

- 2018; Roshan et al. 2018), suggesting that rapid Zn removal into the particulate phase is a key
- feature of biogeochemical cycling in the Southern Ocean.
- The rapid removal of dZn from the surface within the Southern Ocean suggests the possibility for phytoplankton growth to become Zn-limited. While phytoplankton growth in the Southern Ocean is well-known to be primarily limited by Fe availability (Martin 1990; Arrigo et al. 2008), melting icebergs and ice shelves are known to act as external sources of Fe (St-Laurent et al. 2017; Hopwood et al. 2019; Person et al. 2021) with larger Fe inputs expected from increased ice melt in a warming climate. Increased dFe inputs to surface Antarctic waters may act to relieve Fe stress, but would simultaneously support the development of other nutrient 102 limitations. For example, low availabilities of both dZn and vitamin B_{12} have been previously observed to co-limit phytoplankton growth with Fe in the Ross Sea (Bertrand et al. 2007; Kell et al. 2023). A high demand for Zn naturally exists within eukaryotic phytoplankton due to the 105 requirement for Zn^{2+} in numerous metabolic functions— therefore, without similarly enhanced inputs of dZn to the water column, the alleviation of primary Fe limitation could induce Zn stress as the next most in-demand metal micronutrient. Coastal polynyas that form within the Amundsen and Ross Seas during austral spring and summer are particularly primed to experience Zn stress as these regions host highly productive seasonal phytoplankton blooms that 110 act as significant carbon sinks (Arrigo et al. 2012). This high productivity draws $pCO₂$ down to low levels (< 200 ppm), putting pressure on the carbon concentrating mechanism of photosynthetic phytoplankton to acquire CO² and thus to acquire Zn as the predominantly utilized metal cofactor within carbonic anhydrases. The present study enhances our knowledge of what constitutes the anticipated high levels
- of Zn uptake in the Southern Ocean with empirical field data measured within native Southern

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- **2 Materials and methods**
- **2.1 Study area and sample collection**
- Samples were collected during the CICLOPS (Cobalamin and Iron Co-Limitation of
- Phytoplankton Species) expedition (NBP18-01) aboard the RVIB *Nathaniel B. Palmer*,
- December 11, 2017 March 3, 2018 in the Amundsen Sea and Ross Sea of the Southern Ocean
- (**Fig. 1**).

 Figure 1. Map showing the stations sampled over the course of the CICLOPS cruise. Stations marked by red triangles indicate those at which stable 67Zn and 110Cd uptake rate experiments

were performed. An expanded map of stations sampled in the Ross Sea is shown at bottom left,

- while a further expansion of stations sampled in Terra Nova Bay is shown at bottom right.
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- clean (TMC) sampling protocols described previously (Cutter and Bruland 2012). A TMC
- rosette suspended on a Kevlar line and equipped with twelve 8L X-Niskin bottles (Ocean Test
- Equipment) was used to collect seawater at depths ranging from 10 600 m. Continuous

- 156 underway measurements of $pCO₂$ measurements at ~5 m depth were taken using a $pCO₂$
- measurement system from Lamont-Doherty Earth Observatory (LDEO, 0.017/sec rate).
- Hydrography data were collected using sensors deployed on a titanium trace metal rosette
- (TMR) in tandem with TMC niskin bottles. The TMR was equipped with sensors to measure
- temperature, conductivity, pressure, dissolved oxygen, chlorophyll (Chl) fluorescence, altimetry,
- beam transmission, and photosynthetically active irradiance (PAR). Chl fluorescence was
- measured using a WetLabs ECO-FL fluorometer. A complete data report and sensor list are
- available at NBP1801DATA.pdf (rvdata.us). Mixed layer depth (MLD) was determined for each
- station within Terra Nova Bay as the first depth at which the difference between the potential
- 165 density (σ_θ) and reference density (the potential density at 10m, σ_{ref}) was greater than or equal to
- 166 0.125 kg m^3 (Bishop and Wood 2009; Ohnemus et al. 2017).
- **2.2 Preparation of plasticware**
- Polyethylene and polycarbonate sampling and incubation bottles were rigorously cleaned
- to remove trace metal contaminants before use. Bottles with rinsed with Milli-Q water
- (Millipore), soaked for 72h in <1% Citranox detergent, rotated, soaked for an additional 72h, and
- then rinsed five times with Milli-Q water. Bottles were then filled with 10% HCl (Baker instra-
- analyzed) by volume and soaked for a minimum of one week, rotated, and soaked for another
- week. Bottles were then rinsed five times with dilute acid (HCl, pH 2) and stored double-bagged
- in plastic zip bags. All cleaning work was conducted in a Class 100 clean room.
- **2.3 Analyses of total dissolved Cd and Zn using isotope dilution**
- Samples for the analysis of total dissolved Zn, Cd, Fe, Mn, Cu and Ni concentrations
- were collected shipboard by pressure-filtering X-Niskin bottles through an acid-washed 142mm,
- 0.2µM Supor membrane filter (Pall) within 3 hours of rosette recovery using high purity

nitric acid rinse solution (Optima grade, Fisher Chemical), and a 10% nitric acid elution buffer

244 **Figure 2.** Diagram showing the overall workflow used to measure particulate uptake of ¹¹⁰Cd 245 and Zn and total dissolved Cd and Zn, after Cox et al. 2014.

247	Uptake experiments were performed at 18 stations total (Fig. 1). Raw (unfiltered)
248	seawater was collected shipboard over a depth range of $10 - 600$ m into 250 mL TMC
249	polycarbonate incubation bottles. All incubation bottles were filled with minimal headspace such
250	that the total culture volume was \approx 275 mL. Two incubation bottles per depth were filled with
251	raw seawater— one was spiked with ${}^{67}Zn$, the other was spiked with ${}^{110}Cd$. The Cd and Zn
252	isotope spikes were prepared by dissolving ¹¹⁰ CdO and ⁶⁷ ZnO (Oak Ridge National Laboratory)
253	in 5% HNO ₃ (Seastar Baseline) and were diluted using Milli-Q water to minimize added acidity.
254	When added to the filled incubation bottles, the total added (spiked) concentration of Cd was 300
255	pM and the total added concentration of Zn was 2 nM. The chosen total added concentrations
256	were based on the surface ratio of total dissolved Cd (dCd) to total dissolved Zn (dZn) reported
257	previously for the Ross Sea (Fitzwater et al. 2000). Immediately after spiking, incubation bottles

- were sealed, inverted to mix, and transferred to flow-through on-deck incubators for 24hr. Incubators were shielded by black net neutral density screening to allow 20% ambient light
- penetration.

