

Accurate Elucidation of Oxidation Under Heavy Ozone Pollution: A Full Suite of Radical Measurement In the Chemical-complex Atmosphere

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Abstract: The Yangze River Delta (YRD) in China encountered with prolonged ozone pollution in September 2020, which had significant impacts on the respiratory, dermatological, and visual health of local residents. To accurately elucidate the limitations of oxidation processes in the chemical-complex atmosphere, a full suite of radical measurements (OH, HO₂, RO₂, and k_{OH}) was established in YRD region for the first time. The diurnal peaks of radicals exhibited considerable variation due to environmental factors, showing ranges of 3.6 to 27.1×10⁶ cm⁻³ for OH, 2.1 to 33.2×10⁸ cm⁻³ for HO₂, and 4.9 to 30.5×10⁸ cm⁻³ for RO₂. At a heavy ozone pollution episode, the oxidation capacity reached an intensive level compared with other sites, and the simulated OH, HO₂, and RO₂ radicals provided by the RACM2-LIM1 mechanism failed to adequately match the observed data both in radical concentration and experimental budget analysis. Sensitivity tests utilizing a comprehensive set of radical measurements revealed that the higher aldehyde mechanism (HAM) effectively complements the non-traditional regeneration of OH radicals, yielding enhancements of 4.4% - 6.0% compared to the base scenario, while the concentrations of HO₂ and RO₂ radicals have shown increments of about 7.4% and 12.5%, respectively. Notably, RO₂ radical concentrations exhibit a pronounced sensitivity to autooxidation, with the incorporation of additional OVOCs potentially boosting simulated RO₂ radical concentrations by 20% to 40%. The incorporation of larger alkoxy radicals stemming from monoterpenes has refined the consistency between measurements and modeling in the context of ozone production under elevated NO levels, diminishing the disparity from 4.17 to 2.33. This outcome corroborates the hypothesis of sensitivity analysis as it pertains to ozone formation. Moving forward, by implementing a comprehensive radical detection approach, further investigations should concentrate on a broader range of OVOCs to rectify the imbalance associated with RO₂ radicals, thereby providing a more precise understanding of oxidation processes during severe ozone pollution episodes.

Keywords: FAGE-LIF; Full-chain detection; Radical; P(Ox); OVOCs;

1 Introduction

In recent years, China's rapid economic development has led to severe environmental pollution problems, which significantly impacted the respiratory, dermatological, and visual health of local residents (Wang et al., 2022c; Huang et al., 2018). This has raised concerns about the coexistence of regional primary and secondary pollution, making air quality improvement efforts a focal point (Liu et al., 2021; Wang et al., 2022a). In the complex atmosphere, near-surface ozone (O_3) is formed through continuous photochemical reactions between nitrogen oxides ($NO_x \equiv NO + NO_2$) and volatile organic compounds (VOCs) under light conditions, while hydroxyl radicals (OH) serve as the main oxidant in the troposphere, converting VOCs into hydroperoxy (HO_2) and organic peroxy (RO_2) radicals (Rohrer et al., 2014; Hofzumahaus et al., 2009). Additionally, the oxidation of nitric oxide (NO) and peroxy radicals produce nitrogen dioxide (NO_2), which is the sole photochemical source of ozone (Lu et al., 2012; Stone et al., 2012).

Despite numerous experimental and theoretical explorations to establish the radical-cored photooxidation mechanism in the troposphere, field observations were primarily focused on HO_x ($HO_x \equiv OH + HO_2$) radicals due to the limitations of detection technology (Kanaya et al., 2012; Lu et al., 2012; Hofzumahaus et al., 2009; Yugo Kanaya and Tanimoto, 2007; Ren et al., 2008; Stone et al., 2012; Levy, 1971). Recent advancements in detection technology, such as the application of a new LIF technique (ROxLIF), have made the detection of RO_2 radicals possible (Whalley et al., 2013; Tan et al., 2017a). Moreover, the union of comprehensive field campaigns and box model, has proven to be an effective method for verifying the integrity of radical chemistry at local to global scales (Lu et al., 2019b; Tan et al., 2018). Several experiments have indicated that the existing atmospheric chemical mechanism posed challenges in deepening the understanding of the regional pollution explosion (Whalley et al., 2021; Slater et al., 2020; Tan et al., 2017a; Woodward-Massey et al., 2023). For instance, the observation of up to $4 \times 10^9 \text{ cm}^{-3}$ of RO_2 radical in the center of Beijing in 2017 (APHH) was significantly underestimated by the MCM v3.3.1 mechanism (Whalley et al., 2021). Further exploring the unreproducible concentration and the oxidation process in the chemical-complex atmosphere is deemed necessary (Whalley et al., 2021; Woodward-Massey et al., 2023).

The YRD region, situated between the North China Plain (NCP) and Pearl River Delta (PRD), is highly prone to regional transport interactions and aerosol-boundary layer feedback (Jia et al., 2021; Huang et al., 2020). In September 2020, the YRD region experienced a severe episode of secondary pollution, with both the daily maximum 8-hour average ozone (MDA8) and daily average PM_{2.5} concentrations surpassing the pollution threshold, distinguishing it from other megacities (Fig. S1). In an effort to gain a better understanding between the complex radical chemistry and the intensive oxidation level, TROPSTECH-YRD (The experiment on Radical chemistry and Ozone Pollution perSpectively: long-Term Elucidation of the photochemiCal oxidation in the Yangze River Delta) was conducted in Hefei during September 2020. Accurate elucidation of the oxidation process under heavy ozone pollution was provided by a full suite of radical measurement (OH, HO₂, RO₂ and k_{OH}) in the chemical-complex atmosphere.

2 Materials and methods

2.1 Site description and instrumentation

The TROPSTECH observation was conducted from 1 to 20 September 2020 at the Science Island background station (31.9° N, 117.2°E) in Hefei, a typical megacity located in the central region of Anhui Province within the Yangtze River Delta. The station is situated on a peninsula with abundant vegetation to the northwest of urban areas and is in close proximity to Dongpu Lake, which is only 200 meters away, and the main road, positioned 250 meters southward (Fig. 1). Consequently, the relatively enclosed environment exhibits typical suburban characteristics of anthropogenic emissions. The station is located in the transition region between urban and suburban areas, reflecting the regional transpor of pollution in Hefei and its surrounding areas.



Fig. 1. (a) The location of the measurement site (source: © Google Earth).
 (b) The close shot of the measurement site location (source: © Google Earth).
 (c) The actual image for the LIF-Box.

Regarding the instrumentation, a group of oxidation-related instruments were installed on the 7th floor of the Comprehensive Building at the Anhui Institute of Optics and Fine Mechanics (AIOFM), with all sampling outlets positioned more than 20 meters above the ground. The details of the instruments measuring various parameters such as meteorological factors (WS, WD, T, RH, P, Jvalues), gas pollutants (O₃, CO, SO₂, NO, NO₂, HONO, HCHO, PAN), and non-methane hydrocarbons (NMHCs) are provided in Table S1.

The measurement of NO, NO₂, O₃, CO, and SO₂ was carried out using commercial Thermo Electron model series instruments. Thereof, NO and NO₂ were measured using a chemical fluorescence method (CL) with an enhanced trace-level NO-NO₂-NO_x analyzer (PKU-PL), which achieved a detection limit of 50 ppt (Tan et al., 2017a). The detection of O₃ and SO₂ was conducted through Thermo Electron model 49i and 43i, respectively, while Thermo Electron model 48i was utilized for CO detection. Cavity ring-down spectroscopy (CRDS, Picarro-G2401) was employed for CO detection in parallel, and another ultraviolet absorption instrument (Ecotech EC9810B) was for ozone detection. The instrument inlets were placed within 5 meters of each other for comparison.

To ensure measurement accuracy, the instruments in the campaign underwent zero point calibration procedures during the early (August 31st) and late (September 21st) observation periods, and cross-calibrations for O₃ and CO measurements were carried out during the middle (September 9th). Furthermore, additional zero calibration for Thermo 48i CO detection was undertaken daily from 0:00-0:30 to minimize shift correction. The comparison results revealed high consistency within the instrument accuracy range for both CO and O₃ measurements (Fig. S2(a)(b)).

HONO was detected using a home-built instrument by cavity-enhanced absorption spectroscopy (CEAS), while formaldehyde was determined by the Hantzsch method (SDL MODEL 4050) (Duan et al., 2018; Yang et al., 2021a). An automated gas chromatograph equipped with a mass spectrometer and flame ionization detector (GC-FID/MS) was employed for the online measurement of 99 VOCs species. Information table for parts of the VOC monitoring species by online GC-MS/FID was listed in Table S2.

The eight crucial photolysis frequencies ($j(\text{NO}_2)$, $j(\text{H}_2\text{O}_2)$, $j(\text{HCHO_M})$, $j(\text{HCHO_R})$,

j(HONO), j(NO₃_M), j(NO₃_R), j(O¹D)) were directly measured by a photolysis spectrometer (Metcon, Germany). The unmeasured photolysis frequencies of the remaining active species were computed using Eq.(1):

$$j = l \cdot \cos(\chi)^m \cdot e^{-n \cdot \sec(\chi)} \quad (1)$$

The variations in photolysis frequency due to solar zenith angle (χ) were adjusted based on the ratio of observed and simulated j(NO₂). The optimal values for parameters (l , m , and n) for different photolysis frequencies were extensively detailed by the MCM v3.3.1 (http://mcm.york.ac.uk/parameters/photolysis_param.htm) (Jenkin et al., 2003; Jenkin et al., 1997).

2.2 Radical measurement

2.2.1 OH, HO₂, RO₂ Concentrations

The laser-induced fluorescence instrument developed by the Anhui Institute of Optics and Fine Mechanics (AIOFM-LIF) was used to simultaneously detect the concentrations of OH, HO₂, and RO₂ radicals, along with OH reactivity (k_{OH}). The OH radical was directly measured by detecting on-resonance fluorescence excited by a 308 nm laser. An indirect measurement for HO₂ was carried out after converting it to OH at a fixed efficiency (Heard and Pilling, 2003).

