red dash circle) at the source region, indicating an uplift from the surface to 700  
399 hPa (Fig. 10a). This uplift, driven by the strong pressure gradient at the surface and the 'eastward moving  
400 trough system' at the upper level (700 hPa), is a key factor in the eastward and southward transfer of the  
401 dust (Fig. 10b). The high dust fraction reappears at the source region (Fig. 10c) and is transported  
402 eastwardly by the similar upper-level trough (Fig. 10d), causing a long dust belt at 15°N, distributing over  
403 central Asia continental, Taiwan Straits, Taiwan and large part of western Pacific Ocean. (Fig. 10e). On  
404 29 January, the model of E20 clearly predicted that the dust plume moved in the southward direction  
405 toward the South China Sea (Fig. 10f). The dust aerosol was left distributed at a certain part of the northern  
406 South China Sea and the Philippine Sea until it totally dissipated (Fig. 10g, h). This interesting result  
407 suggests the possible EAD at the longer distance at the upper level, which is a topic for further  
408 investigation.

409 The southward high-pressure system responsible for the long-range transport haze episode has  
410 been widely discussed (Chuang et al., 2008; Kong et al., 2021)—however, the upper-level transboundary  
411 transport needs to be addressed more. While focusing on CMAQ\_Dust\_E20, we attempted to characterize  
412 the long-range transport of modeled black carbon at the upper level (700 hPa) (Fig. 10i-p). As shown in  
413 Fig. 10(i), the modeled black carbon concentration is shown to be significantly distributed at central  
414 China. The black carbon transport pattern followed the eastward-moving trough system as the plume  
415 moved eastward and southward (Fig. 10m, n). Interestingly, the long black carbon belt is consistent with  
416 the long dust belt, as shown in Fig. 10(e, f). For instance, both modeled dust and BC were distributed at  
417 the western Pacific Ocean (Fig. 10e, f, m, n) and South China Sea (Fig. 10g, o). This means that the BC  
418 due to the anthropogenic emission and the natural EAD shared a similar transport pattern at the upper  
419 level, driven by the trough system. Such consistency has been verified by the MERRA-2 dust and BC  
420 mass column over the region (red dash rectangular in Fig. S5).

421 The dust aerosol vertical profiles (Fig. 11) show a significant distribution of the large dust fraction  
422 over the Asian Continent under all simulation scenarios (Fig. 11a1-e1), as indicated by the transect drawn



423 in Fig. 1. The westerly winds, depicted in Fig. 10, facilitated the eastward transport of the aerosol plume  
424 towards the western Pacific Ocean, where it accumulated along the 700 hPa altitude. Another plume was  
425 observed across the ocean on the east side of Taiwan Island (Fig. 11b1). On 27 January, showed another  
426 substantial fraction of dust covering the Asian Continent and Western Pacific Ocean, with significantly  
427 higher dust concentrations compared to Fig. 11a1. The plume distributed eastward exhibited a clear dust  
428 dome (Fig. 11a5-e5). These findings have important implications for understanding and predicting dust  
429 aerosol transport patterns and their potential environmental impact.

430 The vertical profile of the modeled BC mirrors the transport pattern of mineral dust, as shown in  
431 Fig. 12. A transparent BC dome was distributed along 700 hPa, echoing the pattern observed for dust.  
432 This simulation suggests the consistency of the “double dome” mechanism of Asian dust and biomass  
433 burning episodes (Dong et al., 2018; Huang et al., 2019). The potential warming effect of such a  
434 mechanism is a topic ripe for future studies. However, it's important to note that the dust dome contains  
435 a higher fraction of concentrations than the black carbon dome. The present simulation suggests that dust  
436 aerosol can reach up to 500 hPa, which is consistent with Kong et al. (2021). On the other hand, the black  
437 carbon plume was slightly lower, with approximately 600 hPa of the maximum height under the same  
438 meteorological condition. This section, which discusses the similarity and distinctiveness of natural dust  
439 and anthropogenic aerosol at the upper level, highlights the need for further study. The present simulation  
440 did not consider the two-way coupling model, and it is strongly suggested for future research.

441 Table 6 shows the modeled deposition and mass concentration for different simulation scenarios  
442 in January 2023. The simulation of the wet deposition and mass concentration for dust aerosol was the  
443 highest by E20. This is consistent with the globally averaged aerosol number concentrations over the  
444 ocean for the large size particle (Emerson et al., 2020). Contrary, P22 was the lowest in simulated wet  
445 deposition and mass concentration. P22 could increase the accumulation mode's  $V_d$  and reduce the  $PM_{2.5}$   
446 over CONUS, which is similar to the present result (Pleim et al., 2022). Moreover, the present simulation  
447 by P22 showed the highest  $V_d$  of the coarse mode that leads to the less simulated  $PM_{10}$ . P22 revised the  
448 impaction collective efficiency, which is the parameterization of  $R_b$ . In order to understand the sensitivity  
449 of  $R_b$  on CMAQ simulation, the  $R_b$  has been scaled up, as shown in Table 2. Generally, the increment of

450  $R_b$  has gradually increased the wet deposition (surface mass concentration) by 13.6 (45.8) %, 25.2 (83.3)  
451 %, and 28.2 (93.7) %, under P22E01, P22E02 and P22E03, respectively. In addition, the increment  
452 intensity at the surface was higher than at the upper level. The simulated dust at western Pacific Ocean  
453 responding to the different dry deposition schemes was shown during 27 January in Figure 11 (red-dash  
454 rectangular box). As  $R_b$  increased by P22E01 and P22E03, the simulated  $PM_{10}$  by base scheme P22 ( $\sim 30$   
455  $\mu g m^{-3}$ ), has increased to  $\sim 40 \mu g m^{-3}$  and  $\sim 50 \mu g m^{-3}$ , respectively. It is worth noted that P22E03 simulated  
456 a similar dust concentration as E20, indicating the importance of revising the  $R_b$ . On the contrary, the wet  
457 deposition and mass concentration were most significant for modeled BC under the S22 dry deposition  
458 scheme (Table 6). P22E01 only showed a minor increment, but it was nearly identical for P22E02 and  
459 P22E03 compared to P22.

#### 460 **4.0 Summary and Conclusions**

461 The chemical transport model is considered sensitive to the dry deposition parameterization besides the  
462 dust emission treatment. The present study demonstrates the impact of the dry deposition  
463 parameterizations (S22, E20, and P22) on aerosol performance in East Asia. It provides a significant  
464 analysis of the transboundary transport of East Asian Dust to Taiwan from a 22-31 January 2023 case  
465 study and multiple heavy dust storm episodes from 12 Mar-20 Apr 2021. Incorporating the latest dust  
466 emission treatment (Kong et al., 2024) into the CMAQ slightly improved the model performance to -  
467 47.01 % from -52.81 %. By implementing the E20 dry deposition scheme, characterized by calibrating  
468 the collection efficiency by Brownian diffusion and interception, the CMAQ simulation of the surface  
469  $PM_{10}$  has been improved by NMB of -41.9 %, as compared to the dry deposition proposed by P22 (-53.90  
470 %). Moreover, the modeled  $PM_{10}$  pattern by E20 at the upper level (700 hPa) was mainly consistent with  
471 the observed  $PM_{10}$ , especially in capturing the peak value. The dry deposition of E20 was correlated well  
472 with the high altitude in situ by 0.55, as compared to S22 (0.54) and P22 (0.46). On the contrary, simulated  
473 surface  $PM_{2.5}$  by S22 has been improved to -8.84 % from -12.63 % after using the latest dust treatment,  
474 and slightly better performance than E20 (-10.65 %) and P22 (-15.22 %). Additionally, the simulations  
475 of the multiple dust episodes in spring 2021 were re-constructed to evaluate the CMAQ performance over  
476 the Asian Continental. The E20 dry deposition scheme outperformed the others with the lowest NMB

477 value in simulating PM<sub>10</sub> (-25.4 %) and AOD (-26.2%). For the modeled PM<sub>2.5</sub>, S22 performed slightly  
478 better than E20, with NMB of -36.29 % and -37.5 %, respectively.

479

480 The previous CMAQ model, modulated by Kong et al. (2021; 2024), showed excessive deposition at the  
481 marine boundary layer, leading to underestimating the modeled surface PM<sub>10</sub>. However, using the E20  
482 scheme over the entire model domain, our updated model has the lowest  $V_d$ . This precise reduction of  $V_d$   
483 of the coarse mode particle, responsible for reducing the PM<sub>10</sub> simulation underestimation, has not just  
484 minimized, but effectively minimized the total loss of aerosol to the surface, leading to a concentration  
485 increment. Furthermore, the low-lying modeled  $V_d$  across the water surface by E20 could be crucial in  
486 reducing the excessive aerosol deposition over the ocean layer.

487

488 It is worth revealing that the transboundary transport of EAD from the Asian continent towards the  
489 western Pacific Ocean at the upper level was associated with the eastward moving trough system. Such  
490 transport mechanisms have been found to bring along black carbon aerosol, which is primarily the main  
491 element of China's human-made emissions. More interestingly, both aerosol profiles created a "long dust-  
492 black carbon belt" along the 15°N. The 'double dome mechanism', a concept proposed by Huang et al.  
493 (2019) that depicts the superposition of the two aerosol types, was also simulated in the present study.  
494 Besides the similarity of both, the discrepancy in the case of the aerosol deposition and mass concentration  
495 was shown. By comparing the base P22 scheme to the revised scheme (P22E01-P22E03), wet deposition  
496 increases and hence increases the dust aerosol. In other ways, black carbon aerosol also increases in a  
497 minimal magnitude, not as much as dust aerosol. This study highlights the importance of dry deposition  
498 schemes for the modeled dust and black carbon concentration and provides a reference for better dry  
499 deposition schemes in CTMs over East Asia.

## 500 **Data Availability**

501 MERRA-2 data are available online through the NASA Goddard Earth Sciences Data Information  
502 Services Center (GES DISC; <https://disc.gsfc.nasa.gov>; last access: 01 August 2024). MODIS data used

503 in this study are available at <https://asdc.larc.nasa.gov/>(last access: 01 August 2024). The observational  
504 data at LABS can be ordered by contacting corresponding authors.

#### 505 **Author Contribution**

506 **Steven Soon-Kai Kong:** Conceptualization; Data curation; Formal analysis; Investigation; Methodology;  
507 Software; Validation; Visualization; Writing – original draft; Writing – review and editing.

508 **Joshua S. Fu:** Conceptualization; Investigation; Methodology; Formal analysis; Writing – review and  
509 editing.

