1	
2	
3	Eddy Covariance Evaluation of Ecosystem Fluxes at a Temperate Saltmarsh in
4	Victoria, Australia Shows Large CO ₂ Uptake
5	
6	Authors
7	
8	Ruth Reef ¹ ,
9	Edoardo Daly ^{2,3} ,
10	Tivanka Anandappa ¹ ,
11	Eboni-Jane Vienna-Hallam ¹ ,
12	Harriet Robertson ¹ ,
13	Matthew Peck ¹ ,
14	Adrien Guyot ^{4,5}
15	
16	Affiliations
17	
18	1 School of Earth, Atmosphere and Environment, Monash University, VIC 3800, Australia
19	2 Department of Civil Engineering, Monash University, VIC 3800, Australia
20	3 WMAwater, Brisbane, QLD 4000, Australia
21	4 Atmospheric Observations Research Group, The University of Queensland, Brisbane,
22	Australia
23	5 Australian Bureau of Meteorology, Melbourne, Australia
24	
25	Corresponding Author
26	
27	Associate Professor Ruth Reef
28	School of Earth Atmosphere and Environment
29	Monash University
30	9 Rainforest Walk, Clayton VIC 3800
31	Australia
32	Email: <u>ruth.reef@monash.edu</u>
33	Ph: +61 3 9905 8309
34	

T: 41

35			
36	Key Points		
37			
38	This is the first study using eddy covariance to measure CO ₂ fluxes at an Australian		
39	temperate saltmarsh, revealing temperature and light limitations to CO2 uptake.		
40			
41	CO_2 fluxes varied seasonally; growing season net ecosystem productivity was 10.54 g CO_2		
42	m ⁻² day ⁻¹ , dropping to 1.64 g CO ₂ m ⁻² day ⁻¹ in winter.		
43			
44	Productivity at the French Island saltmarsh is high relative to global saltmarsh estimates but		Deleted: Annual
45	below global mangrove averages.		Deleted: p
46			Deleted: estimated at 753 g C m ⁻² y ⁻¹ , surpassing
47			
48			
49	Abstract		
50			
51	Recent studies highlight the important role of vegetated coastal ecosystems in atmospheric		
52	carbon sequestration. Saltmarshes constitute 30% of these ecosystems globally and are the		
53	primary intertidal coastal wetland habitat outside the tropics. Eddy covariance (EC) is the		Deleted: vegetation
54	main method for measuring biosphere-atmosphere fluxes, but its use in coastal environments		
55	is rare. At an Australian temperate saltmarsh site on French Island, Victoria, we measured		
56	CO_2 and water gas concentration gradients, temperature, wind speed and radiation. The		
57	marsh was dominated by a dense cover of Sarcocornia quinqueflora. Fluxes were seasonal,		
58	with minima in winter when vegetation is dormant. Net ecosystem productivity (NEP) during		
59	the growing season averaged 10.54 g CO ₂ m- ² day- ¹ decreasing to 1.64 g CO ₂ m- ² day- ¹ in		
60	the dormant period, yet the marsh remained a CO ₂ sink due to some sempervirent species.		
61	Ecosystem respiration rates were lower during the dormant period compared with the		
62	growing season (1.00 vs 1.77 $\mu mol~CO_2~m^{-2}~s^{-1})$ with a slight positive relationship with		
63	temperature. During the growing season, fluxes were significantly influenced by light levels,	/	Deleted: During the growing season, fluxes were significantly influenced by light levels, ambient temperatures
64	ambient temperatures and humidity with cool temperatures and cloud cover limiting NEP.		and humidity. Evapotranspiration peaked at 0.27 mm h^{-1} . We cautiously estimate the annual NEP budget at this marsh to be
65	Ecosystem water use efficiency of 0.86 g C kg ⁻¹ H ₂ O was similar to other C3 intertidal		753 (\pm 112.7) g C m ⁻² y ⁻¹ which is similar to carbon uptake by temperate saltmarshes in Europe and within the range
66	marshes and evapotranspiration averaged 2.48 mm day ⁻¹ during the growing season.	/	measured at some US saltmarshes. This value is higher than the value hypothesised for global saltmarshes of 382 g C m^{-2}
67			y ⁻¹ but is only half the mean value estimated for global mangroves.
68	EGUsphere Topics		Deleted: ¶

84 Emissions, Marine and Freshwater Biogeosciences, Earth System Biogeosciences

85

86 Short Summary

87

88 Studies show that saltmarshes excel at capturing carbon from the atmosphere. In this study,

89 we measured CO₂ flux in an Australian temperate saltmarsh on French Island. The temperate

90 saltmarsh exhibited strong seasonality. During the warmer growing season, the saltmarsh

91 absorbed on average 10.5 grams of CO_2 from the atmosphere per m² daily. Even in winter,

92 when plants were dormant, it continued to be a CO₂ sink, albeit smaller. Cool temperatures

- 93 and high cloud cover inhibit carbon sequestration.
- 94
- 95

96

97

98 1. Introduction 99 100 Despite their relatively small global footprint of 54,650 km² (Mcowen et al., 2017), salt 101 marshes provide a range of ecosystem services, including shoreline protection (Shepard et al., 102 2011), nutrient uptake, nursery grounds for fish populations (Whitfield, 2017) as well as 103 functioning as significant carbon sinks through CO2 uptake and storage in their organic rich 104 sediments (McLeod et al., 2011). These 'blue carbon' habitats are recognised for their 105 significant contribution to the global carbon cycle, as coastal wetlands more broadly are 106 estimated to have accumulated more than a quarter of global organic soil carbon (Duarte, 107 2017). 108 109 Saltmarshes are a widely distributed intertidal habitat but are floristically divergent globally 110 (Adam, 2002), such that commonalities in function and form do not extend across 111 biogeographic realms. US saltmarshes, for example, are extensively dominated by a single 112 grassy species, Spartina alterniflora, as opposed to the dominance of C3 Chenopodioideae 113 species in the southern hemisphere (Adam, 2002). Temperate saltmarshes occupy a latitudinal range spanning from approximately 30° to 60° (Mcowen et al., 2017) and are most 114 115 commonly found along protected coastlines such as bays, estuaries, and lagoons, where they 116 are sheltered from the full force of wave action (Mitsch and Gosselink, 2000). In the 117 Southern Hemisphere, temperate saltmarshes have a strong Gondwanan element with high 118 floristic similarity among the marshes of New Zealand, the southernmost coasts of South 119 America and South Africa and the southern coastlines of Australia (Adam, 1990). These 120 marshes are often associated with extensive seagrass meadows and mudflats, and in parts of 121 their range, mangroves, forming complex coastal mosaics (Huxham et al., 2018). 122 Saltmarshes have been heavily degraded across their range, and it is estimated that perhaps 123 up to 50% of the global saltmarsh area has been lost since 1900 (Gedan et al., 2009), 124 primarily due to land use change. 125

126 In most areas where they occur, seasonality plays a major role in the functioning of temperate

- 127 saltmarshes (Ghosh and Mishra, 2017). These ecosystems experience distinct growing and
- dormant seasons, primarily driven by temperature, light availability, and precipitation 128
- 129 patterns (Adam, 2000). During the growing season (typically spring and summer), increased
- 130 temperatures and longer daylight hours stimulate plant growth, photosynthetic activity, and

Deleted: Seasonality

132	decomposition processes. Photosynthesis typically outpaces decomposition during this	
133	period, resulting in the temperate saltmarsh acting as a net CO ₂ sink (Chmura et al., 2003).	
134	Conversely, the dormant season (usually fall and winter) is characterized by cooler	
135	temperatures and shorter days (Adam, 2000; Howe et al., 2010). These factors lead to	
136	reduced plant growth and photosynthetic activity (Adam, 2000) and while decomposition	
137	processes also slow down due to cooler temperatures, CO2 release through decomposition	
138	often exceeds CO ₂ uptake during this period (Artigas et al., 2015). In Australia, saltmarshes	
139	have been assumed to not exhibit seasonality (Owers et al., 2018) despite there being a	
140	scarcity of data on saltmarsh phenology and the implication this untested assumption could	
141	have on carbon budget estimations.	
142		
143	Gross primary production (GPP) of saltmarshes is the total amount of CO2 uptake by plants	
144	through photosynthesis. Respiration (Re) leads to a CO2 flux directed back to the atmosphere	E
145	due to all respiration processes occurring within the saltmarsh, involving both autotrophs and	
146	heterotrophs. The difference between these two fluxes is the net ecosystem exchange (NEE).	
147	Saltmarsh ecosystems can act as both sources and sinks of carbon dioxide (CO ₂), influencing	
148	atmospheric CO ₂ concentrations (Chmura et al., 2003). However, quantifying their net	
149	exchange remains challenging (Lu et al., 2017) hindering their effective inclusion in Earth	
150	System Models (Ward et al., 2020) and confounding the incorporation of saltmarsh	
151	restoration in emission reduction targets. Eddy covariance (EC) provides a powerful method	
152	for near-continuous, high-frequency monitoring of gas exchange between a vegetated surface	
153	and the atmosphere (Baldocchi, 2003), enabling the determination of net ecosystem exchange	
154	(NEE) of CO ₂ , and identifying the forcings that determine how CO ₂ fluxes will respond to	
155	global climate change (Borges et al., 2006; Cai, 2011).	
156		
157	Previous EC studies in coastal saltmarshes have been focused on the Northern Hemisphere, in	
158	sites in the USA (e.g. Hill and Vargas, 2022; Kathilankal et al., 2008; Moffett et al., 2010;	
159	Nahrawi et al., 2020; Schäfer et al., 2019), France (Mayen et al., 2024), Japan (Otani and	
160	Endo, 2019) and China (Wei et al., 2020) but interest in the southern hemisphere is growing	
161	(Bautista et al., 2023). The NEE values from these studies indicate that there is high inter-site	
162	(as well as interannual, Erickson et al., (2013)) variability in carbon dynamics of saltmarshes,	
163	with a link to species types, salinity, hydrology (Moffett et al., 2010; Nahrawi et al., 2020),	
164	site specific biochemical conditions (Seyfferth et al., 2020) and latitude (Feagin et al., 2020).	
165	While generally considered important carbon sinks (e.g. ranging between 130 to 775 g C $\mathrm{m}^{\text{-2}}$	

Deleted: photosynthetic flux

Deleted: CO_2 from the atmosphere to the land surface, while respiration

Deleted: are limited to

Deleted:).

