





23 gathered from 0-100 cm depth in terraces (containing different crops and different  
24 ecological restoration vegetation) and slopes.

25 **Results:** The results show that terracing can effectively increase soil organic carbon  
26 (SOC) content (7.7 g·kg<sup>-1</sup> in terraced cropland > 4.9 g·kg<sup>-1</sup> in sloping cropland).  
27 Changes in the organic carbon content of the terracing is mainly due to improvements  
28 in soil and water conservation capacity and agricultural activities, loss of soil organic  
29 carbon due to short-term abandonment and an increase in soil organic carbon due to  
30 replanting of fruit trees and crops. The choice of tree species in afforestation policies  
31 has also led to differences in soil organic carbon. *Pinus tabuliformis* Carr. has the  
32 highest SOC content (9.8 g·kg<sup>-1</sup>).

33 **Conclusions:** The SOC content in 0-100 cm of terraced fields planted with wheat was  
34 1.5 times higher than that of sloping fields planted with wheat. Compared with  
35 sloping land, terrace construction significantly increased the SOC content of  
36 cultivated land, especially in the top soil layer (0-30 cm), and converting some sloping  
37 land into terraces would enhance the carbon sequestration capacity. This study has  
38 significant implications for agricultural management and ecological restoration in the  
39 terraced areas of the Loess Plateau and contributes to the development of rational  
40 policies for carbon sequestration on arable land in terraced areas.

41 **Keywords:** Terracing; Soil Organic Carbon; Agricultural activities; Vegetation cover

## 42 **1 Introduction**

43 Soil is considered to be the second largest carbon reservoir next to the ocean



44 (Stockmann et al., 2013), and the soil carbon reservoir is crucial for the global carbon  
45 cycle (Lal, 2004; Houghton, 2007). SOC is is a key element of the global carbon cycle  
46 ( Rossel et al., 2019) and serves as a significant indicator for assessing soil quality and  
47 land productivity (Guillaume et al., 2021; Wang et al., 2012). To address climate  
48 change and lower CO<sub>2</sub> emissions for sustainable development, we should sequester  
49 more carbon in the soil rather than releasing it into the atmosphere. This is already an  
50 urgent and challenging issue for humanity (Bednar et al., 2021). Therefore,  
51 understanding the variation of SOC content in different regions, especially in  
52 anthropogenic landscapes, is of great importance to assess the carbon sequestration  
53 potential of soils and mitigate climate change. The sequestration of large amounts of  
54 CO<sub>2</sub> into soils can be achieved by the ways of land use and management practices  
55 policies (Smith, 2012). However, SOC reserves have been declining worldwide (Jones  
56 et al., 2005). Forest destruction caused by reclaiming farmland is the main reason for  
57 global SOC consumption to date, resulting in increasingly serious ecological damage  
58 (Lal, 2016). Hence, how to increase soil organic carbon to maintain soil quality,  
59 restore ecology, ensure food security and reduce CO<sub>2</sub> emissions is a matter of great  
60 concern in the context of global warming.

61       Agricultural terracing is a crucial landscape engineering measure to reduce soil  
62 erosion and maintain soil fertility and increase agricultural productivity (Doetterl et al.,  
63 2012; Zhu et al, 2021), which is one of the ways to achieve sustainable agricultural  
64 development. On the one hand, the conversion of terraces into slopes increases the  
65 cultivated area significantly. On the other hand, it helps prevent erosion problems



66 (Arnáez et al., 2015) and effectively increases food production (Tarolli et al., 2014).  
67 Terraces are widely distributed and have created environmental benefits in countries  
68 in East Asia, the Mediterranean, and Southeast Asia (Wei et al., 2016). Many studies  
69 have shown that terracing can intercept more than 80% of rainfall runoff and sediment,  
70 and horizontal terracing can retain all rainfall to replenish soil moisture. The positive  
71 benefits of carbon capture generated by terraces come from the collection of eroded  
72 material for sloping soils. However, the conversion of natural vegetation to cropland  
73 inevitably results in a reduction of biomass and therefore a significant loss of SOC  
74 (Aguilera et al., 2018 and 2013). During the construction of terraces period, it is  
75 inevitable stripping of topsoil and exposure of deep soil, and a large amount of new  
76 subsoil covers the surface of the terraces. This severe soil disturbance may alter soil  
77 organic carbon dynamics (Sidle et al., 2006), but the potential long-term benefits of  
78 terrace construction are considerable (Chen et al., 2017). However, many terraces are  
79 experiencing ridge damage and terrace collapse due to a lack of terrace maintenance  
80 or land abandonment, which not only leads to reduce soil and water conservation  
81 benefits but potentially increases erosion and carbon emissions (Arnáez et al., 2015;  
82 Wen et al., 2020).

83       Agricultural land accounts for more than 30% of the global area and has great  
84 potential for carbon sequestration and mitigation of global climate change (Sun et al.,  
85 2010). Agricultural soils can be improved by implementing some regulatory  
86 management practices to improve soil properties and further increase organic carbon  
87 content (Lal et al., 2011). However, studies from different regions have shown a



88 global trend of decreasing organic carbon content in agricultural soils nowadays  
89 (Bellamy et al. 2008; Heikkinen et al. 2013; Yli-Halla et al., 2018). How to increase  
90 the SOC content of farmland is the key to promoting sustainable agriculture and  
91 improving the carbon sink. In the Loess Plateau region, a large amount of arable land  
92 has been converted to woodland and grassland through the implementation of  
93 afforestation and reforestation policies, and these measures have increased the organic  
94 carbon content of the soil (Rong et al., 2021). In the implementation of ecological  
95 restoration, different vegetation types have different benefits in increasing soil carbon  
96 pools (Hong et al., 2020). The agroforestry cropping pattern with other crops stores  
97 more carbon than other traditional agricultural cropping patterns (Smith et al., 2022).  
98 Nair et al. (2009) cited seven studies on soil carbon in tropical agroforestry systems  
99 that show that this type of cropping certainly stores more carbon.