510 **Neng-Huei Lin:** Conceptualization; Visualization; Supervision; Funding acquisition; Resources; Writing  
511 – review and editing.

512 **Guey-Rong Sheu:** Funding acquisition; Resources.

513 **Wei-Syun Huang:** Data curation; Software.

#### 514 **Competing Interests**

515 Some authors are members of the editorial board of journal ACP.

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#### 521 **References:**

522 Albert, M. F., Anguelova, M. D., Manders, A. M., Schaap, M., & Leeuw, G. D.: Parameterization of  
523 oceanic whitecap fraction based on satellite observations. *Atmospheric Chemistry and Physics*, 16, 21,  
524 13725–13751. <https://doi.org/10.5194/acp-16-13725-2016>, 2016.

525 Chang, J. and Hanna, S.: Air quality model performance evaluation, *Meteorol Atmos Phys.*, 87, 167–196,  
526 <https://doi.org/10.1007/s00703-003-0070-7>, 2004.

527 Chuang, M. T., Fu, J. S., Lee, C., Lin, N., Gao, Y., Wang, S., Sheu, G., Hsiao, T., Wang, J., Yen, M., Lin,  
528 T., and Thongboonchoo, N.: The Simulation of Long-Range Transport of Biomass Burning Plume and  
529 Short-Range Transport of Anthropogenic Pollutants to a Mountain Observatory in East Asia during the  
530 7-SEAS / 2010 Dongsha Experiment, 2933–2949, <https://doi.org/10.4209/aaqr.2015.07.0440>, 2016.

531 Chuang, M. T., Fu, J. S., Jang, C. J., Chan, C. C., Ni, P. C., and Lee, C. Te: Simulation of long-range  
532 transport aerosols from the Asian Continent to Taiwan by a Southward Asian high-pressure system, *Sci.*  
533 *Total Environ.*, 406, 168–179, <https://doi.org/10.1016/j.scitotenv.2008.07.003>, 2008.

534 Chuang, M.-T., Ooi, M. C. G., Lin, N.-H., Fu, J. S., Lee, C.-T., Wang, S.-H., Yen, M.-C., Kong, S. S.-K.,  
535 and Huang, W.-S.: Study on the impact of three Asian industrial regions on PM<sub>2.5</sub> in Taiwan and the  
536 process analysis during transport, *Atmos. Chem. Phys.*, 20, 14947–14967, [https://doi.org/10.5194/acp-](https://doi.org/10.5194/acp-20-14947-2020)  
537 [20-14947-2020](https://doi.org/10.5194/acp-20-14947-2020), 2020.

538 Dong, X., Fu, J. S., Huang, K., Tong, D., and Zhuang, G.: Model development of dust emission and  
539 heterogeneous chemistry within the Community Multiscale Air Quality modeling system and its  
540 application over East Asia, *Atmos. Chem. Phys.*, 16, 8157–8180, [https://doi.org/10.5194/acp-16-8157-](https://doi.org/10.5194/acp-16-8157-2016)  
541 [2016](https://doi.org/10.5194/acp-16-8157-2016), 2016.

542 Dong, X., Fu, J. S., Huang, K., Lin, N., Wang, S., and Yang, C.: Analysis of the Co-existence of Long-  
543 range Transport Biomass Burning and Dust in the Subtropical West Pacific Region, *Sci. Rep.*, 1–10,  
544 <https://doi.org/10.1038/s41598-018-27129-2>, 2018.

545 Dong, X., Fu, J. S., Huang, K., Zhu, Q., and Tipton, M.: Regional climate effects of biomass burning and  
546 dust in East Asia: Evidence from modeling and observation, *Geophysical Research Letters*, 46,  
547 <https://doi.org/10.1029/2019GL083894>, 2019.

548 Emerson, E. W., Hodshire, A. L., DeBolt, H. M., Bilsback, K. R., Pierce, J. R., McMeeking, G. R., and  
549 Farmer, D. K.: Revisiting particle dry deposition and its role in radiative effect estimates, *Proc. Natl.*  
550 *Acad. Sci. U. S. A.*, 117, 26076–26082, <https://doi.org/10.1073/pnas.2014761117>, 2020.

551 Foroutan, H., Young, J., Napelenok, S., Ran, L., Appel, K., Gilliam, R., and Pleim, J.: Journal of Advances  
552 in Modeling Earth Systems, *J. Adv. Model. Earth Syst.*, 9, 585–606,  
553 <https://doi.org/10.1002/2013MS000282>. Received, 2017.

554 Gaydos, T. M., Pinder, R., Koo, B., Fahey, K. M., Yarwood, G., and Pandis, S. N.: Development and  
555 application of a three-dimensional aerosol chemical transport model, PMCAMx, *Atmos. Environ.*, 41,  
556 2594–2611, <https://doi.org/10.1016/j.atmosenv.2006.11.034>, 2007.

557 Gelaro, R., McCarty, W., Suárez, M. J., Todling, R., Molod, A., Takacs, L., Randles, C. A., Darmenov,  
558 A., Bosilovich, M. G., Reichle, R., Wargan, K., Coy, L., Cullather, R., Draper, C., Akella, S., Buchard,  
559 V., Conaty, A., da Silva, A. M., Gu, W., Kim, G. K., Koster, R., Lucchesi, R., Merkova, D., Nielsen, J.  
560 E., Partyka, G., Pawson, S., Putman, W., Rienecker, M., Schubert, S. D., Sienkiewicz, M., and Zhao, B.:  
561 The modern-era retrospective analysis for research and applications, version 2 (MERRA-2), *J. Clim.*, 30,  
562 5419–5454, <https://doi.org/10.1175/JCLI-D-16-0758.1>, 2017.

563 Giardina, M. and Buffa, P.: A new approach for modeling dry deposition velocity of particles, *Atmos.*  
564 *Environ.*, 180, 11–22, <https://doi.org/10.1016/j.atmosenv.2018.02.038>, 2018.

565 Gui, K., Yao, W., Che, H., An, L., Zheng, Y., Li, L., Zhao, H., Zhang, L., Zhong, J., Wang, Y., and Zhang,  
566 X.: Record-breaking dust loading during two mega dust storm events over northern China in March 2021:  
567 aerosol optical and radiative properties and meteorological drivers, *Atmos. Chem. Phys.*, 22, 7905–7932,  
568 <https://doi.org/10.5194/acp-22-7905-2022>, 2022.

569 Han, X., Ge, C., Tao, J., Zhang, M., and Zhang, R.: Air quality modeling for a strong dust event in East  
570 Asia in March 2010, *Aerosol Air Qual. Res.*, 12, 615–628, <https://doi.org/10.4209/aaqr.2011.11.0191>,  
571 2012.

572 He, Y., Yi, F., Yin, Z., Liu, F., Yi, Y., and Zhou, J.: Mega Asian dust event over China on 27–31 March  
573 2021 observed with space-borne instruments and ground-based polarization lidar, *Atmos. Environ.*, 285,  
574 119238, <https://doi.org/10.1016/j.atmosenv.2022.119238>, 2022.

575 Hogrefe, C., Bash, J. O., Pleim, J. E., Schwede, D. B., Gilliam, R. C., Foley, K. M., Appel, K. W., and  
576 Mathur, R.: An analysis of CMAQ gas-phase dry deposition over North America through grid-scale and  
577 land-use-specific diagnostics in the context of AQMEII4, *Atmos. Chem. Phys.*, 23, 8119–8147,  
578 <https://doi.org/10.5194/acp-23-8119-2023>, 2023.

579 Huang, H.-Y., Wang, S.-H., Huang, W.-X., Lin, N.-H., Chuang, M.-T., da Silva, A. M., Peng, C.-M.:  
580 Influence of synoptic-dynamic meteorology on the long-range transport of Indochina biomass burning  
581 aerosols, *J. Geophys. Res.*, 111, 125, e2019JD031260. <https://doi.org/10.1029/2019JD031260>, 2020.

582 Huang, K., Fu, J. S., Lin, N.-H., Wang, S.-H., Dong, X., Wang, G.: Superposition of Gobi Dust and  
583 Southeast Asian Biomass Burning: The Effect of Multisource Long - Range Transport on Aerosol Optical  
584 Properties and Regional Meteorology Modification, *J. Geophys. Res.*, 124, 16, 9464-9483,  
585 <https://doi.org/10.1029/2018JD030241>, 2019.

586 Huang, W. S., Griffith, S. M., Lin, Y. C., Chen, Y. C., Lee, C. Te, Chou, C. C. K., Chuang, M. T., Wang,  
587 S. H., and Lin, N. H.: Satellite-based emission inventory adjustments improve simulations of long-range  
588 transport events, *Aerosol Air Qual. Res.*, 21, 1–16, <https://doi.org/10.4209/AAQR.210121>, 2021.

589 Jin, J., Pang, M., Segers, A., Han, W., Fang, L., Li, B., Feng, H., Lin, H. X., and Liao, H.: Inverse  
590 modeling of the 2021 spring super dust storms in East Asia, *Atmos. Chem. Phys.*, 22, 6393–6410,  
591 <https://doi.org/10.5194/acp-22-6393-2022>, 2022.

592 Khan, T. R. and Perlinger, J. A.: Evaluation of five dry particle deposition parameterizations for  
593 incorporation into atmospheric transport models, *Geosci. Model Dev.*, 10, 3861–3888,  
594 <https://doi.org/10.5194/gmd-10-3861-2017>, 2017.

595 Kok, J. F., Parteli, E. J. R., Michaels, T. I., Karam, D. B., and Pierre, U.: The physics of wind-blown sand  
596 and dust, 1–119, n.d.

597 Kong, S. S.-K., Pani, S. K., Griffith, S. M., Ou-Yang, C.-F., Babu, S. R., Chuang, M.-T., Ooi, M. C. G.,  
598 Huang, W.-S., Sheu, G.-R., and Lin, N.-H.: Distinct transport mechanisms of East Asian dust and the  
599 impact on downwind marine and atmospheric environments, *Sci. Total Environ.*, 827, 154255,  
600 <https://doi.org/10.1016/j.scitotenv.2022.154255>, 2022.

601 Kong, S. S., Fu, J. S., Dong, X., Chuang, M., Chel, M., Ooi, G., Huang, W., Griffith, S. M., Kumar, S.,  
602 and Lin, N.: Sensitivity analysis of the dust emission treatment in CMAQv5. 2. 1 and its application to  
603 long-range transport over East Asia, *Atmos. Environ.*, 118441,  
604 <https://doi.org/10.1016/j.atmosenv.2021.118441>, 2021.

