

1           **High metabolic zinc demand within native Amundsen and Ross Sea phytoplankton**  
2                           **communities determined by stable isotope uptake rate measurements**

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13 **Abstract.** Zinc (Zn) is an essential micronutrient for most eukaryotic phytoplankton. Zn uptake  
14 by phytoplankton within the euphotic zone results in nutrient-like dissolved Zn (dZn) profiles  
15 with a large dynamic range. The combination of key biochemical uses for Zn and large vertical  
16 gradients in dZn implies the potential for rapid rates of Zn removal from the surface ocean.

17 However, due to the ease of contamination at sea, direct measurements of dZn uptake within  
18 natural environments have not been previously made. To investigate the demand for dZn and for  
19 dissolved cadmium (dCd; a closely related nutrient-like element) within Southern Ocean  
20 phytoplankton communities, we conducted <sup>67</sup>Zn and <sup>110</sup>Cd tracer uptake experiments within the  
21 Amundsen Sea, Ross Sea, and Terra Nova Bay of the Southern Ocean. We observed a high  
22 magnitude of Zn uptake ( $\rho\text{Zn} > 100 \text{ pmol dZn L}^{-1} \text{ d}^{-1}$ ) into the particulate phase that was  
23 consistent with ambient depleted dZn surface concentrations. High biomass and low seawater

24  $p\text{CO}_2$  appeared to contribute to  $\rho\text{Zn}$ , which also led to increases in  $\rho\text{Cd}$  likely through the  
25 upregulation of shared transport systems. These high  $\rho\text{Zn}$  measurements further imply that only  
26 short timescales are needed to deplete the large winter  $d\text{Zn}$  inventory down to the observed  
27 surface levels in this important carbon-capturing region. Overall, the high magnitude of Zn  
28 uptake into the particulate fraction suggests that even in the Zn-rich waters of the Southern  
29 Ocean, high Zn uptake rates can lead to Zn depletion and potential Zn scarcity.

## 30 **1 Introduction**

31 Zinc (Zn) is an essential trace metal micronutrient for marine phytoplankton with roles in  
32 carbon fixation, organic phosphorus uptake, and transcriptional and translational processes,  
33 among others (Morel et al., 2013, 2020; Shaked et al., 2006; Twining and Baines, 2013).  
34 Nutrient-like depth profiles of total dissolved Zn ( $d\text{Zn}$ ) are characterized by depleted surface  
35 concentrations due to uptake by phytoplankton within the euphotic zone, reflecting this high  
36 biological demand (Fitzwater et al., 2000; Lohan et al., 2002; Middag et al., 2019; Zhao et al.,  
37 2014). Zn is particularly important as a catalytic cofactor in carbonic anhydrase (CA)  
38 metalloenzymes, which catalyze the reversible dehydration of  $\text{HCO}_3^-$  to  $\text{CO}_2$ . As  $\text{HCO}_3^-$   
39 constitutes about 90% of the dissolved inorganic carbon (DIC) pool in the surface ocean, CAs in  
40 marine algae are a critical part of the carbon concentrating mechanism (CCM) that maintains a  
41  $\text{CO}_2$  supply to the carbon-fixing enzyme ribulose-1,5-biphosphate carboxylase/oxygenase  
42 (RUBISCO). Less abundant divalent metal cations such as cobalt ( $\text{Co}^{2+}$ ) and cadmium ( $\text{Cd}^{2+}$ )  
43 can replace  $\text{Zn}^{2+}$  in some algal CA subtypes (Lane et al., 2005), conferring biochemical  
44 flexibility to algae confronted with low Zn bioavailability.

45 While Cd is known to cause toxic effects in most organisms (Brand et al., 1986; Das et  
46 al., 1997),  $d\text{Cd}$  depth profiles are also nutrient-like. As noted above, the biological use of Cd as a

47 catalytic cofactor within Cd-containing carbonic anhydrase ( $\zeta$ -CA, or CDCA) likely contributes  
48 to surface dCd depletion and thus to the observed nutrient-like profiles, though this remains the  
49 only known biological use of Cd to date (Haas et al., 2009; Lee and Morel, 1995; Sunda and  
50 Huntsman, 2000). It has also been proposed that phytoplankton may assimilate Cd abiotically—  
51 this mode of Cd uptake is non-specific, a case of ‘mistaken identity’ in which phytoplankton  
52 bind and store imported Cd inside the cell to avoid toxicity, coupling the cycling of Cd to the  
53 biological cycle of nutrients (Horner et al., 2013). As the beneficial effect of adding Cd to  
54 phytoplankton cultures has only been observed when Zn is limiting (Lee et al., 1995; Price and  
55 Morel, 1990; Xu et al., 2007), it has been speculated that the ability to use Cd in place of Zn in  
56 CDCA may confer a competitive advantage to Zn-limited algae under low pCO<sub>2</sub>. To date,  
57 homolog *cdca* genes have been found exclusively in diatom species (Park et al., 2007, 2008).  
58 However, since the beneficial effect of Cd has also been observed in organisms such as the green  
59 alga *Tetraselmis maculata* and the coccolithophore *Emiliania huxleyi* that lack the *cdca* gene  
60 (Lee and Morel, 1995), it is thought that Cd may have other biochemical functions in  
61 phytoplankton still awaiting discovery.

62         Due to the generally high (> 1nM) dZn concentrations observed in Southern Ocean  
63 waters (Baars and Croot, 2011), Zn has not been considered to be a limiting micronutrient in this  
64 region, despite the fact that Southern Ocean diatom species possess cellular Zn quotas that are 3-  
65 15x higher compared to those of low-latitude species (Twining and Baines, 2013), with model-  
66 inferred Zn uptake rates used to find that Southern Ocean phytoplankton account for 62% of  
67 global Zn uptake (Roshan et al., 2018). Substantial removal of dZn from Southern Ocean surface  
68 waters appears associated with high biomass blooms and low pCO<sub>2</sub> conditions during austral  
69 spring and summer, creating the potential for phytoplankton growth to become Zn and carbon

70 co-limited (Kell et al., 2023; Morel et al., 1994). While phytoplankton growth in the Southern  
71 Ocean is well-known to be primarily limited by Fe availability (Arrigo et al., 2008; Martin,  
72 1990), melting icebergs and ice shelves are known to act as external sources of Fe (Hopwood et  
73 al., 2019; Person et al., 2021; Planquette et al., 2013; St-Laurent et al., 2017) with larger Fe  
74 inputs expected from increased ice melt in a warming climate. The majority of ice-melted Fe  
75 input is sourced from particulate lithogenic material (entrained during grounding of ice shelves  
76 on the continent and sediments). Fe in crustal material is more abundant than Zn (3.5% as Fe  
77 versus 0.0071% as Zn) (Taylor and McLennan, 1985), creating a large inventory of particulate  
78 Fe available that can be partially dissolved by biotic and abiotic processes. Increased dFe inputs  
79 to surface Antarctic waters may act to relieve Fe stress, but would simultaneously support the  
80 development of other nutrient limitations. For example, low availabilities of both dZn and  
81 vitamin B<sub>12</sub> have been previously observed to co-limit phytoplankton growth with Fe in the Ross  
82 Sea (Bertrand et al., 2007; Kell et al., 2023). A high demand for Zn naturally exists within  
83 eukaryotic phytoplankton due to the requirement for Zn<sup>2+</sup> in numerous metabolic functions;  
84 therefore, without similarly enhanced inputs of dZn to the water column, the alleviation of  
85 primary Fe limitation could induce Zn stress as the next most in-demand metal micronutrient.  
86 Coastal polynyas that form within the Amundsen and Ross Seas during austral spring and  
87 summer are particularly primed to experience Zn stress as these regions host highly productive  
88 seasonal phytoplankton blooms that act as significant carbon sinks (Arrigo et al., 2012). This  
89 high productivity draws pCO<sub>2</sub> down to low levels (< 200 ppm), putting pressure on the carbon  
90 concentrating mechanism of photosynthetic phytoplankton to acquire CO<sub>2</sub> and thus to acquire Zn  
91 as the predominantly utilized metal cofactor within carbonic anhydrases.

92           The present study enhances our knowledge of the rate of dZn removal from the surface  
93 Southern Ocean and uptake into the particulate fraction with empirical field data measured  
94 within native Southern Ocean phytoplankton communities. This study developed a field-based,  
95 stable Zn isotope uptake rate method, building on a prior stable Cd uptake rate method (Cox et  
96 al., 2014). While Zn uptake has been measured in laboratory cultures (Sunda and Huntsman,  
97 1992, 1995, 2000), and the influence of grazing and trophic transfer studies have been conducted  
98 using radioactive isotopes (Hutchins and Bruland, 1995, 1994), to our knowledge direct  
99 measurements of Zn uptake in natural marine phytoplankton communities have not been  
100 conducted previously, despite interest in modeling the biogeochemical uptake and cycling of Zn  
101 (Weber et al., 2018). We measured the total uptake rates of Zn and Cd along the shelves of the  
102 Amundsen Sea and Ross Sea during the austral summer of 2017-2018 (December – March). This  
103 was accomplished by introducing  $^{67}\text{Zn}$  and  $^{110}\text{Cd}$  (with natural abundances of 4.10% and 12.5%,  
104 respectively) into short-term (24 hr) incubation experiments. The aim was to quantify the transfer  
105 of dissolved  $^{110}\text{Cd}^{2+}$  and  $^{67}\text{Zn}^{2+}$  into the particulate fraction exceeding 3  $\mu\text{m}$ . Both stable isotopes  
106 can be used as uptake tracers by analysis of isotope abundances that deviate from natural  
107 abundances within the particulate phase. The transfer of added isotopes into the particulate phase  
108 is the combined result of 1) active transport of metal into cells, 2) nonspecific metal adsorption  
109 to cell surfaces, 3) metal adsorption to non-living particulate organic matter, and 4) metal  
110 adsorption to particulate inorganic matter, though we expect active transport into cells to  
111 dominate the measured particulate isotopic signal due to the high abundance of actively growing  
112 autotrophic cells in the photic zone observed in the Southern Ocean during austral summer.  
113 These measurements of uptake rates were then used to infer timescales of surface dZn depletion  
114 in these Antarctic environments. These uptake rates contribute to understanding the biological

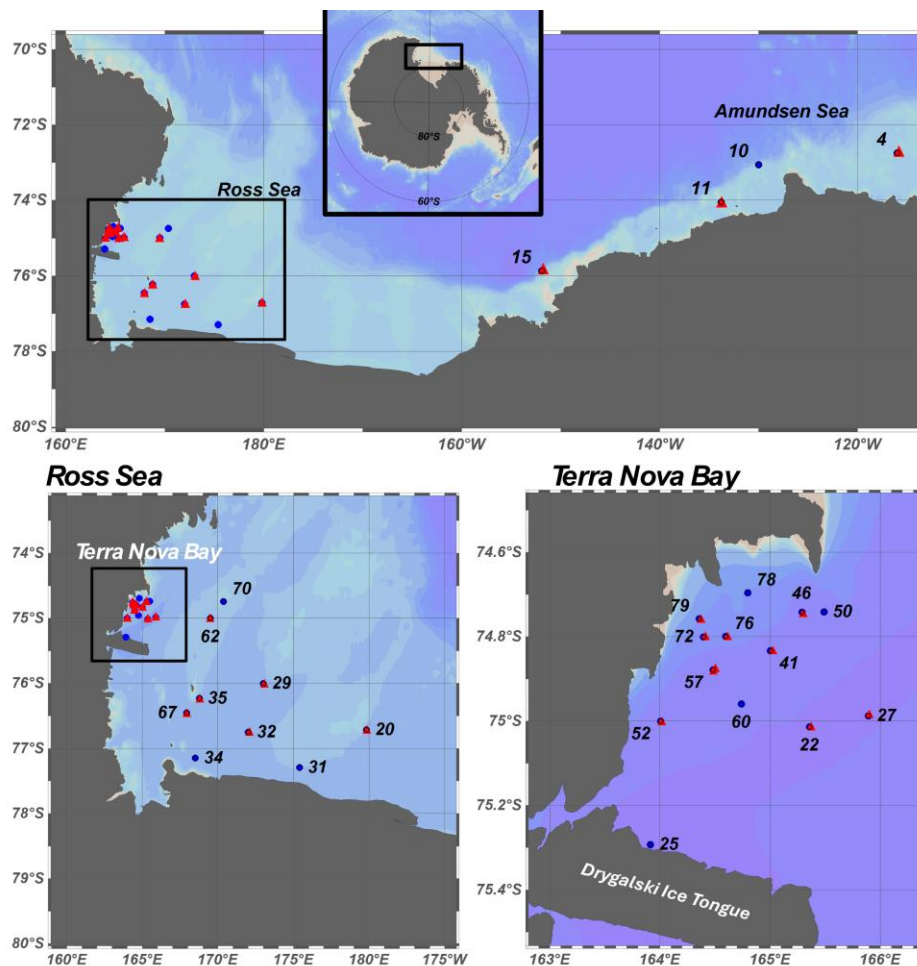
115 demand and potential for Zn limitation of primary productivity in highly productive coastal  
116 environments, such as the polynyas surrounding Antarctica (Kell et al., 2023).

117

## 118 2 Methods

### 119 2.1 Study area and sample collection

120 Samples were collected during the CICLOPS (Cobalamin and Iron Co-Limitation of  
121 Phytoplankton Species) expedition (NBP18-01) aboard the RVIB *Nathaniel B. Palmer*,  
122 December 11, 2017 – March 3, 2018 in the Amundsen Sea and Ross Sea of the Southern Ocean  
123 (Fig. 1).



124

125 **Figure 1.** Map showing the stations sampled over the course of the CICLOPS cruise. Stations  
126 marked by red triangles indicate those at which stable  $^{67}\text{Zn}$  and  $^{110}\text{Cd}$  uptake rate experiments  
127 were performed. An expanded map of stations sampled in the Ross Sea is shown at bottom left,  
128 while a further expansion of stations sampled in Terra Nova Bay is shown at bottom right.  
129

130 Station metadata is given in Table S1. Water samples were collected using trace metal clean  
131 (TMC) sampling protocols described previously (Cutter and Bruland, 2012). A TMC rosette  
132 suspended on a Kevlar line and equipped with twelve 8L X-Niskin bottles (Ocean Test  
133 Equipment) was used to collect seawater at depths ranging from 10 – 600 m. Continuous  
134 underway measurements of  $\text{pCO}_2$  at ~5 m depth were taken using a  $\text{pCO}_2$  measurement system  
135 from Lamont-Doherty Earth Observatory (LDEO, 0.017/sec rate). Hydrography data were  
136 collected using sensors deployed on a titanium trace metal rosette (TMR) in tandem with TMC  
137 niskin bottles. The TMR was equipped with sensors to measure temperature, conductivity,  
138 pressure, dissolved oxygen, chlorophyll (Chl) fluorescence, altimetry, beam transmission, and  
139 photosynthetically active irradiance (PAR). Chl fluorescence was measured using a WetLabs  
140 ECO-FL fluorometer. A complete data report and sensor list are available at [NBPI801DATA.pdf](https://rvdata.us)  
141 ([rvdata.us](https://rvdata.us)). Mixed layer depth (MLD) was determined for each station within Terra Nova Bay as  
142 the first depth at which the difference between the potential density ( $\sigma_\theta$ ) and reference density  
143 (the potential density at 10m,  $\sigma_{\text{ref}}$ ) was greater than or equal to  $0.125 \text{ kg m}^{-3}$  (Bishop and Wood,  
144 2009; Ohnemus et al., 2017).

## 145 **2.2 Preparation of plasticware**

146 Polyethylene and polycarbonate sampling and incubation bottles were rigorously cleaned  
147 to remove trace metal contaminants before use. Bottles with rinsed with Milli-Q water  
148 (Millipore), soaked for 72h in <1% Citranox detergent, rotated, soaked for an additional 72h, and  
149 then rinsed five times with Milli-Q water. Bottles were then filled with 10% HCl (Baker instra-

150 analyzed) by volume and soaked for a minimum of one week, rotated, and soaked for another  
151 week. Bottles were then rinsed five times with dilute acid (HCl, pH 2) and stored double-bagged  
152 in plastic zip bags. All cleaning work was conducted in a Class 100 clean room. Polypropylene  
153 15 mL centrifuge tubes used in sample processing were cleaned of potential metal contamination  
154 by soaking in 10% HCl for 5 days and rinsing with pH 2 HCl prior to use.

### 155 **2.3 Sampling for total dissolved metal analyses**

156 Samples for the analysis of total dissolved Zn, Cd, Fe, Mn, Cu and Ni concentrations  
157 were collected shipboard by pressure-filtering X-Niskin bottles through an acid-washed 142mm,  
158 0.2µM Supor membrane filter (Pall) within 3 hours of rosette recovery using high purity  
159 (99.999%) N<sub>2</sub> gas. Total dissolved water samples were collected into 250 mL TMC polyethylene  
160 bottles and stored double-bagged in plastic zip bags. Seawater samples for <sup>110</sup>Cd and <sup>67</sup>Zn stable  
161 isotope uptake experiments were collected in the same way but without filtering. All sample  
162 collection occurred shipboard within a TMC van containing laminar flow hoods and plastic  
163 sheeting. Samples for total dissolved metal analysis were acidified to pH 1.7 with high purity  
164 HCl (Optima, Fisher Scientific) within 7 months of sampling and were stored acidified at room  
165 temperature for over 1 year prior to analysis.