 Biomass was collected after 24hr by vacuum filtering the entire volume of each incubation sample at 34.5 kPa (5 psi) onto an acid-cleaned 3m, 50mm acrylic copolymer (Versapore) filter (Pall) mounted on an acid-cleaned Teflon (Savillex) filtration rig. Samples 264 were filtered through 3 μ m pore-size filters rather than 0.2 μ m in order to minimize filtration time (and thus time exposed to potential contamination) and to capture the bulk of eukaryotic phytoplankton biomass typically found in the Southern Ocean. An aliquot of 1 mL of 0.2 m filtered surface seawater (collected at 10 m depth) was used to rinse the sample before collecting the filter into an acid-cleaned 2 mL cryovial using acid-rinsed plastic forceps. Filter blanks were 269 duplicate 3 µm acid-clean Versapore (Pall) filters that were placed onto the filtration rig, rinsed with filtered surface seawater, collected, stored, and processed as samples were to correct for any contaminating metals present on the filters themselves. Blanks were collected at each station. 272 Filters were stored frozen at -80 °C in acid-cleaned cryovials until analysis. The filtration rig was rinsed with pH 2 HCl between samples. Polycarbonate incubation bottles were cleaned between stations with a 10% HCl rinse and several rinses in Milli-Q water, followed by a brief soak in 10% HCl followed by a pH 2 HCl rinse. All spike addition and sample filtration procedures were completed in a fabricated shipboard positive-pressure clean room environment made of laminar flow hoods and plastic sheeting. We note that the total Zn and Cd uptake rate values presented in this study represent

 potential uptake rates rather than true uptake rates—this naturally arises as a consequence of 280 adding the spiked tracer ^{67}Zn and ^{110}Cd into raw surface seawater. As this seawater is naturally

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- 304 concentrations were simultaneously measured by ICP-MS and were calibrated to a standard
- 305 curve ranging from 100 3,200 ppb using a 1 ppm certified P stock (Alfa Aesar Specpure).

306 Equation #2 was used for the calculations described above:

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M_{\text{particulate}} = \left[\frac{M_{\text{sample}}}{I n_{\text{sample}}} - \frac{M_{\text{blank}}}{I n_{\text{blank}}}\right] * \frac{I n_{\text{digestion}}}{M_{\text{slope}}} * \frac{V_{\text{digested}}}{V_{\text{fittered}}} \tag{2}
$$

308 where *V*_{filtered} is the total spiked sample volume estimated to have passed through the filter (275 309 mL), *V*_{digested} is the final volume the sample was resuspended in (2.0 mL), M_{sample} is the metal of 310 interest measured in the sample in units of counts per second (cps), *M*blank is the metal of interest 311 measured in the process blanks (cps), *M*slope is the slope of the metal of interest obtained by the standard curve (cps ppb⁻¹), In_{sample} is the In measured in the sample (cps), In_{blank} is the In 313 measured in the process blanks (cps), In_{disp} is the cps of In measured in the 5% HNO₃+1 ppb 314 In digestion solution, and the calculated concentration of the metal of interest $(M_{\text{particular}})$ is in 315 ppb (μ g L⁻¹). This equation is the same as that used by Noble et. al. 2013 for the determination 316 of particulate metal concentrations using ICP-MS (Noble et al. 2013). 317 The Zn spike and Cd spike were also analyzed by ICP-MS using a tenfold dilution of 318 spike solution into 5% $HNO₃$ containing 1 ppb In to determine isotopic compositions and 319 concentrations. When added to filled incubation bottles (275 mL total volume), the added 320 concentrations were 288 pM $\rm{^{110}Cd}$, 4.51 pM $\rm{^{111}Cd}$, and 1.69 pM $\rm{^{114}Cd}$ for Cd spiked bottles, and 321 were 1.91 nM ⁶⁷Zn, 0.045 nM ⁶⁶Zn, and 0.047 nM ⁶⁸Zn for Zn spiked bottles (**Table S2**). For all 322 stations and all depths, ${}^{67}Zn$ and ${}^{110}Cd$ spike concentrations exceeded natural dissolved ${}^{67}Zn$ and $110¹¹⁰$ 323 Cd concentrations, estimated by multiplying the total dissolved Zn and Cd by the natural

324 isotope abundance of ⁶⁷Zn and ¹¹⁰Cd (0.0410 and 0.1249, respectively; see comparisons in **Fig.**

325 **S2**).

2.6 Calculating zinc and cadmium uptake using ⁶⁷Zn and ¹¹⁰ 326 **Cd**

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$$
Cd_{\text{total}} \text{ Uptake Rate (pmol L-1 d-1)} = \frac{[^{110}Cd_{\text{Sample}} (pmol L-1 d-1) - ^{110}Cd_{\text{PEP}} (pmol L-1 d-1)]}{[^{110}Cd_{\text{Spike}} (pmol L-1) + ^{110}Cd_{\text{Natural}} (pmol L-1)]} \times Cd_{\text{total}} (pmol L-1) (3)
$$

 Zn_{total} Uptake Rate (pmol L⁻¹ d⁻¹) = $\frac{[67Zn_{\text{sample}} \text{ (pmol L}^{-1} d^{-1}) - 67Zn_{\text{PEP}} \text{ (pmol L}^{-1} d^{-1})]}{[677n_{\text{max}} \text{ (pmol L}^{-1}) - 67Zn_{\text{PEP}} \text{ (pmol L}^{-1} d^{-1})]}$ 351 Zn_{total} Uptake Rate (pmol L⁻¹ d⁻¹) = $\frac{[67\text{Zn}_{\text{Smple}} \text{ (pmol L} \cdot \text{d} \cdot \text{)} - 67\text{Zn}_{\text{N}} \cdot \text{mol L} \cdot \text{d} \cdot \text{)}}{[67\text{Zn}_{\text{Spike}} \text{ (pmol L} \cdot \text{d} \cdot \text{)} + 67\text{Zn}_{\text{Natural}} \text{ (pmol L} \cdot \text{d} \cdot)]}$ x Zn_{total} (pmol L

2.7 Nutrient analyses

 Seawater samples taken for macronutrient analysis were filtered through 0.2 μm Supor (Pall) membrane filters and frozen at sea in acid-washed 60-mL high-density polyethylene (HDPE) bottles until analysis. Nutrient analyses were conducted by nutrient autoanalyzer by Joe Jennings at Oregon State University using previously described methods (Noble et al. 2012). **2.8 Statistics and plotting** Dissolved ecological stoichiometries were obtained from the slopes of two-way (type II) least squares linear regressions performed using the script lsqfitma.m rewritten from MATLAB to Python by Rebecca Chmiel (https://github.com/rebecca-chmiel/GP15). A correlation matrix of various parameters measured during NBP18-01 was created with SciPy v1.5.2 using the scipy.stats.pearsonr function, yielding Pearson correlation coefficients and p values that were visually represented using Seaborn v.0.11.1 and Matplotlib v3.3.2. Ocean sections were plotted

using Ocean Data View v5.3.0 with gridded bathymetry file ETOPO1_2min. Outliers (see Data

