



1 **Accurate Elucidation of Oxidation Under Heavy Ozone**
2 **Pollution: A Full Suite of Radical Measurement In the**
3 **Chemical-complex Atmosphere**

4 Renzhi Hu¹, Guoxian Zhang^{1,2,×}, Haotian Cai¹, Jingyi Guo¹, Keding Lu⁴, Xin Li⁴,
5 Shengrong Lou⁵, Zhaofeng Tan⁴, Changjin Hu¹, Pinhua Xie^{1,3,××}, Wenqing Liu^{1,3}

6 ¹ Key Laboratory of Environment Optics and Technology, Anhui Institute of Optics and Fine
7 Mechanics, HFIPS, Chinese Academy of Sciences, Hefei, China

8 ² School of Physics and New Energy, Xuzhou University of Technology, Xuzhou, China

9 ³ College of Resources and Environment, University of Chinese Academy of Science, Beijing, China

10 ⁴ State Key Joint Laboratory of Environmental Simulation and Pollution Control, College of
11 Environmental Sciences and Engineering, Peking University, Beijing, China

12 ⁵ State Environmental Protection Key Laboratory of the Formation and Prevention of Urban Air
13 Pollution Complex, Shanghai Academy of Environmental Sciences, Shanghai, China

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15 **×Correspondence to:** Guoxian Zhang, School of Physics and New Energy, Xuzhou
16 University of Technology, Xuzhou, China

17 **××Correspondence to:** Pinhua Xie, Key Laboratory of Environment Optics and
18 Technology, Anhui Institute of Optics and Fine Mechanics, HFIPS, Chinese Academy of
19 Sciences, Hefei, China

20 **Email addresses:** gxzhang@aiofm.ac.cn (Guoxian Zhang); phxie@aiofm.ac.cn (Pinhua
21 Xie)

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23 **Abstract:** The Yangze River Delta (YRD) in China encountered with prolonged ozone
24 pollution in September 2020, which had significant impacts on the respiratory,
25 dermatological, and visual health of local residents. To accurately elucidate the
26 limitations of oxidation processes in the chemical-complex atmosphere, a full suite of
27 radical measurements (OH, HO₂, RO₂, and *k*_{OH}) was established in YRD region for the
28 first time. The diurnal peaks of radicals exhibited considerable variation due to
29 environmental factors, showing ranges of 3.6 to 27.1×10⁶ cm⁻³ for OH, 2.1 to 33.2×10⁸
30 cm⁻³ for HO₂, and 4.9 to 30.5×10⁸ cm⁻³ for RO₂. At a heavy ozone pollution episode, the
31 oxidation capacity reached an intensive level compared with other sites, and the
32 simulated OH, HO₂, and RO₂ radicals provided by the RACM2-LIM1 mechanism failed
33 to adequately match the observed data both in concentration and coordinate ratios.
34 Sensitivity tests based on the full suite of radical measurement revealed that the X
35 mechanism accelerated OH regeneration, and the introduction of larger RO₂
36 isomerization steps alleviated the RO₂-related imbalance by 2 to 4 times. The hypothesis
37 of sensitivity analysis can be chemically validated by the special HCHO contribution to
38 oxidation. Constraining HCHO increased the ChL from 1.94 to 4.45, leading to a 51.54%
39 increase in ozone production during the heavy pollution. The incorporation of complex
40 processes enabled better coordination of HO₂/OH, RO₂/OH, and HO₂/RO₂ ratios
41 comparable to the observed values, and adequately addressed the deficiency in the ozone
42 generation mechanism within a certain range. The full-chain radical detection untangled a
43 gap-bridge between the photochemistry and the intensive oxidation level in the
44 chemical-complex atmosphere, enabling a deeper understanding of the tropospheric
45 radical chemistry at play.

46

47 **Keywords:** FAGE-LIF; Full-chain detection; Radical; P(Ox); HCHO;

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49 **1 Introduction**

50 In recent years, China's rapid economic development has led to severe environmental
51 pollution problems, which significantly impacted the respiratory, dermatological, and visual
52 health of local residents (Wang et al., 2022c; Huang et al., 2018). This has raised
53 concerns about the coexistence of regional primary and secondary pollution, making air
54 quality improvement efforts a focal point (Liu et al., 2021; Wang et al., 2022a). In the
55 complex atmosphere, near-surface ozone (O_3) is formed through continuous
56 photochemical reactions between nitrogen oxides ($NO_x \equiv NO + NO_2$) and volatile
57 organic compounds (VOCs) under light conditions, while hydroxyl radicals (OH) serve
58 as the main oxidant in the troposphere, converting VOCs into hydroperoxy (HO_2) and
59 organic peroxy (RO_2) radicals (Rohrer et al., 2014; Hofzumahaus et al., 2009).
60 Additionally, the oxidation of nitric oxide (NO) and peroxy radicals produce nitrogen
61 dioxide (NO_2), which is the sole photochemical source of ozone (Lu et al., 2012; Stone et
62 al., 2012).

63 Despite numerous experimental and theoretical explorations to establish the
64 radical-cored photooxidation mechanism in the troposphere, field observations were
65 primarily focused on HO_x ($HO_x \equiv OH + HO_2$) radicals due to the limitations of detection
66 technology (Kanaya et al., 2012; Lu et al., 2012; Hofzumahaus et al., 2009; Yugo Kanaya
67 and Tanimoto, 2007; Ren et al., 2008; Stone et al., 2012; Levy, 1971). Recent
68 advancements in detection technology, such as the application of a new LIF technique
69 (ROxLIF), have made the detection of RO_2 radicals possible (Whalley et al., 2013; Tan et
70 al., 2017a). Moreover, the closure experiment, incorporating field campaigns and box
71 model, has proven to be an effective method for verifying the integrity of radical
72 chemistry at local to global scales (Lu et al., 2019; Tan et al., 2018). Several experiments
73 have indicated that the existing atmospheric chemical mechanism posed challenges in
74 deepening the understanding of the regional pollution explosion (Whalley et al., 2021;
75 Slater et al., 2020; Tan et al., 2017a; Woodward-Massey et al., 2023). For instance, the
76 observation of up to $4 \times 10^9 \text{ cm}^{-3}$ of RO_2 radical in the center of Beijing in 2017 (APHH)
77 was significantly underestimated by the MCM v3.3.1 mechanism (Whalley et al., 2021).
78 Further exploring the unreproducible concentration and the oxidation process in the



79 chemical-complex atmosphere is deemed necessary (Whalley et al., 2021;
80 Woodward-Massey et al., 2023).

81 The YRD region, situated between the North China Plain (NCP) and Pearl River
82 Delta (PRD), is highly prone to regional transport interactions and aerosol-boundary layer
83 feedback (Jia et al., 2021; Huang et al., 2020). In September 2020, the YRD region
84 experienced a severe episode of secondary pollution, with both the daily maximum
85 8-hour average ozone (MDA8) and daily average PM_{2.5} concentrations surpassing the
86 pollution threshold, distinguishing it from other megacities (Fig. S1). In an effort to gain
87 a better understanding between the complex radical chemistry and the intensive oxidation
88 level, TROPSTECT-YRD (The experiment on Radical chemistry and Ozone Pollution
89 perSpectively: long-Term Elucidation of the photochemiCal oxidaTion in the Yangze
90 River Delta) was conducted in Hefei during September 2020. Accurate elucidation of the
91 oxidation process under heavy ozone pollution was provided by a full suite of radical
92 measurement (OH, HO₂, RO₂ and *k*_{OH}) in the chemical-complex atmosphere.

93 **2 Materials and methods**

94 **2.1 Site description and instrumentation**

95 The TROPSTECT observation was conducted from 1 to 20 September 2020 at the
96 Science Island background station (31.9° N, 117.2°E) in Hefei, a typical megacity located
97 in the central region of Anhui Province within the Yangtze River Delta. The station is
98 situated on a peninsula with abundant vegetation to the northwest of urban areas and is in
99 close proximity to Dongpu Lake, which is only 200 meters away, and the main road,
100 positioned 250 meters southward (Fig. 1). Consequently, the relatively enclosed
101 environment exhibits typical suburban characteristics of anthropogenic emissions. The
102 station is located in the transition region between urban and suburban areas, reflecting the
103 regional transpor of pollution in Hefei and its surrounding areas.