### 166 **2.4 Analyses of total dissolved metals using isotope dilution**

167 Quantification of dissolved metals in samples and reference seawater was performed for  
168 total dissolved Fe, Ni, Cu, Zn, and Cd using isotope dilution. 15 mL of acidified seawater sample  
169 was spiked with 50 µL of a stable isotope spike solution artificially enriched in <sup>57</sup>Fe, <sup>61</sup>Ni, <sup>65</sup>Cu,  
170 <sup>67</sup>Zn, and <sup>110</sup>Cd ('spike isotopes'). Reference isotopes used in this study were <sup>56</sup>Fe, <sup>60</sup>Ni, <sup>63</sup>Cu,  
171 <sup>66</sup>Zn, and <sup>114</sup>Cd. All spike isotopes were received in solid form (Oak Ridge National  
172 Laboratory). Initial dissolution and all subsequent dilutions were made using concentrated nitric



173 acid (Optima, Fisher Scientific). Concentrations and ratios of isotopes for each metal in the spike  
 174 solution were verified by inductively coupled plasma mass spectrometry (ICP-MS) using a  
 175 multi-element standard curve (SPEX CertiPrep). The composition of the isotope spike addition  
 176 was made such that the target isotope ratios in the total, 15mL spiked sample would be  $^{57}\text{Fe}/^{56}\text{Fe}$   
 177  $= 0.7$ ,  $^{61}\text{Ni}/^{60}\text{Ni} = 0.5$ ,  $^{65}\text{Cu}/^{63}\text{Cu} = 1$ ,  $^{67}\text{Zn}/^{66}\text{Zn} = 0.7$ , and  $^{110}\text{Cd}/^{114}\text{Cd} = 1$  and were verified with  
 178 ICP-MS. These ratios were chosen to minimize the uncertainty introduced by error propagation  
 179 through the isotope dilution equation (Kato et al., 1990; Rudge et al., 2009; Tan et al., 2020; Wu  
 180 and Boyle, 1998). The same spike solution was used to spike all samples from all depths.  
 181 Because it is monoisotopic, total dissolved Mn was calculated using a modified isotope dilution  
 182 equation:

$$183 \quad \text{Mn (nM)} = \frac{{}^{55}\text{Mn}_{\text{spl}} (\text{cps})}{{}^{57}\text{Fe}_{\text{spl}} (\text{cps})} * {}^{57}\text{Fe}_{\text{spike}} (\text{nM}) * {}^{57}\text{Fe}_{\text{slope}} (\text{cps/ppb}) * \frac{1}{({}^{55}\text{Mn}_{\text{slope}}) (\text{cps/ppb})} \quad (1)$$

184 in which  ${}^{55}\text{Mn}_{\text{spl}}$  and  ${}^{57}\text{Fe}_{\text{spl}}$  refer to the blank corrected counts per second (cps) of  ${}^{55}\text{Mn}$  and  ${}^{57}\text{Fe}$   
 185 in the spiked sample,  ${}^{57}\text{Fe}_{\text{spike}}$  is the concentration of  ${}^{57}\text{Fe}$  in the spike,  ${}^{57}\text{Fe}_{\text{slope}}$  is the slope of  
 186 the external standard calibration curve (SPEX curve) relating  ${}^{57}\text{Fe}$  cps to ppb, and  ${}^{55}\text{Mn}_{\text{slope}}$  is the  
 187 slope of the external calibration curve (SPEX curve) relating  ${}^{55}\text{Mn}$  cps to ppb. Due to the  
 188 acidification of seawater prior to ICP-MS analysis, Mn ICP-MS measurements do not include  
 189 contributions from humic-type Mn(III)-ligand complexes (Oldham et al., 2021). Until the  
 190 inclusion of Mn(III) is resolved and intercalibrated, we report these Mn values as Mn(II) and  
 191 note that they are consistent with prior studies employing the same acidification technique  
 192 (Gerringa et al., 2020; Noble et al., 2013; Sedwick et al., 2000).

193 Preconcentration of spiked seawater samples for total dissolved metal analysis was  
 194 performed using the automated solid phase extraction system seaFAST-pico (Elemental  
 195 Scientific) in offline concentration mode with an initial volume of 15 mL and elution volume of

196 500  $\mu$ L (Bown et al., 2017; Jackson et al., 2018; Rapp et al., 2017; Wuttig et al., 2019). The  
197 seaFAST contains a Nobias-chelate PA1 resin column (ethylenediaminetriacetate and  
198 iminodiacetate) suitable for the simultaneous preconcentration of several trace metals (Fe, Mn,  
199 Zn, Cu, Co, Cd, Ni) with high sensitivity and quantitative recovery (Biller and Bruland, 2012;  
200 Sohrin et al., 2008). Adjusted seaFAST software settings were a 17 second load loop time and a  
201 single 10 mL load cycle.

202 Reagents consisted of a 1.5M ammonium acetate pH 6.0 buffer made using glacial acetic  
203 acid and ammonium hydroxide (20-22%) of the highest purity (Optima, Fisher Chemical), a 1%  
204 nitric acid rinse solution (Optima grade, Fisher Chemical), and a 10% nitric acid elution buffer  
205 (Optima grade, Fisher Chemical) with 10 ppb indium ( $^{115}\text{In}$ , SPEX CertiPrep) added as an  
206 internal standard. Solutions were prepared with 18.2  $\Omega$  Milli-Q water (Millipore).

207 Following offline seaFAST preconcentration, multi-elemental quantitative analysis was  
208 performed using an iCAP-Q inductively coupled plasma-mass spectrometer (Thermo Scientific).  
209 To minimize oxide interference on metal isotopes, a cooled spray chamber and helium collision  
210 gas were employed. Analytes were measured in single quadrupole mode (kinetic energy  
211 discrimination [KED]). Concentrations of Mn, Fe, Ni, Cu, Zn and Cd were determined using a  
212 six-point external standard curve of a multi-element standard (SPEX CertiPrep), diluted to range  
213 from 1-10 ppb in 5% nitric acid. An indium standard (SPEX CertiPrep) was similarly added to  
214 these standard stocks, diluted to range 1-10 ppb. Instrument injection blanks consisted of 5%  
215 nitric acid in Milli-Q. Standard curve  $R^2$  values were  $\geq 0.98$  for all metals monitored. Method  
216 accuracy and precision were assessed using the 2009 GEOTRACES coastal surface seawater  
217 (GSC) standard ( $n = 8$ ; Table S2), which produced values consistent with consensus results.

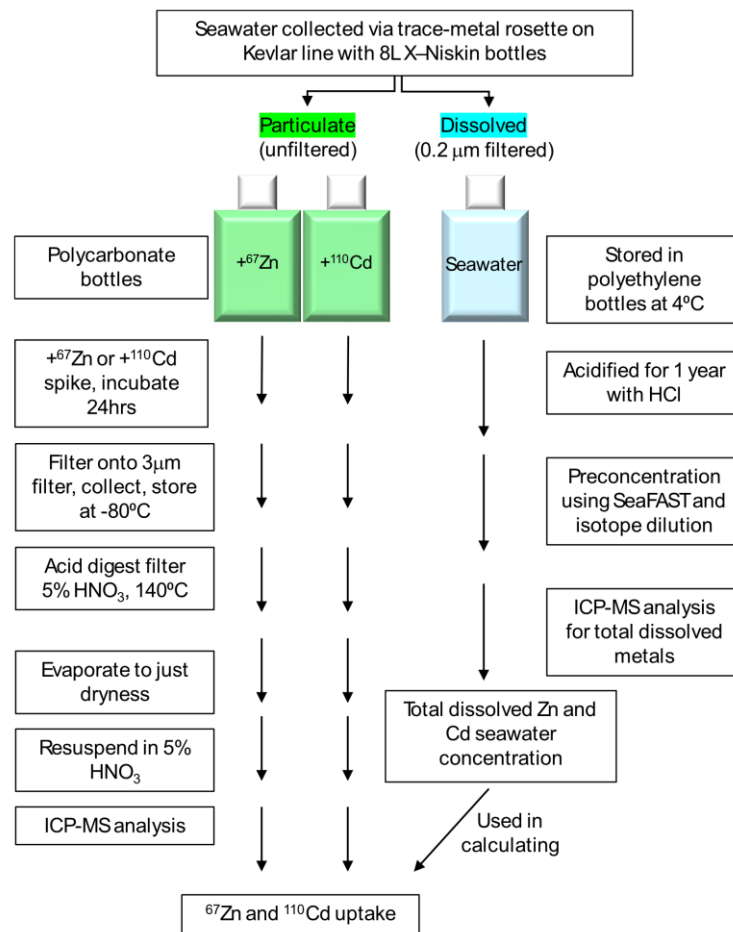
218 **2.5 Procedural blanks and limit of detection (LOD).**

219 Procedural blanks were quantified by preconcentrating 30mL of MilliQ water adjusted to  
 220 pH 2 with HCl (Optima, Fisher Scientific) to 1mL. Metal concentrations were determined using  
 221 an external SPEX multi-element standard as described above. The LOD was calculated as 3 x the  
 222 standard deviation of the blank measurements (13.5 pM Fe, 2.9 pM Ni, 1.6 pM Cu, 38.1 pM Zn,  
 223 and 0.3 pM Cd).

224

225 **2.6 Uptake experiments: <sup>67</sup>Zn and <sup>110</sup>Cd spiking, incubation, and sample collection**

226 <sup>67</sup>Zn and <sup>110</sup>Cd and stable isotope uptake experiments were modeled after those  
 227 conducted by Cox et. al. 2014, with the addition of Zn uptake measurements. An overall  
 228 schematic detailing these experiment workflows is shown in Fig. 2.



229

230 **Figure 2.** Diagram showing the overall workflow used to measure particulate uptake of  $^{110}\text{Cd}$   
231 and  $^{67}\text{Zn}$  and total dissolved Cd and Zn, after Cox et al. 2014.

232

233 Uptake experiments were performed at 18 stations total (Fig. 1). Raw (unfiltered) seawater was  
234 collected shipboard over a depth range of 10 – 600 m into 250 mL TMC polycarbonate  
235 incubation bottles. All incubation bottles were filled with minimal headspace such that the total  
236 culture volume was ~275 mL. Two incubation bottles per depth were filled with raw seawater—  
237 one was spiked with  $^{67}\text{Zn}$ , the other was spiked with  $^{110}\text{Cd}$ . The Cd and Zn isotope spikes were  
238 prepared by dissolving  $^{110}\text{CdO}$  and  $^{67}\text{ZnO}$  (Oak Ridge National Laboratory) in 5%  $\text{HNO}_3$   
239 (Seastar Baseline) and were diluted using Milli-Q water to minimize added acidity. When added  
240 to the filled incubation bottles, the total added (spiked) concentration of Cd was 300 pM and the  
241 total added concentration of Zn was 2 nM. The chosen total added concentrations were based on  
242 the surface ratio of total dissolved Cd (dCd) to total dissolved Zn (dZn) reported previously for  
243 the Ross Sea (Fitzwater et al., 2000). Immediately after spiking, incubation bottles were sealed,  
244 inverted to mix, and transferred to flow-through on-deck incubators for 24hr. Incubators were  
245 shielded by black net neutral density screening to allow 20% ambient light penetration.

246 Biomass was collected after 24hr by vacuum filtering the entire volume of each  
247 incubation sample at 34.5 kPa (5 psi) onto an acid-cleaned  $3\mu\text{m}$ , 50mm acrylic copolymer  
248 (Versapore) filter (Pall) mounted on an acid-cleaned Teflon (Savillex) filtration rig. Samples  
249 were filtered through  $3\mu\text{m}$  pore-size filters rather than  $0.2\mu\text{m}$  in order to minimize filtration  
250 time (and thus time exposed to potential contamination) and to capture the bulk of eukaryotic  
251 phytoplankton biomass typically found in the Southern Ocean. An aliquot of 1 mL of  $0.2\mu\text{m}$   
252 filtered surface seawater (collected at 10 m depth) was used to rinse the sample before collecting  
253 the filter into an acid-cleaned 2 mL cryovial using acid-rinsed plastic forceps. Filter blanks were

254 duplicate 3  $\mu\text{m}$  acid-clean Versapore (Pall) filters that were placed onto the filtration rig, rinsed  
255 with filtered surface seawater, collected, stored, and processed as samples were to correct for any  
256 contaminating metals present on the filters themselves. Blanks were collected at each station.  
257 Filters were stored frozen at  $-80^{\circ}\text{C}$  in acid-cleaned cryovials until analysis. The filtration rig was  
258 rinsed with pH 2 HCl between samples. Polycarbonate incubation bottles were cleaned between  
259 stations with a 10% HCl rinse and several rinses in Milli-Q water, followed by a brief soak in  
260 10% HCl followed by a pH 2 HCl rinse. All spike addition and sample filtration procedures were  
261 completed in a fabricated shipboard positive-pressure clean room environment made of laminar  
262 flow hoods and plastic sheeting.

263 We note that the total Zn and Cd uptake rate values presented in this study represent  
264 potential uptake rates rather than true uptake rates—this naturally arises as a consequence of  
265 adding the spiked tracer  $^{67}\text{Zn}$  and  $^{110}\text{Cd}$  into raw surface seawater. As this seawater is naturally  
266 depleted in both metals, the spike addition artificially increases the total Zn and Cd present and  
267 thus could perturb the response of biology to these additions. It should also be noted that both  
268  $^{67}\text{Zn}$  and  $^{110}\text{Cd}$  spikes were not equilibrated with natural seawater before their addition to  
269 incubation bottles to maintain experimental consistency. Experiments of this nature have been  
270 conducted previously using radioisotopes as tracers (Cullen et al., 1999; Hutchins et al., 1999;  
271 Morel et al., 1994; Sunda and Huntsman, 1995), though we chose to use stable isotopes for ease  
272 of shipboard use and waste disposal.

## 273 **2.7 Filter digestion and particulate ICP-MS analysis**

274 All work was performed in a Class 100 clean room under laminar flow hoods. Sample  
275 filters were retrieved from storage at  $-80^{\circ}\text{C}$ , removed from cryovials using plastic acid-washed  
276 forceps, and transferred into trace metal clean 15 mL PFA vials with 4 mL of 5%  $\text{HNO}_3$

277 (Optima) containing a 1 ppb Indium (In) internal standard. Filters were digested for ~3.5h at  
 278 140°C using a HotBlock® heating block (Environmental Express, USA). Filters were then  
 279 removed and discarded, leaving behind the liquid extract. After evaporating the remaining  
 280 solution to just dryness, the residue was resuspended in 2 mL of 5% HNO<sub>3</sub> (Optima) by light  
 281 vortexing. Process blank filters were digested and processed as sample filters were. Digests were  
 282 analyzed in duplicate by ICP-MS using a Thermo ICAP-Q plasma mass spectrometer calibrated  
 283 to a multi-element standard curve (SPEX Certiprep) over a range of 1 – 20 ppb. Duplicate values  
 284 were in good agreement (Supplementary File 1), and the average value was used in further  
 285 calculations. Natural Cd and Zn isotope abundances of the standards were assumed to calculate  
 286 concentrations of <sup>110</sup>Cd, <sup>111</sup>Cd, <sup>114</sup>Cd, <sup>67</sup>Zn, <sup>66</sup>Zn, and <sup>68</sup>Zn. Digests were analyzed in KED mode  
 287 after an 85s sample uptake window and element mass windows were scanned 3 times during  
 288 measurements. The 1 ppb In internal standard was used to correct for variation in sample  
 289 delivery and plasma suppression between samples. Process blanks were subtracted from  
 290 measured sample concentrations. Phosphorus concentrations were simultaneously measured by  
 291 ICP-MS and were calibrated to a standard curve ranging from 100 – 3,200 ppb using a 1 ppm  
 292 certified P stock (Alfa Aesar Specpure). Equation #2 was used for the calculations described  
 293 above:

$$294 \quad M_{particulate} = \left[ \frac{M_{sample}}{In_{sample}} - \frac{M_{blank}}{In_{blank}} \right] * \frac{In_{digestion}}{M_{slope}} * \frac{V_{digested}}{V_{filtered}} \quad (2)$$

295 where  $V_{filtered}$  is the total spiked sample volume estimated to have passed through the filter (275  
 296 mL),  $V_{digested}$  is the final volume the sample was resuspended in (2.0 mL),  $M_{sample}$  is the metal of  
 297 interest measured in the sample in units of counts per second (cps),  $M_{blank}$  is the metal of interest  
 298 measured in the process blanks (cps),  $M_{slope}$  is the slope of the metal of interest obtained by the  
 299 standard curve (cps ppb<sup>-1</sup>),  $In_{sample}$  is the In measured in the sample (cps),  $In_{blank}$  is the In

300 measured in the process blanks (cps),  $I_{\text{digestion}}$  is the cps of In measured in the 5% HNO<sub>3</sub>+1 ppb  
301 In digestion solution, and the calculated concentration of the metal of interest ( $M_{\text{particulate}}$ ) is in  
302 ppb ( $\mu\text{g L}^{-1}$ ). This equation is the same as that used by Noble et. al. 2013 for the determination  
303 of particulate metal concentrations using ICP-MS (Noble et al., 2013).

304 The Zn spike and Cd spike were also analyzed by ICP-MS using a tenfold dilution of  
305 spike solution into 5% HNO<sub>3</sub> containing 1 ppb In to determine isotopic compositions and  
306 concentrations. When added to filled incubation bottles (275 mL total volume), the added  
307 concentrations were 288 pM <sup>110</sup>Cd, 4.51 pM <sup>111</sup>Cd, and 1.69 pM <sup>114</sup>Cd for Cd spiked bottles, and  
308 were 1.91 nM <sup>67</sup>Zn, 0.045 nM <sup>66</sup>Zn, and 0.047 nM <sup>68</sup>Zn for Zn spiked bottles (Table S3). For all  
309 stations and all depths, <sup>67</sup>Zn and <sup>110</sup>Cd spike concentrations exceeded natural dissolved <sup>67</sup>Zn and  
310 <sup>110</sup>Cd concentrations, estimated by multiplying the total dissolved Zn and Cd by the natural  
311 isotope abundance of <sup>67</sup>Zn and <sup>110</sup>Cd (0.0410 and 0.1249, respectively; see comparisons in Fig.  
312 S2).

