

Foliar nutrient uptake from dust sustains plant nutrition

Anton Lokshin^{1,2*}, Daniel Palchan², Elnatan Golan³, Ran Erel³, Daniele Andronico⁴, and Avner Gross¹

1. The Department of Environment, Geoinformatics and Urban planning Sciences, Ben Gurion University of the Negev; Beer Sheva, Israel.
2. The Department of Civil Engineering, Ariel University; Ariel, Israel.
3. Institute of Soil, Water and Environmental Sciences, Gilat Research Center, Agricultural Research Organization; Gilat, Israel.
4. Istituto Nazionale di Geofisica e Vulcanologia, Sezione di Catania-Osservatorio Etneo, Rome, Italy.

*** Corresponding author:** Lokshinanton@gmail.com

Abstract

Mineral nutrient uptake from soil through the roots is considered the main nutrition pathway for vascular terrestrial plants. Recently, desert dust was discovered as an alternative nutrient source to plants, through direct uptake from dust deposited on their foliage. Here we studied the uptake of nutrients from freshly deposited desert and volcanic dusts by chickpea plants under ambient and future elevated levels of atmospheric CO₂, through the roots and directly through the foliage. We found that within weeks, chickpea plants acquired phosphorus (P) from dust only through foliar uptake under ambient conditions, and P, Iron (Fe) and Nickel (Ni) under elevated CO₂ conditions, significantly increasing their growth. Using additional chickpea variety with contrasting leaf properties we have shown that the foliar nutrient uptake pathway from dust is facilitated by leaf surface chemical and physiological traits such as low pH and trichome densities. We analyzed Nd radiogenic isotopes extracted from plant tissues after dust application to assess the contribution of mineral nutrients that were acquired through the foliage. Our results suggest that foliar mineral nutrient uptake from dust is an important pathway, that may play an even bigger role in an elevated CO₂ world.

Keywords: plant nutrition; Nd isotopes; foliage; elevated CO₂

Introduction

Vascular plants obtain carbon (C) from the atmosphere and most of their mineral nutrients mainly from the soil. Hence, it is generally thought that mineral nutrients such as phosphorus (P), potassium (K), iron (Fe), and other macro and micronutrients are acquired predominantly through the plant's roots system (Marschner et al., 1997). Evidence gathered in recent decades demonstrates that the atmosphere is an important source for mineral-nutrients to terrestrial ecosystems via dust deposition (Chadwick et al., 1999; Goll et al., 2023; Gross et al., 2015; Van Langenhove et al., 2020; Okin et al., 2004; Palchan et al., 2018). The concentration of P (and other nutrients) in mineral atmospheric particles such as desert dust and volcanic ash are enriched relative to most soils and are important plant nutrient sources, especially when soil fertility is low or in dusty regions (Arvin et al., 2017; Bauters et al., 2021; Ciriminna et al., 2022; Eger et al., 2013; Gross et al., 2016b). In a montane environment in California, dust P contribution to plants was documented to outpace the contribution from weathering of host bedrock (Arvin et al., 2017). In a recent study we discovered that certain crop plants can gain P directly from the atmospheric dust, via particles that accumulate on their leaves (Gross et al., 2021a; Lokshin et al., 2024b). Over short time scales, foliar uptake was found as the only P uptake pathway from biomass fire ash particles (while the roots played a negligible role (Lokshin et al., 2024a,b). These recent findings highlights the need to better understand the role of the contribution of nutrient uptake from dust through the foliage (i.e., direct foliar nutrient uptake), a process that has been traditionally overlooked and has never been quantified before, even though foliar fertilization has been a well-known agricultural practice for many decades (Fageria et al., 2009; Ishfaq et al., 2022; Bukovac & Wittwer, 1957; Wittwer & Teubner, 1959). Foliar nutrient uptake occurs through two primary pathways: cuticle penetration and stomatal uptake. The cuticle, while largely hydrophobic, contains aqueous pathways that allow the diffusion of small, polar molecules, particularly under high humidity conditions (Schonherr, 2006). Stomata, which regulate gas exchange, can also act as entry points for hydrophilic solutes and small particles when they are open (Fernández and Eichert, 2009). These pathways can expand or contract dynamically in response to environmental factors, enabling at times solute penetration. Minerals are solid nutrient particles hence need to partially dissolve into the aqueous film on the leaf surface before uptake by the plant. This dissolution can be facilitated by surface moisture, leaf exudates, and microbial activity in the phyllosphere, which enhances the solubility and bioavailability of nutrients (Burkhardt et al., 2012; Fernández et al., 2014, Marschner, 2022).

In the context of climate change, the foliar pathway may be even more pronounced for plants that will grow under elevated CO₂ (eCO₂) conditions because of two documented phenomena: the 'dilution' effect, where accumulation of C exceeds that of mineral nutrients which can lead to stoichiometric imbalance (Loladze, 2014), and partial inhibition of key root uptake mechanisms (Gojon et al., 2023), together with soil fertility degradation (Lal, 2009; St. Clair and Lynch, 2010). These changes may lead to the selection of plant traits that facilitate alternative nutrient uptake pathways. The use of the foliar pathway under eCO₂ may offset the alarming phenomenon where an increasing production of carbohydrates causes dilutes the concentration of macro and micronutrients such as P, Fe, calcium (Ca), magnesium (Mg), K, zinc (Zn), copper (Cu), nickel (Ni) and others that are vital for the floral ecological systems (Clarkson and Hanson, 1980).

In this work we designed two experiments to study both the principal mechanisms and biological functions in the plants that benefit foliar nutrient uptake from dust retained on the foliage, and the chemical composition of the

nutrients transferred in this process. Furthermore, we utilized Nd isotopes to attempt quantification of the ion flux from the foliage to the plant. Then we discuss the larger aspect of this newly discovered pathway and its importance.

2 Materials and Methods

2.1 Experimental design

To study the impact of dust deposition on plant nutrition, we selected two contrasting chickpea genotypes (*Cicer* spp.) from the Hebrew University of Jerusalem chickpea collection. These genotypes were chosen based on previous studies (Gross et al., 2021) that demonstrated differences in their response to foliar dust application. The first genotype, ‘CR934,’ is a non-responsive genotype of the wild progenitor *C. reticulatum*, sampled near Savur, Turkey, showing minimal physiological or nutritional changes following dust application. In contrast, the second genotype, ‘Zehavit,’ is a modern high-yield cultivar widely used by Israeli growers, which exhibits a pronounced response to foliar dust application, such as increase in biomass and total P content. For further biogeochemical analysis of foliar nutrient uptake, the responsive genotype ‘Zehavit’ was used. Experiments were conducted at Gilat Research Center in southern Israel (31°21’ N, 34°42’ E) in two separate glasshouse rooms. Temperature was fixed at 25 ± 3°C and relative humidity at 40–50%. Inside the glasshouse the pots were subjected to natural lighting partially concealed by transparent white walls and roof. Overall, the Photosynthetically Active Radiation (PAR) levels were typical for the southern part of Israel during the months of September to November. In one room we set the CO₂ concentration to the ambient 412 ppm (aCO₂) and in the other room to elevated 850 ppm (eCO₂), simulating current and future earth CO₂ concentrations based on high emissions scenario (business as usual, SSP 8.5, IPCC, 2021). Following germination, plants were cultivated in 72 pots containing inert media (perlite 206, particle size of 0.075–1.5 mm; Agrekal, HaBonim, Israel). The pot size was 3 liter, with sufficient room for root growth during the experimental period. All the pots were supplied with a nutrition solution (fertigation) containing the following elements: nitrogen (N) (50 mg L⁻¹), P (3.5 mg L⁻¹), K (50 mg L⁻¹), Ca (40 mg L⁻¹), Mg (10 mg L⁻¹), Fe (0.8 mg L⁻¹), Mn (0.4 mg L⁻¹), Zn (0.2 mg L⁻¹), boron (B) (0.4 mg L⁻¹), Cu (0.3 mg L⁻¹) and molybdenum (Mo) (0.2 mg L⁻¹). The mineral concentrations were achieved by proportionally dissolving NH₄NO₃, KH₂PO₄, KNO₃, MgSO₄ and NaNO₃. The micronutrients were applied as EDTA (ethylenediaminetetraacetic acid) chelates as commercial liquid fertilizer (Koratin, ICL Ltd). The location of each pot within the glasshouse was randomized at the beginning and changed every two weeks over the course of the experiment. The plants were dripped irrigated 4 times per day for 5 minutes, via an automated irrigation system from the germination stage. At 14 days after germination, when plants were early in the vegetative phase (two or three developed leaves), we changed the nutrient solution of 60 out of the 72 pots to P deficient fertigation (P concentration of 0.1 mg L⁻¹) to create P starvation (-P treatment). Preliminary tests showed that our -P deficient media allows chickpea plants to continue their growth cycle and increase their responsiveness for dust application, reflected in physiological, morphological, and biochemical changes and eCO₂ condition (Gross et al., 2021, Lokshin et al., 2024). The remaining 12 pots continued to receive the full P sufficient nutrient media (+P treatment). Plants fertigated with -P solution started to show P-deficiency symptoms such as chlorosis of mature leaves, slight symptoms of necrotic leaf tips and an overall decrease in biomass accumulation at 35 days after germination. At this stage we applied desert dust and volcanic ash on the -P plants.

Of a total number of plants (72) 48 were treated with dust and 24 served as untreated control group. Twenty-four plants were applied with dust on their foliage by manually sprinkling dust through a 63 μ m sieve in proximity to the foliage and 24 plants received root treatment by applying dust through a 63 μ m sieve on the surface of the pot, followed by gentle mixing of the surface to sink the dust particles deeper to enhance the physical contact between the roots and the particles, thereby increasing the chances of having a more significant impact. The experimental design is a 3-factorial design, considering the effects of dust type (desert or volcanic dust), application method (foliar or root), and CO₂ concentration (412 ppm or 850 ppm). This information is further explained in Table 1, which summarizes the treatments and experimental groups.