The laser utilized for fluorescence excitation is a high-frequency tunable dye laser that emits a 308 nm laser, with the laser power divided into a ratio of 0.45:0.45:0.08:0.02. Of this power, 90% is directed towards fluorescence cells for concentration and reactivity detection via optical fibers, respectively. 8% of the laser power is directed to a reference cell, while the remaining 2% is used to monitor real-time power fluctuations. The laser is transmitted through HO₂, OH, and RO₂ cells in turn via a coaxial optical path. Two photodiodes are set up at the end of the reference cell and RO₂ detection cell, respectively. The voltage signals and power fluctuations are compared synchronously to diagnose the laser stability. To maintain detection efficiency, the power inside the measurement cells should not be less than 10 mW. Sampling nozzles of 0.4 mm are deployed above OH and HO₂ cells for efficient sampling at a flow rate of approximately 1.1 SLM, and the pressure for all fluorescence cells are maintained at 400 Pa. The micro-channel plate (MCP) detects the weak fluorescence signal collected by lens systems with low noise and high gain. Additionally, a digital delay generator (DG645) optimizes the timing control

between the laser output, signal detection, and data acquisition. All of these modules are integrated into a sampling box with constant air conditioning, except for the laser.

The detection of RO₂ radicals is more complex compared to the integrated detection of OH and HO₂ radicals (Whalley et al., 2013). To achieve the complete chemical conversion from RO_x to HO₂, a crucial role is played by a 66 mm×830 mm aluminium flow tube, whose performance has been confirmed through the CHOOSE-2019 field campaign (Li et al., 2020). A mixture of 0.17% CO and 0.7 ppm NO injected into the flow tube facilitates the reduction of heterogeneous radical loss and enhancement of conversion efficiency. The sampling flow is limited to 7 SLM by a 1 mm nozzle, and the tube pressure is maintained at 25 hPa. In contrast to the HO_x cells, the large-diameter nozzle (4 mm) is equipped above the cell, and a high concentration of NO (~300 ppm) facilitates the full magnitude HO₂→OH conversion.

The observation data (H₂O, O₃) is combined with experimental characterization to eliminate ozone photolysis interference, and most interference signals are excluded by utilizing wavelength modulation (Zhang et al., 2022a). A comparison experiment with PKU-LIF demonstrated the consistency of OH measurement in complex atmosphere (Zhang et al., 2022b). An additional atmospheric oxidation observation was conducted in the same location and season in 2022 with a chemical modulation method to determine the chemical background of OH radicals (Fig. S3). During the ozone pollution (2022.9.29-2022.10.3), the daytime peaks of ozone concentration above 75 ppb, accompanied by alkene species approaching ~10 ppb. The diurnal concentration of isoprene was also a high level (>1 ppb). The chemical conditions are more favourable to induce OH interference than in the TROPSTECH campaign, while the OH concentrations achieved by chemical modulation (OH_{chem}) and wavelength modulation (OH_{wav}) were in good agreement. No obvious chemical background was observed by deploying an inlet pre-injector. Therefore, it is not expected that OH measurement in the present study was affected by internal interference.

For HO₂ measurement, lower NO concentration ($\sim 1.6 \times 10^{12} \text{ cm}^{-3}$, corresponding to ~15% conversion efficiency) are selected to limit the RO₂→HO₂ interference to less than 5% (Wang et al., 2021). Since only the total-RO₂ mode is used for the campaign, the additional uncertainty introduced by RO₂/R(OH)O₂ classification is negligible (Tan et al.,

2017b). The observed maximum daily PAN (11:00-14:00) is only 1.15 ± 0.67 ppb, resulting in a calculated PAN-pyrolytic interference for RO_2 measurement of less than 1 ppt (Fuchs et al., 2008). The general applicability of AIOFM-LIF in complex atmosphere has been demonstrated through various campaigns (Zhang et al., 2022b; Wang et al., 2021; Wang et al., 2019a).

To complete the calibration task, a standard source stably generates equal amounts of OH and HO_2 radicals (Wang et al., 2020). The radical source is also capable of yielding specific RO_2 by titrating hydrocarbon with OH. It is noteworthy that CH_3O_2 has the highest mixing ratio in the RO_2 species, thus it was chosen to represent for sensitivity calibration. The instrument is calibrated every two days, except during rainy weather. The limit of detection (LOD) for OH, HO_2 , and RO_2 in different cells with a typical laser power of 10 mW is measured at $3.3 \times 10^5 \text{ cm}^{-3}$, $1.1 \times 10^6 \text{ cm}^{-3}$, and $2.5 \times 10^6 \text{ cm}^{-3}$, respectively (60 s, 1σ). Measurement accuracy for OH, HO_2 , and RO_2 radicals are reported to be 13%, 17%, and 21%, respectively.

2.2.2 OH reactivity(k_{OH})

The detection of k_{OH} in the atmosphere, defined as the reciprocal of OH lifetime, was conducted using a laser flash photolysis laser-induced fluorescence (LP-LIF) instrument (Lou et al., 2010). The configuration structure for k_{OH} measurement has been detailed in a previous study (Liu et al., 2019). The flow tube in the OH production-reaction unit is at ambient pressure, with a gas flow rate of 17 SLM. A pulsed laser beam (266 nm with an average power of 15 mJ) is output from a frequency-quadrupled Nd:YAG laser, which generates stable OH radical through flash photolysis of ambient ozone in the flow tube. Consistent and stable production of OH radicals is ensured by maintaining a stable concentration of reactants, flow field, and laser energy. Under conditions of 80 ppb O_3 and 8000 ppm water vapor concentration, OH radicals produced in the flow tube remains at the concentration order of 10^9 cm^{-3} . Subsequently, the OH radicals are sampled through a nozzle into a fluorescence cell. The OH fluorescence signal is then detected using laser pump and probe techniques and is fitted to calculate the slope of OH decay (k_{OH}). The detection accuracy, achieved with an integration time of 180 s, is 0.3 s^{-1} (1σ).

2.3 Observation-Based Model

The Regional Atmospheric Chemical Mechanism version 2 (RACM2) incorporating

the latest Leuven isoprene mechanism (LIM) was utilized to simulate the concentrations and reactions of OH, HO₂, and RO₂ radicals (Stockwell et al., 1997; Griffith et al., 2013; Peeters et al., 2014). The RACM2-LIM1 mechanism was specifically involved with fewer species compared to the explicit MCM mechanism, thus ensuring higher operational efficiency (Liu et al., 2022). The comprehensive list of model constraints was provided in Table S3. The measured NMHCs include 29 alkanes, 11 alkenes, 15 aromatics, as well as acetylene and isoprene. For the base scenario, boundary conditions were established using the observed species, with assumed concentrations of hydrogen (H₂) and methane (CH₄) at 550 ppb and 1900 ppb, respectively. An ozone-simulation test was conducted to determine the suitable atmospheric lifetime (τ_D) for the base model. At the lifetime of 24 hours, with a corresponding first-order loss rate of 1.1 cm/s (assuming a boundary layer height of 1 km), the simulated ozone concentration closely matched the observed values (Fig. S4). To improve the model-measurement consistency between OH, HO₂ and RO₂ radicals, a series of sensitivity analyses were performed to evaluate the impacts of potential mechanisms, as detailed in Table 1. The time resolution of all constraints was uniformly set to 15 minutes through averaging or linear interpolation. To reinitialize unconstrained species to a steady-state, three days of data were input in advance as the spin-up time.

Table.1. The sensitive test scenarios utilized to improve the model-measurement consistency between OH, HO₂ and RO₂ radicals.

Scenario	Configuration	Purpose
Base	RACM2 updated with isoprene reaction scheme (LIM)	The base case with the species involved in Table S3 are constrained as boundary conditions.
X on	As the base scenario, but add the X mechanism, and the X level is between 0.25 - 0.5 ppb.	To untangle the missing OH source where base scenario failed.
MTS on	As the base scenario, but add a monoterpene source, and the monoterpene level is ~0.4 ppb.	Utilizing monoterpene-derived RO ₂ to represent the alkoxy radicals with rather complex chemical structures.
MTS+X on	As the base scenario, but both the X mechanism and monoterpene source are considered.	To consider both the missing OH and RO ₂ sources.
HAM on	As the base scenario, but add the reactive aldehyde chemistry.	To provide a test of whether the proposed mechanism can explain the missing OH source.
HAM on (4 × ALD)	As the base scenario, but add the reactive aldehyde chemistry, and the concentration of ALD was amplified by a factor of 4.	To quantify the impact of missing aldehyde primary emissions on ROx chemistry.
Ozone simulation	As the base scenario, but remove the constraints of the observed ozone and NO concentrations.	To test the suitable lifetime for the base model.
HCHO simulation	As the base scenario, but remove the constraint of the observed HCHO concentration.	To test the simulation effect of the existing mechanism on formaldehyde

The local formation of ozone can be accurately quantified through the online measurement of ROx radicals (Tan et al., 2018). To overcome the interference from NO, the total oxidant (Ox), which is defined as the sum of NO₂ and O₃, can serve as a reliable parameter to indicate the level of oxidation. Eq.(2) shows that the rate of NO oxidation by peroxy radicals is equivalent to the production of O₃, denoted as F(Ox):

$$F(O_x) = k_{HO_2+NO}[NO][HO_2] + \sum_i k_{RO_2^i+NO}[NO]RO_2^i \quad (2)$$

The major loss pathways for Ox encompass ozone photolysis, ozonolysis reactions, and radical-related reactions (OH/HO₂+O₃, OH+NO₂), represented as D(Ox) in Eq.(3):

$$D(O_x) = \varphi_{OH}j(O^1D)[O_3] + \sum_i \{ \varphi_{OH}^i k_{Alkenes+O_3}^i [Alkenes][O_3] \} + (k_{O_3+OH}[OH] + k_{O_3+HO_2}[HO_2])[O_3] + k_{OH+NO_2}[OH][NO_2] \quad (3)$$

Here, the OH yields from ozone photolysis and ozonolysis reactions are denoted as φ_{OH} and φ_{OH}^i , respectively.