171	yr ⁻¹ in the USA, according to Kathilankal et al. (2008) and Wang et al,(2016) respectively)	
172	and globally hypothesised to average 382 g C $m^{\text{-2}}$ y $^{\text{-1}}$ (Alongi, 2020), some EC studies	
173	revealed saltmarshes to be net sources of CO2 to the atmosphere (Vázquez-Lule and Vargas,	
174	2021) especially in temperate saltmarshes that experience long dormant periods.	
175		
176	The aim of this study is to estimate CO ₂ and water fluxes in a temperate saltmarsh in	
177	Victoria, southern Australia, to better characterise the effect of seasonality and environmental	
178	variables on the saltmarsh CO2 budgets. This is the first study in an Australian coastal	
179	saltmarsh where CO ₂ fluxes are estimated using the EC method.	
180		
181	2. Methods	
182		
183	2.82 Site Description	Deleted: 2.1
184		Formatted: List Paragraph, Outline numbered + Level: 2 Numbering Style: 1, 2, 3, + Start at: 82 + Alignment: I
185	Ecosystem flux measurements were collected at the Tortoise Head Ramsar coastal wetland on	+ Aligned at: 0.63 cm + Indent at: 1.27 cm
186	French Island, Victoria (38.388°S, 145.278°E, Fig. 1) within the Western Port embayment.	
187	French Island is within the Cfb climate zone (temperate oceanic climate) and experiences	
188	distinct seasonal variations in temperature and precipitation. Long term (30 year) climate data	
189	averaged from the nearby Cerberus Station (Australian Bureau of Meteorology, site 86361)	
190	indicated that summers, spanning from December through February, are generally mild to	
191	warm, with maximum temperatures typically ranging from 17°C to 25°C although occasional	
192	heatwaves lead to temporary spikes in temperature that can exceed 30°C. Winters, from June	
193	to September, are cooler, with maximum temperatures ranging between 7°C and 14°C and a	
194	mean minimum temperature of 6°C. Frost is infrequent due to maritime influence, though	
195	crisp mornings below 0°C occur 10% of the time in winter. Rainfall, evenly distributed	
196	throughout the year, averages ca. 715 mm y-1, although in 2020 the site Jed higher than	Deleted: experienc
197	average rainfall (860 mm y ⁻¹). The island is exposed to weather patterns influenced by the	
198	Southern Ocean and Bass Strait, leading to occasional storm systems, particularly in winter,	
199	bringing gusty winds and increased precipitation. Western Port has semi-diurnal tides with a	
200	range of nearly 3 m, resulting in wide intertidal flats occupied by mangroves of the species	
201	Avicennia marina and saltmarshes.	
202		





207 Figure 1: a) The location of French Island along the Bass Strait coast of Australia, and b) The location of the flux tower on French Island as well as the nearby Cerberus meteorological 208 209 station (Bureau of Meteorology, Australia), © Google Earth. c) An image of the saltmarsh 210 within the flux tower footprint during the growing season (with the tower and the author in the background), taken in February 2020 by Prudence Perry. d) an image of the saltmarsh 211 212 during the dormant period, taken at the same location in September 2020 by Ruth Reef. 213 214 The site at French Island is dominated by an extensive temperate coastal saltmarsh 215 community that is a particularly good natural representation of a broader biogeographic 216 saltmarsh grouping which covers an area of ca. 7000 ha along Victoria's central coast 217 embayments (Navarro et al., 2021). While the wetland at the site is a saltmarsh-mangrove-218 seagrass wetland system, the footprint of the flux tower was limited to the saltmarsh alone,

- 219 which extends more than a kilometre from the shoreline in places. This geography provided
- 220 the critical horizontally homogenous area with flat terrain required for ecosystem flux
- 221 measurements. Floristically this saltmarsh is species poor, dominated by Sarcocornia
- 222 quingeflora. Stands of Tecticornia arbuscula are common in this saltmarsh, while Atriplex
- 223 cinerea,7pprox.7aaustralia and Distichis distichophylla can be prevalent depending on

Deleted: , Suaeda

225	elevation and soil drainage conditions. Sarcocornia quinqeflora is a perennial succulent and	
226	at the temperate ranges of its distribution it has a distinct growing season from October to	
227	May (Fig. 1c) when the stems turn red, followed by a woody and fibrous dormant period	
228	during the colder months of June through September (Fig. 1d). The height of the dominant	
229	vegetation ranged between 0.3, m.	 Deleted: -0.6
230		
231	2.83 2.2 Data Collection and Analysis	 Formatted: List Paragraph, Outline numbered + Level: 2 + Numbering Style: 1, 2, 3, + Start at: 82 + Alignment: Left
232		+ Aligned at: 0.63 cm + Indent at: 1.27 cm
233	Eddy covariance measurements were made between November 2019 and August 2021	
234	capturing both the saltmarsh growing season (October-May) as well as a dormant period	 Deleted: March
235	(June-September). An array of standard micro-meteorological instruments included a 3-	 Deleted: April
236	dimensional sonic anemometer (CSAT3, Campbell Scientific, USA), an open-path infra-red	
237	carbon dioxide (CO ₂) gas and water vapour (H ₂ O) analyser (Li-7500, Li-Cor, USA) and 2	
238	data-loggers. The tower was powered by a solar array with two accompanying 12V DC	
239	storage batteries. The sonic anemometer was mounted 2.3 m above ground. The $CO_2/H2O$	
240	gas analyser was mounted 0.11 m longitudinally displaced from the anemometer. A CR3000	
241	datalogger (Campbell Scientific, USA), recorded the Li-7500, anemometer, short- and long-	
242	wave radiation (CNR4, Klip & Zonen, the Netherlands), air temperature and humidity (083E,	
243	Met One, USA) readings at 10 Hz frequency. Due to the location of the site in the Bass Strait	
244	(a region that experiences regular winter storms, high wind speeds and higher than national	
245	average cloud cover) the tower sustained damage due to winter storms several times during	
246	the deployment, as well as suffered periods of poor power supply due to short day lengths	
247	and high cloud cover; this was exacerbated by poor accessibility to the remote location during	
248	COVID-19 travel restrictions. The analysis thus focused on extended periods of continuous	 Deleted: ,
249	daily records and periods with large gaps in the dataset were removed.	
250		
251	Ecosystem fluxes were calculated for 30 min intervals using Eddy Pro software v.7 (LI-COR	
252	Inc., USA) Express Mode protocols. This processing step includes coordinate axis rotation	
253	correction, trend correction, data synchronisation, statistical tests for quality, density	
254	corrections and spectrum corrections. As part of this step, flux quality flags were assigned to	
255	the calculated CO_2 fluxes using the 0–2 flag policy 'Mauder and Foken 2004', based on the	
256	steady state test and the developed turbulent conditions test. The steady state test checks if	
257	fluxes remain consistent over the 30-minute averaging period by comparing the mean and	
·		

262	standard deviation (SD) of fluxes in the first and second halves of the period. The developed
263	turbulent conditions test ensures turbulence is well-developed and its energy spectra fits the
264	Kolmogorov spectrum. Both tests assign partial flags that are combined into a single flag (0-
265	2) in Eddy Pro, indicating the overall data quality. Only data that met the criteria of being in
266	quality class 0 ('best quality fluxes') for CO ₂ flux were chosen for further analysis. We
267	further removed anomalous data points defined as values that exceed four <u>SDs</u> from the mean
268	CO_2 flux; this resulted in the additional loss of ca. 1% of the dataset. Gap filling was not
269	applied. Additional filtering was applied to nighttime data due to known weak convection at
270	night, thus CO ₂ flux data during periods of atmospheric stability, i.e. when night friction wind
271	velocities (u*) were below 0.2 m s ⁻¹ , were excluded, following inspection of the nightly NEE
272	vs. u* curve to detect the threshold where NEE fall-off occurs. 0.2 m s ⁻¹ is the typical
273	threshold value used in eddy-covariance studies (Davis et al., 2003). This resulted in a dataset
274	of 674 day-time and 606 nighttime flux measurements during the dormant period and 4124
275	day-time and 3020 nighttime flux measurements for the growing period (Table 1). The
276	growing season dataset included 90 days with 85% or more flux data coverage, while the
277	dormant season dataset included 18 days, and these days were used for 24-hour flux
278	integrations.
279	
280	Table 1: Mean (±SD) net ecosystem exchange (µmol CO ₂ m ⁻² s ⁻¹) during day- and nighttime
281	respectively, as well as the corresponding number of half hourly measurements from each
282	month, following filter applications (n). Pink shading indicates the dormant season at the
283	French Island saltmarsh.