Availability) were excluded from ocean sectional plots. Mixed layer depth was calculated using

the potential density function (pden) within the python-seawater module (v3.3.4). Figures were

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367 made using matplotlib (v3.3.2), Ocean Data View (v5.5.2), Excel (2019), and RStudio
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(v1.3.1093). ODV color palettes (https://doi.org/10.5281/zenodo.1243862) are inverse 'roma' for

trace metal and macronutrient concentrations, 'thermal' for Zn and Cd uptake rates, and 'algae'

- for total fluorescence (Crameri 2023).
- **3 Results**
- **3.1 Amundsen Sea**

 Zn and Cd uptake rate experiments were conducted at 18 stations. We define 3 groups of stations based on location: the Amundsen Sea, Ross Sea, and Terra Nova Bay (TNB) groups (**Fig. 1**). Uptake rates were assessed at 3 stations (4, 11 and 15) within the Amundsen Sea group, 6 stations (20, 29, 32, 35, 62, and 67) within the Ross Sea group, and 9 stations (22, 27, 41, 46, 377 52, 57, 72, 76 and 79) within the TNB group spanning $\sim 10 - 250$ m depth for a total of 18 stations and 125 samples. An overall schematic detailing these experiment workflows is shown $\frac{379}{10}$ in Fig. 2. The experimental design was validated by comparison of surface particulate ⁶⁷Zn:⁶⁸Zn and 110 Cd: 114 Cd ratios measured in spiked samples with those measured in control (unspiked) 381 samples. Samples spiked with ${}^{67}Zn$ had particulate ${}^{67}Zn$:⁶⁸ Zn ratios larger than natural abundance 382 ratios at all stations (as was also true for ¹¹⁰Cd spiked samples and Cd natural abundances; **Fig. S1**), indicative of uptake of the spike into the particulate phase. The Amundsen Sea stations represented a linear cruise track, and we report total 385 dissolved metal concentrations (dMetal_T) and uptake rates (*ρ*Metal) over time in order of station sampling date (**Fig. 3a**).

 Figure 3. Total fluorescence and trace metal concentrations measured at Amundsen Sea stations shown over time. (a) Map showing station locations, (b) total chlorophyll (Chl) fluorescence, (c) 390 total Zn uptake rates, (d) total Cd uptake rates, (e) total dissolved Zn, (f) total dissolved Cd, (g) total dissolved Fe, and (h) total dissolved Mn measured in the upper 250 m represented in color scale. Uptake experiments were not performed at station 10. Metal concentrations measured to 500 m depth are shown in Figure S3. dZnT, total dissolved Zn; dCdT, total dissolved Cd; dFeT, total dissolved Fe; dMnT, total dissolved Mn.

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396 Among these stations, total Chl fluorescence was lowest at station 4 and increased 
397 moving westward along the transect to a Chl maximum of 41.8 mg m<sup>-3</sup> at station 15, 10 m (Fig.
398 3b). Maximum surface concentrations of dZn, dCd and dMn were highest at station 4 (3.5 nM,
399 639 pM, and 2.6 nM at 10 m, respectively; Fig. 3e, f, h), likely reflecting the relatively smaller
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 amount of total biomass (as indicated by total Chl fluorescence; **Fig. 3b**) at this station. Concentrations of dZn, dCd and dMn decreased moving westward along the transect (**Fig. 3e, f, h**) as total Chl fluorescence increased (**Fig. 3b**). Total Zn uptake rates (*ρ*Zn) and total Cd uptake 403 rates (ρ Cd) were highest at station 11 (158 and 21 pmol L⁻¹ d⁻¹, respectively, at 10 m; **Fig. 3c,d; Fig. 4b**). Among the three Amundsen Sea stations, the largest movement of both Zn and Cd into 405 the particulate phase therefore occurred at station 11, concurrent with the relatively higher dF e_T surface values observed at station 11 (0.2 nM dFe compared to 0.01 nM at station 4, 10 m; **Fig. 3g, Fig. S3c**). The dFe concentrations exceeding 1 nM near the seafloor are consistent with a sedimentary or subglacial source (**Fig. S3c**). Overall, *ρ*Zn and *ρ*Cd profiles exhibited trends in which values were highest within the upper 50 m at all three stations and decreased with depth, following the trend in Chl a or total Chl fluoresence (**Fig. 4**). Vertical sections of dZn and dCd through the water column mirrored these trends (**Fig. 4**), demonstrating the movement of these dissolved metal micronutrients into the particulate phase.

 Figure 4. Depth profiles of total Zn and Cd uptake rates, total chlorophyll fluorescence (or chlorophyll *a*) and total dissolved metal measured in the upper 250 m at (a) station 4, (b) station 11, and (c) station 15 sampled along the Amundsen Sea shelf. Total chlorophyll (Chl) fluorescence is reported for stations where chlorophyll *a* (Chl a) data was not measured.

3.2 Ross Sea

We next investigated the dissolved Zn and Cd demand of the natural phytoplankton

community at stations sampled over the Ross Sea shelf. Data collected from this group is

- presented over time, in order of sampling date (**Fig. 5**).
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 Figure 5. Total fluorescence and trace metal concentrations measured at Ross Sea stations shown over a latitudinal transect. (a) Map showing station locations, (b) total chlorophyll (Chl) fluorescence**,** (c) total Zn uptake rates, (d) total Cd uptake rates, (e) total dissolved Zn, (f) total dissolved Cd, (g) total dissolved Fe, and (h) total dissolved Mn measured in the upper 250 m represented in color scale. Uptake experiments were not performed at stations 31, 34, and 70. Metal concentrations measured to 800 m depth are shown in Figure S4. dZnT, total dissolved Zn; dCdT, total dissolved Cd; dFeT, total dissolved Fe; dMnT, total dissolved Mn.

We note that unlike the Amundsen Sea sector, the stations sampled in this group did not

 Figure 6. Depth profiles of total Zn and Cd uptake rates, total chlorophyll fluorescence (or, 453 where available chlorophyll a), and total dissolved metal $(dMetal_T)$ measured in the upper 250 m at (a) station 20, (b) station 29, (c) station 32, (d) station 35, (e) station 62, and (f) station 67 sampled along the Ross Sea shelf. Total chlorophyll (Chl) fluorescence is reported for stations where chlorophyll a (Chl a) data was not measured.

3.3 Terra Nova Bay

Zinc and Cd uptake rate data collected from stations sampled in Terra Nova Bay (TNB)

were visualized over time due to repeated sampling within a small geographic region and similar

- timeframe (**Fig. 7a**). This allowed for an analysis of how dissolved metal concentrations and
- metal uptake rates changed throughout January-February 2018 within the same spatial area.
- Station data is presented in order of sampling date, from the earliest (station 22, sampled in early
- January) to the latest (station 79, sampled in late February).