The net photochemical Ox production rate in the troposphere, denoted as P(Ox) in Eq.(4), can therefore be calculated as the difference between Eqs. (2) and (3):

$$P(O_x) = F(O_x) - D(O_x) \quad (4)$$

2.4 Experimental budget analysis

In this study, an experimental radical budget analysis was also conducted (Eqs. (5) - (12)). Unlike model studies, this method relies solely on field measurements (concentrations and photolysis rates) and chemical kinetic data, without depending on concentrations calculated by models (Whalley et al., 2021; Tan et al., 2019b). Given the short-lived characteristics of OH, HO₂, and RO₂ radicals, it is expected that the concentrations are in a steady state, with total production and loss rates being balanced (Lu et al., 2019a). By comparing the known sources and sinks for radicals, unknown processes for initiation, transformation and termination can be determined.

$$P(OH) = j_{HONO}[HONO] + \varphi_{OH}j(O^1D)[O_3] + \sum_i \{ \varphi_{OH}^i k_{Alkenes+O_3}^i [Alkenes][O_3] \} + (k_{HO_2+NO}[NO] + k_{HO_2+O_3}[O_3])[HO_2] \quad (5)$$

$$D(OH) = [OH] \times k_{OH} \quad (6)$$

$$P(HO_2) = 2 \times j_{HCHO_R}[HCHO] + \Sigma i \{ \varphi_{HO_2}^i k_{Alkenes+O_3}^i [Alkenes][O_3] \} \\ + (k_{HCHO+OH}[HCHO] + k_{CO+OH}[CO])[OH] \\ + \alpha k_{RO_2+NO}[NO][RO_2] \quad (7)$$

$$D(HO_2) = (k_{HO_2+NO}[NO] + k_{HO_2+O_3}[O_3] + k_{HO_2+RO_2}[RO_2] \\ + 2 \times k_{HO_2+HO_2}[HO_2])[HO_2] \quad (8)$$

$$P(RO_2) = \Sigma i \{ \varphi_{RO_2}^i k_{Alkenes+O_3}^i [Alkenes][O_3] \} \\ + k_{OH}[VOCs][OH] \quad (9)$$

$$D(RO_2) = \{ (\alpha + \beta) k_{RO_2+NO}[NO] + (2 \times k_{RO_2+RO_2}[RO_2] \\ + k_{HO_2+RO_2}[HO_2]) \} [RO_2] \quad (10)$$

$$P(RO_x) = \Sigma i \{ (\varphi_{OH}^i + \varphi_{HO_2}^i + \varphi_{RO_2}^i) k_{Alkenes+O_3}^i [Alkenes][O_3] \} + j_{HONO}[HONO] \\ + \varphi_{OH} j(O^1D)[O_3] + 2 \times j_{HCHO_R}[HCHO] \quad (11)$$

$$D(RO_x) = (k_{OH+NO_2}[NO_2] + k_{OH+NO}[NO])[OH] + \beta k_{RO_2+NO}[NO] \\ + 2 \times (k_{RO_2+RO_2}[RO_2][RO_2] + k_{HO_2+RO_2}[HO_2][RO_2] \\ + k_{HO_2+HO_2}[HO_2][HO_2]) \quad (12)$$

278 In which, $j(HONO)$, $j(O^1D)$ are the measured photolysis rates of HONO and O_3 ,
 279 respectively, and j_{HCHO_R} is the measured photolysis rate for the channel of
 280 formaldehyde photolysis generating HO_2 . φ_{OH} represent the OH yield in the O_3
 281 photolysis reaction. φ_{OH}^i , $\varphi_{HO_2}^i$ and $\varphi_{RO_2}^i$ are the yields for the ozonolysis reaction
 282 producing OH, HO_2 , and RO_2 , respectively. α is the proportion of RO_2 radicals reacting
 283 with NO that are converted to HO_2 , and β is the proportion of alkyl nitrates formation,
 284 which are set to 1 and 0.05, respectively(Tan et al., 2019b).

285 3 Results

286 3.1 Overview of Measurement

287 During the observation period, the meteorological parameters and trace gas
 288 concentrations were plotted in Fig. S5. The timeseries revealed that the peak temperature
 289 exceeded 30°C, and the humidity levels remained between 30 – 50% during the daytime.
 290 The photolysis rates were observed to peak at noon (11:00 – 13:00), with $j(O^1D)$ and
 291 $j(NO_2)$ reaching approximately $3 \times 10^{-5} s^{-1}$ and $8 \times 10^{-3} s^{-1}$, respectively. Brief rainfall events
 292 temporarily happened on September 10th, 15th, and 17th, but totally favorable
 293 meteorologies induced the prolonged ozone pollution. The daily maximum 8-hour

average ozone concentration (MDA8), as depicted in Fig. 2, consistently exceeded the Chinese Grade I national air quality standard (GB3095-2012) throughout the observation, with nine days exceeding the Grade II standard.

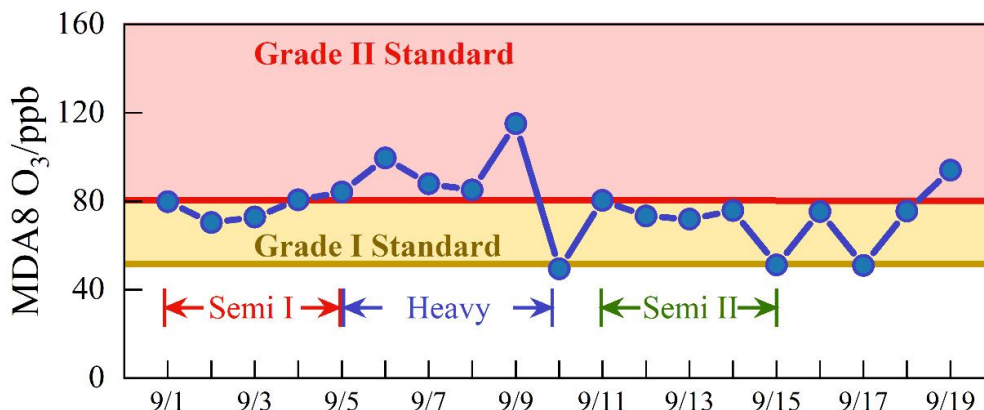


Fig. 2. The daily maximum 8 h average O₃ during the campaign. The yellow and red lines denote the Grade I and Grade II national standards for O₃, respectively. Brief rainfall events temporarily happened on 10, 15, and 17 Sep.

The ozone pollution can be categorized into three continuous periods based on pollution levels, which disclose transitional ‘Semi - Heavy - Semi’ pollution characteristics. Fig. 3 depicts daily variations in meteorological and trace gas concentrations for different periods. During the Semi I (1 to 5 September) and Semi II (11 to 14 September) periods, the MDA8 levels exceeded Grade I standard, with an average value of 75.92 ± 5.14 ppb and 75.45 ± 3.73 ppb, respectively. Notably, NO levels peaked around 9:00 and rapidly decreased to a few hundred ppt due to photochemistry. In addition, HONO and NO₂ exhibited bimodal variations, with diurnal concentration ranges of 0.09 – 0.50 ppb and 3.35 – 13.77 ppb, respectively. The HONO/NO₂ ratios during both Semi periods were consistent with previous urban/suburban observations, with daytime values of 0.049 ± 0.014 and 0.035 ± 0.012 , respectively (Yang et al., 2021b; Shi et al., 2020; Hu et al., 2022). Isoprene levels accumulated during the day and decreased at night during both Semi pollution episodes, with a diurnal average concentration in Semi II only 49.3% of that in Semi I (0.71 ± 0.087 ppb vs 0.35 ± 0.073 ppb). Formaldehyde, as the key oxidation species, exhibited a concentration profile mirroring that of isoprene, with significantly higher concentrations ranging from 1.20 to 36.34 ppb compared to other urban regions (Ma et al., 2022; Yang et al., 2022; Tan et al., 2017b; Yang et al., 2021a). Heavy pollution episodes from 5 to 9 September resulted in daytime ozone concentration as high as 129.9 ppb, and oxidation-related species such as HCHO, HONO, NO_x, and

319 VOCs increased synchronously compared to other days.

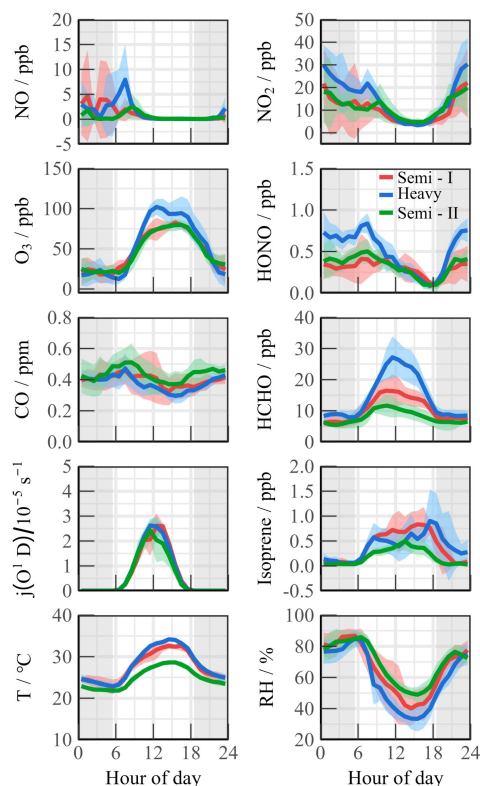


Fig. 3. Mean diurnal profiles of observed meteorological and chemical parameters during the campaign. Three periods were divided for subsequent study (Semi I, Heavy, and Semi II).