Month	<u>Daytime Mean NEE (SD); n</u>	Nighttime Mean NEE (SD); n
October 2019	<u>-2.29 (3.08); 121</u>	<u>2.04 (1.28); 70</u>
November 2019	<u>-1.84 (3.89); 151</u>	<u>2.85 (1.75); 110</u>
December 2019	<u>-3.33 (4.59); 96</u>	<u>1.14 (1.70); 15</u>
January 2020	<u>-1.31 (3.31); 63</u>	<u>2.10 (0.79); 27</u>
February 2020	<u>-3.83 (4.11); 540</u>	<u>1.89 (1.10); 280</u>
March 2020	<u>-3.86 (3.90); 494</u>	<u>1.63 (0.78); 351</u>
August 2020	<u>0.05 (2.05); 150</u>	<u>1.76 (1.22); 39</u>
September 2020	<u>-0.98 (2.04); 147</u>	<u>1.27 (0.96); 101</u>
January 2021	<u>-4.81 (5.04); 602</u>	<u>2.15 (1.55); 373</u>
February 2021	<u>-3.62 (4.27); 615</u>	<u>2.00 (1.19); 423</u>
March 2021	<u>-3.07 (3.95); 660</u>	<u>1.76 (1.20); 556</u>
April 2021	<u>-2.08 (3.02); 409</u>	<u>1.15 (0.87); 403</u>

Deleted:	standard	deviations

Deleted: night-time

Deleted: .

Deleted: night-time

Deleted: night-time Deleted: .

<u>May 2021</u>	<u>-0.98 (2.57); 377</u>	<u>1.14 (1.04); 423</u>
<u>June 2021</u>	<u>0.58 (1.67); 271</u>	<u>0.93 (1.30); 328</u>
July 2021	<u>1.07 (1.38); 102</u>	<u>0.82 (0.62); 127</u>

2	9	1
_	_	_

292 Half-hourly average CO2 flux was measured in µmol m⁻² s⁻¹, with positive fluxes indicating a 293 294 flux direction from the Earth's surface to the atmosphere. Net ecosystem exchange (NEE) 295 was defined as the net flux of CO2 from the atmosphere to the marsh and was often negative 296 during daytime, indicating that Gross Primary Productivity (GPP) was larger than ecosystem 297 respiration (Re). Evapotranspiration (ET) was calculated by Eddy Pro as the ratio between the 298 latent heat flux (LE) and latent heat of vaporisation (λ). Ecosystem water use efficiency 299 (WUEe) was then expressed as the ratio between daytime net ecosystem productivity in g 300 CO₂ m⁻² h⁻¹ and evapotranspiration in mm h⁻¹. 301 302 A two-dimensional footprint estimation was provided according to the simple footprint 303 parameterisation described in Kljun et al. (2015) calculating the ground position of the 304 cumulative fraction of flux source contribution by distance for each 30-minute interval. We 305 assessed the short-term effects of environmental factors on CO₂ fluxes at a half-hourly time 306 scale (e.g. the effects of light, air temperature and vapour pressure deficit) using a series of 307 non-linear or linear models. These analyses were limited to the growing season, when the 308 plants were actively photosynthesising. To calculate the daily-integrated CO₂ and H₂O fluxes. 309 the daily sum of these fluxes was determined for days with at least 85% data coverage. This 310 involved using the trapezoid rule to estimate the area under the curve for each of these 24-311 hour periods. The trapezoid rule approximates the total flux by dividing the day into smaller 312 intervals, each lasting 1,800 seconds. For each interval, the area is calculated by averaging 313 the flux values at the beginning and end of the interval, then multiplying by the interval 314 duration. These areas are then summed to obtain the total daily flux. This method ensures that 315 even with some missing data points, a reliable estimate of the daily flux can be obtained. All 316 post-processing and statistical analyses were performed in R 4.3.2 (R Core Team, 2024) 317 including the packages ggplto2, clifro, MASS, dismo, amerifluxr, rmarkdown, geosphere, 318 ggmap and gbm. 319

- 320 Because of the large data gaps, it was not possible to model the partition of the NEE in GEP
- 321 and Re using common partitioning methods (Lasslop et al., 2010). For simplicity, it was

Deleted: The

 Deleted: over time (i.e.,

 Deleted: CO2 or H2O flux) were calculated

 Deleted: complete records (

 Deleted: density>80%) as

 Deleted: period according to the trapezoid rule.



respiration using <u>a</u> linear slope of the relationship between <u>nighttime</u> NEE and temperature.

For the CO₂ budget, Net Ecosystem Production (NEP), <u>was</u> defined as NEP=-NEE,

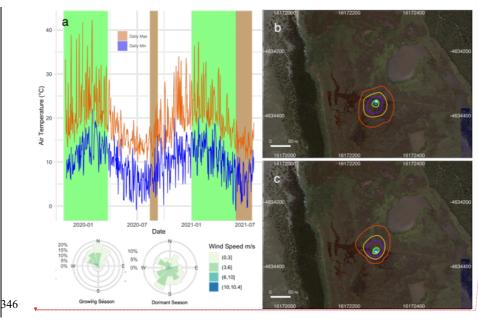
331

332 3. Results

333

334 The observations were divided into a growing season and a dormant season to reflect the 335 seasonal phenology of the dominant vegetation type within the flux tower footprint, which 336 has a relatively short growing season during the summer. During the growing season, mean 337 temperature averaged 22.3°C. Several heatwaves occurred during this period, with 338 temperatures exceeding 40°C on a few occasions in 2019. The dormant season was 339 significantly colder and windier, with frequent southerly winds (Fig. 2a). Footprint models 340 showed a slight variation in flux source between the two seasons, although in both cases the 341 size of the footprint and the vegetation composition within the footprint was similar (Figs. 2b 342 and 2c), but the shape was skewed to the north during winter due to the prevalent southerly 343 winds in that season (Fig. 2a). 70% of the flux measurement source was from within 50 m of 344 the tower, while the maximum length of the source location was 73 m.

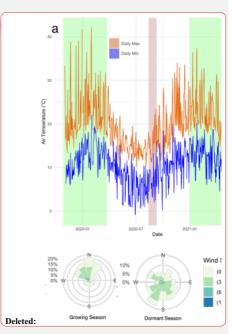




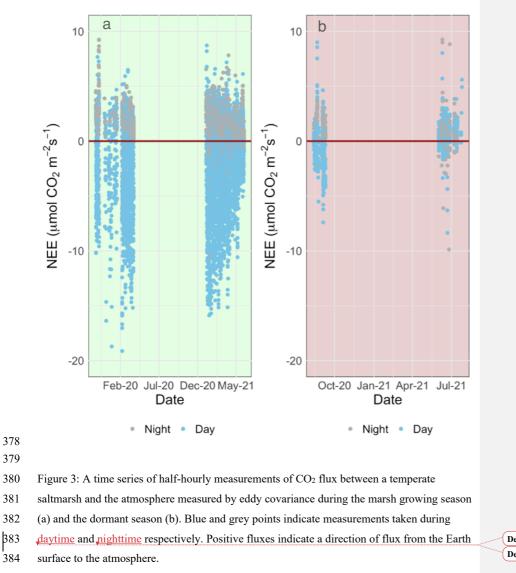
Deleted: the

Deleted: night-time

Deleted: , and Gross Ecosystem Production (GEP), defined as GEP=-GPP, were used



352			
353	Figure 2: a) The minimum and maximum daily temperature recorded at the Cerberus		
354	meteorological station (Bureau of Meteorology, Fig. 1b) during 2019-2021. The marsh		
355	growing (October-May) and dormant (June-September) periods observed during this study	(Deleted: Nov-Mar
356	are shaded in green and pink respectively. A corresponding wind rose diagram summarises	(Deleted: Aug-Sep
357	the wind speeds and directions measured at the tower site during the observation periods. The	(Deleted: highlighted
358	flux source footprint surrounding the tower during the dormant season (b) and the growing		
359	season (c) shows the cumulative flux source contribution to the flux measurements, with the		
360	outer red line representing the distance by which 90% of the calculated flux is sourced and		
361	the other isolines from the tower outwards correspond to 10%, 20%, 40%, 60% and 80% of		
362	the flux.		
363			
364	The growing season dataset included 90 days with <u>\$5</u> % or more flux data coverage, while the	(Deleted: 80
365	dormant season dataset included 18 days. There was a strong temporal variability in net		
366	ecosystem exchange (NEE) across both short (daily) and long (seasonal) temporal scales		
367	(Fig. 3). Daytime fluxes were defined as flux points where the global radiation values in the		
368	flux averaging half-hour interval were >12 W m ⁻² (as per EddyPro methodology). At the	(Deleted: .
369	diurnal scale, saltmarsh NEE were negative mostly during the day and positive mostly during		
370	the night and ranged between -19.1 and 10.86 $\mu mol\ m^{-2}\ s^{-1}$ across the measurement periods.		
371	Monthly averages and data coverage are shown in Table 1.		
372			



