



1 **A conservative resource use strategy in agricultural grasslands counteracts lower productivity**
2 **and water use efficiency under drought conditions**

3 Elena Tello-García^{1*}, Erich Tasser^{1,2}, Ursula Peintner³, Ulrike Tappeiner¹, Georg Leitinger¹

4 ¹ Department of Ecology, Universität Innsbruck, 6020 Innsbruck, Austria

5 ² Eurac research, Institute for Alpine Environment, 39100 Bozen, Italy

6 ³ Department of Microbiology, Universität Innsbruck, 6020 Innsbruck, Austria

7 **Correspondence to:* Elena Tello-García (elena.tello-garcia@uibk.ac.at)

8 Erich Tasser: Erich.Tasser@uibk.ac.at

9 Ursula Peintner: Ursula.Peintner@uibk.ac.at

10 Ulrike Tappeiner: Ulrike.Tappeiner@uibk.ac.at

11 Georg Leitinger: Georg.Leitinger@uibk.ac.at

12



13 Abstract

14 Grassland response to changes in water availability is closely tied to the traits of the plant community
15 which determine plant water uptake and resource use. Plants can adopt either moderate and efficient
16 (conservative) or rapid and demanding (acquisitive) resource use strategies. These strategies combined
17 with the plant interactions with microbes, such as arbuscular mycorrhiza fungi (AMF), determine the
18 grassland productivity and efficiency. This study aims to compare the drought response of two
19 agricultural grasslands that differ in their resource use strategies. In an experimental garden 12 small-
20 scale lysimeters were installed with two different agricultural grassland types (conservative and
21 acquisitive) and two different irrigation levels (wet and drought). We measured water fluxes, above- and
22 belowground phytomass productivity and AMF spore productivity in these two grassland types
23 throughout a drought of 64 days. Despite differing resource use strategies, both grassland types exhibited
24 similar reductions in evapotranspiration and aboveground phytomass under drought. However, the
25 conservative grassland showed higher water use efficiency (WUE) when considering only aboveground
26 phytomass, and a less pronounced compositional shift towards greater grass phytomass. Furthermore,
27 in acquisitive grasslands the root:shoot ratio of grasses and AMF spores abundance in the soil were
28 greatly reduced than in conservative grasslands. We also identified differences in legume productivity,
29 rooting system, and AMF spores community composition as key factors influencing WUE. In a changing
30 climate with greater frequency and severity of droughts in the European Alps, opting for grassland
31 mixtures with more species with conservative characteristics should be considered, as they i) reach the
32 productivity of acquisitive grassland even under wetter conditions and ii) show higher efficiency and
33 longer vitality under drought conditions.

34 1. Introduction

35 The presence and abundance of functional groups in grasslands have a significant influence on the
36 quantity and quality of agricultural yield. Forb species increase the mineral content of forage and
37 improve the palatability and digestibility of fodder (Evers, 2011). Legumes are particularly important
38 as protein carriers and for nitrogen fixation from the air. Legumes improve the nitrogen supply of the
39 entire canopy through infection of roots by effective *Rhizobium* strains, which causes the formation of
40 functional root nodules (Reich et al., 2003). Farmers directly or indirectly determine the composition of
41 plant species through management measures to build up economically and agro-production optimized
42 grassland. Ideally, these should have a canopy with an optimal plant composition ratio of 60 to 70 %
43 grasses and 30 to 40 % legumes and other forbs (Potsch, 1999; Sanderson et al., 2013; Knot et al., 2015).
44 Such mixtures are also best suited to deliver high yields not only now, but also in a warmer climate of
45 the future (Elgersma and Sørensen, 2016). Thus, the need for functionally optimized grasslands and to
46 improve our knowledge of the reaction of species and functional groups to stresses like drought.
47 Functional diversity may not only determine these responses, but also may itself be affected by drought
48 (Miller et al., 2019; Tello-García et al., 2020) through changes in plant species composition (Griffin-



49 Nolan et al., 2019), and simplification of communities, which may lead to species reduction (Peralta et
50 al., 2019).

51 According to the individual traits involved in resource use, plants can range from fast/acquisitive to
52 slow/conservative strategies (Diaz et al., 2004; Reich, 2014; Wright et al., 2004). Acquisitive species
53 construct energetically cheap but short-lived tissues optimized for fast resource acquisition, whereas
54 conservative species produce more resource-expensive but persistent tissues with a slower rate of return.
55 Species following an acquisitive strategy are usually characterised by high specific leaf area (SLA), low
56 leaf dry-matter content (Wright et al., 2004) and long, thin roots (Roumet et al., 2016; Reich, 2014).
57 Fine roots make rapid use of resources with minimal biomass investment, resulting in high metabolic
58 and growth rates under favourable conditions (Ostonen et al., 2007; Reich, 2014; Bergmann et al., 2020).
59 Overall, acquisitive species have a high photosynthetic rate, but also high water consumption (Manzoni
60 et al., 2013). Acquisitive traits may be detrimental under extreme drought conditions as such species are
61 insensitive to progressive soil drying until very low soil water content is reached (e.g. the permanent
62 wilting point) (Brilli et al., 2011). Conversely, conservative species typically have smaller, thicker leaves
63 and shorter, coarser roots (Roumet et al., 2016; Tardy et al., 2017). Conservative species reduce water
64 uptake earlier under drought, economise nutrients and use them slowly, which results in low growth
65 rates under favourable conditions. Conservative and acquisitive strategies mark the ends of the spectrum,
66 a wide range exists between them, defined by the specific traits of a species and its plasticity (Perez-
67 Ramos et al., 2019).

68 Belowground components (e.g. root biomass and characteristics) have a significant impact on plant
69 resources uptake, as they are in contact with the soil and are one of the main pathways for all resources
70 (Ma et al., 2018; McCormack and Iversen, 2019). Plant roots and associated rhizosphere organisms
71 influence many important ecosystem processes, including nutrient and carbon storage in the soil, soil
72 structural stability (e.g. erosion resistance and porosity (Bardgett et al., 2014)), water uptake (Gross et
73 al., 2008), and interactions with other soil organisms. In this context, plants response to drought not only
74 on their functional traits and position on the conservative-acquisitive spectrum but also on their
75 associated rhizosphere. Belowground traits are intrinsically connected to the way plants benefit from
76 soil microbes. Therefore, a deeper knowledge of root-associated mutualists, such as arbuscular
77 mycorrhiza fungi (AMF), may partially account for variation along the spectrum (Bergmann et al.,
78 2020).

79 AMF are obligate symbionts with plant roots meaning they obtain most of their required organic carbon
80 from the host plant in return for inorganic N, P and water (Parniske, 2008). Moreover, AMF can alter
81 plant traits, such as stomatal conductance, triggering mechanisms to protect the plant from drought
82 (Augé, 2001). They improve plant growth even when root colonisation is reduced (Bahmani et al., 2018;
83 Hartman and Tringe, 2019; Tuo et al., 2017; Zhang et al., 2018b). Decrease on AMF productivity
84 includes reduced extraradical mycelial network and reproduction (i.e. sporulation) (Bahadur et al., 2019;



85 Hashem et al., 2019; Kilpeläinen et al., 2017; Maitra et al., 2019). Interestingly, AMF spore abundance
86 is primarily influenced by seasonal climate and host-independent physico-chemical soil properties, and
87 less so by the composition of host plants, as reported for Chilean Mediterranean-type ecosystems (Silva-
88 Flores et al., 2019). Although AMF spore abundances may not be an accurate indicator for mycorrhizal
89 activity, they can be useful as a proxy for stress. Under stress, the photobionts must reduce the amounts
90 of nutrients provided to their mycorrhizal partner, thus limiting the seasonally induced AMF spore
91 formation. Therefore, when comparing plant drought strategies in the same soil, we expect AMF spore
92 abundance to decrease proportionally to the stress a plant community is experiencing.

93 The community-weighted mean (CWM) of the individual plant traits and plant interactions (i.e. with
94 AMF) determine the strategy followed by the community under drought, and therefore the productivity
95 and water use efficiency (WUE). Acquisitive plant species that can use water and nutrients quickly but
96 wastefully during short wet periods are better prepared to cope with irregular rainfall (Querejeta et al.,
97 2018). These traits may be advantageous for successful colonization in dry regions with extreme
98 fluctuations in resource availability. However, in humid to semi-arid climates or in scenarios with brief
99 but intense dry periods this strategy breaks down. The economical use of water of conservative species
100 through dynamic control of water uptake via stomatal behaviour is more advantageous in the long term
101 (Zhao et al., 2016; Lu et al., 2020; Brodribb et al., 2020). Therefore, the traits of the plant communities
102 used in farming must be in accordance with the climatic conditions, as they will determine the tolerance
103 of crops to drought and their productivity.

104 This study aimed to (i) compare the effects of drought on two grassland seed mixtures commonly sown
105 in wet and dry sites in mountain farming in the European Alps, and to (ii) identify the main parameters
106 affecting water use efficiency of these agricultural grasslands. The seed mixtures studied differed
107 primarily on the resource use strategies of the plant community (i.e. more conservative or acquisitive).
108 In addition, we wanted to understand any impacts on agricultural yields under dry conditions. To analyse
109 the mechanisms behind the expected differences, we focused on the water fluxes, functional
110 composition, phytomass productivity, and AMF spore productivity of these two grassland types in a
111 drought experiment. We hypothesized that the agricultural grassland with more traits related to
112 conservative strategies (hereafter referred to as conservative grassland) would show a faster response to
113 drought. Expected responses included closing of stomata and higher investment in root and AMF
114 sporulation. We expected that these responses would result in a higher WUE under drought conditions.
115 In contrast, we expected that the grassland with more traits related to acquisitive strategies (hereafter
116 referred to as acquisitive grassland) would be less responsive to drought. We expected plants to use the
117 available water under drought more quickly and then have less time to produce phytomass and AMF
118 biomass before wilting, resulting in a lower WUE.

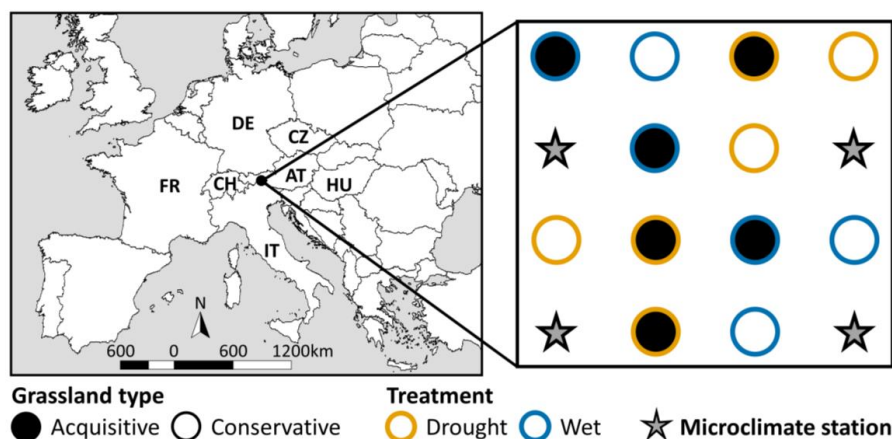
119



120 2. Material and Methods

121 The experimental garden was established in the long-term socio-ecological research (LTSER) site
122 “Stubai valley” (LTSER platform “Tyrolean Alps”), close to Neustift im Stubaital (Austria), located on
123 the valley floor in the montane belt at 972 m above the sea level (WGS84: 47.115833 N, 11.320556 E)
124 in a meadow used for hay production (Fig. 1). It has an annual mean temperature of 6.5 °C and annual
125 mean precipitation of 1097 mm (Frenck et al., 2018). There, 12 small-scale lysimeters were installed
126 with two different agricultural grassland types and two different irrigation levels. Investigations took
127 place in 2018 during part of the growing season (June - August).

128



129

130 **Figure 1. Schematic overview of the experimental design. Three replicates were used for each combination**
131 **of grassland type and irrigation level. © EuroGeographics for the administrative boundaries.**

132

133 2.1. Investigated grassland types: community composition and plant traits

134 We linked this investigation to applied grassland farming using two seed mixtures commonly sown in
135 European Alpine grassland farming (Schwarzenberger Samen & Gartenbedarf, Austria). Both seed
136 mixtures were sown directly in 12 small-scale lysimeters (Fig. 1). The highly productive ‘seed mixture
137 for humid locations (SR037)’ is typically sown in more humid areas of the Central and Northern Alps
138 (> 800 mm annual precipitation), and the ‘seed mixture for dry locations (SR032)’ is sown in the drier
139 areas of the Southern Alps (< 800 mm annual precipitation). The two seed mixtures were characterized
140 as acquisitive and conservative, respectively, based on their traits’ CWM being closer to one or another
141 end of the resource use spectrum. For all species present in the mixtures (Table 1), main traits for
142 resource use classification were searched in the literature (only data on traits from plants growing in
143 favourable conditions were selected, and a mean was calculated when more than one paper or value was



144 found): root-shoot ratio, shoot C:N ratio, SLA, leaf N concentration (LNC) and rooting index (Wright
145 et al., 2004; Roumet et al., 2016; Diaz et al., 2016; Bu et al., 2018; Chacón-Madrigo et al., 2018; Reich,
146 2014). We then calculated a CWM proxy for each trait of the two seed mixtures (Table 1). Due to
147 difficulties in differentiating species in their vegetative state, we calculated the CWM proxy based on
148 the seed abundance in the mixture rather than on actual biomass or cover. The highly productive ‘seed
149 mixture for humid locations (SR037)’ displayed higher SLA and LNC and lower rooting, C:N in shoots
150 and root-shoot ratio; and therefore was defined as the acquisitive grassland. The ‘seed mixture for dry
151 locations (SR032)’ with opposite traits was defined as the conservative grassland (Wright et al., 2004;
152 Roumet et al., 2016). Plant biomass was cut and separated into functional groups prior to the start of the
153 drought experiment to confirm the relation between biomass and seed share. Furthermore, SLA was
154 calculated for the three main species of each grassland type to confirm that mean SLA in conservative
155 grassland was lower ($28.9 \text{ m}^2 \text{ kg}^{-1}$) than in acquisitive grassland ($33.9 \text{ m}^2 \text{ kg}^{-1}$).



156 **Table 1** Plant species and corresponding functional group present in the seed mixtures with their
 157 respective plant traits, i.e. rooting index, specific leaf area (SLA), leaf nitrogen content (LNC), C:N ratio
 158 in shoots and root-shoot ratio; the share (%) of each species in the seed mixtures. Rooting index was
 159 calculated as the product of the root density (classes) × main rooting depth (m) × root lateral extension
 160 (m) based on species-specific data according to Tasser and Tappeiner (2005). The grassland types were
 161 characterised based on these two seed mixtures, utilising the community-weighted means, i.e. mean
 162 values of the traits and seed percentages.