3.2 ROx radical concentrations and budgets

The observed and modeled timeseries for OH, HO₂, RO₂, and k_{OH} during the observation time are depicted in Fig. 4. The diurnal peaks of radicals exhibited a wide span due to changes in environmental conditions, with ranges of $3.6 - 27.1 \times 10^6 \text{ cm}^{-3}$ for OH, $2.1 - 33.2 \times 10^8 \text{ cm}^{-3}$ for HO₂, and $4.9 - 30.5 \times 10^8 \text{ cm}^{-3}$ for RO₂. Continuous data for k_{OH} observation were acquired within a range of $8.6 - 30.2 \text{ s}^{-1}$. Fig. S6 presents the diurnal profiles of the observed and modeled values during different episodes. The diurnal maximum of OH radical at noon differed between Semi I and Semi II, with $9.28 \times 10^6 \text{ cm}^{-3}$ and $5.08 \times 10^6 \text{ cm}^{-3}$, respectively, while total peroxy radicals (HO₂+RO₂) remained at similar levels with $19.43 \times 10^8 \text{ cm}^{-3}$ and $18.38 \times 10^8 \text{ cm}^{-3}$. Additionally, the distribution of peroxy radicals are not similar in the two Semi periods, with HO₂/RO₂ ratios of 1.69:1 and 0.76:1, respectively, which reflects the uneven oxidation levels between Semi I and Semi II. During the Heavy ozone pollution, the averaged OH, HO₂, and RO₂ concentrations were 1.90, 2.15, and 1.98 times higher than those in the Semi

periods, suggesting a stronger oxidation capacity, with k_{OH} in Heavy being 26.43% and 9.56% higher than in Semi I and Semi II, respectively. Limited anthropogenic emissions in the suburban environment reduced the oxidation contribution by NO_x and CO (27.59%). During the heavy pollution, organic species exhibited dominant behavior regarding diurnal reactivity (9.22 s^{-1} for 69.79%), and anthropogenic hydrocarbons were not major k_{OH} sources. With an abundant level ($\sim 1\text{ ppb}$), isoprene contributed more than 10% of the reactivity in the diurnal cycle. Therefore, the effect of BVOCs species (such as monoterpenes, limonene, etc.) on radical chemistry cannot be ignored (Ma et al., 2022; Wang et al., 2022b). k_{OVOCs} are categorized into three groups: $k_{OVOCs(Obs)}$, $k_{OVOCs(Model)}$, and k_{HCHO} . Given the significance of formaldehyde photolysis, the contribution of HCHO to k_{OVOCs} is distinguished. $k_{OVOCs(Obs)}$ encompasses species observed in addition to formaldehyde, such as acetaldehyde (ACD) and the oxidation products of isoprene (MACR and MVK). Intermediates generated by the model, including glyoxal (GLY), methylglyoxal (MGLY), higher aldehydes (ALD), ketones (KET), methyl ethyl ketone (MEK), and methanol (MOH), are classified as $k_{OVOCs(Model)}$. Upon considering $k_{OVOCs(Model)}$, the reactivity calculated prior to September 10th aligns quite well with the observed OH reactivity.

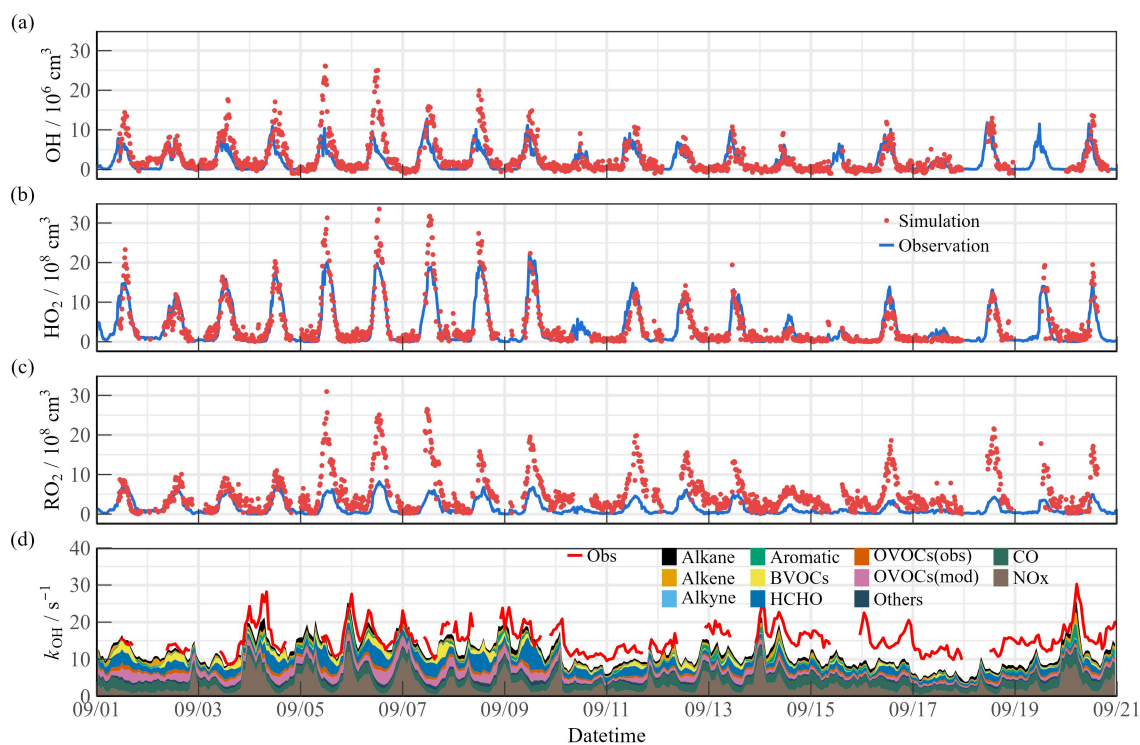


Fig. 4. Timeseries of the observed and modelled parameters for OH, HO₂ and *k*_{OH} during the observation period.
(a) OH, (b) HO₂, (c) *k*_{OH}.

The significant variations in oxidation can be inferred from the disparities during different pollution periods (Fig. S6). During Semi I, there was a good agreement between the measurement and model for peroxy radicals during the daytime. The RACM2-LIM1 mechanism effectively replicated the morning OH radical concentration. However, following 10:00, NO gradually declined, and the increasing OH concentration could not be accounted for by the HO₂+NO formation channel, resulting in a maximum underestimation of $5.85 \times 10^6 \text{ cm}^{-3}$ (Hofzumahaus et al., 2009; Lu et al., 2012). In the Semi II episode, OH was not underestimated in the low-NO regime, with a slight overestimation of HO₂ concentration. However, the simulated RO₂ concentration was only $3.78 \times 10^8 \text{ cm}^{-3}$, whereas observations were 2.77 times larger than the simulation, indicating the existence of additional reaction pathways that likely propagated the OH→RO₂ conversion efficiency. A significant discrepancy of radicals existed in the heavy ozone concentration, with OH, HO₂, and RO₂ radicals concurrently underestimated at noon by $8.23 \times 10^6 \text{ cm}^{-3}$, $3.94 \times 10^8 \text{ cm}^{-3}$ and $11.59 \times 10^8 \text{ cm}^{-3}$, respectively. The observed HO₂/RO₂ ratio approached 1:1, while the model reflected an unreasonable ratio of 3:1, indicating deficiencies in both primary sources and secondary propagation. The calculated reactivity seems to compare well with the observed OH reactivity at the start of the measurement period, but then there is evidence of missing OH reactivity after September 10th (Fig. 4(d)). Due to the limitations of available instruments, this observation only measured a limited number of OVOCs species, making it difficult to accurately quantify the contribution of larger aldehydes and ketones, carboxylic acids, nitrophenols, and other multifunctional species to *k*_{OH} (Wang et al., 2024). Since the MCM mechanism considers more secondary formation reactions than the RACM2 mechanism, it can qualitatively assess the photochemical role of unmeasured OVOCs species in the atmosphere (Wang et al., 2022d). The additional modeled OVOCs by the MCM v3.3.1 mechanism contributed $\sim 2.4 \text{ s}^{-1}$ to the missing OH reactivity (Fig. S7). During Heavy period, the reactivity of more model oxidation products increased the daytime *k*_{OH} by about 5.1 s^{-1} . Therefore, the observed *k*_{OH} can serve as an upper limit for sensitivity tests, thereby the full suite of radical measurement can be performed to explore the missing oxidation properties and ozone formation (Section 4.1).

Fig. 5 displays the diurnal profiles of the ROx budget during different episodes. In Semi I, formaldehyde photolysis showed a higher contribution (38.6%), while HONO photolysis (21.0%) and ozone photolysis (24.7%) accounted for similar proportions in primary sources. The contribution of photolysis from other OVOCs was comparable to that of ozonolysis reactions (7.2% vs. 4.8%). However, in Semi II, the decreased oxidation level was attributed to lower ROx sources, despite the similar proportions. During the Heavy period, the primary sources dramatically increased (up to ~10 ppb/h), with HCHO photolysis contributing the most, alongside other sources at common levels (ranging between 1.74 – 2.66 ppb/h) in the YRD region (Ma et al., 2022). Fast HCHO oxidation dominated the radical primary source during heavy ozone pollution, which contrasts with the dominant role of HONO/O₃ in other megacities (Yang et al., 2022; Tan et al., 2017b; Yang et al., 2021a).

The radical removal rate during the daytime was generally balanced with production contributions. In the morning, owing to high NO_x concentrations, radical termination was mainly dominated by OH+NO₂, OH+NO, RO₂+NO, and RO₂+NO₂. Furthermore, the formation of peroxy nitrate accounted for a certain proportion (~5%). As NO_x concentrations decreased after 10:00, self-reactions in peroxy radicals became significant.