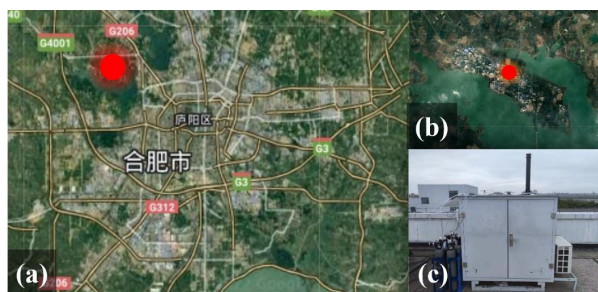


Fig. 1. (a) The location of the measurement site (source: © Google Earth).
(b) The close shot of the measurement site location (source: © Google Earth).
(c) The actual image for the LIF-Box.

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Regarding the instrumentation, a group of oxidation-related instruments were installed on the 7th floor of the Comprehensive Building at the Anhui Institute of Optics and Fine Mechanics (AIOFM), with all sampling outlets positioned more than 20 meters above the ground. The details of the instruments measuring various parameters such as meteorological factors (WS, WD, T, RH, P, Jvalues), gas pollutants (O_3 , CO, SO_2 , NO, NO_2 , HONO, HCHO, PAN), and non-methane hydrocarbons (NMHCs) are provided in Table S1.

115 The measurement of NO, NO_2 , O_3 , CO, and SO_2 was carried out using commercial
116 Thermo Electron model series instruments. Thereof, NO and NO_2 were measured using a
117 chemical fluorescence method (CL) with an enhanced trace-level NO- NO_2 -NO_x analyzer
118 (PKU-PL), which achieved a detection limit of 50 ppt (Tan et al., 2017a). The detection
119 of O_3 and SO_2 was conducted through Thermo Electron model 49i and 43i, respectively,
120 while Thermo Electron model 48i was utilized for CO detection. Cavity ring-down
121 spectroscopy (CRDS, Picarro-G2401) was employed for CO detection in parallel, and
122 another ultraviolet absorption instrument (Ecotech EC9810B) was for ozone detection.
123 The instrument inlets were placed within 5 meters of each other for comparison.

124 To ensure measurement accuracy, the instruments in the campaign underwent zero
125 point calibration procedures during the early (August 31st) and late (September 21st)
126 observation periods, and cross-calibrations for O_3 and CO measurements were carried out
127 during the middle (September 9th). Furthermore, additional zero calibration for Thermo
128 48i CO detection was undertaken daily from 0:00-0:30 to minimize shift correction. The
129 comparison results revealed high consistency within the instrument accuracy range for
130 both CO and O_3 measurements (Fig. S2(a)(b)).

131 HONO was detected using a home-built instrument by cavity-enhanced absorption



132 spectroscopy (CEAS), while formaldehyde was determined by the Hantzsch method
133 (SDL MODEL 4050) (Duan et al., 2018; Yang et al., 2021a). An automated gas
134 chromatograph equipped with a mass spectrometer and flame ionization detector
135 (GC-FID/MS) was employed for the online measurement of 99 VOCs species.

136 The eight crucial photolysis frequencies ($j(\text{NO}_2)$, $j(\text{H}_2\text{O}_2)$, $j(\text{HCHO})$, $j(\text{HONO})$,
137 $j(\text{NO}_2)$, $j(\text{NO}_3)$, $j(\text{O}^1\text{D})$) were directly measured by a photolysis spectrometer (Metcon,
138 Germany). The unmeasured photolysis frequencies of the remaining active species were
139 computed using Eq.(1):

$$j = l \cdot \cos(\chi)^m \cdot e^{-n \cdot \sec(\chi)} \quad (1)$$

140 The variations in photolysis frequency due to solar zenith angle (χ) were adjusted based
141 on the ratio of observed and simulated $j(\text{NO}_2)$. The optimal values for parameters (l , m ,
142 and n) for different photolysis frequencies were extensively detailed by the MCM v3.3.1
143 (http://mcm.york.ac.uk/parameters/photolysis_param.htm) (Jenkin et al., 2003; Jenkin et
144 al., 1997).

145 2.2 Radical measurement

146 2.2.1 OH, HO₂, RO₂ Concentrations

147 The laser-induced fluorescence instrument developed by the Anhui Institute of Optics
148 and Fine Mechanics (AIOFM-LIF) was used to simultaneously detect the concentrations
149 of OH, HO₂, and RO₂ radicals, along with OH reactivity (k_{OH}). The OH radical was
150 directly measured by detecting on-resonance fluorescence excited by a 308 nm laser. An
151 indirect measurement for HO₂ was carried out after converting it to OH at a fixed
152 efficiency (Heard and Pilling, 2003).

153 The laser utilized for fluorescence excitation is a high-frequency tunable dye laser
154 that emits a 308 nm laser, with the laser power divided into a ratio of 0.45:0.45:0.08:0.02.
155 Of this power, 90% is directed towards fluorescence cells for concentration and reactivity
156 detection via optical fibers, respectively. 8% of the laser power is directed to a reference
157 cell, while the remaining 2% is used to monitor real-time power fluctuations. The laser is
158 transmitted through HO₂, OH, and RO₂ cells in turn via a coaxial optical path. Two
159 photodiodes are set up at the end of the reference cell and RO₂ detection cell, respectively.
160 The voltage signals and power fluctuations are compared synchronously to diagnose the
161 laser stability. To maintain detection efficiency, the power inside the measurement cells



162 should not be less than 10 mW. Sampling nozzles of 0.4 mm are deployed above OH and
163 HO₂ cells for efficient sampling at a flow rate of approximately 1.1 SLM, and the
164 pressure for all fluorescence cells are maintained at 400 Pa. The micro-channel plate
165 (MCP) detects the weak fluorescence signal collected by lens systems with low noise and
166 high gain. Additionally, a digital delay generator (DG645) optimizes the timing control
167 between the laser output, signal detection, and data acquisition. All of these modules are
168 integrated into a sampling box with constant air conditioning, except for the laser.

169 The detection of RO₂ radicals is more complex compared to the integrated detection
170 of OH and HO₂ radicals (Whalley et al., 2013). To achieve the complete chemical
171 conversion from RO_x to HO₂, a crucial role is played by a 66 mm×830 mm aluminium
172 flow tube, whose performance has been confirmed through the CHOOSE-2019 field
173 campaign (Li et al., 2020). A mixture of 0.17% CO and 0.7 ppm NO injected into the
174 flow tube facilitates the reduction of heterogeneous radical loss and enhancement of
175 conversion efficiency. The sampling flow is limited to 7 SLM by a 1 mm nozzle, and the
176 tube pressure is maintained at 25 hPa. In contrast to the HO_x cells, the large-diameter
177 nozzle (4 mm) is equipped above the cell, and a high concentration of NO (~300 ppm)
178 facilitates the full magnitude HO₂→OH conversion.

179 The observation data (H₂O, O₃) is combined with experimental characterization to
180 eliminate ozone photolysis interference, and most interference signals are excluded by
181 utilizing wavelength modulation (Zhang et al., 2022a). A comparison experiment with
182 PKU-LIF demonstrated the consistency of OH measurement in complex atmosphere
183 (Zhang et al., 2022b). For HO₂ measurement, lower NO concentration (~1.6 × 10¹² cm⁻³,
184 corresponding to ~15% conversion efficiency) are selected to limit the RO₂→HO₂
185 interference to less than 5% (Wang et al., 2021). Since only the total-RO₂ mode is used
186 for the campaign, the additional uncertainty introduced by RO₂/R(OH)O₂ classification is
187 negligible (Tan et al., 2017b). The observed maximum daily PAN (11:00-14:00) is only
188 1.15±0.67 ppb, resulting in a calculated PAN-pyrolytic interference for RO₂ measurement
189 of less than 1 ppt (Fuchs et al., 2008). The general applicability of AIOFM-LIF in
190 complex atmosphere has been demonstrated through various campaigns (Zhang et al.,
191 2022b; Wang et al., 2021; Wang et al., 2019a).