 Figure 7. Total fluorescence and trace metal concentrations measured at Terra Nova Bay (TNB) stations shown over time. (a) Map showing station locations, (b) total chlorophyll (Chl) 469 fluorescence, (c) total Zn uptake rates, (d) total Cd uptake rates, (e) total dissolved Zn, (f) total dissolved Cd, (g) total dissolved Fe, and (h) total dissolved Mn measured in the upper 250 m dissolved Cd, (g) total dissolved Fe, and (h) total dissolved Mn measured in the upper 250 m represented in color scale. Uptake experiments were not performed at stations 70 and 34. Metal concentrations measured to 600 m depth are shown in Figure S5. dZnT, total dissolved Zn; dCdT, total dissolved Cd; dFeT, total dissolved Fe; dMnT, total dissolved Mn.

475 Surface Chl fluorescence was highest in early January $({\sim}18 \text{ mg m}^{-3})$ and waned into

February (**Fig. 7b**), similar to observed trends in Zn and Cd uptake rates (**Fig. 7 c,d**). Of all TNB

stations, stations 22 and 27, sampled in January, had the highest maximum Zn uptake rates of

478 89.9 pmol L^{-1} d⁻¹ and 46.0 pmol L^{-1} d⁻¹, respectively, at 10 m (**Fig. 8a,b**). Cd uptake rates were

 Figure 8. Depth profiles of total Zn and Cd uptake rates, total chlorophyll fluorescence (or, where available, chlorophyll a), and total dissolved metal (dMetalT) measured in the upper 250 m at (a) station 22, (b) station 27, (c) station 41, (d) station 45, (e) station 52, (f) station 57, (g) station 72, (h) station 76, and (i) station 79 within Terra Nova Bay. Total chlorophyll (Chl) fluorescence is reported for stations where chlorophyll a (Chl a) data was not measured.

 The increased surface concentrations of dZn and dCd and macronutrients, as well as the persistence of measurable uptake rates at deeper depths, at these late TNB stations may be attributed to the deepening of the mixed layer (**Fig. S7**). Vertical mixing was evidenced by more uniform potential densities, temperatures, dissolved oxygen (O2) concentrations, salinity, and beam transmission measurements at the late TNB stations within the upper 200 m (**Fig. S7**). Higher (>0.5 nM) dFe concentrations were also observed below 100 m at these late stations (**Fig. 7g**) and increased with depth $(>2 \text{ nM})$, as did dZn and dMn concentrations, possibly due to sedimentary inputs (Giordano et al. 1999) (**Fig. S5**). At these late stations (Station 76, 78, 79) mixing replenished surface concentrations of both macronutrients (**Fig. S6**) and dZn (**Fig. S5a**),

- but dZn was replenished to a lower extent. For example, comparing 50 m "replenished" surface
- values of P, N+N, and Si to deepwater (200 m) values at Station 79, percent changes from deep
- to surface values were -0.35% for P, -0.30% for N+N, and -0.26% for Si (a % change of 0 would
- indicate complete replenishment; i.e, if values at 200 m and at 50 m were equal). In contrast, the
- percent change from deep (200m) to surface (50m) dZn at Stn79 was lower, -0.71%. Hence, dZn
- was apparently replenished to a lesser extent compared to macronutrients, which may reflect a
- sustained high demand for Zn generating a dearth of this micronutrient despite macronutrient
- replenishment.
- **4 Discussion**
- **4.1 Overview of Zn and Cd uptake at 18 stations**

 Maximum Zn and Cd uptake rates observed at each station (all of which were observed at 525 \leq 10 m depth; **Fig. 9a,b**) with uptake rates normalized to Chl a (μ g/L) as a proxy for biomass (**Fig. 9c,d).**

Figure 9. Unnormalized (a) maximum Zn uptake rates (ρZn_{Max}) and (b) maximum Cd uptake 529 rates (ρCd_{Max}) at each station grouped by area (Amundsen Sea, Ross Sea, Terra Nova Bay). (c) *530 ρ*Zn_{Max} and (d) *ρ*Cd_{Max} normalized to chlorophyll a (μ g L⁻¹) measured at each station. (e) Depth integrated (10 m-250 m) *ρ*Zn and *ρ*Cd values at each station. ND, no data (chlorophyll a not measured).

4.2 Use of metal uptake rates to determine depletion timeframes

 The measurement of total dissolved metal concentrations over large latitudinal or longitudinal areas allows for the characterization of metal inventories, though these are snapshots of inventories observed at specific times. The measurement of metal uptake rates allows us to gain new insight into how these inventories came to be and the timeframes over which they are

 consumed and replenished. Due to the resetting of surface dissolved metal concentrations to those of deepwater values during austral winter with deep winter mixing, the Ross Sea of the Southern Ocean is particularly applicable to this type of timeframe study (Sedwick and DiTullio 1997; Sedwick et al. 2011).

 Using the Zn uptake rates measured in this study, we can estimate the time required for the high levels of primary production observed in the Southern Ocean to draw down surface dZn from high (deep water) winter concentrations to the surface concentrations observed during austral summer 2017. The Southern Ocean growing season typically spans October-March, with primary productivity peaking November-January and the area of open (ice-free) water over the Ross Sea shelf linearly increasing from November-mid January (Sedwick et al. 2011). Vertical profiles of nutrients and micronutrients in coastal Antarctic ecosystems such as the Ross Sea are reset and become uniform with depth during the winter months due to whole-water column mixing and an absence of photosynthetic activity during the dark winter under the sea ice (Noble et al. 2013). As a result, the drawdown of nutrients in the upper water column observed during the spring and summer seasons is the result of less than one year's biological influence. For this 571 simple calculation, we ignore the upward flux of Zn (upwelling $= 0$) and assume a high export ratio of 0.8 due to bloom productivity being dominated by diatoms and *Phaeocystis antarctica*, both of which sink rapidly and thus contribute substantially to carbon export flux (Asper and Smith 1999; DiTullio et al. 2000). The depletion of dZn from a surface box was therefore estimated as:

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\left(\frac{dZn}{dt}\right)_{surface\ box} = -\rho Zn + (Rf * \rho Zn) + upwelling
$$

Where Rf is the remineralization factor equal to 1 - export ratio.

- Taking station 11, for which the highest Zn uptake rate was observed, as an extreme case:
- 579 with a maximum Zn uptake rate of 158 pmol $L^{-1} d^{-1}$, it would take only 25 days to deplete a
- surface winter concentration of 4.8 nM (that is, the average deepwater (< 200 m) dZn
- concentration for all stations measured in this study) down to the observed surface concentration
- of 1.7 nM at station 11, assuming a constant uptake rate and no additional inputs of dissolved Zn
- (**Fig. 10**).

 Figure 10. A simple model estimating the time (in days) required to deplete the estimated average winter surface concentration of dZn (4.8 nM) over a range of various Zn uptake rates (UR). 158 pmol/L/d was the maximum Zn uptake rate observed in this study (station 11, 10 m).