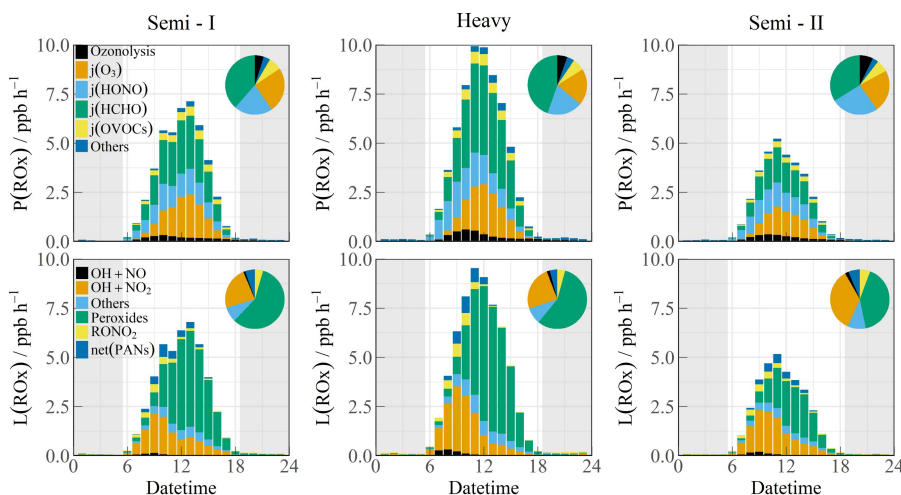


Fig. 5. The diurnal profiles of ROx budget during different polluted episodes (Semi I, Heavy, and Semi II). The pie chart denotes proportions in different parts during the daytime (10:00-15:00). The grey areas denote nighttime.

By comparing the known sources and sinks for radicals, unknown processes for initiation, transformation and termination can be determined in the experimental budget analysis (Fig. S8). During the Semi I period, the production and destruction rates of HO₂, RO₂, and total ROx radicals were very consistent, but a significant lack of a source term

for OH radicals was existed after 10:00. This missing source became more pronounced during the Heavy period, reaching 16 ppb/h at noon, which is close to the results observed by AIRPRO, but three times that observed by Heshan in PRD region(Tan et al., 2019b; Whalley et al., 2021). The ratio of OH production-to-destruction rate during the Semi II period was close to 1, indicating consistency between the observed results of OH, HO₂, k_{OH} , and other precursors(Whalley et al., 2018). However, the generation of HO₂ radicals in the morning was about twice as high as the removal rate, suggesting that there are contributions from unconsidered HO₂ radical removal channels (such as heterogeneous reactions)(Song et al., 2021). During the Heavy period, there was a rapid total removal rate of RO₂ radicals, reflecting the dominated HO₂ generation by the reaction of RO₂ radicals with NO. Although the P(HO₂) and D(HO₂) were quite in balance, the removal rate of RO₂ radicals far exceeded the known production rate (especially before 12:00). Previous work has shown that halogen chemistry (such as photolysis of nitryl chloride (ClNO₂)) could be an important source in the morning time, but this was not included in the calculation of RO_x or RO₂ budget in this campaign(Tan et al., 2017b). The steady-state analysis for HO₂ radical in the London campaign emphasized that only by significantly reducing the observed RO₂-to-HO₂ propagation rate to just 15% could balance both P(HO₂) and D(HO₂), indicating that the RO₂-related mechanism for propagation to other radical species may not be fully understood(Whalley et al., 2018). Therefore, based on the current knowledge seems unlikely to explain the required source-sink difference of nearly 25 ppb/h in the RO₂ budget. Sensitivity analysis is needed to further infer the causes of the difference for the experimental budget analysis.

3.3 Oxidation comparison

The concentration of OH radicals during the daytime is a crucial indicator of atmospheric oxidation levels (Liu et al., 2021). Table 2 summarized radicals and related parameters for regions with similar latitudes ($32.0^\circ \pm 2^\circ$ N, $j(O^1D) \approx 2.5 \pm 0.5 \times 10^{-5} \text{ s}^{-1}$). The joint influence of solar radiation and local photochemistry resulted in megacities exhibiting intense oxidation levels in summer/autumn, characterized by OH radicals being maintained at approximately $10.0 \times 10^6 \text{ cm}^{-3}$ at noon. Notably, an observation in Houston revealed an OH concentration of nearly $20.0 \times 10^6 \text{ cm}^{-3}$, with k_{OH} of 10 s^{-1} (Mao

et al., 2010). In areas such as Los Angeles, Pasadena, and Tokyo, the propagation efficiency of radicals was restricted due to fresh anthropogenic emissions. OH concentrations were only half of those observed in other megacities, with higher inorganic-dominated k_{OH} recorded (Pasadena, $\sim 20 \text{ s}^{-1}$) (George et al., 1999; Griffith et al., 2016; Yugo Kanaya et al., 2007). In the TROPSTECH observation, the observed k_{OH} exceeded the mean value at the same latitude ($>15 \text{ s}^{-1}$). Additionally, during the Heavy episode, higher OH concentration ($13.5 \times 10^6 \text{ cm}^{-3}$) was found, comparable to the highest level at regions with similar latitude (Houston 2000/2006, (Mao et al., 2010)). Synchronous elevation in radical concentration and reactivity indicated a strong oxidation level in the YRD region.

The observations in the YRD region showed a stable conversion factor ($OH-j(O^1D)$) of $4 \pm 1 \times 10^{11} \text{ cm}^{-3} \text{ s}$, which was comparable to other megacities in the PRD, NCP, and SCB regions (Ma et al., 2022; Tan et al., 2019a). The corresponding slope between OH concentration and solar radiation was used to quantify the oxidation efficiency from photolysis, and it was observed that a higher slope of $5.3 \times 10^{11} \text{ cm}^{-3} \text{ s}$ during the Heavy period indicated an active radical chemistry. This implies that there is a strong oxidation efficiency from photolysis in the YRD region.

During summer and autumn seasons, photochemical pollution is a common occurrence, as noted by (Tan et al., 2021). Analysis of radical concentration across different regions reveals that the YRD region exhibited concentrations higher than 10^7 cm^{-3} , slightly lower than in Guangzhou in 2006 but consistent with observations in other megacities (Ma et al., 2022; Tan et al., 2017a; Lu et al., 2012; Yang et al., 2021a). Conversely, winter is characterized by haze pollution (Ma et al., 2019). An urban site in Shanghai reported a peak OH concentration of $2.6 \times 10^6 \text{ cm}^{-3}$, closely resembling the $1.7 - 3.1 \times 10^6 \text{ cm}^{-3}$ range found in polluted winter atmospheres (Zhang et al., 2022a). Although no significant regional disparities in oxidation levels were detected in agglomerations, attention should be directed to the YRD region due to its elevated radical concentration, reactivity, and photolysis efficiency, signaling the need to investigate its role in radical chemistry.

Table 2. Summary of radical concentrations and related species concentrations at regions with similar latitude and megapolitan areas in China. All data are listed as the average in noontime (11:00–13:00).

Location	Latitude	Year	OH (10^6 cm^{-3})	k_{OH} (s^{-1})	$j(O^1D)$ (10^{-5} s^{-1})	Slope ($10^{11} \text{ cm}^{-3} \text{ s}$)	X (ppb)	References
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Regions with similar latitude								
Los Angeles	34.1° N	Sep 1993	6.0	-	-	-	-	(George et al., 1999)
Nashville	36.2° N	Jun–Jul 1999	10.0	10.2	3.0	3.3 ^c	-	(Martinez et al., 2003)
Houston	29.7° N	Aug 2000	20.0	9.0 ^b	3.0	6.7 ^c	-	(Mao et al., 2010)
Tokyo	35.6° N	Jul–Aug 2004	6.3	-	2.5	3.0	-	(Yugo Kanaya et al., 2007)
Houston	29.7° N	Sep 2006	15.0	11.0	3.1	5.0 ^c	-	(Mao et al., 2010)
Pasadena	34.1° N	May–Jun 2010	4.0	20.0	2.5	1.6 ^c	-	(Griffith et al., 2016)
Taizhou	32.6° N	May–Jun 2018	10.6	10.8 ^a	2.1	4.8	0.10	(Ma et al., 2022)
Chengdu	30.7° N	Aug 2019	10.0	8.0	2.2	4.1	0.25	(Yang et al., 2021a)
TROPSTECT (Heavy)	31.9° N	Sep 2020	13.5	16.0	2.6	5.3	0.50	This work
TROPSTECT (Semi)	31.9° N	Sep 2020	7.2	14.2	2.4	3.1	0.25	This work
Regions in megapolitan areas in China								
Guangzhou (PRD)	23.5° N	Jul 2006	12.6	17.9	3.5 ^b	4.5	0.85	(Lu et al., 2012)
Wangdu (NCP)	38.7° N	Jun–Jul 2014	8.3	15.0	1.8	4.5	0.10	(Tan et al., 2017b)
Beijing (NCP)	39.9° N	May–Jun 2017	9.0	30.0	2.4	3.8 ^c	~0	(Whalley et al., 2021)
Taizhou (YRD)	32.6° N	May–Jun 2018	10.6	10.8 ^a	2.1	4.8	0.10	(Ma et al., 2022)
Shenzhen (PRD)	22.6° N	Sep–Oct 2018	4.5	21.0	1.8	2.4	0.10	(Yang et al., 2022)
Chengdu (SCB)	30.7° N	Aug 2019	9.0	8.0	2.2	4.0	0.25	(Yang et al., 2021a)
Hefei (YRD)	31.9° N	Sep 2020	10.4	14.3	2.4	4.4	0.30	This work

473 ^a The modeled k_{OH} .

474 ^b Value only in the afternoon.

475 ^c Using the ratio of OH / j(O¹D)

476 4 Discussion

477 4.1 Measurement–model reconciliation for radicals

478 4.1.1 OH underestimation

479 Full suite of OH, HO₂, RO₂ and k_{OH} was utilized in the TROPSTECT campaign to
480 untangle a thorough understanding of oxidation mechanisms where base model failed.
481 One specific phenomenon was the absence of an OH source in situations where NO
482 levels gradually decreased after 10:00. A sensitivity test was conducted introducing a
483 species X, analogous to NO, to enhance OH regeneration (Fig. 6, RO₂→HO₂ and
484 HO₂→OH) (Hofzumahaus et al., 2009). It was found that the addition of as little as 0.25
485 ppb X was sufficient to compensate for the full magnitude of the OH underestimation in
486 the low NO region (Fig. 6). The employment of the X mechanism not only accelerated
487 OH regeneration but also augmented the removal channel of peroxy radicals, which
488 consequently led to a reduction in both HO₂ and RO₂ radical concentrations compared to
489 the base scenario.