605 Kong, S. S. K., Ravindra Babu, S., Wang, S. H., Griffith, S. M., Chang, J. H. W., Chuang, M. T., Sheu,  
606 G. R., and Lin, N. H.: Expanding the simulation of East Asian super dust storms: physical transport

607 mechanisms impacting the western Pacific, *Atmos. Chem. Phys.*, 24, 1041–1058,  
608 <https://doi.org/10.5194/acp-24-1041-2024>, 2024.

609 Li, M., Zhang, Q., Kurokawa, J.-I., Woo, J.-H., He, K., Lu, Z., Ohara, T., Song, Y., Streets, D. G.,  
610 Carmichael, G. R., Cheng, Y., Hong, C., Huo, H., Jiang, X., Kang, S., Liu, F., Su, H., and Zheng, B.:  
611 MIX: a mosaic Asian anthropogenic emission inventory under the international collaboration framework  
612 of the MICS-Asia and HTAP, *Atmos. Chem. Phys.*, 17, 935–963, [https://doi.org/10.5194/acp-17-935-](https://doi.org/10.5194/acp-17-935-2017)  
613 2017, 2017.

614 Liang, L., Han, Z., Li, J., Xia, X., Sun, Y., Liao, H., Liu, R., and Liang, M.: Science of the Total  
615 Environment Emission, transport, deposition, chemical and radiative impacts of mineral dust during  
616 severe dust storm periods in March 2021 over East Asia, *Sci. Total Environ.*, 852, 158459,  
617 <https://doi.org/10.1016/j.scitotenv.2022.158459>, 2022.

618 Liu, S., Xing, J., Sahu, S. K., Liu, X., Liu, S., Jiang, Y., Zhang, H., Li, S., Ding, D., Chang, X., and Wang,  
619 S.: Wind-blown dust and its impacts on particulate matter pollution in Northern China: Current and future  
620 scenarios, *Environ. Res. Lett.*, 16, 114041, <https://doi.org/10.1088/1748-9326/ac31ec>, 2021.

621 Massad, R. S., Nemitz, E., and Sutton, M. A.: Review and parameterisation of bi-directional ammonia  
622 exchange between vegetation and the atmosphere, *Atmos. Chem. Phys.*, 10, 10359–10386,  
623 <https://doi.org/10.5194/acp-10-10359-2010>, 2010.

624 Nemitz, E., Milford, C., and Sutton, M. A.: A two-layer canopy compensation point model for describing  
625 bi-directional biosphere-atmosphere exchange of ammonia, *Q. J. R. Meteorol. Soc.*, 127, 815–833,  
626 <https://doi.org/10.1256/smsqj.57305>, 2001.

627 Ooi, M., Chuang, M.-T., Fu, J., Kong, S., Huang, W.-S., Wang, S.-H., Chan, A., Pani, S., and Lin, N.-H.:  
628 Improving prediction of trans-boundary biomass burning plume dispersion: from northern peninsular  
629 Southeast Asia to downwind western north Pacific Ocean, *Atmos. Chem. Phys.*, 20, 14947–14967,  
630 <https://doi.org/10.5194/acp-2020-1283>, 2021.

631 Pani, S. K., Wang, S. H., Lin, N. H., Lee, C. Te, Tsay, S. C., Holben, B. N., Janjai, S., Hsiao, T. C.,  
632 Chuang, M. T., and Chantara, S.: Radiative effect of springtime biomass-burning aerosols over northern  
633 indochina during 7-SEAS/BASELInE 2013 campaign, *Aerosol Air Qual. Res.*, 16, 2802–2817,  
634 <https://doi.org/10.4209/aaqr.2016.03.0130>, 2016.



635 Pani, S. K., Wang, S. H., Lin, N. H., Chantara, S., Lee, C. Te, and Thepnuan, D.: Black carbon over an  
636 urban atmosphere in northern peninsular Southeast Asia: Characteristics, source apportionment, and  
637 associated health risks, *Environ. Pollut.*, 259, 113871, <https://doi.org/10.1016/j.envpol.2019.113871>,  
638 2020.

639 Ramanathan, V and Carmichael, G.: Climate change due to BC, *Nat. Geosci.*, 1, 221–227, 2008.

640 Ravindra Babu, S., Ou-Yang, C. F., Griffith, S. M., Pani, S. K., Kong, S. S. K., and Lin, N. H.: Transport  
641 pathways of carbon monoxide from Indonesian fire pollution to a subtropical high-Altitude mountain site  
642 in the western North Pacific, *Atmos. Chem. Phys.*, 23, 4727–4740, [https://doi.org/10.5194/acp-23-4727-](https://doi.org/10.5194/acp-23-4727-2023)  
643 2023, 2023.

644 Ryu, Y. H. and Min, S. K.: Improving Wet and Dry Deposition of Aerosols in WRF-Chem: Updates to  
645 Below-Cloud Scavenging and Coarse-Particle Dry Deposition, *J. Adv. Model. Earth Syst.*, 14,  
646 <https://doi.org/10.1029/2021MS002792>, 2022.

647 Saylor, R. D., Baker, B. D., Lee, P., Tong, D., Pan, L., and Hicks, B. B.: The particle dry deposition  
648 component of total deposition from air quality models: right, wrong or uncertain?, *Tellus, Ser. B Chem.*  
649 *Phys. Meteorol.*, 71, 1–22, <https://doi.org/10.1080/16000889.2018.1550324>, 2019.

650 Shu, Q., Koo, B., Yarwood, G., and Henderson, B. H.: Strong influence of deposition and vertical mixing  
651 on secondary organic aerosol concentrations in CMAQ and CAMx, *Atmos. Environ.*, 171, 317–329,  
652 <https://doi.org/10.1016/j.atmosenv.2017.10.035>, 2017.

653 Slinn, W. G. N.: Predictions for particle deposition to vegetative canopies, *Atmos. Environ.*, 16, 1785–  
654 1794, [https://doi.org/10.1016/0004-6981\(82\)90271-2](https://doi.org/10.1016/0004-6981(82)90271-2), 1982.

655 Tang, W., Dai, T., Cheng, Y., Wang, S., and Liu, Y.: A Study of a Severe Spring Dust Event in 2021 over  
656 East Asia with WRF-Chem and Multiple Platforms of Observations, *Remote Sens.*, 14, 3795,  
657 <https://doi.org/10.3390/rs14153795>, 2022.

658 Tav, J., Masson, O., Burnet, F., Paulat, P., Bourriane, T., Conil, S. and Pourcelot, L.: Determination of  
659 Fog-Droplet Deposition Velocity from a Simple Weighing Method. *Aerosol Air Qual. Res.* 18: 103-113.  
660 <https://doi.org/10.4209/aaqr.2016.11.0519>, 2018.

661 Wang, W., Zhou, H., Lyu, R., Shao, L., Li, W., Xing, J., Zhao, Z., Li, X., Zhou, X., and Zhang, D.:  
662 Organic Carbon and Elemental Carbon in Two Dust Plumes at a Coastal City in North China, *Aerosol*  
663 *Air Qual. Res.*, 24, <https://doi.org/10.4209/aaqr.240002>, 2024.

664 Wesley, M. L.: Parameterization of Surface Resistances to Gaseous Dry Deposition in Regional-Scale  
665 Numerical Models, *Atmos. Environ.*, 23, 1293–1304, 1989.

666 Xiao, H. W., Xu, Y., and Xiao, H. Y.: Source apportionment of black carbon aerosols in winter across  
667 China, *Atmos. Environ.*, 298, <https://doi.org/10.1016/j.atmosenv.2023.119622>, 2023.

668 Zeng, Y., Wang, M., Zhao, C., Chen, S., Liu, Z., Huang, X., and Gao, Y.: WRF-Chem v3.9 simulations  
669 of the East Asian dust storm in May 2017: Modeling sensitivities to dust emission and dry deposition  
670 schemes, *Geosci. Model Dev.*, 13, 2125–2147, <https://doi.org/10.5194/gmd-13-2125-2020>, 2020.

671 Zhang, L., Gong, S., Padro, J., and Barrie, L.: A size-segregated particle dry deposition scheme for an  
672 atmospheric aerosol module, *Atmos. Environ.*, 35, 549–560, [https://doi.org/10.1016/S1352-](https://doi.org/10.1016/S1352-2310(00)00326-5)  
673 [2310\(00\)00326-5](https://doi.org/10.1016/S1352-2310(00)00326-5), 2001.

674 Zhang, L., Zhang, H., Li, Q., Cai, X., and Song, Y.: Vertical dispersion mechanism of long-range  
675 transported dust in Beijing: Effects of atmospheric turbulence, *Atmos. Res.*, 269, 106033,  
676 <https://doi.org/10.1016/j.atmosres.2022.106033>, 2022.

677 Zheng, B., Tong, D., Li, M., Liu, F., Hong, C., Geng, G., Li, H., Li, X., Peng, L., Qi, J., Yan, L., Zhang,  
678 Y., Zhao, H., Zheng, Y., He, K., and Zhang, Q.: Trends in China's anthropogenic emissions since 2010  
679 as the consequence of clean air actions, *Atmos. Chem. Phys.*, 18, 14095–14111,  
680 <https://doi.org/10.5194/acp-18-14095-2018>, 2018.

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692 **Table 1.** Model settings.

Model setting	Descriptions
Period	12 March-20 April 2021 and 22-31 January 2023
Domain	d01, d02, and d03 with 45 KM, 15 KM, and 5 KM of the resolutions, respectively
Boundary condition	NCEP FNL lateral boundary condition
Surface and land surface model	NOAH
Numerical weather model	WRF v40, including grid and observation nudging at d01.
Chemical transport model	CMAQ v5.4
Gas-phase chemistry and aerosol mechanism	CB06e51 + AE7
Emission Inventory	MICS-ASIA III emission in 2023, adjusted from the emission in 2017 (Zhang et al., 2018) based on the OMI-NO <sub>x</sub> satellite (Huang et al., 2021).
Online dust treatment	The refined windblown dust treatment suggested by Kong et al. (2024).
Dry deposition option	STAGE (S22, E20 and P22).

693

694 **Table 2.** Simulation scenarios used in this present study.