100 In the Loess Plateau, the erosion problem of sloping land has a great impact on  
101 agricultural production (Ran et al., 2020). Terracing has become the most important  
102 ways to solve the erosion problem on sloping lands. By the end of 2012, there were  
103 37,100 km<sup>2</sup> of terraced fields on the Loess Plateau. With the further development of  
104 terrace construction, changes in surface morphology and soil properties also lead to  
105 dynamic changes in soil carbon pools, and such changes will lead to changes in soil  
106 carbon storage capacity. Meanwhile, how the land use pattern of terraces and human  
107 activity factors such as abandonment and crop type affect the soil carbon pool of  
108 terraces still needs to be investigated. The implementation of afforestation policies in  
109 terraced areas will also lead to changes in soil carbon pools. Consequently, we

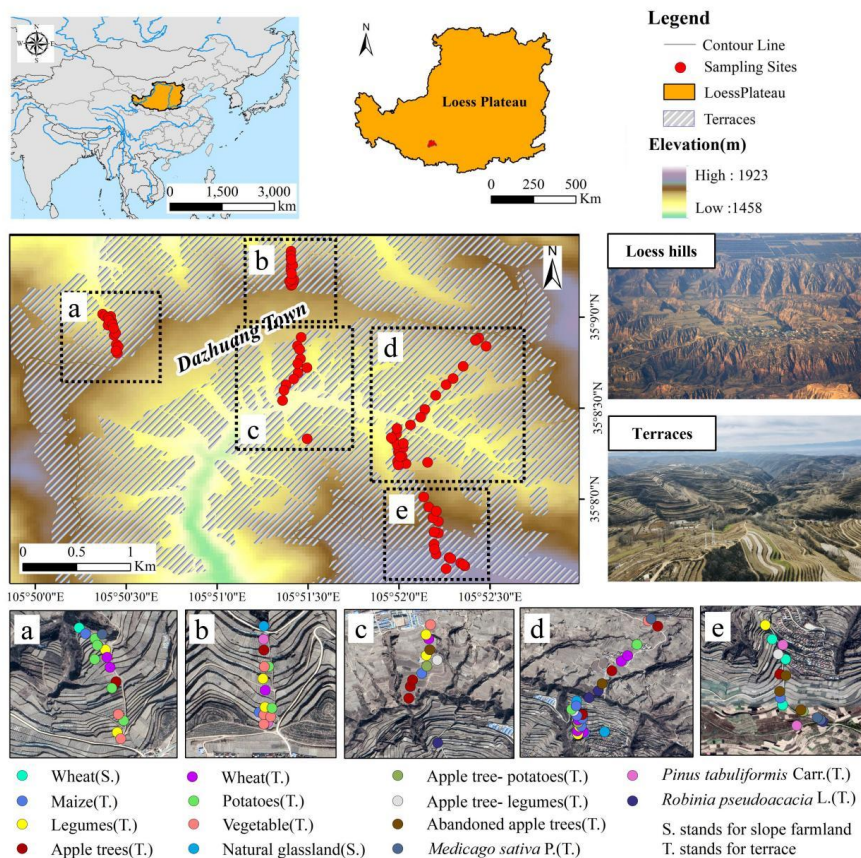


110 gathered soil samples from terraces and slopes, including terraces with varying land  
111 use and crop types, to investigate (1) the impacts of terrace construction on SOC in  
112 the Loess Plateau region and (2) the effects of different vegetation cover and tillage  
113 activities on the carbon sink capacity of terraces. This study will provide a reference  
114 for terrace management and is also a guide for soil carbon sequestration.

## 115 **2 Data and Methods**

### 116 **2.1 Study Area**

117 The study area is the typical terrace construction region of the Loess Plateau  
118 (Zhuanglang terracing) and the construction of this area began in the 1960s. By 2005,  
119 14790 km<sup>2</sup> of terraces have been built, accounting for 95.3% of the total arable land in  
120 the region. The structure of the terraces is mainly horizontal terraces. Zhuanglang  
121 terraces belong to the loess hilly terrain area with gullies and complex topography,  
122 and the elevation is between 1521m-1784m. The climate type is temperate continental,  
123 with warm, humid summers and cold, dry winters. More than 60% of precipitation  
124 occurs in summer and autumn (July - October), with an annual rainfall of 542mm and  
125 an average annual temperature of 7.5°C. The dominant soil type in this area is fine  
126 loessial soil, the natural vegetation is mainly herbaceous, shrubs, coniferous forests,  
127 and locust trees, and the crops are wheat, maize, potatoes, and apple trees. Due to the  
128 constraints of soil properties and irrigation water sources in this region, the growth of  
129 crops depends on natural rainfall.



130

131

Fig.1 Study area

## 132 2.2 Terraced soil sample collection

133 We carried out soil sample collection at the terrace observation system in  
 134 Yangpota mountain, Dazhuang Town, Zhuanglang County, with the sampling date  
 135 being October 2020. The construction of terraces in this area started in 1964 and  
 136 based on interviews with local farmers, it was confirmed that all the agricultural fields  
 137 in the area had been constructed in 1991, and all the terraces were constructed 30  
 138 years ago. The main terracing structure built in the region is the horizontal terrace,



139 which is an agricultural field with stepped sections along contours on the slopes of  
140 loess hills. The width of the terraces varies from 1.6 to 6 meters, and the height of the  
141 terrace steps varies from 0.3 to 0.8 meters. The terraced slopes are slightly  
142 counter-sloped to collect more precipitation and the slope ranges from 0% to 11%  
143 (Chen et al., 2020).

144 We randomly set up 84 sampling sites in the study area, with 77 terraced  
145 sampling sites and 7 slope sampling sites. Crop type, cropping pattern, and  
146 agricultural abandonment all affect the soil carbon pool of the terraces, so the terraces  
147 included sampling points for different cropping patterns of apple trees (9), number of  
148 sampling points), vegetable (9), wheat (9), legume (9), potato (9), maize (9), apple  
149 tree-legume (3) and apple- potato (3) (Appendix, Fig. A1). Five abandoned apple tree  
150 terraces were included in the terraces and the apple trees were not removed from the  
151 terraces and there was a large amount of weed growth. Three types of restored  
152 vegetation were planted on the terraces: *Robinia pseudoacacia* L. (4), *Pinus*  
153 *tabuliformis* Carr. (4) and *Medicago sativa* L. (4). The seven slope sites included four  
154 wheat plantations and three natural grassland sites (Table 1).