Species	Functional group	Plant traits					Species abundance in seed mixture (%)		
		Rooting index	SLA (m ² kg ⁻¹)	LNC (mg g ⁻¹)	C:N of shoot	Root-shoot ratio	for dry locations (SR032)	for humid locations (SR037)	
		<i>Agrostis capillaris</i> L.	grass	1.42 ^a	29.31 ^{b,c,d}	23.65 ^d	20.32 ^m	0.41 ^{aa,ab}	0.0
<i>Alopecurus pratensis</i> L.	grass	0.81 ^a	24.08 ^{b,c}	41.50 ^d	17.47 ^{m,n,o}	0.39 ^{ac}	0.0	6.1	
<i>Arrhenatherum elatius</i> P. Beauv.	grass	4.62 ^a	29.68 ^{ef}	29.62 ^d	26.00 ^o	0.74 ^{ab}	4.4	0.0	
<i>Dactylis glomerata</i> L.	grass	0.15 ^a	24.35 ^{d,f,g,h,i}	24.30 ^d	19.63 ^p	0.79 ^{ab,ad}	6.7	7.0	
<i>Festuca arundinacea</i> Schreber	grass	3.88 ^a	17.40 ^{i,j}	18.00 ^d	36.33 ^{q,r,s}	0.91 ^{ad}	50.6	0.0	
<i>Festuca pratensis</i> Hudson	grass	0.64 ^a	22.29 ^{d,f}	24.95 ^d	20.00 ^o	0.68 ^{ae}	2.7	18.1	
<i>Festuca rubra</i> L.	grass	0.64 ^a	11.62 ^{d,g}	34.00 ^d	29.57 ^t	0.43 ^{af}	5.0	12.0	
<i>Lolium perenne</i> L.	grass	0.66 ^a	22.60 ^{f,i,j}	33.00 ^d	34.21 ^{t,u}	0.70 ^{aa,ab,ad,ag}	6.9	13.5	
<i>Lotus corniculatus</i> L.	legume	0.25 ^a	18.8 ^{d,e,k}	28.98 ^d	22.60 ^v	0.62 ^{ah}	0.0	2.5	
<i>Medicago sativa</i> L.	legume	0.76 ^a	28.50 ^j	55.50 ^d	21.60 ^v	0.24 ^{ai,aj}	3.9	0.0	
<i>Phleum pratense</i> L.	grass	0.40 ^a	28.41 ^{d,j}	25.42 ^d	18.30 ^w	0.86 ^{ab}	5.8	8.0	
<i>Poa pratensis</i> L.	grass	0.31 ^a	20.29 ^{d,l}	34.50 ^d	29.37 ^t	0.67 ^{ad,ag}	5.6	12.1	
<i>Trifolium hybridum</i> L.	legume	0.84 ^a	33.68 ^{g,i}	43.20 ^d	14.29 ^x	0.27 ^{ak}	0.0	4.0	
<i>Trifolium pratense</i> L.	legume	0.21 ^a	25.07 ^{d,g,j}	34.86 ^d	18.10 ^v	0.30 ^{ad,ah}	2.8	2.5	
<i>Trifolium repens</i> L.	legume	0.84 ^a	40.84 ^{d,g,i}	49.10 ^d	14.10 ^v	0.36 ^{aa,ad}	5.6	6.0	
<i>Trisetum flavescens</i> (L.) P. Beauv.	grass	0.62 ^a	29.00 ^{d,h}	21.55 ^d	17.75	0.81 ^{al}	0.0	4.1	
Grassland type	Community-weighted means								
Conservative (seed mixture for dry location)		2.39	21.37	25.99	30.07	0.76			
Acquisitive (seed mixture for humid location)		0.59	23.67	31.61	23.24	0.60			

163 Literature: a) Tasser and Tappeiner (2005), b) Kleyer et al. (2008), c) Kattge et al. (2020), d) Fontana et
 164 al. (2017), e) Freschet et al. (2015), f) Arredondo and Schnyder (2003), g) Wohlfahrt et al. (1999), h)
 165 Guyonnet et al. (2018), i) Prieto et al. (2017), j) Mason et al. (2016), k) Ordóñez et al. (2010), l) Meziane



166 and Shipley (2001), m) Baxendale et al. (2014), n) De Long et al. (2019), o) Arfin Khan et al. (2016),
167 p) Ryser and Eek (2000), q) Chen et al. (2017), r) Jacob et al. (2020), s) Bowden et al. (2007), t) Evers
168 et al. (2020), u) Shahzad et al. (2012), v) Talgre et al. (2017), w) Hempel et al. (2009), x) Davis (1991),
169 y) Kaleem Abbasi et al. (2015), z) Gubsch et al. (2011), aa) Cook et al. (1996), ab) Crush et al. (2005),
170 ac) Gloser (1993), ad) Davidson (1969), ae) Malinowski (1995), af) Paterson et al. (2005), ag) Pande
171 and Singh (1981), ah) Foulds (1977), ai) Philippot et al. (1991), aj) Zhang et al. (2018a), ak) Murray
172 (2012), al) Baillard et al. (2020).

173

174 **2.2. Experimental Set-up**

175 Twelve small-scale lysimeters (Smart-Field-Lysimeter, UMS/Meter Group Munich, Germany) with a
176 depth and diameter of 0.3 m were used for the experiment (Fig. 1). Each lysimeter was filled with a
177 standard horticultural soil. The soil had $6.03\% \pm 0.07\%$ of organic matter and a pH of 7.21 ± 0.07 .
178 More information on the soil characterization can be found in the appendice. All lysimeters were
179 equipped with an independent bi-directional pumping system between the lysimeters and drainage
180 containers, thus ensuring tension-controlled soil moisture conditions. An electronic weighing platform
181 (PL-50, UMS/Meter Group Munich, Germany) under each lysimeter recorded the weight every minute.
182 With the same frequency, the weight of each water drainage container was recorded (used electronic
183 platform: PL-10, UMS/Meter Group Munich, Germany). Both measurements were applied for the
184 calculation of evapotranspiration (ET). Soil water content (SWC) was recorded every 10 minutes by
185 EC-5 sensors (Decagon Devices, Inc.), at a soil depth of 5, 15 and 25 cm (Frenck et al., 2018) and the
186 daily mean of the three sensors was calculated per lysimeter. Sensors were calibrated for repeatability
187 and accuracy over a period of one month prior to the start of the experiment. This ensured they
188 consistently provided the same values under identical conditions, thus minimising random noise. To do
189 this, they were calibrated to the mean as a reference by creating a sensor-specific linear calibration curve
190 between the measured values and the mean values (see also Schwamback et al. (2023)). These sensor-
191 specific calibration functions were then applied throughout the experiment.

192 The seed mixtures were sown in the lysimeters in the previous year and kept in a greenhouse at 20°C.
193 After installation in the experimental garden, the lysimeters were managed according to the scheme of
194 the surrounding meadow to settle (i.e. cutting, and exposed to the natural precipitation regime of the
195 study site where mean annual precipitation is 1100 mm). The density of plant individuals at the time of
196 the experiment was 2512.4 ± 349.4 ind. m⁻² (legumes: 589.5 ± 123.0 ind. m⁻², grasses: 1922.9 ± 330.2
197 ind. m⁻²). The experiment started directly after the first harvest of the season and all lysimeters were
198 covered with a UV-transmissive transparent polyethylene film (Lumisol Clear AF. 88-92 % light



199 transmission) to exclude natural precipitation. The shelters were approximately 2.5 m tall and were
200 closed from 0.5 to 2.5 m, but open on the sides facing the main wind directions to account for typical
201 wind conditions at the site. Lysimeters were subjected to two irrigation levels (i.e. wet and drought
202 treatments, Fig. 1) from 8 June to 8 August (64 days). For the wet treatment, soil moisture was
203 maintained at field capacity by controlled watering to avoid waterlogging and ensure sufficient water
204 supply for maximum productivity. For the drought treatment, soil was maintained in a water deficit. For
205 the first 26 days, the lower boundary was maintained at a constant matric potential of -1 kPa. This was
206 achieved by a bi-directional pumping system that delivered water in and out of the lysimeter to mimic
207 groundwater inflow and capillary rise from deeper soil layers. From thereon until the end of the
208 experiment (3 July to 8 August), no watering was applied and extreme drought was simulated.

209

210 **2.3. Measurements**

211 *Water fluxes*

212 ET was calculated by the weight differences between the lysimeter and the drainage container, and
213 applying the AWAT (Adaptive Window Adaptive Threshold) filter to separate significant from
214 insignificant mass changes (Peters et al., 2014; Peters et al., 2016; Peters et al., 2017). The integrated
215 WUE of the grasslands was calculated by dividing dry phytomass by the accumulated ET for the
216 experimental period. Here, aboveground dry phytomass was used for calculating WUE_{ap} and total dry
217 phytomass (i.e., aboveground and belowground) was used to calculate WUE_{ip} (Kirkham, 2005).

218 Starting on 3 July, weekly stomatal conductance measurements were taken on the leaf underside of
219 grasses and legumes in the lysimeters using a DECAGON SC-1 Leaf Porometer (Decagon Devices;
220 2016) in four measurement cycles from 8:00 h to 18:00 h in sun-exposed individuals. The measurements
221 could not be carried out at the species level due to the young age of the canopy and the resulting non-
222 definitive species recognition. Three replicates per functional group in each lysimeter were used to
223 calculate a daily mean, resulting in 12 single measurements per day for each functional group and
224 lysimeter.

225 *Phytomass*

226 At the end of the experiment, individuals of grasses and legumes were counted. Between 5 and 20
227 individuals of legumes and grasses were randomly selected and excavated to determine the root-shoot
228 ratio, depending on how strongly a group was represented in the lysimeter. Root extraction was done
229 manually by washing the roots under running water. The low stand density and loose soil made
230 individual roots easily distinguishable during washing, as they were not tightly intertwined. Afterwards,
231 roots were dried at 80°C , and weighed to determine the average root weight per functional group in each
232 lysimeter. The remaining aboveground phytomass in the lysimeters was harvested, separated into
233 functional groups, dried at 80°C , and weighed. We included individuals used for root-shoot ratio



234 assessment. Biomass and necromass proportions were visually estimated at five time points (19 June, 9
235 July, 18 July, 1 August, and 7 August) in each lysimeter during the experiment.

236 *Spore productivity of arbuscular mycorrhizal fungi*

237 The abundance of AMF spores in the initial soil mixture used represented the amount of available AMF
238 inoculum in the soil at the beginning of the season. End-of-season sporulation was considered an
239 approximation of AMF reproduction. The extraradical spores produced by AMF in the soil were
240 extracted by the differential water-sucrose centrifugation method (Ianson and Allen, 1986). First, 30 g
241 of 2 mm-sieved dry soil was mixed with 60 ml of water to allow rehydration, the solution was
242 centrifuged at 2,000 rpm for 10 min, and the supernatant containing organic material was discarded. The
243 soil was re-suspended in 60 ml of a 2 M sucrose solution with a dispersant (2 % Calgon) and centrifuged
244 at 2,000 rpm at room temperature for 10 min. Next, to remove soil particles remaining in suspension, a
245 sucrose-Calgon solution was added to the supernatant containing the AMF spores to reach an
246 approximate volume of 20 ml. The mixture was sieved through a 500 μm sieve and centrifuged again
247 for 10 min at 2,000 rpm. The material retained on the 500 μm sieve was examined under a stereo
248 microscope to check for the presence of large AMF spores. To separate the spores into three additional
249 size classes (i.e. 51 – 100 μm , 100 – 200 μm , and 200 – 500 μm), the spore-containing supernatant was
250 consecutively sieved through a 200 μm and then a 100 μm sieve cascade. The spores retained on each
251 sieve were separately washed off with distilled water and poured into the filter funnel of a vacuum line
252 covered with a 51 μm mesh tissue. Spores of each size class retained on the mesh were counted under
253 the stereo microscope at 63-fold magnification. Size classes and pigmentation of AMF spores were
254 recorded as they are common criteria for identification (Invam, 2023; Schüßler and Walker, 2010).
255 Although no molecular AMF species classification was carried out, the main AMF classes could be
256 discriminated by the combination of size and pigmentation. This allowed us to detect compositional
257 changes in AMF communities. We considered small (51 - 100 μm), medium (100 – 200 μm), and large
258 (200 - 500 μm) spore sizes and black, brown and yellow (pale) spore pigmentation. The density of AMF
259 spores was expressed on a soil dry weight basis.

260

261 **2.4. Data processing and statistics**

262 Most of the statistical analyses were performed using the R statistical programming language (R Core
263 Team, 2022). To compare SWC, ET, and necromass between irrigation levels and grassland types,
264 ANCOVAs were used. Date was included as a covariate, and grassland type as well as irrigation levels
265 were independent variables. In addition, interactions between variables were tested. For stomatal
266 conductance, functional group was added as independent value. Finally, legumes and grasses stomatal
267 conductance were tested also independently. ANOVAs with a significance level of 0.05 were performed
268 to compare aboveground, belowground phytomass, abundance of AMF spores, WUE_{ap} , and WUE_{ip} ,
269 between the different irrigation levels and grassland types. All ANOVAs included grassland type and



270 irrigation levels as independent variables, and interactions between variables were tested. Above- and
271 belowground phytomass also included plant functional groups as independent variables, whilst AMF
272 spores included the different size and pigmentation classes of AMF spores. When interactions between
273 variables were significant, a post hoc Tukey test was performed to test the differences between groups.
274 The homogeneity of variances and distribution of residuals were analysed using residual plots.
275 Additionally, the Levene test with “car” package (Fox and Weisberg, 2019) was used to examine the
276 homogeneity of variances and Shapiro-Wilk’s with “stats” package (R Core Team, 2022) was used for
277 analysing distribution of residuals. Aboveground and belowground phytomass were log-transformed to
278 meet the assumptions of normality and homogeneity of variances.

279 Finally, we estimated the relationship between WUE and the above- and belowground plant
280 characteristics, parameters of the water balance and AMF. We ran a factor analysis using varimax
281 rotation and reduced our 32 variables to five factors (Table S3). This allowed us to assess
282 multicollinearities between the variables as well as in terms of trade-offs and redundancies. Linear
283 regression analysis was then used to test the effect of these factors on WUE_{ap} and WUE_{tp} . These analyses
284 were performed in SPSS Statistics (IBM SPSS Statistics 28).

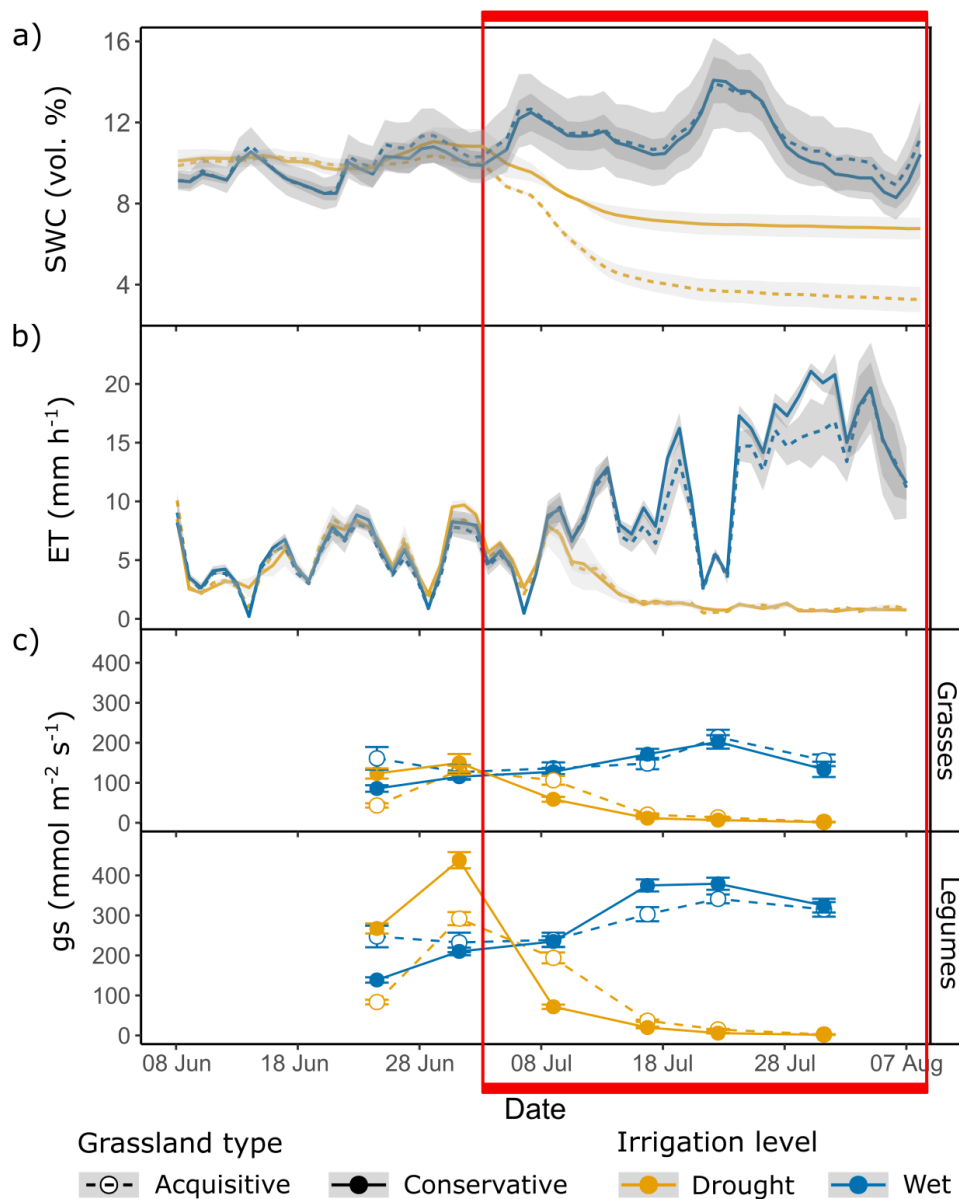
285

286 **3. Results**

287 **3.1. Drought intensity and water fluxes**

288 The drought treatment effectively induced a severe drought, where soil water content (SWC) decreased
289 immediately after the start of the extreme drought (3 July) for both grassland types (Table S1, irrigation
290 level, $p < 0.001$). However, the decrease was more pronounced in acquisitive grassland (up to $3.27 \pm$
291 0.62 %) than in conservative grassland (up to 6.76 ± 0.54 %; irrigation level * grassland type, $p < 0.001$)
292 under drought. In contrast, lysimeters in wet treatment generally maintained high and similar SWC (Fig.
293 2a).