### 313 **2.8 Calculating zinc and cadmium uptake using <sup>67</sup>Zn and <sup>110</sup>Cd**

314 Total Zn and Cd uptake was calculated using Eq. (3) and Eq. (4), respectively. <sup>110</sup>Cd<sub>Sample</sub>  
315 and <sup>67</sup>Zn<sub>Sample</sub> are the particulate <sup>110</sup>Cd and <sup>67</sup>Zn measured by ICP-MS analysis of the 3  $\mu\text{m}$   
316 sample filter (using the digestion protocol described in section 2.7) normalized to the total  
317 culture volume (275 mL) and 24 hr of incubation. <sup>110</sup>Cd<sub>Sample</sub> and <sup>67</sup>Zn<sub>Sample</sub> already in the  
318 particulate fraction (that is, the pCd and pZn that existed in the water column upon collection of  
319 the raw seawater samples) were accounted for by subtracting these pre-existing particulate <sup>110</sup>Cd  
320 and <sup>67</sup>Zn values, <sup>110</sup>Cd<sub>PEP</sub> and <sup>67</sup>Zn<sub>PEP</sub>. The pre-existing particulate value for <sup>110</sup>Cd was obtained  
321 from incubation bottles that had Zn added, but no Cd spike. Likewise, the pre-existing particulate  
322 value for <sup>67</sup>Zn was obtained from incubation bottles that had Cd added, but no Zn spike. The

323  $^{67}\text{Zn}$  spike solution was confirmed to contain virtually no  $^{110}\text{Cd}$ ,  $^{111}\text{Cd}$ , nor  $^{114}\text{Cd}$ . The  $^{110}\text{Cd}$   
324 spike was likewise confirmed to contain virtually no  $^{67}\text{Zn}$ ,  $^{64}\text{Zn}$ , nor  $^{66}\text{Zn}$ . As a result, we  
325 assumed that the added  $^{67}\text{Zn}$  spike did not affect the pre-existing Cd, nor did the  $^{110}\text{Cd}$  spike  
326 affect the pre-existing Zn. It is assumed that the pre-existing particulate blank was in steady  
327 state, i.e. that it represented the Cd or Zn already in the particulate fraction and that any possible  
328 natural uptake that could occur during incubation for 24 h was negligible. The total dissolved  
329 pool of each metal isotope (denominator of each equation) is the sum of the dissolved  $^{110}\text{Cd}$  or  
330  $^{67}\text{Zn}$  added as the spike ( $^{110}\text{Cd}_{\text{Spike}}$ ,  $^{67}\text{Zn}_{\text{Spike}}$ ) plus the natural, pre-existing dissolved  $^{110}\text{Cd}$  or  $^{67}\text{Zn}$   
331 that was in the raw seawater ( $^{110}\text{Cd}_{\text{Natural}}$ ,  $^{67}\text{Zn}_{\text{Natural}}$ ) collected at each depth. To calculate  
332  $^{110}\text{Cd}_{\text{Natural}}$  and  $^{67}\text{Zn}_{\text{Natural}}$ , the total dissolved Cd or Zn measured by isotope dilution-ICP-MS  
333 ( $\text{Cd}_{\text{Total}}$ ,  $\text{Zn}_{\text{Total}}$ ) was multiplied by the natural abundance of  $^{110}\text{Cd}$  and  $^{67}\text{Zn}$  (12.49% and 4.10%,  
334 respectively). Dividing the particulate  $^{110}\text{Cd}$  and  $^{67}\text{Zn}$  by the total dissolved  $^{110}\text{Cd}$  and  $^{67}\text{Zn}$  yields  
335 the fraction of these metal isotopes that moved from the dissolved pool to the particulate pool per  
336 day (equation 3 and equation 4, respectively):

$$337 \quad \text{Cd}_{\text{total}} \text{ Uptake Rate (pmol L}^{-1} \text{ d}^{-1}) = \frac{[^{110}\text{Cd}_{\text{Sample}} (\text{pmol L}^{-1} \text{ d}^{-1}) - ^{110}\text{Cd}_{\text{PEP}} (\text{pmol L}^{-1} \text{ d}^{-1})]}{[^{110}\text{Cd}_{\text{Spike}} (\text{pmol L}^{-1}) + ^{110}\text{Cd}_{\text{Natural}} (\text{pmol L}^{-1})]} \times \text{Cd}_{\text{total}} (\text{pmol L}^{-1}) \quad (3)$$

$$338 \quad \text{Zn}_{\text{total}} \text{ Uptake Rate (pmol L}^{-1} \text{ d}^{-1}) = \frac{[^{67}\text{Zn}_{\text{Sample}} (\text{pmol L}^{-1} \text{ d}^{-1}) - ^{67}\text{Zn}_{\text{PEP}} (\text{pmol L}^{-1} \text{ d}^{-1})]}{[^{67}\text{Zn}_{\text{Spike}} (\text{pmol L}^{-1}) + ^{67}\text{Zn}_{\text{Natural}} (\text{pmol L}^{-1})]} \times \text{Zn}_{\text{total}} (\text{pmol L}^{-1}) \quad (4)$$

## 339 **2.9 Nutrient analyses**

340 Seawater samples taken for macronutrient analysis were filtered through 0.2  $\mu\text{m}$  Supor  
341 (Pall) membrane filters and frozen at sea in acid-washed 60-mL high-density polyethylene  
342 (HDPE) bottles until analysis. Nutrient analyses were conducted by nutrient autoanalyzer by Joe  
343 Jennings at Oregon State University using previously described methods (Noble et al., 2012).

## 344 **2.10 Statistics and plotting**



345 Dissolved ecological stoichiometries were obtained from the slopes of two-way (type II)  
346 least squares linear regressions performed using the script lsqfitma.m rewritten from MATLAB  
347 to Python by Rebecca Chmiel (<https://github.com/rebecca-chmiel/GP15>). A correlation matrix of  
348 various parameters measured during NBP18-01 was created with SciPy v1.5.2 using the  
349 ‘scipy.stats.pearsonr’ function, yielding Pearson correlation coefficients and p values that were  
350 visually represented using Seaborn v.0.11.1 and Matplotlib v3.3.2. Ocean sections were plotted  
351 using Ocean Data View v5.3.0 with gridded bathymetry file ETOPO1\_2min. Outliers (see Data  
352 Availability) were excluded from ocean sectional plots. Mixed layer depth was calculated using  
353 the potential density function (pden) within the python-seawater module (v3.3.4). Depth-  
354 integrated uptake rates were calculated using the ‘auc’ function within the Scikit-learn (v0.23.2)  
355 Python library. Figures were made using matplotlib (v3.3.2), Ocean Data View (v5.5.2), Excel  
356 (2019), and RStudio (v1.3.1093). ODV color palettes (<https://doi.org/10.5281/zenodo.1243862>)  
357 are inverse ‘roma’ for trace metal and macronutrient concentrations, ‘thermal’ for Zn and Cd  
358 uptake rates, and ‘algae’ for total fluorescence (Crameri, 2023).

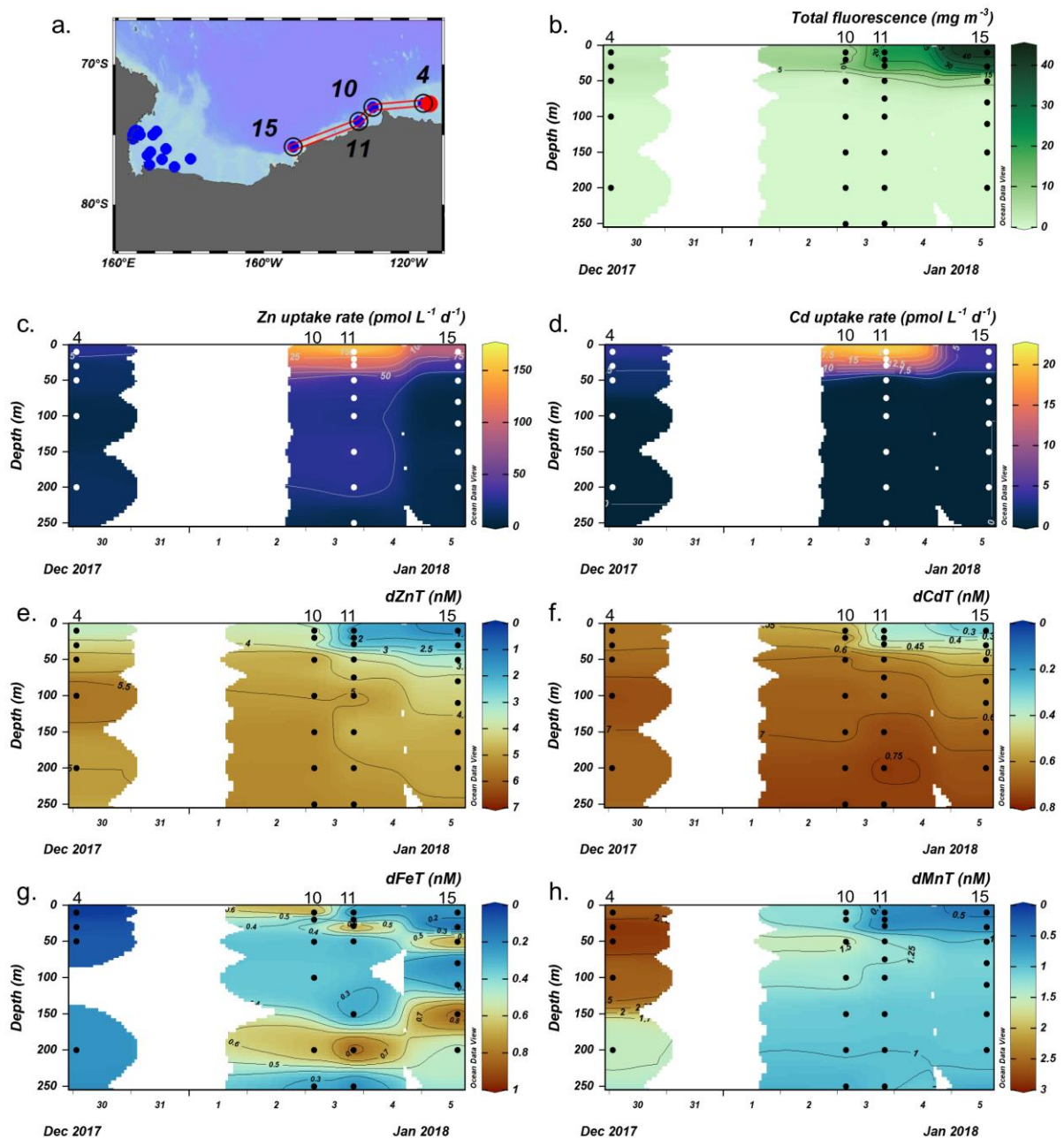
### 359 **3 Results**

#### 360 **3.1 Amundsen Sea**

361 Zn and Cd uptake rate experiments were conducted at 18 stations. We define 3 groups of  
362 stations based on location: the Amundsen Sea, Ross Sea, and Terra Nova Bay (TNB) groups  
363 (Fig. 1). Uptake rates were assessed at 3 stations (4, 11 and 15) within the Amundsen Sea group,  
364 6 stations (20, 29, 32, 35, 62, and 67) within the Ross Sea group, and 9 stations (22, 27, 41, 46,  
365 52, 57, 72, 76 and 79) within the TNB group spanning ~10 – 250 m depth for a total of 18  
366 stations and 125 samples. An overall schematic detailing these experiment workflows is shown  
367 in Fig. 2. The experimental design was validated by comparison of surface particulate  $^{67}\text{Zn}$ : $^{68}\text{Zn}$

368 and  $^{110}\text{Cd}:^{114}\text{Cd}$  ratios measured in spiked samples with natural abundance ratios. Samples  
369 spiked with  $^{67}\text{Zn}$  had particulate  $^{67}\text{Zn}:^{68}\text{Zn}$  ratios larger than natural abundance ratios at all  
370 stations (as was also true for  $^{110}\text{Cd}$  spiked samples and Cd natural abundances; Fig. S1),  
371 indicative of uptake of the spike into the particulate phase.

372           The Amundsen Sea stations represented a linear cruise track, and we report total  
373 dissolved metal concentrations ( $\text{dMetal}_T$ ) and uptake rates ( $\rho\text{Metal}$ ) over time in order of station  
374 sampling date (Fig. 3a).

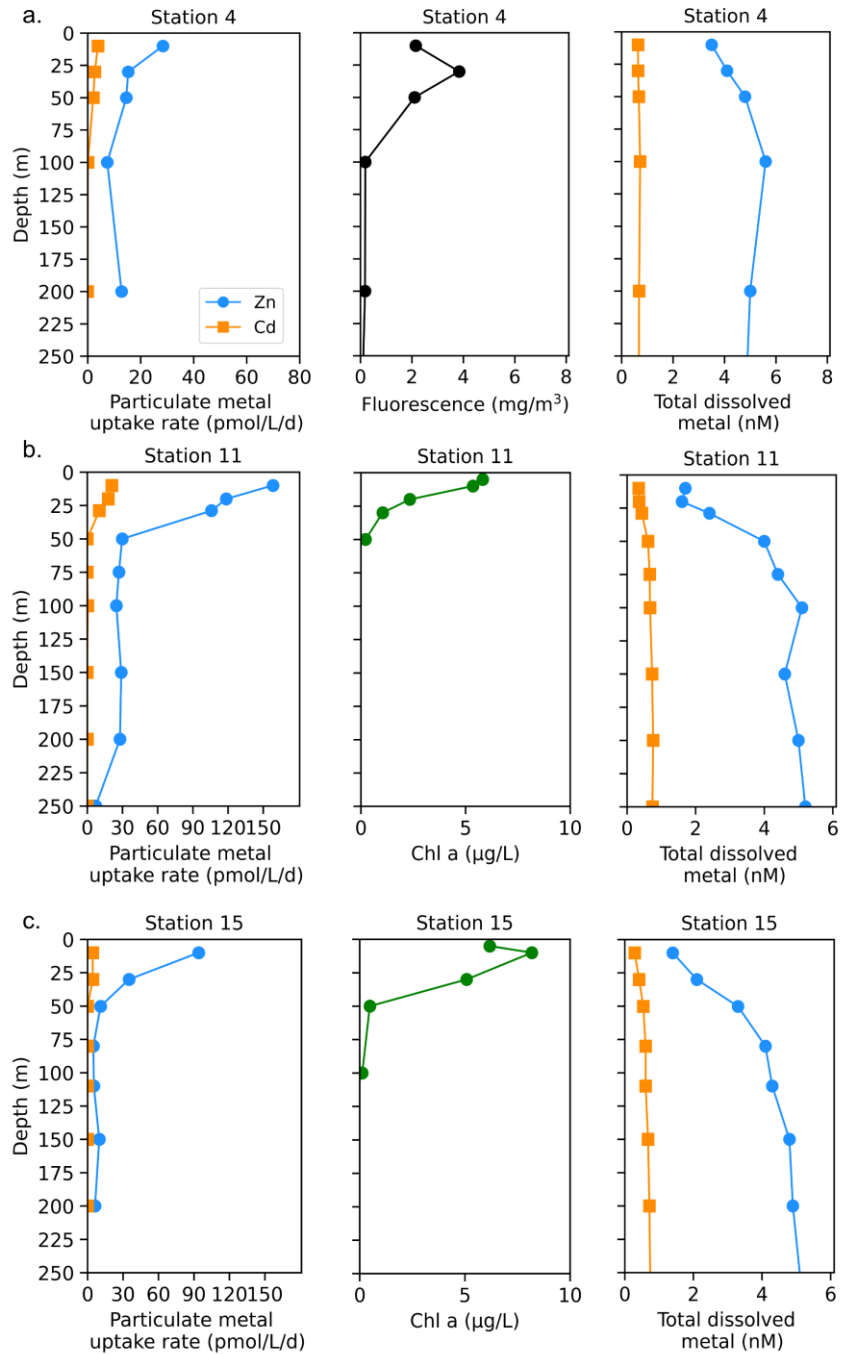


375  
 376 **Figure 3.** Total fluorescence and trace metal concentrations measured at Amundsen Sea stations  
 377 shown over time. (a) Map showing station locations, (b) total chlorophyll (Chl) fluorescence, (c)  
 378 total Zn uptake rates, (d) total Cd uptake rates, (e) total dissolved Zn, (f) total dissolved Cd, (g)  
 379 total dissolved Fe, and (h) total dissolved Mn measured in the upper 250 m represented in color  
 380 scale. Uptake experiments were not performed at station 10. Metal concentrations measured to  
 381 500 m depth are shown in Figure S3. dZnT, total dissolved Zn; dCdT, total dissolved Cd; dFeT,  
 382 total dissolved Fe; dMnT, total dissolved Mn.  
 383

384 Among these stations, total Chl fluorescence was lowest at station 4 and increased moving  
385 westward along the transect to a Chl maximum of  $41.8 \text{ mg m}^{-3}$  at station 15, 10 m (Fig. 3b).  
386 Maximum surface concentrations of dZn, dCd and dMn were highest at station 4 (3.5 nM, 639  
387 pM, and 2.6 nM at 10 m, respectively; Fig. 3e, f, h), likely reflecting the relatively smaller  
388 amount of total biomass (as indicated by total Chl fluorescence; Fig. 3b) at this station.  
389 Concentrations of dZn, dCd and dMn decreased moving westward along the transect (Fig. 3e, f,  
390 h) as total Chl fluorescence increased (Fig. 3b). Total Zn uptake rates ( $\rho\text{Zn}$ ) and total Cd uptake  
391 rates ( $\rho\text{Cd}$ ) were highest at station 11 ( $158$  and  $21 \text{ pmol L}^{-1} \text{ d}^{-1}$ , respectively, at 10 m; Fig. 3c,d;  
392 Fig. 4b). Among the three Amundsen Sea stations, the largest movement of both Zn and Cd into  
393 the particulate phase therefore occurred at station 11, concurrent with the relatively higher dFe<sub>T</sub>  
394 surface values observed at station 11 (0.2 nM dFe compared to 0.01 nM at station 4, 10 m; Fig.  
395 3g, Fig. S3c). The dFe concentrations exceeding 1 nM near the seafloor are consistent with a  
396 sedimentary or subglacial source (Fig. S3c). Overall,  $\rho\text{Zn}$  and  $\rho\text{Cd}$  profiles exhibited trends in  
397 which values were highest within the upper 50 m at all three stations and decreased with depth,  
398 following the trend in Chl a or total Chl fluorescence (Fig. 4). Vertical sections of dZn and dCd  
399 through the water column mirrored these trends (Fig. 4), demonstrating the movement of these  
400 dissolved metal micronutrients into the particulate phase.