- 385
- $386 \qquad \mbox{Flux rates varied across the day, with CO_2 uptake peaking at 11:00 during the growing}$
- season, and later in the day (14:00) during the dormant period (Fig. 4). Ecosystem respiration
- rates (R_e, defined as <u>nighttime</u> CO₂ flux) were on average (\pm SD) 1.77 (\pm 1.12) µmol m⁻² s⁻¹
- during the growing season and 1.0 (\pm 0.93) µmol m⁻² s⁻¹ during the dormant period. The

Deleted: day-time Deleted: night-time

Deleted: night-time

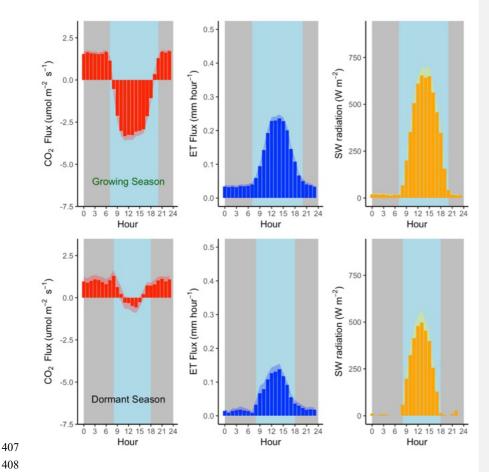
13

- 393 difference in ecosystem respiration between the growing and dormant seasons is highly
- 394 significant (t-test, p<0.01). Daytime CO₂ flux was on average (\pm SD) -3.53 (\pm 4.15) µmol m⁻²
- 395 s⁻¹ during the growing season and -0.25 (± 2.18) µmol m⁻² s⁻¹ during the dormant season.
- 396 Thus, we derive that the maximum Gross Primary Productivity (GPP) of this ecosystem from
- NEE and temperature-corrected Re, (Fig. 5), measured during the growing season, is ca. -5.34

- 400
- 401 Mean (\pm SD) daily evapotranspiration was 2.48 mm (\pm 2.79 mm) during the growing season
- 402 and 0.97 mm (\pm 1.35 mm) during the dormant season (Fig. 4). Evapotranspiration peaked at
- 403 noon AEST during the growing season (0.26 mm h⁻¹), and later in the day (14:00 AEST)
- 404 during the dormant season (0.14 mm h^{-1}).

405

Deleted: ,





409 Figure 4: Mean hourly CO2 and H2O flux (evapotranspiration) rates during the growing

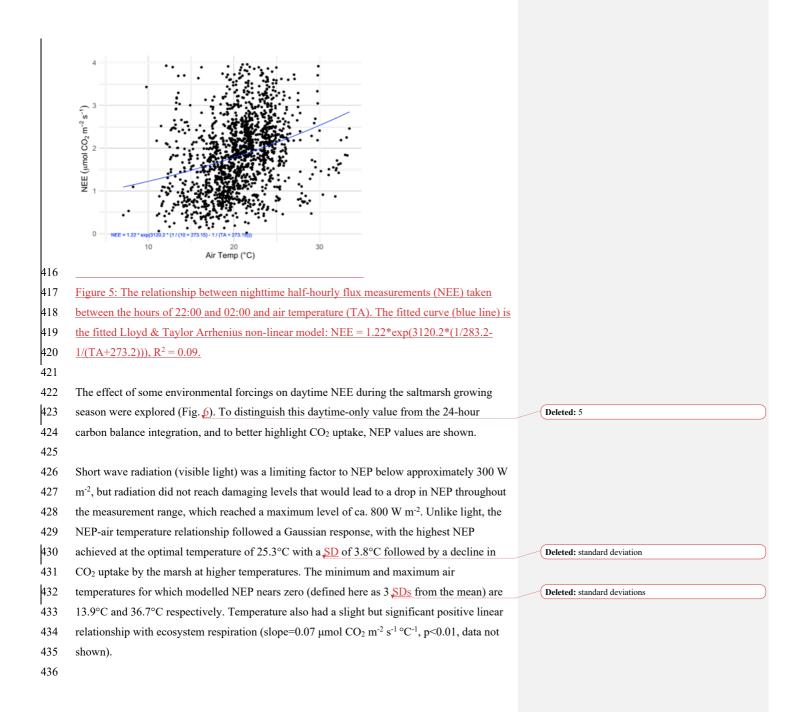
410 season (top) and the dormant season (bottom) alongside mean short wave incoming radiation.

411 Shading corresponds to 1 standard deviation (SD) around the mean. Grey plot background

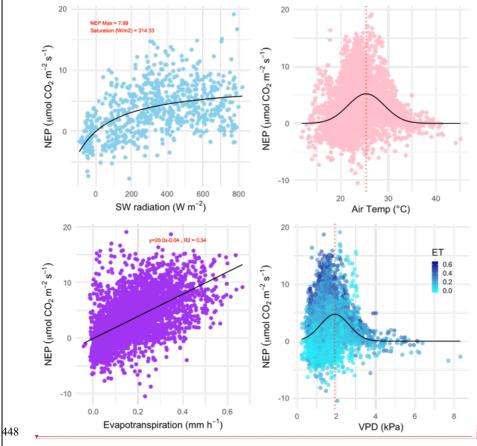
412 approximates nighttime periods, while light blue approximates daytime (actual day length

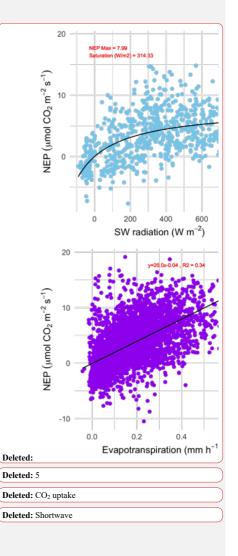
- 413 varies within each season).
- 414

Deleted: night-time

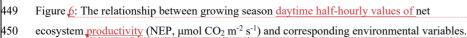


- $440 \qquad \text{NEP was positively correlated with evapotranspiration during the growing season (Pearson r$
- 441 = 0.59, Fig. (c). The slope of the NEP/ET relationship was 20.0, indicating an ecosystem
- 442 water use efficiency (WUE_e) of 0.86 g C kg⁻¹ H₂O ($R^2 = 0.34$, p<0.001). The response of
- 443 NEP to atmospheric vapour pressure deficit (VPD) fit a Gaussian relationship (the commonly
- 444 observed inverse U-shaped curve relationship in response to VPD in plants), with NEP
- 445 declining rapidly when VPD exceeded 2.39 kPa. The optimal range of VPD within which
- 446 NEP was maximised in this ecosystem was 1.92 kPa (± 0.73 kPa).



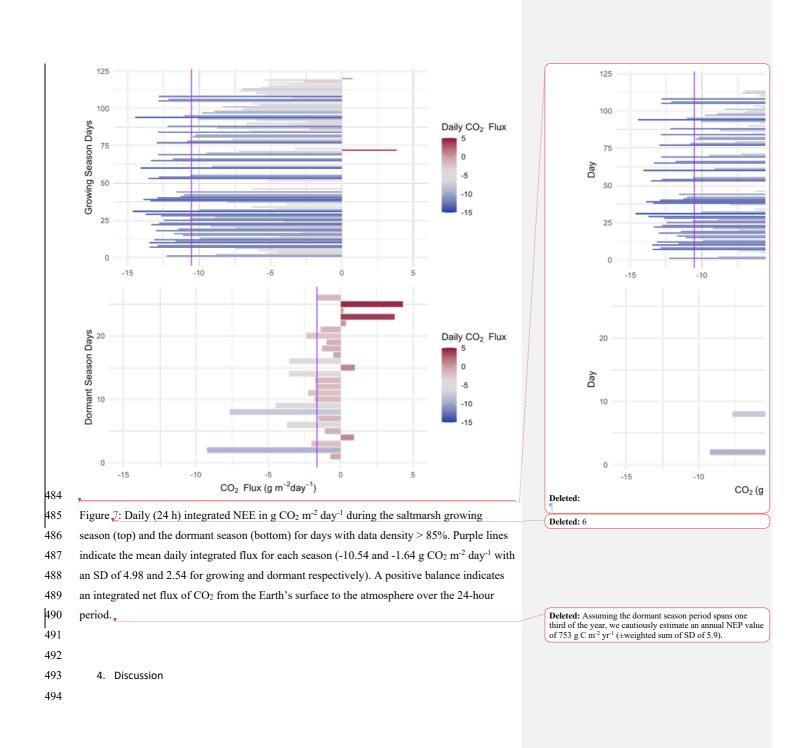


Deleted: 5



451 a) <u>Net shortwave (SW)</u> radiation (visible light); black line is the Michaelis-Menten model of