155 Measuring SOC concentrations in the surface layer of the soil (10 or 20 cm)  
156 alone does not imply soil changes due to tillage management, so we designed the  
157 sampling depth as 1 m. Samples were collected at 0-10 cm, 10-20 cm, 20-30 cm,  
158 30-40 cm, 40-50 cm, 50-60 cm, 60-70 cm, 70-80 cm, 80- 90cm, and 90-100cm  
159 samples. A (2 × 2) m<sup>2</sup> sample square was randomly delineated in each identified  
160 sample plot and sampled along the diagonal line. The soil samples collected three





161 times were mixed according to different soil layers after removing plant roots and  
162 debris, and 840 mixed soil samples were finally obtained.



163

164 Table1 Number of different types of sampling plots, vegetation types, number of

165

samples, and soil property data

Land types	Planting method	Vegetation types	Number of soil profiles	Soil depth(cm)	Soil texture fractions			Soil moisture(%)		SOC(g•kg <sup>-1</sup> )			
					Clay (%)	Silt (%)	Sand (%)	Mean	SD	Mean	SD		
Terrace	Single vegetation	Wheat	9	0-100	11.1	80.9	8.0	24.2	10.36	7.7	2.78		
		Apple trees	9	0-100	—	—	—	24.3	9.32	7.1	2.39		
		Potatoes	9	0-100	—	—	—	26.8	12.98	5.2	2.32		
		Legumes	9	0-100	—	—	—	23.2	15.21	4.5	2.10		
		Maize	9	0-100	—	—	—	21.4	10.35	7.7	2.48		
	Multiple vegetation	<i>Robinia pseudoacacia</i> L.		4	0-100	10.8	79.4	9.8	18.2	5.92	8.0	2.77	
			<i>Pinus tabulaeformis</i> Carr.		4	0-100	9.9	84.0	6.1	19.3	6.33	9.8	4.44
					4	0-100	10.9	80.0	9.1	20.4	13.62	5.6	1.57
		Vegetable		9	0-100	—	—	—	22.1	5.31	6.5	1.80	
			Apple tree-legumes		3	0-100	—	—	—	20.2	7.36	5.4	1.65
				Apple tree-potatoes	3	0-100	—	—	—	22.7	10.68	6.7	1.77
		Sloping land	Single vegetation	Wheat	4	0-100	10.0	78.0	12.0	21.3	13.44	4.9	1.07
				Grassland	3	0-100	10.4	80.1	9.5	21.0	9.62	6.0	2.26
		Abandoned terraces	Multiple vegetation	Apple trees and weeds	5	0-100	10.6	78.5	10.8	23.7	11.35	6.4	1.85

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Note: “—” represents not measured. All soil attribute data values are average values, average soil texture, soil moisture, and SOC of the 0-100 cm profiles, derived from 10 samples for each profile (10-cm depth intervals)



## 169 **2.3 Experimental analysis and data statistics**

170 The collected soil samples were placed in sealed plastic bags and pre-weighed  
171 aluminum boxes. The soil samples in the aluminum box were dried in the bake oven  
172 at 105 °C for 24 hours to measure the soil moisture. After the samples were  
173 completely shade-dried in the laboratory, gravels and plant roots were removed from  
174 the samples using a sieve with a particle size of 2 mm. A 0.2 g soil sample was  
175 weighed and the concentration of SOC was measured using a wet oxidation method  
176 with dichromate (Nelson and Sommers, 1982 ). The soil texture, including sand, silt,  
177 and clay content, was analyzed through a laser diffraction technique utilizing a  
178 Mastersizer 2000 (Malvern Instruments, Malvern, England).

179 All data in this paper were analyzed by SPSS 21 statistical software. All  
180 collected data underwent normality testing using the Kolmogorov-Smirnov test and  
181 were assessed for homogeneity of variance with Levene's test, ensuring  $P > 0.05$ .  
182 Comparative analysis of various sample point types was performed utilizing a  
183 one-way ANOVA, with significance considered at  $P \leq 0.05$ . All data are expressed as  
184 means  $\pm$  standard deviation. Graphs were made using Origin 2021 software.

## 185 **3 Results**

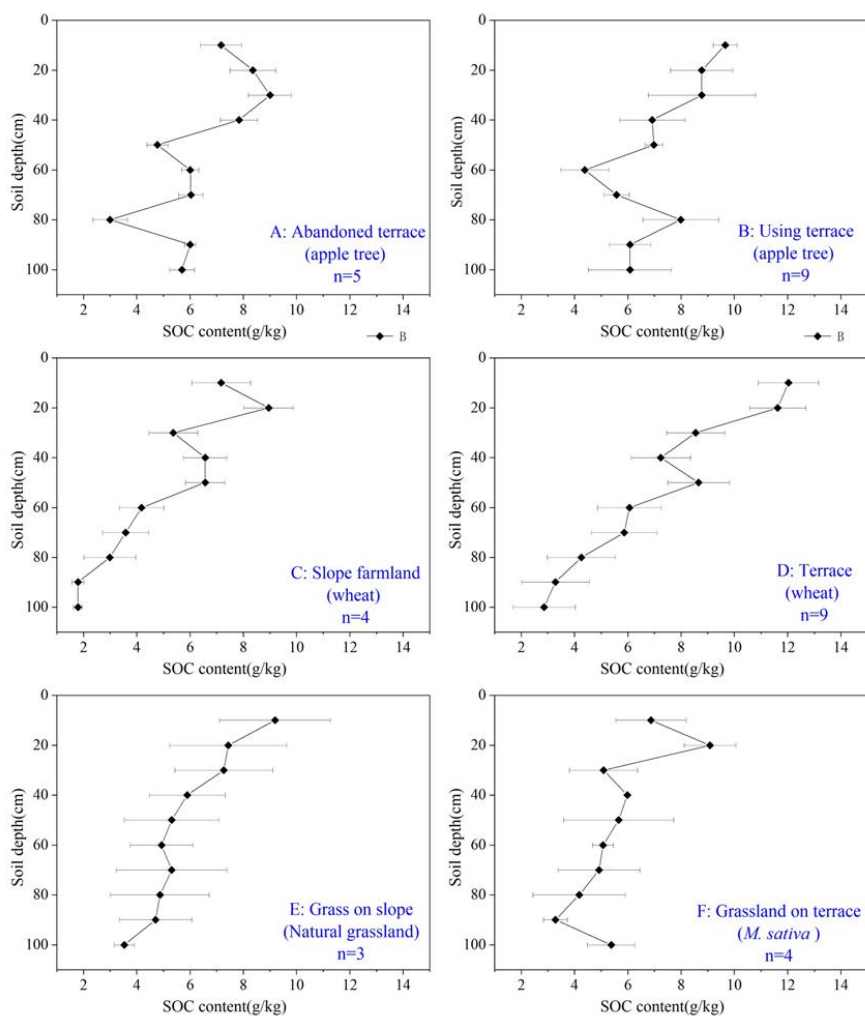
### 186 **3.1 SOC characteristics of different land use types**