Table 1: A summary of the treatments and experimental groups.

Treatment Group	P Concentration	Dust Application Method	CO ₂ Condition	Replicates (n)
-P treatment (P-deficient)	0.1 mg/L	Foliar (Desert Dust)	412 ppm	6
-P treatment (P-deficient)	0.1 mg/L	Foliar (Volcanic Ash)	412 ppm	6
-P treatment (P-deficient)	0.1 mg/L	Root (Desert Dust)	412 ppm	6
-P treatment (P-deficient)	0.1 mg/L	Root (Volcanic Ash)	412 ppm	6
+P treatment (P-sufficient)	Full P	No Dust	412 ppm	6
-P control (untreated)	0.1 mg/L	No Dust	412 ppm	6
-P treatment (P-deficient)	0.1 mg/L	Foliar (Desert Dust)	850 ppm	6
-P treatment (P-deficient)	0.1 mg/L	Foliar (Volcanic Ash)	850 ppm	6
-P treatment (P-deficient)	0.1 mg/L	Root (Desert Dust)	850 ppm	6
-P treatment (P-deficient)	0.1 mg/L	Root (Volcanic Ash)	850 ppm	6
+P treatment (P-sufficient)	Full P	No Dust	850 ppm	6
-P control (untreated)	0.1 mg/L	No Dust	850 ppm	6

Table 1: Experimental design summarizing treatments, P concentrations, dust application, CO₂ conditions, and replicates.

To mimic dust deposition which typically occurs during a few major desert storms or volcanic eruptions each year, we applied the dust in two equivalent doses between 35-42 days after germination. We calculated the average foliar area of the chickpea pots, taking into consideration the planting density, and correcting values for the area covered by an individual plants. These values were used to determine the total application mass. In total, the average application mass was 3 g per pot, simulating the total dust deposition per square meter over the growing season in southern Israel. This method was based on our previous studies (Gross et al, 2021, Starr et al, 2023,

Lokshin et al, 2024). The dust treatments were done either directly on the foliage while covering the pot, preventing the dust from touching the roots, or directly on the roots where the pots were subsequently covered with nylon to equalize conditions with the foliage treated plants. Afterwards, the plants were left undisturbed with the settled dust particles on their foliage or surface of the root area.

Among the control plants, 12 plants received the +P fertigation and 12 additional plants received -P fertigation. Each treatment group was divided into two CO₂ levels, 36 plants in each CO₂ growing room. The plants were harvested 10 days after the last dust application (55 days after germination). To ensure that nutrients from dust particles were not washed by the irrigation during the experiment, we monitored the total P (i.e., P that dissolves in strong acid) in the water that drained from the pots (Longo et al., 2014; Gross et al., 2015) throughout the experiment.

We performed a parallel experiment under aCO₂ where we grew six additional plants, in larger 5 L pots, filled with soil, to test whether our findings also apply to natural soil conditions (Fig. S1).

2.2 Mineral dust material

We applied plant foliage and the area near plants' roots, with desert dust and volcanic ash, the two main mineral dust types in the atmosphere (Langmann, 2013b). To achieve enough mass for our experiment, we produced dust analogs from surface desert soil and surface volcanic ash soil, following procedures described by others (Gross et al., 2021; Stockdale et al., 2016). The desert dust analog surface soil was collected from the southern Israel Negev desert (30°32'N 34°55'E) (Gross et al., 2021). Chemical and mineralogical properties of the resulted dust are comparable to dust collected in the Sahara and other places in the Middle East (Gross et al., 2016a; Palchan et al., 2018). The volcanic ash analog was collected from Mount Etna (Sicily, Italy) two months after the eruption of 21 February 2022. The ash was taken from the upper cable car station "Funivia dell'Etna" (37°70'N, 14°99'E). The samples were then processed through a setup of sieves to achieve a particle size smaller than 63µm that are considered windblown (Guieu et al., 2010). The chemical and mineralogical properties of the dust analogs are presented in Table 2.

2.3 Plant biomass and elemental analysis

After harvesting, the plants were separated for roots and shoots, washed in 0.1M HCl and rinsed three times in distilled water to remove dust particle residue (Gross et al., 2021; Lokshin et al., 2024a). To ensure that the washing procedure removed all the applied dust particles from the leaf surfaces, we scanned surfaces of randomly selected dusted and washed leaves with SEM-EDS which combines scanning electron microscope and energy-dispersive X-ray spectroscopy to detect and analyze materials. After washing, plant tissue was dried, weighed and root and shoot biomass were recorded. Afterwards, the dry shoot material was ground to powder and dry ashed at 550 C° in a furnace for four hours (Tiwari et al., 2022). Approximately 1g of the ashed material was subsequently dissolved using 1 mL concentrated HNO₃ to achieve a clear solution. To prepare the dust types for elemental analysis, the samples were dissolved on a hotplate by sequential dissolution using concentrated HNO₃, HF, and HCl, resulting in clear solutions (Palchan et al., 2018). The elemental composition of the plants, dusts and nutrient

solution were analyzed at the Hebrew University using ICP-MS (Agilent 8900cx; Agilent Technology). Prior to analysis, the ICP-MS was calibrated with a series of multi-element standard solutions (1 pg/mL - 100 ng/mL Merck ME VI) and standards of major metals (300 ng/mL - 3 mg/mL). Internal standard (50 ng/mL Sc and 5 ng/mL Re and Rh) was added to every standard and sample for drift correction. Standard reference solutions (USGS SRS T-207, T-209) were examined at the beginning and end of the calibration to determine accuracy. The calculated accuracies for the major and trace elements are 3% and 2%, respectively.

2.4 Leaf surface pH

Leaf surface pH was measured by manually attaching a portable pH electrode designed for flat surfaces (HI-1413; HANNA pH instruments) onto the surface of three leaves from each plant. The measurements were performed four times throughout the growing season (19, 24, 35 and 40 DAG) in the morning, two hours after sunrise.

2.5 Trichome density

Trichome density was determined in four young, fully developed leaves from four different plants per variety in the P- treatment only (n=16). Leaves were scanned in a scanning electron microscope (VEGA3; Tescan, Czech Republic). From each leaf, three photos of a 1mm² field were taken, and glandular and regular trichomes were counted.

2.6 Leaf exudates

For analysis of the organic exudates, 2g of fresh leaves were sampled randomly from the P+ and P- treatments before harvesting. The leaves were rinsed in 2 ml of distilled water and methanol (50:50) for 10 s. The extracted surface metabolites were supplied with 50 µL of internal standard (ribitol, 0.2 mg mL⁻¹) and stored at -80°C until analysis. Before analysis, the extracted samples were vacuum dried overnight at 35°C. The dried material was redissolved in 40 µL of 20 mg mL⁻¹ methoxamine hydrochloride (CH₃ONH₂ HCl) in pyridine (C₅H₅N) and derivatized for 90 min at 37°C, followed by a spike of 70 µL MSTFA (*N*-methyl-*N* (trimethylsilyl) trifluoroacetamide (CF₃CON(CH₃)Si(CH₃)₃) at 37°C for 30 min. The dissolved metabolites were then introduced to a mass spectrometry gas chromatograph (Agilent 6850 GC/5795C; Agilent Technology) for analysis. The metabolites were detected by a mass spectrometer, where 1 µL of each sample was injected in split-less mode at 230°C to a helium carrier gas at a flow rate of 0.6 mL min⁻¹. GC processing was carried out using an HP-5MS capillary column (30 m 9 0.250 mm 9 0.25 µm) and the spectrum was scanned for *m/z* 50–550 at 2.4 Hz. The ion chromatograms and mass spectra obtained were evaluated using the MSD CHEMSTATION (E.02.00.493) software, and sugars and amino acids were identified via comparison of retention times and mass spectra with certified GC plant metabolite standards (Sigma Aldrich).

2.7 Nd isotopes

We characterized the radiogenic Nd isotope compositions of desert dust, volcanic ash, and the different treated and control plants to evaluate the ion flux from the foliage into the plant. The use of Nd isotopes is common in

fingerprinting studies in various fields (Aciego et al., 2017; Arvin et al., 2017; Chadwick et al., 1999) as it is a refractory element that has virtually no fractionation and reflects the composition of its source. In our experiments the control plants reflect the original composition of the plant and deviation from it can only occur if ions were supplemented through nutrient uptake. Nd was extracted by ion chromatography using specialized resins, first TRU resin to separate the REEs followed by LN-spec to separate Nd from Sm (following Palchan et al., 2013). The isotope ratios were measured using a Thermo Neptune multi-collector ICP-mass spectrometer at the Weizmann Institute of Science. Along with the samples we measured several standards to ensure quality and accuracy of the measurements. Standard JNdi bracketed every five samples, resulting with $^{143}\text{Nd}/^{144}\text{Nd}$ value of 0.512035 ± 10 (2σ , $n=60$). We normalized the data to $^{143}\text{Nd}/^{144}\text{Nd} = 0.512115$ (Tanaka et al., 2000). Standard BCR-2 was dissolved and analyzed along with the plant and dust samples yielding $^{143}\text{Nd}/^{144}\text{Nd}$ value of 0.512628 ± 6 (2σ , $n=3$) that agrees with $^{143}\text{Nd}/^{144}\text{Nd} = 0.512637 \pm 13$ (Jweda et al., 2016). We present the Nd isotope ratios in the epsilon notation:

$$\epsilon_{Nd} = \left(\frac{\left(\frac{^{143}\text{Nd}}{^{144}\text{Nd}} \right)_{\text{Sample}}}{\left(\frac{^{143}\text{Nd}}{^{144}\text{Nd}} \right)_{\text{CHUR}}} - 1 \right) \times 10,000$$

where the value of $^{143}\text{Nd}/^{144}\text{Nd} = 0.512638$ in CHUR (Wasserburg et al., 1981). A sample isotopic characterization is given in SI Table 4. We calculated the % of foliar contribution using simple mixing equation of two components:

$$\% \text{ Foliar contribution} = \frac{\epsilon Nd_{\text{sample}} - \epsilon Nd_{\text{control}}}{\epsilon Nd_{\text{end member}} - \epsilon Nd_{\text{control}}} * 100$$

where $\epsilon Nd_{\text{sample}}$ refers to plants that were treated either with desert dust or volcanic ash with, $\epsilon Nd_{\text{control}}$ refers to the untreated control plants and $\epsilon Nd_{\text{end member}}$ are the measured end member values of -10.3 for desert dust or 4.5 value for volcanic ash (Table SI-4 & Fig. 4). The calculation was done separately on the two different treatments (i.e., desert dust and volcanic ash).