457	best fit. The coefficient of saturation is at 314 W m ⁻² and maximum net productivity is 8.0	
458	$\mu mol \ CO_2 \ m^{-2} \ s^{-1}. \ b)$ Air temperature; black line is a Gaussian model of best fit with a	
459	temperature optimum at 25.3 °C. c) Evapotranspiration; linear model ($R^2 = 0.34$) has a slope	
460	of 20.0. d) Vapour Pressure Deficit; black line is a Gaussian model of best fit with a VPD	
461	optimum at 1.92 kPa, points are coloured by the level of evapotranspiration during the half	 Deleted: .
462	hourly NEP measurement.	
463		
464	When integrated over a 24-hour period, the saltmarsh is on average a $\underline{\text{daily}}$ CO ₂ sink during	
465	all canopy phenological phases (Fig. \mathcal{J}), although during the dormant season the sink is	 Deleted: 6
466	weaker, with an average uptake of -2.42 g CO_2 m^-2 day^-1 (± 2.54). During the growing season	
467	(defined as the non-dormant period and thus reflecting several phenological stages), the	
468	marsh is a substantial sink with a mean (±SD) daily NEP of 10.95 g CO ₂ m ⁻² day ⁻¹ (±4.98)	
469	over a 24-hour period (ranging between -22.8 and 4.3 g of CO_2 emission to the atmosphere	
470	$m^{\text{-}2}\text{day}^{\text{-}1})$. The daily CO_2 budget during the growing season showed some variability among	
471	days (CV=0.46, Fig. 2) and days with lower average light levels (i.e. cloudy days) had a	 Deleted: 6
472	significant negative impact on the CO ₂ budget (multiple linear regression, $p < 0.02$, $R^2 =$	
473	0.27). Daily maximum air temperatures did not have a significant impact on the daily CO_2	
474	budget ($p = 0.77$) at this location, although NEE was significantly affected by temperature at	
475	finer temporal scales (Figure <u>6).</u>	 Deleted: 5). Assuming the dormant period spans a third of the year, we cautiously estimate an annual NEP value of 753
476		the year, we cautiously estimate an annual NEP value of 753 (± 112.7) g C m ⁻² yr ⁻¹ .
477		



501	The study provided high-frequency measurements of an abundant greenhouse gas (CO2)	
502	using a precise technique (eddy covariance flux) in an ecosystem with limited historical	
503	measurements. Time series analysis was performed on CO2 flux measurements across various	
504	scales (daily, nightly, diel, half-hourly, hourly, seasonally) to assess the impacts of ET, SW	
505	Rad, VPD, and Tair on CO2 flux and how these relationships change throughout the year.	
506	Seasonality was observed for the first time in an Australian saltmarsh and had a significant	
507	effect on carbon and water flux. Growing season net ecosystem productivity was five times	
508	greater than during the dormant period. Seasonality in Australian marshes has not been	
509	previously reported in the scientific literature, and contradicts previous assumptions, that	~
510	Australian saltmarshes do not exhibit the growing and dormant phenology observed on other	
511	continents (Clarke and Jacoby, 1994). Seasonality had a significant impact on the carbon	<
512	budget in this marsh and is an important characteristic of this habitat that has been	
513	overlooked (Owers et al., 2018). Seasonality can also have other broader implications yet to	
514	be considered in Australian marshes. For example, in the USA, the saltmarsh greening up	
515	period was shown to be an important range-wide timing event for migratory birds (Smith et	
516	al., 2020) with plant-growth metrics predicting the timing of nest initiation for shorebirds.	
517	Saltmarshes in Australia are important roosting and feeding sites along the East Asian	
518	Australasian Flyway, particularly for waders, thus potentially a similar relationship between	
519	migration timing and saltmarsh phenology could be occurring. Seasonality also affects other	
520	significant ecosystem functions such as the bio-geomorphological feedback between	
521	saltmarshes, coastal hydrodynamics and landscape evolution (Reents et al., 2022).	
522		
523	We derived the light-response and associated coefficients of light regulation of saltmarsh	
524	NEE using the Michaelis Menten model (Chen et al., 2002). Quantum (or production)	
525	efficiency is the predominant input in remote sensing techniques to model productivity, and is	
526	specific to the biome (Hilker et al., 2010). While not directly comparable to leaf level	
527	quantum efficiency measurements, the quantum efficiency (α) of the NEP light response	
528	curve was estimated from the slope of the Michaelis-Menten model to be 0.025 $\mu mol \ CO_2 \ J^{-1}.$	
529	The ecosystem reached light saturation at an insolation of 314 W m^{-2} , but daytime insolation	
530	was below this value more than 50% of the time suggesting that light might be a significant	
531	limiting factor to NEP at this marsh, especially during winter. The level of light limitation we	
532	observed is an underestimation, due to the loss of high-quality EC data during periods of rain.	
533	The solar geometry at this latitude and the length of day result in an annual average top of	

Deleted: At this temperate saltmarsh, seasonality

Deleted: ,

Deleted: were made

Deleted: ¶

Deleted: might be an overlooked

Deleted: and in addition to affecting flux estimations,

540	atmosphere SW radiation of 250 W m ⁻² , but clouds can strongly modulate the SW radiation
541	balance (SWCRE), and apart from the months of January and February when cloudy days are
542	less frequent (10-12 days per month), cloudy days are frequent at this site, averaging 15-17
543	days per month (Bureau of Meteorology) and could significantly impact on NEP.
544	
545	Temperature is another forcing that significantly impacts NEE at this marsh, with an optimal
546	range for maximum NEP at 25.3°C (21.5°C-29.1°C). Data for Australian saltmarshes is not
547	available, but this optimal temperature response range is similar to that measured
548	experimentally in a saltmarsh species in an equivalent climate zone (e.g. Georgia,
549	(Giurgevich and Dunn, 1981)) and to the values hypothesised for the habitat from data
550	collected along the US Atlantic Coast, (Feher et al., 2017). The long-term average maximum
551	daytime temperature at this site is 19.2°C, which is cooler than the optimal range for NEE
552	suggesting temperature can be a significant limiting factor to productivity, especially during
553	the dormancy period where average monthly maximum temperatures are only 13.7°C to
554	16.6°C (Bureau of Meteorology). During the growing season the average maximum
555	temperatures are within the range of optimal NEE (20.6°C to 23.1°C), although hot days
556	(>30°C) significantly depress NEE and depending on the year, can be common during
557	summer months (averaging 2-6 days per month). Within the diversity of saltmarsh species
558	found globally, some species have C4 photosynthetic pathways (Drake, 1989). C4
559	photosynthesis plants often exhibit higher optimum temperature ranges (30-35°C, Berry and
560	Björkman, 1980) than C3 photosynthesis plants (20-25), and the cooler conditions at this site
561	could explain the absence of C4 plants from this bioregion. The parabolic relationship
562	between NEP and air temperature and NEP and VPD suggest that higher air temperatures and
563	<u>VPD</u> (which are expected with climate change) could negatively impact CO ₂ uptake by these
564	coastal ecosystems. High VPD was related to lower NEP, and to a lesser extent, lower ET
565	(Fig. 6d). However, VPD increases atmospheric demand for water, increasing the evaporation
566	from the saturated marsh surfaces in the footprint, and this atmospheric demand could be
567	forcing ET at high VPD rather than plant moderation via reduced transpiration, even if
568	transpiration is reduced. Thus, despite maintained ET during VPD periods we cannot
569	conclude a non-closure of stomata. NEP also reduced below a VPD of 1.92 KPa, but at our
570	field site low VPD correlated with low temperatures ($r = 0.88$), and low temperatures were
571	shown to limit NEP.
572	
573	Υ

Deleted: ¶

575	In saltmarshes, evapotranspiration occurs from plant mediated transpiration but also from soil	
576	pores (which tend to be saturated), wetted leaves and open water. We observed average	
577	evaporation rates of 2.48 mm day-1 during the growing season and 0.97 mm day-1 during the	
578	dormant season. Actual evapotranspiration in this region modelled using the CMRSET	
579	algorithm is estimated to range between 0.6 and 3.2 mm day-1 during winter and summer	
580	respectively (McVicar et al., 2022); our field measurements support the model. Overall,	Deleted: these values are consistent with
581	rainfall is in excess of the requirements for maintaining ET at this site, although deficits can	
582	develop for short periods during the growing season, when ET is higher, perhaps explaining	Deleted: . ¶
583	the drier saltmarsh surface during this period. Conversely, long term rainfall excess could be	Long
584	contributing to the complicated hydrology at this location, where inundation is not strictly	
585	associated with tidal stage (data not shown) and our observation of long (5-day) periods of	Deleted:).
586	inundation during winter.	
587		
588	Growing season ET rates are significantly higher than those of the dormant season, partly due	
589	to the solar configuration in winter as opposed to summer, but also due to phenological	
590	changes. A big leaf model estimation of evapotranspiration from saltmarshes in New South	
591	Wales estimates ET to be highly sensitive to vegetation height, increasing by more than 1 mm	
592	day-1 as vegetation height increases from 0.1 to 0.4 m (Hughes et al., 2001) and transpiration	
593	in saltmarsh plants in the cold season has been shown to account for only 20% of the annual	
594	transpiration budget (Giurgevich and Dunn, 1981) following the same pattern as the seasonal	
595	distribution of productivity.	
596		
597	The rate of carbon uptake per unit of water loss (WUE) is a key ecosystem characteristic,	
598	which is a result of a suite of physical and canopy physiological forcings, and has direct	
599	implications for ecosystem function and global water and carbon cycling. Mean water use	
600	efficiency (WUEe) of this saltmarsh was estimated at 0.86 g C kg ⁻¹ H ₂ O, which is markedly	
601	lower than for grass dominated saltmarshes in China (2.9 g C kg ⁻¹ H ₂ O, Xiao et al. (2013))	
602	but similar to the value for WUEe based on NEP and ET in mangroves (0.77 g C kg ⁻¹ H ₂ O,	
603	Krauss et al. (2022)), which are also C3 plants. The Chinese saltmarshes studied in Xiao et al.	Deleted: (2022)). The
604	(2013) are dominated by Spartina alterniflora, a C4 perennial grass. C4 plants have higher	
605	(often double) water use efficiencies than C3 plants due to CO2 concentrating mechanisms	
606	(Osborne and Freckleton, 2009). The saltmarsh at French Island includes only C3 plants, and	
607	the dominant chenopod Sarcocornia quinqueflora has been suspected to have higher	
608	evapotranspiration rates than saltmarsh by approx. 15% (Hughes et al., 2001), but while	Deleted: grasses