294



295

296 **Figure 2. a) Mean soil water content (SWC), b) evapotranspiration (ET) and, c) stomatal conductance (gs)**
 297 **over the course of the experiment for conservative and acquisitive grasslands under wet and drought**
 298 **treatment. The shaded area represents the standard error and the red box indicates the duration of the**
 299 **extreme drought. Stomatal conductance values are presented as $\bar{x} \pm \text{s.e.}$**

300 Before the extreme drought event, evapotranspiration (ET) dynamics were similar between conservative
 301 and acquisitive grasslands, irrespective of the irrigation level (Fig. 2b). With the onset of the extreme



302 drought, ET changed over time (Table S1, date, $p = 0.001$), and differed between irrigation levels ($p <$
303 0.001). Wet lysimeters exhibited increased ET, while drought lysimeters showed a decreased. The
304 impact of the drought treatment on ET became evident on day 6 of the extreme drought (9 July) (Fig.
305 2b).

306 Measuring stomatal conductance throughout the experiment (Fig. 2c) revealed that legumes had higher
307 conductance rates than grasses (Table S1, functional group, $p < 0.001$) and reacted differently to drought
308 (irrigation level * functional group, $p < 0.001$). Legumes in acquisitive grassland only reduced their
309 conductance after two weeks of extreme drought, whereas in conservative grassland substantial
310 reductions were observed one week earlier (Table S1, irrigation level * grassland type, $p < 0.001$). For
311 grasses, drought reduced stomatal conductance (Table S1, irrigation level, $p < 0.001$) after two weeks
312 of extreme drought in both grassland types. We also observed that grasses in the conservative grassland
313 tended to reduce stomatal conductance earlier, although this difference was smaller than that in legumes.

314 **3.2. Aboveground and belowground impacts**

315 In wet treatment, both grassland types produced similar aboveground phytomass ($1.01 \pm 0.15 \text{ kg m}^{-2}$
316 and $1.00 \pm 0.11 \text{ kg m}^{-2}$ for acquisitive and conservative grasslands, respectively). In the drought
317 treatment, aboveground phytomass production dropped (Table S2, irrigation level, $p = 0.001$) to $0.26 \pm$
318 0.03 kg m^{-2} in acquisitive and $0.38 \pm 0.03 \text{ kg m}^{-2}$ in conservative grassland. Grasses and legumes differed
319 significantly in their total aboveground phytomass (Fig. 3a, functional group $p < 0.001$), and on the
320 grassland types (grassland type * functional group, $p = 0.004$), and they reacted differently to drought
321 (irrigation level * functional group, $p < 0.001$). In conservative grassland, the proportion of grasses
322 increased from 2.2 % in wet to 9.7 % in drought treatment, whilst in acquisitive grassland, it increased
323 from 3.9 to 19.9 %. The proportion of legumes decreased with drought, with a smaller reduction in
324 conservative grassland (7.5 % vs 16.0 % reduction from wet to drought treatment for conservative and
325 acquisitive grassland, respectively).

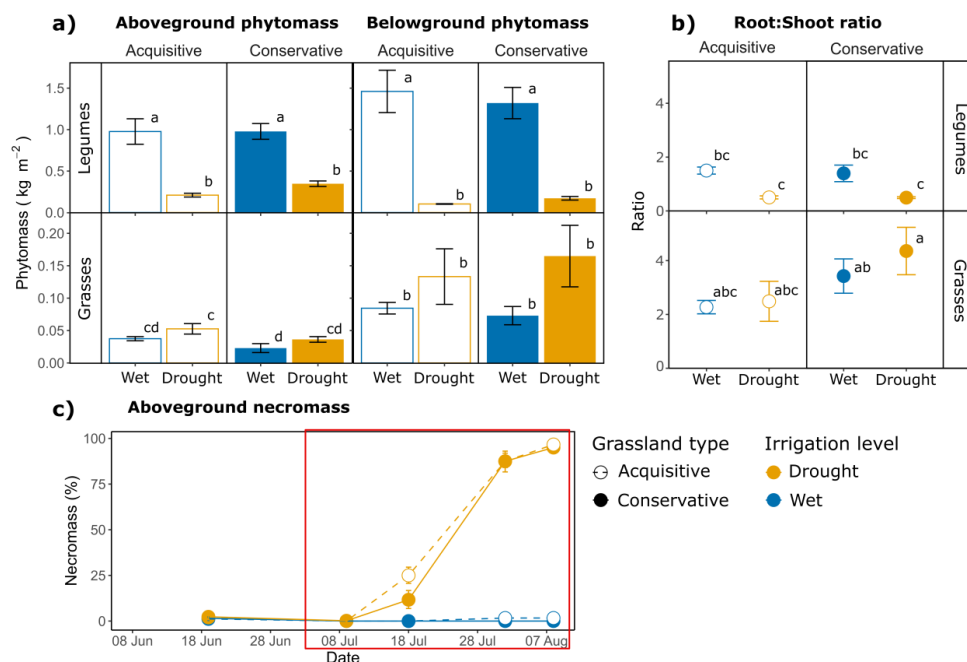
326 Belowground phytomass did not differ between grassland types (Table S2, grassland type, $p = 0.46$).
327 Acquisitive grassland reduced belowground phytomass from $1.45 \pm 0.26 \text{ kg m}^{-2}$ in wet to 0.24 ± 0.04
328 kg m^{-2} in drought treatment. Whilst conservative grassland reduced belowground phytomass from 1.39
329 $\pm 0.20 \text{ kg m}^{-2}$ to $0.37 \pm 0.03 \text{ kg m}^{-2}$ between wet and dry treatments. However, legumes and grasses were
330 affected differently by drought (functional group, $p < 0.001$). Legumes decreased their belowground
331 phytomass after drought, while grasses increased it (Fig. 3a). Accordingly, the root-shoot ratio (Fig. 3b)
332 increased for grasses and decreased for legumes (Table S2, irrigation level * functional group, $p =$
333 0.043). Additionally, grasses exhibited a higher root-shoot ratio in conservative than acquisitive
334 grasslands (grassland type, $p = 0.052$), independently of the irrigation level.

335 At the end of the extreme drought period, the proportion of necromass aboveground in both grassland
336 types affected by drought was more than 90 %. Almost all aboveground parts of the legumes and



337 approximately 80 % of the grasses had died (Fig. 3c). Die-off commenced in the first third of the drought
 338 period and increased with duration (Table S1, date * irrigation level, $p < 0.001$). After two weeks of
 339 extreme drought (i.e. 18 July), the proportion of necromass in the conservative grassland was 13 %, and
 340 in the acquisitive grassland was 30 %.

341



342

343 **Figure 3. a) Above- and belowground phytomass ($\bar{x} \pm$ s.e.), b) root:shoot ratio ($\bar{x} \pm$ s.e.), and c) proportion of necromass on the aboveground phytomass ($\bar{x} \pm$ s.e.) of the conservative and acquisitive grasslands under**
 344 **wet versus drought treatments. The red box represents the duration of the extreme drought. Letters indicate the significant differences found by post-hoc analysis following three different ANOVAs ($p < 0.05$): i)**
 345 **wet versus drought treatments. The red box represents the duration of the extreme drought. Letters indicate the significant differences found by post-hoc analysis following three different ANOVAs ($p < 0.05$): i)**
 346 **aboveground phytomass, ii) belowground phytomass, and iii) root:shoot ratio.**
 347

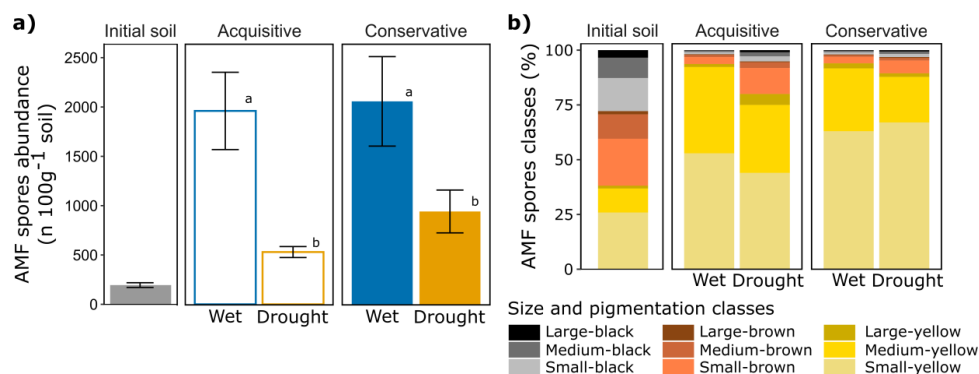
348 3.3. Arbuscular mycorrhizal fungi spore abundance

349 In all lysimeters, the total abundance of AMF spores increased over the course of the experiment
 350 compared to the initial AMF inoculum present in the soil (Fig. 4a). The drought treatment showed lower
 351 AMF spore abundance (Table S2, irrigation level, $p < 0.001$) and resulted in changes in the detected
 352 AMF size and pigmentation classes (Fig. 4b, irrigation level * size and pigmentation classes, $p = 0.008$).
 353 Results also indicated higher AMF spore productivity for the conservative grassland compared to
 354 acquisitive grassland after the drought treatment, but differences were not statistically significant (Fig.
 355 4a).



356 Nine different combinations of size and pigmentation in AMF spores were identified and AMF
 357 abundances significantly varied amongst classes (size and pigmentation classes, $p < 0.001$). Of these
 358 nine combinations, small and yellow spores were the most abundant (Fig. 4b). In conservative grassland,
 359 differences between irrigation levels were small. However, in acquisitive grassland, black and brown
 360 spores, especially the small-brown class, increased their abundance in drought conditions. The evenness
 361 of spore classes was high in the initial soil (0.81 ± 0.02), i.e. all abundances are the same when evenness
 362 is one. After the drought experiment, observed evenness generally decreased but this effect differed
 363 between irrigation levels and grassland types. Conservative grassland maintained similar evenness
 364 between irrigation levels (0.49 ± 0.03 for wet and 0.55 ± 0.06 for drought treatment). However, evenness
 365 of the AMF spores classes increased under drought in acquisitive grasslands from 0.52 ± 0.03 for wet
 366 to 0.66 ± 0.02 in drought (irrigation level, $p = 0.006$; grassland type, $p = 0.045$).

367



368

369 **Figure 4. a) Abundance of arbuscular mycorrhiza fungi (AMF) spores ($\bar{x} \pm$ s.e.) in conservative and**
 370 **acquisitive grasslands after drought or wet treatments. The initial soil corresponds to the AMF spores**
 371 **present in the standardized soil mixture used for the lysimeter setup. Spore counts are provided as AMF**
 372 **spores per 100 g of dry soil. Letters indicate the significant differences found by post-hoc analysis following**
 373 **ANOVAs ($p < 0.05$) comparing AMF spores numbers at the end of the experiment. b) Relative abundance**
 374 **of different AMF spore size and pigmentation classes in the soil of lysimeters with conservative or acquisitive**
 375 **grassland types under drought or wet treatment, and in the initial soil before the experiment. Size and**
 376 **pigmentation classes were combinations of small (51 - 100 μ m), medium (100 - 200 μ m), or large (200 - 500**
 377 **μ m) with black, brown and yellow.**

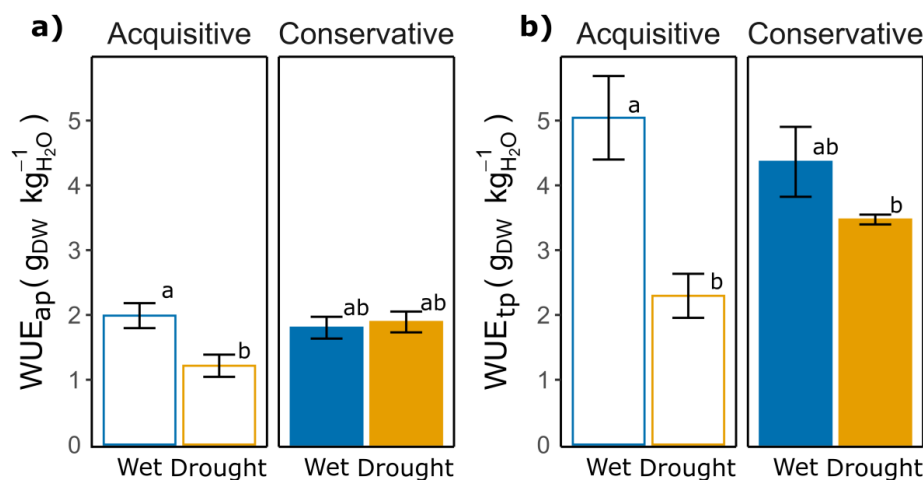
378 3.4. Water use efficiency (WUE) of two grassland types

379 The two calculated WUEs showed different trends (Fig. 5). Despite similarities in ET and aboveground
 380 phytomass productivity between grassland types under drought, a difference in WUE_{ap} was found at the
 381 end of the experiment (Fig. 5a). Drought WUE_{ap} decreased only in acquisitive grassland (Table S2,
 382 irrigation level * grassland type, $p = 0.038$), whilst conservative grassland showed no difference between
 383 wet and drought conditions. WUE_{tp} decreased with drought treatment (Table S2, irrigation level, $p =$



384 0.004), regardless of the grassland type (Fig. 5b). However, the decrease is much larger for acquisitive
 385 (-54.6 %) than for conservative grassland (-22.2 %).