2.8 Mineralogical analysis

Mineralogical composition of the dusts was determined with an X-ray powder diffraction (XRD) using a Panalytical Empyrean Powder Diffractometer equipped with a position sensitive X'Celerator detector. Cu K α radiation ($k = 1.54178_{\text{\AA}}$) at 40 kV and 30 mA. Scans were done over a 2h period, between 5° and 65° with an approximate step size of 0.033°.

2.9 Statistical Analysis

Treatment comparisons for all measured parameters were tested using post-hoc Tukey honest significant difference (HSD) tests ($P < 0.05$). The significant differences are denoted using different letters in the figures.

The standard errors of the mean in the vertical bars (in the figures) were calculated using GraphPad Prism version 9.0.0.

3 Results

3.1 Plant biomass and total P under aCO₂ and eCO₂

The P contribution in the dusted plants was expressed in increased aboveground plant biomass rather than in increased P concentrations in tissues (Table 2), probably as plants distributed the extra P to support growing biomass.. Thus, the treatment effects are reflected by changes in total plant P (concentration multiplied by shoot biomass). The impact of desert and volcanic dust application on plants' foliage was reflected by the increase of their total P content through shoot biomass gain rather than through changes in shoot P concentration. Under aCO₂ conditions, desert dust application resulted in shoot biomass and total P content increases of 35% and 21%, respectively, and volcanic ash application resulted in 28% and 35% increases, respectively (Fig. 1 d,f). The root-treated plants did not show any increases in the shoot biomass or total P content (Fig. 1 c,e). These trends are also seen in the eCO₂ conditions of 850 ppm atmospheric CO₂ experiment. Desert dust application resulted in shoot biomass and total P content increases of 29% and 20%, respectively, and volcanic ash application resulted in 62% and 51% shoot biomass increases, respectively (Fig. 2 d, f). Similarly, the root-treated plants did not show any increases in the shoot biomass or total P content (Fig. 2 c, e). Unlike the shoots, no significant changes of the biomass of the roots were detected across all treatments, thus changes in the root shoot ratio reflect variations in shoot biomass rather than root biomass (Table 2).

Root treatment 412 ppm

Foliar treatment 412 ppm

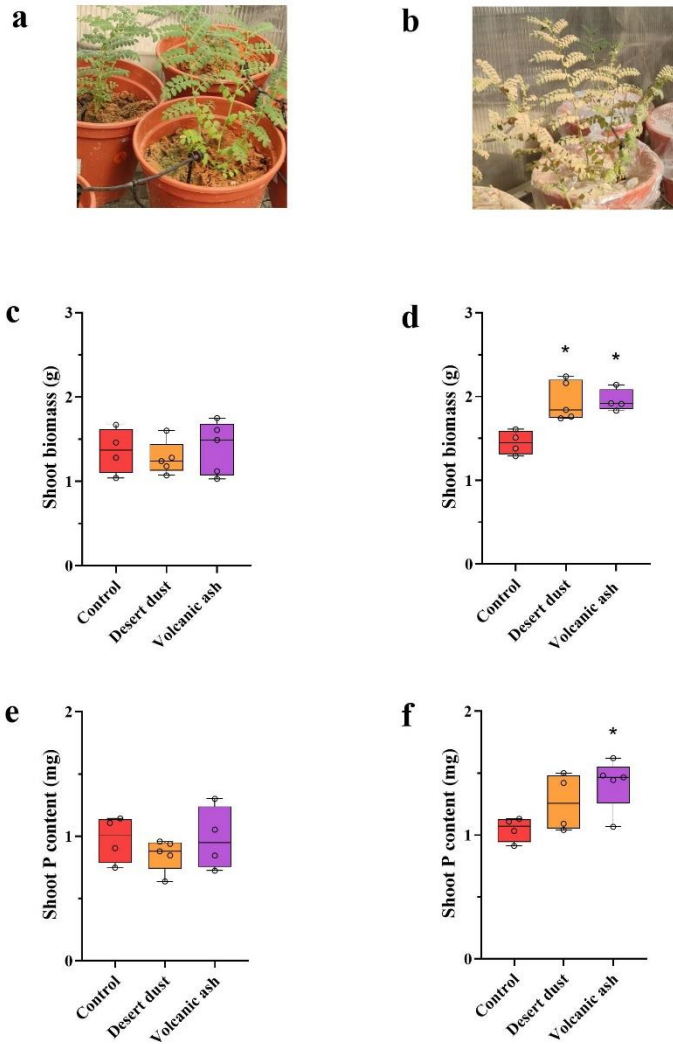
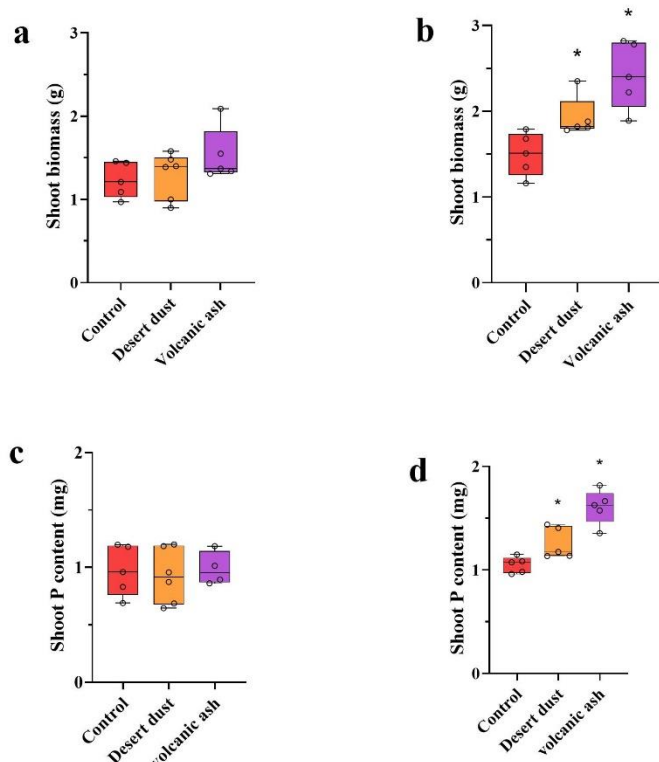


Figure 1 (a-f): Biomass and total P content increases due to dust application treatments at aCO₂ of 412ppm. (a) Image of experiment setting of the root treatment. Image was taken immediately after the dust application. (The actual amount of dust remaining on the plant leaves at the end of the experiment was significantly smaller than what is depicted in the picture image). (b) Image of experiment setting of foliar treatment. (c) Shoot biomass of root treated plants. (d) Shoot biomass of foliar treated plants. (e) Shoot total P content of root treated plants. (f) Shoot total P content of foliar treated plants. The red color represents control plants, orange desert dust treatment and purple volcanic ash treatment. Asterisks represent statistically significant differences between bars (P<0.05, Tukey test). The box and whiskers plots represent the distribution of the data. The central line indicates the median, the edges of the box correspond to the 25th (Q1) and 75th (Q3) percentiles, and the whiskers span from the minimum to the maximum values. Individual data points (n = 5) are overlaid on the plot to illustrate the full distribution.

Root treatment 850 ppm

Foliar treatment 850 ppm



337

Figure 2 (a-d): Biomass and total P content increases due to dust application treatments at eCO₂ of 850ppm. (a) Shoot biomass of root treated plants. (b) shoot biomass of foliar treated plants. (c) Shoot total P content of root treated plants. (d) Shoot P content of foliar treated plants. The biomass and total P content in the root treated plants do not show increases compared with the control groups. The red color represents control plants, orange desert dust treatment and purple volcanic ash treatment. Asterisks represent statistically significant differences between bars (P<0.05, Tukey test). The box and whiskers plots represent the distribution of the data. The central line indicates the median, the edges of the box correspond to the 25th (Q1) and 75th (Q3) percentiles, and the whiskers span from the minimum to the maximum values. Individual data points (n = 5) are overlaid on the plot to illustrate the full distribution.

3.2 Elemental analysis of the plants

The concentrations of selected micro- and macronutrients that comprise the plant ionome, along with shoot biomass, are presented in Table 2. This table includes data for the control groups (-P and +P) as well as the treatment groups. The results, based on a post hoc Tukey test, show that +P plants are significantly larger than -P plants. Among +P plants, those grown under elevated eCO₂ conditions are significantly larger than those grown under ambient aCO₂ conditions ($p < 0.001$). In contrast, -P plants grown under eCO₂ exhibit similar biomass to -P plants grown under aCO₂ ($p = 0.99$). A detailed breakdown of elemental concentrations in all the treated plants based on ICP-MS analysis, is provided in Supplementary Table S2.