615	Sarcocornia quinqueflora dominates at this site, the footprint is a mix of species, and the
616	lower WUEe cannot be directly linked to the presence of Sarcocornia quinqueflora.
617	Furthermore, like most wetlands, the wetland surface is a mixed composition of emergent
618	vegetation, unsaturated soil and water bodies thus the spatial scale at which WUEe is
619	determined encompasses both the canopy (Ec) as well as any open water present in the
620	footprint. Transpiration is predicted to account for only 55% of ET in these systems (Hughes
621	et al., 2001), which is an Ec to ET ratio similar to that of mangroves (Krauss et al., 2022) but
622	significantly lower than terrestrial forests where more than 90% of ET can be attributed to
623	transpiration. Thus, regional variations in WUEe can be attributed to multiple forcings that
624	form complex spatiotemporal patterns.
625	
626	Saltmarshes are considered among the most productive ecosystems on Earth with an
627	estimated global NEP of 634 Tg C y ⁻¹ (Fagherazzi et al., 2013) and 601 634 Tg C y ⁻¹
628	(Rosentreter et al., 2023). Productivity of southern Australian marshes was previously
629	estimated at 0.8 kg m ⁻² y ⁻¹ by repeated measurements of above ground standing crops (Clarke
630	and Jacoby, 1994), which if not accounting for season, equates to 2,2 g C m ⁻² d ⁻¹ . Similar
631	studies on saltmarshes in France report lower productivity (483 g C m ⁻² y ⁻¹ , (Mayen et al.,
632	2024)) and daily growing season rates of 1.53 g C m ⁻² d ⁻¹ , but mid-latitude saltmarsh sites in
633	the USA and China show productivity rates of 775 g C m ⁻² y ⁻¹ , (Wang et al., 2016) and 668 g
634	C m ⁻² y ⁻¹ , (Xiao et al., 2013 <u>) respectively.</u> It is clear that productivity across climate zones
635	and biogeographic regions varies widely with some studies even reporting net emissions over
636	an annual period from some marshes and a global average estimated between 382 (Alongi,
637	2020) and 1,585 g C m ⁻² y ⁻¹ (Chmura et al., 2003), albeit based on a small subset of studies.
638	An analysis of GPP across latitudes in the USA show that warmer sites (including mangrove
639	wetlands in southern USA) had significantly higher GPP than mid-latitude saltmarshes such
640	as the one on French Island (Feagin et al., 2020). Mangroves have higher NEE than
641	saltmarshes, estimated by Krauss et al. (2022) to average 1200 g C m ⁻² y ⁻¹ . While our data
642	does not provide enough coverage for a long-term annual estimate of carbon flux, our daily
643	values of an average of 2.88 g C m ⁻² d ⁻¹ during the growing season, combined with the
644	relatively short dormant season relative to other temperate locations, suggest a high carbon
645	sequestration rate for this ecosystem type. In another southern hemisphere study, growing
646	season rates at an EC tower site in Argentina, are extrapolated by us to average 1.6 g C m ⁻² d
647	¹ (Bautista et al., 2023) but in that saltmarsh, flooding reduced vegetation biomass and
648	productivity.

Formatted: Not Superscript/ Subscript

Deleted: is remarkably similar
Deleted: the values reported here, where we extrapolate an approximate annual mean of 0.75 kg C m^{-1}
Deleted: y
Formatted: Not Superscript/ Subscript
Deleted: than the marshes at French Island (-
Deleted: but our values are within the range reported for
Deleted: (-
Deleted:))
Deleted: China (-
Deleted:)).
Deleted: -
Deleted: -

 $\label{eq:Deleted: Krauss et al., 2022} \ensuremath{\text{Deleted: (Krauss et al., 2022) to average 1200 g C m^{-2} y^{-1}}.$

662			
663	The data presented here is the exchange of carbon between the land surface and the		
664	atmosphere, but saltmarshes, like other marine connected communities, exchange carbon also		
665	through dissolved carbon pathways, which can be significant (Cai, 2011). Thus, the fluxes		
666	presented here do not constitute the entire carbon budget of this ecosystem.		
667			
668	5. Conclusions		
669			
670	The response of the French Island saltmarsh to environmental drivers is indicative of the		
671	complex interactions determining saltmarsh productivity. The unique long-term, high-		Deleted: While the overall carbon sequestration rate we measured was in the range of other temperate saltmarsh
672	resolution record enabled us to derive temperature, VPD and light response functions, thus	\setminus	estimates (ca. 750 g C m ⁻² y ⁻¹),
673	formulating equations that describe how climate-change sensitive parameters such as	Ŷ	Deleted: t
674	temperature, relative humidity, and cloud cover, affect CO2 uptake, respiration and		
675	evapotranspiration. The marsh operated as a CO_2 sink throughout the various canopy		
676	phenological phases, but during the dormant period, CO_2 uptake was less than 25% that of		
677	the growing season. Seasonality of greenhouse gas fluxes in Australian saltmarshes is an	<	Deleted: has not been previously considered
678	understudied but important aspect of global carbon budgeting.		Deleted: and it should not be overlooked when estimating saltmarsh
679			Deleted: budgets
680	Competing interests		
681			
682 683	The contact author has declared that none of the authors has any competing interests.	(Formatted: Font: Times New Roman, 12 pt
684	Acknowledgments		
685			
686	The work was carried out with the permission of Parks Victoria (Permit 10008684). We thank		
687	Phil and Yuko Bock for logistic support and accommodation on French Island. We thank		
688	Leigh Burgess, Kiri Mason and Ian McHugh for technical support and the Australian OzFlux		
689	community for ongoing collaboration. This work was funded by an Australian Research		
690	Council Discovery Award to RR and ED (DP220102873) as well as a Monash University		
691	Networks of Excellence award to RR.		
692			
693	Data Availability		

- 702 Data used for this analysis is available at https://figshare.com/s/ba62aafd1a4049248a08 (note
- that this is a temporary private link to an embargoed dataset which will be replaced with a 703
- 704 publicly available DOI upon publication).
- 705
- 706 Author contribution
- 707 RR conceptualised the study, acquired funding, prepared the manuscript, designed and
- 708 carried out the field campaign, and performed the analysis. ED acquired funding, developed
- 709 methodology and prepared the manuscript. AG developed methodology and prepared the
- 710 manuscript. TA, EJVH, HR and MP were involved in the field investigation and
- 711 administration of the project and provided edits on the manuscript.
- 712
- 713 References
- 714
- 715 Adam, P.: Saltmarsh Ecology, Cambridge University Press, 1990.
- 716 Adam, P.: Morecambe Bay saltmarshes: 25 years of change, in: British Saltmarshes, Forrest Text, Cardigan, UK, 81-107, 2000. 717
- 718 Adam, P.: Saltmarshes in a time of change, Environ. Conserv., 29, 39-61,
- 719 https://doi.org/10.1017/S0376892902000048, 2002.
- 720 Alongi, D. M.: Carbon balance in salt marsh and mangrove ecosystems: A global synthesis, J. 721 Mar. Sci. Eng., 8, 767, 2020.
- Artigas, F., Shin, J. Y., Hobble, C., Marti-Donati, A., Schäfer, K. V. R., and Pechmann, I.: 722
- 723 Long term carbon storage potential and CO2 sink strength of a restored salt marsh in New

Jersey, Agric. For. Meteorol., 200, 313-321, https://doi.org/10.1016/j.agrformet.2014.09.012, 724 725 2015.