187 The SOC content of the abandoned apple tree terraces were lower than that of the  
188 in-use apple tree terraces, but the difference was small at 0.7 g·kg<sup>-1</sup>. The SOC content



189 of the wheat-grown sloping fields was significantly lower than that of the  
190 wheat-grown terraces, with the SOC content of the terraces being 1.5 times higher  
191 than that of the sloping fields, with a difference of  $3.8 \text{ g}\cdot\text{kg}^{-1}$  between the two ( $7.7$   
192  $\text{g}\cdot\text{kg}^{-1} > 4.9 \text{ g}\cdot\text{kg}^{-1}$ ). The SOC content of natural grassland was slightly higher ( $6.0$   
193  $\text{g}\cdot\text{kg}^{-1} > 5.6 \text{ g}\cdot\text{kg}^{-1}$ ) compared to planted grassland, although natural grassland with  
194 weeds had not been terraced. Slopes with natural vegetation were higher in SOC ( $6.0$   
195  $\text{g}\cdot\text{kg}^{-1}$ ) than those under cultivation ( $4.9 \text{ g}\cdot\text{kg}^{-1}$ ) (Table 1).

196 The vertical variation of SOC at 0-100 cm depth varied significantly among land  
197 use types (Fig. 2). Except for the abandoned land, all land use types showed an  
198 irregular decreasing change pattern from the surface layer to the deep layer of the soil.  
199 In the 0-10cm soil layer, the highest SOC content was found in terraces planted with  
200 wheat and the lowest SOC content was found in terraces planted with *M. sativa*.  
201 Terraces planted with fruit trees had significantly higher SOC content at 80-100cm  
202 depth than terraces planted with crops, *M. sativa*, and natural grassland. The vertical  
203 variation of SOC in sloping fields and terraces planted with wheat was consistent,  
204 with their greatest SOC content occurring at 0-20 cm depth and their smallest SOC  
205 content at 90-100 cm depth. In the abandoned terraces, SOC varied between 0-80 cm,  
206 with the smallest SOC content occurring in the 80 cm soil layer.



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208

Fig.2 SOC Vertical variation of SOC content by land use type

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Bars denote the standard deviation of the mean, n represents the number of soil profiles.

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### 3.2 Characteristics of SOC in terraces of different planting patterns

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Differences in SOC were smaller in terraces planted with apple trees and greater

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between terraces planted with a single crop. Among all crop types, legumes had the

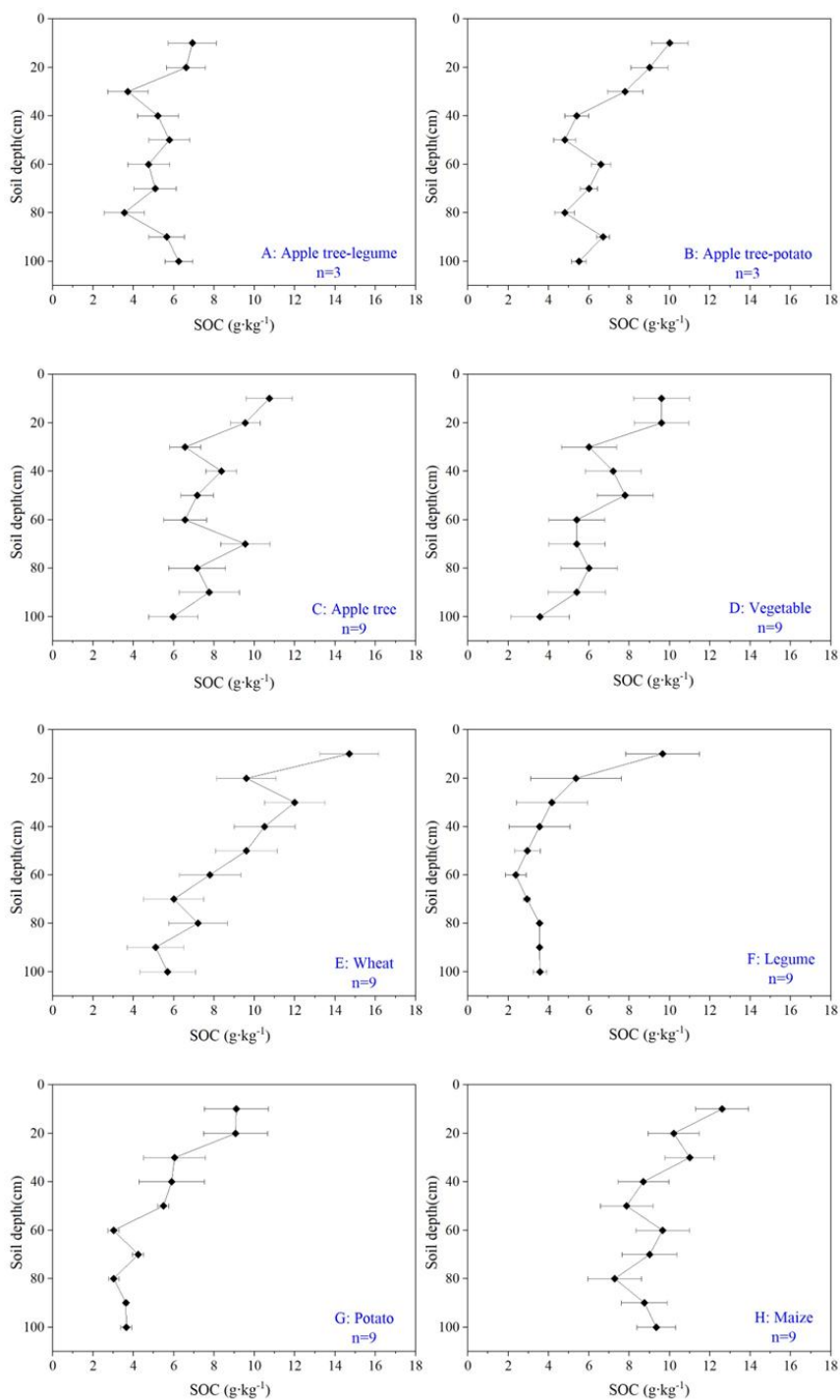
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lowest SOC content ( $4.5 \text{ g}\cdot\text{kg}^{-1}$ ) and maize had the highest SOC content ( $7.7 \text{ g}\cdot\text{kg}^{-1}$ ).