Table 2 Total elemental analysis of plants (*Cicer arietinum* cv. ‘Zehavit’), fertilizers, and dust (ICP-MS analysis). The concentrations of micro- and macronutrients are presented in µg/g or mg/g, respectively, while plant biomass is shown in grams. The p-values refer to biomass comparisons. For aCO₂ treatments, all groups were compared to the -P control grown under aCO₂, and for eCO₂ treatments, all groups were compared to the -P control grown under eCO₂.

Plant material	Shoot biomass (g)	P_value Biomass	Root biomass (g)	Root/Shoot ratio	Mg (mg/g)	P (µg/g)	K (mg/g)	Ca (mg/g)	Mn (µg/g)	Fe (µg/g)	Ni (µg/g)	Cu (µg/g)	Zn (µg/g)
Control '-P' 412 ppm CO ₂ <u>average value</u>	1.36		1.17	0.86	2.75	726	21.70	7.24	37	90	1.7	4.7	23
Control '-P' 412 ppm CO ₂ <u>standard deviation</u>	0.22		0.22	0.07	0.14	78	1.49	0.43	12	17	0.7	3.4	0.7
Control '-P' 850 CO ₂ ppm <u>average value</u>	1.50		1.22	0.82	3.05	712	24.54	7.53	37	79	1.3	2.9	28.5
Control '-P' 850 CO ₂ ppm <u>standard deviation</u>	0.25		0.20	0.05	0.50	89	3.14	0.90	18	11	0.8	1.0	6.1
'-P' + foliar application of desert dust 412 ppm CO ₂ - <u>average value</u>	1.95	0.004	1.57	0.81	2.42	833	20.02	6.64	34	108	0.9	2.7	21.0
'-P' + foliar application of desert dust 412 ppm CO ₂ - <u>standard deviation</u>	0.24		0.21	0.08	0.15	366	2.77	0.37	9	12	0.2	0.5	1.5
'-P' + foliar application of desert dust 850 ppm CO ₂ - <u>average value</u>	1.93	0.013	1.55	0.81	2.48	663	22.10	7.25	36	145	2.4	3.3	20.8
'-P' + foliar application of desert dust 850 ppm CO ₂ - <u>standard deviation</u>	0.24		0.26	0.16	0.41	125	2.08	1.03	6	17	0.6	0.3	3.1
'-P' + foliar application of volcanic ash 412 ppm CO ₂ - <u>average value</u>	1.84	0.0055	1.39	0.75	2.66	769	21.73	7.12	44	188	1.1	3.2	22.3
'-P' + foliar application of volcanic ash 412 ppm CO ₂ - <u>standard deviation</u>	0.27		0.27	0.04	0.15	57	1.98	0.72	7	74	0.4	0.3	2.9
'-P' + foliar application of volcanic ash 850 ppm CO ₂ - <u>average value</u>	2.42	0.0008	2.05	0.84	2.53	674	23.54	6.89	48	203	1.0	3.2	23.1
'-P' + foliar application of volcanic ash 850 ppm CO ₂ - <u>standard deviation</u>	0.39		0.44	0.07	0.19	105	1.69	0.24	12	97	0.8	0.2	4.0
Control '+P' 412 ppm CO ₂ <u>average value</u>	10.74	<0.0001	4.01	0.37	4.55	2123	26.96	8.10	55	105.6	0.6	4.4	38.6
Control '+P' 412 ppm CO ₂ <u>standard deviation</u>	0.71		0.66	0.04	1.28	225	2.83	1.06	16	28.9	0.3	0.4	8.0
Control '+P' 850 ppm CO ₂ <u>average value</u>	15.99	<0.0001	7.98	0.50	4.02	2417	27.41	9.05	74	99	0.8	4.7	36.3
Control '+P' 850 ppm CO ₂ <u>standard deviation</u>	1.86		1.94	0.10	1.01	363	2.22	1.62	18	7.7	0.1	0.9	9.5
Fertilizers and dusts (ppm)					Mg (mg/g)	P (µg/g)	K (mg/g)	Ca (mg/g)	Mn (mg/g)	Fe (µg/g)	Ni (µg/g)	Cu (µg/g)	Zn (µg/g)
+P fertilizer					1.26	713	6.00	0.010	0.08	0.15	0.4	5.6	50.1
-P fertilizer					1.21	35	7.80	0.007	0.07	0.13	0.4	5.1	47.5
Desert dust					6.50	1387	8.60	136	0.24	12.74	18.0	10.0	39.0
Volcanic ash #1					23.50	1669	12.50	64.4	1.10	63.70	49.5	118.6	80.0
Volcanic ash #2					22.60	1788	12.00	61.6	1.06	61.90	48.3	115.6	73.7

3.3 Biochemical properties of chickpea varieties

The domesticated variety 'Zehavit' showed a strong response to the foliar treatment with up 35% increased biomass compared to the control group, whereas the wild variety CR934 showed up to 5% increases compared with the control group (Fig. 3a). The leaf pH of the Zehavit was 1.15 and of the CR934 it was 2.7 (Fig. 3b), trichome density, both glandular and non-glandular, were higher in the Zehavit compared to the CR934 (Fig. 3c-e). The exudates of oxalic, malic, and citric acids were significantly higher at the Zehavit in comparison to CR934 (Fig. 3f). The results indicate increased biomass, lower pH, higher trichome density, and higher exudate levels in the 'Zehavit' variety.

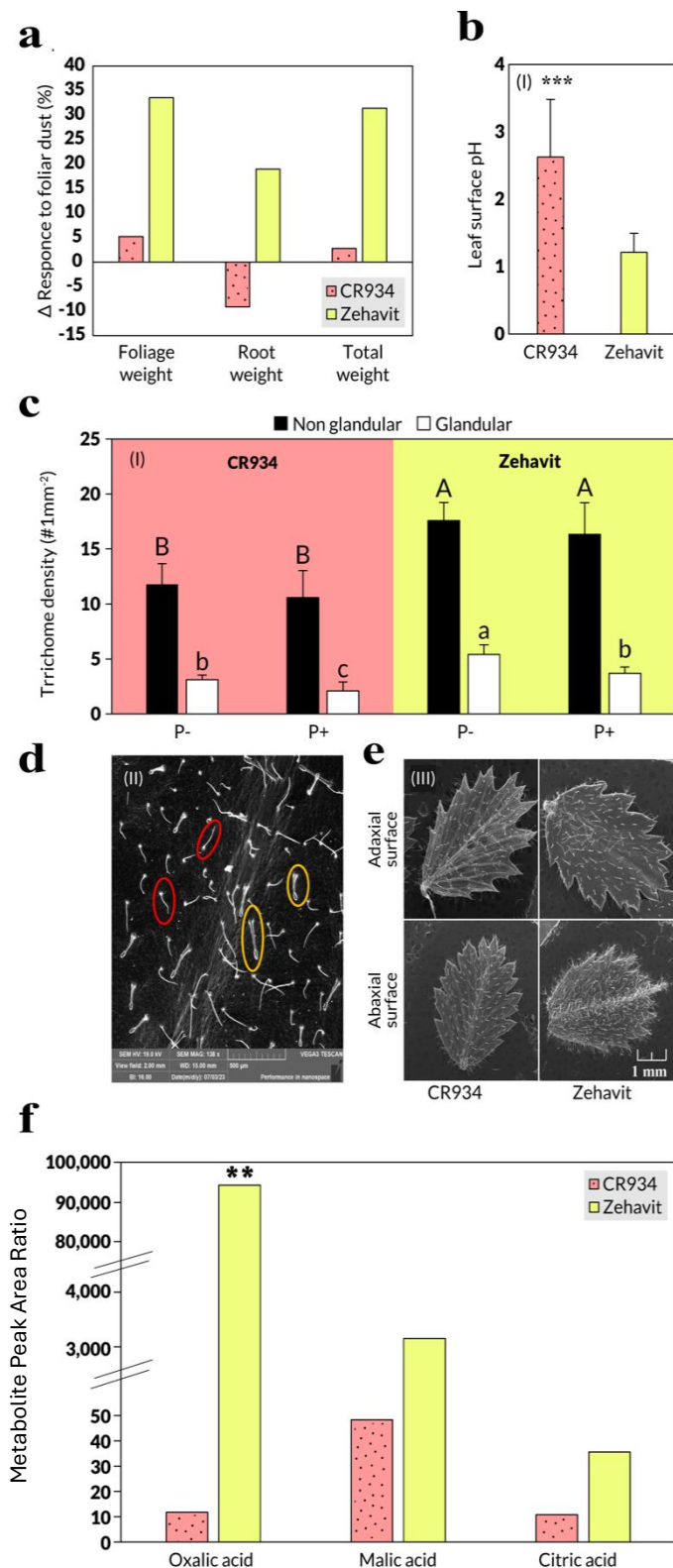
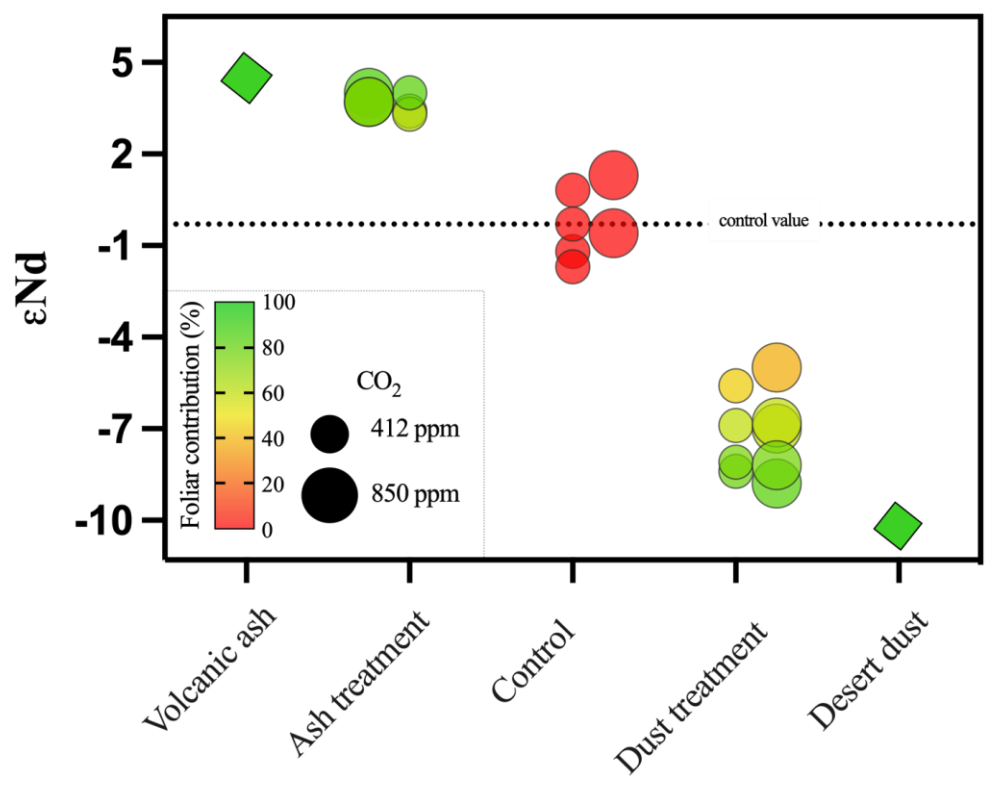