401

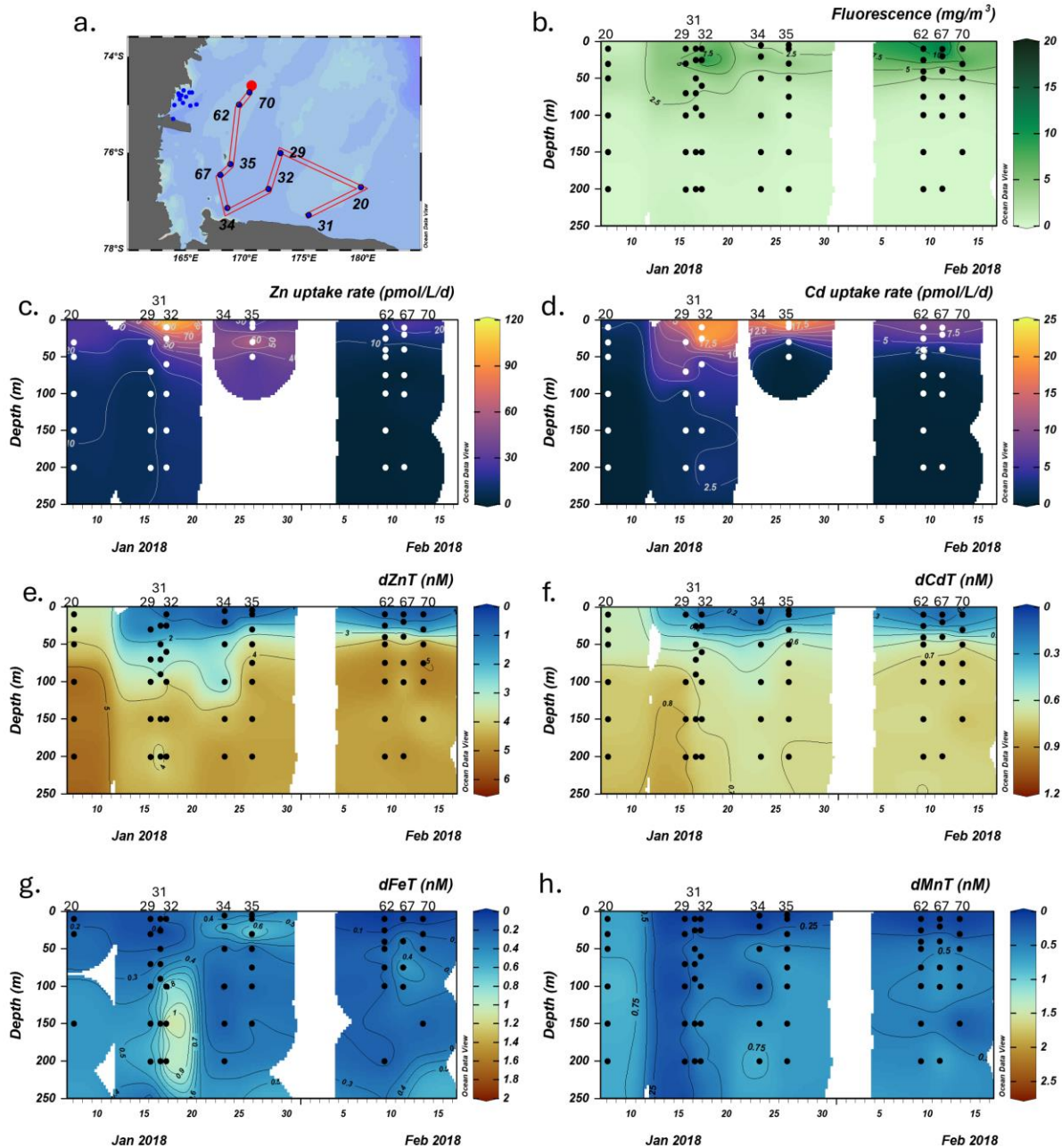
402



403  
 404 **Figure 4.** Depth profiles of total Zn and Cd uptake rates, total chlorophyll fluorescence (or  
 405 chlorophyll *a*) and total dissolved metal measured in the upper 250 m at (a) station 4, (b) station  
 406 11, and (c) station 15 sampled along the Amundsen Sea shelf. Total chlorophyll (Chl)  
 407 fluorescence is reported for stations where chlorophyll *a* (Chl *a*) data was not measured.  
 408

409 **3.2 Ross Sea**

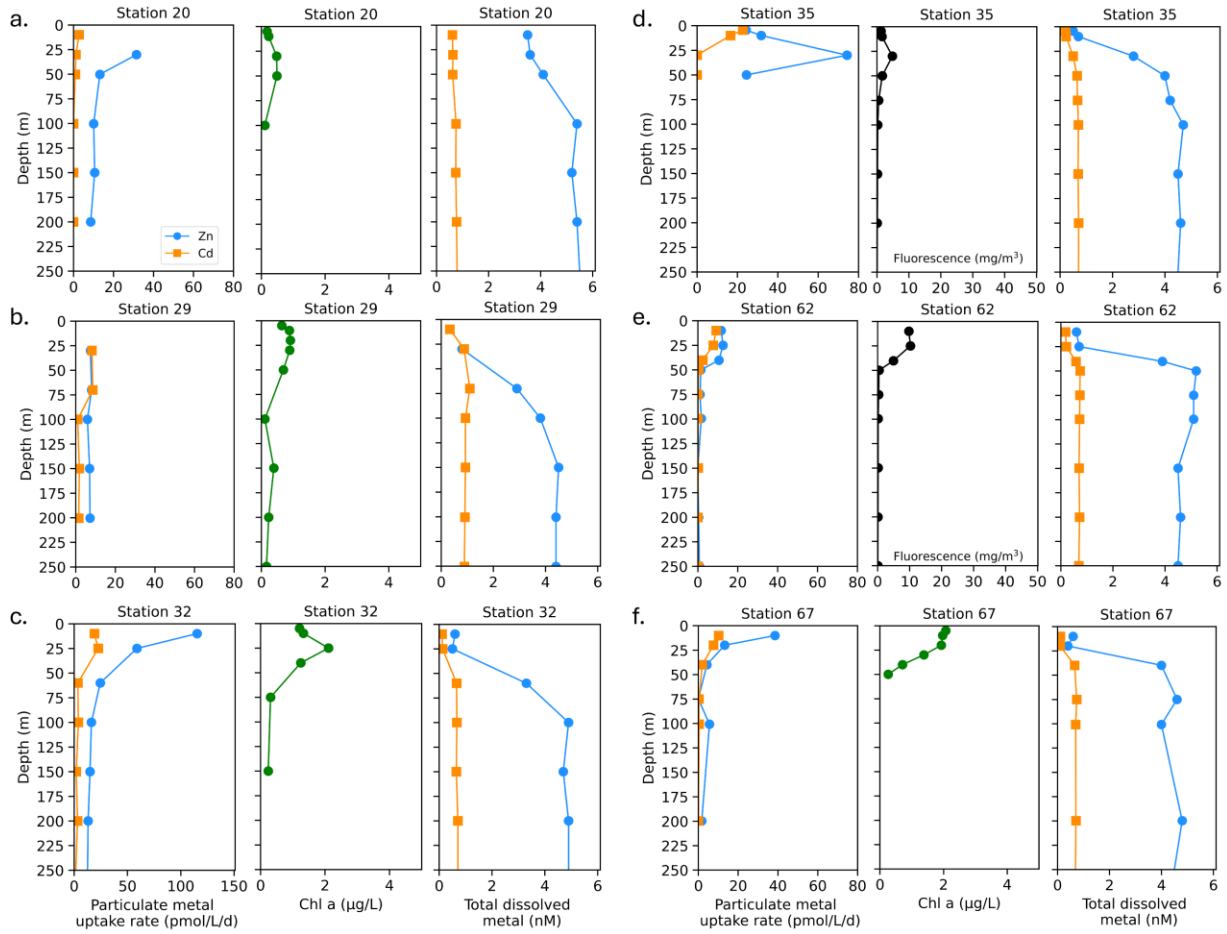
410 We next investigated the dissolved Zn and Cd demand of the natural phytoplankton  
 411 community at stations sampled over the Ross Sea shelf. Data collected from this group is  
 412 presented over time, in order of sampling date (Fig. 5).



413  
 414 **Figure 5.** Total fluorescence and trace metal concentrations measured at Ross Sea stations  
 415 shown over a latitudinal transect. (a) Map showing station locations, (b) total chlorophyll (Chl)  
 416 fluorescence, (c) total Zn uptake rates, (d) total Cd uptake rates, (e) total dissolved Zn, (f) total  
 417 dissolved Cd, (g) total dissolved Fe, and (h) total dissolved Mn measured in the upper 250 m

418 represented in color scale. Uptake experiments were not performed at stations 31, 34, and 70.  
419 Metal concentrations measured to 800 m depth are shown in Figure S4. dZnT, total dissolved Zn;  
420 dCdT, total dissolved Cd; dFeT, total dissolved Fe; dMnT, total dissolved Mn.  
421

422 We note that unlike the Amundsen Sea sector, the stations sampled in this group did not  
423 follow a linear cruise track, thus we cannot make inferences regarding latitudinal or longitudinal  
424 changes. Surface Chl fluorescence was highest at stations 32 and 67 with maximum values of  
425 15.8 and 14.6 mg m<sup>-3</sup> at 25 m and 10 m, respectively (Fig. 5b). With the exception of station 20,  
426 dZn and dCd demonstrated high levels of surface depletion within the upper 25 m (Fig. 5e,f)  
427 with average concentrations of  $0.63 \pm 0.13$  nM and  $0.19 \pm 0.09$  nM respectively at  $\leq 10$  m.  
428 Compared to 100 m values (i.e., below the MLD), concentrations at 25 m were equivalent to  
429 87%, 34%, 85%, 85%, 77%, and 88% decreases in dZn and 83%, 19%, 84%, 67%, 64%, and  
430 75% decreases in dCd at stations 32, 20, 67, 35, 29, and 62, respectively. Measured dMn  
431 concentrations were also highly depleted within the upper 250 m at all Ross Sea stations  
432 (average 10 m dMn =  $0.18 \pm 0.26$  nM; Fig. 5h). While dFe was depleted within the upper 10 m at  
433 all stations (average dFe concentration at 10 m =  $0.12 \pm 0.12$  nM), concentrations exceeding 1  
434 nM were observed below 100 m at station 32 (Fig. 5g) and extended down to 650 m (Fig. S4c),  
435 implying a sedimentary source. The largest Zn uptake rate measured among all stations in this  
436 group ( $115 \text{ pmol L}^{-1} \text{ d}^{-1}$ ) was observed at station 32, 10 m (Fig. 6c). As observed in the  
437 Amundsen Sea,  $\rho\text{Zn}$ ,  $\rho\text{Cd}$  and total Chl a or Chl fluorescence profiles exhibited surface maxima  
438 and became depleted with depth and were again mirrored by nutrient-like dZn and dCd depth  
439 profiles (Fig. 6), indicative of uptake of these metals into the particulate phase in surface waters.



440

441 **Figure 6.** Depth profiles of total Zn and Cd uptake rates, total chlorophyll fluorescence (or,  
 442 where available, chlorophyll a), and total dissolved metal ( $dMetal_T$ ) measured in the upper 250 m  
 443 at (a) station 20, (b) station 29, (c) station 32, (d) station 35, (e) station 62, and (f) station 67  
 444 sampled along the Ross Sea shelf. Total chlorophyll (Chl) fluorescence is reported for stations  
 445 where chlorophyll a (Chl a) data was not measured.  
 446

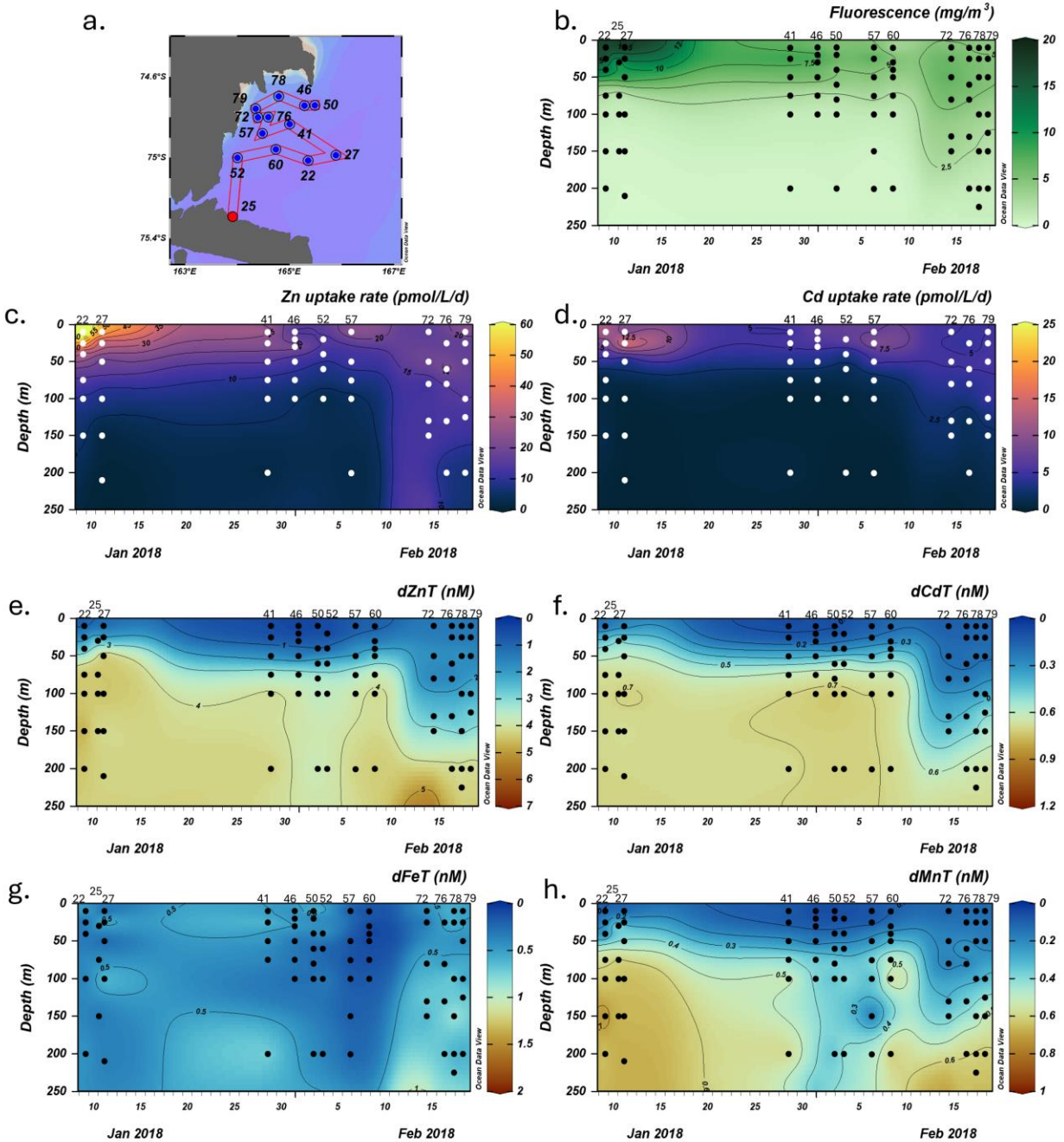
### 447 3.3 Terra Nova Bay

448 Zinc and Cd uptake rate data collected from stations sampled in Terra Nova Bay (TNB)  
 449 were visualized over time due to repeated sampling within a small geographic region and similar  
 450 timeframe (Fig. 7a). This allowed for an analysis of how dissolved metal concentrations and  
 451 metal uptake rates changed throughout January-February 2018 within the same spatial area.



452 Station data is presented in order of sampling date, from the earliest (station 22, sampled in early  
 453 January) to the latest (station 79, sampled in late February).

