



1 **Surface Observation Constrained High Frequency Coal**
2 **Mine Methane Emissions in Shanxi China Reveal More**
3 **Emissions than Inventories, Consistency with Satellite**
4 **Inversion**

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11 **Abstract.** This work focuses on Changzhi, Shanxi China, a city and surrounding rural region with one
12 of the highest atmospheric concentrations of methane (CH₄) world-wide (campaign-wide
13 minimum/mean/standard deviation/max observations: 2.0, 2.9, 1.3, and 16 ppm) due to a rapid increase
14 in the mining, production, and use of coal over the past decade. An intensive 15-day surface observation
15 campaign of CH₄ is used to drive a new analytical, mass-conserving method to compute and attribute
16 CH₄ emissions. Observations made in concentric circles at 1km, 3km, and 5km around a high production
17 high gas coal mine yielded emissions of 0.73, 0.28, and 0.15 ppm min⁻¹ respectively. Attribution used a
18 2-box mass conserving model to identify the known mine's emissions from 0.042-5.3 ppm min⁻¹, and a
19 previously unidentified mine's emission from 0.22-7.9 ppm min⁻¹. These results demonstrate the
20 importance of simultaneously quantifying both the spatial and temporal distribution of CH₄ to better
21 control regional-scale CH₄ emissions. Results of the attribution are used in tandem with observations of
22 boundary layer height to quantify policy-relevant emissions from the two coal mines as 13670±7400 kg
23 h⁻¹ and 5070±2270 kg h⁻¹ respectively. Both mines display a fat tail distribution, with respective 25th,
24 median, and 75th percentile values of [870, 7500, 38700] kg h⁻¹ and [431, 1590, 7000] kg h⁻¹. These
25 findings are demonstrated to be higher than CH₄ emissions from equivalent oil and gas operations in the
26 USA, with one about double and the other similar to day-to-day emissions inverted over 5-years using
27 TROPOMI over the same region.

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31 **1. Introduction**

32 Emissions of Methane (CH₄) contribute the second most to direct anthropogenic longwave radiative
33 forcing (Etminan et al., 2016; Li et al., 2022). Since CH₄ has a lifetime from 9.5 to 12.5 years (Li et al.,
34 2022; Prather et al., 2012), controlling CH₄ emissions can provide an opportunity to mitigate peak
35 loading and slow the rate of net global warming. Fossil fuel CH₄ is one of the largest sources of
36 anthropogenic CH₄ emissions (Kirschke et al., 2013; Saunois et al., 2020a). Since China is the world's
37 largest producer and consumer of coal (Bournazian, 2016), Coal mines CH₄ (CMM) possibly contributes
38 up to 33%-40% of China's total CH₄ emissions (Janssens-Maenhout et al., 2017; Miller et al., 2019; Peng
39 et al., 2016). Although China enacted CMM regulations in 2010, CMM continues grow (Kerr and Yang,
40 2009; Miller et al., 2019). CH₄ emission estimates are highly uncertain in both space and time (Brandt et
41 al., 2014; Saunois et al., 2020b). They also generally have a fat tail distribution, wherein a small number
42 of samples have extremely large emissions that overwhelm emissions under average conditions (Duren
43 et al., 2019; Plant et al., 2022). For these reasons, new approaches to quantify, reduce uncertainty, and
44 attribute CH₄ emissions are necessary and can provide support for policies aiming to control and mitigate
45 CMM (Cao, 2017).

46 Bottom-up (BU) quantification of emissions requires a priori knowledge of source locations and
47 diversity, which tends to not represent real-world conditions. Top-down (TD) approaches analyze
48 concentration data with improving accuracy (Allen, 2014; Rigby et al., 2019; Varon et al., 2018; Vaughn
49 et al., 2018), specifically combining surface (Heerah et al., 2021; Katzenstein et al., 2003; Shi et al.,
50 2023), aircraft (Karion et al., 2013; Shi et al., 2022; Tong et al., 2023; Vinković et al., 2022), and/or
51 satellite (Wecht et al., 2014) CH₄ observations with atmospheric models. Some TD approaches use
52 physically realistic but complex chemical transport models (Bloom et al., 2017), others use plume
53 models (Goldsmith et al., 2012), and others still use data driven approaches (Buchwitz et al., 2017).
54 Uncertainties are rarely assessed holistically or in detail (Cohen and Prinn, 2011; Cohen and Wang, 2014).