The underdetermined radical sources in China were corresponding to the oxidation level (Ma et al., 2022; Tan et al., 2017a; Lu et al., 2012; Yang et al., 2021a; Wang et al., 2019b). The required X level typically ranged from 0.1 to 0.3 ppb, with the exception of the Backgarden observation which required 0.85 ppb X, as indicated in Table 2 (Lu et al., 2012). A minimum limit of 0.1 ppb X was established to account for any missing reactivity (Ma et al., 2022). Notably, throughout the entire observation, a strong agreement between the modeled and observed OH was achieved when a mixture of 0.25 ppb X was incorporated into the base scenario, consistent with the the order of magnitude in Chengdu (Yang et al., 2021a). During the Heavy period, the augmented photochemistry resulted in complex oxidation, necessitating an additional missing OH source equivalent to 0.5 ppb X to fully address the underestimation of OH.

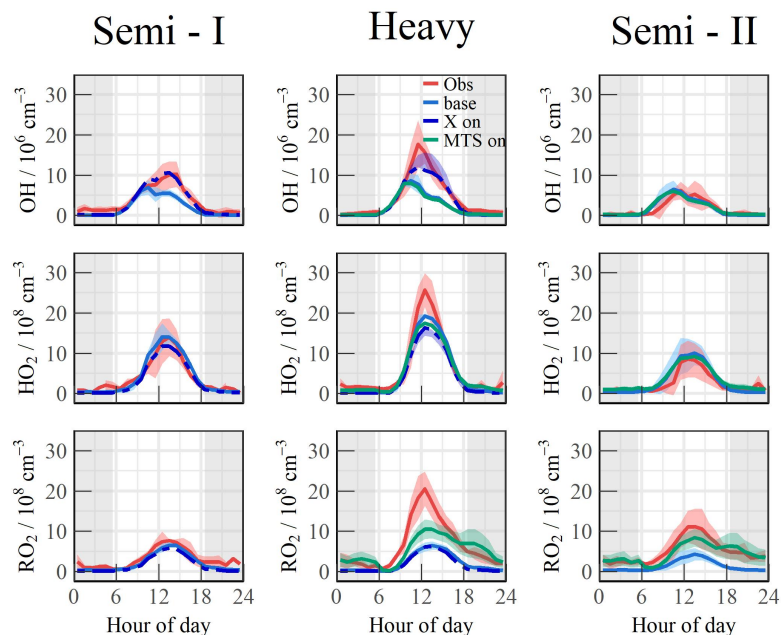


Fig. 6. The mean diurnal profiles of measured and modeled OH, HO₂ and RO₂ concentrations at different scenarios. Sensitivity tests included three scenarios (Scenario 1: base case; Scenario 2: X mechanism on. The dashed line represented the performance of 0.25 ppb X introduced in the Semi I and Heavy episodes, and the blue shadow denoted the upper limit for X influence (0.5 ppb); Scenario 3: monoterpene mechanism on; Both API and LIM were added into the base model as upper and lower limits for the influence of monoterpenes, and the mean of the two values was taken as the average effect. The grey areas denote nighttime.

Missing OH sources are closely related to the chemistry of OVOCs (Yang et al., 2024a; Qu et al., 2021). Reactive aldehyde chemistry, particularly the autoxidation of carbonyl organic peroxy radicals (R(CO)O₂) derived from higher aldehydes, is a significant OH regeneration mechanism that has been shown to contribute importantly to OH sources in regions with abundant natural and anthropogenic emissions during warm

seasons (Yang et al., 2024b). In this study, the higher aldehyde mechanism (HAM) by Yang et al was parameterized into the base model to test new insights into the potential missing radical chemistry (Fig. 7). The results indicate that the contribution of the HAM mechanism to OH radicals in different episodes ranged between 4.4% - 6.0%, while the concentrations of HO₂ and RO₂ radicals increased by approximately 7.4% and 12.5%, respectively.

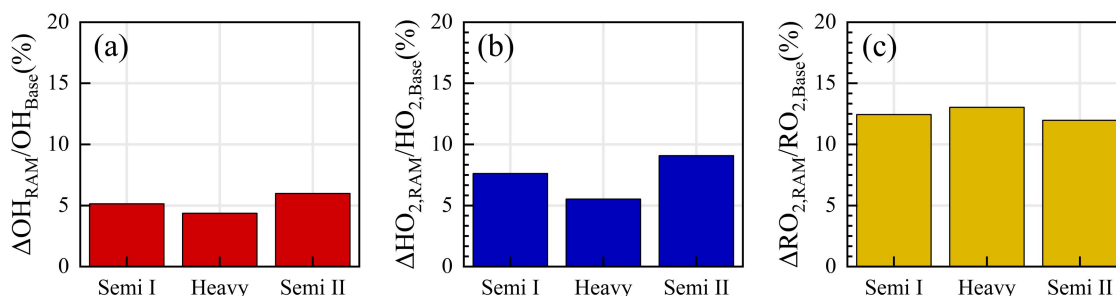


Fig. 7. The response of (a) OH, (b) HO₂ and (c) RO₂ radicals to the Higher Aldehyde Mechanism (HAM) in different episodes (Semi I, Heavy, and Semi II) in diurnal time (10:00-15:00).

4.1.2 RO₂ underestimation

The base scenario in Semi II is capable of accurately reproducing the concentrations of OH and HO₂ radicals within the data uncertainty. However, the simulated RO₂ concentration by the base model is only $3.78 \times 10^8 \text{ cm}^{-3}$, which does not align with the observed oxidation levels in YRD, indicating a clear discrepancy. This underestimation is similarly evident in the APHH observation in Beijing, as the highest observed concentration of RO₂ radicals reached $5.5 \times 10^9 \text{ cm}^{-3}$, far exceeding the level predicted by the MCM v3.3.1 mechanism (Whalley et al., 2021). The failure to reproduce the RO₂ concentration reflects the inadequacy of the mechanisms related to RO₂ radicals due to diverse oxidation reactions. This issue is further elucidated by previous studies, which highlighted the possibility of certain VOCs undergoing more intricate isomerization or fragmentation steps to sustain the long lifetime of RO₂ radicals (Whalley et al., 2018; Whalley et al., 2021).

The union of k_{OH} and RO₂ measurement can help reveal the magnitude of missing RO₂ as a hypothesis of sensitivity analysis. An additional reaction was added to the base model in a previous research, converting OH into C₉H₁₆O₂ (the oxidation product of α -pinene) with a reaction rate equal to the missing reactivity, to explore the source of the missing RO₂ radicals (Whalley et al., 2021). Discrepancy of OH reactivity ($\sim 3 - 5 \text{ s}^{-1}$)

between measurement and model suggested that an additional driving force was necessary to complete the OH to RO₂ step. In the TROPSPECT campaign, approximately 0.4 ppb of monoterpene was introduced into the base scenario as the chemical reactions of complex alkoxy radicals, which is similar to an atmospheric level in the EXPLORE-2018 campaign, the YRD region (Wang et al., 2022b). The RACM2 mechanism identified α -pinene (API) and limonene (LIM) as representative monoterpene species. Sensitivity tests were conducted by incorporating API and LIM into the 'MTS on' and 'MTS+X on' scenarios, respectively (Ma et al., 2022). The mean of these values was considered the average effect of monoterpene chemistry, and depicted as the green line in Fig. 6. In the 'MTS on' scenario, the chemistry of peroxy radicals in Semi II was reasonably described by introducing the source of complex alkoxy radicals, and the obs-to-mod ratio of peroxy radicals decreased from 2.2 to 1.3. Furthermore, the introduction of additional complex alkoxy radicals had minimal impact on HOx chemistry, with changes in daytime OH and HO₂ concentrations of less than $5 \times 10^5 \text{ cm}^{-3}$ and $2.5 \times 10^7 \text{ cm}^{-3}$, respectively. This demonstrates the robustness of HOx radical in response to potential monoterpene.

Higher aldehyde chemistry is a concrete manifestation of verifying the aforementioned hypothesis for RO₂ sources (Yang et al., 2024b). The autoxidation process of R(CO)O₂, encompasses a hydrogen migration process that transforms it into the $\cdot\text{OOR}(\text{CO})\text{OOH}$ radical (Wang et al., 2019b). This radical subsequently reacts with NO to yield the $\cdot\text{OR}(\text{CO})\text{OOH}$ radical. The $\cdot\text{OR}(\text{CO})\text{OOH}$ radical predominantly undergoes two successive rapid hydrogen migration reactions, ultimately resulting in the formation of HO₂ radicals and hydroperoxy carbonyl (HPC). Consequently, the HAM mechanism extends the lifetime of the RO₂ radical, providing a valuable complement to the unaccounted sources of RO₂ radicals. As depicted in Fig. 7, the incorporation of the HAM mechanism results in an approximate 7.4% and 12.5% increase in the concentrations of HO₂ and RO₂ radicals, respectively. It is important to note that the total concentrations of primary emitted aldehydes and the HPC group may be underestimated, which could lead to the aforementioned analysis being conservative in nature. Further exploration of the unaccounted sources of RO₂ radicals will be presented in Section 4.3.