Figure 3 (a-f): Comparison of two chickpea varieties - CR934 (dotted, pink) and Zehavit (yellow) and their leaf properties under dust foliar fertilization. (a) Biomass and P uptake response to foliar dust P. Each column indicates the difference Δ (%) between the foliar dusted plants and the control untreated plants (n=6). (b) (I) Leaf surface pH. Each value indicates an average of five measurements on a plant throughout the growth season in control treatment (n=90), and two measurements in foliar dust treatment (n=10). Three asterisks indicate significant differences between treatments using a T-test, and a one-way ANOVA ($P \leq 0.001$). (c) Leaf non-Glandular (black column) and glandular (white column) trichome density in CR934 and Zehavit control plants (-P and +P). Different letters indicate significant differences between varieties and treatments using Tukey-HSD test ($P \leq 0.05$) (n=12). Capital letters refer to non-glandular trichomes and small letters refer to glandular trichomes. (d). SEM scans of non-glandular (red circles) and glandular (yellow circles) trichomes of typical Zehavit leaf. (e). SEM scans of leaves of CR934 (left) and Zehavit (right) varieties. (f). Exudates of organic acids. Each column indicates the average of leaf washing from four plants, in P- control treatment (n=4). Values are normalized to an internal standard (ribitol) for comparability. Two asterisks indicate significant differences between treatments using a T-test, and a one-way ANOVA ($P \leq 0.01$). Values are concentrations compared with an internal standard.



395

396

397

398

399

400

401

402

403

404

405

406

407

408

409

410

411

412

413

414

Figure 4: Quantification of dust mineral-nutrient flux from foliage. Radiogenic isotopic ratios of $^{143}\text{Nd}/^{144}\text{Nd}$ in the different sample groups (x-axis) expressed in ϵNd values. Diamonds represent the two applied mineral fractions of volcanic ash and desert dust; circles represent plants treated with the desert and volcanic dusts and the control groups. Large circles represent plants growing in the 850 ppm eCO_2 and small circles represent the 412 ppm aCO_2 . The color scale reflects the % contribution of Nd originating from the dusts via the foliage, which was calculated using a two-component mixing model. The control plants' Nd signature reflects the inheritance value from the seed, where a value of $\epsilon\text{Nd} = -0.3$ is set as the control, $\epsilon\text{Nd} = -10.3$ as the desert dust value, and $\epsilon\text{Nd} = 4.6$ as the volcanic ash value. A foliar contribution of more than 60% is evident in the plants applied with desert dust and more than 70% in the plants applied with volcanic ash. Standard errors on the isotopic values are all smaller than the depicted data points.

4 Discussion

4.1 Foliar mineral-nutrients uptake

In our experiments, we simulated desert dust and volcanic ash deposition by manually applying them on chickpea plants (*Cicer arietinum* cv 'Zehavit'). The dust was applied separately either on the surface of the pot near the roots, or on its foliage (Fig. 1), while control plants were not treated with dust. After several weeks, a significant impact of the foliar treatment was already noticeable where shoot biomass and total P content in the foliage-treated plants had increased, following dusts treatment, compared with the control group. In contrast, the root-treated plants did not show any increases in the biomass or P content, suggesting that over short timescales (i.e., several

weeks), foliar uptake is the only nutrient uptake pathway from freshly deposited dust (Fig. 1c, e). These results were then replicated when a similar experiment was conducted with plants grown on sandy soil, in bigger pots (Fig. S1), emphasizing that our observations are not limited to the specific artificial experimental conditions in perlite (which may bias root behavior), but also apply for real soil conditions (Fig. S1).

4.2 Foliar mineral-nutrient uptake mechanisms

Most of the P in dust is incorporated in the mineral lattice of minerals such as apatite (Dam et al., 2021), which is largely insoluble under the natural rhizosphere pH range (Hinsinger, 2001). Hence, P in volcanic or desert dust has low bioavailability for root uptake as was also shown in Lokshin et al, (2024a) with fire ash. On the leaf surface however, chemical, morphological, and microbial modifications may promote nutrient solubility and bioavailability and thus enable uptake through the leaf surface (Gross et al., 2021; Muhammad et al., 2019). Examining two chickpea varieties with contrasting responses to dust application: wild variety CR934, and common domesticated variety Zehavit, we found a few properties that facilitate foliar P acquisition from dust (Fig. 3). These include structural, morphological, and chemical modifications that are comparable to those reported in the rhizosphere (Hinsinger, 2001). The foliar-uptake-efficient variety Zehavit has significantly more acidic leaf surface (pH ~ 1, Fig. 3b), and thus promotes both dissolution and mobility of P from the pH sensitive mineral apatite (Gross et al., 2015), as well as other mineral-nutrients in the dust (Bradl, 2004; Gross et al., 2021; Muhammad et al., 2019). Nutrient uptake through the leaf is mediated by two primary pathways: the cuticle and stomata (Eichert and Fernández, 2022). The cuticle, while a hydrophobic barrier, contains dynamic aqueous pathways that allow solutes to diffuse, particularly under high humidity. Stomata act as regulated pores, facilitating the direct uptake of hydrophilic solutes and dissolved nutrients. These mechanisms likely complement the observed properties in Zehavit, where acidic exudates such as oxalate and malate further facilitated P uptake by promoting the dissolution of insoluble P forms (Lambers et al., 2019; Tiwari et al., 2022). Similarly, increased sugar levels, such as glucose and sucrose (Fig. 3f, fig. S2), likely stimulated the activity of nutrient-solubilizing microbes on the phyllosphere (Shakir et al., 2021). Additionally, Zehavit displayed higher trichome density on both leaf axial and adaxial surfaces (Fig. 3 c,d,e). These trichomes not only enhance metabolite release but also improve dust adhesion, increasing the contact time for nutrient solubilization and uptake (Gross et al., 2021) (fig. S3). Together, these traits align with established foliar uptake pathways and highlight the synergistic roles of chemical, morphological, and microbial modifications in facilitating nutrient acquisition from dust."

. We postulate that other plant species share comparable leaf traits that enhance dust capture and solubility such as wheat and various tree species that showed strong responses to foliar dust fertilization (Gross et al., 2021; Starr et al., 2023). Overall, our results suggest that the combination of leaf surface acidification, secretion of organic acids and additional exudations combined with an increased trichome density enhances foliar dust capture and nutrient uptake in chickpeas.

4.3 Dust Shading, Photosynthesis, and Elevated CO₂ Effects

Desert dust deposition on plant foliage exhibits both beneficial and harmful effects, influenced by its physical and chemical attributes as well as environmental conditions. Starr et al. (2023) demonstrated that dust application can

impact trees both positively and negatively: while it increased phosphorus (P) concentrations in some species, total plant P content showed only mild or stagnant increases, suggesting limited P utilization due to antagonistic dust–leaf interactions. On the harmful side, dust reduced final biomass across tree species, likely through disruptions in photosynthesis, stomatal conductance, and other physiological processes. Species-specific responses were noted, such as significant reductions in carbon assimilation and biomass in *Schinus* and a notable 58% biomass reduction in *Quercus*. Similarly, Lokshin et al. (2024b) observed an initial reduction in carbon assimilation following dust application, likely due to shading effects or stomatal occlusion. However, the increased content of rare earth elements (REE) resulting from foliar dust application subsequently contributed to improving carbon assimilation. In contrast, Gross et al. (2021) found no significant impact on plant biomass when using inert silica dust, suggesting that the effects of dust strongly depend on its chemical composition. They also proposed that negative impacts might be mitigated by plants increased internal P demand.