55 Airborne remote sensing is a highly technical and costly approach to record CH₄ fluxes from
56 landfills, coal basins, and oil and gas production (Krautwurst et al., 2021; Krautwurst et al., 2017;
57 Kuhlmann et al., 2023), which suffers from not being able to monitor CH₄ emissions over long periods
58 of time or in regions where the source is not well constrained (Brandt et al., 2014; Gorchoy Negron et
59 al., 2020; Hiller et al., 2014; Mehrotra et al., 2017; Molina et al., 2010). Satellite remote sensing can
60 measure CH₄ under specific orbits where the source is known and identified (Jacob et al., 2016; Jacob et



61 al., 2022; Plant et al., 2022; Varon et al., 2018; Zhang et al., 2020), but only after being calibrated by
62 upward looking remotely sensed measurements (Tu et al., 2022), and only when the atmosphere is rain,
63 cloud and aerosol free (Cohen and Prinn, 2011; Reuter et al., 2019; Sadavarte et al., 2021). TROPOMI
64 and GOSAT have both been shown to be data-rich at times (Butz et al., 2012; Hu et al., 2018; Jacob et
65 al., 2016), but severely limited at other times (Butz et al., 2012; Kuze et al., 2009). Even when these
66 satellites have sufficient data to compute emissions from other species, frequently CH₄ cannot be
67 computed (Li et al., 2023; Qin et al., 2023b) due to insufficient signal strength, and uncertainties which
68 are both non-understood and mis-constrained (Povey and Grainger, 2015).

69 Ground-based remote sensing provides higher accuracy versus satellite observations (Heerah et al.,
70 2021; Luther et al., 2022; Tu et al., 2022). EM27/SUN measurements have approximated CH₄ emissions
71 in Poland (Luther et al., 2019; Luther et al., 2022). However, these instruments are expensive, require
72 calibration, and have limited data collection due to solar signal strength.

73 This work employs a high-frequency surface-based observation platform of CH₄ concentration
74 which is portable, economical, and unaffected by most environmental factors. The observations are
75 combined with a new mass-conserving methodology based on temporal transformation of the spatially
76 derived mass-conserving framework successfully applied to NO₂ (Li et al., 2023; Qin et al., 2023b). This
77 work focuses on Shanxi, one of the densest coal mining regions in the world, accounting for
78 approximately 10% of total global coal production (Lin and Liu, 2010; Qin et al., 2023a). Continuous
79 observations were made around known coal mines, unknown sources, and of background conditions.
80 High-frequency emissions calculated using these data were used to drive a 2-box model to attribute
81 emissions to the known mine and a second low production mine previously thought insignificant. The
82 results provide insights into the spatial distribution of CH₄ emissions, demonstrate rapid adoption of
83 practical methods globally, and enable source attribution.