Experiments	Online dust emission treatment by Kong et al. (2024)	Dry deposition treatment	Surface resistance (R <sub>b</sub> ) at the smooth surface
CMAQ_Off_S22	Off	S22	Default
CMAQ_Dust_S22	On	S22	Default
CMAQ_Dust_E20	On	E20	Default
CMAQ_Dust_P22	On	P22	Default
CMAQ_Dust_P22E01	On	P22	Increased by a factor of 10
CMAQ_Dust_P22E02	On	P22	Increased by a factor of 50
CMAQ_Dust_P22E03	On	P22	Increased by a factor of 100

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707 **Table 3.** Statistical evaluation for PM<sub>10</sub> and PM<sub>2.5</sub> concentrations during 22-31 January 2023 for Cape  
 708 Fuguei under the multiple simulation scenarios.

Benchmark		CMAQ						
		Off_ S22	Dust_ S22	Dust_ E20	Dust_ P22	Dust_ P22E01	Dust_ P22E02	Dust_ P22E03
<b>PM<sub>10</sub></b>								
MeanObs		49.97	49.97	49.97	49.97	49.97	49.97	49.97
MeanMod		23.58	26.48	29.04	23.04	25.99	27.36	27.69
NMSE		0.66	0.56	0.49	0.71	0.57	0.53	0.52
NMB	± 85%	-52.81	-47.01	-41.90	-53.90	-47.99	-45.24	-44.58
R	> 0.35	0.43	0.46	0.52	0.42	0.48	0.51	0.52
NMBF		-1.12	-0.89	-0.72	-1.17	-0.92	-0.83	-0.80
<b>PM<sub>2.5</sub></b>								
MeanObs		15.52	15.52	15.52	15.52	15.52	15.52	15.52
MeanMod		13.56	14.15	13.86	13.16	13.26	13.22	13.20
NMSE		0.30	0.30	0.29	0.31	0.30	0.30	0.30
NMB	± 85%	-12.63	-8.84	-10.65	-15.22	-14.54	-14.80	-14.92
R	> 0.35	0.50	0.53	0.53	0.52	0.53	0.53	0.53
NMBF		-0.14	-0.20	-0.12	-0.18	-0.17	-0.17	-0.18

709 Note: the definition of the statistical formulas NMSE: Normalized Mean Square Error; NMB: Normalized  
 710 Mean Bias; R: Correlation Coefficient and NMBF: Normalized Mean Bias Factor

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 712 **Table 4.** CMAQ evaluation for PM<sub>10</sub> and PM<sub>2.5</sub> against the averaged 100 observation sites across  
 713 mainland China (Fig. S1) and AOD against MODIS daily observation near the dust source region (above  
 714 30°N) with Normalized Mean Bias (NMB) under the multiple simulation scenarios (Fig. S3). Spring  
 715 2021, 3.15, 3.27, and 4.18 represent the evaluation period by 12 March-20 April 2021, 14-16 March 2021,  
 716 26-28 March 2021, and 17-19 April 2021, respectively.

Parameters	Period	CMAQ						
		Off_ S22	Dust_ S22	Dust_ E20	Dust_ P22	Dust_ P22E01	Dust_ P22E02	Dust_ P22E03
PM <sub>10</sub>	Spring 2021	-75.00	-45.97	-25.43	-59.82	-45.09	-35.42	-32.92
PM <sub>2.5</sub>	Spring 2021	-55.56	-36.29	-37.50	-42.47	-41.20	-41.51	-41.66
AOD	3.15	-80.49	-46.41	-38.97	-48.45	-44.80	-41.66	-40.80
	3.27	-80.92	-41.84	-36.39	-44.52	-41.60	-39.30	-38.72
	4.18	-83.09	-7.83	-3.20	-14.52	-9.45	-7.18	-6.67
	Mean AOD	-81.50	-32.03	-26.19	-35.83	-31.95	-29.38	-28.73

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725 **Table 5.** Average deposition velocity in January 2023 (S22\_2023, E20\_2023, and P22\_2023) and Spring  
 726 2021 (S22\_2021, E20\_2021, and P22\_2021) for Aitken, Accumulation, and Coarse modes over land and  
 727 ocean boundary layer, respectively.

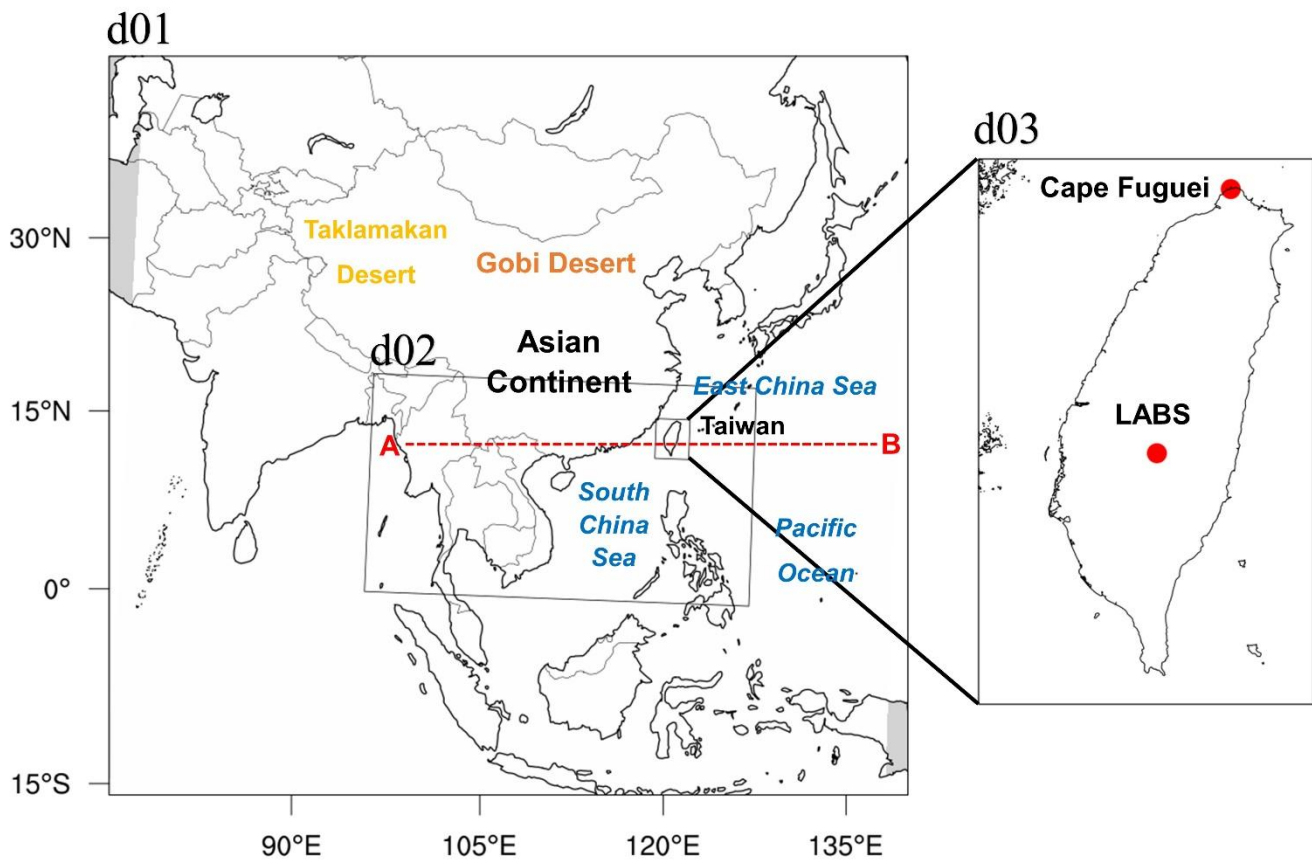
Dry deposition schemes (cm s <sup>-1</sup> )	Aitken		Accumulation		Coarse	
	Land	Ocean	Land	Ocean	Land	Ocean
S22_2023	0.219	0.117	0.120	0.064	0.078	0.085
E20_2023	0.090	0.074	0.065	0.040	0.139	0.060
P22_2023	0.085	0.062	0.072	0.043	0.290	0.116
S22_2021	0.308	0.100	0.109	0.042	0.077	0.060
E20_2021	0.139	0.063	0.063	0.026	0.142	0.053
P22_2021	0.119	0.047	0.072	0.025	0.265	0.070

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754 **Table 6.** Model averaged dry, wet deposition and mass concentration for dust and BC aerosols in January  
 755 2023 (10-days averaged) for different simulation scenarios.

Dust (ASOIL)	Dry deposition (mg m <sup>-2</sup> )	Wet deposition (mg m <sup>-2</sup> )	Mass concentration at the surface (μg m <sup>-3</sup> )	Mass concentration at 700 hPa (μg m <sup>-3</sup> )
S22	0.267	0.112	6.34	3.62
E20	0.167	0.136	10.25	4.40
P22	0.300	0.103	4.79	3.56
P22E01	0.243	0.117	7.00	3.79
P22E02	0.196	0.129	8.78	4.13
P22E03	0.183	0.132	9.28	4.22
BC (AECI + AECJ)	Dry deposition (μg m <sup>-2</sup> )	Wet deposition (μg m <sup>-2</sup> )	Mass concentration at the surface (ng m <sup>-3</sup> )	Mass concentration at 700 hPa (ng m <sup>-3</sup> )
S22	5.13	50.49	492	60.04
E20	8.09	48.27	471	57.73
P22	17.79	40.96	411	50.95
P22E01	16.88	41.64	415	51.23
P22E02	16.82	41.67	415	51.27
P22E03	16.82	41.67	415	51.27