4.4 Dust impact on plant nutrient status under eCO₂

Numerous studies reported that eCO₂ conditions reduce the concentrations of several nutrients in plant tissues such as Fe, Zn, Cu, Mn, Ni and others (Loladze, 2002; Fernando et al., 2014; Myers *et al.*, 2014; Gojon *et al.*, 2023). The reduction in shoot nutrient concentrations was also observed in our experiments (fig. 5). In accordance with previous knowledge (Loladze, 2002), plants that were grown under eCO₂ in our experiment showed a significant reduction of 10-50% in the concentrations of nutrients such as Mg, K, Ca, Mn, Zn and Fe, with even more significant reductions in Cu and Ni (72% and 90%, respectively), (Fig. 5). Although we did not observe statistically significant differences in biomass between control plants grown under aCO₂ and eCO₂ conditions ($P = 0.4$), the reduction in essential macro- and micronutrient concentrations may be partly explained by the effect of nutrient dilution. Another potential reason for the nutrient decline under eCO₂ could be related to reduced efficiency in mineral nutrient absorption through the root system (Gojon et al., 2023). We found that foliar application of both volcanic and desert dust on plants that were grown under eCO₂ replenished their Fe and Ni concentrations (both essential micronutrients for plant growth and in the human diet) compared with the control group (fig. 5a,b). Desert dust treated plants showed increases of Fe and Ni concentrations of 44% and 46%, respectively (Fig. 5a). Volcanic ash treated plants showed Fe elevated concentrations of 66% (Fig. 5b). The Ni concentrations had more moderate increases from volcanic ash, with 40% higher than in the aCO₂. These increases returned Fe and Ni back to standard, nontoxic levels (Shahzad et al. 2018). These results suggest that the role of foliar uptake of atmospheric nutrients in the mineral nutrition level of plants may increase under eCO₂, potentially offsetting some of the nutrient reduction driven by the dilution effect and the downregulation of the root nutrient uptake pathway (Zhu et al., 2018). Despite adhering strictly to the washing protocol described in Gross et al. (2021) and Lokshin et al. (2024a), we acknowledge that some dust particles may remain on the plant surfaces. However, their influence on the results is negligible, as the contribution of any residual particles to the measured values is minimal and does not affect the overall interpretation of the results.

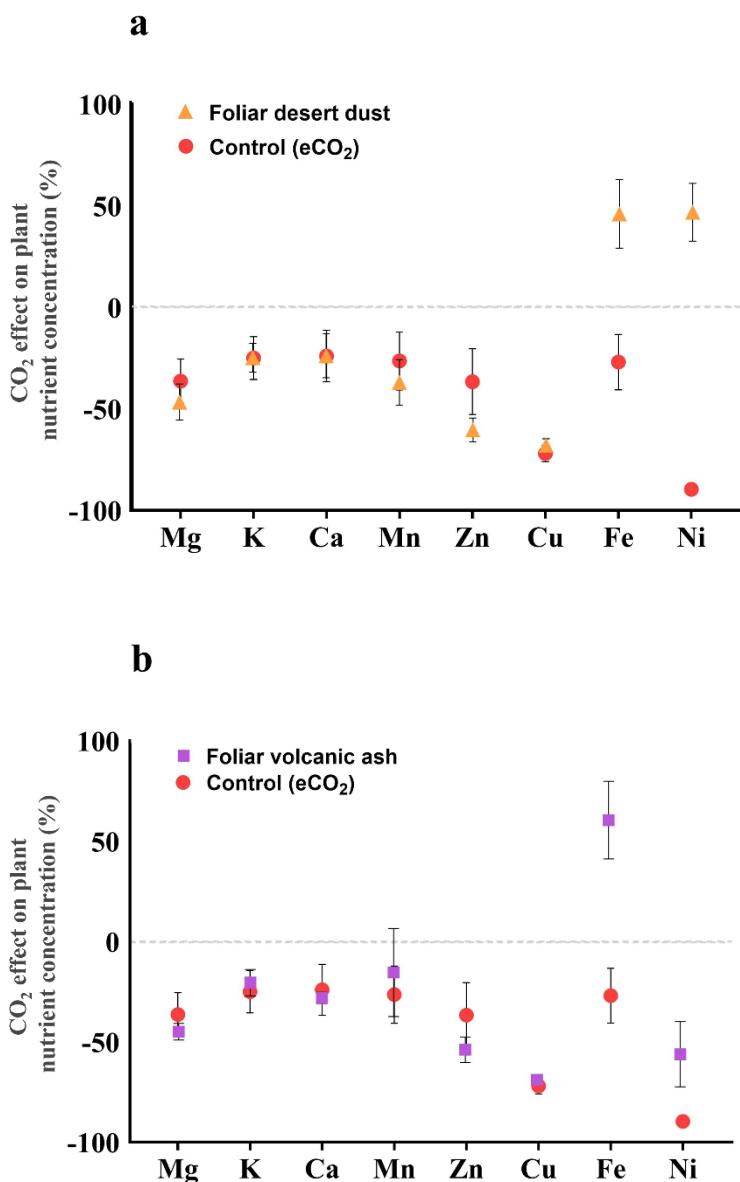


Figure 5 (a-b) Comparison of the % change in plant nutrient concentration under eCO₂ compared with aCO₂ control plants. The comparison was conducted as follows: the average value of each nutrient in plants grown under aCO₂ was calculated, and then each nutrient in individual chickpea plants grown under eCO₂ levels was expressed as a ratio relative to the average under aCO₂ conditions (eCO₂ plant (each individual plant) /aCO₂ plant (average of all the control plants)). Changes in nutrient concentrations of the control eCO₂ plants (red circles) show that eCO₂ conditions deteriorate plant nutritional status significantly. (a) The effect of foliar treatment of desert dust (orange triangles). (b) The effect of foliar treatment of volcanic ash (purple squares). Error bars represent standard deviations (n = 5).

4.5 Quantifying the contribution of foliar nutrient uptake from dust

Traditionally, radiogenic Nd isotopes serve as excellent tracers for sources of magmatic rocks (Stein and Goldstein, 1996), sediment archives (Chadwick et al., 1999; Palchan et al., 2018), and water bodies (Farmer et al., 2019). Since Nd is found in high concentration in nutrient bearing minerals (Aciego et al., 2017; Arvin et al., 2017; Chadwick et al., 1999), Nd isotopes were recently used to trace P sources in plant tissues, where it was

shown that the contribution of dust outpaces the weathering of the local bedrock over geological time scales (Aciego et al., 2017; Arvin et al., 2017). While the use of Nd isotopes to other elements such as P provides new knowledge on their sources, it should be done cautiously because different elements have differing speciation, uptake mechanisms, and transport kinetics in plant tissue. Here, we utilized the ratio of $^{143}\text{Nd}/^{144}\text{Nd}$ in the ϵNd notation to trace the source of Nd in our experiments and quantify its flux to plant tissue from dust. From this measurement we can approximate the flux of P, Fe and Ni via foliar pathway (Fig. 4). We used a two-component mixing model, where the average ϵNd value of the control plants, -0.3, which arise from the Nd “inheritance” (i.e., the Nd composition of the seed, since the amount of Nd in the chemical fertilizer was negligible) is regarded as one end member, and dust ϵNd values are regarded as the second end member, with values of -11 (desert dust) and 5 (volcanic ash). We found that desert dust treated plants were characterized with ϵNd values of -8.8 to -5, significantly different than the inheritance value of the control group. Similarly, the volcanic ash treated plants were characterized with ϵNd values of 3.4 to 4, significantly different than the inheritance value of -0.3. Thus, it is evident that the ϵNd of the foliage-treated plants comprise a mixture of the inheritance and the type of dust applied. Based on the mixing model, the chickpea plant acquired over 60% of its Nd from desert dust deposited on the foliage. Volcanic ash deposited on the foliage contributed over 70% of its Nd (Fig. 4). However, Nd isotopes do not show the increased supplement of Fe and Ni in plants that were grown under eCO_2 . Thus, more data on the relation between Nd and other nutrients uptake will advance its use in future studies to quantify the immediate contribution of freshly deposited dust on plants nutrition in field and lab experimental settings.

In conclusion, we showed here that dust nutrient uptake via the foliar pathway in responsive chickpea plants play a significant role in their nutrition under P limited conditions. Plant foliage captures and dissolves freshly deposited dust particles, making atmospheric mineral nutrients more accessible through the foliage on a short time scale than via the roots. Most of the P in the dust is incorporated in the mineral lattice of minerals such as apatite (Dam et al., 2021), which is largely insoluble under the natural rhizosphere pH range (Hinsinger, 2001). Hence, P in dust has low bioavailability for root uptake. On the leaf surface however, chemical, morphological, and microbial modifications may promote nutrient solubility and bioavailability and facilitate uptake through the leaf surface (Gross et al., 2021; Muhammad et al., 2019). Thus, our findings highlight that dust serves as an alternative source of nutrients to plants from the foliage on short timescales of a few weeks. Furthermore, foliar dust acquisition compensates for the reduction in nutrients such as Fe and Ni, induced by eCO_2 conditions (Gojon et al., 2023). The broader aspect of our findings emphasizes the central role of dust in plant nutrition through the foliar pathway and to global biogeochemical cycles. Our findings suggest that foliar uptake from natural dust could be a relevant pathway under future elevated CO_2 conditions, and that this pathway may be a target for novel fertilization techniques to compensate for the expected decline in the crops’ nutritional value.

Acknowledgments:

We thank Dr. Yigal Erel and Ofir Tirosh from the Hebrew University of Jerusalem for their support in ICP-MS analyses, and Dr. Yael Kiro from Weismann Institute for conducting isotopic chromatography in her lab, and Dr. Stephen Fox for his support in MC-ICP-MS analyses.