192 To complete the calibration task, a standard source stably generates equal amounts



193 of OH and HO₂ radicals (Wang et al., 2020). The radical source is also capable of
194 yielding specific RO₂ by titrating hydrocarbon with OH. It is noteworthy that CH₃O₂ has
195 the highest mixing ratio in the RO₂ species, thus it was chosen to represent for sensitivity
196 calibration. The instrument is calibrated every two days, except during rainy weather. The
197 limit of detection (LOD) for OH, HO₂, and RO₂ in different cells with a typical laser
198 power of 10 mW is measured at $3.3 \times 10^5 \text{ cm}^{-3}$, $1.1 \times 10^6 \text{ cm}^{-3}$, and $2.5 \times 10^6 \text{ cm}^{-3}$,
199 respectively (60 s, 1 σ). Measurement accuracy for OH, HO₂, and RO₂ radicals are
200 reported to be 13%, 17%, and 21%, respectively.

201 **2.2.2 OH reactivity (k_{OH})**

202 The detection of k_{OH} in the atmosphere, defined as the reciprocal of OH lifetime, was
203 conducted using a laser flash photolysis laser-induced fluorescence (LP-LIF) instrument
204 (Lou et al., 2010). The configuration structure for k_{OH} measurement has been previously
205 detailed in a study by (Liu et al., 2019). A pulsed laser beam (266 nm with an average
206 power of 15 mJ) is output from a frequency-quadrupled Nd:YAG laser, which generates
207 stable OH radical through flash photolysis in the flow tube. Subsequently, the OH
208 radicals are sampled through a nozzle into a fluorescence cell. The OH fluorescence
209 signal is then detected using laser pump and probe techniques and is fitted to calculate the
210 slope of OH decay (k_{OH}). The detection accuracy, achieved with an integration time of
211 180 s, is 0.3 s^{-1} (1 σ).

212 **2.3 Observation-Based Model**

213 The Regional Atmospheric Chemical Mechanism version 2 (RACM2) incorporating
214 the latest Leuven isoprene mechanism (LIM) was utilized to simulate the concentrations
215 and reactions of OH, HO₂, and RO₂ radicals (Stockwell et al., 1997; Griffith et al., 2013;
216 Peeters et al., 2014). The RACM2-LIM1 mechanism was specifically involved with
217 fewer species compared to the explicit MCM mechanism, thus ensuring higher
218 operational efficiency (Liu et al., 2022). For the base scenario, boundary conditions were
219 established using the observed species listed in Table S1, with assumed concentrations of
220 hydrogen (H₂) and methane (CH₄) at 550 ppb and 1900 ppb, respectively. Moreover, the
221 atmospheric lifetime (τ_D) for all the simulated species was set to 24 hours, and a
222 corresponding first-order loss rate of 1.1 cm/s (assuming a boundary layer height of 1
223 km). At this lifetime, the simulated ozone concentration closely matched the observed



224 values (Fig.S3). The time resolution of all constraints was uniformly set to 15 minutes
 225 through averaging or linear interpolation. To reinitialize unconstrained species to a
 226 steady-state, three days of data were input in advance as the spin-up time.

227 The local formation of ozone can be accurately quantified through the online
 228 measurement of ROx radicals (Tan et al., 2018). To overcome the interference from NO,
 229 the total oxidant (Ox), which is defined as the sum of NO₂ and O₃, can serve as a reliable
 230 parameter to indicate the level of oxidation. Eq.(2) shows that the rate of NO oxidation
 231 by peroxy radicals is equivalent to the production of O₃, denoted as F(Ox):

$$232 \quad F(O_x) = k_{HO_2+NO}[NO][HO_2] + \sum_i k_{RO_2^i+NO}[NO]RO_2^i \quad (2)$$

233 The major loss pathways for Ox encompass ozone photolysis, ozonolysis reactions,
 234 and radical-related reactions (OH/HO₂+O₃, OH+NO₂), represented as D(Ox) in Eq.(3):

$$235 \quad D(O_x) = \varphi_{OH}j(O^1D)[O_3] + \sum_i \{ \varphi_{OH}^i k_{Alkenes+O_3}^i [Alkenes][O_3] \} + (k_{O_3+OH}[OH] +$$

$$236 \quad k_{O_3+HO_2}[HO_2])[O_3] + k_{OH+NO_2}[OH][NO_2] \quad (3)$$

237 The net photochemical Ox production rate in the troposphere, denoted as P(Ox) in
 238 Eq.(4), can therefore be calculated as the difference between Eqs. (2) and (3):

$$239 \quad P(O_x) = F(O_x) - D(O_x) \quad (4)$$

240 Chain Length (ChL) is positively correlated with NO concentration, illustrating the
 241 competitive relationship between chain propagation and chain termination reactions (Eq.
 242 5). ChL is defined as the ratio of OH regeneration to the primary source of OH
 243 production (Emmerson et al., 2007; Martinez et al., 2003):

$$244 \quad ChL = \frac{k_{HO_2+NO}[HO_2][NO] + k_{HO_2+O_3}[HO_2][O_3]}{j_{HONO}[HONO] + \varphi_{OH}j(O^1D)[O_3] + \sum (\varphi_{OH}^i k_{alkene}^i [alkene]_i [O_3])} \quad (5)$$

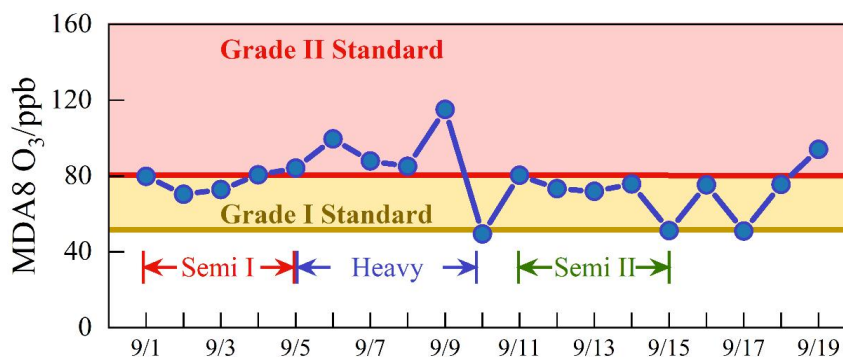
245 **3 Results**

246 **3.1 Overview of Measurement**

247 During the observation period, the meteorological parameters and trace gas
 248 concentrations were plotted in Fig. S4. The timeseries revealed that the peak temperature
 249 exceeded 30°C, and the humidity levels remained between 30 – 50% during the daytime.
 250 The photolysis rates were observed to peak at noon (11:00 – 13:00), with j(O¹D) and
 251 j(NO₂) reaching approximately 3×10⁻⁵ s⁻¹ and 8×10⁻³ s⁻¹, respectively. Brief rainfall events
 252 temporarily happened on September 10th, 15th, and 17th, but totally favorable



253 meteorologies induced the prolonged ozone pollution. The daily maximum 8-hour
 254 average ozone concentration (MDA8), as depicted in Fig.2, consistently exceeded the
 255 Chinese Grade I national air quality standard (GB3095-2012) throughout the observation,
 256 with nine days exceeding the Grade II standard.