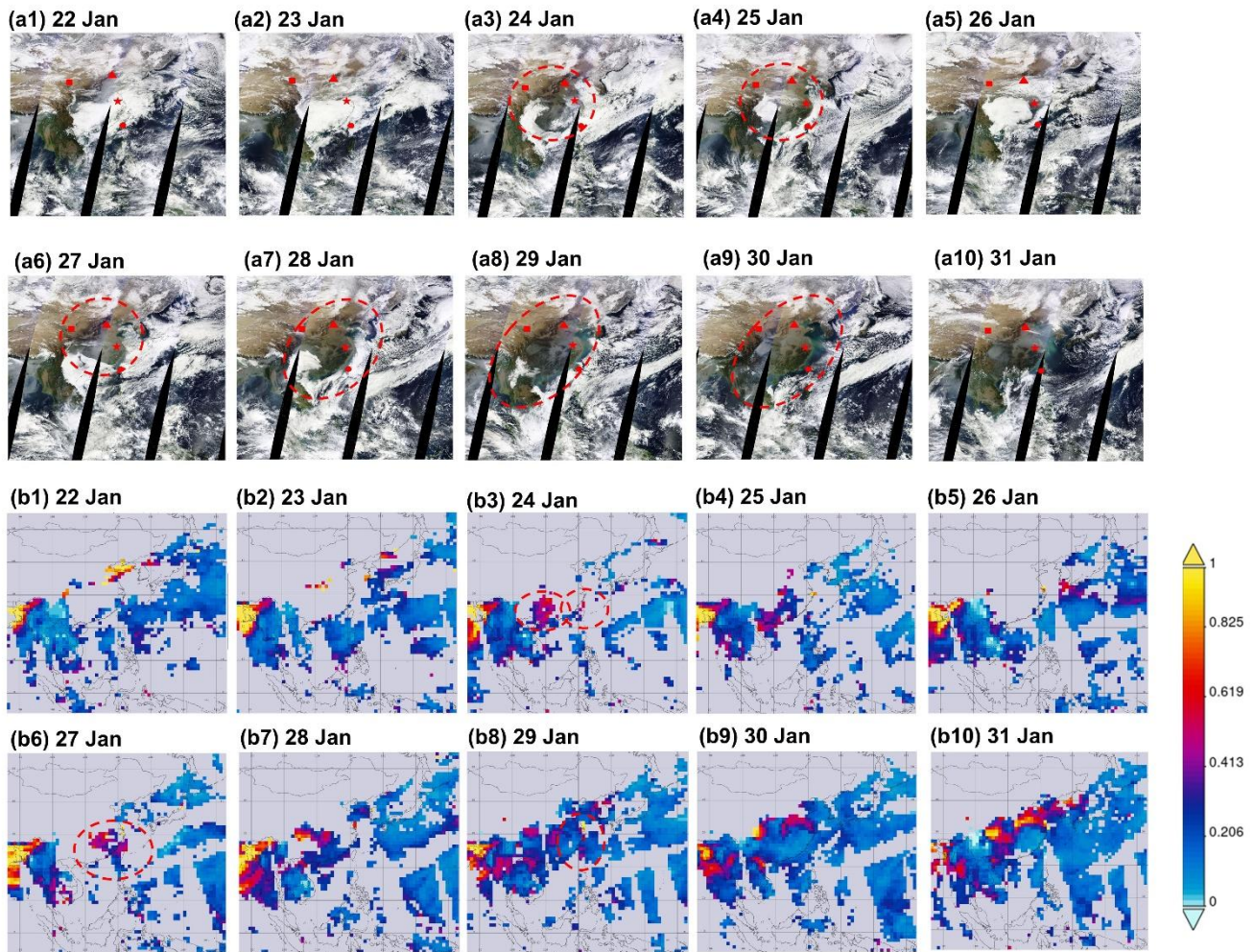
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758 **Figure 1:** Modeling domain configuration in East Asia. Ground-based air quality stations in Taiwan at  
 759 Cape Fuguei and Lulin Atmospheric Background Station (LABS) are shown in the zoomed panel. The  
 760 red dash line (A→B) represents the transects that the aerosol plumes traveled along in this study and that  
 761 are discussed in Section 3.4;

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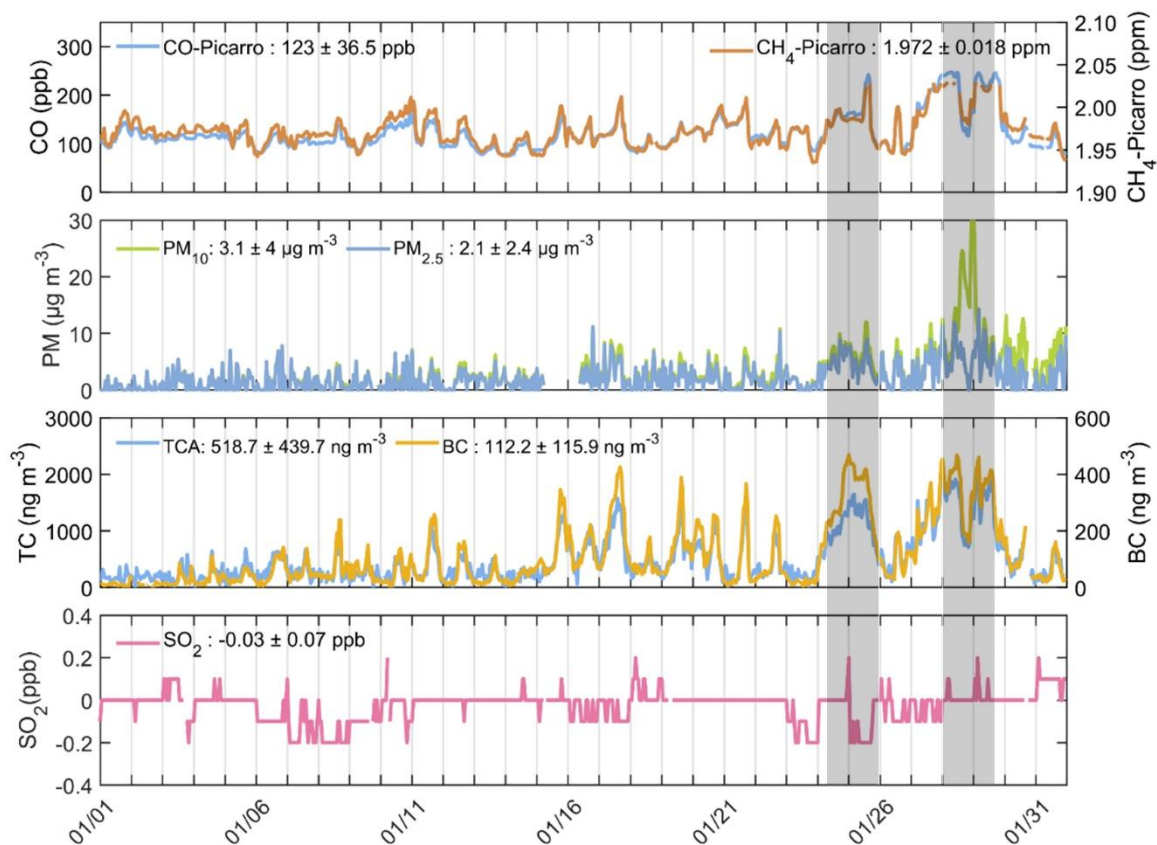
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764 **Figure 2:** MODIS Terra images (a1-a10) and MODIS aerosol optical depth AOD at 550 nm (b1-b10)  
 765 showing dust outbreak across East Asia during 22-31 January 2023. Red Rectangular, triangle, star and  
 766 circle indicate Lanzhou, Beijing, Shanghai and Taiwan. The red circle with dash line indicates the dust  
 767 plume.

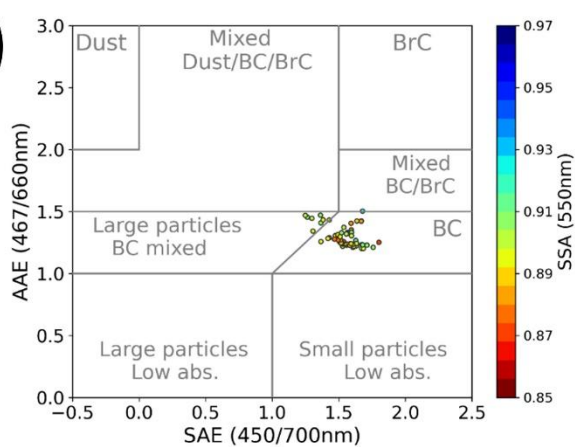
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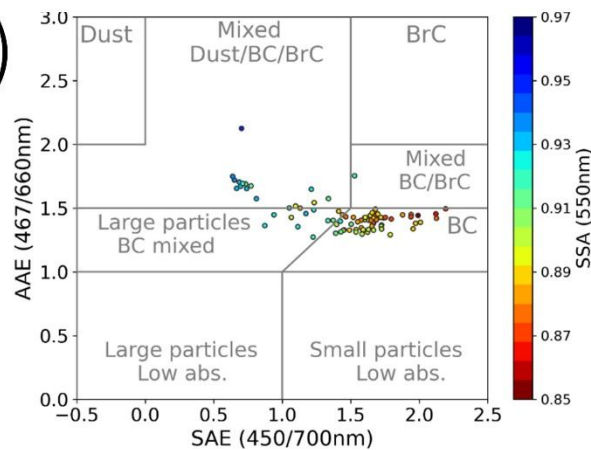
(a)



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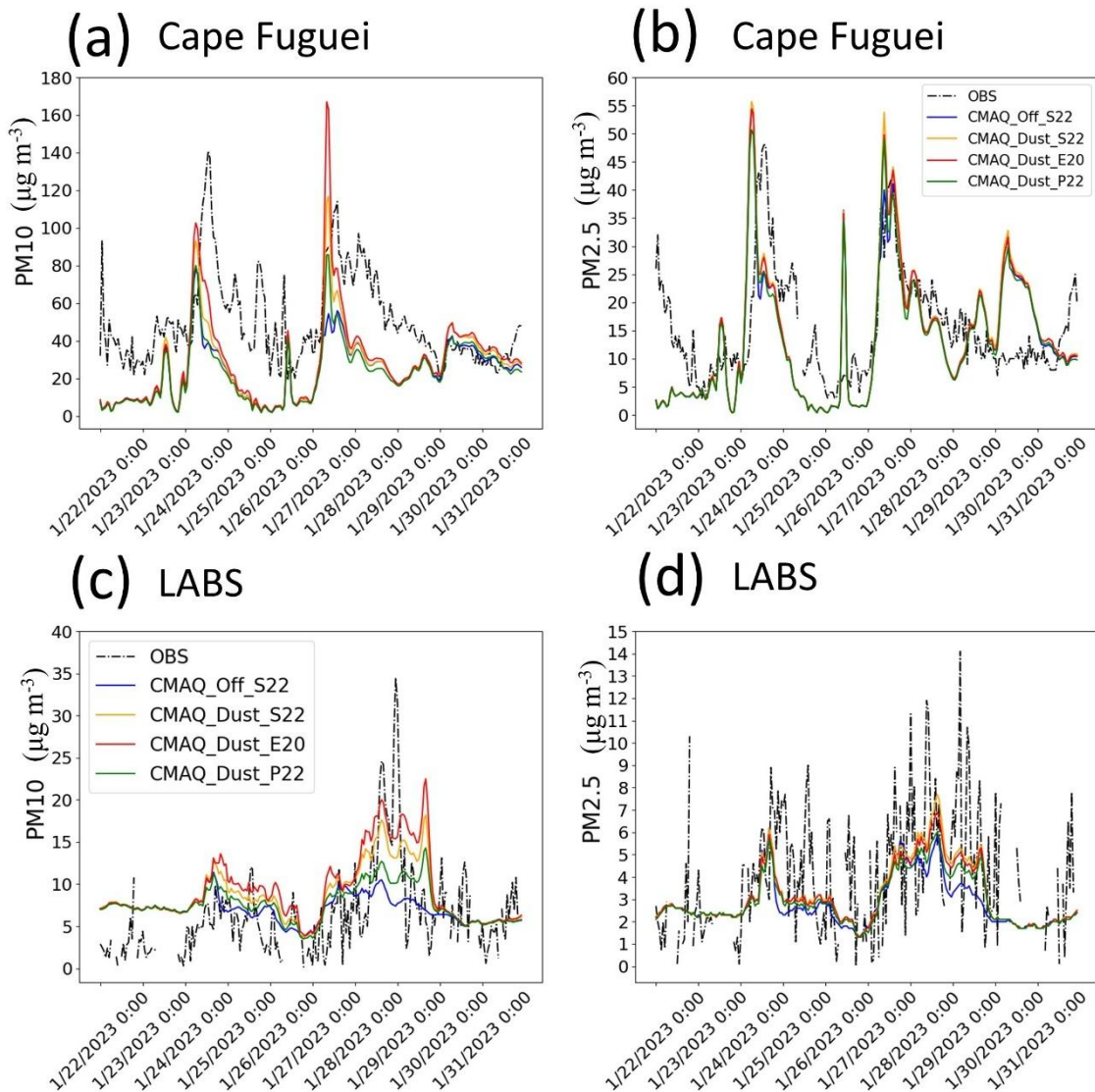