 Given that dZn surface depletion to sub-nanomolar levels was observed throughout much of the CICLOPS expedition, prolonged high levels of Zn uptake and export that overwhelm replenishment by vertical mixing and/or remineralization are likely key to giving rise to the observed extent of seasonal surface dZn depletion. These calculations were conducted as a proof- of-concept to determine if uptake rates were sufficient to draw down the otherwise abundant dZn inventory on seasonal timescales. Future studies could conduct mesoscale modeling of the region, replacing upwelling including eddy diffusion and advection. Notably, any dZn upwelling

- flux into the euphotic zone would require even higher Zn uptake rates to create the seasonal
- surface Zn depletion we observed on this expedition.
-

4.3 Influences on Zn and Cd uptake

- We next consider the factors driving the magnitude of *ρ*Zn and *ρ*Cd. As noted above, *ρ*Zn
- and *ρ*Cd were positively correlated with total Chl fluorescence or Chl a at every station (**Fig. 4;**
- **Fig. 6; Fig. 8**), demonstrating the influence of total autotrophic biomass on uptake rates. A
- Pearson correlation analysis comparing the abundance of individual algal pigments to *ρ*Zn and
- *ρ*Cd throughout the water column for all stations revealed significant, positive correlations
- 605 (Pearson correlation coefficient > 0.50 , $p \le 1.2e-4$) between pZn and Chl *a*, Chl *b*, and Chl c1, c2
- 606 and c3. Pearson correlation coefficients are normally symbolized as rho (ρ) , but to avoid
- confusion with our uptake symbol (*ρ*), and with p-values (p), they are herein referred to as 'cc'
- 608 values. The correlation between pZn and Chl *b* was strongest (cc= 0.77, p = 3.8e-10) of any
- pigment (**Fig. 11e**).

 Figure 11. Relationships comparing seawater CO² partial pressure (pCO2) at 5 m depth to (a) Zn 612 uptake rates (ρZn ; n=15, R² = 0.63) and (b) Cd uptake rates (ρ Cd; n=15) measured at surface 613 (\leq 10 m) depths. (c) Relationship between ρZn and ρCd for all depths (n=121, R² = 0.64). (d) Visual representation of the correlation matrix comparing all water column parameters measured with depth with warm and cool colors indicative of positive and inverse correlations, respectively. Pearson correlation coefficients and p values are shown. (e) Representation of the correlation matrix comparing *ρ*Zn and *ρ*Cd to various phytoplankton pigments. Fe, Mn, Ni, Cu, Zn, Cd, and Co labels correspond to total dissolved metal concentrations. PO4, N+N, and Si correspond to total dissolved concentrations of phosphate, the sum of nitrate+nitrite, and silicate. Temp, temperature; cond, conductivity; O2, dissolved oxygen; Fsu, total fluorescence; PAR, photosynthetically active radiation; Sal, salinity. Chl_a, chlorophyll *a*; Chl_b, chlorophyll *b*; Chl_c1, chlorophyll c1; Chl_c2, chlorophyll c2; Chl_c3, chlorophyll c3; chl_lide, chlorophyllide; Fuco, fucoxanthin; Hex_19, 19'-hexanoyloxyfucoxanthin; Diadino, diadinoxanthin; Diato, diatoxanthin.

- In bottle incubation experiments conducted at station 27, the addition of Zn alone resulted
- in the positive growth response of Chl *b*-containing algae (small green algae such as

- 651 surface Chl fluorescence (19.3 mg m⁻³ at 10 m), low pCO₂ (221 µatm at 5 m), and high
- 652 maximum Zn and Cd uptake rates (46 and 20 pmol L^{-1} d⁻¹, respectively), demonstrating a high
- 653 algal demand for Zn that likely created pressure for Cd uptake.
- 654 We next consider the effect of the depleted seawater $pCO₂$ levels induced by the high 655 biomass conditions observed on this expedition. Previously, a strong correlation between 656 dissolved δ^{114} Cd and dissolved CO₂ was documented in the Atlantic Sector of the Southern 657 Ocean (de Baar et al. 2017), suggesting significant Cd isotope fractionation due to biological 658 uptake into the particulate phase. A relationship between total surface Cd uptake rates at 10 m 659 and surface pCO2 (underway, measured at 5 m) was not observed in the present study (**Fig. 11b**). 660 The present study includes measurements of total Cd uptake (that is, the sum of all Cd isotopes) 661 using an added Cd isotope tracer, and hence did not explore natural isotope fractionation effects. 662 However, we did observe a significant negative linear relationship between total Zn uptake rates 663 and seawater pCO₂ (m = -0.58; R² = 0.63; Fig. 11a) consistent with an increased demand for Zn^{2+} 664 to power the carbon concentrating mechanism of photosynthetic algae under lower $CO₂$ 665 availability. *ρ*Zn and *ρ*Cd furthermore shared a significant positive linear relationship with each 666 other (m = 0.13; R^2 = 0.64; **Fig. 11c**) (as was also reflected in the Pearson correlation test; PCC = 667 0.69, p = 1.7e-7, **Fig. 11d**) implying that as demand for Zn increased, demand for Cd also 668 increased, consistent with laboratory studies showing their co-transport in marine algae (Sunda 669 and Huntsman 2000). We also note that *ρ*Cd:*ρ*Zn uptake ratios were higher (> 0.4) at the surface 670 where total dissolved dCd:dZn ratios were comparatively higher (> 0.3) (**Fig. 12a,b**). The strong 671 positive linear relationship shared between these ratios $(R^2 = 0.82; Fig. 12c)$ further suggests that 672 dZn levels were depleted enough to induce increased Cd uptake rates, and is consistent with their 673 known biochemical substitution within marine algae.

 Figure 12. (a) Cd:Zn uptake ratios (*ρ*Cd:*ρ*Zn) and (b) total dissolved Cd:Zn ratios (dCd:dZn) for all stations during the CICLOPS expedition measured in the upper 250 m represented in color scale and over time of sampling. (c) Two-way linear regression showing the positive relationship 678 between dCd:dZn and ρ Cd: ρ Zn inclusive of all stations and depth (n=111, R² = 0.82).