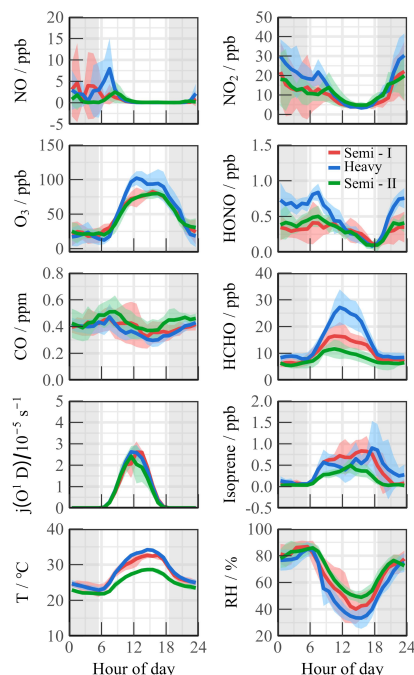
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Fig.2. The daily maximum 8 h average O₃ during the campaign. The yellow and red lines denote the Grade I and Grade II national standards for O₃, respectively. Brief rainfall events temporarily happened on 10, 15, and 17 Sep.

261 The ozone pollution can be categorized into three continuous periods based on
 262 pollution levels, which disclose transitional ‘Semi - Heavy - Semi’ pollution
 263 characteristics. Fig.3 depicts daily variations in meteorological and trace gas
 264 concentrations for different periods. During the Semi I (1 to 5 September) and Semi II (11
 265 to 14 September) periods, the MDA8 levels exceeded Grade I standard, with an average
 266 value of 75.92 ± 5.14 ppb and 75.45 ± 3.73 ppb, respectively. Notably, NO levels peaked
 267 around 9:00 and rapidly decreased to a few hundred ppt due to photochemistry. In
 268 addition, HONO and NO₂ exhibited bimodal variations, with diurnal concentration ranges
 269 of 0.09 – 0.50 ppb and 3.35 – 13.77 ppb, respectively. The HONO/NO₂ ratios during both
 270 Semi periods were consistent with previous urban/suburban observations, with daytime
 271 values of 0.049 ± 0.014 and 0.035 ± 0.012 , respectively (Yang et al., 2021b; Shi et al., 2020;
 272 Hu et al., 2022). Isoprene levels accumulated during the day and decreased at night
 273 during both Semi pollution episodes, with a diurnal average concentration in Semi II only
 274 49.3% of that in Semi I (0.71 ± 0.087 ppb vs 0.35 ± 0.073 ppb). Formaldehyde, as the key
 275 oxidation species, exhibited a concentration profile mirroring that of isoprene, with
 276 significantly higher concentrations ranging from 1.20 to 36.34 ppb compared to other
 277 urban regions (Ma et al., 2022; Yang et al., 2022; Tan et al., 2017b; Yang et al., 2021a).
 Heavy pollution episodes from 5 to 9 September resulted in daytime ozone concentration



278 as high as 129.9 ppb, and oxidation-related species such as HCHO, HONO, NO_x, and
 279 VOCs increased synchronously compared to other days.



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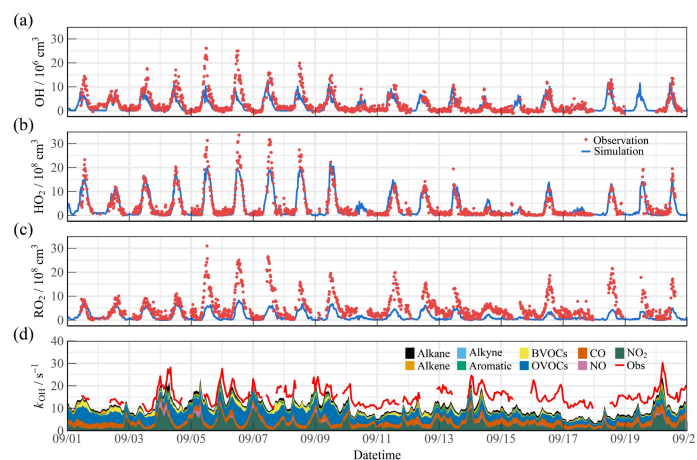
281 **Fig. 3.** Mean diurnal profiles of observed meteorological and chemical parameters during the campaign. Three periods
 282 were divided for subsequent study (Semi I, Heavy, and Semi II).

283 3.2 RO_x radical concentrations and budgets

284 The observed and modeled timeseries for OH, HO₂, RO₂, and *k*_{OH} during the
 285 observation time are depicted in Fig.4. The diurnal peaks of radicals exhibited a wide
 286 span due to changes in environmental conditions, with ranges of 3.6 – 27.1×10⁶ cm⁻³ for
 287 OH, 2.1 – 33.2×10⁸ cm⁻³ for HO₂, and 4.9 – 30.5×10⁸ cm⁻³ for RO₂. Continuous data for
 288 *k*_{OH} observation were acquired within a range of 8.6 – 30.2 s⁻¹. Fig.S5 presents the diurnal
 289 profiles of the observed and modeled values during different episodes. The diurnal
 290 maximum of OH radical at noon differed between Semi I and Semi II, with 9.28×10⁶ cm⁻³
 291 and 5.08×10⁶ cm⁻³, respectively, while total peroxy radicals (HO₂+RO₂) remained at
 292 similar levels with 19.43×10⁸ cm⁻³ and 18.38×10⁸ cm⁻³. Additionally, the distribution of
 293 peroxy radicals are not similar in the two Semi periods, with HO₂/RO₂ ratios of 1.69:1
 294 and 0.76:1, respectively, which reflects the uneven oxidation levels between Semi I and
 295 Semi II. During the Heavy ozone pollution, the averaged OH, HO₂, and RO₂



296 concentrations were 1.90, 2.15, and 1.98 times higher than those in the Semi periods,
 297 suggesting a stronger oxidation capacity, with k_{OH} in Heavy being 26.43% and 9.56%
 298 higher than in Semi I and Semi II, respectively. Limited anthropogenic emissions in the
 299 suburban environment reduced the oxidation contribution by NO_x and CO (27.59%).
 300 During the heavy pollution, organic species exhibited dominant behavior regarding
 301 diurnal reactivity (9.22 s^{-1} for 69.79%), and anthropogenic hydrocarbons were not major
 302 k_{OH} sources. With an abundant level ($\sim 1 \text{ ppb}$), isoprene contributed more than 10% of the
 303 reactivity in the diurnal cycle. Therefore, the effect of BVOCs species (such as
 304 monoterpenes, limonene, etc.) on radical chemistry cannot be ignored (Ma et al., 2022;
 305 Wang et al., 2022b).



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307 **Fig. 4.** Timeseries of the observed and modelled parameters for OH, HO₂ and k_{OH} during the observation period.

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(a) OH, (b) HO₂, (c) k_{OH} .

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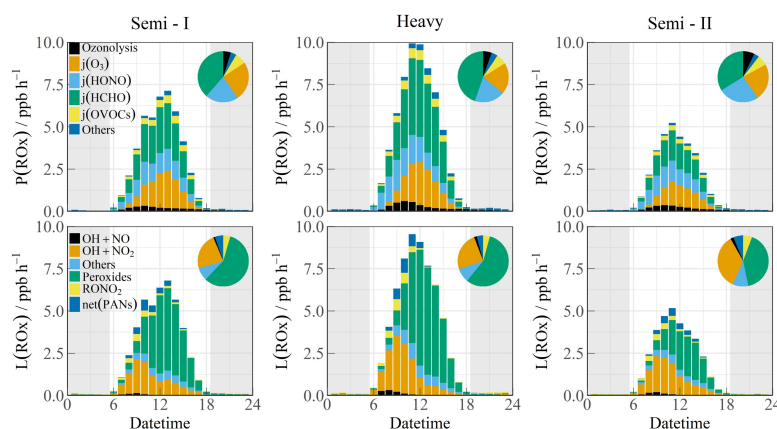
310 The significant variations in oxidation can be inferred from the disparities during
 311 different pollution periods (Fig.S5). During Semi I, there was a good agreement between
 312 the measurement and model for peroxy radicals during the daytime. The RACM2-LIM1
 313 mechanism effectively replicated the morning OH radical concentration. However,
 314 following 10:00, NO gradually declined, and the increasing OH concentration could not
 315 be accounted for by the HO₂+NO formation channel, resulting in a maximum
 316 underestimation of $5.85 \times 10^6 \text{ cm}^{-3}$ (Hofzumahaus et al., 2009; Lu et al., 2012). In the
 317 Semi II episode, OH was not underestimated in the low-NO regime, with a slight
 318 overestimation of HO₂ concentration. However, the simulated RO₂ concentration was
 only $3.78 \times 10^8 \text{ cm}^{-3}$, whereas observations were 2.77 times larger than the simulation,



319 indicating the existence of additional reaction pathways that likely propagated the
320 OH→RO₂ conversion efficiency. A significant discrepancy of radicals existed in the
321 heavy ozone concentration, with OH, HO₂, and RO₂ radicals concurrently underestimated
322 at noon by $8.23 \times 10^6 \text{ cm}^{-3}$, $3.94 \times 10^8 \text{ cm}^{-3}$ and $11.59 \times 10^8 \text{ cm}^{-3}$, respectively. The observed
323 HO₂/RO₂ ratio approached 1:1, while the model reflected an unreasonable ratio of 3:1,
324 indicating deficiencies in both primary sources and secondary propagation. Sensitivity
325 tests based on the full suite of radical measurement are performed to explore the missing
326 oxidation properties (Section 4.1).