386



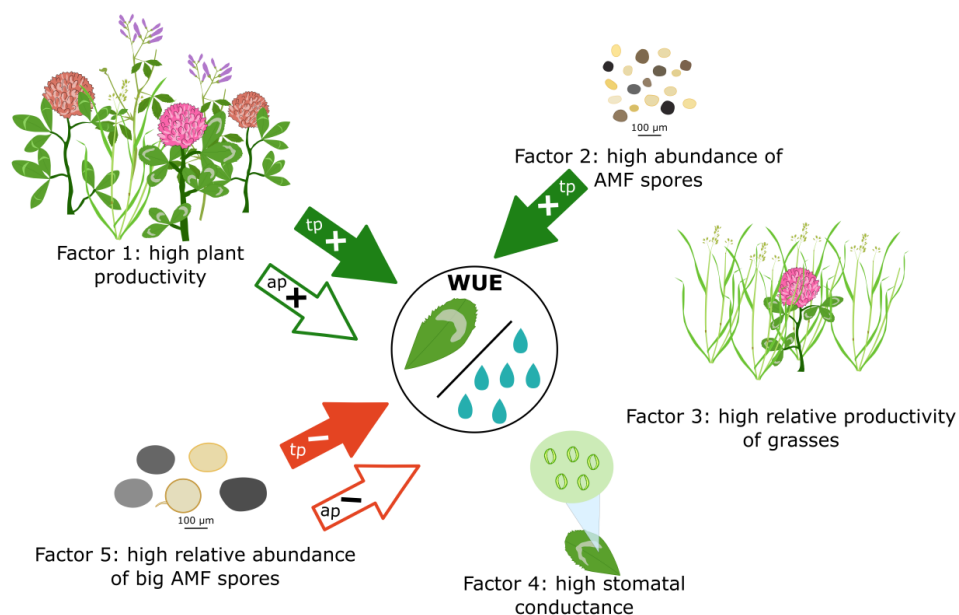
387

388 **Figure 5. Water use efficiency calculated at the end of the experiment as the amount of a) aboveground**
 389 **phytomass (WUE_{ap}) and b) total phytomass. Total phytomass is the sum of aboveground and belowground**
 390 **phytomass (WUE_{tp}) produced (gDW) per quantity of used water (kg H₂O) from conservative or acquisitive**
 391 **grasslands under drought and wet treatment ($\bar{x} \pm s.e.$). Letters indicate the significant differences found by**
 392 **post-hoc analysis following ANOVAs ($p < 0.05$).**

393 To estimate the relationship between WUE and the 32 potential driving variables (Table S3), firstly the
 394 variables were condensed into 5 factors: factor 1 related to high plant productivity, factor 2 related to a
 395 generally high abundance of AMF spores, factor 3 related to high productivity of grasses, factor 4 related
 396 to high stomatal conductance and factor 5 related to high abundance of large AMF spores. This grouping
 397 allowed us to see how AMF spores with specific size and pigmentation (e.g. small-black to medium-
 398 black spores) are associated with legumes, and others with grasses (small-brown spores). Secondly, the
 399 relationship of the factors with the two indices WUE_{ap} and WUE_{tp} were estimated. The results (Table
 400 S4, Fig. 6) show that WUE_{ap} was strongly correlated with the created factors ($R^2 = 0.68$, $p = 0.027$).
 401 Factor 1 (high total productivity) was positively correlated, and factor 5 (high abundance of large AMF
 402 spores) was negatively correlated with WUE_{ap} ($p = 0.005$ and $p = 0.020$, respectively). On the other
 403 hand, WUE_{tp} was also well explained by the considered factors ($R^2 = 0.96$, $p < 0.001$) and it was
 404 positively correlated with overall high productivity (factor 1, $p < 0.001$) and a general high abundance
 405 of AMF (factor 2, $p = 0.001$). However, WUE_{tp} was negatively correlated with the high abundance of
 406 large spores of AMF (factor 5, $p = 0.009$).



407



408

409 **Figure 6. Significant relationships between water use efficiency (WUE) and the five factors into which the**
410 **indicators were condensed based on a factor analysis. Filled arrows represent the statistically significant**
411 **correlations between the factor and WUE_{tp} and empty arrows between the factor and WUE_{ap}. Green**
412 **arrows indicate a positive relationship and red arrows show a negative relationship.**

413

414 **4. Discussion**

415 Our results provide experimental evidence of the effects of drought on two grassland types used for
416 hay production in two climatically different areas of the European Alps. These two grasslands were
417 classified as conservative and acquisitive according to their plant resource use strategies. In this study,
418 the drought treatment provoked an extreme water deficit for the plants. The conditions simulated may
419 not be observed in nature in the near future, however, as Reichstein et al. (2013) recommend,
420 experiments towards climate extremes are necessary to better understand the ecosystem processes.

421 **4.1. Consequences of water stress on aboveground productivity**

422 Although there were some differences in ET between grassland types in wet conditions, ET generally
423 increased over the season alongside the temperature and evaporative demand. In the drought treatment,
424 ET dropped following the decrease in SWC at the start of the extreme drought. Although we could not
425 find differences in how the two grassland types modulated ET to drought, differences were observed
426 when we looked more closely at the stomatal conductance of the main functional groups. Reduction in



427 stomatal conductance in both legumes and grasses within the conservative grassland preceded those in
428 the acquisitive grassland. The first detected measure where individuals in the acquisitive grassland
429 started limiting stomatal conductance occurred when SWC was approximately 4 %. In conservative
430 grassland, the first observed reduction started at a SWC of approximately 9 %. In agreement with
431 Leitinger et al. (2015) and Frenck et al. (2018), the acquisitive grassland type, which represents the one
432 sown in humid areas of the European Alps, did not react to drought until water scarcity became severe.
433 In other words, legumes in the acquisitive grassland showed little reduced stomatal conductance until
434 the soil became very dry. However, the results did not show as great a difference as we would expect
435 from seed mixtures specifically designed for such different climates.

436 The conservative grassland had numerically higher phytomass productivity than the acquisitive
437 grassland in drought treatment. However, the differences were not statistically significant, possibly due
438 to low sampling size ($n = 3$ per grassland type \times irrigation level). Nevertheless, we found that plants wilt
439 earlier in extreme drought conditions in acquisitive grassland. In this grassland type, stomatal
440 conductance was restricted later, therefore a higher water loss was experienced, and an earlier death was
441 expected (Martin-Stpaul et al., 2017). In this experiment, drought was long enough to provoke wilt in
442 most of the grassland community, but in a scenario with shorter drought periods, conservative grassland
443 would show higher living biomass.

444 The two grassland types studied in this paper differed from the seed mixture slightly on the percentage
445 of grasses (typically more conservative) and legumes (typically more acquisitive). However, it must be
446 noted that some grasses present in the acquisitive grassland showed traits that would be towards the
447 acquisitive end of the gradient, e.g. *Agrostis capillaris* and *Alopecurus pratensis* (Table 1). As in the
448 study by Tello-García et al. (2020), which utilised a commercial seed mixture typically used for
449 intensively cultivated hay grasslands, drought favoured the grass species over legume species. This
450 effect resulted in a shift towards a predominance of grasses in community composition. After drought,
451 the conservative grassland had a higher proportion of legumes (98 %) than the acquisitive one (80 %).
452 Legumes tend to use resources acquisitively, which may make them stronger competitors in favourable
453 conditions but worse as the conditions get more severe (Thakur et al., 2022; Komainda et al., 2019;
454 Carlsson et al., 2017). Accordingly, acquisitive grassland suffered a stronger decrease on SWC, which
455 is expected to translate into greater physiological stress. Even if the proportion of legumes remains high,
456 the decrease is evident and has a big impact on the total phytomass productivity. Legumes are plants
457 known to enrich soil with nitrogen due to their mutualistic relationship with nitrogen-fixing bacteria,
458 and thus provide valuable ecosystem services. These ecosystem services can be interrupted by drought
459 (Dollete et al., 2023) and decreased by the reduction of legumes biomass. Hence, a greater loss of
460 legumes results in i) a reduction in soil fertility (Ledgard and Steele, 1992; Wei et al., 2019); ii) greater
461 loss of an important food source for grazing animals, affecting their diet (Cole et al., 2022); iii) stronger
462 changes in plant community composition, possibly further declining soil fertility through species loss



463 (Spehn et al., 2002; Cong et al., 2014); and finally iv) greater loss of pollinators, as legumes are
464 important food sources for pollinators including bees and butterflies (Cole et al., 2022).

465 **4.2. Higher investment belowground for the conservative grassland**

466 Aboveground components of both grassland types did not always differ significantly under drought
467 conditions. This effect illustrates the importance of belowground components in shaping grasslands
468 drought strategies. Young individuals, such those sown here, adapt particularly well to changing
469 conditions so are especially well suited for such studies, as life-span plasticity decreases in the plant
470 ageing process (Henn and Damschen, 2021). In addition, the root turnover in young individuals is
471 generally short, meaning that rapid adaptations can also be recognised here (Sun et al., 2016; Wang et
472 al., 2020).

473 Besides a general reduction of phytomass productivity, drought usually changes biomass allocation, and
474 increases root-shoot ratio (Zhou et al., 2020). Investing in the root system under stress conditions
475 increases the ability to absorb water and nutrients, as the results of the grasses in this study support.
476 However, it is known that root responses to drought are very heterogeneous (Lozano et al., 2020).
477 Accordingly, the opposite root system effect was also observed in the legumes in this experiment, as
478 they drastically reduced root phytomass and thus the root-shoot ratio. Our findings confirm that legumes
479 usually follow acquisitive strategies, using the available resources to build fast-growing and short-lived
480 aboveground tissues, rather than larger root systems. A deeper analysis of the rooting system in the
481 direction of architectural and morphological traits could allow us to link the functional groups or
482 grassland rooting system to a higher or lower efficiency on carbon, nutrients input, plant uptake or
483 mineralisation of nutrients (Bardgett et al., 2014).

484 The abundance of AMF spores in the soil was measured at the end of the season as an indicator of stress.
485 Congruently, at the end of the drought experiment AMF spore abundance was significantly lower in
486 both grassland types compared to the wet treatment. The presence of AMF is known to improve drought
487 tolerance (Bahadur et al., 2019), and reduce plant stress and results in increased plant biomass (Sun et
488 al., 2017; Jacott et al., 2017), rooting system (Ji et al., 2019) and plant fitness either via direct effects on
489 stomatal conductance or enhanced nutrition and water-uptake (Hartman and Tringe, 2019). Although
490 we could not causally link increased plant fitness with the productivity of AMF spores, our results imply
491 that plants in the conservative grassland (with slightly higher biomass production) can allocate a greater
492 proportion of their resources into their associated AMF partners under drought. This allows them to
493 produce more spores (slightly higher spore abundance) than acquisitive grasslands. Other studies carried
494 out with similar main plant species than in our lysimeters reported an increased fitness of plants with
495 mycorrhiza symbionts under drought. This includes mycorrhized roots of *Trifolium repens* having
496 higher concentrations of proline, soluble proteins, flavonoids (Tuo et al., 2017), and mycorrhized
497 *Festuca arundinacea* leaves having a higher chlorophyll content and thus an enhanced nutrient uptake
498 under drought (Mahdavi et al., 2018).



499 Besides a slightly higher number of AMF spores produced, the differences on AMF spore classes
500 (defined based on spore size and pigmentation) provide insights in potential changes in AMF community
501 composition due to stress or a different reaction of plants with different strategies. Although we did not
502 further identify these classes to species level, we used this as a proxy for species composition, as AMF
503 spores with different size and pigmentation typically represent different taxonomical units (Invam, 2023;
504 Schüßler and Walker, 2010). Changes in the AMF community composition of acquisitive grassland
505 under drought could cause the proportion of some specific spore classes to change and evenness to
506 increase, indicating a homogenization of abundances. According to our results, these compositional
507 changes may be related to changes in the functional group composition of each grassland type, as
508 suggested by the factor analysis. AMF mutualism can be species-specific or genus-specific to associated
509 plant species (Zafari et al., 2017; Ben Laouane et al., 2019; Bhandari et al., 2020). Thus, mycorrhizal
510 fungal communities often follow plant compositional changes (Kruger et al., 2017). Another reason for
511 this shift in community composition could be adaptation to drought. The species *Funneliformis mossae*
512 and *Glomus margarita* (Ji et al., 2019) or taxa of the genus *Rhizophagus* (Emery et al., 2022) are
513 examples of AMF with increased abundance under drought. Molecular identification on species level
514 and further *in-vitro* experiments would help to better understand the changes on the AMF community
515 and the factors causing these changes.

516 **4.3. Conservative strategies lead to higher water use efficiency under drought**

517 All plant traits and characteristics listed above (i.e. ET, stomatal conductance, plant productivity
518 (aboveground and belowground) and AMF sporulation) respond to drought with the main goal of greater
519 water use efficiency by balancing water consumption and phytomass productivity (e.g. stomatal closure
520 (Brodribb et al., 2020), investing in root phytomass (Li et al., 2022), or maintaining AMF symbiosis
521 (Zhao et al., 2015)). Conservative and acquisitive grasslands did not show all the significant differences
522 we expected but followed different trends in some aspects. Conservative grassland under drought
523 reduced stomatal conductance of legumes earlier, had higher root-shoot ratio of grasses and higher AMF
524 activity. These trends were supported by the significant difference in WUE_{ap} between both grassland
525 types after drought. In conservative grassland, WUE_{ap} was similar in wet and drought treatment.
526 Therefore the grassland appeared to react to changes in water availability through specific mechanisms
527 to maintain efficiency, as suggested by Lubbe et al. (2021). However, the acquisitive grassland did not
528 react to the decrease in water availability, resulting in a drop in WUE_{ap} , which we expected to be caused
529 by a high water loss by the later closing of the stomata (Brodribb et al., 2020). However, WUE_{ap} was
530 positively related to the total productivity led by legumes and AMF spores related to legumes (small-
531 and medium-black, and large-brown), and negatively to some specific classes of large spores (black and
532 yellow) (Fig. 6). We detected changes on the proportion of these spores at the different irrigation levels
533 of the acquisitive grassland. Therefore, a deeper understand on the changes on the mycorrhizal
534 community could help to disentangle the decrease on WUE_{ap} in acquisitive grassland under drought.



535 However, when we expand the focus to the whole phytomass productivity and calculate WUE_{tp} , the
536 greater efficiency of conservative grassland under drought is no longer clear. At this scale, differences
537 in wet vs drought become higher, as expected for grasslands under severe drought (Gang et al., 2016;
538 Zhao et al., 2020). Therefore, when interpreting the commonly used WUE_{ap} , we must consider that the
539 relevant belowground part and the changes on phytomass allocation are missing (Quan et al., 2020). In
540 addition, at least in this study, WUE_{ap} for grasslands under drought was overestimated. This
541 overestimation is obvious in our results as root phytomass was more strongly affected than aboveground
542 phytomass and differences between wet and drought were greater. These responses on the belowground
543 phytomass were not strong enough to equal WUE_{tp} of the wet treatment, but pinpoint a higher WUE_{tp} in
544 conservative than in acquisitive grassland under drought treatment. Besides the factors related to WUE_{ap} ,
545 WUE_{tp} is positively related to total AMF spores abundance and spores not linked to grasses nor legumes.
546 This reinforces the potential importance of AMF on WUE in agricultural grasslands.

547 **4.4. Implications for grassland farming**

548 In addition to disentangling plant mechanisms during drought, we focus this study on productivity and
549 efficiency which have the greatest impact on grassland management. We can confirm the decrease in
550 productivity due to drought, but cannot statistically separate grassland types. However, two other results
551 need to be considered: vitality and WUE. A more acquisitive strategy could result in less vigorous
552 grassland with lower efficiency which should lead to a decline in productivity in the future. We should
553 also point out that under wet treatment, conservative grassland reached similar productivity to
554 acquisitive grassland. This finding is contrary to the expectations that acquisitive grassland would grow
555 more and faster in favourable conditions (Reich, 2014). This may indicate a high plasticity of the
556 conservative grassland. However, further analysis of the nutritional value and palatability of the forage
557 would be interesting to ensure that quality is not reduced in more conservative grasslands.

558 Although it appears in this study that a conservative strategy would be better to tolerate drought, it is
559 known that diversity of strategies among coexisting species is more beneficial to the community (Pérez-
560 Ramos et al., 2013). For example, Griffin-Nolan et al. (2019) and Blessing et al. (2018) revealed in
561 different experiments that, on one hand, functional diversity decreased in dry years because conservative
562 grasses grew more and maintained higher soil moisture. On the other hand, acquisitive species grew less
563 or changed their phenology. However, in wet years or under short periods of drought, the acquisitive
564 species grew faster, increasing aboveground phytomass production and functional diversity. This
565 finding underlines that a mix of conservative and acquisitive species would be expected to perform
566 better than communities where only conservative species remain, which would have very low
567 productivity (Lavorel and Grigulis, 2012). We must consider that in this study the strategy attributed to
568 each grassland type was calculated as the CWM proxy of all the species coexisting, which could be
569 more acquisitive or conservative. Therefore, smaller differences would be expected than in other studies



570 comparing communities of exclusively conservative/slow species versus acquisitive/fast species (Oram
571 et al., 2023).