454



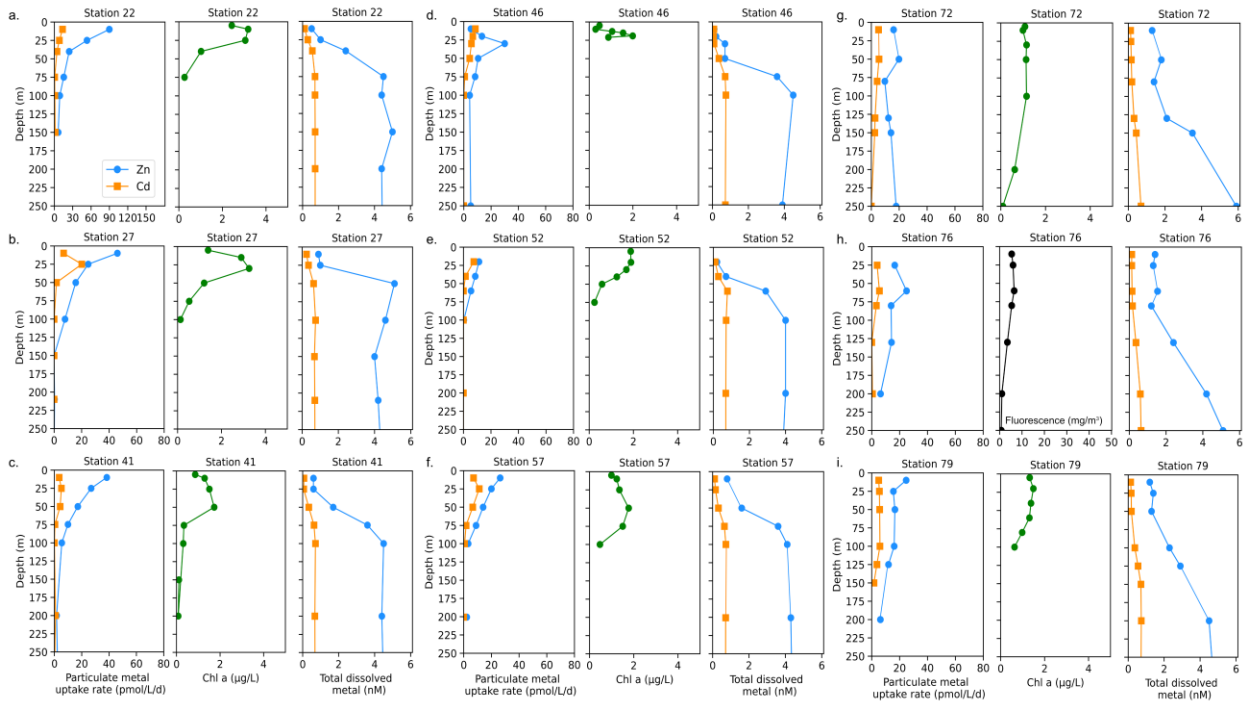
455 **Figure 7.** Total fluorescence and trace metal concentrations measured at Terra Nova Bay (TNB)  
 456 stations shown over time. (a) Map showing station locations, (b) total chlorophyll (Chl)  
 457 fluorescence, (c) total Zn uptake rates, (d) total Cd uptake rates, (e) total dissolved Zn, (f) total  
 458 dissolved Cd, (g) total dissolved Fe, and (h) total dissolved Mn measured in the upper 250 m  
 459 represented in color scale. Uptake experiments were not performed at stations 70 and 34. Metal  
 460

461 concentrations measured to 600 m depth are shown in Figure S5. dZnT, total dissolved Zn;  
462 dCdT, total dissolved Cd; dFeT, total dissolved Fe; dMnT, total dissolved Mn.  
463

464 Surface Chl fluorescence was highest in early January ( $\sim 18 \text{ mg m}^{-3}$ ) and waned into  
465 February (Fig. 7b), similar to observed trends in Zn and Cd uptake rates (Fig. 7 c,d). Of all TNB  
466 stations, stations 22 and 27, sampled in January, had the highest maximum Zn uptake rates of  
467  $89.9 \text{ pmol L}^{-1} \text{ d}^{-1}$  and  $46.0 \text{ pmol L}^{-1} \text{ d}^{-1}$ , respectively, at 10 m (Fig. 8a,b). Cd uptake rates were  
468 also highest at these stations with values of  $13.4 \text{ pmol L}^{-1} \text{ d}^{-1}$  and  $20.1 \text{ pmol L}^{-1} \text{ d}^{-1}$  (Fig. 8a,b). At  
469 the final station (station 79, sampled in late February) maximum uptake rates of both metals had  
470 sharply decreased to  $24.7 \text{ pmol Zn L}^{-1} \text{ d}^{-1}$  and  $5.0 \text{ pmol Cd L}^{-1} \text{ d}^{-1}$  (Fig. 8i). Overall, maximum  
471 uptake rates of both metals decreased over time within TNB (Fig. 7c,d), consistent with the  
472 decrease in total Chl fluorescence (Fig. 7b) likely due to the aging and decline of the  
473 phytoplankton bloom.

474 Surface depletion of dZn, dCd, and dMn was observed at all stations with average  
475 dissolved concentrations of  $0.82 \pm 0.47 \text{ nM Zn}$ ,  $0.13 \pm 0.06 \text{ nM Cd}$ , and  $0.08 \pm 0.04 \text{ nM Mn}$  at  
476 10 m depth (Fig. 7e,f,h). Notably, increased surface concentrations of dZn, dCd, and dMn were  
477 apparent at the late stations 72, 76, 78 and 79, with dZn  $\sim 2 \text{ nM}$ , dCd  $\sim 300 \text{ pM}$ , and dMn  $\sim 0.2$   
478  $\text{nM}$  (Fig. 7e,f,h; Fig. S5). Dissolved macronutrient (phosphate, nitrate and nitrite, and silicate)  
479 concentrations also followed this trend, with increased surface concentrations at the late stations  
480 (Fig. S6). As with the Amundsen and Ross Sea station groups, Zn and Cd uptake rates within  
481 TNB tended to be highest at the surface  $\leq 50 \text{ m}$  as also observed in total Chl fluorescence trends  
482 and mirrored the decrease in total dissolved Zn and Cd (Fig. 8). Unlike the Amundsen and Ross  
483 Sea stations, where Cd uptake consistently became negligible ( $\sim 0 \text{ pM L}^{-1} \text{ d}^{-1}$ ) by 100 m (Fig. 4;  
484 Fig. 6), measurable Cd uptake persisted in TNB to 150 m at stations 72 and 79 (Fig. 8g,i).

485 Measurable Zn uptake rates were also captured at deeper depths at these late TNB stations (Fig.  
 486 8g,h,i).



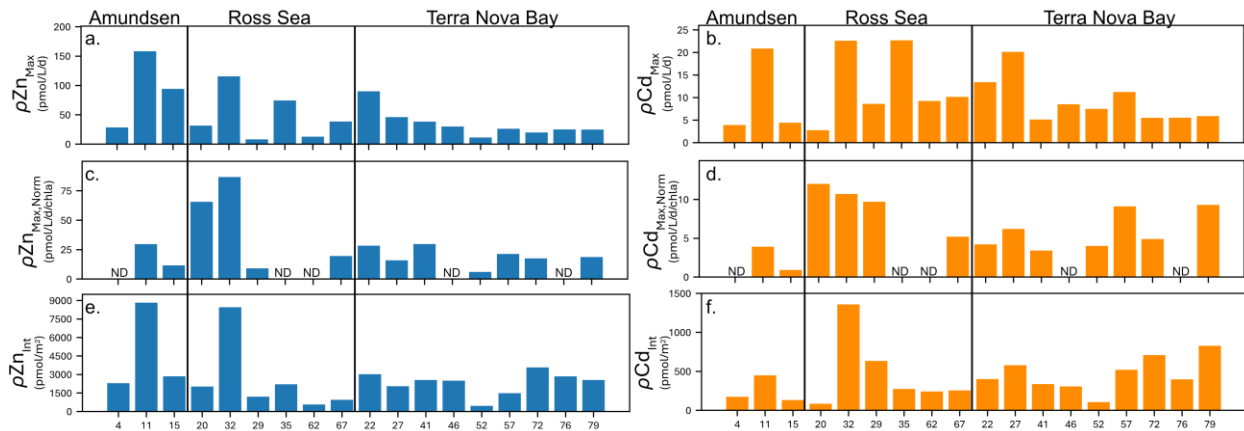
487  
 488 **Figure 8.** Depth profiles of total Zn and Cd uptake rates, total chlorophyll fluorescence (or,  
 489 where available, chlorophyll a), and total dissolved metal (dMetalT) measured in the upper 250  
 490 m at (a) station 22, (b) station 27, (c) station 41, (d) station 45, (e) station 52, (f) station 57, (g)  
 491 station 72, (h) station 76, and (i) station 79 within Terra Nova Bay. Total chlorophyll (Chl)  
 492 fluorescence is reported for stations where chlorophyll a (Chl a) data was not measured.  
 493

494 The increased surface concentrations of dZn and dCd and macronutrients, as well as the  
 495 persistence of measurable uptake rates at deeper depths, at these late TNB stations may be  
 496 attributed to the deepening of the mixed layer (Fig. S7). Vertical mixing was evidenced by more  
 497 uniform potential densities, temperatures, dissolved oxygen (O<sub>2</sub>) concentrations, salinity, and  
 498 beam transmission measurements at the late TNB stations within the upper 200 m (Fig. S7).  
 499 Higher (>0.5 nM) dFe concentrations were also observed below 100 m at these late stations (Fig.  
 500 7g) and increased with depth (>2 nM), as did dZn and dMn concentrations, possibly due to  
 501 sedimentary inputs (Giordano et al., 1999) (Fig. S5). At these late stations (Station 76, 78, 79)

502 mixing replenished surface concentrations of both macronutrients (Fig. S6) and dZn (Fig. S5a),  
 503 but dZn was replenished to a lower extent. For example, comparing 50 m “replenished” surface  
 504 values of P, N+N, and Si to deepwater (200 m) values at Station 79, percent changes from deep  
 505 to surface values were -0.35% for P, -0.30% for N+N, and -0.26% for Si (where a percent change  
 506 of 0 would indicate complete replenishment; i.e, if nutrient values at 200 m and at 50 m were  
 507 equal). In contrast, the percent change from deep (200m) to surface (50m) dZn at Stn79 was  
 508 lower, -0.71%. Hence, dZn was apparently replenished to a lesser extent compared to  
 509 macronutrients, which may reflect a sustained high demand for Zn generating a dearth of this  
 510 micronutrient despite macronutrient replenishment.

### 511 3.4 Overview of Zn and Cd uptake at 18 stations

512 We next summarize the maximum Zn and Cd uptake rates observed at each station (all of  
 513 which were observed at  $\leq 10$  m depth; Fig. 9a,b) with uptake rates normalized to Chl a ( $\mu\text{g/L}$ ) as  
 514 a proxy for biomass (Fig. 9c,d).



515

516 **Figure 9.** Unnormalized (a) maximum Zn uptake rates ( $\rho\text{Zn}_{\text{Max}}$ ) and (b) maximum Cd uptake  
 517 rates ( $\rho\text{Cd}_{\text{Max}}$ ) at each station grouped by area (Amundsen Sea, Ross Sea, Terra Nova Bay). (c)  
 518  $\rho\text{Zn}_{\text{Max}}$  and (d)  $\rho\text{Cd}_{\text{Max}}$  normalized to chlorophyll a ( $\mu\text{g L}^{-1}$ ) measured at each station. (e) Depth  
 519 integrated (10 m-250 m)  $\rho\text{Zn}$  and  $\rho\text{Cd}$  values at each station. ND, no data (chlorophyll a not  
 520 measured).  
 521

522 Overall, high ( $>25 \text{ pmol L}^{-1} \text{ d}^{-1} \text{ Chl a (ug/L)}^{-1}$ ) Chl a-normalized Zn uptake rates were measured  
523 at station 11 in the Amundsen Sea and at stations 20 and 32 in the Ross Sea (Fig. 9c). The  
524 highest Chl a-normalized Cd uptake rates among all 18 stations were also measured at stations  
525 20 and 32 (Fig. 9d). Across TNB, Chl a-normalized maximum Zn and Cd uptake ranged from  
526  $6.0 - 28.3 \text{ pmol L}^{-1} \text{ d}^{-1} \text{ Chl a}^{-1}$  for Zn, and  $3.4 - 9.3 \text{ pmol L}^{-1} \text{ d}^{-1} \text{ Chl a}^{-1}$  for Cd; Fig. 9c,d).  
527 Integrated (10 m-250 m) uptake rate values were highest for Zn at stations 11 and 32, and highest  
528 for Cd at station 32 (Fig. 9e,f). Increases in integrated Cd and Zn uptake at the late stations 72,  
529 76 and 79 reflected the deeper depths to which uptake rates of these metals remained measurable,  
530 likely reflecting deepened mixed layers (Fig. S7) and/or sinking of the phytoplankton  
531 community, as seen in the fluorescence data to beyond 150m depth (Fig. 7b). The presence of  
532 Chl a (Fig. 8g,i) implies these deep phytoplankton communities may have been alive, if not  
533 actively photosynthesizing. We previously identified ZCRP-B, a membrane-associated protein  
534 involved in high-affinity Zn transport (Kellogg et al., 2022). These proteins have a single  
535 transmembrane domain, implying function as a membrane-tethered ligand to assist in the  
536 acquisition of Zn from seawater in cooperation with adjacent zinc transporters (ZIP transporters).  
537 Hence ZCRP-B could be a potential site of Zn binding and ‘uptake’, as our uptake rate  
538 measurements do not discern between extracellular and intracellular Zn, even if the  
539 phytoplankton are inactive due to a lack of photosynthetic energy at these depths.

540 Notably, maximum Cd uptake rates measured in the present study were 3.4, 3.7, and 3.3  
541 times higher in the Amundsen Sea, Ross Sea, and Terra Nova Bay, respectively, compared to the  
542 maximum Cd uptake rate of  $6.1 \text{ pmol L}^{-1} \text{ d}^{-1}$  measured previously within the Costa Rica Dome  
543 using identical methods (Cox et al., 2014), demonstrating the influence of high productivity and  
544 the native community on the flux of dCd into the particulate phase.

## 545 **4 Discussion**

### 546 **4.1 Use of metal uptake rates to determine depletion timeframes**

547           The measurement of total dissolved metal concentrations over large latitudinal or  
548 longitudinal areas allows for the characterization of metal inventories, though these are snapshots  
549 of inventories observed at specific times. The measurement of metal uptake rates allows us to  
550 gain new insight into how these inventories came to be and the timeframes over which they are  
551 consumed and replenished. Due to the resetting of surface dissolved metal concentrations to  
552 those of deepwater values during austral winter with deep winter mixing, the Ross Sea of the  
553 Southern Ocean is particularly applicable to this type of timeframe study (Sedwick and DiTullio  
554 1997; Sedwick et al. 2011).

555           Using the Zn uptake rates measured in this study, we can estimate the time required for  
556 the high levels of primary production observed in the Southern Ocean to draw down surface dZn  
557 from high (deep water) winter concentrations to the surface concentrations observed during  
558 austral summer 2017. The Southern Ocean growing season typically spans October-March, with  
559 primary productivity peaking November-January and the area of open (ice-free) water over the  
560 Ross Sea shelf linearly increasing from November-mid January (Sedwick et al., 2011). Vertical  
561 profiles of nutrients and micronutrients in coastal Antarctic ecosystems such as the Ross Sea are  
562 reset and become uniform with depth during the winter months due to whole water column  
563 mixing and an absence of photosynthetic activity during the dark winter under the sea ice (Noble  
564 et al., 2013). As a result, the drawdown of nutrients in the upper water column observed during  
565 the spring and summer seasons is the result of less than one year's biological influence. For this  
566 simple calculation, we ignore the upward flux of Zn (upwelling = 0) and assume a high export  
567 ratio of 0.8 due to bloom productivity being dominated by diatoms and *Phaeocystis antarctica*,

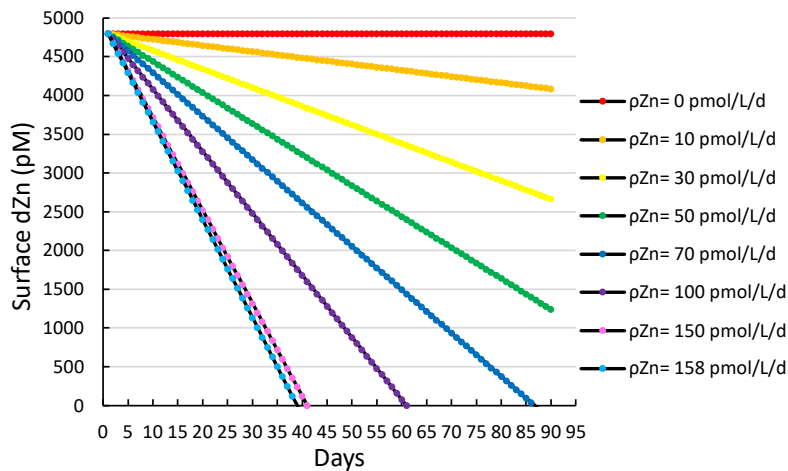
568 both of which sink rapidly and thus contribute substantially to carbon export flux (Asper and  
 569 Smith 1999; DiTullio et al. 2000). The depletion of dZn from a surface box was therefore  
 570 estimated as:

$$571 \quad \left(\frac{dZn}{dt}\right)_{\text{surface box}} = -\rho Zn + (Rf * \rho Zn) + \text{upwelling}$$

572 Where Rf is the remineralization factor equal to 1 - export ratio.

573 Taking station 11, for which the highest Zn uptake rate was observed, as an extreme case:  
 574 with a maximum Zn uptake rate of 158 pmol L<sup>-1</sup> d<sup>-1</sup>, it would take only 25 days to deplete a  
 575 surface winter concentration of 4.8 nM (that is, the average deepwater (< 200 m) dZn  
 576 concentration for all stations measured in this study) down to the observed surface concentration  
 577 of 1.7 nM at station 11 (Fig. 4b), assuming a constant uptake rate and no additional inputs of  
 578 dissolved Zn (Fig. 10).