680 Algal Cd uptake rates are known to be inversely related to both Mn^{2+} and Zn^{2+}

concentrations in culture (Lee et al. 1995; Sunda and Huntsman 1996), which is thought to

reflect the uptake of Cd by two separate inducible transport systems. Cadmium is taken up

683 competitively by the high-affinity Zn uptake system under low Zn^{2+} conditions, as demonstrated

- 684 above, while Cd, Zn, and Mn share the same low-affinity Mn uptake system under high Zn^{2+}
- conditions (Lee et al. 1995; Sunda and Huntsman 1998a; b, 2000; Xu et al. 2007). With the
- exception of the Amundsen Sea stations, dMn was consistently observed at concentrations of
- only 0.1-0.5 nM within the upper 50 m (**Fig. 3h; Fig. 5h; Fig. 7h).** Low surface dMn

4.4 dZn and dCd relationships with macronutrients

 The growth of phytoplankton and bacteria in the shallow euphotic zone results in the removal of bioactive trace metals and macronutrients from the dissolved phase into the particulate phase, resulting in dissolved metal:macronutrient relationships that reflect their collective stoichiometry (Horner et al. 2021). Positive linear slopes result generally indicate the co-cycling of the metal and the macronutrient via uptake and remineralization, though slope values can vary widely by basin as they are a function of the metal:macronutrient uptake and remineralization stoichiometry of the native community and overall nutrient availability. Two- way linear regressions (see Methods) were used to investigate the relationships between dZn and dissolved silicate (dSi), dZn and dissolved phosphate (dP), and dCd and dP for the Amundsen Sea, Ross Sea, and TNB station groups (**Fig. 13**).

 Figure 13. Relationships between (Top row) total dissolved Zn and silicate (dSi), (Middle row) 710 total dissolved Zn and phosphate $(dPO₄³)$, and (Bottom row) total dissolved Cd and $dPO₄³$ for surface (blue squares) and deep ocean (orange circles) arranged by station group (Amundsen Sea, Ross Sea, and Terra Nova Bay). Depth thresholds were manually chosen to optimize the 713 linear fit of the surface and deep ocean trends. Regressions with an $R^2 \ge 0.50$ are shown as a 714 solid line, and those with an R^2 < 0.50 are shown as a dotted line. See **Table S4** for stoichiometric parameters and values. Regression outliers are marked with an 'x'. Data originally plotted in Chmiel et al. 2023 and reprised here for ease of comparison with dZn:Si data.

 The dZn:dP and dCd:dP relationships from this expedition were originally presented in Chmiel et al. 2023 for comparison to dCo:dP, while they are included in the present study for ease of comparison with dZn:dSi relationships presented for the first time. For these analyses, the depth threshold that separates the surface and deep ocean was manually defined in order to optimize the linear fit of the surface versus deep trends— this threshold depth can be thought of

as an inflection point that represents the largest change in trace metal concentration with respect

to dP or dSi concentration (Chmiel et al. 2023).

 As noted above, the near-linear global dZn:dSi relationship (Bruland et al. 1978; Vance et al. 2017; Middag et al. 2019) has been posited to arise, in part, from faster drawdown of Zn 727 and Si relative to $dPO₄³$ into Southern Ocean diatoms that leaves surface waters Zn and Si depleted (Vance et al. 2017). We observed distinct differences in dissolved dZn:dSi ecological stoichiometries comparing Amundsen Sea, Ross Sea and Terra Nova Bay station groups (**Fig. 13; Table S4**). A positive linear dZn:dSi relationship with a steep (m = 0.23 ± 0.05; **Table S4**) slope observed in the upper ocean of the Amundsen Sea contrasted starkly with the shallow slopes observed in the upper ocean of the Ross Sea and Terra Nova Bay. A bloom of non- silicifying *Phaeocystis antarctica* was present during our passage through the Amundsen Sea, consistent with abundant silicic acid yet rapid drawn down of Zn, which is known to be used by this organism (Saito and Goepfert 2008). In contrast, the shallow slopes in the Ross Sea and 736 Terra Nova Bay resulted from the persistence of dSi concentrations $\geq 30 \mu M$ in the upper 30 m 737 while dZn was reduced to sub-nanomolar concentrations (average $dZn = 0.87 \pm 0.42$ nM in TNB at 10 m depth, n= 11), highlighting the intense drawdown of dZn by biota in this region to meet a high metabolic dZn demand. Similar trends were observed for dZn:dP and dCd:P, which exhibited shallow slopes within the upper ocean of the Ross Sea and Terra Nova Bay. Southern Ocean diatoms are known to have Zn:P uptake ratios that are up to an order of magnitude greater than the average for

- oceanic phytoplankton (Twining and Baines 2013; Vance et al. 2017; Sieber et al. 2020). The
- increased presence of diatoms (as indicated by higher fucoxanthin concentrations) at the late
- stations within Terra Nova Bay therefore likely exacerbated the surface decoupling of dZn and

768 ⁶⁷Zn:⁶⁸Zn and ¹¹⁰Cd:¹¹⁴Cd sample ratios relative to controls, and surface depletion of total

- areas with the potential for Zn scarcity during highly productive bloom events, but also reveal
- dynamic changes in the cycling of Cd and Co as a consequence of high Zn demand.
-
- *Data availability*
- CICLOPS (NBP18-01) CTD hydrography data (including pressure, temperature, total
- dissolved oxygen, conductivity, fluorescence, and beam transmission; https://www.bco-
- dmo.org/dataset-deployment/783917) in addition to total dissolved metal, Zn and Cd uptake rate,
- macronutrient, and pigment datasets are available through the NSF Biological and Chemical
- Oceanography Data Management Office (BCO-DMO) repository (https://www.bco-
- 801 dmo.org/deployment/778919). Underway pCO₂ data collected during cruise NBP1801 is
- available through R2R, https://doi.org/10.7284/139318.
- *Author contributions*
- Conceptualization and analysis of the study was carried out by RMK and MAS. This work
- was supervised by MAS and GRD. Funding was acquired by MAS and GRD. All co-authors
- contributed to data collection. RMK and MAS wrote the manuscript with review and editing
- contributions from all co-authors.
- *Competing interests*
- The authors declare that they have no conflict of interest.
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- Baars, O., W. Abouchami, S. J. G. Galer, M. Boye, and P. L. Croot. 2014. Dissolved cadmium in
- the Southern Ocean: Distribution, speciation, and relation to phosphate. Limnology and
- Oceanography **59**: 385–399. doi:10.4319/LO.2014.59.2.0385
- Bertrand, E. M., M. A. Saito, J. M. Rose, C. R. Riesselman, M. C. Lohan, A. E. Noble, P. A.
- Lee, and G. R. DiTullio. 2007. Vitamin B 12 and iron colimitation of phytoplankton
- growth in the Ross Sea. Limnology and Oceanography **52**: 1079–1093.
- doi:10.4319/lo.2007.52.3.1079
- Biller, D. V., and K. W. Bruland. 2012. Analysis of Mn, Fe, Co, Ni, Cu, Zn, Cd, and Pb in
- seawater using the Nobias-chelate PA1 resin and magnetic sector inductively coupled
- 848 plasma mass spectrometry (ICP-MS). Marine Chemistry.
- doi:10.1016/j.marchem.2011.12.001
- Bishop, J. K. B., and T. J. Wood. 2009. Year-round observations of carbon biomass and flux
- variability in the Southern Ocean. Global Biogeochemical Cycles.
- doi:10.1029/2008GB003206
- Bown, J., P. Laan, S. Ossebaar, K. Bakker, P. Rozema, and H. J. W. de Baar. 2017. Bioactive
- trace metal time series during Austral summer in Ryder Bay, Western Antarctic
- Peninsula. Deep-Sea Research Part II: Topical Studies in Oceanography.
- doi:10.1016/j.dsr2.2016.07.004
- Boyle, E. A. 1988. Cadmium: Chemical tracer of deepwater paleoceanography.
- Paleoceanography. doi:10.1029/PA003i004p00471
- Boyle, E. A., F. Sclater, and J. M. Edmond. 1976. On the marine geochemistry of cadmium.
- Nature. doi:10.1038/263042a0