572 **5. Conclusions**

573 This study compares responses of two different grassland types to drought. We focus on their resource
574 use strategies and the impact on water fluxes, productivity, and AMF spore productivity. Although some
575 of the results are in line with our expectations and previous findings, (i.e. decrease of ET, phytomass
576 productivity, vitality, and AMF spore abundance) our results showed that the selected grassland types
577 differ in their reactions to drought according to their more acquisitive or conservative strategies. Our
578 predominantly conservative grassland revealed earlier stomata closure in legumes, a higher percentage
579 of legumes, slightly higher root-shoot ratio for grasses, and a slightly higher AMF spore productivity.
580 In particular, differences in legume productivity and abundance of certain AMF spores resulted in
581 greater WUE and prolonged vitality under extreme drought in comparison to the acquisitive grassland.

582 The use of seed mixtures adapted to prevailing climatic conditions should ensure high productivity in
583 grassland management. Therefore, our results suggest that increasing the abundance of species with
584 more conservative traits in a future climate with more frequent droughts should be considered. This
585 recommendation is based on findings that i) under favourable conditions they achieve the productivity
586 of acquisitive grassland, and ii) under drought conditions they show a lower yield loss, longer vitality
587 and higher efficiency. Besides these findings, it also became evident that the apparent aboveground
588 phytomass is not a good indicator of overall productivity or WUE. Therefore, root traits and plant-soil
589 interactions must also be considered for an overall assessment of drought tolerance.

590 **6. Data availability**

591 The datasets generated are available in B2SHARE with DOI:
592 10.23728/b2share.33d070b2e3d541ab8d147e80ebba9eb4

593 **7. Author contribution**

594 The study was conceived by ET, GL and UT. GL, ET, UP and ETG conceived the paper. Field
595 measurements, data collection and analysis were performed by ETG instructed by GL and ET on field
596 measurements and by UP on laboratory analysis. The first draft of the manuscript was written by ETG
597 and all authors commented on previous versions of the manuscript and made suggestions for its
598 structural and scientific improvement. All authors read and approved the final manuscript.

599 **8. Competing interests**

600 The authors declare that they have no conflict of interest.



601 **9. Acknowledgments**

602 We thank our collaborators at the Botanical Garden of the Universität of Innsbruck for their support;
603 Andre Peters and Lisa Ambrosi for his support with the AWAT filter routine; Ryan Stewart and Laura
604 Barraclough for constructive comments on the manuscript; and Alina Miler, Lisa Ambrosi, and Regina
605 Kuhnert for their help with soil analysis. This research was partially funded by the Austrian Science
606 Fund (FWF): [I4969-B], the project ClimAgro (Autonome Provinz Bozen – Südtirol, Abteilung
607 Bildungsförderung, Universität und Forschung) and the Austrian Federal Ministry of Science, Research
608 and Economy with the HRSM cooperation project KLIMAGRO. This study was conducted on the
609 LTSER site ‘Stubai’ (LTSER platform Tyrolean Alps) belonging to the national and international long-
610 term ecological research networks (LTER-Austria, LTER Europe and ILTER). The study site is part of
611 the Alpine Research Sites (ARS) at the Faculty of Biology of the Universität Innsbruck. UT and GL are
612 part of the Research Area ‘Mountain Regions’ and the Research Centre ‘Ecology of the Alpine Region’
613 at the Universität Innsbruck. ETG was supported by the Universität Innsbruck through a doctoral
614 scholarship (‘Doktoratsstipendium aus der Nachwuchsförderung der Universität Innsbruck’) and is
615 student member of the Innsbruck Doctoral College (IDC) ‘Alpine Biology and Global Change’. This
616 work is part of ETG’s PhD at the Universität Innsbruck.

617

618 **10. References**

619 Arfin Khan, M. A. S., Kreyling, J., Beierkuhnlein, C., and Jentsch, A.: Ecotype-specific improvement
620 of nitrogen status in European grasses after drought combined with rewetting, *Acta Oecologica*, 77, 118-
621 127, doi:10.1016/j.actao.2016.10.004, 2016.

622 Arredondo, J. T. and Schnyder, H.: Components of leaf elongation rate and their relationship to specific
623 leaf area in contrasting grasses, *New Phytologist*, 158, 305-314, doi:10.1046/j.1469-8137.2003.00745.x,
624 2003.

625 Augé, R. M.: Water relations, drought and vesicular-arbuscular mycorrhizal symbiosis, *Mycorrhiza*, 11,
626 3-42, doi:10.1007/s005720100097, 2001.

627 Bahadur, A., Batool, A., Nasir, F., Jiang, S., Mingsen, Q., Zhang, Q., Pan, J., Liu, Y., and Feng, H.:
628 Mechanistic insights into arbuscular mycorrhizal fungi-mediated drought stress tolerance in plants,
629 *International Journal of Molecular Sciences*, 20, doi:10.3390/ijms20174199, 2019.



- 630 Bahmani, M., Naghdi, R., and Kartoolinejad, D.: Milkweed seedlings tolerance against water stress:
631 Comparison of inoculations with *Rhizophagus irregularis* and *Pseudomonas putida*, *Environmental*
632 *Technology & Innovation*, 10, 111-121, doi:10.1016/j.eti.2018.01.001, 2018.
- 633 Baillard, V., Sulmon, C., Bittebiere, A. K., Mony, C., Couee, I., Gouesbet, G., Delignette-Muller, M.
634 L., Devin, S., and Billoir, E.: Effect of interspecific competition on species sensitivity distribution
635 models: Analysis of plant responses to chemical stress, *Ecotoxicology and Environmental Safety*, 200,
636 110722, doi:10.1016/j.ecoenv.2020.110722, 2020.
- 637 Bardgett, R. D., Mommer, L., and De Vries, F. T.: Going underground: root traits as drivers of ecosystem
638 processes, *Trends in Ecology & Evolution*, 29, 692-699, doi:10.1016/j.tree.2014.10.006, 2014.
- 639 Baxendale, C., Orwin, K. H., Poly, F., Pommier, T., and Bardgett, R. D.: Are plant-soil feedback
640 responses explained by plant traits?, *New Phytologist*, 204, 408-423, doi:10.1111/nph.12915, 2014.
- 641 Ben Laouane, R., Meddich, A., Bechtaoui, N., Oufdou, K., and Wahbi, S.: Effects of arbuscular
642 mycorrhizal fungi and rhizobia symbiosis on the tolerance of *Medicago Sativa* to salt stress, *Gesunde*
643 *Pflanzen*, 71, 135-146, doi:10.1007/s10343-019-00461-x, 2019.
- 644 Bergmann, J., Weigelt, A., van der Plas, F., Laughlin, D. C., Kuyper, T. W., Guerrero-Ramirez, N.,
645 Valverde-Barrantes, O. J., Bruelheide, H., Freschet, G. T., Iversen, C. M., Kattge, J., McCormack, M.
646 L., Meier, I. C., Rillig, M. C., Roumet, C., Semchenko, M., Sweeney, C. J., van Ruijven, J., York, L.
647 M., and Mommer, L.: The fungal collaboration gradient dominates the root economics space in plants,
648 *Science Advances*, 6, eaba3756, doi:10.1126/sciadv.aba3756, 2020.
- 649 Bhandari, K. B., West, C. P., and Acosta-Martinez, V.: Assessing the role of interseeding alfalfa into
650 grass on improving pasture soil health in semi-arid Texas High Plains, *Applied Soil Ecology*, 147,
651 103399, doi:10.1016/j.apsoil.2019.103399, 2020.
- 652 Blessing, C. H., Mariette, A., Kaloki, P., and Bramley, H.: Profligate and conservative: water use
653 strategies in grain legumes, *Journal of Experimental Botany*, 69, 349-369, doi:10.1093/jxb/erx415,
654 2018.



- 655 Bowden, C., Spargo, J., and Evanylo, G.: Mineralization and N fertilizer equivalent value of composts
656 as assessed by tall fescue (*Festuca arundinacea*), *Compost Science & Utilization*, 15, 111-118,
657 doi:10.1080/1065657X.2007.10702320, 2007.
- 658 Brilli, F., Hortnagl, L., Hammerle, A., Haslwanter, A., Hansel, A., Loreto, F., and Wohlfahrt, G.: Leaf
659 and ecosystem response to soil water availability in mountain grasslands, *Agricultural and Forest
660 Meteorology*, 151, 1731-1740, doi:10.1016/j.agrformet.2011.07.007 2011.
- 661 Brodribb, T. J., Sussmilch, F., and McAdam, S. A. M.: From reproduction to production, stomata are
662 the master regulators, *The Plant Journal*, 101, 756-767, doi:10.1111/tpj.14561, 2020.
- 663 Bu, W., Huang, J., Xu, H., Zang, R., Ding, Y., Li, Y., Lin, M., Wang, J., and Zhang, C.: Plant functional
664 traits are the mediators in regulating effects of abiotic site conditions on aboveground carbon stock-
665 Evidence from a 30 ha tropical forest plot, *Frontiers in Plant Science*, 9, 1958,
666 doi:10.3389/fpls.2018.01958, 2018.
- 667 Carlsson, M., Merten, M., Kayser, M., Isselstein, J., and Wrage-Mönnig, N.: Drought stress resistance
668 and resilience of permanent grasslands are shaped by functional group composition and N fertilization,
669 *Agriculture, Ecosystems & Environment*, 236, 52-60, doi:10.1016/j.agee.2016.11.009, 2017.
- 670 Chacón-Madrugal, E., Wanek, W., Hietz, P., and Dullinger, S.: Traits indicating a conservative resource
671 strategy are weakly related to narrow range size in a group of neotropical trees, *Perspectives in Plant
672 Ecology, Evolution and Systematics*, 32, 30-37, doi:10.1016/j.ppees.2018.01.003, 2018.
- 673 Chen, W., Liu, H., Wurihan, Gao, Y., Card, S. D., and Ren, A.: The advantages of endophyte-infected
674 over uninfected tall fescue in the growth and pathogen resistance are counteracted by elevated CO₂,
675 *Scientific Reports*, 7, 6952, doi:10.1038/s41598-017-07183-y, 2017.
- 676 Cole, L. J., Baddeley, J. A., Robertson, D., Topp, C. F. E., Walker, R. L., and Watson, C. A.: Supporting
677 wild pollinators in agricultural landscapes through targeted legume mixtures, *Agriculture, Ecosystems
678 & Environment*, 323, 107648, doi:10.1016/j.agee.2021.107648, 2022.



- 679 Cong, W.-F., van Ruijven, J., Mommer, L., De Deyn, G. B., Berendse, F., and Hoffland, E.: Plant species
680 richness promotes soil carbon and nitrogen stocks in grasslands without legumes, *Journal of Ecology*,
681 102, 1163-1170, doi:10.1111/1365-2745.12280, 2014.
- 682 Cook, A., Marriott, C. A., Seel, W., and Mullins, C. E.: Effects of soil mechanical impedance on root
683 and shoot growth of *Lolium perenne* L., *Agrostis capillaris* and *Trifolium repens* L, *Journal of*
684 *Experimental Botany*, 47, 1075-1084, doi:10.1093/jxb/47.8.1075, 1996.
- 685 Crush, J. R., Waller, J. E., and Care, D. A.: Root distribution and nitrate interception in eleven temperate
686 forage grasses, *Grass and Forage Science*, 60, 385-392, doi:10.1111/j.1365-2494.2005.00488.x, 2005.
- 687 Davidson, R. L.: Effect of root/leaf temperature differentials on root/shoot ratios in some pasture grasses
688 and clover, *Annals of Botany*, 33, 561-569, doi:10.1093/oxfordjournals.aob.a084308, 1969.
- 689 Davis, M. R.: The comparative phosphorus requirements of some temperate perennial legumes, *Plant*
690 *and Soil*, 133, 17-30, doi:10.1007/BF00011895, 1991.
- 691 De Long, J. R., Semchenko, M., Pritchard, W. J., Cordero, I., Fry, E. L., Jackson, B. G., Kurnosova, K.,
692 Ostle, N. J., Johnson, D., Baggs, E. M., and Bardgett, R. D.: Drought soil legacy overrides maternal
693 effects on plant growth, *Functional Ecology*, 33, 1400-1410, doi:10.1111/1365-2435.13341, 2019.
- 694 Diaz, S., Kattge, J., Cornelissen, J. H., Wright, I. J., Lavorel, S., Dray, S., Reu, B., Kleyer, M., Wirth,
695 C., Prentice, I. C., Garnier, E., Bonisch, G., Westoby, M., Poorter, H., Reich, P. B., Moles, A. T., Dickie,
696 J., Gillison, A. N., Zanne, A. E., Chave, J., Wright, S. J., Sheremet'ev, S. N., Jactel, H., Baraloto, C.,
697 Cerabolini, B., Pierce, S., Shipley, B., Kirkup, D., Casanoves, F., Joswig, J. S., Gunther, A., Falczuk,
698 V., Ruger, N., Mahecha, M. D., and Gorne, L. D.: The global spectrum of plant form and function,
699 *Nature*, 529, 167-171, doi:10.1038/nature16489, 2016.
- 700 Diaz, S., Hodgson, J. G., Thompson, K., Cabido, M., Cornelissen, J. H. C., Jalili, A., Montserrat-Marti,
701 G., Grime, J. P., Zarrinkamar, F., Asri, Y., Band, S. R., Basconcelo, S., Castro-Diez, P., Funes, G.,
702 Hamzehee, B., Khoshnevi, M., Perez-Harguindeguy, N., Perez-Rontome, M. C., Shirvany, F. A.,
703 Vendramini, F., Yazdani, S., Abbas-Azimi, R., Bogaard, A., Boustani, S., Charles, M., Dehghan, M., de



- 704 Torres-Espuny, L., Falczuk, V., Guerrero-Campo, J., Hynd, A., Jones, G., Kowsary, E., Kazemi-Saeed,
705 F., Maestro-Martinez, M., Romo-Diez, A., Shaw, S., Siavash, B., Villar-Salvador, P., and Zak, M. R.:
706 The plant traits that drive ecosystems: Evidence from three continents, *Journal of Vegetation Science*,
707 15, 295-304, doi:10.1111/j.1654-1103.2004.tb02266.x, 2004.
- 708 Dollete, D., Lumactud, R. A., Carlyle, C. N., Szczyglowski, K., Hill, B., and Thilakarathna, M. S.: Effect
709 of drought stress on symbiotic nitrogen fixation, soil nitrogen availability and soil microbial diversity in
710 forage legumes, *Plant and Soil*, doi:10.1007/s11104-023-06348-1, 2023.
- 711 Elgersma, A. and Sjøgaard, K.: Effects of species diversity on seasonal variation in herbage yield and
712 nutritive value of seven binary grass-legume mixtures and pure grass under cutting, *European Journal*
713 *of Agronomy*, 78, 73-83, doi:10.1016/j.eja.2016.04.011, 2016.
- 714 Emery, S. M., Bell-Dereske, L., Stahlheber, K. A., and Gross, K. L.: Arbuscular mycorrhizal fungal
715 community responses to drought and nitrogen fertilization in switchgrass stands, *Applied Soil Ecology*,
716 169, 104218, doi:10.1016/j.apsoil.2021.104218, 2022.
- 717 Evers, G. W.: Forage legumes: forage quality, fixed nitrogen, or both, *Crop Science*, 51, 403-409,
718 doi:10.2135/cropsci2010.06.0380, 2011.
- 719 Evers, M., Kroon, H., Visser, E., and Caluwe, H.: Carbon accumulation of cool season sports turfgrass
720 species in distinctive soil layers, *Agronomy Journal*, 112, 3435-3449, doi:10.1002/agj2.20231, 2020.
- 721 Fontana, V., Kohler, M., Niedrist, G., Bahn, M., Tappeiner, U., and Frenck, G.: Decomposing the land-
722 use specific response of plant functional traits along environmental gradients, *Science of The Total*
723 *Environment*, 599-600, 750-759, doi:10.1016/j.scitotenv.2017.04.245, 2017.
- 724 Foulds, W.: The physiological response to moisture supply of cyanogenic and acyanogenic phenotypes
725 of *Trifolium repens* L. and *Lotus corniculatus* L, *Heredity*, 39, 219-234, doi:10.1038/hdy.1977.62, 1977.
- 726 Fox, J. and Weisberg, S.: *An {R} Companion to Applied Regression*, Sage [code], 2019.