327 Fig.5 displays the diurnal profiles of the ROx budget during different episodes. In
328 Semi I, formaldehyde photolysis showed a higher contribution (38.6%), while HONO
329 photolysis (21.0%) and ozone photolysis (24.7%) accounted for similar proportions in
330 primary sources. The contribution of photolysis from other OVOCs was comparable to
331 that of ozonolysis reactions (7.2% vs. 4.8%). However, in Semi II, the decreased
332 oxidation level was attributed to lower ROx sources, despite the similar proportions.
333 During the Heavy period, the primary sources dramatically increased (up to ~10 ppb/h),
334 with HCHO photolysis contributing the most, alongside other sources at common levels
335 (ranging between 1.74 – 2.66 ppb/h) in the YRD region (Ma et al., 2022). Fast HCHO
336 oxidation dominated the radical primary source during heavy ozone pollution, which
337 contrasts with the dominant role of HONO/O₃ in other megacities (Yang et al., 2022; Tan
338 et al., 2017b; Yang et al., 2021a).

339 The radical removal rate during the daytime was generally balanced with production
340 contributions. In the morning, owing to high NOx concentrations, radical termination was
341 mainly dominated by OH+NO₂, OH+NO, RO₂+NO, and RO₂+NO₂. Furthermore, the
342 formation of peroxy nitrate accounted for a certain proportion (~5%). As NOx
343 concentrations decreased after 10:00, self-reactions in peroxy radicals became significant.



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Fig. 5. The diurnal profiles of ROx budget during different polluted episodes (Semi I, Heavy, and Semi II). The pie chart denotes proportions in different parts during the daytime (10:00-15:00). The grey areas denote nighttime.

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3.3 Oxidation comparison

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The concentration of OH radicals during the daytime is a crucial indicator of atmospheric oxidation levels (Liu et al., 2021). Table 1 summarized radicals and related parameters for regions with similar latitudes ($32.0^\circ \pm 2^\circ$ N, $j(\text{O}^1\text{D}) \approx 2.5 \pm 0.5 \times 10^{-5} \text{ s}^{-1}$). The joint influence of solar radiation and local photochemistry resulted in megacities exhibiting intense oxidation levels in summer/autumn, characterized by OH radicals being maintained at approximately $10.0 \times 10^6 \text{ cm}^{-3}$ at noon. Notably, an observation in Houston revealed an OH concentration of nearly $20.0 \times 10^6 \text{ cm}^{-3}$, with k_{OH} of 10 s^{-1} (Mao et al., 2010). In areas such as Los Angeles, Pasadena, and Tokyo, the propagation efficiency of radicals was restricted due to fresh anthropogenic emissions. OH concentrations were only half of those observed in other megacities, with higher inorganic-dominated k_{OH} recorded (Pasadena, $\sim 20 \text{ s}^{-1}$) (George et al., 1999; Griffith et al., 2016; Yugo Kanaya et al., 2007). In the TROPSTECH observation, the observed k_{OH} exceeded the mean value at the same latitude ($>15 \text{ s}^{-1}$). Additionally, during the Heavy episode, higher OH concentration ($13.5 \times 10^6 \text{ cm}^{-3}$) was found, comparable to the highest level at regions with similar latitude (Houston 2000/2006, (Mao et al., 2010)). Synchronous elevation in radical concentration and reactivity indicated a strong oxidation level in the YRD region.

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The observations in the YRD region showed a stable conversion factor ($\text{OH} \cdot j(\text{O}^1\text{D})$) of $4 \pm 1 \times 10^{11} \text{ cm}^{-3} \text{ s}$, which was comparable to other megacities in the PRD, NCP, and SCB regions (Ma et al., 2022; Tan et al., 2019). The corresponding slope between OH

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368 concentration and solar radiation was used to quantify the oxidation efficiency from
 369 photolysis, and it was observed that a higher slope of $5.3 \times 10^{11} \text{ cm}^{-3} \text{ s}$ during the Heavy
 370 period indicated an active radical chemistry. This implies that there is a strong oxidation
 371 efficiency from photolysis in the YRD region.

372 During summer and autumn seasons, photochemical pollution is a common
 373 occurrence, as noted by (Tan et al., 2021). Analysis of radical concentration across
 374 different regions reveals that the YRD region exhibited concentrations higher than 10^7
 375 cm^{-3} , slightly lower than in Guangzhou in 2006 but consistent with observations in other
 376 megacities (Ma et al., 2022; Tan et al., 2017a; Lu et al., 2012; Yang et al., 2021a).
 377 Conversely, winter is characterized by haze pollution (Ma et al., 2019). An urban site in
 378 Shanghai reported a peak OH concentration of $2.6 \times 10^6 \text{ cm}^{-3}$, closely resembling the $1.7 -$
 379 $3.1 \times 10^6 \text{ cm}^{-3}$ range found in polluted winter atmospheres (Zhang et al., 2022a). Although
 380 no significant regional disparities in oxidation levels were detected in agglomerations,
 381 attention should be directed to the YRD region due to its elevated radical concentration,
 382 reactivity, and photolysis efficiency, signaling the need to investigate its role in radical
 383 chemistry.

384 **Table 1.** Summary of radical concentrations and related species concentrations at regions with similar latitude and
 385 megapolitan areas in China. All data are listed as the average in noontime (11:00–13:00).

Location	Latitude	Year	OH (10^6 cm^{-3})	k_{OH} (s^{-1})	$j(\text{O}^1\text{D})$ (10^{-5} s^{-1})	Slope ($10^{11} \text{ cm}^{-3} \text{ s}$)	X (ppb)	References
Regions with similar latitude								
Los Angeles	34.1° N	Sep 1993	6.0	-	-	-	-	(George et al., 1999)
Nashville	36.2° N	Jun–Jul 1999	10.0	10.2	3.0	3.3 ^c	-	(Martinez et al., 2003)
Houston	29.7° N	Aug 2000	20.0	9.0 ^b	3.0	6.7 ^c	-	(Mao et al., 2010)
Tokyo	35.6° N	Jul–Aug 2004	6.3	-	2.5	3.0	-	(Yugo Kanaya et al., 2007)
Houston	29.7° N	Sep 2006	15.0	11.0	3.1	5.0 ^c	-	(Mao et al., 2010)
Pasadena	34.1° N	May–Jun 2010	4.0	20.0	2.5	1.6 ^c	-	(Griffith et al., 2016)
Taizhou	32.6° N	May–Jun 2018	10.6	10.8 ^a	2.1	4.8	0.10	(Ma et al., 2022)
Chengdu	30.7° N	Aug 2019	10.0	8.0	2.2	4.1	0.25	(Yang et al., 2021a)
TROPSTECT (Heavy)	31.9° N	Sep 2020	13.5	16.0	2.6	5.3	0.50	This work
TROPSTECT (Semi)	31.9° N	Sep 2020	7.2	14.2	2.4	3.1	0.25	This work
Regions in megapolitan areas in China								
Guangzhou (PRD)	23.5° N	Jul 2006	12.6	17.9	3.5 ^b	4.5	0.85	(Lu et al., 2012)
Wangdu (NCP)	38.7° N	Jun–Jul 2014	8.3	15.0	1.8	4.5	0.10	(Tan et al., 2017b)
Beijing (NCP)	39.9° N	May–Jun 2017	9.0	30.0	2.4	3.8 ^c	~0	(Whalley et al., 2021)
Taizhou (YRD)	32.6° N	May–Jun 2018	10.6	10.8 ^a	2.1	4.8	0.10	(Ma et al., 2022)
Shenzhen (PRD)	22.6° N	Sep–Oct 2018	4.5	21.0	1.8	2.4	0.10	(Yang et al., 2022)



Chengdu (SCB)	30.7° N	Aug 2019	9.0	8.0	2.2	4.0	0.25	(Yang et al., 2021a)
Hefei (YRD)	31.9° N	Sep 2020	10.4	14.3	2.4	4.4	0.30	This work

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^a The modeled k_{OH} .