- Cullen, J. T., and R. M. Sherrell. 2005. Effects of dissolved carbon dioxide, zinc, and manganese
- on the cadmium to phosphorus ratio in natural phytoplankton assemblages. Limnology
- and Oceanography **50**: 1193–1204. doi:10.4319/lo.2005.50.4.1193
- Cutter, G. A., and K. W. Bruland. 2012. Rapid and noncontaminating sampling system for trace
- elements in global ocean surveys. Limnology and Oceanography: Methods.
- doi:10.4319/lom.2012.10.425
- 888 Das, P., S. Samantaray, and G. R. Rout. 1997. Studies on cadmium toxicity in plants: A review.
- Environmental Pollution. doi:10.1016/S0269-7491(97)00110-3
- DiTullio, G. R., N. Garcia, S. F. Riseman, and P. N. Sedwick. 2007. Effects of iron concentration
- on pigment composition in Phaeocystis antarctica grown at low irradiance.

Biogeochemistry **83**: 71–81. doi:10.1007/s10533-007-9080-8

- DiTullio, G. R., M. E. Geesey, A. Leventer, and M. P. Lizotte. 2003. Algal pigment ratios in the
- Ross Sea: Implications for Chemtax analysis of Southern Ocean data, p. 35–51. *In*.
- DiTullio, G. R., and W. O. Smith. 1995. Relationship between dimethylsulfide and
- phytoplankton pigment concentrations in the Ross Sea, Antarctica. Deep-Sea Research
- **Part I. doi:10.1016/0967-0637(95)00051-7**
- Ellwood, M. J. 2008. Wintertime trace metal (Zn, Cu, Ni, Cd, Pb and Co) and nutrient
- distributions in the Subantarctic Zone between 40-52°S; 155-160°E. Marine Chemistry.
- doi:10.1016/j.marchem.2008.07.008
- Ellwood, M. J., and K. A. Hunter. 2000. The incorporation of zinc and iron into the frustule of
- the marine diatom Thalassiosira pseudonana. Limnology and Oceanography.
- doi:10.4319/lo.2000.45.7.1517

Global Biogeochemical Cycles **35**: e2020GB006814. doi:10.1029/2020GB006814

- Hutchins, D. A., and K. W. Bruland. 1995. Fe, Zn, Mn and N transfer between size classes in a
- coastal phytoplankton community: Trace metal and major nutrient recycling compared.
- issn: 0022-2402 **53**: 297–313. doi:10.1357/0022240953213197
- Hutchins, D. A., W. X. Wang, M. A. Schmidt, and N. S. Fisher. 1999. Dual-labeling techniques
- for trace metal biogeochemical investigations in aquatic plankton communities. Aquatic
- Microbial Ecology. doi:10.3354/ame019129
- Hutchins, D., and K. Bruland. 1994. Grazer-mediated regeneration and assimilation of Fe, Zn
- and Mn from planktonic prey. Mar. Ecol. Prog. Ser. **110**: 259–269.
- doi:10.3354/meps110259
- Irving, B. H., and R. J. P. Williams. 1953. The Stability of Transition-metal Complexes. Journal 937 of the Chemical Society (Resumed). doi:10.1039/JR9530003192
- Jackson, S. L., J. Spence, D. J. Janssen, A. R. S. Ross, and J. T. Cullen. 2018. Determination of
-
- Mn, Fe, Ni, Cu, Zn, Cd and Pb in seawater using offline extraction and triple quadrupole
- ICP-MS/MS. Journal of Analytical Atomic Spectrometry **33**: 304–313.
- doi:10.1039/c7ja00237h
- Kell, R. M., A. V. Subhas, N. L. Schanke, and others. 2023. Zinc stimulation of phytoplankton in
- a low carbon dioxide, coastal Antarctic environment.doi:10.1101/2023.11.05.565706
- Kellogg, R. M., M. A. Moosburner, N. R. Cohen, and others. 2022. Adaptive responses of
- marine diatoms to zinc scarcity and ecological implications. Nature Communications
- 2022 13:1 **13**: 1–13. doi:10.1038/s41467-022-29603-y
- Lane, E. S., D. M. Semeniuk, R. F. Strzepek, J. T. Cullen, and M. T. Maldonado. 2009. Effects
- of iron limitation on intracellular cadmium of cultured phytoplankton: Implications for

- surface dissolved cadmium to phosphate ratios. Marine Chemistry **115**: 155–162.
- doi:10.1016/J.MARCHEM.2009.07.008
- Lane, T. W., M. A. Saito, G. N. George, I. J. Pickering, R. C. Prince, and F. M. M. Morel. 2005.
- A cadmium enzyme from a marine diatom. Nature **435**: 42–42. doi:10.1038/435042a
- Latour, P., K. Wuttig, P. van der Merwe, and others. 2021. Manganese biogeochemistry in the
- Southern Ocean, from Tasmania to Antarctica. Limnology and Oceanography **66**: 2547–
- 2562. doi:10.1002/lno.11772
- Lee, J. G., S. B. Roberts, and F. M. M. Morel. 1995. Cadmium: A nutrient for the marine diatom
- Thalassiosira weissflogii. Limnology and Oceanography. doi:10.4319/lo.1995.40.6.1056
- Lee, J., and F. Morel. 1995. Replacement of zinc by cadmium in marine phytoplankton. Marine

Ecology Progress Series **127**: 305–309. doi:10.3354/meps127305

- Lohan, M. C., P. J. Statham, and D. W. Crawford. 2002. Total dissolved zinc in the upper water
- column of the subarctic North East Pacific. Deep Sea Research Part II: Topical Studies in Oceanography **49**: 5793–5808. doi:10.1016/S0967-0645(02)00215-1
- Martin, J. H. 1990. Glacial‐interglacial CO2 change: The Iron Hypothesis. Paleoceanography.
- doi:10.1029/PA005i001p00001
- Middag, R., H. J. W. Baar, and K. W. Bruland. 2019. The relationships netween dissolved zinc
- and major nutrients phosphate and silicate along the GEOTRACES GA02 transect in the
- West Atlantic Ocean. Global Biogeochemical Cycles **33**: 63–84.
- doi:10.1029/2018GB006034
- Morel, F. M. M., P. J. Lam, and M. A. Saito. 2020. Trace metal substitution in marine
- phytoplankton. Annual Review of Earth and Planetary Sciences **48**: 491–517.
- doi:10.1146/annurev-earth-053018-060108