215 The SOC content of concurrent apple tree - legumes and apple tree - potatoes was  
216 higher than that of legumes and potatoes grown alone ( $5.4 \text{ g}\cdot\text{kg}^{-1} > 4.5 \text{ g}\cdot\text{kg}^{-1}$  and  $6.7$   
217  $\text{g}\cdot\text{kg}^{-1} > 5.2 \text{ g}\cdot\text{kg}^{-1}$ ) (Table 1).

218 The vertical distribution of SOC content across a depth of 0-100 cm showed a  
219 consistent decline from the surface soil to the deeper layers for all crops. The SOC  
220 content ( $15.1 \text{ g}\cdot\text{kg}^{-1}$ ) of wheat cultivated terraces was the highest among all crops at  
221 the soil surface (0-10 cm). Terraces planted with apple trees, maize, or wheat had  
222 higher SOC content in deeper soils (30-100 cm). Beans and potatoes had lower SOC  
223 content at 50-100 cm depth than terraces planted with other crops. The difference in  
224 SOC content between potato terraces planted alone and apple tree-potato terraces at  
225 0-20 cm depth was not significant. However, below 20 cm depth, the SOC content of  
226 the apple tree-potatoes combination was significantly higher than that of the terraces  
227 planted with potatoes alone. This difference is also reflected in the legumes and apple  
228 tree-legumes (Fig.3).



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Fig.3 Vertical distribution of SOC content in different crop types



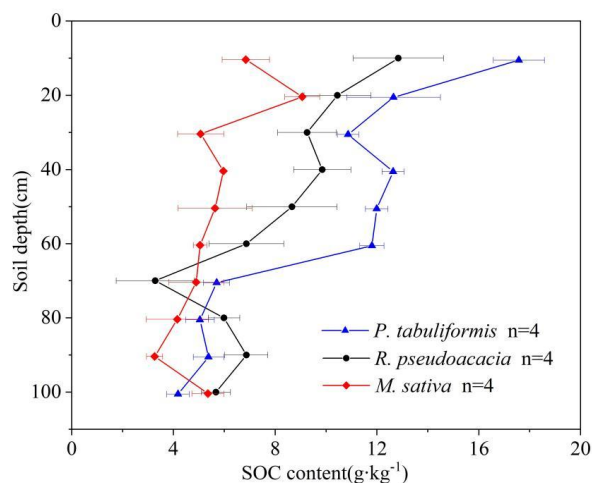
231 Bars denote the standard deviation of the mean, n represents the number of soil  
232 profiles.

### 233 **3.3 Characteristics of SOC in terraces with different ecologically restored** 234 **vegetation**

235 Comparing the SOC characteristics of the three types of ecologically restored  
236 vegetation after terracing, the average SOC content of terraces planted with trees  
237 0-100 cm was higher than that of terraces planted with forage (Table 1). This  
238 difference varied at different depths, with the SOC content of alfalfa being  
239 significantly lower than that of the two trees in the 0-10 cm soil surface layer,  
240 becoming smaller in the 10-20 cm depth, but increasing again in the 20-60 cm depth.  
241 At 70-100 cm depth this difference became smaller and the SOC content between the  
242 three vegetation species became close.

243 The difference in SOC between different silvicultural species was higher in *P.*  
244 *tabuliformis* than in *R. pseudoacacia* at 0-100 cm depth, with a difference of 1.83  
245 g·kg<sup>-1</sup>. The significant difference in SOC between the two species was mainly at 0-70  
246 cm depth, where *P. tabuliformis* had a higher SOC content than *R. pseudoacacia*. The  
247 difference in SOC content between the two species became smaller at the depth of  
248 70-100 cm, and the SOC content of *R. pseudoacacia* was slightly higher than that of *P.*  
249 *tabuliformis* (Fig.4).





250

251 Fig.4 SOC vertical variation of different ecologically restored vegetation.

252 Bars denote the standard deviation of the mean, n represents the number of soil

253 profiles.

## 254 4 Discussion

### 255 4.1 Effect of terrace construction on SOC

256 In the Loess Plateau area, the average SOC content of terraces 0-100cm is 1.4  
257 times higher than that of sloping farmland (Table 1). In Zhang et al. (2013), the SOC  
258 stock at 0-100 cm depth was 4.97 kg·m<sup>-2</sup> in terraces and 3.09 kg·m<sup>-2</sup> in sloping fields,  
259 which is 1.6 times higher than the soil organic carbon stock in sloping fields, which is  
260 consistent with the results of this study. Terracing is considered an important practice  
261 to prevent water erosion and minimize the loss of SOC (Nie et al., 2017). There is a  
262 positive feedback relationship between soil moisture and soil carbon (Green et al.,  
263 2019). Horizontal terraces are the most widespread terraces in the Loess Plateau,