579



580 **Figure 10.** A simple model estimating the time (in days) required to deplete the estimated  
 581 average winter surface concentration of dZn (4.8 nM) over a range of various Zn uptake rates  
 582 (ρZn). 158 pmol/L/d was the maximum Zn uptake rate observed in this study (station 11, 10 m).  
 583  
 584

585 Given that dZn surface depletion to sub-nanomolar levels was observed throughout much  
 586 of the CICLOPS expedition, prolonged high levels of Zn uptake and export that overwhelm

587 replenishment by vertical mixing and/or remineralization are likely key to giving rise to the  
588 observed extent of seasonal surface dZn depletion. These calculations were conducted as a proof-  
589 of-concept to determine if uptake rates were sufficient to draw down the otherwise abundant dZn  
590 inventory on seasonal timescales. An important caveat to this calculation is that the regulation  
591 and production of Zn sensors that modulate Zn uptake, export, and storage will naturally  
592 fluctuate in response to changing dZn and therefore cannot be assumed as a constant. Future  
593 studies could conduct mesoscale modeling of the region, including eddy diffusion and advection.  
594 Notably, any dZn upwelling flux into the euphotic zone would require even higher Zn uptake  
595 rates to create the seasonal surface Zn depletion we observed on this expedition.

596

#### 597 **4.2 Influences on Zn and Cd uptake**

598 We next consider the factors driving the magnitude of  $\rho\text{Zn}$  and  $\rho\text{Cd}$ . As noted above,  $\rho\text{Zn}$   
599 and  $\rho\text{Cd}$  were positively correlated with total Chl fluorescence or Chl *a* at every station (Fig. 4;  
600 Fig. 6; Fig. 8), demonstrating the influence of total autotrophic biomass on uptake rates. A  
601 Pearson correlation analysis comparing the abundance of individual algal pigments to  $\rho\text{Zn}$  and  
602  $\rho\text{Cd}$  throughout the water column for all stations revealed significant, positive correlations  
603 (Pearson correlation coefficient  $> 0.50$ ,  $p \leq 1.2e-4$ ) between  $\rho\text{Zn}$  and Chl *a*, Chl *b*, and Chl *c*1, *c*2  
604 and *c*3. Pearson correlation coefficients are normally symbolized as rho ( $\rho$ ), but to avoid  
605 confusion with our uptake rate symbol ( $\rho$ ), and with p-values ( $p$ ), they are herein referred to as  
606 ‘PCC’ values. The correlation between  $\rho\text{Zn}$  and Chl *b* was strongest (PCC= 0.77,  $p = 3.8e-10$ ) of  
607 any pigment (Fig. 11e).





624 In bottle incubation experiments conducted at station 27, the addition of Zn alone resulted  
625 in the positive growth response of Chl *b*-containing algae (small green algae such as  
626 prasinophytes; (Kell et al., 2023)), corroborating this finding.  $\rho$ Cd also positively correlated with  
627 these Chl pigments but with slightly lower correlation Pearson correlation coefficients (PCC =  
628 0.3-0.51;  $p \leq .043$ ). Fucoxanthin (fuco) concentrations were more highly correlated with  $\rho$ Cd  
629 (PCC = 0.57,  $p = 4.3e-5$ ) than with  $\rho$ Zn (PCC = 0.32,  $p = 2.9e-2$ ), while the opposite was  
630 observed for 19'-Hex (19'-hexanoyloxyfucoxanthin; PCC = 0.54,  $p = 1.1e-4$  for Zn; not  
631 significant for Cd) (Fig. 11e). Fuco and 19'-Hex are used as taxonomic indicators of diatoms and  
632 *Phaeocystis*, respectively, in the Ross Sea (DiTullio et al., 2003, 2007; DiTullio and Smith,  
633 1995; Wright et al., 2010). The higher correlation coefficient between  $\rho$ Zn and *Phaeocystis*  
634 abundance (as indicated by 19'-Hex) implies that Zn uptake was driven largely by *Phaeocystis*.  
635 This finding is consistent with the detection of *Phaeocystis* ZCRP-A, a protein characterized as  
636 an algal  $Zn^{2+}$  metallochaperone (Kellogg et al., 2022), in metaproteomic data collected from both  
637 the incubation experiment and throughout the water column at station 27 (Kell et al., 2023). The  
638 positive correlation between  $\rho$ Cd and the abundance of diatoms (as indicated by fuco) is  
639 consistent with the diatomic utilization of Cd as a nutrient within CDCA metalloenzymes, as  
640 *cdca* genes have, to date, been found exclusively in diatom species (Park et al., 2007, 2008).  
641 While it is likely that both *Phaeocystis* and diatoms contributed to the Cd and Zn uptake rates  
642 measured here, it is currently unknown if *Phaeocystis* can utilize Cd as a nutrient. Overall, any  
643 potential growth benefit conferred by our Cd spike additions may only have been applicable to  
644 diatoms that 1) possessed the *cdca* gene and 2) faced selection pressure to utilize Cd as a  
645 cofactor in CDCA due to low seawater  $pCO_2$  (as documented on this expedition) creating  
646 enhanced demand for dZn. The presence of Cd-utilizing diatoms in the water column at station

647 27 was demonstrated by the detection of CDCA transcripts with closest taxonomic matches to  
648 the diatom genera *Chaetoceros* and *Corethron* (Kell et al., 2023). Station 27 also exhibited high  
649 surface Chl fluorescence ( $19.3 \text{ mg m}^{-3}$  at 10 m), low  $\text{pCO}_2$  ( $221 \text{ } \mu\text{atm}$  at 5 m), and high  
650 maximum Zn and Cd uptake rates ( $46$  and  $20 \text{ pmol L}^{-1} \text{ d}^{-1}$ , respectively), demonstrating a high  
651 algal demand for Zn that likely created pressure for Cd uptake.

652 We next consider the effect of the depleted seawater  $\text{pCO}_2$  levels induced by the high  
653 biomass conditions observed on this expedition. Previously, a strong correlation between  
654 dissolved  $\delta^{114}\text{Cd}$  and dissolved  $\text{CO}_2$  was documented in the Atlantic Sector of the Southern  
655 Ocean (de Baar et al., 2017), suggesting significant Cd isotope fractionation due to biological  
656 uptake into the particulate phase. A relationship between total surface Cd uptake rates at 10 m  
657 and surface  $\text{pCO}_2$  (underway, measured at 5 m) was not observed in the present study (Fig. 11b).  
658 The present study includes measurements of total Cd uptake (that is, the sum of all Cd isotopes)  
659 using an added Cd isotope tracer, and hence did not explore natural isotope fractionation effects.  
660 However, we did observe a significant negative linear relationship between total Zn uptake rates  
661 and seawater  $\text{pCO}_2$  ( $m = -0.58$ ;  $R^2 = 0.63$ ; Fig. 11a) consistent with an increased demand for  $\text{Zn}^{2+}$   
662 to power the carbon concentrating mechanism of photosynthetic algae under lower  $\text{CO}_2$   
663 availability.  $\rho\text{Zn}$  and  $\rho\text{Cd}$  furthermore shared a significant positive linear relationship with each  
664 other ( $m = 0.13$ ;  $R^2 = 0.64$ ; Fig. 11c) (as was also reflected in the Pearson correlation test;  $\text{PCC} =$   
665  $0.69$ ,  $p = 1.7\text{e-}7$ , Fig. 11d) implying that as demand for Zn increased, demand for Cd also  
666 increased, consistent with laboratory studies showing their co-transport in marine algae (Sunda  
667 and Huntsman, 2000). We also note that  $\rho\text{Cd}:\rho\text{Zn}$  uptake ratios were higher ( $> 0.4$ ) at the surface  
668 where total dissolved  $\text{dCd}:\text{dZn}$  ratios were comparatively higher ( $> 0.3$ ) (Fig. S8a,b). The strong  
669 positive linear relationship shared between these ratios ( $R^2 = 0.82$ ; Fig. S8c) further suggests that

670 dZn levels were depleted enough to induce increased Cd uptake rates, and is consistent with their  
671 known biochemical substitution within marine algae.

672 Algal Cd uptake rates are known to be inversely related to both  $Mn^{2+}$  and  $Zn^{2+}$   
673 concentrations in culture (Lee et al., 1995; Sunda and Huntsman, 1996), which is thought to  
674 reflect the uptake of Cd by two separate inducible transport systems. Cd is taken up  
675 competitively by the high-affinity Zn uptake system under low  $Zn^{2+}$  conditions, as demonstrated  
676 above, while Cd, Zn, and Mn share the same low-affinity Mn uptake system under high  $Zn^{2+}$   
677 conditions (Lee et al., 1995; Sunda and Huntsman, 1998b, a, 2000; Xu et al., 2007). With the  
678 exception of the Amundsen Sea stations, dMn was consistently observed at concentrations of  
679 only 0.1-0.5 nM within the upper 50 m (Fig. 3h; Fig. 5h; Fig. 7h). Low surface dMn  
680 concentrations within the Southern Ocean have been documented previously and were attributed  
681 to a combination of biological uptake at the surface causing depletion and low resupply due to  
682 few external sources (Latour et al., 2021). While  $\rho Cd$  was negatively correlated with dMn (PCC  
683 = -0.34,  $p = 0.02$ ) considering all stations and all depths,  $\rho Cd$  was more strongly negatively  
684 correlated with dZn (PCC = -0.76,  $p = 1.1e-9$ ), which was the strongest negative correlation  
685 comparing all measured parameters to  $\rho Cd$  (Fig. 11d). This finding is consistent with decreased  
686 dCd uptake where dZn availability is sufficient. Overall, these results are consistent with biology  
687 (total biomass) and  $pCO_2$  acting as primary influences on  $\rho Zn$ , with increases in  $\rho Zn$  leading to  
688 increases in  $\rho Cd$  through the upregulation of a shared transport system.

#### 689 **4.3 Effects on dCo cycling**

690 In addition to Cd, the intense Zn demand captured by these uptake rates also appears to  
691 have shifted the demand for the trace metal micronutrient cobalt (Co) (Chmiel et al., 2023). Due  
692 to their similar charge and atomic radii,  $Zn^{2+}$ ,  $Co^{2+}$  and  $Cd^{2+}$  cations often share the same

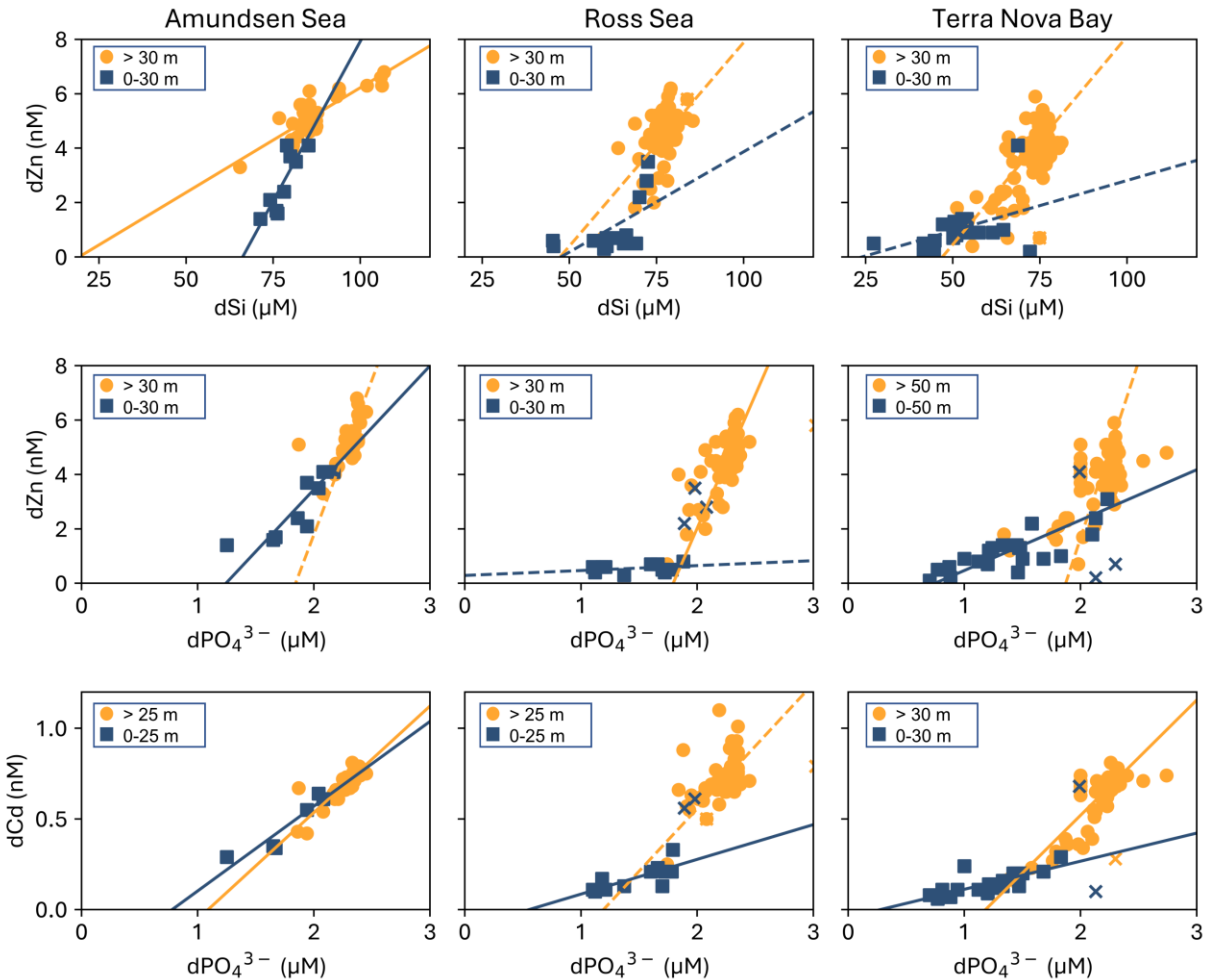
693 transporter uptake systems. An organism's ability to utilize these metals as metabolic cofactors is  
694 influenced by their environment and the affinity of the uptake ligands for each metal cation  
695 (Irving and Williams, 1953; Sunda and Huntsman, 1992). When dZn availability is low, more  
696 dCd and dCo are able to bind transport ligands. Therefore, dZn concentrations and cycling can  
697 influence the cycling of dCd and dCo, particularly in low dZn environments as documented for  
698 dCd in the present study. The influence of dZn cycling on dCo distributions in this region was  
699 also documented during this expedition, with evidence for high rates of biological Co uptake in  
700 the Ross Sea driven by dZn (and vitamin B<sub>12</sub>) scarcity (Chmiel et al., 2023). The high Zn uptake  
701 rates measured in this study therefore also reveal dynamic changes in the cycling of Cd and Co  
702 as a consequence of high Zn demand.

703

#### 704 **4.4 dZn and dCd relationships with macronutrients**

705 The growth of phytoplankton and bacteria in the shallow euphotic zone results in the  
706 removal of bioactive trace metals and macronutrients from the dissolved phase into the  
707 particulate phase, resulting in dissolved metal:macronutrient relationships that reflect their  
708 collective stoichiometry (Horner et al., 2021). Positive linear slopes result generally indicate the  
709 co-cycling of the metal and the macronutrient via uptake and remineralization, though slope  
710 values can vary widely by basin as they are a function of the metal:macronutrient uptake and  
711 remineralization stoichiometry of the native community and overall nutrient availability. Two-  
712 way linear regressions (see Methods) were used to investigate the relationships between dZn and  
713 dissolved silicate (dSi), dZn and dissolved phosphate (dP), and dCd and dP for the Amundsen  
714 Sea, Ross Sea, and TNB station groups (Fig. 12).

715



716  
 717 **Figure 12.** Relationships between (top row) total dissolved Zn and silicate (dSi) (top row), total  
 718 dissolved Zn and phosphate (dPO<sub>4</sub><sup>3-</sup>) (middle row), and total dissolved Cd and dPO<sub>4</sub><sup>3-</sup> (bottom  
 719 row) for surface (blue squares) and deep ocean (orange circles) arranged by station group  
 720 (Amundsen Sea, Ross Sea, and Terra Nova Bay). Depth thresholds were manually chosen to  
 721 optimize the linear fit of the surface and deep ocean trends. Regressions with an  $R^2 \geq 0.50$  are  
 722 shown as a solid line, and those with an  $R^2 < 0.50$  are shown as a dotted line. See Table S4 for  
 723 stoichiometric parameters and values. Regression outliers are marked with an 'x'. Data originally  
 724 plotted in Chmiel et al. 2023 and reprised here for ease of comparison with dZn:Si data.  
 725

726 The dZn:dP and dCd:dP relationships from this expedition were originally presented in  
 727 Chmiel et al. 2023 for comparison to dCo:dP, while they are included in the present study for  
 728 ease of comparison with dZn:dSi relationships presented for the first time. For these analyses, the  
 729 depth threshold that separates the surface and deep ocean was manually defined in order to

730 optimize the linear fit of the surface versus deep trends— this threshold depth can be thought of  
731 as an inflection point that represents the largest change in trace metal concentration with respect  
732 to dP or dSi concentration (Chmiel et al., 2023).

733         The near-linear global dZn:dSi relationship (Bruland et al., 1978; Middag et al., 2019;  
734 Vance et al., 2017) has been posited to arise, in part, from faster drawdown of Zn and Si relative  
735 to  $\text{dPO}_4^{3-}$  into Southern Ocean diatoms that leaves surface waters Zn and Si depleted (Vance et  
736 al., 2017). We observed distinct differences in dissolved dZn:dSi ecological stoichiometries  
737 comparing Amundsen Sea, Ross Sea and Terra Nova Bay station groups (Fig. 12; Table S4). A  
738 positive linear dZn:dSi relationship with a steep ( $m = 0.23 \pm 0.05$ ; Table S4) slope observed in  
739 the upper ocean of the Amundsen Sea contrasted starkly with the shallow slopes observed in the  
740 upper ocean of the Ross Sea and Terra Nova Bay. A bloom of non-silicifying *Phaeocystis*  
741 *antarctica* was present during our passage through the Amundsen Sea, consistent with abundant  
742 silicic acid yet rapid drawdown of Zn, which is known to be used by this organism (Saito and  
743 Goepfert, 2008). In contrast, the shallow slopes in the Ross Sea and Terra Nova Bay resulted  
744 from the persistence of dSi concentrations  $\geq 30 \mu\text{M}$  in the upper 30 m, while dZn was reduced to  
745 sub-nanomolar concentrations (average dZn =  $0.87 \pm 0.42 \text{ nM}$  in TNB at 10 m depth,  $n = 11$ ),  
746 highlighting the intense drawdown of dZn by biota in this region to meet a high metabolic dZn  
747 demand.