- Park, H., B. Song, and F. M. M. Morel. 2007. Diversity of the cadmium-containing carbonic
- anhydrase in marine diatoms and natural waters. Environmental Microbiology.
- 995 doi:10.1111/j.1462-2920.2006.01151.x
- Person, R., M. Vancoppenolle, O. Aumont, and M. Malsang. 2021. Continental and Sea Ice Iron
- Sources Fertilize the Southern Ocean in Synergy. Geophysical Research Letters **48**:
- e2021GL094761. doi:10.1029/2021GL094761
- Price, N. M., and F. M. M. Morel. 1990. Cadmium and cobalt substitution for zinc in a marine diatom. Nature **344**: 658–660. doi:10.1038/344658a0
- Rapp, I., C. Schlosser, D. Rusiecka, M. Gledhill, and E. P. Achterberg. 2017. Automated
- preconcentration of Fe, Zn, Cu, Ni, Cd, Pb, Co, and Mn in seawater with analysis using
- high-resolution sector field inductively-coupled plasma mass spectrometry. Analytica
- Chimica Acta. doi:10.1016/j.aca.2017.05.008
- Roshan, S., T. DeVries, J. Wu, and G. Chen. 2018. The Internal Cycling of Zinc in the Ocean.
- Global Biogeochemical Cycles **32**: 1833–1849. doi:10.1029/2018GB006045
- Rudge, J. F., B. C. Reynolds, and B. Bourdon. 2009. The double spike toolbox. Chemical
- Geology. doi:10.1016/j.chemgeo.2009.05.010
- Saito, M. A., and T. J. Goepfert. 2008. Zinc-cobalt colimitation of *Phaeocystis antarctica*.
- Limnology and Oceanography **53**: 266–275. doi:10.4319/lo.2008.53.1.0266
- Sedwick, P. N., G. R. Di Tullio, and D. J. Mackey. 2000. Iron and manganese in the Ross Sea,
- Seasonal iron limitation in Antarctic. Journal of Geophysical Research: Oceans.
- doi:10.1029/2000JC000256
- Sedwick, P. N., C. M. Marsay, B. M. Sohst, and others. 2011. Early season depletion of
- dissolved iron in the Ross Sea polynya: Implications for iron dynamics on the Antarctic

- Sunda, W. G., and S. A. Huntsman. 1995. Cobalt and zinc interreplacement in marine
- phytoplankton: Biological and geochemical implications. Limnology and Oceanography
- **40**: 1404–1417. doi:10.4319/lo.1995.40.8.1404
- Sunda, W. G., and S. A. Huntsman. 1996. Antagonisms between cadmium and zinc toxicity and
- manganese limitation in a coastal diatom. Limnology and Oceanography **41**: 373–387.
- doi:10.4319/lo.1996.41.3.0373
- Sunda, W. G., and S. A. Huntsman. 1998a. Processes regulating cellular metal accumulation and
- physiological effects: Phytoplankton as model systems. Science of The Total
- Environment **219**: 165–181. doi:10.1016/S0048-9697(98)00226-5
- Sunda, W. G., and S. A. Huntsman. 1998b. Control of Cd Concentrations in a Coastal Diatom by
- Interactions among Free Ionic Cd, Zn, and Mn in Seawater. Environmental Science &
- Technology **32**: 2961–2968. doi:10.1021/es980271y
- Sunda, W. G., and S. A. Huntsman. 2000. Effect of Zn, Mn, and Fe on Cd accumulation in
- phytoplankton: Implications for oceanic Cd cycling. Limnology and Oceanography **45**:
- 1501–1516. doi:10.4319/lo.2000.45.7.1501
- Tan, D., W. Xu, Z. Zhu, S. Li, G. Wu, and H. Qin. 2020. Optimizing the ratio of the spike to
- sample for isotope dilution analysis: a case study with selenium isotopes. Acta
- Geochimica. doi:10.1007/s11631-019-00390-6
- Twining, B. S., and S. B. Baines. 2013. The trace metal composition of marine phytoplankton.
- Annual review of marine science **5**: 191–215. doi:10.1146/annurev-marine-121211-
- 172322

- Twining, B. S., S. B. Baines, and N. S. Fisher. 2004. Element stoichiometries of individual
- plankton cells collected during the Southern Ocean Iron Experiment (SOFeX).
- Limnology and Oceanography **49**: 2115–2128. doi:10.4319/LO.2004.49.6.2115
- Twining, B. S., S. D. Nodder, A. L. King, and others. 2014. Differential remineralization of
- major and trace elements in sinking diatoms. Limnology and Oceanography **59**: 689–704.
- doi:10.4319/LO.2014.59.3.0689
- Vance, D., S. H. Little, G. F. De Souza, S. Khatiwala, M. C. Lohan, and R. Middag. 2017.
- Silicon and zinc biogeochemical cycles coupled through the Southern Ocean. Nature Geoscience **10**: 202–206. doi:10.1038/NGEO2890
- Weber, T., S. John, A. Tagliabue, and T. DeVries. 2018. Biological uptake and reversible

scavenging of zinc in the global ocean. Science **361**: 72–76.

- doi:10.1126/SCIENCE.AAP8532
- Wright, S. W., R. L. van den Enden, I. Pearce, A. T. Davidson, F. J. Scott, and K. J. Westwood.
- 2010. Phytoplankton community structure and stocks in the Southern Ocean (30-80°E)
- determined by CHEMTAX analysis of HPLC pigment signatures. Deep-Sea Research

Part II: Topical Studies in Oceanography. doi:10.1016/j.dsr2.2009.06.015

- Wu, J., and E. A. Boyle. 1998. Determination of iron in seawater by high-resolution isotope
- dilution inductively coupled plasma mass spectrometry after Mg(OH)2 coprecipitation.
- Analytica Chimica Acta **367**: 183–191. doi:10.1016/S0003-2670(98)00145-7
- Wuttig, K., A. T. Townsend, P. van der Merwe, and others. 2019. Critical evaluation of a
- seaFAST system for the analysis of trace metals in marine samples. Talanta.
- doi:10.1016/j.talanta.2019.01.047

- Xu, Y., D. Tang, Y. Shaked, and F. M. M. Morel. 2007. Zinc, cadmium, and cobalt
- interreplacement and relative use efficiencies in the coccolithophore Emiliania huxleyi.
- Limnology and Oceanography **52**: 2294–2305. doi:10.4319/lo.2007.52.5.2294
- Zhao, Y., D. Vance, W. Abouchami, and H. J. W. de Baar. 2014. Biogeochemical cycling of zinc
- and its isotopes in the Southern Ocean. Geochimica et Cosmochimica Acta.
- doi:10.1016/j.gca.2013.07.045