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770 **Figure 3:** (a) Time series of observed pollutants over LABS during January 2023. The aerosol radiation  
771 properties during (b) 24-26 January and (c) 27-30 January 2023.

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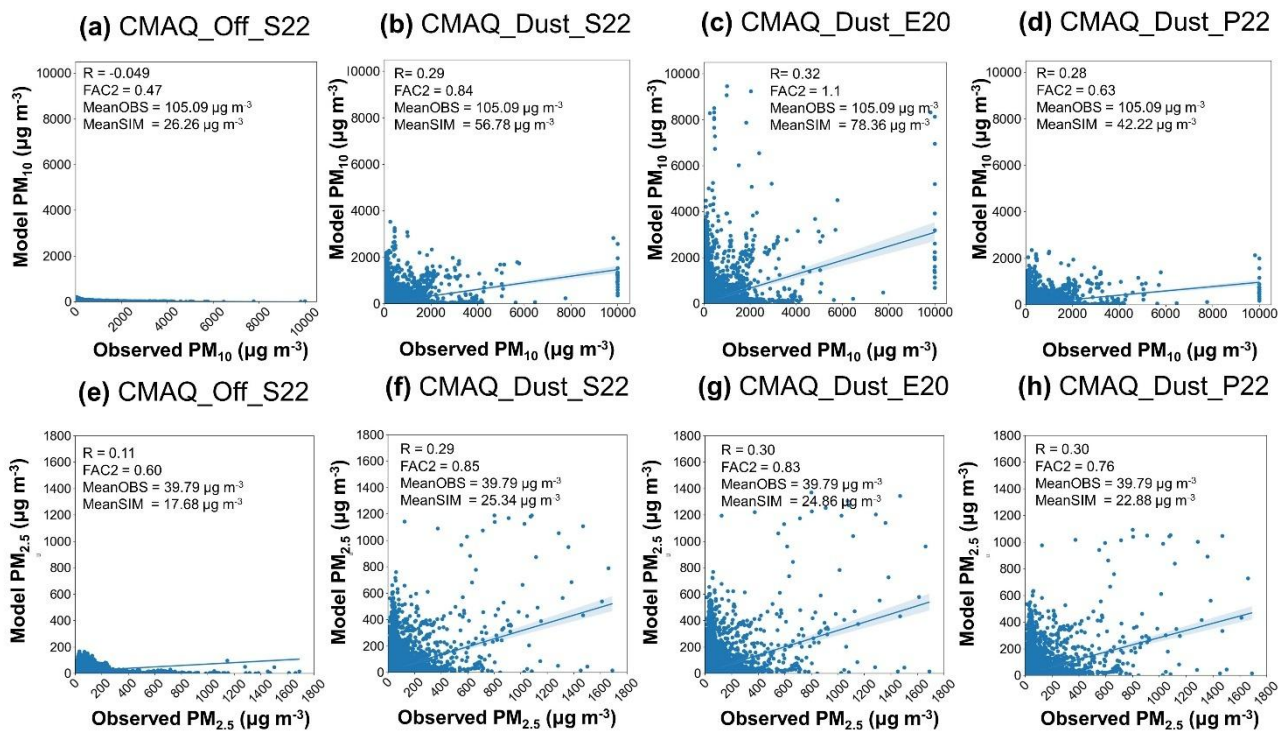
775 **Figure 4:** Time series of PM<sub>10</sub> (left panel) and PM<sub>2.5</sub> (right panel) concentrations during 22-31 January  
 776 2023 under multiple deposition schemes over the Cape Fuguei (upper panel) and LABS (lower panel),  
 777 representing the surface and high altitude, respectively.

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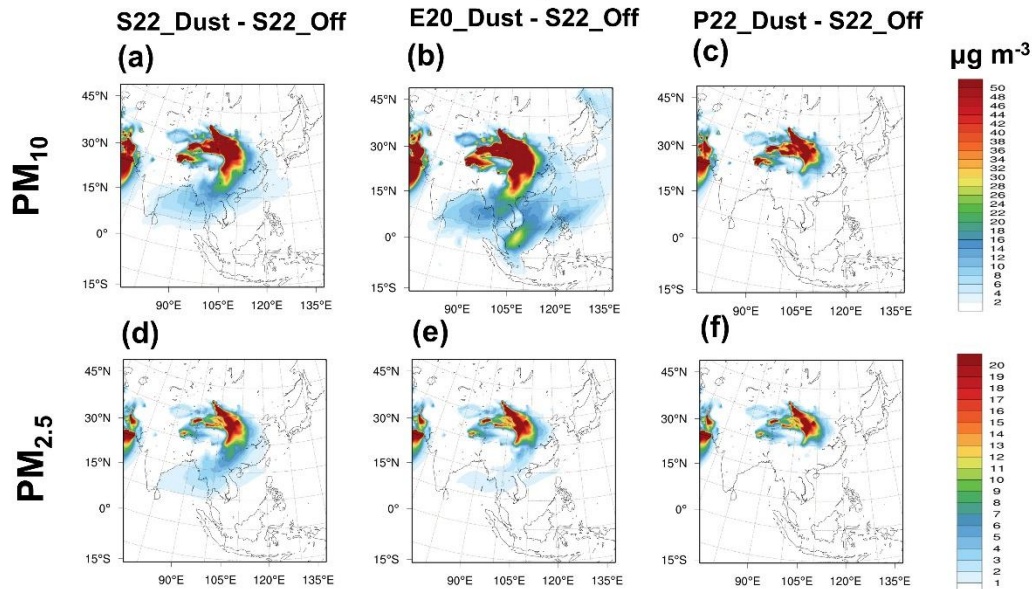
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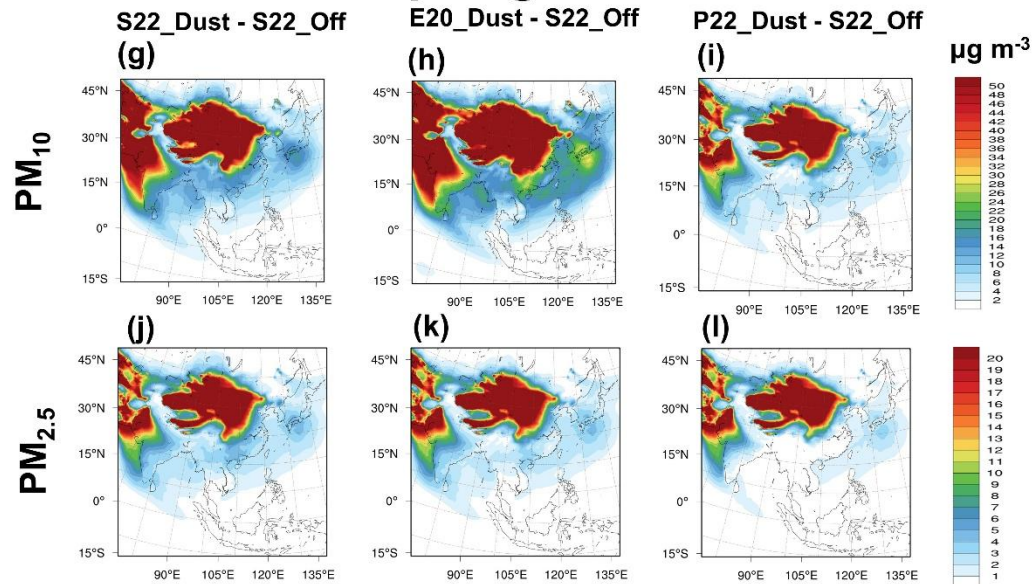
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784 **Figure 5:** The scatter plot of the observed against modeled  $PM_{10}$  (a-d) and  $PM_{2.5}$  (e-h) for  
 785 CMAQ\_Off\_S22 (a, e), CMAQ\_Dust\_S22 (b, f), CMAQ\_Dust\_E20 (c, g) and CMAQ\_Dust\_P22 (d, h)  
 786 at the 100 sites of the mainland China on 12 March-20 April 2021 (<http://>). R is the correlation coefficient  
 787 between the observation and model; FAC2 is the factor of two; MeanOBS and MeanSIM are the mean of  
 788 PM from observation and model, respectively.