748         Similar trends were observed for dZn:dP and dCd:P, which exhibited shallow slopes  
749 within the upper ocean of the Ross Sea and Terra Nova Bay. Southern Ocean diatoms are known  
750 to have Zn:P uptake ratios that are up to an order of magnitude greater than the average for  
751 oceanic phytoplankton (Sieber et al., 2020; Twining and Baines, 2013; Vance et al., 2017). The  
752 increased presence of diatoms (as indicated by higher fucoxanthin concentrations) at the late

753 stations within Terra Nova Bay therefore likely exacerbated the surface decoupling of dZn and  
754 dP due to their high dZn demand. The maximum uptake rates of 158, 115, and 89 pmol Zn L<sup>-1</sup> d<sup>-1</sup>  
755 measured in this study for the Amundsen Sea, Ross Sea, and Terra Nova Bay groups,  
756 respectively, contextualize the high Zn uptake rates hypothesized to contribute to the high dZn:  
757 dP uptake ratios observed in Southern Ocean diatoms. These rates are indicative of total potential  
758 biological uptake, likely influenced by a depleted labile Zn pool and residual of complexed Zn,  
759 that then results in low dZn:dP ratios in shallow waters.

760 Like dZn and dSi, dCd and dP concentrations are known to share strong correlations in  
761 both deep and surface seawater (de Baar et al., 1994; Boyle, 1988; Boyle et al., 1976), with the  
762 vertical distribution of Cd controlled by phytoplankton uptake in surface waters and sinking of  
763 particulate organic matter and subsequent remineralization at depth. Observations of enhanced  
764 Cd uptake within the Fe-limited Southern Ocean (Cullen, 2006) are consistent with observations  
765 of increased Cd uptake by marine algal species under Fe limitation in both the field (Baars et al.,  
766 2014; Cullen et al., 2003; Cullen and Sherrell, 2005) and in culture (Lane et al., 2009; Sunda and  
767 Huntsman, 2000), thought to be due to the increased use of Cd in biochemical processes or  
768 inadvertent uptake due to the upregulation of metal transporters (Cullen, 2006; Sunda and  
769 Huntsman, 2000). In these coastal regions, dCd:dP had the same regional and depth trends as  
770 dZn:dP, further demonstrating their close biogeochemical association.

## 771 **5 Conclusions**

772 We have quantified the movement of the trace metals Zn and Cd from the dissolved to  
773 the particulate phase within the phytoplankton >3 μm size fraction collected in the Amundsen  
774 Sea, Ross Sea, and Terra Nova Bay of the Southern Ocean during austral summer 2017-2018.  
775 Our study adapts the stable Cd isotope tracer uptake method (Cox et al., 2014) to the



776 measurement of Zn uptake, and represents the first time series measurements of dZn and Zn  
777 uptake in a coastal environment during a bloom event. We have found that these high observed  
778 uptake rates were sufficient to draw down the otherwise abundant dZn inventory on seasonal  
779 timescales. Our analysis suggests that enhanced total biomass and low pCO<sub>2</sub> act as primary  
780 influences on  $\rho$ Zn, and that high  $\rho$ Zn results in dynamic changes in the cycling of both Cd and  
781 Co as a consequence of high Zn demand. Low dZn:Si slope values observed in the Ross Sea and  
782 Terra Nova Bay further highlighted the intense drawdown of dZn by biota to meet a high  
783 metabolic dZn demand. Overall, our study demonstrates that Zn demand is high and rapid  
784 enough to depress the inventory of Zn available to phytoplankton, suggesting that Zn has been  
785 overlooked as a dynamic limiting micronutrient in primary productivity modeling.

#### 786 *Data availability*

787       CICLOPS (NBP18-01) CTD hydrography data (including pressure, temperature, total  
788 dissolved oxygen, conductivity, fluorescence, and beam transmission; [https://www.bco-](https://www.bco-dmo.org/dataset-deployment/783917)  
789 [dmo.org/dataset-deployment/783917](https://www.bco-dmo.org/dataset-deployment/783917)) in addition to total dissolved metal, Zn and Cd uptake rate,  
790 macronutrient, and pigment datasets are available through the NSF Biological and Chemical  
791 Oceanography Data Management Office (BCO-DMO) repository ([https://www.bco-](https://www.bco-dmo.org/deployment/778919)  
792 [dmo.org/deployment/778919](https://www.bco-dmo.org/deployment/778919)). Underway pCO<sub>2</sub> data collected during cruise NBP1801 is  
793 available through R2R, <https://doi.org/10.7284/139318>.

#### 794 *Author contributions*

795       Conceptualization and analysis of the study was carried out by RMK and MAS. This work  
796 was supervised by MAS and GRD. Funding was acquired by MAS and GRD. RJC and DR  
797 contributed dCo data and discussion. DMM, MRM, NLS, IS, and RBD aided in sampling and data  
798 collection. TJH contributed to analysis and discussion. RMK and MAS wrote the manuscript with  
799 review and editing contributions from all co-authors.

800 *Competing interests*

801 The authors declare that they have no conflict of interest.

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818 **References**

819 Arrigo, K. R., van Dijken, G. L., and Bushinsky, S.: Primary production in the Southern Ocean,  
820 1997–2006, *Journal of Geophysical Research*, 113, C08004,  
821 <https://doi.org/10.1029/2007JC004551>, 2008.

822 Arrigo, K. R., Lowry, K. E., and van Dijken, G. L.: Annual changes in sea ice and phytoplankton  
823 in polynyas of the Amundsen Sea, Antarctica, *Deep Sea Research Part II: Topical Studies in*  
824 *Oceanography*, 71–76, 5–15, <https://doi.org/10.1016/j.dsr2.2012.03.006>, 2012.

- 825 de Baar, H. J. W., Saager, P. M., Nolting, R. F., and van der Meer, J.: Cadmium versus  
826 phosphate in the world ocean, *Marine Chemistry*, [https://doi.org/10.1016/0304-4203\(94\)90082-](https://doi.org/10.1016/0304-4203(94)90082-5)  
827 5, 1994.
- 828 de Baar, H. J. W., van Heuven, S. M. A. C., Abouchami, W., Xue, Z., Galer, S. J. G.,  
829 Rehkämper, M., Middag, R., and van Ooijen, J.: Interactions of dissolved CO<sub>2</sub> with cadmium  
830 isotopes in the Southern Ocean, *Marine Chemistry*, 195, 105–121,  
831 <https://doi.org/10.1016/J.MARCHEM.2017.06.010>, 2017.
- 832 Baars, O. and Croot, P. L.: The speciation of dissolved zinc in the Atlantic sector of the Southern  
833 Ocean, *Deep Sea Research Part II: Topical Studies in Oceanography*, 58, 2720–2732,  
834 <https://doi.org/10.1016/j.dsr2.2011.02.003>, 2011.
- 835 Baars, O., Abouchami, W., Galer, S. J. G., Boye, M., and Croot, P. L.: Dissolved cadmium in the  
836 Southern Ocean: Distribution, speciation, and relation to phosphate, *Limnology and*  
837 *Oceanography*, 59, 385–399, <https://doi.org/10.4319/LO.2014.59.2.0385>, 2014.
- 838 Bertrand, E. M., Saito, M. A., Rose, J. M., Riesselman, C. R., Lohan, M. C., Noble, A. E., Lee,  
839 P. A., and DiTullio, G. R.: Vitamin B 12 and iron colimitation of phytoplankton growth in the  
840 Ross Sea, *Limnology and Oceanography*, 52, 1079–1093,  
841 <https://doi.org/10.4319/lo.2007.52.3.1079>, 2007.
- 842 Biller, D. V. and Bruland, K. W.: Analysis of Mn, Fe, Co, Ni, Cu, Zn, Cd, and Pb in seawater  
843 using the Nobias-chelate PA1 resin and magnetic sector inductively coupled plasma mass  
844 spectrometry (ICP-MS), *Marine Chemistry*, <https://doi.org/10.1016/j.marchem.2011.12.001>,  
845 2012.
- 846 Bishop, J. K. B. and Wood, T. J.: Year-round observations of carbon biomass and flux variability  
847 in the Southern Ocean, *Global Biogeochemical Cycles*, <https://doi.org/10.1029/2008GB003206>,  
848 2009.
- 849 Bown, J., Laan, P., Ossebaar, S., Bakker, K., Rozema, P., and de Baar, H. J. W.: Bioactive trace  
850 metal time series during Austral summer in Ryder Bay, Western Antarctic Peninsula, *Deep-Sea*  
851 *Research Part II: Topical Studies in Oceanography*, <https://doi.org/10.1016/j.dsr2.2016.07.004>,  
852 2017.
- 853 Boyle, E. A.: Cadmium: Chemical tracer of deepwater paleoceanography, *Paleoceanography*,  
854 <https://doi.org/10.1029/PA003i004p00471>, 1988.
- 855 Boyle, E. A., Sclater, F., and Edmond, J. M.: On the marine geochemistry of cadmium, *Nature*,  
856 <https://doi.org/10.1038/263042a0>, 1976.
- 857 Brand, L. E., Sunda, W. G., and Guillard, R. R. L.: Reduction of marine phytoplankton  
858 reproduction rates by copper and cadmium, *Journal of Experimental Marine Biology and*  
859 *Ecology*, [https://doi.org/10.1016/0022-0981\(86\)90205-4](https://doi.org/10.1016/0022-0981(86)90205-4), 1986.
- 860 Bruland, K. W., Knauer, G. A., and Martin, J. H.: Zinc in north-east Pacific water, *Nature*, 271,  
861 741–743, <https://doi.org/10.1038/271741a0>, 1978.

862 Chmiel, R. J., Kell, R. M., Rao, D., Moran, D. M., DiTullio, G. R., and Saito, M. A.: Low cobalt  
863 inventories in the Amundsen and Ross seas driven by high demand for labile cobalt uptake  
864 among native phytoplankton communities, *Biogeosciences*, 20, 3997–4027,  
865 <https://doi.org/10.5194/bg-20-3997-2023>, 2023.

866 Cox, A. D., Noble, A. E., and Saito, M. A.: Cadmium enriched stable isotope uptake and  
867 addition experiments with natural phytoplankton assemblages in the Costa Rica Upwelling  
868 Dome, *Marine Chemistry*, <https://doi.org/10.1016/j.marchem.2014.09.009>, 2014.

869 Crameri, F.: Scientific colour maps, , <https://doi.org/10.5281/ZENODO.1243862>, 2023.

870 Cullen, J. T.: On the nonlinear relationship between dissolved cadmium and phosphate in the  
871 modern global ocean: Could chronic iron limitation of phytoplankton growth cause the kink?,  
872 *Limnology and Oceanography*, 51, 1369–1380, <https://doi.org/10.4319/lo.2006.51.3.1369>, 2006.

873 Cullen, J. T. and Sherrell, R. M.: Effects of dissolved carbon dioxide, zinc, and manganese on  
874 the cadmium to phosphorus ratio in natural phytoplankton assemblages, *Limnology and*  
875 *Oceanography*, 50, 1193–1204, <https://doi.org/10.4319/lo.2005.50.4.1193>, 2005.

876 Cullen, J. T., Lane, T. W., Morel, F. M. M., and Sheerell, R. M.: Modulation of cadmium uptake  
877 in phytoplankton by seawater CO<sub>2</sub> concentration, *Nature*, <https://doi.org/10.1038/46007>, 1999.

878 Cullen, J. T., Chase, Z., Coale, K. H., Fitzwater, S. E., and Sherrell, R. M.: Effect of iron  
879 limitation on the cadmium to phosphorus ratio of natural phytoplankton assemblages from the  
880 Southern Ocean, *Limnology and Oceanography*, <https://doi.org/10.4319/lo.2003.48.3.1079>,  
881 2003.

882 Cutter, G. A. and Bruland, K. W.: Rapid and noncontaminating sampling system for trace  
883 elements in global ocean surveys, *Limnology and Oceanography: Methods*,  
884 <https://doi.org/10.4319/lom.2012.10.425>, 2012.

885 Das, P., Samantaray, S., and Rout, G. R.: Studies on cadmium toxicity in plants: A review,  
886 [https://doi.org/10.1016/S0269-7491\(97\)00110-3](https://doi.org/10.1016/S0269-7491(97)00110-3), 1997.

887 DiTullio, G. R. and Smith, W. O.: Relationship between dimethylsulfide and phytoplankton  
888 pigment concentrations in the Ross Sea, Antarctica, *Deep-Sea Research Part I*,  
889 [https://doi.org/10.1016/0967-0637\(95\)00051-7](https://doi.org/10.1016/0967-0637(95)00051-7), 1995.

890 DiTullio, G. R., Geesey, M. E., Leventer, A., and Lizotte, M. P.: Algal pigment ratios in the Ross  
891 Sea: Implications for Chemtax analysis of Southern Ocean data, 35–51,  
892 <https://doi.org/10.1029/078ARS03>, 2003.

893 DiTullio, G. R., Garcia, N., Riseman, S. F., and Sedwick, P. N.: Effects of iron concentration on  
894 pigment composition in *Phaeocystis antarctica* grown at low irradiance, *Biogeochemistry*, 83,  
895 71–81, <https://doi.org/10.1007/s10533-007-9080-8>, 2007.

896 Fitzwater, S. E., Johnson, K. S., Gordon, R. M., Coale, K. H., and Smith, W. O.: Trace metal  
897 concentrations in the Ross Sea and their relationship with nutrients and phytoplankton growth,

- 898 Deep-Sea Research Part II: Topical Studies in Oceanography, 47, 3159–3179,  
899 [https://doi.org/10.1016/S0967-0645\(00\)00063-1](https://doi.org/10.1016/S0967-0645(00)00063-1), 2000.
- 900 Gerringa, L. J. A., Alderkamp, A.-C., van Dijken, G., Laan, P., Middag, R., and Arrigo, K. R.:  
901 Dissolved Trace Metals in the Ross Sea, *Frontiers in Marine Science*, 7,  
902 <https://doi.org/10.3389/fmars.2020.577098>, 2020.
- 903 Giordano, R., Lombardi, G., Ciaralli, L., Beccaloni, E., Sepe, A., Ciprotti, M., and Costantini, S.:  
904 Major and trace elements in sediments from Terra Nova Bay, Antarctica, *Science of The Total*  
905 *Environment*, 227, 29–40, [https://doi.org/10.1016/S0048-9697\(98\)00402-1](https://doi.org/10.1016/S0048-9697(98)00402-1), 1999.
- 906 Haas, C. E., Rodionov, D. A., Kropat, J., Malasarn, D., Merchant, S. S., and de Crécy-Lagard,  
907 V.: A subset of the diverse COG0523 family of putative metal chaperones is linked to zinc  
908 homeostasis in all kingdoms of life, *BMC Genomics*, 10, 470, [https://doi.org/10.1186/1471-](https://doi.org/10.1186/1471-2164-10-470)  
909 [2164-10-470](https://doi.org/10.1186/1471-2164-10-470), 2009.
- 910 Hopwood, M. J., Carroll, D., Höfer, J., Achterberg, E. P., Meire, L., Le Moigne, F. A. C., Bach,  
911 L. T., Eich, C., Sutherland, D. A., and González, H. E.: Highly variable iron content modulates  
912 iceberg-ocean fertilisation and potential carbon export, *Nature Communications* 2019 10:1, 10,  
913 1–10, <https://doi.org/10.1038/s41467-019-13231-0>, 2019.
- 914 Horner, T. J., Lee, R. B. Y., Henderson, G. M., and Rickaby, R. E. M.: Nonspecific uptake and  
915 homeostasis drive the oceanic cadmium cycle, *Proceedings of the National Academy of*  
916 *Sciences*, 110, 2500–2505, <https://doi.org/10.1073/pnas.1213857110>, 2013.
- 917 Horner, T. J., Little, S. H., Conway, T. M., Farmer, J. R., Hertzberg, J. E., Janssen, D. J., Lough,  
918 A. J. M., McKay, J. L., Tessin, A., Galer, S. J. G., Jaccard, S. L., Lacan, F., Paytan, A., Wuttig,  
919 K., and GEOTRACES–PAGES Biological Productivity Working Group Members: Bioactive  
920 Trace Metals and Their Isotopes as Paleoproductivity Proxies: An Assessment Using  
921 GEOTRACES-Era Data, *Global Biogeochemical Cycles*, 35, e2020GB006814,  
922 <https://doi.org/10.1029/2020GB006814>, 2021.
- 923 Hutchins, D. and Bruland, K.: Grazer-mediated regeneration and assimilation of Fe, Zn and Mn  
924 from planktonic prey, *Mar. Ecol. Prog. Ser.*, 110, 259–269, <https://doi.org/10.3354/meps110259>,  
925 1994.
- 926 Hutchins, D. A. and Bruland, K. W.: Fe, Zn, Mn and N transfer between size classes in a coastal  
927 phytoplankton community: Trace metal and major nutrient recycling compared, *issn: 0022-2402*,  
928 53, 297–313, <https://doi.org/10.1357/0022240953213197>, 1995.
- 929 Hutchins, D. A., Wang, W. X., Schmidt, M. A., and Fisher, N. S.: Dual-labeling techniques for  
930 trace metal biogeochemical investigations in aquatic plankton communities, *Aquatic Microbial*  
931 *Ecology*, <https://doi.org/10.3354/ame019129>, 1999.
- 932 Irving, B. H. and Williams, R. J. P.: The Stability of Transition-metal Complexes, *Journal of the*  
933 *Chemical Society (Resumed)*, <https://doi.org/10.1039/JR9530003192>, 1953.