4.1.3 P(Ox) underestimation

Upon completing the hypothetical investigation into the radical underestimation, a sensitivity comparison between observed and modeled $P(O_x)$ was conducted across the entire range of NO concentrations, as depicted in Fig. 8(a)(b). With increasing NO concentration, the overall $P(O_x)$ amplified, reaching a maximum of approximately 30 ppb/h. This variation has been validated through multiple observations in Wangdu, APHH, and other studies (Tan et al., 2017b; Whalley et al., 2021; Whalley et al., 2018). However, the imperfect understanding of the mechanisms related to peroxy radicals ultimately leads to misjudgment of the ozone production process in high NO regimes, with a degree of underestimation close to 10 times, as illustrated in Fig. 8(b).

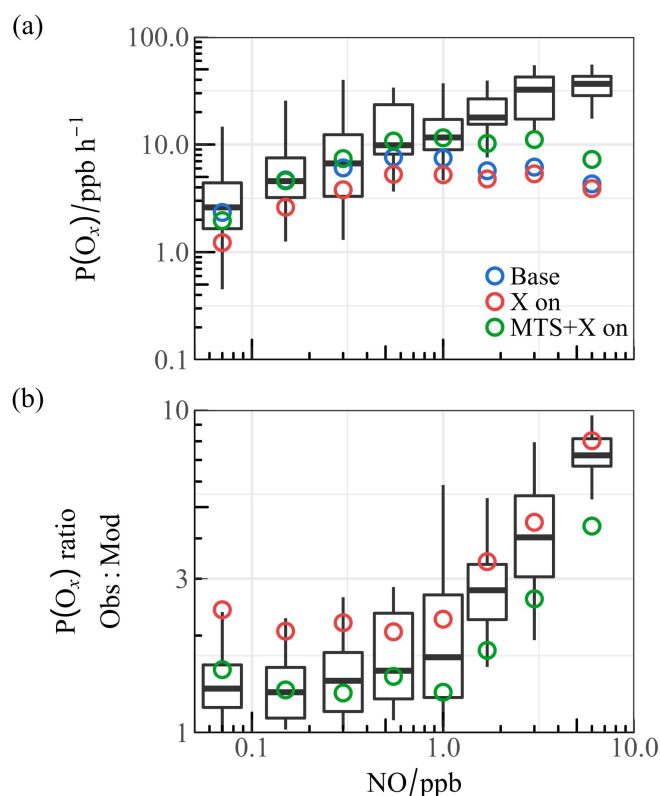


Fig. 8. The relationship between NO and (a) $P(O_x)$, (b) $P(O_x)$ (Obs:Mod). Boxplot diagrams are used to illustrate the minimum, 25th percentile, median, 75th percentile, and maximum values of the observed dataset. The circles represent the median values for the base model as well as for different mechanisms added to the model within various ranges.

Although the inclusion of the X mechanism improves the agreement between simulated and observed OH concentrations in the low-NO range, it has a negative effect on the $P(O_x)$ simulation. The introduction of a major source of RO_2 can help address the underestimation problem in the base scenario, as the lack of RO_2 species and related reaction rates is an important factor leading to deviations in the simulation of ozone

production rates (Tan et al., 2017a). The combination of the X mechanism and monoterpene chemistry is shown to better harmonize the relationship between HO₂ and RO₂. Notably, the deficiency in the ozone generation mechanism was adequately explained within a certain range in the 'MTS+X On' scenario, leading to an enhancement in the simulation performance of P(Ox) in the high NO_x region (Fig. 8(b)). Therefore, reasonable simulation of the concentration of peroxy radicals is key to accurately quantifying the process of ozone generation.

4.2 Effect of mechanism reconciliation on oxidation

Both radical concentration and oxidation coordinating deficiency are worthy of examine (Fig. S9). To eliminate the influence of non-photolytic processes, only the daytime concentration range with $j(\text{O}^1\text{D})$ greater than $5 \times 10^{-6} \text{ s}^{-1}$ was selected. The boxplots illustrate the ratio of observation to simulation (base model), with the circles representing the average values after integrating different mechanisms into the base scenario. In the low NO regime ($\text{NO} < 1 \text{ ppb}$), the OH underestimation was consistently prominent as NO concentration decreased, and the base model was able to reasonably reflect the HO₂ distribution contrastly. As NO levels increased, the simulated OH concentration aligned well with the observation, but both HO₂ and RO₂ concentrations exhibited underprediction. RO₂ underestimation extended across the entire NO range, and could rise to over 10 times when NO levels reached about 10 ppb. Sensitivity tests based on the full suite of radical measurement revealed that the X mechanism accelerated OH regeneration, and the introduction of larger RO₂ alleviated the absence of certain sources by 2 to 4 times.

The coordinate ratios of radical serves as another test for RO_x propagation (Fig. 9). The observed HO₂/OH ratio is approximately 100, declining to some extent as the concentration of NO increases, which is consistent with previous studies (Griffith et al., 2016; Griffith et al., 2013). However, the base model does not accurately replicate the curve depicting the change in HO₂/OH ratio, as shown in Fig. 9 (a). At low NO levels, the ratio significantly overestimated and shows a steeper decline compared to the base scenario as NO levels increase. Furthermore, the observed RO₂/OH ratios remain around 100, whereas the predicted values are significantly underestimated when NO exceeds 1 ppb (refer to Fig. 9(b)). In terms of the observed HO₂/RO₂ ratio, it maintains a relatively

constant trend within the range of 0.5 – 1.5, while the model overestimated by more than twice, highlighting an inconsistency between the conversion of $\text{RO}_2 \rightarrow \text{HO}_2$. The incorporation of the X mechanism has proven to be effective in a balanced HO_2/OH ratio as illustrated in Fig.9(a), but amplifying the termination pathway for HO_2 and RO_2 , which altered the coordination between RO_2 and OH across the entire NO range (Fig. 9(b)). The connection between unconditional OH source and larger RO_2 isomerization in chemically complex environments is key to fully understanding tropospheric chemistry, and a better coordination of HO_2/OH , RO_2/OH , and HO_2/RO_2 ratios are established by incorporating additional mechanisms.

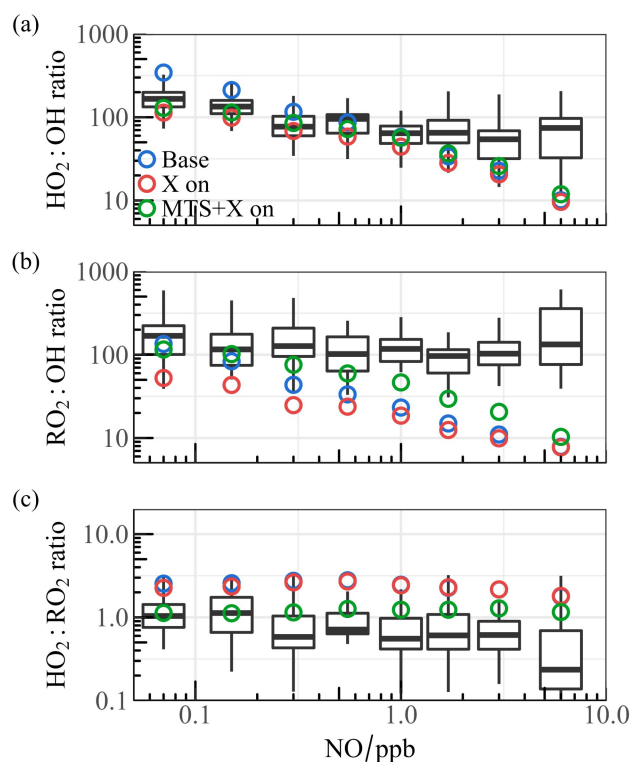


Fig. 9. The ratios for (a) HO_2/OH , (b) RO_2/OH , and (c) HO_2/RO_2 show a correlation with NO levels. Boxplot diagrams are used to illustrate the minimum, 25th percentile, median, 75th percentile, and maximum values of the observed dataset. The circles represent the median values for the base model as well as for different mechanisms added to the model within various ranges.

The HO_2/RO_2 parameter was utilized to explore the transformation relationship between HO_2 and RO_2 radicals. If HO_2 is formed from an RO_2 radical, it would result in an HO_2/RO_2 radical concentration ratio of approximately 1. The HO_2/RO_2 ratios derived from radical concentrations measured by laser-induced fluorescence instruments and calculated using the MCM or RACM models were summarized in Fig. 10. In field studies, the observed HO_2/RO_2 ratios were between 0.2 - 1.7 under low- NO conditions ($\text{NO} < 1$

ppb) and only 0.1 - 0.8 under high-NO conditions ($3 < \text{NO} < 6$ ppb). From the perspective of model-observation matching, except for three measurements in ClearfLo, ICOZA and AIRPRO-summer campaigns, the HO_2/RO_2 ratios in other regions could be reasonably reflected by the MCM or RACM2 mechanisms (Woodward-Massey et al., 2023; Whalley et al., 2021; Whalley et al., 2018; Färber et al., 2024). However, the ratio is generally underestimated under high NO conditions, reaching up to 5 times in ClearfLo. According to the latest chamber experiments, the HO_2/RO_2 radical concentration ratios for VOCs forming HO_2 are 0.6 for both one-step and two-step reactions. Therefore, the extremely low HO_2/RO_2 ratios observed in field campaigns can only be explained if almost all RO_2 radicals undergo multiple-step reactions before forming HO_2 . During the TROPSTECT campaign, the observed HO_2/RO_2 remains at 1.1 and 0.8 under low-NO and high-NO conditions, respectively. After considering the complex sources of complex alkoxy radicals in the 'MTS+X' scenario, the simulated values of HO_2/RO_2 in both low-NO and high-NO regions match the observed values well.