# January 2023



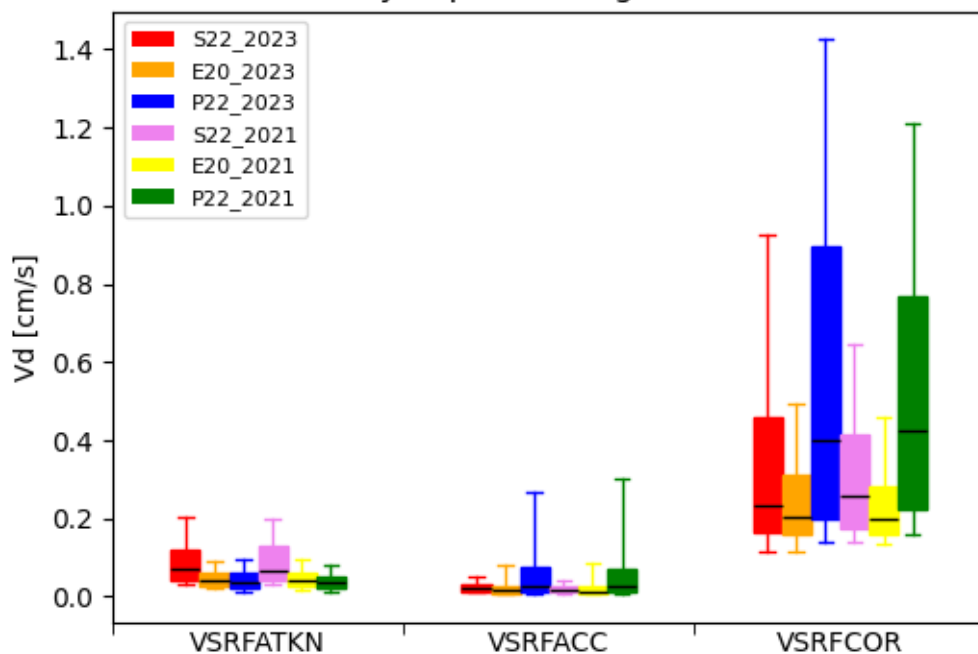
# Spring 2021



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790 **Figure 6:** CMAQ estimated 10 days (January 2023) (a-f) and 40 days (Spring 2021) (g-l) averaged mean  
 791 (a, b, c, g, h, i) PM<sub>10</sub> and (d, e, f, j, k, l) PM<sub>2.5</sub> for the concentration changes using (a, d, g, j) S22, (b, e, h,  
 792 k) E20 and (c, f, i, l) P22 schemes, as relative to the CMAQ\_Off\_S22 scenarios.

## Dry deposition algorithms



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794 **Figure 7:** 10-days (2023) and 40-days (2021) averaged dry  $V_d$  predicted by CMAQ for the Aitken,  
795 accumulation, and coarse particle modes using the 2023\_S22 (red), 2023\_E20 (orange), 2023\_P22 (blue),  
796 2021\_S22 (violet), 2021\_E20 (yellow) and 2021\_P22 (green) particle dry deposition schemes. The  
797 variability illustrated by the boxes and whiskers corresponds to spatial variability in annually averaged  
798 values throughout the CMAQ domain.

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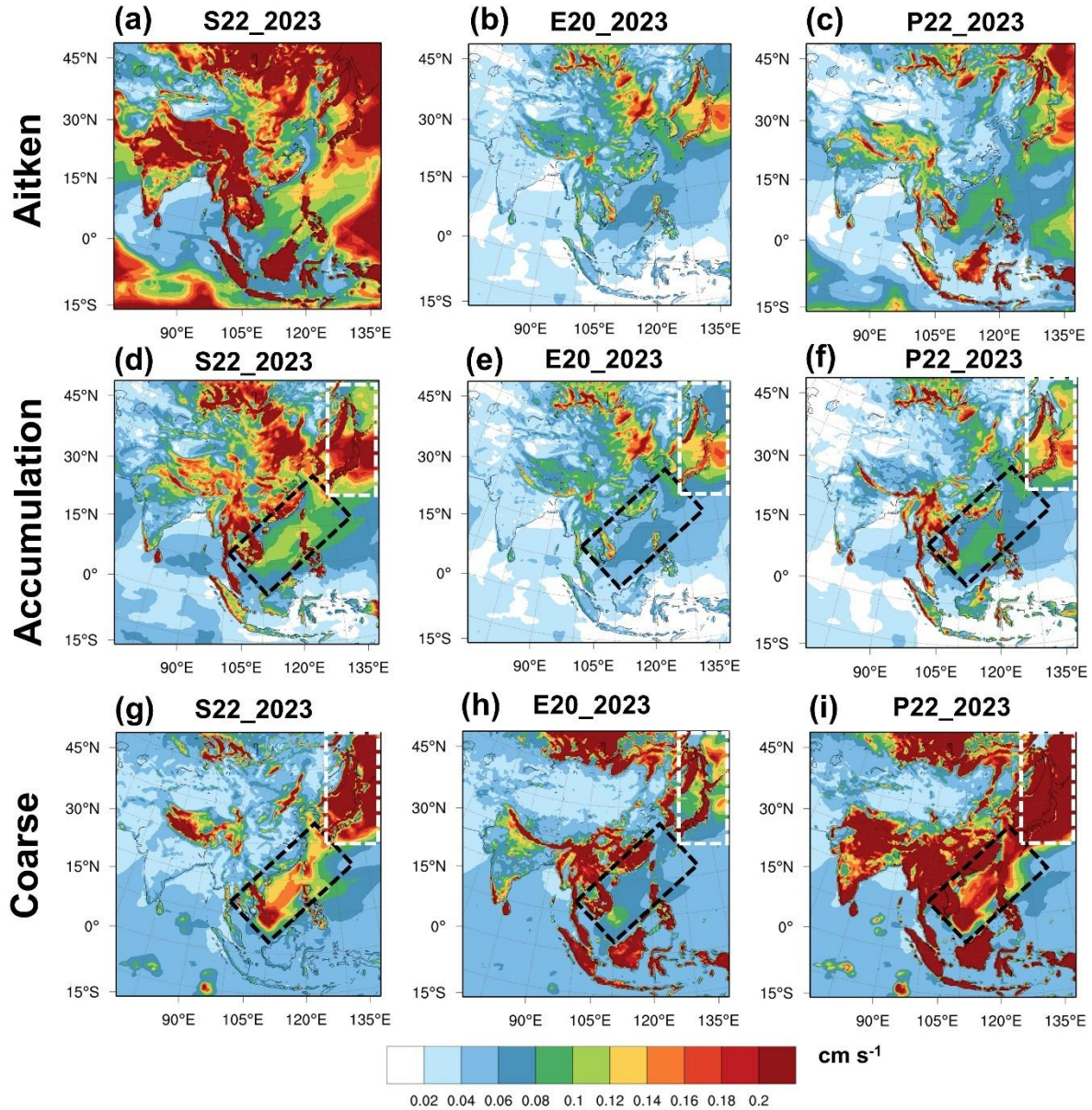
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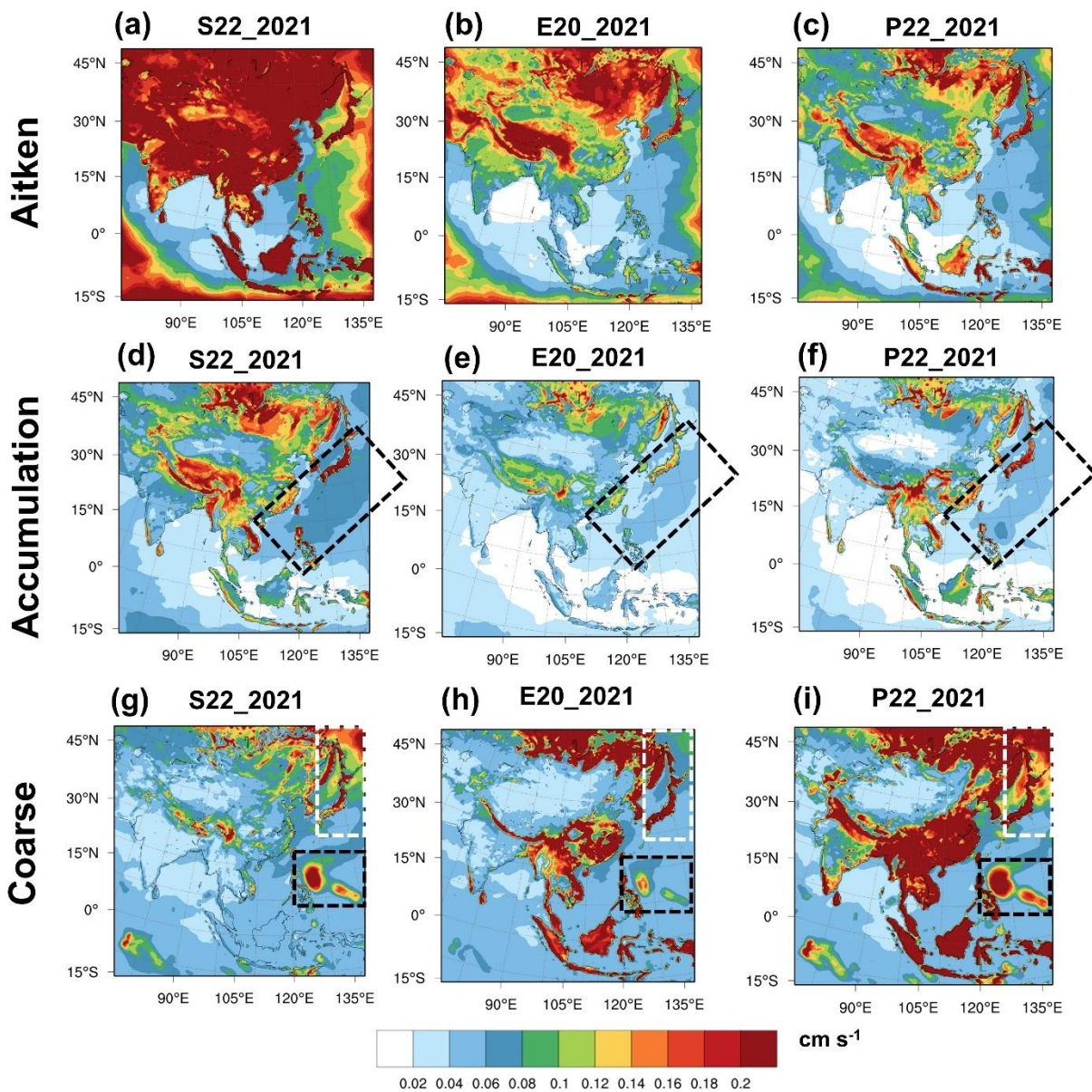
# January 2023