934 Jackson, S. L., Spence, J., Janssen, D. J., Ross, A. R. S., and Cullen, J. T.: Determination of Mn,  
935 Fe, Ni, Cu, Zn, Cd and Pb in seawater using offline extraction and triple quadrupole ICP-  
936 MS/MS, *Journal of Analytical Atomic Spectrometry*, 33, 304–313,  
937 <https://doi.org/10.1039/c7ja00237h>, 2018.

938 Kato, T., Nakamura, S., and Morita, M.: Determination of Nickel, Copper, Zinc, Silver,  
939 Cadmium and Lead in Seawater by Isotope Dilution Inductively Coupled Plasma Mass  
940 Spectrometry, *ANAL. SCI.*, 6, 623–626, <https://doi.org/10.2116/analsci.6.623>, 1990.

941 Kell, R. M., Subhas, A. V., Schanke, N. L., Lees, L. E., Chmiel, R. J., Rao, D., Brisbin, M. M.  
942 M., Moran, D. M., McIlvin, M. R., Bolinesi, F., Mangoni, O., Casotti, R., Balestra, C., Horner,  
943 T., Dunbar, R. B., Allen, A. E., DiTullio, G. R., and Saito, M. A.: Zinc stimulation of  
944 phytoplankton in a low carbon dioxide, coastal Antarctic environment,  
945 <https://doi.org/10.1101/2023.11.05.565706>, 5 November 2023.

946 Kellogg, R. M., Moosburner, M. A., Cohen, N. R., Hawco, N. J., McIlvin, M. R., Moran, D. M.,  
947 DiTullio, G. R., Subhas, A. V., Allen, A. E., and Saito, M. A.: Adaptive responses of marine  
948 diatoms to zinc scarcity and ecological implications, *Nature Communications* 2022 13:1, 13, 1–  
949 13, <https://doi.org/10.1038/s41467-022-29603-y>, 2022.

950 Lane, E. S., Semeniuk, D. M., Strzepek, R. F., Cullen, J. T., and Maldonado, M. T.: Effects of  
951 iron limitation on intracellular cadmium of cultured phytoplankton: Implications for surface  
952 dissolved cadmium to phosphate ratios, *Marine Chemistry*, 115, 155–162,  
953 <https://doi.org/10.1016/J.MARCHEM.2009.07.008>, 2009.

954 Lane, T. W., Saito, M. A., George, G. N., Pickering, I. J., Prince, R. C., and Morel, F. M. M.: A  
955 cadmium enzyme from a marine diatom, *Nature*, 435, 42–42, <https://doi.org/10.1038/435042a>,  
956 2005.

957 Latour, P., Wuttig, K., van der Merwe, P., Strzepek, R. F., Gault-Ringold, M., Townsend, A. T.,  
958 Holmes, T. M., Corkill, M., and Bowie, A. R.: Manganese biogeochemistry in the Southern  
959 Ocean, from Tasmania to Antarctica, *Limnology and Oceanography*, 66, 2547–2562,  
960 <https://doi.org/10.1002/lno.11772>, 2021.

961 Lee, J. and Morel, F.: Replacement of zinc by cadmium in marine phytoplankton, *Marine*  
962 *Ecology Progress Series*, 127, 305–309, <https://doi.org/10.3354/meps127305>, 1995.

963 Lee, J. G., Roberts, S. B., and Morel, F. M. M.: Cadmium: A nutrient for the marine diatom  
964 *Thalassiosira weissflogii*, *Limnology and Oceanography*,  
965 <https://doi.org/10.4319/lo.1995.40.6.1056>, 1995.

966 Lohan, M. C., Statham, P. J., and Crawford, D. W.: Total dissolved zinc in the upper water  
967 column of the subarctic North East Pacific, *Deep Sea Research Part II: Topical Studies in*  
968 *Oceanography*, 49, 5793–5808, [https://doi.org/10.1016/S0967-0645\(02\)00215-1](https://doi.org/10.1016/S0967-0645(02)00215-1), 2002.

969 Martin, J. H.: Glacial-interglacial CO<sub>2</sub> change: The Iron Hypothesis, *Paleoceanography*,  
970 <https://doi.org/10.1029/PA005i001p00001>, 1990.

- 971 Middag, R., Baar, H. J. W., and Bruland, K. W.: The relationships between dissolved zinc and  
972 major nutrients phosphate and silicate along the GEOTRACES GA02 transect in the West  
973 Atlantic Ocean, *Global Biogeochemical Cycles*, 33, 63–84,  
974 <https://doi.org/10.1029/2018GB006034>, 2019.
- 975 Morel, F. M. M., Reinfelder, J. R., Roberts, S. B., Chamberlain, C. P., Lee, J. G., and Yee, D.:  
976 Zinc and carbon co-limitation of marine phytoplankton, *Nature*, 369, 740–742,  
977 <https://doi.org/10.1038/369740A0>, 1994.
- 978 Morel, F. M. M., Milligan, A. J., and Saito, M. A.: Marine Bioinorganic Chemistry: The Role of  
979 Trace Metals in the Oceanic Cycles of Major Nutrients, in: *Treatise on Geochemistry: Second*  
980 *Edition*, <https://doi.org/10.1016/B978-0-08-095975-7.00605-7>, 2013.
- 981 Morel, F. M. M., Lam, P. J., and Saito, M. A.: Trace metal substitution in marine phytoplankton,  
982 *Annual Review of Earth and Planetary Sciences*, 48, 491–517, [https://doi.org/10.1146/annurev-](https://doi.org/10.1146/annurev-earth-053018-060108)  
983 [earth-053018-060108](https://doi.org/10.1146/annurev-earth-053018-060108), 2020.
- 984 Noble, A. E., Lamborg, C. H., Ohnemus, D. C., Lam, P. J., Goepfert, T. J., Measures, C. I.,  
985 Frame, C. H., Casciotti, K. L., DiTullio, G. R., Jennings, J., and Saito, M. A.: Basin-scale inputs  
986 of cobalt, iron, and manganese from the Benguela-Angola front to the South Atlantic Ocean,  
987 *Limnology and Oceanography*, <https://doi.org/10.4319/lo.2012.57.4.0989>, 2012.
- 988 Noble, A. E., Moran, D. M., Allen, A. E., and Saito, M. A.: Dissolved and particulate trace metal  
989 micronutrients under the McMurdo Sound seasonal sea ice: basal sea ice communities as a  
990 capacitor for iron, *Frontiers in Chemistry*, 1, <https://doi.org/10.3389/fchem.2013.00025>, 2013.
- 991 Ohnemus, D. C., Rauschenberg, S., Cutter, G. A., Fitzsimmons, J. N., Sherrell, R. M., and  
992 Twining, B. S.: Elevated trace metal content of prokaryotic communities associated with marine  
993 oxygen deficient zones, *Limnology and Oceanography*, <https://doi.org/10.1002/lno.10363>, 2017.
- 994 Oldham, V. E., Chmiel, R., Hansel, C. M., DiTullio, G. R., Rao, D., and Saito, M.: Inhibited  
995 Manganese Oxide Formation Hinders Cobalt Scavenging in the Ross Sea, *Global*  
996 *Biogeochemical Cycles*, 35, <https://doi.org/10.1029/2020GB006706>, 2021.
- 997 Park, H., Song, B., and Morel, F. M. M.: Diversity of the cadmium-containing carbonic  
998 anhydrase in marine diatoms and natural waters, *Environmental Microbiology*,  
999 <https://doi.org/10.1111/j.1462-2920.2006.01151.x>, 2007.
- 1000 Park, H., McGinn, P. J., and Morel, F. M. M.: Expression of cadmium carbonic anhydrase of  
1001 diatoms in seawater, *Aquatic Microbial Ecology*, <https://doi.org/10.3354/ame01192>, 2008.
- 1002 Person, R., Vancoppenolle, M., Aumont, O., and Malsang, M.: Continental and Sea Ice Iron  
1003 Sources Fertilize the Southern Ocean in Synergy, *Geophysical Research Letters*, 48,  
1004 [e2021GL094761](https://doi.org/10.1029/2021GL094761), <https://doi.org/10.1029/2021GL094761>, 2021.
- 1005 Planquette, H., Sherrell, R. M., Stammerjohn, S., and Field, M. P.: Particulate iron delivery to the  
1006 water column of the Amundsen Sea, Antarctica, *Marine Chemistry*, 153, 15–30,  
1007 <https://doi.org/10.1016/j.marchem.2013.04.006>, 2013.

- 1008 Price, N. M. and Morel, F. M. M.: Cadmium and cobalt substitution for zinc in a marine diatom,  
1009 Nature, 344, 658–660, <https://doi.org/10.1038/344658a0>, 1990.
- 1010 Rapp, I., Schlosser, C., Rusiecka, D., Gledhill, M., and Achterberg, E. P.: Automated  
1011 preconcentration of Fe, Zn, Cu, Ni, Cd, Pb, Co, and Mn in seawater with analysis using high-  
1012 resolution sector field inductively-coupled plasma mass spectrometry, *Analytica Chimica Acta*,  
1013 <https://doi.org/10.1016/j.aca.2017.05.008>, 2017.
- 1014 Rudge, J. F., Reynolds, B. C., and Bourdon, B.: The double spike toolbox, *Chemical Geology*,  
1015 <https://doi.org/10.1016/j.chemgeo.2009.05.010>, 2009.
- 1016 Saito, M. A. and Goepfert, T. J.: Zinc-cobalt colimitation of *Phaeocystis antarctica*, *Limnology*  
1017 and *Oceanography*, 53, 266–275, <https://doi.org/10.4319/lo.2008.53.1.0266>, 2008.
- 1018 Sedwick, P. N., Di Tullio, G. R., and Mackey, D. J.: Iron and manganese in the Ross Sea,  
1019 Seasonal iron limitation in Antarctic, *Journal of Geophysical Research: Oceans*,  
1020 <https://doi.org/10.1029/2000JC000256>, 2000.
- 1021 Sedwick, P. N., Marsay, C. M., Sohst, B. M., Aguilar-Islas, A. M., Lohan, M. C., Long, M. C.,  
1022 Arrigo, K. R., Dunbar, R. B., Saito, M. A., Smith, W. O., and DiTullio, G. R.: Early season  
1023 depletion of dissolved iron in the Ross Sea polynya: Implications for iron dynamics on the  
1024 Antarctic continental shelf, *Journal of Geophysical Research*, 116, C12019,  
1025 <https://doi.org/10.1029/2010JC006553>, 2011.
- 1026 Shaked, Y., Xu, Y., Leblanc, K., and Morel, F. M. M.: Zinc availability and alkaline phosphatase  
1027 activity in *Emiliania huxleyi*: Implications for Zn-P co-limitation in the ocean, *Limnology and*  
1028 *Oceanography*, 51, 299–309, <https://doi.org/10.4319/lo.2006.51.1.0299>, 2006.
- 1029 Sieber, M., Conway, T. M., de Souza, G. F., Hassler, C. S., Ellwood, M. J., and Vance, D.:  
1030 Cycling of zinc and its isotopes across multiple zones of the Southern Ocean: Insights from the  
1031 Antarctic Circumnavigation Expedition, *Geochimica et Cosmochimica Acta*,  
1032 <https://doi.org/10.1016/j.gca.2019.09.039>, 2020.
- 1033 Sohrin, Y., Urushihara, S., Nakatsuka, S., Kono, T., Higo, E., Minami, T., Norisuye, K., and  
1034 Umetani, S.: Multielemental determination of GEOTRACES key trace metals in seawater by  
1035 ICPMS after preconcentration using an ethylenediaminetriacetic acid chelating resin, *Analytical*  
1036 *Chemistry*, <https://doi.org/10.1021/ac800500f>, 2008.
- 1037 St-Laurent, P., Yager, P. L., Sherrell, R. M., Stammerjohn, S. E., and Dinniman, M. S.: Pathways  
1038 and supply of dissolved iron in the Amundsen Sea (Antarctica), *Journal of Geophysical*  
1039 *Research: Oceans*, 122, 7135–7162, <https://doi.org/10.1002/2017JC013162>, 2017.
- 1040 Sunda, W. G. and Huntsman, S. A.: Feedback interactions between zinc and phytoplankton in  
1041 seawater, *Limnology and Oceanography*, 37, 25–40, <https://doi.org/10.4319/lo.1992.37.1.0025>,  
1042 1992.



- 1043 Sunda, W. G. and Huntsman, S. A.: Cobalt and zinc interreplacement in marine phytoplankton:  
1044 Biological and geochemical implications, *Limnology and Oceanography*, 40, 1404–1417,  
1045 <https://doi.org/10.4319/lo.1995.40.8.1404>, 1995.
- 1046 Sunda, W. G. and Huntsman, S. A.: Antagonisms between cadmium and zinc toxicity and  
1047 manganese limitation in a coastal diatom, *Limnology and Oceanography*, 41, 373–387,  
1048 <https://doi.org/10.4319/lo.1996.41.3.0373>, 1996.
- 1049 Sunda, W. G. and Huntsman, S. A.: Control of Cd Concentrations in a Coastal Diatom by  
1050 Interactions among Free Ionic Cd, Zn, and Mn in Seawater, *Environmental Science &*  
1051 *Technology*, 32, 2961–2968, <https://doi.org/10.1021/es980271y>, 1998a.
- 1052 Sunda, W. G. and Huntsman, S. A.: Processes regulating cellular metal accumulation and  
1053 physiological effects: Phytoplankton as model systems, *Science of The Total Environment*, 219,  
1054 165–181, [https://doi.org/10.1016/S0048-9697\(98\)00226-5](https://doi.org/10.1016/S0048-9697(98)00226-5), 1998b.
- 1055 Sunda, W. G. and Huntsman, S. A.: Effect of Zn, Mn, and Fe on Cd accumulation in  
1056 phytoplankton: Implications for oceanic Cd cycling, *Limnology and Oceanography*, 45, 1501–  
1057 1516, <https://doi.org/10.4319/lo.2000.45.7.1501>, 2000.
- 1058 Tan, D., Xu, W., Zhu, Z., Li, S., Wu, G., and Qin, H.: Optimizing the ratio of the spike to sample  
1059 for isotope dilution analysis: a case study with selenium isotopes, *Acta Geochimica*,  
1060 <https://doi.org/10.1007/s11631-019-00390-6>, 2020.
- 1061 Taylor, S. R. and McLennan, S. M.: *The Continental Crust: Its Composition and Evolution.*,  
1062 Blackwell Scientific Publications, Oxford, 312 pp., 1985.
- 1063 Twining, B. S. and Baines, S. B.: The trace metal composition of marine phytoplankton., *Annual*  
1064 *review of marine science*, 5, 191–215, <https://doi.org/10.1146/annurev-marine-121211-172322>,  
1065 2013.
- 1066 Vance, D., Little, S. H., De Souza, G. F., Khatiwala, S., Lohan, M. C., and Middag, R.: Silicon  
1067 and zinc biogeochemical cycles coupled through the Southern Ocean, *Nature Geoscience*, 10,  
1068 202–206, <https://doi.org/10.1038/NGEO2890>, 2017.
- 1069 Weber, T., John, S., Tagliabue, A., and DeVries, T.: Biological uptake and reversible scavenging  
1070 of zinc in the global ocean, *Science*, 361, 72–76, <https://doi.org/10.1126/SCIENCE.AAP8532>,  
1071 2018.
- 1072 Wright, S. W., van den Enden, R. L., Pearce, I., Davidson, A. T., Scott, F. J., and Westwood, K.  
1073 J.: Phytoplankton community structure and stocks in the Southern Ocean (30–80°E) determined  
1074 by CHEMTAX analysis of HPLC pigment signatures, *Deep-Sea Research Part II: Topical*  
1075 *Studies in Oceanography*, <https://doi.org/10.1016/j.dsr2.2009.06.015>, 2010.
- 1076 Wu, J. and Boyle, E. A.: Determination of iron in seawater by high-resolution isotope dilution  
1077 inductively coupled plasma mass spectrometry after Mg(OH)<sub>2</sub> coprecipitation, *Analytica*  
1078 *Chimica Acta*, 367, 183–191, [https://doi.org/10.1016/S0003-2670\(98\)00145-7](https://doi.org/10.1016/S0003-2670(98)00145-7), 1998.

- 1079 Wuttig, K., Townsend, A. T., van der Merwe, P., Gault-Ringold, M., Holmes, T., Schallenberg,  
1080 C., Latour, P., Tonnard, M., Rijkenberg, M. J. A., Lannuzel, D., and Bowie, A. R.: Critical  
1081 evaluation of a seaFAST system for the analysis of trace metals in marine samples, *Talanta*,  
1082 <https://doi.org/10.1016/j.talanta.2019.01.047>, 2019.
- 1083 Xu, Y., Tang, D., Shaked, Y., and Morel, F. M. M.: Zinc, cadmium, and cobalt interreplacement  
1084 and relative use efficiencies in the coccolithophore *Emiliana huxleyi*, *Limnology and*  
1085 *Oceanography*, 52, 2294–2305, <https://doi.org/10.4319/lo.2007.52.5.2294>, 2007.
- 1086 Zhao, Y., Vance, D., Abouchami, W., and de Baar, H. J. W.: Biogeochemical cycling of zinc and  
1087 its isotopes in the Southern Ocean, *Geochimica et Cosmochimica Acta*,  
1088 <https://doi.org/10.1016/j.gca.2013.07.045>, 2014.
- 1089