550 **Financial support:**

551 This research has been supported by the Keren Kayemeth LeIsrael-Jewish National Fund (KKL-JNF) climate
552 team doctoral fellowship and the Israeli Science Foundation (ISF) (grant no. 144/19 and grant no. 267/24). In
553 addition, this research was partly supported by Research Grant Award No. IS-5443-21 R, from BARD, The United
554 States - Israel Binational Agricultural Research and Development fund.

555 **Author Contributions:**

556 Conceptualization: DP, AG, RE

557 Dust sampling: AL, DA, AG

558 Methodology: DP, AG, RE

559 Investigation: AL, EG, SF

560 Visualization: DP, AL, EG

561 Funding acquisition: AG, RE, AL

562 Project administration: DP, AG

563 Supervision: DP, AG, RE

564 Writing – original draft: DP, AG, AL, RE

565 **Competing Interest Statement:** The authors declare no competing interests.

566 **Classification:** Physical Sciences - Earth, Atmospheric, and Planetary Sciences; Biological Sciences - Plant
567 Biology.

568

569 **References**

570 Aciego, S. M., Riebe, C. S., Hart, S. C., Blakowski, M. A., Carey, C. J., Aarons, S. M., Dove, N. C., Botthoff, J.
571 K., Sims, K. W. W., and Aronson, E. L.: Dust outpaces bedrock in nutrient supply to montane forest ecosystems,
572 Nat Commun, 8, 14800, <https://doi.org/10.1038/ncomms14800>, 2017.

573 Arvin, L. J., Riebe, C. S., Aciego, S. M., and Blakowski, M. A.: Global patterns of dust and bedrock nutrient
574 supply to montane ecosystems, Sci Adv, 3, eaao1588, <https://doi.org/10.1126/sciadv.aao1588>, 2017.

575 Bauters, M., Drake, T. W., Wagner, S., Baumgartner, S., Makelele, I. A., Bodé, S., Verheyen, K., Verbeeck, H.,
576 Ewango, C., Cizungu, L., Van Oost, K., and Boeckx, P.: Fire-derived phosphorus fertilization of African tropical
577 forests, Nat Commun, 12, <https://doi.org/10.1038/S41467-021-25428-3>, 2021.

578 Bradl, H. B.: Adsorption of heavy metal ions on soils and soils constituents, J Colloid Interface Sci, 277, 1–18,
579 <https://doi.org/10.1016/J.JCIS.2004.04.005>, 2004.

580 Burkhardt, J., Basi, S., Paryar, S., Hunsche, M.: Stomatal penetration by aqueous solutions- an update involving
581 leaf surface particles. New Phytologist, 196, 774-787, 2012.

582 Burkhardt, J., and Eichert, T.: Chapter 4 - Uptake and Release of Elements by Leaves and Other Aerial Plant
583 Parts, in: Marschner's Mineral Nutrition of Higher Plants, 4th ed., edited by: Rengel, Z., Cakmak, I., and White,
584 P. J., Elsevier, London, pp. 71–112, 2022.

585 Chadwick, O. A., Derry, L. A., Vitousek, P. M., Huebert, B. J., and Hedin, L. O.: Changing sources of nutrients
586 during four million years of ecosystem development, *Nature*, 397, 491–497, <https://doi.org/10.1038/17276>, 1999.

587 Ciriminna, R., Scurria, A., Tizza, G., and Pagliaro, M.: Volcanic ash as multi-nutrient mineral fertilizer: Science
588 and early applications, *JSFA Reports*, 2, 528–534, <https://doi.org/10.1002/JSF2.87>, 2022.

589 Clarkson, D. T. and Hanson, J. B.: The mineral nutrition of higher plants, *Ann. Rev. Plant Physiol*, 31, 239–98,
590 1980.

591 Dam, T. T. N., Angert, A., Krom, M. D., Bigio, L., Hu, Y., Beyer, K. A., Mayol-Bracero, O. L., Santos-Figueroa,
592 G., Pio, C., and Zhu, M.: X-ray Spectroscopic Quantification of Phosphorus Transformation in Saharan Dust
593 during Trans-Atlantic Dust Transport, *Cite This: Environ. Sci. Technol*, 55, 12694–12703,
594 <https://doi.org/10.1021/acs.est.1c01573>, 2021.

595 Eger, A., Almond, P. C., and Condon, L. M.: Phosphorus fertilization by active dust deposition in a super-humid,
596 temperate environment—Soil phosphorus fractionation and accession processes, *Global Biogeochem Cycles*, 27,
597 108–118, <https://doi.org/10.1002/GBC.20019>, 2013.

598 Farmer, J. R., Hönisch, B., Haynes, L. L., Kroon, D., Jung, S., Ford, H. L., Raymo, M. E., Jaume-Seguí, M., Bell,
599 D. B., Goldstein, S. L., Pena, L. D., Yehudai, M., and Kim, J.: Deep Atlantic Ocean carbon storage and the rise
600 of 100,000-year glacial cycles, *Nature Geoscience*, 12:5, 12, 355–360, [https://doi.org/10.1038/s41561-019-0334-](https://doi.org/10.1038/s41561-019-0334-6)
601 6, 2019.

602 Eichert, T. and Fernández, V.: Chapter 4: Foliar nutrient absorption: principles and prospects, in: Marschner's
603 Mineral Nutrition of Higher Plants, 4th Edn., edited by: Marschner, P., Academic Press, London, UK, pp. 123–
604 150, 2022.

605 Fernández, V., Guzmán, P., Peirce, C.A., McBeath, T.M., Khayet, M., McLaughlin, M.J.: Effect of wheat
606 phosphorus status on leaf surface properties and permeability to foliar-applied phosphorus. *Plant Soil* 384, 7-20,
607 2014a. Gojon, A., Cassan, O., Bach, L., Lejay, L., and Martin, A.: The decline of plant mineral nutrition under
608 rising CO₂: physiological and molecular aspects of a bad deal, *Trends Plant Sci*, 28, 185–198,
609 <https://doi.org/10.1016/J.TPLANTS.2022.09.002>, 2023.

610 Goll, D. S., Bauters, M., Zhang, H., Ciais, P., Balkanski, Y., Wang, R., and Verbeeck, H.: Atmospheric
611 phosphorus deposition amplifies carbon sinks in simulations of a tropical forest in Central Africa, *New*
612 *Phytologist*, 237, 2054–2068, <https://doi.org/10.1111/NPH.18535>, 2023.

613 Gross, A., Goren, T., Pio, C., Cardoso, J., Tirosh, O., Todd, M. C., Rosenfeld, D., Weiner, T., Custodio, D., and
614 Angert, A.: Variability in Sources and Concentrations of Saharan Dust Phosphorus over the Atlantic Ocean,
615 *Environ Sci Technol Lett*, 2, 31–37, <https://doi.org/10.1021/ez500399z>, 2015.

616 Gross, A., Palchan, D., Krom, M. D., and Angert, A.: Elemental and isotopic composition of surface soils from
617 key Saharan dust sources, *Chem Geol*, 442, 54–61, <https://doi.org/10.1016/j.chemgeo.2016.09.001>, 2016a.

618 Gross, A., Turner, B. L., Goren, T., Berry, A., and Angert, A.: Tracing the Sources of Atmospheric Phosphorus
619 Deposition to a Tropical Rain Forest in Panama Using Stable Oxygen Isotopes, *Environ Sci Technol*, 50, 1147–
620 1156, <https://doi.org/10.1021/ACS.EST.5B04936>, 2016b.

621 Gross, A., Tiwari, S., Shtein, I., and Erel, R.: Direct foliar uptake of phosphorus from desert dust, *New Phytologist*,
622 230, 2213–2225, <https://doi.org/10.1111/nph.17344>, 2021.

623 Guieu, C., Dulac, F., Desboeufs, K., Wagener, T., Pulido-Villena, E., Grisoni, J. M., Louis, F., Ridame, C., Blain,
624 S., Brunet, C., Bon Nguyen, E., Tran, S., Labiadh, M., and Dominici, J. M.: Large clean mesocosms and simulated
625 dust deposition: A new methodology to investigate responses of marine oligotrophic ecosystems to atmospheric
626 inputs, *Biogeosciences*, 7, 2765–2784, <https://doi.org/10.5194/BG-7-2765-2010>, 2010.

627 Hinsinger, P.: Bioavailability of soil inorganic P in the rhizosphere as affected by root-induced chemical changes:
628 A review, *Plant Soil*, 237, 173–195, <https://doi.org/10.1023/A:1013351617532/METRICS>, 2001.

629 Jweda, J., Bolge, L., Class, C., and Goldstein, S. L.: High Precision Sr-Nd-Hf-Pb Isotopic Compositions of USGS
630 Reference Material BCR-2, *Geostand Geoanal Res*, 40, 101–115, [https://doi.org/10.1111/j.1751-](https://doi.org/10.1111/j.1751-908X.2015.00342.x)
631 908X.2015.00342.x, 2016.

632 Kok, J. F., Adebisi, A. A., Albani, S., Balkanski, Y., Checa-Garcia, R., Chin, M., Colarco, P. R., Hamilton, D. S.,
633 Huang, Y., Ito, A., Klose, M., Li, L., Mahowald, N. M., Miller, R. L., Obiso, V., Pérez García-Pando, C., Rocha-
634 Lima, A., and Wan, J. S.: Contribution of the world’s main dust source regions to the global cycle of desert dust,
635 *Atmos Chem Phys*, 21, 8169–8193, <https://doi.org/10.5194/ACP-21-8169-2021>, 2021.

636 Lal, R.: Soil degradation as a reason for inadequate human nutrition, *Food Secur*, 1, 45–57,
637 <https://doi.org/10.1007/S12571-009-0009-Z>, 2009.