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^b Value only in the afternoon.

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^c Using the ratio of OH / j(O¹D)

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4 Discussion

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4.1 Measurement–model reconciliation for radicals

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4.1.1 OH underestimation

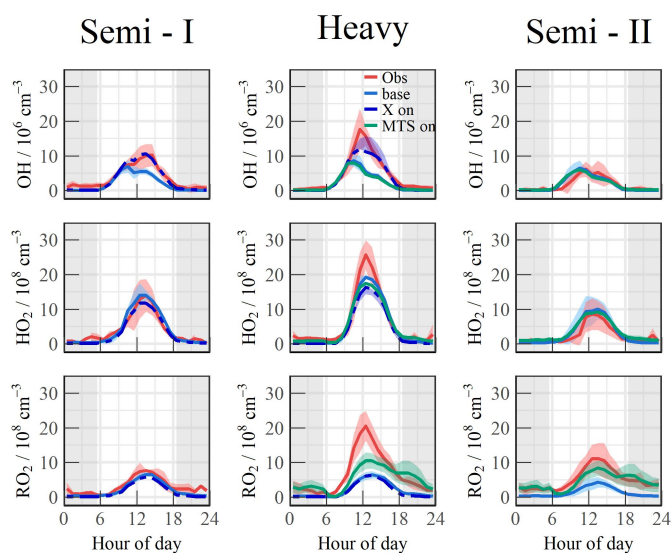
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Full suite of OH, HO₂, RO₂ and k_{OH} was utilized in the TROPSTECT campaign to untangle a thorough understanding of oxidation mechanisms where base model failed. One specific phenomenon was the absence of an OH source in situations where NO levels gradually decreased after 10:00. A sensitivity test was conducted introducing a species X, analogous to NO, to enhance OH regeneration (Fig. 6, RO₂→HO₂ and HO₂→OH) (Hofzumahaus et al., 2009). It was found that the addition of as little as 0.25 ppb X was sufficient to compensate for the full magnitude of the OH underestimation in the low NO region (Fig. 6). The employment of the X mechanism not only accelerated OH regeneration but also augmented the removal channel of peroxy radicals, which consequently led to a reduction in both HO₂ and RO₂ radical concentrations compared to the base scenario.

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The underdetermined radical sources in China were corresponding to the oxidation level (Ma et al., 2022; Tan et al., 2017a; Lu et al., 2012; Yang et al., 2021a; Wang et al., 2019b). The required X level typically ranged from 0.1 to 0.3 ppb, with the exception of the Backgarden observation which required 0.85 ppb X, as indicated in Table 1 (Lu et al., 2012). A minimum limit of 0.1 ppb X was established to account for any missing reactivity (Ma et al., 2022). Notably, throughout the entire observation, a strong agreement between the modeled and observed OH was achieved when a mixture of 0.25 ppb X was incorporated into the base scenario, consistent with the the order of magnitude in Chengdu (Yang et al., 2021a). During the Heavy period, the augmented photochemistry resulted in complex oxidation, necessitating an additional missing OH source equivalent to 0.5 ppb X to fully address the underestimation of OH.

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Fig. 6. The mean diurnal profiles of measured and modeled OH, HO₂ and RO₂ concentrations at different scenarios. Sensitivity tests included three scenarios (Scenario 1: base case; Scenario 2: X mechanism on. The dashed line represented the performance of 0.25 ppb X introduced in the Semi I and Heavy episodes, and the blue shadow denoted the upper limit for X influence (0.5 ppb); Scenario 3: monoterpene mechanism on; Both API and LIM were added into the base model as upper and lower limits for the influence of monoterpenes, and the mean of the two values was taken as the average effect. The grey areas denote nighttime.

4.1.2 RO₂ underestimation

The base scenario in Semi II is capable of accurately reproducing the concentrations of OH and HO₂ radicals within the data uncertainty. However, the simulated RO₂ concentration by the base model is only $3.78 \times 10^8 \text{ cm}^{-3}$, which does not align with the observed oxidation levels in YRD, indicating a clear discrepancy. This underestimation is similarly evident in the APHH observation in Beijing, as the highest observed concentration of RO₂ radicals reached $5.5 \times 10^9 \text{ cm}^{-3}$, far exceeding the level predicted by the MCM v3.3.1 mechanism (Whalley et al., 2021). The failure to reproduce the RO₂ concentration reflects the inadequacy of the mechanisms related to RO₂ radicals due to diverse oxidation reactions. This issue is further elucidated by previous studies, which highlighted the possibility of certain VOCs undergoing more intricate isomerization or fragmentation steps to sustain the long lifetime of RO₂ radicals (Whalley et al., 2018; Whalley et al., 2021).

The union of k_{OH} and RO₂ measurement can help reveal the magnitude of missing RO₂ as a hypothesis of sensitivity analysis. During the Semi II period, discrepancy of OH reactivity ($\sim 3 - 5 \text{ s}^{-1}$) between measurement and model suggested that an additional driving force was necessary to complete the OH to RO₂ step. The derivative of α -Pinene,

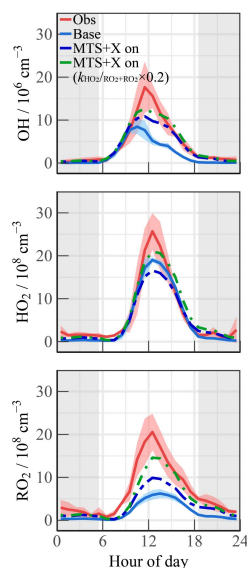


438 C₉O₂, could undergo four RO₂→RO₂ propagations before termination, thereby making it
439 a plausible source of the underestimation of RO₂ (Whalley et al., 2021). In the
440 EXPLORE-2018 campaign in the YRD region, the diurnal average of monoterpenes
441 peaked at 0.55 ppb between 11:00 and 13:00, with an oxidation efficiency of
442 approximately 1.5 ppb/h (Wang et al., 2022b). To compensate for the missing k_{OH} within
443 a reasonable range, approximately 0.4 ppb of monoterpene was introduced into the base
444 scenario. The RACM2 mechanism identified α -pinene (API) and limonene (LIM) as
445 representative monoterpenes species. Sensitivity tests were conducted by incorporating
446 API and LIM into models as upper and lower limits for the influence of monoterpenes,
447 with reaction rates of $5.3 \times 10^{-11} \text{ cm}^{-3} \text{ s}^{-1}$ and $1.6 \times 10^{-10} \text{ cm}^{-3} \text{ s}^{-1}$ at 298 K, respectively
448 (Ma et al., 2022). The mean of these values was considered the average effect of
449 monoterpenes chemistry, and depicted as the green line in Fig.6. In the 'MTS on' scenario,
450 the chemistry of peroxy radicals in Semi II was reasonably described by decreasing the
451 obs-to-mod ratio of peroxy radicals from 2.2 to 1.3. Furthermore, the introduction of
452 additional RO₂ isomerization steps had minimal impact on HO_x chemistry, with changes
453 in daytime OH and HO₂ concentrations of less than $5 \times 10^5 \text{ cm}^{-3}$ and $2.5 \times 10^7 \text{ cm}^{-3}$,
454 respectively. This demonstrates the robustness of HO_x radical in response to potential
455 monoterpene.