- Baldocchi, D. D.: Assessing the eddy covariance technique for evaluating carbon dioxide 726
- 727 exchange rates of ecosystems: past, present and future, Glob. Change Biol., 9, 479-492, https://doi.org/10.1046/j.1365-2486.2003.00629.x, 2003. 728
- 729 Bautista, N. E., Gassmann, M. I., and Pérez, C. F.: Gross primary production, ecosystem
- 730 respiration, and net ecosystem production in a southeastern South American salt marsh. 731
- Estuaries Coast, 46, 1923-1937, https://doi.org/10.1007/s12237-023-01224-8, 2023. 732

733 Berry, J., and Björkman, O.: Photosynthetic response and adaptation to temperature in higher

- 734 plants, Ann. Rev. Plant Physiol., 31, 491-543,
- https://doi.org/10.1146/annurev.pp.31.060180.002423, 1980. 735 736
- Borges, A. V., Schiettecatte, L.-S., Abril, G., Delille, B., and Gazeau, F.: Carbon dioxide in 737
- European coastal waters, Trace Gases Eur. Coast. Zone, 70, 375–387, 738
- 739 https://doi.org/10.1016/j.ecss.2006.05.046, 2006.

- 740 Cai, W.-J.: Estuarine and coastal ocean carbon paradox: CO₂ sinks or sites of terrestrial
- 741 carbon incineration?, Annu. Rev. Mar. Sci., 3, 123–145, https://doi.org/10.1146/annurev-742 marine-120709-142723, 2011.
- 743 Chen, J., Falk, M., Euskirchen, E., Paw U, K. T., Suchanek, T. H., Ustin, S. L., Bond, B. J.,
- 744 Brosofske, K. D., Phillips, N., and Bi, R.: Biophysical controls of carbon flows in three
- 745 successional Douglas-fir stands based on eddy-covariance measurements, Tree Physiol., 22,
- 746 169–177, https://doi.org/10.1093/treephys/22.2-3.169, 2002.
- Chmura, G. L., Anisfeld, S. C., Cahoon, D. R., and Lynch, J. C.: Global carbon sequestrationin tidal, saline wetland soils, Glob. Biogeochem. Cycles, 17,
- 749 https://doi.org/10.1029/2002GB001917, 2003.
- Clarke, P., J. and Jacoby, C. A.: Biomass and above-ground productivity of salt-marsh plants
 in South-eastern Australia, Aust. J. Mar. Freshw. Res., 45, 1521–1528, 1994.
- 752 Davis, K. J., Bakwin, P. S., Yi, C., Berger, B. W., Zhao, C., Teclaw, R. M., and Isebrands, J.
- <u>G.: The annual cycles of CO₂ and H₂O exchange over a northern mixed forest as observed</u>
 <u>from a very tall tower, Glob. Change Biol., 9, 1241-1332, https://doi.org/10.1046/j.1365-</u>
- 755 <u>2486.2003.00672.x, 2003.</u>
- Drake, B. G.: Photosynthesis of salt marsh species, Aquat. Bot., 34, 167-180, https://doi.org/10.1016/0304-3770(89)90055-7, 1989.
- Duarte, C. M.: Reviews and syntheses: Hidden forests, the role of vegetated coastal habitats
 in the ocean carbon budget, Biogeosciences, 14, 301–310, https://doi.org/10.5194/bg-14-301 2017, 2017.
- Fickson, J. E., Peresta, G., Montovan, K. J., and Drake, B. G.: Direct and indirect effects of elevated atmospheric CO₂ on net ecosystem production in a Chesapeake Bay tidal wetland,
- 764 Glob. Change Biol., 19, 3368–3378, 2013.

765 Fagherazzi, S., Wiberg, P. L., Temmerman, S., Struyf, E., Zhao, Y., and Raymond, P. A.:

- Fluxes of water, sediments, and biogeochemical compounds in salt marshes, Ecol. Process.,
 2, 3, https://doi.org/10.1186/2192-1709-2-3, 2013.
- 768 Feagin, R. A., Forbrich, I., Huff, T. P., Barr, J. G., Ruiz-Plancarte, J., Fuentes, J. D., Najjar,
- 769 R. G., Vargas, R., Vázquez-Lule, A., Windham-Myers, L., Kroeger, K. D., Ward, E. J.,
- 770 Moore, G. W., Leclerc, M., Krauss, K. W., Stagg, C. L., Alber, M., Knox, S. H., Schäfer, K.
- V. R., Bianchi, T. S., Hutchings, J. A., Nahrawi, H., Noormets, A., Mitra, B., Jaimes, A.,
 Hinson, A. L., Bergamaschi, B., King, J. S., and Miao, G.: Tidal wetland gross primary
- Hinson, A. L., Bergamaschi, B., King, J. S., and Miao, G.: Tidal wetland gross primary
 production across the continental United States, 2000–2019, Glob. Biogeochem. Cycles, 34,
- e2019GB006349, https://doi.org/10.1029/2019GB006349, 2020.
- 775 Feher, L. C., Osland, M. J., Griffith, K. T., Grace, J. B., Howard, R. J., Stagg, C. L.,
- 776 Enwright, N. M., Krauss, K. W., Gabler, C. A., Day, R. H., and Rogers, K.: Linear and
- nonlinear effects of temperature and precipitation on ecosystem properties in tidal saline
- 778 wetlands, Ecosphere, 8, e01956, https://doi.org/10.1002/ecs2.1956, 2017.

- 779 Gedan, K. B., Silliman, B. R., and Bertness, M. D.: Centuries of human-driven change in salt
- 780 marsh ecosystems, Annu. Rev. Mar. Sci., 1, 117-141,
- 781 https://doi.org/10.1146/annurev.marine.010908.163930, 2009.
- Ghosh, S. and Mishra, D. R.: Analyzing the long-term phenological trends of salt marsh 782
- ecosystem across coastal Louisiana, Remote Sens., 9, https://doi.org/10.3390/rs9121340, 783 784 2017.
- 785 Giurgevich, J. R. and Dunn, E. L .: A comparative analysis of the CO2 and water vapor
- 786 responses of two Spartina species from Georgia coastal marshes, Estuar. Coast. Shelf Sci.,
- 12, 561-568, https://doi.org/10.1016/S0302-3524(81)80082-5, 1981. 787
- 788 Hilker, T., Hall, F. G., Coops, N. C., Lyapustin, A., Wang, Y., Nesic, Z., Grant, N., Black, T.
- 789 A., Wulder, M. A., Kljun, N., Hopkinson, C., and Chasmer, L.: Remote sensing of
- photosynthetic light-use efficiency across two forested biomes: Spatial scaling, Remote Sens. 790
- Environ., 114, 2863-2874, https://doi.org/10.1016/j.rse.2010.07.004, 2010. 791
- 792 Hill, A. C. and Vargas, R.: Methane and carbon dioxide fluxes in a temperate tidal salt marsh:
- 793 comparisons between plot and ecosystem measurements, J. Geophys. Res. Biogeosciences,
- 127, e2022JG006943, https://doi.org/10.1029/2022JG006943, 2022. 794
- 795 Howe, A. J., Rodríguez, J. F., Spencer, J., MacFarlane, G. R., and Saintilan, N.: Response of 796 estuarine wetlands to reinstatement of tidal flows, Mar. Freshw. Res., 61, 702-713, 2010.
- 797 Hughes, C. E., Kalma, J. D., Binning, P., Willgoose, G. R., and Vertzonis, M.: Estimating
- 798 evapotranspiration for a temperate salt marsh, Newcastle, Australia, Hydrol. Process., 15, 957-975, https://doi.org/10.1002/hyp.189, 2001. 799
- 800 Huxham, M., Whitlock, D., Githaiga, M., and Dencer-Brown, A.: Carbon in the coastal
- 801 seascape: how interactions between mangrove forests, seagrass meadows and tidal marshes influence carbon storage, Curr. For. Rep., 4, 101-110, https://doi.org/10.1007/s40725-018-
- 802 0077-4, 2018. 803
- 804 Kathilankal, J. C., Mozdzer, T. J., Fuentes, J. D., D'Odorico, P., McGlathery, K. J., and 805 Zieman, J. C.: Tidal influences on carbon assimilation by a salt marsh, Environ. Res. Lett., 3, 806 044010, https://doi.org/10.1088/1748-9326/3/4/044010, 2008.
- 807 Kljun, N., Calanca, P., Rotach, M. W., and Schmid, H. P.: A simple two-dimensional
- 808 parameterisation for Flux Footprint Prediction (FFP), Geosci Model Dev, 8, 3695-3713, 809 https://doi.org/10.5194/gmd-8-3695-2015, 2015.
- Krauss, K. W., Lovelock, C. E., Chen, L., Berger, U., Ball, M. C., Reef, R., Peters, R., 810
- Bowen, H., Vovides, A. G., Ward, E. J., and others: Mangroves provide blue carbon 811 812 ecological value at a low freshwater cost, Sci. Rep., 12, https-doi, 2022.
- 813
- Lasslop, G., Reichstein, M., Papale, D., Richardson, A. D., Arneth, A., BARR, A., STOY, P., 814
- and WOHLFAHRT, G.: Separation of net ecosystem exchange into assimilation and 815 respiration using a light response curve approach: critical issues and global evaluation, Glob.
- Change Biol., 16, 187-208, https://doi.org/10.1111/j.1365-2486.2009.02041.x, 2010. 816