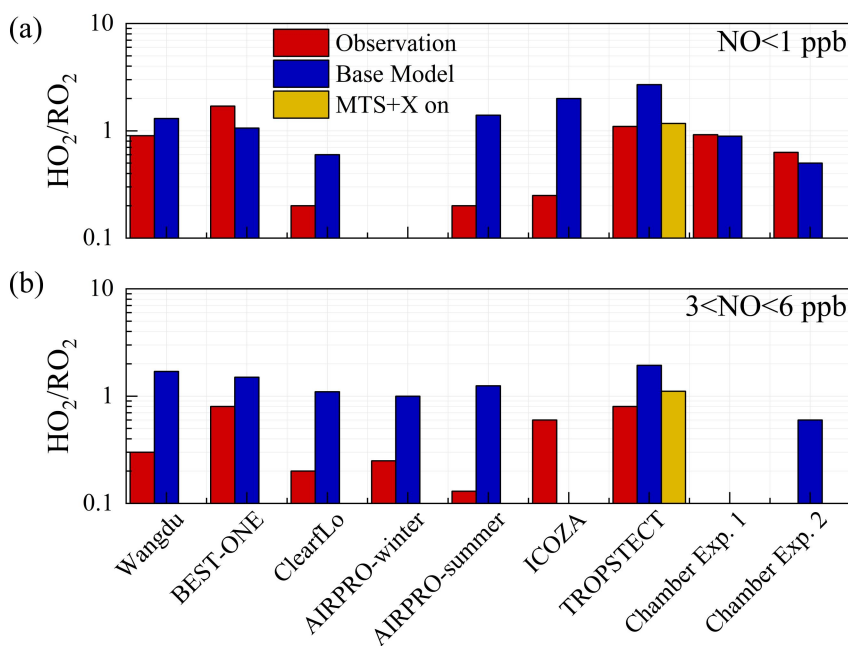


Fig. 10. Summary of the HO_2/RO_2 ratios derived from radical concentrations measured by laser-induced fluorescence instruments and calculated using the MCM or RACM models under (a) low-NO and (b) high-NO conditions. Chamber Exp. 1 and Chamber Exp. 2 denotes the parameters by single-step HO_2 formation and multi-step HO_2 formation determined in the chamber by (Färber et al., 2024).

4.3 Missing OVOCs sources influence ozone production

The consistency between model predictions and observed measurements for ozone

production, akin to the concentration ratio of HO_2/RO_2 , is depicted in Fig. 11(a)(b). In areas with low NO levels, the ratio of modeled to actual ozone production ranges from 0.5 to 2, with the exception of the ClearfLo and AIRPRO-summer datasets (Woodward-Massey et al., 2023; Whalley et al., 2021). Conversely, under high NO conditions (with NO concentrations between 3 and 6 ppbv), the ozone production rate ($\text{P}(\text{Ox})$) derived from measured radical concentrations typically exceeds that of the base model's predictions by more than threefold. Laboratory experiments focusing on the oxidation of representative VOCs suggest that ozone production can be enhanced by approximately 25% for the anthropogenic VOCs under investigation (Färber et al., 2024). The MTS+X scenario represents an effort to enhance the congruence between modeled and measured radical concentrations. The incorporation of OVOCs and larger alkoxy radicals derived from monoterpenes has refined the model-measurement agreement for ozone formation under high NO conditions, reducing the discrepancy from 4.17 to 2.33. This substantiates the hypothesis of sensitivity analysis concerning ozone generation, as detailed in Section 4.2 and illustrated in Fig. S10.

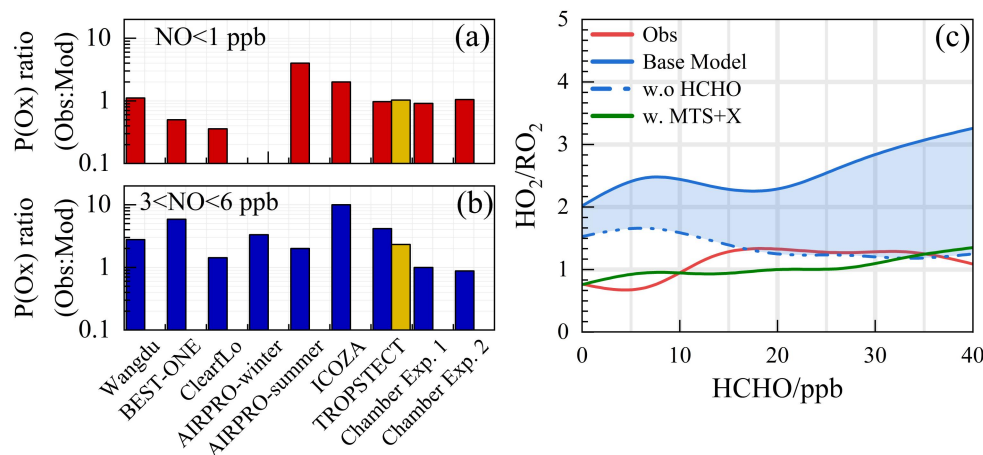


Fig. 11. Summary of the $\text{P}(\text{Ox})_{\text{Obs}}/\text{P}(\text{Ox})_{\text{Mod}}$ under (a) low-NO and (b) high-NO conditions. (c) The ratios for HO_2/RO_2 show a correlation with HCHO levels. The blue shading represents the range of variation from constrained to unconstrained formaldehyde conditions. Chamber Exp. 1 and Chamber Exp. 2 denotes the parameters by single-step HO_2 formation and multi-step HO_2 formation determined in the chamber by (Färber et al., 2024).

The reasons for the discrepancy between simulated and observed values for ozone production deserve further investigation. As depicted in Fig. 11(c), the simulated HO_2/RO_2 ratios display a robust positive correlation with photochemical activity, fluctuating between 2 and 4. A notable feature during severe ozone pollution is the intense distribution of formaldehyde, with an average concentration of 21.81 ± 4.57 ppb

(11:00 – 13:00). While formaldehyde acts as a precursor for HO₂ radicals, it does not directly generate RO₂ radicals. The contributions of OVOCs to the RO_x radical do not exhibit the same intensity as formaldehyde, and the current mechanism encounters difficulties in replicating formaldehyde concentrations (Fig. S11). The simulation of formaldehyde concentrations using the MCM v3.3.1 mechanism has shown improvement, indicating that the secondary formation of unmeasured species, such as OVOCs, will feedback on RO₂ radical levels. When formaldehyde levels are unconstrained, the simulated HO₂/RO₂ ratios align with observations, suggesting that under the prevailing chemical mechanism, the photochemical efficiency of formaldehyde and other OVOCs is similar. Therefore, an empirical hypothesis is proposed to amplify the concentration of higher-order aldehydes by a factor of about 4, which is the proportion of formaldehyde concentration underestimated by the model. The qualitative assessment of the impact of missing aldehyde primary emissions on RO₂ radical concentrations was combined with the HAM mechanism across the entire photochemical spectrum (Fig.S12). Enhanced impact of aldehyde autoxidation in the presence of weak photochemical conditions could alter the simulated levels of OH and HO₂ radicals by approximately 13.9% and 18.1%, respectively. However, higher ALD concentrations will be achieved under intensive photochemical conditions, leading to the gradual dominance of the sink channels for OH + OVOCs, with the effect of autoxidation mechanisms gradually decreasing. RO₂ radical concentrations is notably more sensitive to the HAM mechanism, where incorporates additional OVOCs, can enhance the simulation of RO₂ radical concentrations by 20 - 40%. Consequently, although limiting formaldehyde can partially offset the HO₂ radical cycle and enhance the precision of HO_x radical chemistry studies, additional measurements should be undertaken for other OVOCs, coupled with the deployment of full-chain radical detection systems, to accurately elucidate the oxidation processes under severe ozone pollution conditions.

5 Conclusion

The full suite radical measurement of OH, HO₂, RO₂ and k_{OH} was first deployed in the YRD region (TROPSTECT) and encountered with a prolonged ozone pollution in September 2020. The diurnal peaks of radicals exhibited considerable variation due to

environmental factors, showing ranges of 3.6 to $27.1 \times 10^6 \text{ cm}^{-3}$ for OH, 2.1 to $33.2 \times 10^8 \text{ cm}^{-3}$ for HO₂, and 4.9 to $30.5 \times 10^8 \text{ cm}^{-3}$ for RO₂. Continuous k_{OH} data fell within a range of $8.6 - 30.2 \text{ s}^{-1}$, demonstrating the dominant behavior of organic species in diurnal reactivity. Furthermore, observations in the YRD region were found to be similar to those in other megacities, suggesting no significant regional differences in oxidation levels were observed in agglomerations overall.

At a heavy ozone pollution episode, the oxidation level reached intensive compared with other sites, and the simulated OH, HO₂, and RO₂ radicals provided by the RACM2-LIM1 mechanism failed to adequately match the observed data both in radical concentration and experimental radical budget. Sensitivity tests based on the full suite of radical measurement revealed that the X mechanism accelerated OH regeneration, and the introduction of larger alkoxy radicals alleviated the RO₂-related imbalance. The HAM mechanism effectively complements the non-traditional regeneration of OH radicals, improving by 4.4% - 6.0% compared to the base scenario, while the concentrations of HO₂ and RO₂ radicals increased by approximately 7.4% and 12.5%, respectively. The incorporation of complex processes enabled better coordination of HO₂/OH, RO₂/OH, and HO₂/RO₂ ratios, and adequately addressed the deficiency in the ozone generation mechanism within a certain range. Incorporation of OVOCs and larger alkoxy radicals derived from monoterpenes improved the measurement-model consistency for ozone formation under high NO conditions, reducing the discrepancy from 4.17 to 2.33, which corroborates the hypothesis of sensitivity analysis in the context of ozone generation. This study enabled a deeper understanding of the tropospheric radical chemistry at play. Notably,

- ✓ A full suite of radical measurement can untangle the gap-bridge for the base model in more chemically-complex environments as an hypothesis of sensitivity tests.
- ✓ Additional measurements targeting more OVOCs should also be conducted to fulfill the RO₂-related imbalance, and then accurately elucidating the oxidation under severe ozone pollution.

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Data availability

The data used in this study are available upon request (rzhu@aiofm.ac.cn).

Author contributions

WQ Liu, PH Xie, RZ Hu contributed to the conception of this study. RZ Hu and GX Zhang performed the data analyses and manuscript writing. All authors contributed to measurements, discussed results, and commented on the paper.

Competing interests

The contact author has declared that none of the authors has any competing interests.

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