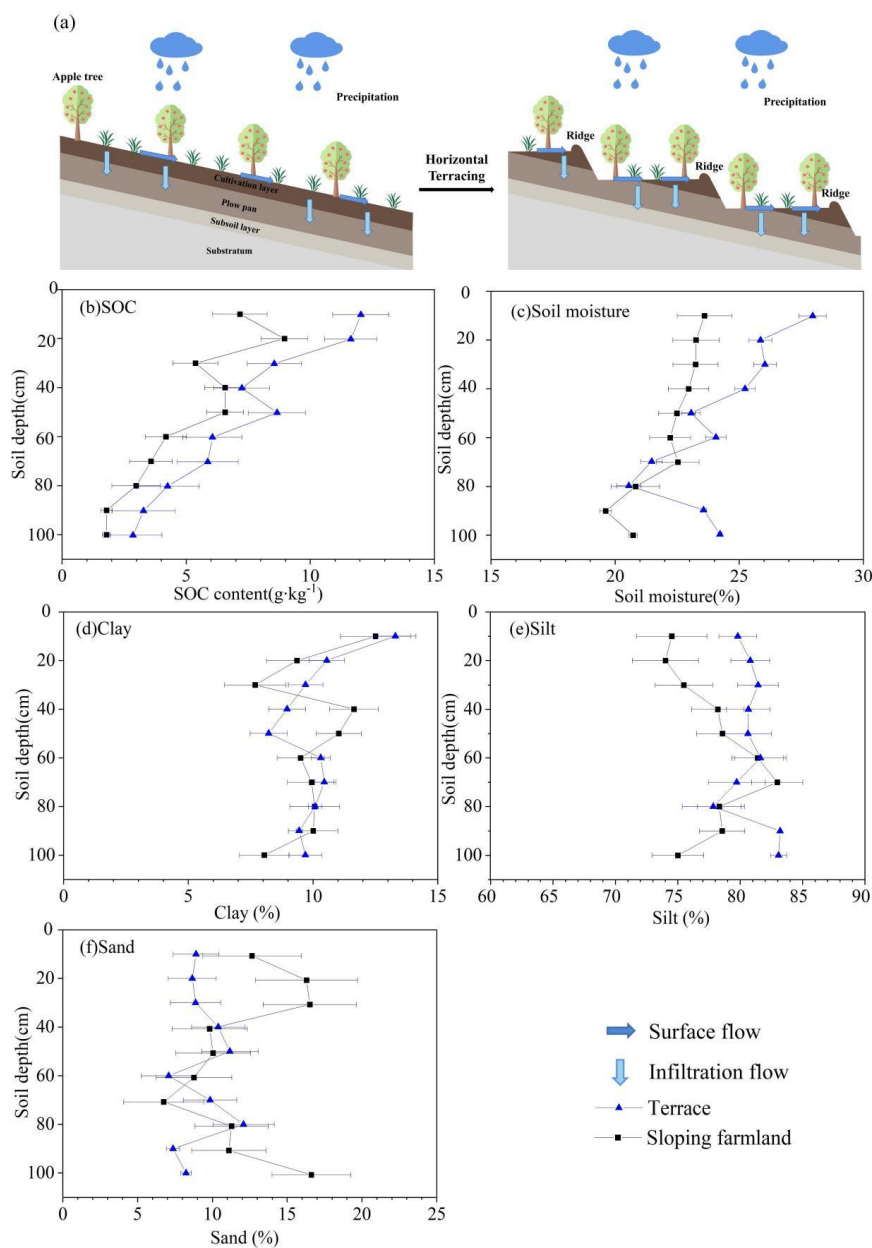


264 which has changed the surface morphology, increased the time of rainfall storage at  
265 the surface, and enhanced soil moisture in rain-dependent farming regions of the  
266 Loess Plateau (Xu et al., 2021). Ecological stress caused by soil water deficit leads to  
267 a decrease in biomass and net primary plant productivity. Conversely, an increase in  
268 soil water has a positive effect on biological growth (McDowell et al., 2015). The  
269 interception of precipitation on the terrace surface provides water for plant growth,  
270 increases plant biomass, and increases the organic matter put into the soil, thus having  
271 an impact on the SOC content. The interception of precipitation by the terraces means  
272 that precipitation will no longer carry large amounts of fine soil particles from the soil,  
273 which will increase the content of clay particles in the soil. Soil clay particles have a  
274 larger specific area, which can adsorb more soil organic carbon and enhance the  
275 accumulation of organic carbon (Post et al., 1982). Compared to sloping land, terraces  
276 have a higher content of both clay and silt in the soil. The terraces therefore further  
277 contribute to carbon accumulation in the terraces by protecting the fine particles in the  
278 soil. In a study on the Loess Plateau, the SOC content of 0-100 cm in untterraced date  
279 palm orchards was  $2.6 \text{ g}\cdot\text{kg}^{-1}$ , which was lower than the soc content of terraced  
280 orchards. This evidence further demonstrates the positive effect of terracing on soil  
281 organic carbon sequestration (Gao et al., 2017).

282 The SOC varies significantly in terms of the amount of plant and animal residues  
283 entering the soil and the depth of the soil under agricultural cultivation (Koga et al.,  
284 2020). The impact of agricultural activities on the surface soil levels was stronger  
285 compared to the deeper soil levels (Li et al., 2020). In this study, we observed a



286 significant increase in SOC in terraces than in sloping lands, particularly in the 0-30  
287 cm soil layer (Fig.5). Post-terracing, SOC sequestration in deeper soils lagged behind  
288 that in surface soils. Furthermore, the rate of SOC change was more pronounced in  
289 the surface layer (0-20 cm) compared to the deeper layer (20-100 cm). Precipitation in  
290 the region is limited and cannot replenish deep soil water, and the erosion of  
291 precipitation on the slope surface also mainly takes away the top soil layer. Therefore,  
292 the soil and water conservation effect brought by terrace construction is limited, so for  
293 the soil depth increases, this effect will become smaller. The impact of terracing on  
294 SOC sequestration diminishes as soil depth increases (Deng, Liu, and Shangguan.,  
295 2014). As soil depth increases, the water stored in the terraces cannot penetrate deeper  
296 soils and deeper soils will maintain their properties. Therefore, the management and  
297 conservation of terrace topsoil are important to ensure local food production and  
298 enhance the carbon sink function (Li et al., 2014).



299

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Fig.5 Effect of terrace construction on SOC, soil moisture, and soil grades.

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(a): Variation in surface morphology by terrace construction; (b): variation in SOC

302

content; (d): variation in soil moisture; (d), (e), and(f): variation in soil grades. The



303 number of profiles is 9 for terraces and 4 for sloping fields. Bars denote the standard  
304 deviation of the mean.