- 727 Frenck, G., Leitinger, G., Obojes, N., Hofmann, M., Newesely, C., Deutschmann, M., Tappeiner, U.,
728 and Tasser, E.: Community-specific hydraulic conductance potential of soil water decomposed for two
729 Alpine grasslands by small-scale lysimetry, *Biogeosciences*, 15, 1065-1078, doi:10.5194/bg-15-1065-
730 2018, 2018.
- 731 Freschet, G. T., Swart, E. M., and Cornelissen, J. H.: Integrated plant phenotypic responses to
732 contrasting above- and below-ground resources: key roles of specific leaf area and root mass fraction,
733 *New Phytologist*, 206, 1247-1260, doi:10.1111/nph.13352, 2015.
- 734 Gang, C., Wang, Z., Chen, Y., Yang, Y., Li, J., Cheng, J., Qi, J., and Odeh, I.: Drought-induced dynamics
735 of carbon and water use efficiency of global grasslands from 2000 to 2011, *Ecological Indicators*, 67,
736 788-797, doi:10.1016/j.ecolind.2016.03.049, 2016.
- 737 Gloser, J.: Growth rate and total nonstructural saccharides content in *Alopecurus pratensis* L, *Biologia*
738 *Plantarum*, 35, 37-42, doi:10.1007/BF02921116, 1993.
- 739 Griffin-Nolan, R. J., Blumenthal, D. M., Collins, S. L., Farkas, T. E., Hoffman, A. M., Mueller, K. E.,
740 Ocheltree, T. W., Smith, M. D., Whitney, K. D., and Knapp, A. K.: Shifts in plant functional composition
741 following long-term drought in grasslands, *Journal of Ecology*, 107, 2133-2148, doi:10.1111/1365-
742 2745.13252, 2019.
- 743 Gross, N., Robson, T. M., Lavorel, S., Albert, C., Le Bagousse-Pinguet, Y., and Guillemin, R.: Plant
744 response traits mediate the effects of subalpine grasslands on soil moisture, *New Phytologist*, 180, 652-
745 662, doi:10.1111/j.1469-8137.2008.02577.x, 2008.
- 746 Gubsch, M., Buchmann, N., Schmid, B., Schulze, E. D., Lipowsky, A., and Roscher, C.: Differential
747 effects of plant diversity on functional trait variation of grass species, *Annals of Botany*, 107, 157-169,
748 doi:10.1093/aob/mcq220, 2011.
- 749 Guyonnet, J. P., Cantarel, A. A. M., Simon, L., and Haichar, F. E. Z.: Root exudation rate as functional
750 trait involved in plant nutrient-use strategy classification, *Ecology and Evolution*, 8, 8573-8581,
751 doi:10.1002/ece3.4383, 2018.



- 752 Hartman, K. and Tringe, S. G.: Interactions between plants and soil shaping the root microbiome under
753 abiotic stress, *Biochemical Journal*, 476, 2705-2724, doi:10.1042/BCJ20180615, 2019.
- 754 Hashem, A., Kumar, A., Al-Dbass, A. M., Alqarawi, A. A., Al-Arjani, A. F., Singh, G., Farooq, M., and
755 Abd Allah, E. F.: Arbuscular mycorrhizal fungi and biochar improves drought tolerance in chickpea,
756 *Saudi Journal of Biological Sciences*, 26, 614-624, doi:10.1016/j.sjbs.2018.11.005, 2019.
- 757 Hempel, S., Stein, C., Unsicker, S. B., Renker, C., Auge, H., Weisser, W. W., and Buscot, F.: Specific
758 bottom-up effects of arbuscular mycorrhizal fungi across a plant-herbivore-parasitoid system,
759 *Oecologia*, 160, 267-277, doi:10.1007/s00442-009-1294-0, 2009.
- 760 Henn, J. J. and Damschen, E. I.: Plant age affects intraspecific variation in functional traits, *Plant*
761 *Ecology*, 222, 669-680, doi:10.1007/s11258-021-01136-2, 2021.
- 762 Ianson, D. C. and Allen, M. F.: The effects of soil texture on extraction of vesicular-arbuscular
763 mycorrhizal fungal spores from arid sites, *Mycologia*, 78, 164-168,
764 doi:10.1080/00275514.1986.12025227, 1986.
- 765 INVAM: Species descriptions from reference cultures. International culture collection of (vesicular)
766 arbuscular mycorrhizal fungi: <https://invam.ku.edu/species-descriptions>, last access: July 2023.
- 767 Jacob, V., Zhang, H., Churchill, A. C., Yang, J., Choat, B., Medlyn, B. E., Power, S. A., and Tissue, D.
768 T.: Warming reduces net carbon gain and productivity in *Medicago sativa* L. and *Festuca arundinacea*,
769 *Agronomy*, 10, 1601, doi:10.3390/agronomy10101601, 2020.
- 770 Jacott, C., Murray, J., and Ridout, C.: Trade-offs in arbuscular mycorrhizal symbiosis: Disease
771 resistance, growth responses and perspectives for crop breeding, *Agronomy*, 7, 75,
772 doi:10.3390/agronomy7040075, 2017.
- 773 Ji, L., Tan, W., and Chen, X.: Arbuscular mycorrhizal mycelial networks and glomalin-related soil
774 protein increase soil aggregation in Calcaric Regosol under well-watered and drought stress conditions,
775 *Soil and Tillage Research*, 185, 1-8, doi:10.1016/j.still.2018.08.010, 2019.



776 Kaleeem Abbasi, M., Mahmood Tahir, M., Sabir, N., and Khurshid, M.: Impact of the addition of
777 different plant residues on nitrogen mineralization–immobilization turnover and carbon content of a soil
778 incubated under laboratory conditions, *Solid Earth*, 6, 197-205, doi:10.5194/se-6-197-2015, 2015.

779 Kattge, J. and Bonisch, G. and Diaz, S. and Lavorel, S. and Prentice, I. C. and Leadley, P. and
780 Tautenhahn, S. and Werner, G. D. A. and Aakala, T. and Abedi, M. and Acosta, A. T. R. and Adamidis,
781 G. C. and Adamson, K. and Aiba, M. and Albert, C. H. and Alcantara, J. M. and Alcazar, C. C. and
782 Aleixo, I. and Ali, H. and Amiaud, B. and Ammer, C. and Amoroso, M. M. and Anand, M. and
783 Anderson, C. and Anten, N. and Antos, J. and Apgaua, D. M. G. and Ashman, T. L. and Asmara, D. H.
784 and Asner, G. P. and Aspinwall, M. and Atkin, O. and Aubin, I. and Baastrup-Spohr, L. and Bahalkeh,
785 K. and Bahn, M. and Baker, T. and Baker, W. J. and Bakker, J. P. and Baldocchi, D. and Baltzer, J. and
786 Banerjee, A. and Baranger, A. and Barlow, J. and Barneche, D. R. and Baruch, Z. and Bastianelli, D.
787 and Battles, J. and Bauerle, W. and Bauters, M. and Bazzato, E. and Beckmann, M. and Beeckman, H.
788 and Beierkuhnlein, C. and Bekker, R. and Belfry, G. and Belluau, M. and Beloiu, M. and Benavides, R.
789 and Benomar, L. and Berdugo-Lattke, M. L. and Berenguer, E. and Bergamin, R. and Bergmann, J. and
790 Bergmann Carlucci, M. and Berner, L. and Bernhardt-Romermann, M. and Bigler, C. and Bjorkman, A.
791 D. and Blackman, C. and Blanco, C. and Blonder, B. and Blumenthal, D. and Bocanegra-Gonzalez, K.
792 T. and Boeckx, P. and Bohlman, S. and Bohning-Gaese, K. and Boisvert-Marsh, L. and Bond, W. and
793 Bond-Lamberty, B. and Boom, A. and Boonman, C. C. F. and Bordin, K. and Boughton, E. H. and
794 Boukili, V. and Bowman, D. and Bravo, S. and Brendel, M. R. and Broadley, M. R. and Brown, K. A.
795 and Bruelheide, H. and Brumlich, F. and Bruun, H. H. and Bruy, D. and Buchanan, S. W. and Bucher,
796 S. F. and Buchmann, N. and Buitenwerf, R. and Bunker, D. E. and Burger, J. and Burrascano, S. and
797 Burslem, D. and Butterfield, B. J. and Byun, C. and Marques, M. and Scalon, M. C. and Caccianiga, M.
798 and Cadotte, M. and Cailleret, M. and Camac, J. and Camarero, J. J. and Campany, C. and Campetella,
799 G. and Campos, J. A. and Cano-Arboleda, L. and Canullo, R. and Carbognani, M. and Carvalho, F. and
800 Casanoves, F. and Castagneyrol, B. and Catford, J. A. and Cavender-Bares, J. and Cerabolini, B. E. L.
801 and Cervellini, M. and Chacon-Madrigal, E. and Chapin, K. and Chapin, F. S. and Chelli, S. and Chen,
802 S. C. and Chen, A. and Cherubini, P. and Chianucci, F. and Choat, B. and Chung, K. S. and Chytry, M.
803 and Ciccarelli, D. and Coll, L. and Collins, C. G. and Conti, L. and Coomes, D. and Cornelissen, J. H.



804 C. and Cornwell, W. K. and Corona, P. and Coyea, M. and Craine, J. and Craven, D. and Crowsigt, J.
805 and Csecserits, A. and Cufar, K. and Cuntz, M. and da Silva, A. C. and Dahlin, K. M. and Dainese, M.
806 and Dalke, I. and Dalle Fratte, M. and Dang-Le, A. T. and Danihelka, J. and Dannoura, M. and Dawson,
807 S. and de Beer, A. J. and De Frutos, A. and De Long, J. R. and Dechant, B. and Delagrangé, S. and
808 Delpierre, N. and Derroire, G. and Dias, A. S. and Diaz-Toribio, M. H. and Dimitrakopoulos, P. G. and
809 Dobrowolski, M. and Doktor, D. and Drevojan, P. and Dong, N. and Dransfield, J. and Dressler, S. and
810 Duarte, L. and Ducouret, E. and Dullinger, S. and Durka, W. and Duursma, R. and Dymova, O. and A,
811 E. V. and Eckstein, R. L. and Ejtehadi, H. and Elser, J. and Emilio, T. and Engemann, K. and Erfanian,
812 M. B. and Erfmeier, A. and Esquivel-Muelbert, A. and Esser, G. and Estiarte, M. and Domingues, T. F.
813 and Fagan, W. F. and Fagundez, J. and Falster, D. S. and Fan, Y. and Fang, J. and Farris, E. and
814 Fazlioglu, F. and Feng, Y. and Fernandez-Mendez, F. and Ferrara, C. and Ferreira, J. and Fidelis, A. and
815 Finegan, B. and Firn, J. and Flowers, T. J. and Flynn, D. F. B. and Fontana, V. and Forey, E. and
816 Forgiarini, C. and Francois, L. and Frangipani, M. and Frank, D. and Frenette-Dussault, C. and Freschet,
817 G. T. and Fry, E. L. and Fyllas, N. M. and Mazzochini, G. G. and Gachet, S. and Gallagher, R. and
818 Ganade, G. and Ganga, F. and Garcia-Palacios, P. and Gargaglione, V. and Garnier, E. and Garrido, J.
819 L. and de Gasper, A. L. and Gea-Izquierdo, G. and Gibson, D. and Gillison, A. N. and Giroldo, A. and
820 Glasenhardt, M. C. and Gleason, S. and Gliesch, M. and Goldberg, E. and Goldel, B. and Gonzalez-
821 Akre, E. and Gonzalez-Andujar, J. L. and Gonzalez-Melo, A. and Gonzalez-Robles, A. and Graae, B. J.
822 and Granda, E. and Graves, S. and Green, W. A. and Gregor, T. and Gross, N. and Guerin, G. R. and
823 Gunther, A. and Gutierrez, A. G. and Haddock, L. and Haines, A. and Hall, J. and Hambuckers, A. and
824 Han, W. and Harrison, S. P. and Hattingh, W. and Hawes, J. E. and He, T. and He, P. and Heberling, J.
825 M. and Helm, A. and Hempel, S. and Hentschel, J. and Herault, B. and Heres, A. M. and Herz, K. and
826 Heuertz, M. and Hickler, T. and Hietz, P. and Higuchi, P. and Hipp, A. L. and Hirons, A. and Hock, M.
827 and Hogan, J. A. and Holl, K. and Honnay, O. and Hornstein, D. and Hou, E. and Hough-Snee, N. and
828 Hovstad, K. A. and Ichie, T. and Iqic, B. and Illa, E. and Isaac, M. and Ishihara, M. and Ivanov, L. and
829 Ivanova, L. and Iversen, C. M. and Izquierdo, J. and Jackson, R. B. and Jackson, B. and Jactel, H. and
830 Jagodzinski, A. M. and Jandt, U. and Jansen, S. and Jenkins, T. and Jentsch, A. and Jespersen, J. R. P.
831 and Jiang, G. F. and Johansen, J. L. and Johnson, D. and Jokela, E. J. and Joly, C. A. and Jordan, G. J.



832 and Joseph, G. S. and Junaedi, D. and Junker, R. R. and Justes, E. and Kabzems, R. and Kane, J. and
833 Kaplan, Z. and Kattenborn, T. and Kavelenova, L. and Kearsley, E. and Kempel, A. and Kenzo, T. and
834 Kerkhoff, A. and Khalil, M. I. and Kinlock, N. L. and Kissling, W. D. and Kitajima, K. and Kitzberger,
835 T. and Kjoller, R. and Klein, T. and Kleyer, M. and Klimesova, J. and Klipel, J. and Kloeppe, B. and
836 Klotz, S. and Knops, J. M. H. and Kohyama, T. and Koike, F. and Kollmann, J. and Komac, B. and
837 Komatsu, K. and Konig, C. and Kraft, N. J. B. and Kramer, K. and Kreft, H. and Kuhn, I. and
838 Kumarathunge, D. and Kuppler, J. and Kurokawa, H. and Kurosawa, Y. and Kuyah, S. and Laclau, J. P.
839 and Lafleur, B. and Lallai, E. and Lamb, E. and Lamprecht, A. and Larkin, D. J. and Laughlin, D. and
840 Le Bagousse-Pinguet, Y. and le Maire, G. and le Roux, P. C. and le Roux, E. and Lee, T. and Lens, F.
841 and Lewis, S. L. and Lhotsky, B. and Li, Y. and Li, X. and Lichstein, J. W. and Liebergesell, M. and
842 Lim, J. Y. and Lin, Y. S. and Linares, J. C. and Liu, C. and Liu, D. and Liu, U. and Livingstone, S. and
843 Llusia, J. and Lohbeck, M. and Lopez-Garcia, A. and Lopez-Gonzalez, G. and Lososova, Z. and Louault,
844 F. and Lukacs, B. A. and Lukes, P. and Luo, Y. and Lussu, M. and Ma, S. and Maciel Rabelo Pereira,
845 C. and Mack, M. and Maire, V. and Makela, A. and Makinen, H. and Malhado, A. C. M. and Mallik, A.
846 and Manning, P. and Manzoni, S. and Marchetti, Z. and Marchino, L. and Marcilio-Silva, V. and
847 Marcon, E. and Marignani, M. and Markesteijn, L. and Martin, A. and Martinez-Garza, C. and Martinez-
848 Vilalta, J. and Maskova, T. and Mason, K. and Mason, N. and Massad, T. J. and Masse, J. and Mayrose,
849 I. and McCarthy, J. and McCormack, M. L. and McCulloh, K. and McFadden, I. R. and McGill, B. J.
850 and McPartland, M. Y. and Medeiros, J. S. and Medlyn, B. and Meerts, P. and Mehrabi, Z. and Meir, P.
851 and Melo, F. P. L. and Mencuccini, M. and Meredieu, C. and Messier, J. and Meszaros, I. and
852 Metsaranta, J. and Michaletz, S. T. and Michelaki, C. and Migalina, S. and Milla, R. and Miller, J. E. D.
853 and Minden, V. and Ming, R. and Mokany, K. and Moles, A. T. and Molnar, A. t. and Molofsky, J. and
854 Molz, M. and Montgomery, R. A. and Monty, A. and Moravcova, L. and Moreno-Martinez, A. and
855 Moretti, M. and Mori, A. S. and Mori, S. and Morris, D. and Morrison, J. and Mucina, L. and Mueller,
856 S. and Muir, C. D. and Muller, S. C. and Munoz, F. and Myers-Smith, I. H. and Myster, R. W. and
857 Nagano, M. and Naidu, S. and Narayanan, A. and Natesan, B. and Negoita, L. and Nelson, A. S. and
858 Neuschulz, E. L. and Ni, J. and Niedrist, G. and Nieto, J. and Niinemets, U. and Nolan, R. and
859 Nottebrock, H. and Nouvellon, Y. and Novakovskiy, A. and Nutrient, N. and Nystuen, K. O. and