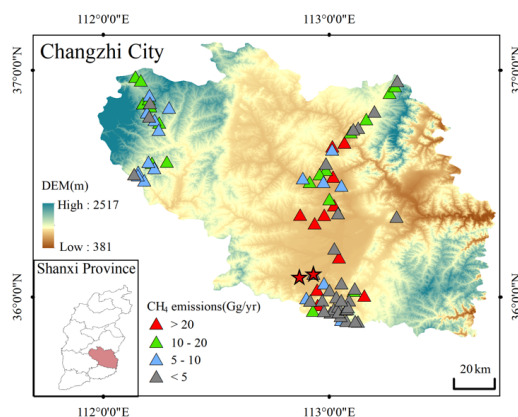
84 **2. Methods and Data**

85 **2.1 Study Site and Campaign Design**

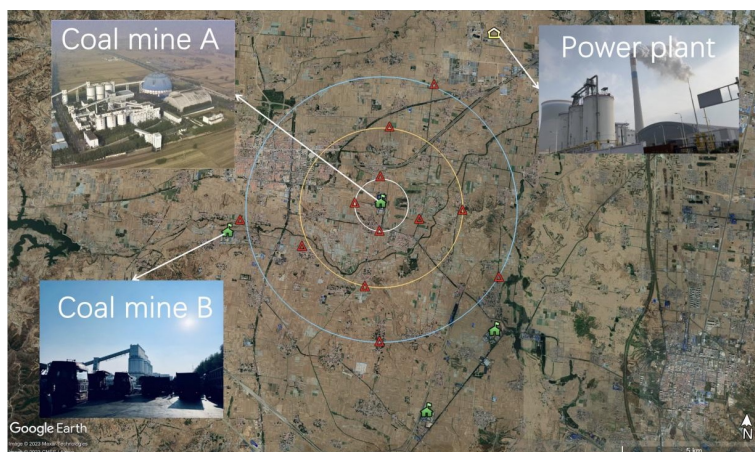
86 Changzhi, Shanxi is located in a basin, with coal mines densely distributed throughout both flat
87 central regions and around the mountainous edges (Figure 1), many of which are classified as high CH₄
88 emitting mines. Due to this combination, province-wide background CH₄ concentrations are very high
89 and have large variation in time. This study mainly focuses on two coal mines: one mine is classified as



90 having high amounts of CH₄ per unit of production and an annual coal production of 4 million tons (CM-
91 A), and the other is unclassified for CH₄ per unit of production and having an annual coal production of
92 3 million tons (CM-B) (Qin et al., 2023a). Observations were positioned along concentric circles located
93 1km, 3km, and 5km from CM-A, over an approximation of the four ordinal directions: east, west, south,
94 north (Figure 2). All locations were planned to be far away from known anthropogenic sources, leading
95 to a net total 12 measurement points. As later discovered, CM-B is located approximately 1km southwest
96 from the measurement point located at 5km west.



97
98 **Figure 1. Topographic map of Changzhi, and its location in Shanxi Province (bottom left). The triangles**
99 **represent the locations of all individual coal mines (including underground and abandoned mines), where the**
100 **triangle color represents the emissions amount: high (red), middle (green), low (blue), and very-low (grey).**
101 **The red stars represent the two coal mines in this work.**



102
103 **Figure 2. Locations of four individual coal mines (Green filled houses), a power plant (Yellow outlined house),**
104 **and the 12 observation locations presented in this work (red double-outlined triangles). Distance from CM-A**



105 are given as concentric circles at 1km (white), 3km (yellow), and 5km (blue).

106 **2.2 Measuring CH₄ concentration**

107 Atmospheric CH₄ concentrations at 5m above the surface were observed daily at 1 Hz from 8:30 to
 108 17:00 local time in August 2022 using two portable greenhouse gas analyzers (LGR-915-0011, California,
 109 USA). Three different locations (1 km, 3 km and 5 km) were selected daily along a single direction from
 110 the CM-A, allowing a more consistent and precise calculation of the spatial gradient (Table 1). In order
 111 to reduce the time errors, using two portable greenhouse gas analyzers randomly selecting the three
 112 observation points during the daily measurements and without fixed sequence.

113 The CH₄ data was averaged minute-by-minute to match observed wind data, and subsequently used
 114 to compute CH₄ emissions. As show in Figure 3, the CH₄ concentration data is highly correlated with
 115 rapid changes in both the wind speed and direction.