#### 305 **4.2 Effect of terraces abandonment on SOC**

306 As in other parts of the world, industrialization and urbanization have led to a  
307 large population flock from rural to urban areas as in China, resulting in the  
308 abandonment of a large number of productive potential farmlands (Wiesmeier et al.,  
309 2012; Cai et al., 2016). Moreover, a large amount of arable land in rain-fed  
310 agricultural areas in the Loess Plateau region has been abandoned due to water  
311 resource constraints or due to declines in soil fertility (Cao et al., 2020). Secondary  
312 succession of vegetation after terraces abandonment leads to an increase in soil carbon  
313 content as a potential pathway for climate change mitigation (Bell et al., 2021). When  
314 the terraced fields were abandoned in this research, the SOC content of the abandoned  
315 terraces was lower than that of the terraces in use. This is caused by the short  
316 abandoned time. To produce significant environmental benefits, the land must remain  
317 abandoned for an extended period to accumulate substantial amounts of both plant  
318 biomass and the species that constitute intact ecological communities. This process  
319 can take decades to reach levels of carbon sequestration or biodiversity comparable to  
320 those of undisturbed ecosystems (Crawford et al., 2022; Poorter et al., 2016). Due to  
321 the limited water resources available in semi-arid areas, a longer natural or assisted  
322 recovery time is required. Therefore, the duration of land abandonment is a crucial  
323 factor influencing the dynamic changes SOC (Djuma et al., 2020; Badalamenti et al.,



324 2019). In related studies in other regions, soil carbon stocks increased by 13% and  
325 16% in cropland abandoned for 15 and 35 years, respectively (Novara et al., 2014).  
326 With the abandonment of disposal time extended, vegetation types gradually  
327 transition to grassland, scrub, and forest and the death of plants and animals return to  
328 the soil as organic matter, increasing the number of soil aggregates and further  
329 increasing the carbon content of the soil (Liu et al., 2020). Therefore, ecological  
330 restoration of newly abandoned terraces should be carried out as soon as possible.  
331 After short-term abandonment, the terraced fields showed a special change pattern at  
332 different depths in this study. SOC content first decreased and then increased with  
333 increasing soil depth. The decrease in surface SOC was controlled by the decrease in  
334 agricultural fertilizer inputs, while the increase in deep SOC was caused by the  
335 inability to utilize deep soil nutrients due to the death of crop roots.

#### 336 **4.3 Effect of vegetation type and planting patterns on SOC in terraces**

337 Vegetation types can influence SOC by modifying the soil's physicochemical  
338 structure and altering both the input and decomposition rates of SOC (Du et al., 2022;  
339 Wiesmeier et al., 2012; Wan et al., 2019). Our study demonstrated that, compared to  
340 terraced fields, the SOC content of afforested land at a 0-100 cm depth was higher and  
341 that the forest litter biomass was more than that of farmland, which was the main  
342 reason for this difference. Planted forest land reduces soil temperature, soil moisture  
343 evaporation, and soil erosion while increasing the quantity and quality of organic  
344 matter input to compensate for carbon decomposition from crop cultivation (Liu et al.,



345 2020). The afforested land is terraced forests, and the effect of preventing soil erosion  
346 is more significant. Some study shows that the SOC in immature forests (10 years old)  
347 is 17.91% higher than that in terraced cropland. The SOC concentration of a  
348 30-year-old forest is significantly higher than that in other land covers (Xin et al.,  
349 2016). These studies further proved the carbon sequestration effect of reforestation.

350 Due to the problem of ecological degradation and soil erosion, various ecological  
351 measures have been taken in the Loess Plateau area, such as returning farmland to  
352 forest and grass and planting trees (Hong et al., 2020). Considering the climate and  
353 soil quality factors, the main species selected in the Loess Plateau region are  
354 drought-tolerant types of trees, and the carbon accumulation effect of different species  
355 selection also differs significantly (Li et al., 2018). *P. tabuliformis* has a higher SOC  
356 content than *R. pseudoacacia*, especially in the 0-50 cm soil layer. The pine species  
357 selected in this region is larch, with the arrival of winter a large number of pine  
358 needles and fruits are into the soil, increasing the input of organic matter in the  
359 surface layer so that the SOC content of pine forests is higher in the surface layer of  
360 the soil (0-10 cm). In humid areas, some studies also show that the soil organic carbon  
361 density of fir conifer forests is the largest among the different 11 middle forest  
362 vegetation types (Chen et al., 2007). Other studies have shown that tree species such  
363 as *P. koraiensis*, *L. gmelinii*, and *P. tabuliformis* increase soil organic carbon stocks  
364 more as silvicultural species (Hong et al., 2020). The biomass of the herbaceous  
365 plants themselves is much lower than that of trees, and the limited amount of organic  
366 matter entering the soil, and the fact that *M. sativa* is mainly used as a source of



367 fodder for the animals raised by farmers in the region, leads to a lower SOC content in  
368 terraces planted with *M. sativa* than in those undergoing afforestation.

369 The SOC content of grassland at a depth of 0-100cm is lower than that of  
370 farmland. Although the grassland has organic matter after the withered herbs enter the  
371 soil, the main planting type in the terrace area is apple trees, and a large amount of  
372 fruit tree leaves will also enter the soil. Grassland is a sloping land that has not been  
373 terraced, leading to slope erosion that removes a significant amount of organic matter  
374 from the soil surface. As a result, the SOC content in grassland is lower than in  
375 terraced fields (Fig.2). The ecological advantages of sequestering SOC and enhancing  
376 soil fertility could be significant, largely thanks to the widespread implementation of  
377 reforestation and various land use strategies in terraced fields across China and  
378 numerous other mountainous areas globally (Hong et al., 2020).

379 Crops may differ in their ability to increase SOC content due to differences in  
380 their photosynthetic capacity and root characteristics (Wegener et al., 2015). The  
381 pattern of intercropping in this area is typical of Agroforestry systems (AFS), where  
382 other crops are planted between the rows of apple trees. The SOC content of apple  
383 trees in combination with other crops was higher than in monocultures, especially in  
384 the lower and middle layers of the soil (30-100 cm). The amount of tree litter and root  
385 decomposition are important reasons for this (Pardon et al., 2017). The fallen leaves  
386 of fruit trees and some rotting apples are not removed, and these organic materials  
387 decompose to replenish SOC after entering the soil. In addition, carbon input can be  
388 achieved by decomposing (fine) tree roots and root secretions (Nair et al., 2009). For





389 soils below 30 cm depth, tree roots produce an important role in the accumulation of  
390 soil organic carbon. When potato or legume crops are harvested, all the fruit and plant  
391 roots are removed and these lands will be tilled to grow other crops, so the input of  
392 organic matter is very limited. Agroforestry systems increase the distribution of roots  
393 in the soil and increase the recalcitrant compounds which slow the rate of  
394 mineralization through the input of organic matter (Recous et al., 2008).