456 During the heavy ozone pollution, a more pronounced level of oxidation was
457 observed, as evidenced by the diurnal peak radical concentrations being 2.58, 1.22, and
458 3.02 times higher than the simulated values for OH, HO₂ and RO₂, respectively.
459 Incorporating monoterpenes into the base model led to a substantial 1.93-fold increase in
460 the modeled RO₂. However, the model was unable to reproduce over half of the
461 concentration due to unidentified oxidation processes, as illustrated in Fig. 7. Previous
462 researches highlighted the significance of investigating whether the sink pathways
463 accurately reflect conditions in the actual atmosphere (Tan et al., 2017a;
464 Woodward-Massey et al., 2023). In regions with low NO_x levels, peroxy radical removal
465 is primarily achieved through self-reaction. Manipulating the self-reaction rate of peroxy
466 radicals by approximately five-fold, and the extended lifetime counterbalance their
467 supplementary consumption by non-traditional regeneration mechanisms. Consequently,
468 the concentrations of OH, HO₂, and RO₂ radicals increased simultaneously, and the



469 obs-to-mod ratio of daytime RO_2 shifted from 3.02 to 1.37.

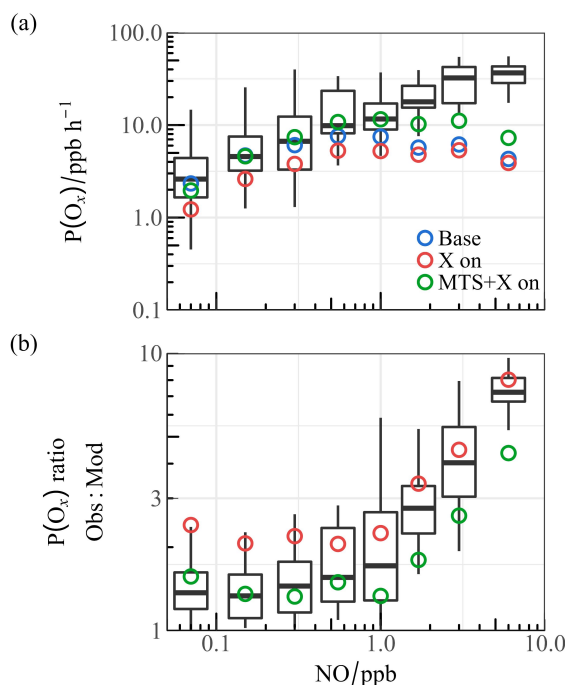


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471 **Fig. 7.** The mean diurnal profiles of measured and modeled OH, HO_2 and RO_2 concentrations in the Heavy period.
 472 The dashed blue line indicates the scenario synchronously considering monoterpene chemistry and non-traditional
 473 regeneration mechanisms (MTS+X). The dashed green line represents artificially reducing the rate of peroxy radical
 474 self-reaction (about 5 times) on top of the MTS+X scenario.

475 4.1.3 P(Ox) underestimation

476 Upon completing the hypothetical investigation into the radical underestimation, a
 477 sensitivity comparison between observed and modelled P(Ox) was conducted across the
 478 entire range of NO concentrations, as depicted in Fig. 8(a)(b). With increasing NO
 479 concentration, the overall P(Ox) amplified, reaching a maximum of approximately 30
 480 ppb/h. This variation has been validated through multiple observations in Wangdu, APHH,
 481 and other studies (Tan et al., 2017b; Whalley et al., 2021; Whalley et al., 2018). However,
 482 the imperfect understanding of the mechanisms related to peroxy radicals ultimately leads
 483 to misjudgment of the ozone production process in high NO regimes, with a degree of
 484 underestimation close to 10 times, as illustrated in Fig. 8(b).



485

486 **Fig. 8.** The relationship between NO and (a) P(Ox), (b) P(Ox) (Obs:Mod). Boxplot diagrams are used to illustrate
 487 the minimum, 25th percentile, median, 75th percentile, and maximum values of the observed dataset. The circles
 488 represent the median values for the base model as well as for different mechanisms added to the model within various
 489 ranges.

490 Although the inclusion of the X mechanism improves the agreement between
 491 simulated and observed OH concentrations in the low-NO range, it has a negative effect
 492 on the P(Ox) simulation. The introduction of a major source of RO₂ can help address the
 493 underestimation problem in the base scenario, as the lack of RO₂ species and related
 494 reaction rates is an important factor leading to deviations in the simulation of ozone
 495 production rates (Tan et al., 2017a). The combination of the X mechanism and
 496 monoterpene chemistry is shown to better harmonize the relationship between HO₂ and
 497 RO₂. Notably, the deficiency in the ozone generation mechanism was adequately
 498 explained within a certain range in the 'MTS+X On' scenario, leading to an enhancement
 499 in the simulation performance of P(Ox) in the high NO_x region (Fig. 8(b)). Therefore,
 500 reasonable simulation of the concentration of peroxy radicals is key to accurately
 501 quantifying the process of ozone generation.

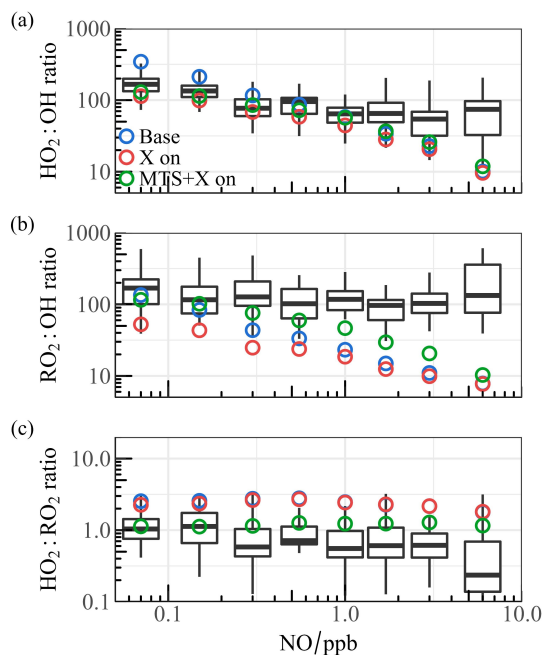
502 4.2 Effect of mechanism reconciliation on oxidation

503 Both radical concentration and oxidation coordinating deficiency are worthy of



504 examine (Fig. S6). To eliminate the influence of non-photolytic processes, only the
505 daytime concentration range with $j(\text{O}^1\text{D})$ greater than $5 \times 10^{-6} \text{ s}^{-1}$ was selected. The
506 boxplots illustrate the ratio of observation to simulation (base model), with the circles
507 representing the average values after integrating different mechanisms into the base
508 scenario. In the low NO regime ($\text{NO} < 1 \text{ ppb}$), the OH underestimation was consistently
509 prominent as NO concentration decreased, and the base model was able to reasonably
510 reflect the HO_2 distribution contrastly. As NO levels increased, the simulated OH
511 concentration aligned well with the observation, but both HO_2 and RO_2 concentrations
512 exhibited underprediction. RO_2 underestimation extended across the entire NO range, and
513 could rise to over 10 times when NO levels reached about 10 ppb. Sensitivity tests based
514 on the full suite of radical measurement revealed that the X mechanism accelerated OH
515 regeneration, and the introduction of larger RO_2 isomerization steps alleviated the
516 absence of certain sources by 2 to 4 times.