638 Lambers, H., Albornoz, F. E., Arruda, A. J., Barker, T., Finnegan, P. M., Gille, C., Gooding, H., Png, K.,
639 Ranathunge, K., and Zhong, H.: Nutrient-acquisition strategies, in: *A Jewel in the crown of a global biodiversity*
640 *hotspot*, edited by: Lambers, H., Kwongan Foundation and the Western Australian Naturalists’ Club Inc., Perth,
641 2019.

642 Lokshin, A., Palchan, D., and Gross, A.: Direct foliar phosphorus uptake from wildfire ash, *Biogeosciences*, 21,
643 2355–2365, <https://doi.org/10.5194/BG-21-2355-2024>, 2024a.

644 Lokshin, A., Gross, A., Dor, Y. Ben, and Palchan, D.: Rare earth elements as a tool to study the foliar nutrient
645 uptake phenomenon under ambient and elevated atmospheric CO₂ concentration, *Science of The Total*
646 *Environment*, 948, 174695, <https://doi.org/10.1016/J.SCITOTENV.2024.174695>, 2024b.

647 Loladze, I.: Rising atmospheric CO₂ and human nutrition: toward globally imbalanced plant stoichiometry,
648 *Trends Ecol Evol*, 17, 457–461, [https://doi.org/10.1016/S0169-5347\(02\)02587-9](https://doi.org/10.1016/S0169-5347(02)02587-9), 2002.

649 Loladze, I.: Hidden shift of the ionome of plants exposed to elevated CO₂ depletes minerals at the base of human
650 nutrition, *Elife*, 2014, <https://doi.org/10.7554/ELIFE.02245>, 2014.

651 Longo, A. F., Ingall, E. D., Diaz, J. M., Oakes, M., King, L. E., Nenes, A., Mihalopoulos, N., Violaki, K., Avila,
652 A., and Benitez-Nelson, C. R.: P-NEXFS analysis of aerosol phosphorus delivered to the Mediterranean Sea,
653 *Geophys Res Lett*, 41, 4043–4049, 2014.

654 Lowe, N. M.: The global challenge of hidden hunger: perspectives from the field, *Proceedings of the Nutrition*
655 *Society*, 80, 283–289, <https://doi.org/10.1017/S0029665121000902>, 2021.

656 Marschner, H., Kirkby, E. A., and Engels, C.: Importance of Cycling and Recycling of Mineral Nutrients within
657 Plants for Growth and Development, *Botanica Acta*, 110, 265–273, [https://doi.org/10.1111/J.1438-](https://doi.org/10.1111/J.1438-8677.1997.TB00639.X)
658 8677.1997.TB00639.X, 1997.

659 Muhammad, S., Wuyts, K., and Samson, R.: Atmospheric net particle accumulation on 96 plant species with
660 contrasting morphological and anatomical leaf characteristics in a common garden experiment, *Atmos Environ*,
661 202, 328–344, <https://doi.org/10.1016/J.Atmosenv.2019.01.015>, 2019.

662 Myers, S. S., Zanobetti, A., Kloog, I., Huybers, P., Leakey, A. D. B., Bloom, A. J., Carlisle, E., Dietterich, L. H.,
663 Fitzgerald, G., Hasegawa, T., Holbrook, N. M., Nelson, R. L., Ottman, M. J., Raboy, V., Sakai, H., Sartor, K. A.,
664 Schwartz, J., Seneweera, S., Tausz, M., and Usui, Y.: Increasing CO₂ threatens human nutrition, *Nature*, 510,
665 139–142, <https://doi.org/10.1038/nature13179>, 2014. Nakamaru, Y., Nanzyo, M., and Yamasaki, S. I.: Utilization
666 of apatite in fresh volcanic ash by pigeonpea and chickpea, *Soil Sci Plant Nutr*, 46, 591–600,
667 <https://doi.org/10.1080/00380768.2000.10409124>, 2000.

668 Okin, G. S., Mahowald, N., Chadwick, O. A., and Artaxo, P.: Impact of desert dust on the biogeochemistry of
669 phosphorus in terrestrial ecosystems, *Global Biogeochem Cycles*, 18, <https://doi.org/10.1029/2003GB002145>,
670 2004.

671 Palchan, D., Stein, M., Almogi-Labin, A., Erel, Y., and Goldstein, S. L.: Dust transport and synoptic conditions
672 over the Sahara–Arabia deserts during the MIS6/5 and 2/1 transitions from grain-size, chemical and isotopic
673 properties of Red Sea cores, *Earth Planet Sci Lett*, 382, 125–139, <https://doi.org/10.1016/j.epsl.2013.09.013>,
674 2013.

675 Palchan, D., Erel, Y., and Stein, M.: Geochemical characterization of contemporary fine detritus in the Dead Sea
676 watershed, *Chem Geol*, 494, 30–42, <https://doi.org/10.1016/J.Chemgeo.2018.07.013>, 2018.

677 Parasuraman, P., Pattnaik, S., and Busi, S.: Chapter 10 - Phyllosphere Microbiome: Functional Importance in
678 Sustainable Agriculture, in: *Phytomicrobiome Interactions and Sustainable Agriculture*, edited by: Rajarshi, K.
679 and Johnson, I., Elsevier, Amsterdam, pp. 185–207, 2022.

680 Shakir, S., Zaidi, S. S. e. A., de Vries, F. T., and Mansoor, S.: Plant Genetic Networks Shaping Phyllosphere
681 Microbial Community, <https://doi.org/10.1016/j.tig.2020.09.010>, 1 April 2021.

682 Schonherr, J.: Characterization of aqueous pores in plant cuticles and permeation of ionic solutes. *J. Exp. Bot.* 57,
683 2471–2491, 2006.

684 Starr, M., Klein, T., and Gross, A.: Direct foliar acquisition of desert dust phosphorus fertilizes forest trees despite
685 reducing photosynthesis, *Tree Physiol*, 43, 794–804, <https://doi.org/10.1093/treephys/tpad012>, 2023.

686 St.Clair, S. B. and Lynch, J. P.: The opening of Pandora’s Box: climate change impacts on soil fertility and crop
687 nutrition in developing countries, 335, 101–115, <https://doi.org/10.1007/s>, 2010.

688 Stein, M. and Goldstein, S. L.: From plume head to continental lithosphere in the Arabian-Nubian shield, *Nature*,
689 382, 773–778, 1996.

690 Stockdale, A., Krom, M. D., Mortimer, R. J. G., Benning, L. G., Carslaw, K. S., Herbert, R. J., Shi, Z.,
691 Myriokefalitakis, S., Kanakidou, M., and Nenes, A.: Understanding the nature of atmospheric acid processing of
692 mineral dusts in supplying bioavailable phosphorus to the oceans, *Proceedings of the National Academy of*
693 *Sciences*, 113, 14639–14644, 2016.

694 Tanaka, T., Togashi, S., Kamioka, H., Amakawa, H., Kagami, H., Hamamoto, T., Yuhara, M., Orihashi, Y.,
695 Yoneda, S., Shimizu, H., Kunimaru, T., Takahashi, K., Yanagi, T., Nakano, T., Fujimaki, H., Shinjo, R., Asahara,
696 Y., Tanimizu, M., and Dragusanu, C.: JNdi-1: a neodymium isotopic reference in consistency with LaJolla
697 neodymium, *Chem Geol*, 168, 279–281, [https://doi.org/10.1016/S0009-2541\(00\)00198-4](https://doi.org/10.1016/S0009-2541(00)00198-4), 2000.

698 Tiwari, S., Erel, R., and Gross, A.: Chemical processes in receiving soils accelerate solubilisation of phosphorus
699 from desert dust and fire ash, *Eur J Soil Sci*, 73, <https://doi.org/10.1111/EJSS.13270>, 2022.

700 Van Langenhove, L., Verryckt, L. T., Bréchet, L., Courtois, E. A., Stahl, C., Hofhansl, F., Bauters, M., Sardans,
701 J., Boeckx, P., Fransen, E., Peñuelas, J., and Janssens, I. A.: Atmospheric deposition of elements and its relevance
702 for nutrient budgets of tropical forests, 149, 175–193, <https://doi.org/10.1007/s10533-020-00673-8>, 2020.

703 Van Oss, R., Abbo, S., Eshed, R., Sherman, A., Coyne, C. J., Vandemark, G. J., Zhang, H. Bin, and Peleg, Z.:
704 Genetic relationship in cicer Sp. expose evidence for geneflow between the cultigen and its wild progenitor, *PLoS*
705 *One*, 10, <https://doi.org/10.1371/journal.pone.0139789>, 2015.

706 Wasserburg, G. J., Jacobsen, S. B., DePaolo, D. J., McCulloch, M. T., and Wen, T.: Precise determination of
707 Sm/Nd ratios, Sm and Nd isotopic abundances in standard solutions, *Geochim Cosmochim Acta*, 45, 2311–2323,
708 [https://doi.org/10.1016/0016-7037\(81\)90085-5](https://doi.org/10.1016/0016-7037(81)90085-5), 1981.

709 Zhu, C., Kobayashi, K., Loladze, I., Zhu, J., Jiang, Q., Xu, X., Liu, G., Seneweera, S., Ebi, K. L., Drewnowski,
710 A., Fukagawa, N. K., and Ziska, L. H.: Carbon dioxide (CO₂) levels this century will alter the protein,
711 micronutrients, and vitamin content of rice grains with potential health consequences for the poorest rice-
712 dependent countries, *Sci Adv*, 4, <https://doi.org/10.1126/Sciadv.aag1012>, 2018.

713

714

715

716

717

718

719

720