- Lu, W., Xiao, J., Liu, F., Zhang, Y., Liu, C., and Lin, G.: Contrasting ecosystem CO₂ fluxes
 of inland and coastal wetlands: a meta-analysis of eddy covariance data, Glob. Change Biol.,
- 819 23, 1180–1198, https://doi.org/10.1111/gcb.13424, 2017.
- 820 Massamann, A., Gentine, P., and Lin, C.: When does vapour pressure deficit drive or reduce
- 821 evapotranspiration? J. Adv. Model. Earth Syst., 11, 3305-3320, 2019.
- 822 https://doi.org/10.1029/2019MS001790
- 823 Mayen, J., Polsenaere, P., Lamaud, É., Arnaud, M., Kostyrka, P., Bonnefond, J.-M., Geairon,
- 824 P., Gernigon, J., Chassagne, R., and Lacoue-Labarthe, T.: Atmospheric CO₂ exchanges
- 825 measured by eddy covariance over a temperate salt marsh and influence of environmental
- 826 controlling factors, Biogeosciences, 21, 993–1016, 2024.
- 827 McLeod, E., Chmura, G. L., Bouillon, S., Salm, R., Björk, M., Duarte, C. M., Lovelock, C.
- E., Schlesinger, W. H., and Silliman, B. R.: A blueprint for blue carbon: toward an improved
 understanding of the role of vegetated coastal habitats in sequestering CO₂, Front. Ecol.
- 822 understanding of the fore of vegetated coastal habitats in sequester.
 830 Environ., 9, 552–560, https://doi.org/10.1890/110004, 2011.
- Mcowen, C. J., Weatherdon, L. V., Bochove, J.-W. V., Sullivan, E., Blyth, S., Zockler, C.,
 Stanwell-Smith, D., Kingston, N., Martin, C. S., Spalding, M., and Fletcher, S.: A global map
- 633 of saltmarshes, Biodivers. Data J., 5, e11764, https://doi.org/10.3897/BDJ.5.e11764, 2017.
- McVicar, T., Vleeshouwer, J., Van Niel, T., Guerschman, J., and Peña-Arancibia, J. L.:
 Actual Evapotranspiration for Australia using CMRSET algorithm. Version 1.0, 2022.
- Mitsch, W. J. and Gosselink, J. G.: The value of wetlands: importance of scale and landscape
 setting, Ecol. Econ., 35, 25–33, https://doi.org/10.1016/S0921-8009(00)00165-8, 2000.
- 838 Moffett, K. B., Wolf, A., Berry, J. A., and Gorelick, S. M.: Salt marsh-atmosphere exchange
- 839 of energy, water vapor, and carbon dioxide: Effects of tidal flooding and biophysical controls,
- 840 Water Resour. Res., 46, 2010.
- 841 Nahrawi, H., Leclerc, M. Y., Pennings, S., Zhang, G., Singh, N., and Pahari, R.: Impact of
- 842 tidal inundation on the net ecosystem exchange in daytime conditions in a salt marsh, Agric.
- 843 For. Meteorol., 294, 108133, https://doi.org/10.1016/j.agrformet.2020.108133, 2020.
- 844 Navarro, A., Young, M., Macreadie, P. I., Nicholson, E., and Ierodiaconou, D.: Mangrove
- and saltmarsh distribution mapping and land cover change assessment for south-eastern
- 846 Australia from 1991 to 2015, Remote Sens., 13, https://doi.org/10.3390/rs13081450, 2021.
- Osborne, C. P. and Freckleton, R. P.: Ecological selection pressures for C4 photosynthesis in
 the grasses. Proc. Roc. Soc. B, 276, https://doi.org/10.1098/rspb.2008.1762, 2009.
- Otani, S. and Endo, T.: CO₂ flux in tidal flats and salt marshes, Blue Carbon Shallow Coast.
 Ecosyst. Carbon Dyn. Policy Implement., 223–250, 2019.
- 851 Owers, C. J., Rogers, K. and Woodroffe, C. D.: Spatial variation of above-ground carbon
- storage in temperate coastal wetlands. Estuar. Coast. Shelf Sci., 210, 55-67,
- 853 https://doi.org/10.1016/j.ecss.2018.06.002, 2018
- 854
- <u>R Core Team: R: A Language Environment for Statistical Computing. Vienna, Australia,</u>
 <u>2024.</u>

- 857 Reents, S., Möller, I., Evans, B. R., Schoutens, K., Jensen, K., Paul, M., Bouma, T. J.,
- 858 Temmerman, S., Lustig, J., Kudella, M., and Nolte, S.: Species-specific and seasonal
- differences in the resistance of salt-marsh vegetation to wave impact, Front. Mar. Sci., 9,2022.
- 861 Rosentreter, J. A., Laruelle, G. G., Bange, H. W., Bianchi, T. S., Busecke, J. J. M., Cai, W. J.
- Eyre, B. D., Forbich, I., Kwon, E. Y., Maavara, T., Moosdorf, N., Najjar, R. G, Sarma, V. V.
 S. S., Van Dam, B. and Regnier, P.: Coastal vegetation and estuaries are collectively a
- greenhouse gas sink. Nat. Clim. Chang, 13, 579–587. https://doi.org/10.1038/s41558-023 01682-9, 2023.
- 866

- 869 Meteorol., 275, 223–232, https://doi.org/10.1016/j.agrformet.2019.05.026, 2019.
- 870 Seyfferth, A. L., Bothfeld, F., Vargas, R., Stuckey, J. W., Wang, J., Kearns, K., Michael, H.
- A., Guimond, J., Yu, X., and Sparks, D. L.: Spatial and temporal heterogeneity of
- geochemical controls on carbon cycling in a tidal salt marsh, Geochim. Cosmochim. Acta,
 282, 1–18, 2020.
- 874 Shepard, C. C., Crain, C. M., and Beck, M. W.: The protective role of coastal marshes: a
- 875 systematic review and meta-analysis, PLoS ONE, 6, e27374,
- 876 https://doi.org/10.1371/journal.pone.0027374, 2011.
- 877 Smith, J. A. M., Regan, K., Cooper, N. W., Johnson, L., Olson, E., Green, A., Tash, J., Evers,
- 878 D. C., and Marra, P. P.: A green wave of saltmarsh productivity predicts the timing of the
- annual cycle in a long-distance migratory shorebird, Sci. Rep., 10, 20658,
- 880 https://doi.org/10.1038/s41598-020-77784-7, 2020.
- 881 Vázquez-Lule, A. and Vargas, R.: Biophysical drivers of net ecosystem and methane
- 882 exchange across phenological phases in a tidal salt marsh, Agric. For. Meteorol., 300,
- 883 108309, https://doi.org/10.1016/j.agrformet.2020.108309, 2021.

Wang, Z. A., Kroeger, K. D., Ganju, N. K., Gonneea, M. E., and Chu, S. N.: Intertidal salt
marshes as an important source of inorganic carbon to the coastal ocean, Limnol. Oceanogr.,
61, 1916–1931, https://doi.org/10.1002/lno.10347, 2016.

- Ward, N. D., Megonigal, J. P., Bond-Lamberty, B., Bailey, V. L., Butman, D., Canuel, E. A.,
 Diefenderfer, H., Ganju, N. K., Goñi, M. A., and Graham, E. B.: Representing the function
- and sensitivity of coastal interfaces in Earth system models, Nat. Commun., 11, 2458, 2020.
- Wei, S., Han, G., Jia, X., Song, W., Chu, X., He, W., Xia, J., and Wu, H.: Tidal effects on
 ecosystem CO2 exchange at multiple timescales in a salt marsh in the Yellow River Delta,
 Estuar. Coast. Shelf Sci., 238, 106727, 2020.
- Whitfield, A. K.: The role of seagrass meadows, mangrove forests, salt marshes and reed
 beds as nursery areas and food sources for fishes in estuaries, Rev. Fish Biol. Fish., 27, 75–
 110, https://doi.org/10.1007/s11160-016-9454-x, 2017.
- Xiao, J., Sun, G., Chen, J., Chen, H., Chen, S., Dong, G., Gao, S., Guo, H., Guo, J., Han, S.,
 Kato, T., Li, Y., Lin, G., Lu, W., Ma, M., McNulty, S., Shao, C., Wang, X., Xie, X., Zhang,

Formatted: Font: (Default) Times New Roman
Formatted: Font: (Default) Times New Roman
Formatted: Font: (Default) Times New Roman
Formatted: Font: (Default) Times New Roman
Formatted: Font: (Default) Times New Roman, Not Italic
Formatted: Font: (Default) Times New Roman
Formatted: Font: (Default) Times New Roman, Not Italic
Formatted: Font: (Default) Times New Roman
Formatted: Font: (Default) Times New Roman, Not Bold
Formatted: Font: (Default) Times New Roman
Formatted: Font: (Default) Times New Roman
Formatted: Hyperlink, Font: (Default) Times New Roman, Font colour: Auto, Pattern: Clear
Field Code Changed
Deleted: ¶

<sup>Schäfer, K. V. R., Duman, T., Tomasicchio, K., Tripathee, R., and Sturtevant, C.: Carbon
dioxide fluxes of temperate urban wetlands with different restoration history, Agric. For.</sup>

- X., Zhang, Z., Zhao, B., Zhou, G., and Zhou, J.: Carbon fluxes, evapotranspiration, and water use efficiency of terrestrial ecosystems in China, Agric. For. Meteorol., 182–183, 76–90, https://doi.org/10.1016/j.agrformet.2013.08.007, 2013.