860 O'Grady, A. and O'Hara, K. and O'Reilly-Nugent, A. and Oakley, S. and Oberhuber, W. and Ohtsuka,
861 T. and Oliveira, R. and Ollerer, K. and Olson, M. E. and Onipchenko, V. and Onoda, Y. and Onstein,
862 R. E. and Ordonez, J. C. and Osada, N. and Ostonen, I. and Ottaviani, G. and Otto, S. and Overbeck, G.
863 E. and Ozinga, W. A. and Pahl, A. T. and Paine, C. E. T. and Pakeman, R. J. and Papageorgiou, A. C.
864 and Parfionova, E. and Partel, M. and Patacca, M. and Paula, S. and Paule, J. and Pauli, H. and Pausas,
865 J. G. and Peco, B. and Penuelas, J. and Perea, A. and Peri, P. L. and Petisco-Souza, A. C. and Petraglia,
866 A. and Petritan, A. M. and Phillips, O. L. and Pierce, S. and Pillar, V. D. and Pisek, J. and Pomogaybin,
867 A. and Poorter, H. and Portsmouth, A. and Poschlod, P. and Potvin, C. and Pounds, D. and Powell, A. S.
868 and Power, S. A. and Prinzing, A. and Puglielli, G. and Pysek, P. and Raavel, V. and Rammig, A. and
869 Ransijn, J. and Ray, C. A. and Reich, P. B. and Reichstein, M. and Reid, D. E. B. and Rejou-Mechain,
870 M. and de Dios, V. R. and Ribeiro, S. and Richardson, S. and Riibak, K. and Rillig, M. C. and Riviera,
871 F. and Robert, E. M. R. and Roberts, S. and Robroek, B. and Roddy, A. and Rodrigues, A. V. and
872 Rogers, A. and Rollinson, E. and Rolo, V. and Romermann, C. and Ronzhina, D. and Roscher, C. and
873 Rosell, J. A. and Rosenfield, M. F. and Rossi, C. and Roy, D. B. and Royer-Tardif, S. and Ruger, N. and
874 Ruiz-Peinado, R. and Rumpf, S. B. and Rusch, G. M. and Ryo, M. and Sack, L. and Saldana, A. and
875 Salgado-Negret, B. and Salguero-Gomez, R. and Santa-Regina, I. and Santacruz-Garcia, A. C. and
876 Santos, J. and Sardans, J. and Schamp, B. and Scherer-Lorenzen, M. and Schleuning, M. and Schmid,
877 B. and Schmidt, M. and Schmitt, S. and Schneider, J. V. and Schowanek, S. D. and Schrader, J. and
878 Schrodt, F. and Schuldt, B. and Schurr, F. and Selaya Garvizu, G. and Semchenko, M. and Seymour, C.
879 and Sfair, J. C. and Sharpe, J. M. and Sheppard, C. S. and Sheremetiev, S. and Shiodera, S. and Shipley,
880 B. and Shovon, T. A. and Siebenkas, A. and Sierra, C. and Silva, V. and Silva, M. and Sitzia, T. and
881 Sjomani, H. and Slot, M. and Smith, N. G. and Sodhi, D. and Soltis, P. and Soltis, D. and Somers, B. and
882 Sonnier, G. and Sorensen, M. V. and Sosinski, E. E., Jr. and Soudzilovskaia, N. A. and Souza, A. F. and
883 Spasojevic, M. and Sperandii, M. G. and Stan, A. B. and Stegen, J. and Steinbauer, K. and Stephan, J.
884 G. and Sterck, F. and Stojanovic, D. B. and Strydom, T. and Suarez, M. L. and Svenning, J. C. and
885 Svitkova, I. and Svitok, M. and Svoboda, M. and Swaine, E. and Swenson, N. and Tabarelli, M. and
886 Takagi, K. and Tappeiner, U. and Tarifa, R. and Tauougourdeau, S. and Tavsanoğlu, C. and Te Beest, M.
887 and Tedersoo, L. and Thiffault, N. and Thom, D. and Thomas, E. and Thompson, K. and Thornton, P.



888 E. and Thuiller, W. and Tichy, L. and Tissue, D. and Tjoelker, M. G. and Tng, D. Y. P. and Tobias, J.
889 and Torok, P. and Tarin, T. and Torres-Ruiz, J. M. and Tothmeresz, B. and Treurnicht, M. and
890 Trivellone, V. and Trolliet, F. and Trotsiuk, V. and Tsakalos, J. L. and Tsiripidis, I. and Tysklind, N.
891 and Umehara, T. and Usoltsev, V. and Vadeboncoeur, M. and Vaezi, J. and Valladares, F. and Vamosi,
892 J. and van Bodegom, P. M. and van Breugel, M. and Van Cleemput, E. and van de Weg, M. and van der
893 Merwe, S. and van der Plas, F. and van der Sande, M. T. and van Kleunen, M. and Van Meerbeek, K.
894 and Vanderwel, M. and Vanselow, K. A. and Varhammar, A. and Varone, L. and Vasquez Valderrama,
895 M. Y. and Vassilev, K. and Vellend, M. and Veneklaas, E. J. and Verbeeck, H. and Verheyen, K. and
896 Vibrans, A. and Vieira, I. and Villacis, J. and Violle, C. and Vivek, P. and Wagner, K. and Waldram,
897 M. and Waldron, A. and Walker, A. P. and Waller, M. and Walther, G. and Wang, H. and Wang, F. and
898 Wang, W. and Watkins, H. and Watkins, J. and Weber, U. and Weedon, J. T. and Wei, L. and Weigelt,
899 P. and Weiher, E. and Wells, A. W. and Wellstein, C. and Wenk, E. and Westoby, M. and Westwood,
900 A. and White, P. J. and Whitten, M. and Williams, M. and Winkler, D. E. and Winter, K. and Womack,
901 C. and Wright, I. J. and Wright, S. J. and Wright, J. and Pinho, B. X. and Ximenes, F. and Yamada, T.
902 and Yamaji, K. and Yanai, R. and Yankov, N. and Yguel, B. and Zanini, K. J. and Zanne, A. E. and
903 Zeleny, D. and Zhao, Y. P. and Zheng, J. and Zheng, J. and Zieminska, K. and Zirbel, C. R. and Zizka,
904 G. and Zo-Bi, I. C. and Zotz, G. and Wirth, C.: TRY plant trait database - enhanced coverage and open
905 access, *Global Change Biology*, 26, 119-188, doi:10.1111/gcb.14904, 2020.

906 Kilpeläinen, J., Barbero-López, A., Vestberg, M., Heiskanen, J., and Lehto, T.: Does severe soil drought
907 have after-effects on arbuscular and ectomycorrhizal root colonisation and plant nutrition?, *Plant and*
908 *Soil*, 418, 377-386, doi:10.1007/s11104-017-3308-8, 2017.

909 Kirkham, M. B., Hillel, D. (Ed.): *Water-use efficiency*, *Encyclopedia of Soils in the Environment*
910 Elsevier, Oxford2005.

911 Kleyer, M., Bekker, R. M., Knevel, I. C., Bakker, J. P., Thompson, K., Sonnenschein, M., Poschlod, P.,
912 van Groenendael, J. M., Klimeš, L., Klimešová, J., Klotz, S., Rusch, G. M., Hermy, M., Adriaens, D.,
913 Boedeltje, G., Bossuyt, B., Dannemann, A., Endels, P., Götzenberger, L., Hodgson, J. G., Jackel, A. K.,
914 Kühn, I., Kunzmann, D., Ozinga, W. A., Römermann, C., Stadler, M., Schlegelmilch, J., Steendam, H.



- 915 J., Tackenberg, O., Wilmann, B., Cornelissen, J. H. C., Eriksson, O., Garnier, E., and Peco, B.: The
916 LEDA Traitbase: a database of life-history traits of the Northwest European flora, *Journal of Ecology*,
917 96, 1266-1274, doi:10.1111/j.1365-2745.2008.01430.x, 2008.
- 918 Knot, P., Skladanka, J., Hrabě, F., Slama, P., Nawrath, A., and Kvasnovsky, M.: Changes in the species
919 diversity of grassland communities during secondary succession, *Bulgarian Journal of Agricultural
920 Science*, 21, 822–828, 2015.
- 921 Komainda, M., Küchenmeister, K., Küchenmeister, F., Breitsameter, L., Wrage-Mönnig, N., Kayser,
922 M., and Isselstein, J.: Forage legumes for future dry climates: Lower relative biomass losses of minor
923 forage legumes compared to *Trifolium repens* under conditions of periodic drought stress, *Journal of
924 Agronomy and Crop Science*, 205, 460-469, doi:10.1111/jac.12337, 2019.
- 925 Kruger, C., Kohout, P., Janouskova, M., Puschel, D., Frouz, J., and Rydlova, J.: Plant communities
926 rather than soil properties structure arbuscular mycorrhizal fungal communities along primary
927 succession on a mine spoil, *Frontiers in Microbiology*, 8, 719, doi:10.3389/fmicb.2017.00719, 2017.
- 928 Lavorel, S. and Grigulis, K.: How fundamental plant functional trait relationships scale-up to trade-offs
929 and synergies in ecosystem services, *Journal of Ecology*, 100, 128-140, doi:10.1111/j.1365-
930 2745.2011.01914.x, 2012.
- 931 Ledgard, S. F. and Steele, K. W.: Biological nitrogen fixation in mixed legume/grass pastures, *Plant and
932 Soil*, 141, 137-153, doi:10.1007/BF00011314, 1992.
- 933 Leitinger, G., Ruggenthaler, R., Hammerle, A., Lavorel, S., Schirpke, U., Clement, J. C., Lamarque, P.,
934 Obojes, N., and Tappeiner, U.: Impact of droughts on water provision in managed alpine grasslands in
935 two climatically different regions of the Alps, *Ecology*, 8, 1600-1613, doi:10.1002/eco.1607,
936 2015.
- 937 Li, W., Wang, L., Qian, S., He, M., Cai, X., and Ding, J.: Root characteristics explain greater water use
938 efficiency and drought tolerance in invasive *Compositae* plants, *Plant and Soil*, 483, 209-223,
939 doi:10.1007/s11104-022-05734-5, 2022.



- 940 Lozano, Y. M., Aguilar-Trigueros, C. A., Flaig, I. C., and Rillig, M. C.: Root trait responses to drought
941 are more heterogeneous than leaf trait responses, *Functional Ecology*, 34, 2224-2235, doi:10.1111/1365-
942 2435.13656, 2020.
- 943 Lu, Y., Duursma, R. A., Farrior, C. E., Medlyn, B. E., and Feng, X.: Optimal stomatal drought response
944 shaped by competition for water and hydraulic risk can explain plant trait covariation, *New Phytologist*,
945 225, 1206-1217, doi:10.1111/nph.16207, 2020.
- 946 Lubbe, F. C., Bitomský, M., Hajek, T., de Bello, F., Doležal, J., Jandová, V., Janeček, Š., Bartušková,
947 A., Lanta, V., and Klimešová, J.: A tale of two grasslands: how belowground storage organs coordinate
948 their traits with water-use traits, *Plant and Soil*, 465, 533-548, doi:10.1007/s11104-021-05031-7, 2021.
- 949 Ma, Z., Guo, D., Xu, X., Lu, M., Bardgett, R. D., Eissenstat, D. M., McCormack, M. L., and Hedin, L.
950 O.: Evolutionary history resolves global organization of root functional traits, *Nature*, 555, 94-97,
951 doi:10.1038/nature25783, 2018.
- 952 Mahdavi, S. M. E., Salehi, H., and Zarei, M.: Can arbuscular mycorrhizal fungi ameliorate the adverse
953 effects of deficit irrigation on tall fescue (*Festuca arundinacea* Schreb.)?, *Journal of Soil Science and*
954 *Plant Nutrition*, 0-0, doi:10.4067/s0718-95162018005001902, 2018.
- 955 Maitra, P., Zheng, Y., Chen, L., Wang, Y.-L., Ji, N.-N., Lü, P.-P., Gan, H.-Y., Li, X.-C., Sun, X., Zhou,
956 X.-H., and Guo, L.-D.: Effect of drought and season on arbuscular mycorrhizal fungi in a subtropical
957 secondary forest, *Fungal Ecology*, 41, 107-115, doi:10.1016/j.funeco.2019.04.005, 2019.
- 958 Malinowski, D.: Rhizomatous ecotypes and symbiosis with endophytes as new possibilities of
959 improvement in competitive ability of meadow fescue (*Festuca pratensis* Huds.). Doctoral Thesis, Swiss
960 Federal Institute of Technology, Zürich, doi:10.3929/ethz-a-001575829, 1995.
- 961 Manzoni, S., Vico, G., Palmroth, S., Porporato, A., and Katul, G.: Optimization of stomatal conductance
962 for maximum carbon gain under dynamic soil moisture, *Advances in Water Resources*, 62, 90-105,
963 doi:10.1016/j.advwatres.2013.09.020, 2013.