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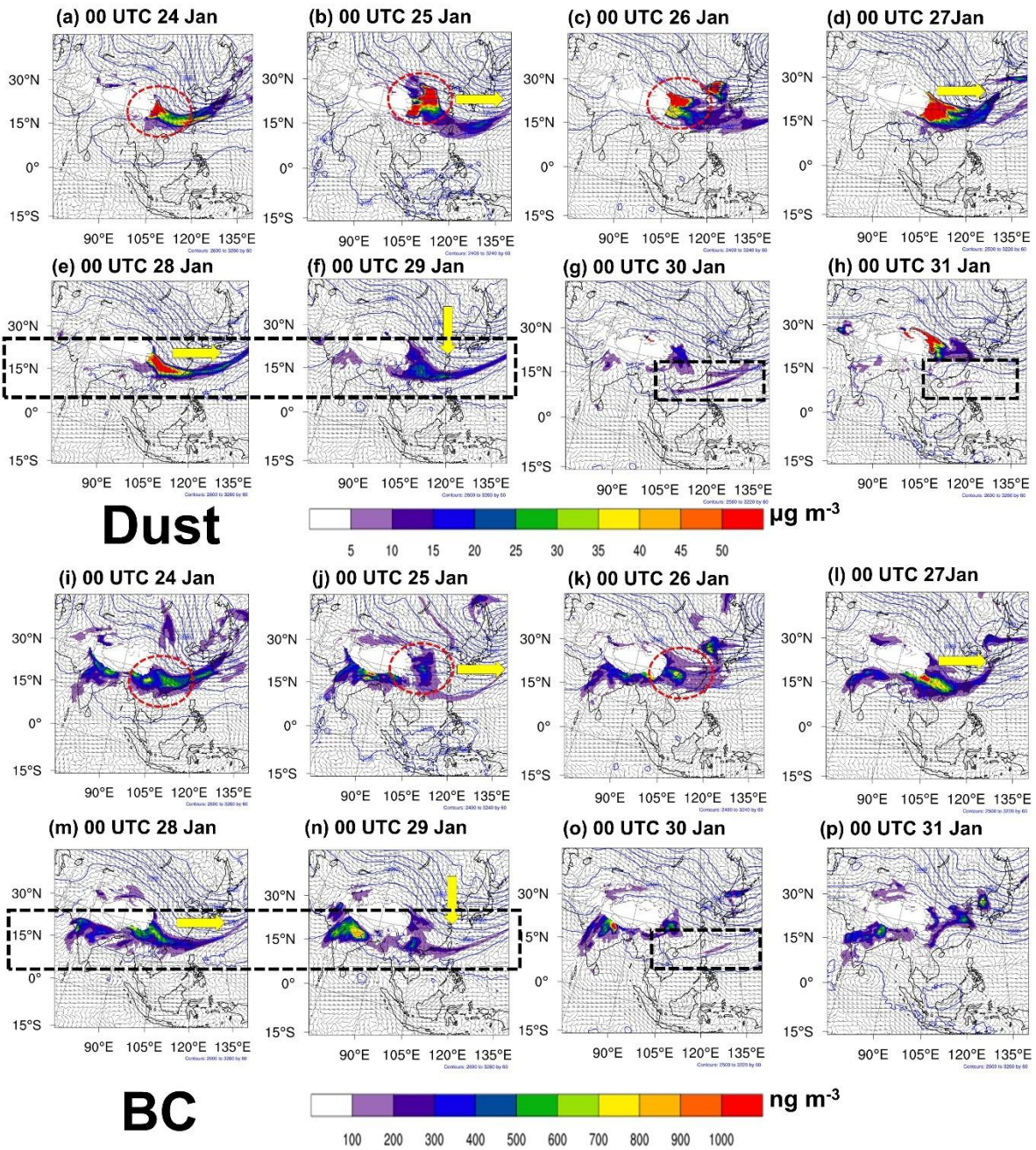
809 **Figure 8:** CMAQ estimated 10 days (22-31 January 2023) averaged for the (a-c) Aitken, (d-f)  
810 accumulation, and (g-i) coarse particle modes for (a, d, g) S22, (b, e, h) E20, and (c, f, i) P22 dry deposition  
811 schemes. White-dash rectangular indicates the region across northwest China; Black-dash rectangular  
812 indicates the marine boundary layer at the western Pacific.

# Spring 2021



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814 **Figure 9:** CMAQ estimated 40 days (12 Mar-20 April 2021) averaged for the (a-c) Aitken, (d-f)  
 815 accumulation, and (g-i) coarse particle modes for (a, d, g) S22, (b, e, h) E20, and (c, f, i) P22 dry deposition  
 816 schemes. White-dash rectangular indicates the region across northwest China; Black-dash rectangular  
 817 indicates the marine boundary layer at the western Pacific.



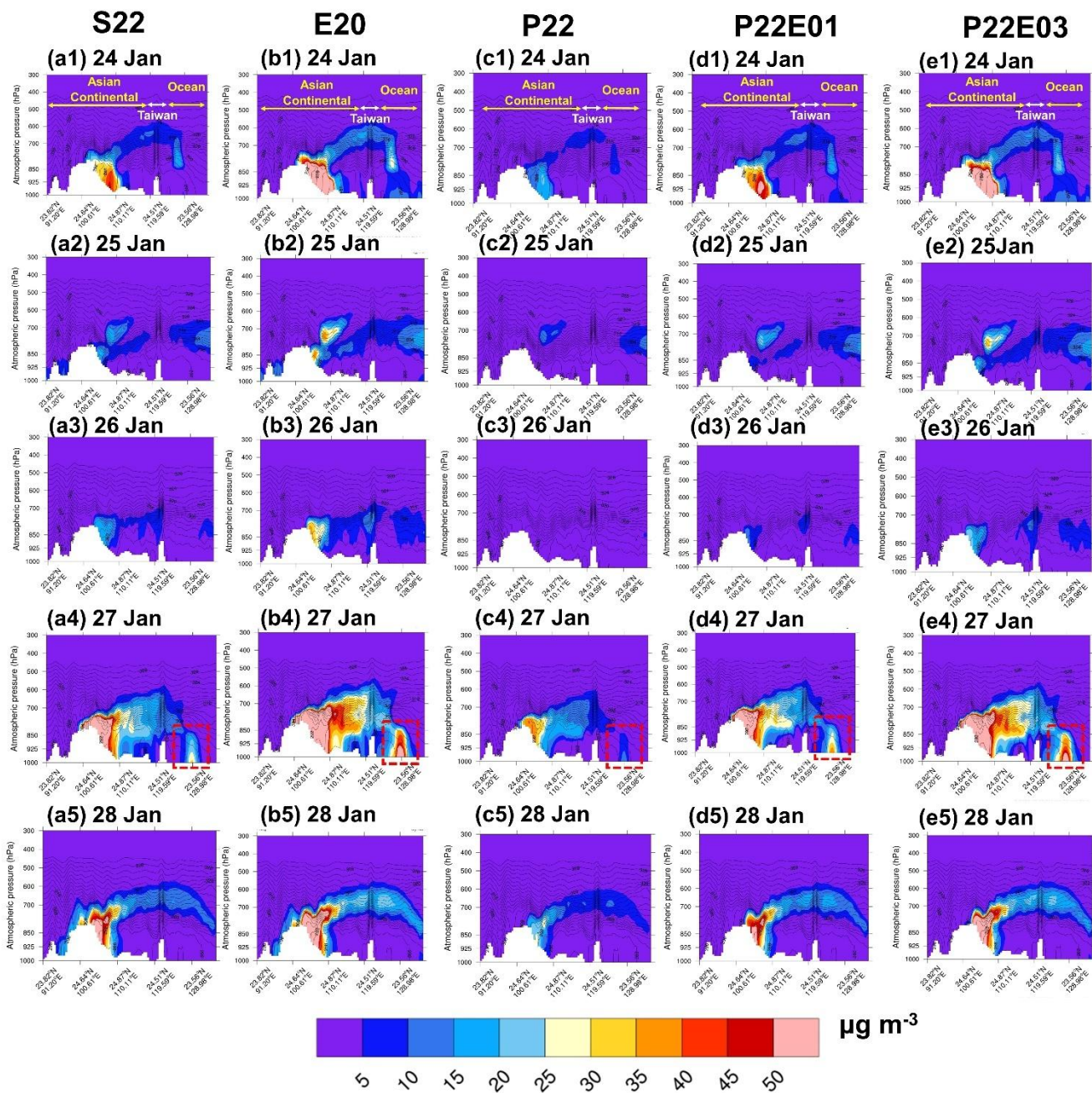
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819 **Figure 10:** CMAQ\_Dust\_E20 simulated mineral dust (a-h) and BC aerosol (i-p) concentrations at the  
 820 700 hPa during 12 UTC 24-31 January 2023. The yellow arrows highlight the trough moving direction.

821 The dash-black rectangular box highlights the aerosol belt.

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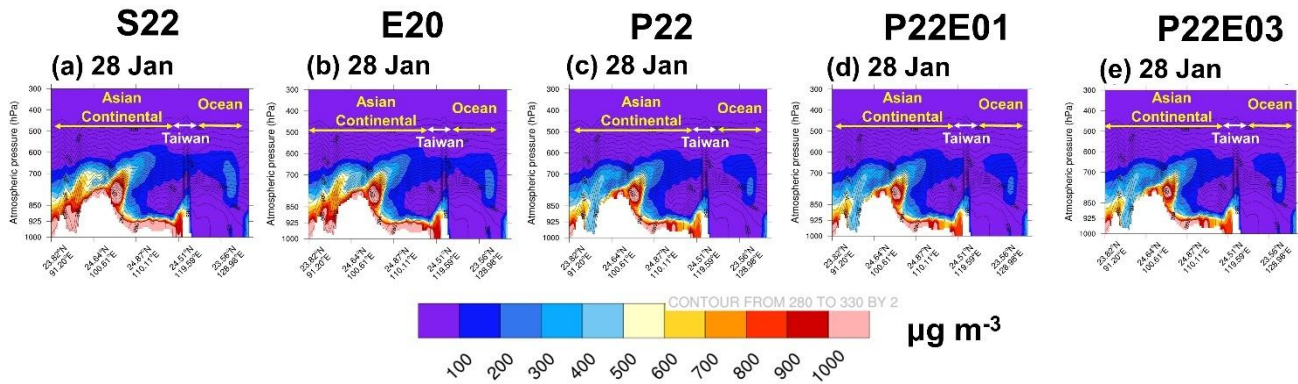


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824 **Figure 11:** Vertical cross section of the simulated dust aerosol for the CMAQ\_DUST (S22, E20, P22,  
 825 P22E01 and P22E03) during 12 UTC 24–28 January 2023.

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829 **Figure 12:** Vertical cross section of the simulated BC aerosol for the CMAQ\_DUST (S22, E20, P22,  
 830 P22E01 and P22E03) during 00 UTC 28 January 2023.

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