#### 395 **4.4 Study limitations**

396 The results of the study are based on field data collected over a relatively short  
397 period of time. Due to the complexity of field conditions, the number of soil profiles  
398 in some of the comparative studies in the sampling frame design was not entirely  
399 consistent, and future studies will need to expand the study area to achieve balanced  
400 sampling. This study examined differences in SOC at individual time points, and  
401 follow-up assessments are needed to confirm long-term trends. Factors such as soil  
402 bulk density, soil pH, root biomass, fertilizer management, and tillage practices also  
403 affect soil organic carbon in terraced areas, and more indicator measurements and  
404 studies are necessary. We need to do more work to understand the SOC characteristics  
405 of terraced agricultural areas and how to better utilize the terraces for carbon storage  
406 and realize the economic and ecological value of terraces.

#### 407 **5 Conclusions**

408 The results showed that the SOC content in 0-100 cm of terraced fields planted



409 with wheat was 1.5 times higher than that of sloping fields planted with wheat.  
410 Compared with sloping land, terrace construction significantly increased the SOC  
411 content of cultivated land, especially in the top soil layer (0-30 cm), and converting  
412 some sloping land into terraces would enhance the carbon sequestration capacity.  
413 Abandonment, vegetation type and planting structure affect the SOC of terraces.  
414 planting other crops between rows of apple trees can increase the SOC content. Since  
415 vegetation restoration takes a long time, short-term abandonment will lead to a  
416 decrease in terrace SOC, and some abandoned terraces can be planted with ecological  
417 restoration vegetation. Among the ecologically restored plant species, the vegetation  
418 with the highest SOC content is *Pinus oleifera*. The SOC content of terraces planted  
419 with artificial forage is lower than that of natural grassland, so it is necessary to  
420 protect the natural grassland left behind and choose tree species with better ecological  
421 benefits when planting trees. In the face of China's huge food pressure and the goal of  
422 increasing carbon sinks to mitigate global climate change, terraces have significance  
423 and importance. Continuous strengthening of terraces management will give full play  
424 to their carbon sequestration role.

#### 425 **Data Availability Statement**

426 The data that support the findings of this study are available on request from the  
427 corresponding author, soil organic carbon data are not publicly available due to  
428 privacy or ethical restrictions.



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437 **Conflict of Interest Statement**

438 The authors declare no conflicts of interest.

439 **Author contributions**

440 Guofeng Zhu and Qinqin Wang conceived the idea of the study; Siyu Lu, Xiaoyu Qi  
441 and Ling Zhao analyzed the data; Dongdong Qiu, Longhu Chen and Rui Li were  
442 responsible for field sampling; Qinqin Wang and Yuanxiao Xu participated in the  
443 experiment; Yinying Jiao, Gaojia Meng and Wenmin Li participated in the drawing;  
444 Qinqin Wang wrote the paper; Yuhao Wang, Wentong Li and Eenwei Huang checked  
445 and edited language. All authors discussed the results and revised the manuscript.

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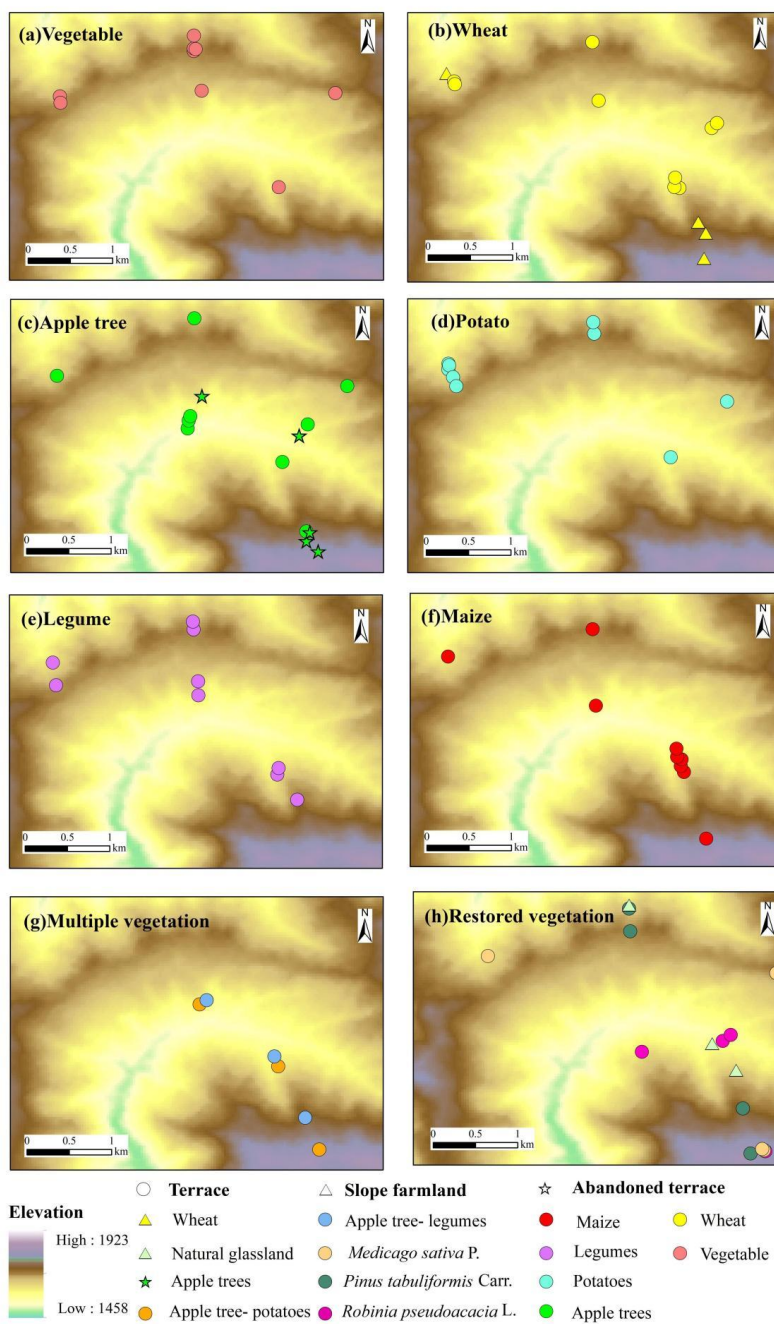
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701 **Appendix**



702

703 Fig.A1 Detailed distribution of randomly placed sampling points in the study area. These sampling





704 points cover various crop types, including apple trees, vegetable, wheat, legume, potato,  
705 maize, apple tree-legume, and apple-potato. Additionally, the figure also displays the  
706 distribution of sampling points under different planting patterns, specifically terraced  
707 sampling points, sloping land sampling points, and abandoned terraced sampling points.