- 964 Martin-StPaul, N., Delzon, S., and Cochard, H.: Plant resistance to drought depends on timely stomatal
965 closure, *Ecology Letters*, 20, 1437-1447, doi:10.1111/ele.12851, 2017.
- 966 Mason, N. W., Orwin, K., Lambie, S., Woodward, S. L., McCreedy, T., and Mudge, P.: Leaf economics
967 spectrum-productivity relationships in intensively grazed pastures depend on dominant species identity,
968 *Ecology and Evolution*, 6, 3079-3091, doi:10.1002/ece3.1964, 2016.
- 969 McCormack, M. L. and Iversen, C. M.: Physical and functional constraints on viable belowground
970 acquisition strategies, *Frontiers in Plant Science*, 10, 1215, doi:10.3389/fpls.2019.01215, 2019.
- 971 Meziane, D. and Shipley, B.: Direct and indirect relationships between specific leaf area, leaf nitrogen
972 and leaf gas exchange. Effects of irradiance and nutrient supply, *Annals of Botany*, 88, 915-927,
973 doi:10.1006/anbo.2001.1536, 2001.
- 974 Miller, J. E. D., Li, D., LaForgia, M., and Harrison, S.: Functional diversity is a passenger but not driver
975 of drought-related plant diversity losses in annual grasslands, *Journal of Ecology*, 107, 2033-2039,
976 doi:10.1111/1365-2745.13244, 2019.
- 977 Murray, L. K.: Early development characteristics of different clover species during establishment,
978 Lincoln University, Christchurch, New Zealand, 2012.
- 979 Oram, N. J., Ingrisch, J., Bardgett, R. D., Brennan, F., Dittmann, G., Gleixner, G., Illmer, P., Praeg, N.,
980 and Bahn, M.: Drought intensity alters productivity, carbon allocation and plant nitrogen uptake in fast
981 versus slow grassland communities, *Journal of Ecology*, doi:10.1111/1365-2745.14136, 2023.
- 982 Ordoñez, J. C., van Bodegom, P. M., Witte, J. P., Bartholomeus, R. P., van Hal, J. R., and Aerts, R.:
983 Plant strategies in relation to resource supply in mesic to wet environments: does theory mirror nature?,
984 *American Naturalist*, 175, 225-239, doi:10.1086/649582, 2010.
- 985 Ostonen, I., Püttsepp, Ü., Biel, C., Alberton, O., Bakker, M. R., Löhmus, K., Majdi, H., Metcalfe, D.,
986 Olsthoorn, A. F. M., Pronk, A., Vanguelova, E., Weih, M., and Brunner, I.: Specific root length as an



- 987 indicator of environmental change, *Plant Biosystems - An International Journal Dealing with all Aspects*
988 of *Plant Biology*, 141, 426-442, doi:10.1080/11263500701626069, 2007.
- 989 Pande, H. and Singh, J. S.: Comparative biomass and water status of four range grasses grown under
990 two soil water conditions, *Society for Range Management*, doi:10.2307/3898103, 1981.
- 991 Parniske, M.: Arbuscular mycorrhiza: the mother of plant root endosymbioses, *Nature Reviews*
992 *Microbiology*, 6, 763-775, doi:10.1038/nrmicro1987, 2008.
- 993 Paterson, E., Thornton, B., Midwood, A. J., and Sim, A.: Defoliation alters the relative contributions of
994 recent and non-recent assimilate to root exudation from *Festuca rubra*, *Plant, Cell & Environment*, 28,
995 1525-1533, doi:10.1111/j.1365-3040.2005.01389.x, 2005.
- 996 Peralta, A. M. L., Sánchez, A. M., Luzuriaga, A. L., Bello, F., and Escudero, A.: Evidence of functional
997 species sorting by rainfall and biotic interactions: A community monolith experimental approach,
998 *Journal of Ecology*, 107, 2772-2788, doi:10.1111/1365-2745.13210, 2019.
- 999 Perez-Ramos, I. M., Matias, L., Gomez-Aparicio, L., and Godoy, O.: Functional traits and phenotypic
1000 plasticity modulate species coexistence across contrasting climatic conditions, *Nature Communications*,
1001 10, 2555, doi:10.1038/s41467-019-10453-0, 2019.
- 1002 Pérez-Ramos, I. M., Volaire, F., Fattet, M., Blanchard, A., and Roumet, C.: Tradeoffs between
1003 functional strategies for resource-use and drought-survival in Mediterranean rangeland species,
1004 *Environmental and Experimental Botany*, 87, 126-136, doi:10.1016/j.envexpbot.2012.09.004, 2013.
- 1005 Peters, A., Nehls, T., and Wessolek, G.: Technical note: Improving the AWAT filter with interpolation
1006 schemes for advanced processing of high resolution data, *Hydrology and Earth System Sciences*, 20,
1007 2309-2315, doi:10.5194/hess-20-2309-2016, 2016.
- 1008 Peters, A., Nehls, T., Schonsky, H., and Wessolek, G.: Separating precipitation and evapotranspiration
1009 from noise – a new filter routine for high-resolution lysimeter data, *Hydrology and Earth System*
1010 *Sciences*, 18, 1189-1198, doi:10.5194/hess-18-1189-2014, 2014.



- 1011 Peters, A., Groh, J., Schrader, F., Durner, W., Vereecken, H., and Pütz, T.: Towards an unbiased filter
1012 routine to determine precipitation and evapotranspiration from high precision lysimeter measurements,
1013 *Journal of Hydrology*, 549, 731-740, doi:10.1016/j.jhydrol.2017.04.015, 2017.
- 1014 Philippot, S., Allirand, J. M., Cartier, M., and Gosse, G.: The role of different daily irradiations on shoot
1015 growth and root/shoot ratio in lucerne (*Medicago sativa* L.), *Annals of Botany*, 68, 329-335,
1016 doi:10.1093/oxfordjournals.aob.a088260, 1991.
- 1017 Potsch, M. E.: Kreislaufwirtschaft im Alpenländischen Grünlandbetrieb, Ökosoziales Forum und BAL
1018 Gumpenstein, Vienna1999.
- 1019 Prieto, I., Litrico, I., Violle, C., and Barre, P.: Five species, many genotypes, broad phenotypic diversity:
1020 When agronomy meets functional ecology, *American Journal of Botany*, 104, 62-71,
1021 doi:10.3732/ajb.1600354, 2017.
- 1022 Quan, Q., Zhang, F., Meng, C., Ma, F., Zhou, Q., Sun, F., and Niu, S.: Shifting biomass allocation
1023 determines community water use efficiency under climate warming, *Environmental Research Letters*,
1024 15, 094041, doi:10.1088/1748-9326/aba472, 2020.
- 1025 Querejeta, J. I., Prieto, I., Torres, P., Campoy, M., Alguacil, M. M., and Roldán, A.: Water-spender
1026 strategy is linked to higher leaf nutrient concentrations across plant species colonizing a dry and nutrient-
1027 poor epiphytic habitat, *Environmental and Experimental Botany*, 153, 302-310,
1028 doi:10.1016/j.envexpbot.2018.06.007, 2018.
- 1029 R Core Team: R: A language and environment for statistical computing. R Foundation for Statistical
1030 Computing [code], 2022.
- 1031 Reich, P. B.: The world-wide ‘fast-slow’ plant economics spectrum: a traits manifesto, *Journal of*
1032 *Ecology*, 102, 275-301, doi:10.1111/1365-2745.12211, 2014.
- 1033 Reich, P. B., Buschena, C., Tjoelker, M. G., Wrage, K., Knops, J., Tilman, D., and Machado, J. L.:
1034 Variation in growth rate and ecophysiology among 34 grassland and savanna species under contrasting



- 1035 N supply: a test of functional group differences, *New Phytologist*, 157, 617-631, doi:10.1046/j.1469-
1036 8137.2003.00703.x, 2003.
- 1037 Reichstein, M., Bahn, M., Ciais, P., Frank, D., Mahecha, M. D., Seneviratne, S. I., Zscheischler, J., Beer,
1038 C., Buchmann, N., Frank, D. C., Papale, D., Rammig, A., Smith, P., Thonicke, K., van der Velde, M.,
1039 Vicca, S., Walz, A., and Wattenbach, M.: Climate extremes and the carbon cycle, *Nature*, 500, 287-295,
1040 doi:10.1038/nature12350, 2013.
- 1041 Roumet, C., Birouste, M., Picon-Cochard, C., Ghestem, M., Osman, N., Vrignon-Brenas, S., Cao, K. F.,
1042 and Stokes, A.: Root structure-function relationships in 74 species: evidence of a root economics
1043 spectrum related to carbon economy, *New Phytologist*, 210, 815-826, doi:10.1111/nph.13828, 2016.
- 1044 Ryser, P. and Eek, L.: Consequences of phenotypic plasticity vs. interspecific differences in leaf and
1045 root traits for acquisition of aboveground and belowground resources, *American Journal of Botany*, 87,
1046 402-411, doi:10.2307/2656636, 2000.
- 1047 Sanderson, M. A., Brink, G., Stout, R., and Ruth, L.: Grass–legume proportions in forage seed mixtures
1048 and effects on herbage yield and weed abundance, *Agronomy Journal*, 105, 1289-1297,
1049 doi:10.2134/agronj2013.0131, 2013.
- 1050 Schübler, A. and Walker, C.: *The Glomeromycota. A species list with new families and new genera.*,
1051 Gloucester, England 2010.
- 1052 Schwamback, D., Persson, M., Berndtsson, R., Bertotto, L. E., Kobayashi, A. N. A., and Wendland, E.
1053 C.: Automated low-cost soil moisture sensors: trade-off between cost and accuracy, *Sensors (Basel)*, 23,
1054 10.3390/s23052451, 2023.
- 1055 Shahzad, T., Chenu, C., Repinçay, C., Mougín, C., Ollier, J.-L., and Fontaine, S.: Plant clipping
1056 decelerates the mineralization of recalcitrant soil organic matter under multiple grassland species, *Soil
1057 Biology and Biochemistry*, 51, 73-80, doi:10.1016/j.soilbio.2012.04.014, 2012.



- 1058 Silva-Flores, P., Bueno, C. G., Neira, J., and Palfner, G.: Factors affecting arbuscular mycorrhizal fungi
1059 spore density in the Chilean Mediterranean-type ecosystem, *Journal of Soil Science and Plant Nutrition*,
1060 19, 42-50, doi:10.1007/s42729-018-0004-6, 2019.
- 1061 Spehn, E. M., Scherer-Lorenzen, M., Schmid, B., Hector, A., Caldeira, M. C., Dimitrakopoulos, P. G.,
1062 Finn, J. A., Jumpponen, A., O'Donovan, G., Pereira, J. S., Schulze, E. D., Troumbis, A. Y., and Körner,
1063 C.: The role of legumes as a component of biodiversity in a cross-European study of grassland biomass
1064 nitrogen, *Oikos*, 98, 205-218, doi:10.1034/j.1600-0706.2002.980203.x, 2002.
- 1065 Sun, K., McCormack, M. L., Li, L., Ma, Z., and Guo, D.: Fast-cycling unit of root turnover in perennial
1066 herbaceous plants in a cold temperate ecosystem, *Scientific Reports*, 6, 19698, doi:10.1038/srep19698,
1067 2016.
- 1068 Sun, X., Shi, J., and Ding, G.: Combined effects of arbuscular mycorrhiza and drought stress on plant
1069 growth and mortality of forage sorghum, *Applied Soil Ecology*, 119, 384-391,
1070 doi:10.1016/j.apsoil.2017.07.030, 2017.
- 1071 Talgre, L., Roostalu, H., Mäeorg, E., and Lauringson, E.: Nitrogen and carbon release during
1072 decomposition of roots and shoots of leguminous green manure crops, *Agronomy Research*, 15, 594-
1073 601, 2017.
- 1074 Tardy, F., Damour, G., Dorel, M., and Moreau, D.: Trait-based characterisation of soil exploitation
1075 strategies of banana, weeds and cover plant species, *PLoS One*, 12, e0173066,
1076 doi:10.1371/journal.pone.0173066, 2017.
- 1077 Tasser, E. and Tappeiner, U.: New model to predict rooting in diverse plant community compositions,
1078 *Ecological Modelling*, 185, 195-211, doi:10.1016/j.ecolmodel.2004.11.024, 2005.
- 1079 Tello-García, E., Huber, L., Leitinger, G., Peters, A., Newesely, C., Ringler, M.-E., and Tasser, E.:
1080 Drought- and heat-induced shifts in vegetation composition impact biomass production and water use
1081 of alpine grasslands, *Environmental and Experimental Botany*, 169, 103921,
1082 doi:10.1016/j.envexpbot.2019.103921, 2020.



- 1083 Thakur, M. P., Risch, A. C., and van der Putten, W. H.: Biotic responses to climate extremes in terrestrial
1084 ecosystems, *iScience*, 25, 104559, doi:10.1016/j.isci.2022.104559, 2022.
- 1085 Tuo, X.-Q., He, L., and Zou, Y.-N.: Alleviation of drought stress in white clover after inoculation with
1086 arbuscular mycorrhizal fungi, *Notulae Botanicae Horti Agrobotanici Cluj-Napoca*, 45, 220-224,
1087 doi:10.15835/nbha45110709, 2017.
- 1088 Wang, P., Huang, K., and Hu, S.: Distinct fine-root responses to precipitation changes in herbaceous
1089 and woody plants: a meta-analysis, *New Phytologist*, 225, 1491-1499, doi:10.1111/nph.16266, 2020.
- 1090 Wei, X., Reich, P. B., and Hobbie, S. E.: Legumes regulate grassland soil N cycling and its response to
1091 variation in species diversity and N supply but not CO₂, *Global Change Biology*, 25, 2396-2409,
1092 doi:10.1111/gcb.14636, 2019.
- 1093 Wohlfahrt, G., Bahn, M., Haubner, E., Horak, I., Michaeler, W., Rottmar, K., Tappeiner, U., and
1094 Cernusca, A.: Inter-specific variation of the biochemical limitation to photosynthesis and related leaf
1095 traits of 30 species from mountain grassland ecosystems under different land use, *Plant, Cell &*
1096 *Environment*, 22, 1281-1296, doi:10.1046/j.1365-3040.1999.00479.x, 1999.
- 1097 Wright, I. J., Reich, P. B., Westoby, M., Ackerly, D. D., Baruch, Z., Bongers, F., Cavender-Bares, J.,
1098 Chapin, T., Cornelissen, J. H., Diemer, M., Flexas, J., Garnier, E., Groom, P. K., Gulias, J., Hikosaka,
1099 K., Lamont, B. B., Lee, T., Lee, W., Lusk, C., Midgley, J. J., Navas, M. L., Niinemets, U., Oleksyn, J.,
1100 Osada, N., Poorter, H., Poot, P., Prior, L., Pyankov, V. I., Roumet, C., Thomas, S. C., Tjoelker, M. G.,
1101 Veneklaas, E. J., and Villar, R.: The worldwide leaf economics spectrum, *Nature*, 428, 821-827,
1102 doi:10.1038/nature02403, 2004.
- 1103 Zafari, M., Ebadi, A., and Jahanbakhsh Godehkahriz, S.: Effect of seed inoculation on alfalfa tolerance
1104 to water deficit stress, *Notulae Botanicae Horti Agrobotanici Cluj-Napoca*, 45, 82-88,
1105 doi:10.15835/nbha45110424, 2017.



- 1106 Zhang, C., Shi, S., Wang, B., and Zhao, J.: Physiological and biochemical changes in different drought-
1107 tolerant alfalfa (*Medicago sativa* L.) varieties under PEG-induced drought stress, *Acta Physiologiae*
1108 *Plantarum*, 40, doi:10.1007/s11738-017-2597-0, 2018a.
- 1109 Zhang, T., Hu, Y., Zhang, K., Tian, C., and Guo, J.: Arbuscular mycorrhizal fungi improve plant growth
1110 of *Ricinus communis* by altering photosynthetic properties and increasing pigments under drought and
1111 salt stress, *Industrial Crops and Products*, 117, 13-19, doi:10.1016/j.indcrop.2018.02.087, 2018b.
- 1112 Zhao, J., Xu, T., Xiao, J., Liu, S., Mao, K., Song, L., Yao, Y., He, X., and Feng, H.: Responses of water
1113 use efficiency to drought in Southwest China, *Remote Sensing*, 12, 199, doi:10.3390/rs12010199, 2020.
- 1114 Zhao, R., Guo, W., Bi, N., Guo, J., Wang, L., Zhao, J., and Zhang, J.: Arbuscular mycorrhizal fungi
1115 affect the growth, nutrient uptake and water status of maize (*Zea mays* L.) grown in two types of coal
1116 mine spoils under drought stress, *Applied Soil Ecology*, 88, 41-49, doi:10.1016/j.apsoil.2014.11.016,
1117 2015.
- 1118 Zhao, W., Liu, B., Chang, X., Yang, Q., Yang, Y., Liu, Z., Cleverly, J., and Eamus, D.:
1119 Evapotranspiration partitioning, stomatal conductance, and components of the water balance: A special
1120 case of a desert ecosystem in China, *Journal of Hydrology*, 538, 374-386,
1121 doi:10.1016/j.jhydrol.2016.04.042, 2016.
- 1122 Zhou, L., Hong, Y., Li, C., Lu, C., He, Y., Shao, J., Sun, X., Wang, C., Liu, R., Liu, H., Zhou, G., and
1123 Zhou, X.: Responses of biomass allocation to multi-factor global change: A global synthesis,
1124 *Agriculture, Ecosystems & Environment*, 304, 107115, doi:10.1016/j.agee.2020.107115, 2020.
- 1125