517 The coordinate ratios of radical serves as another test for ROx propagation (Fig. 9).
518 The observed HO_2/OH ratio is approximately 100, declining to some extent as the
519 concentration of NO increases, which is consistent with previous studies (Griffith et al.,
520 2016; Griffith et al., 2013). However, the base model does not accurately replicate the
521 curve depicting the change in HO_2/OH ratio, as shown in Fig. 9 (a). At low NO levels, the
522 ratio significantly overestimated and shows a steeper decline compared to the base
523 scenario as NO levels increase. Furthermore, the observed RO_2/OH ratios remain around
524 100, whereas the predicted values are significantly underestimated when NO exceeds 1
525 ppb (refer to Fig. 9(b)). In terms of the observed HO_2/RO_2 ratio, it maintains a relatively
526 constant trend within the range of 0.5 – 1.5, while the model overestimated by more than
527 twice, highlighting an inconsistency between the conversion of $\text{RO}_2 \rightarrow \text{HO}_2$. The
528 incorporation of the X mechanism has proven to be effective in a balanced HO_2/OH ratio
529 as illustrated in Fig.9(a), but amplifying the termination pathway for HO_2 and RO_2 , which
530 altered the coordination between RO_2 and OH across the entire NO range (Fig. 9(b)). The
531 connection between unconditional OH source and larger RO_2 isomerization in chemically
532 complex environments is key to fully understanding tropospheric chemistry, and a better
533 coordination of HO_2/OH , RO_2/OH , and HO_2/RO_2 ratios are established by incorporating
534 additional mechanisms.



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Fig. 9. The ratios for (a) HO_2/OH , (b) RO_2/OH , and (c) HO_2/RO_2 show a correlation with NO levels. Boxplot diagrams are used to illustrate the minimum, 25th percentile, median, 75th percentile, and maximum values of the observed dataset. The circles represent the median values for the base model as well as for different mechanisms added to the model within various ranges.

540

4.3 Special HCHO contribution accelerate ozone production

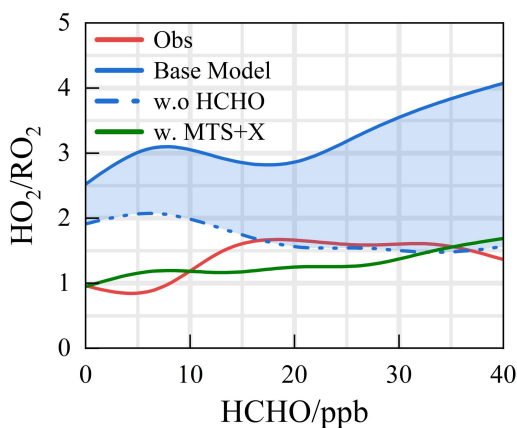
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Amplification of radical cycling can be distinguished by the OVOCs photochemical contribution (Qu et al., 2021). The special formaldehyde distribution was discovered during severe ozone pollution, with the average value of 21.81 ± 4.57 ppb (11:00 – 13:00), embodying the fast oxidation feedback to accelerate ozone production. The distribution ratio between peroxy radicals shows that the observed HO_2/RO_2 remains stable around ~ 1 , whereas the simulated values exhibit a strong positive correlation with photochemistry, ranging between 2-4 (Fig.10). When formaldehyde is unconstrained, the simulated HO_2/RO_2 contrarily aligned with observations, indicating that the photochemistry efficiencies of formaldehyde and other OVOCs is comparable under the current chemical mechanism. Since formaldehyde is a source of HO_2 radicals but does not directly produce RO_2 radicals, constraining formaldehyde can partially compensate for the HO_2 radical cycle, thereby enhancing the accuracy of HOx radical chemistry research. However, the base model underestimates the OVOCs feedback to RO_2 radicals during the heavy pollution. The introduction of larger RO_2 isomerization steps (possibly caused by OVOCs,

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555 monoterpenes) kept both the HO_2/RO_2 ratio and the $\text{P}(\text{Ox})$ comparable to observed
 556 results in the MTS+X scenario during different pollution periods, which chemically
 557 validated the hypothesis of the sensitivity analysis (Section 4.2&Fig.S7).



558

559 **Fig. 10.** The ratios for HO_2/RO_2 show a correlation with HCHO levels. The blue shading represents the range of
 560 variation from constrained to unconstrained formaldehyde conditions.

561 The relationship between radical cycling and ozone pollution is established through
 562 chain length (ChL) and ozone production rate ($\text{P}(\text{Ox})$) (Fig.S8). Overall, the average
 563 chain length during different periods ranges from 1.4 to 6.1, similar to results from
 564 multiple photochemical pollution seasons (Yang et al., 2021a; Martinez et al., 2003;
 565 Shirley et al., 2006). To evaluate the impact of high formaldehyde concentrations on
 566 severe pollution formation, the peak chain length decreased by 38.22%, 55.91%, and
 567 39.15% after removing formaldehyde constraints from the base model. $\text{P}(\text{Ox})$ and ChL
 568 exhibit overall temporal consistency. Constraining formaldehyde increased the peak
 569 radical cycling rate from 1.94 to 4.45 during heavy pollution periods, and the rapid
 570 radical cycling accelerates secondary pollution formation, with peak ozone production
 571 rates increasing by 51.54%. With the implementation of full-chain radical detection,
 572 additional measurements targeting more OVOCs should also be conducted to untangle
 573 the RO_2 -related imbalance, and then accurately elucidating the oxidation under severe
 574 ozone pollution.

575 **5 Conclusion**

576 The full suite radical measurement of OH, HO_2 , RO_2 and k_{OH} was first deployed in
 577 the YRD region (TROPSTECT) and encountered with a prolonged ozone pollution in



578 September 2020. The diurnal peaks of radicals exhibited considerable variation due to
579 environmental factors, showing ranges of 3.6 to $27.1 \times 10^6 \text{ cm}^{-3}$ for OH, 2.1 to 33.2×10^8
580 cm^{-3} for HO₂, and 4.9 to $30.5 \times 10^8 \text{ cm}^{-3}$ for RO₂. Continuous k_{OH} data fell within a range
581 of $8.6 - 30.2 \text{ s}^{-1}$, demonstrating the dominant behavior of organic species in diurnal
582 reactivity. Furthermore, observations in the YRD region were found to be similar to those
583 in other megacities, suggesting no significant regional differences in oxidation levels
584 were observed in agglomerations overall.

585 At a heavy ozone pollution episode, the oxidation level reached intensive compared
586 with other sites, and the simulated OH, HO₂, and RO₂ radicals provided by the
587 RACM2-LIM1 mechanism failed to adequately match the observed data both in radical
588 concentration and coordinate ratios. Sensitivity tests based on the full suite of radical
589 measurement revealed that the X mechanism accelerated OH regeneration, and the
590 introduction of larger RO₂ isomerization steps alleviated the RO₂-related imbalance. The
591 incorporation of complex processes enabled better coordination of HO₂/OH, RO₂/OH,
592 and HO₂/RO₂ ratios, and adequately addressed the deficiency in the ozone generation
593 mechanism within a certain range. The hypothesis of sensitivity analysis can be
594 chemically validated by the special HCHO contribution to oxidation. Constraining
595 HCHO increased the ChL from 1.94 to 4.45, leading to a 51.54% increase in ozone
596 production during the heavy pollution. This study enabled a deeper understanding of the
597 tropospheric radical chemistry at play. Notably,

598 ✓ A full suite of radical measurement can untangle the gap-bridge for the base model in
599 more chemically-complex environments as an hypothesis of sensitivity tests.

600 ✓ Additional measurements targeting more OVOCs should also be conducted to fulfill
601 the RO₂-related imbalance, and then accurately elucidating the oxidation under
602 severe ozone pollution.

603



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610 **Data availability**

611 The data used in this study are available upon request (rzhu@aiofm.ac.cn).

612 **Author contributions**

613 WQ Liu, PH Xie, RZ Hu contributed to the conception of this study. RZ Hu and GX
614 Zhang performed the data analyses and manuscript writing. All authors contributed to
615 measurements, discussed results, and commented on the paper.

616 **Competing interests**

617 The contact author has declared that none of the authors has any competing interests.

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