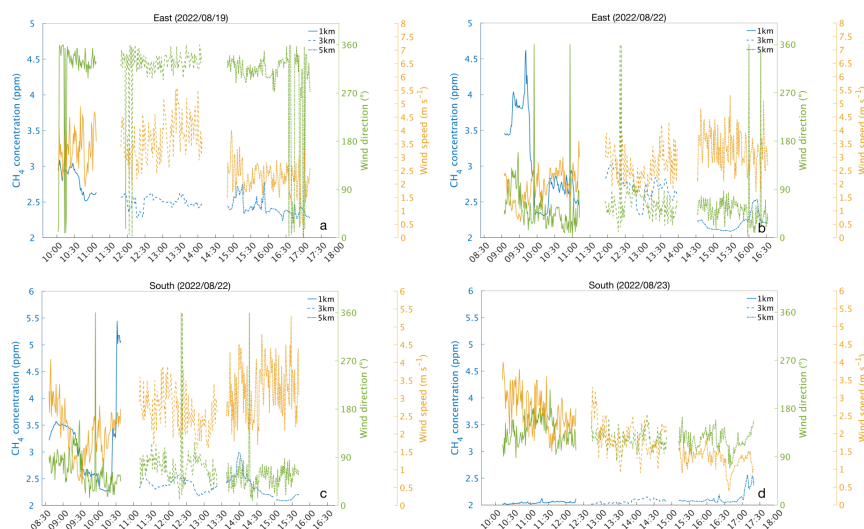
116 **Table 1. Detailed Information Summarizing the Dates, Times, and Locations Observed in the Field**

Date	Direction of measured site	Sample period (min)		
		1 km	3 km	5 km
10 August 2022	North	88	35	47
12 August 2022	North	78	66	61
13 August 2022	North	168	181	178
14 August 2022	North	125	128	122
15 August 2022	North	121	122	119
17 August 2022	North, West	122, 124	119, 120	120, 117
18 August 2022	North, West	143, 119	135, 118	125, 120
19 August 2022	East	126	139	140
21 August 2022	North	140	127	125
22 August 2022	East, South	129, 123	124, 133	121, 124
23 August 2022	South	126	129	129

117 Observations made in clean locations with a wind direction not from the mine are subsequently
 118 considered for background sites. The lowest and least variable CH₄ observations are found on August 23
 119 in the south (2.08±0.08 ppm) (Figure 3). It is important to note that although this site has the minimum
 120 concentrations observed in this work, these values are significantly higher than the global latitude-band
 121 background. Three other locations and days were observed with relatively low mean and not significantly
 122 large variation: August 19 in the east (2.63±0.35 ppm), August 22 in the east (2.65±0.51 ppm), and
 123 August 22 in the south (2.60±0.55 ppm) (Figure 3). These results imply that the practice adopted by the
 124 community to separate a plume from the global latitude band or climatological background state is not
 125 applicable in the locations sampled in this paper (Buchwitz et al., 2017; Irakulis-Loitxate et al., 2021;
 126 Lauvaux et al., 2022; Sadavarte et al., 2021). For this reason, a new quantitative approach is presented
 127 to understand and quantify what is actually a source and what is not. This approach is applicable under
 128 conditions both encountered globally as well as those under the uniquely high and variable conditions



129 observed herein.

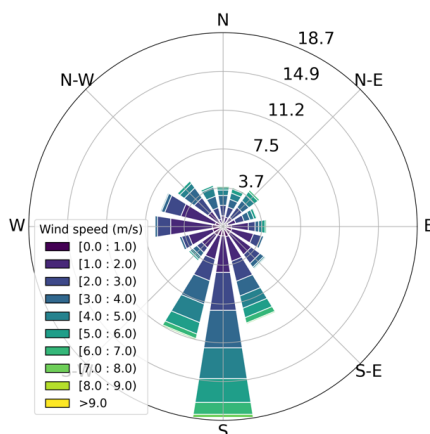


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 131 **Figure 3.** Time series of CH₄ concentration (ppm) (blue), wind speed (m s⁻¹) (yellow) and wind direction [°]
 132 (orange lines) measured at 1km (solid), 3km (dashed) and 5km (dash-dot) located east (top) and south (bottom)
 133 of CM-A on three different days.

134 **2.3 Meteorological Data**

135 The wind speed and direction were obtained from local meteorological stations with a temporal
 136 frequency of 1min. As show in Figures 4 and 5 the overall wind was dominated by a southerly direction
 137 (38.0% of observations between 150° and 210°) and found to be moderately slow (69.9% of observations
 138 were between 1 m s⁻¹ and 4 m s⁻¹). The 10th and 90th percentiles of wind direction (54° and 312°) and
 139 wind speed (1 m s⁻¹ and 5.1 m s⁻¹) respectively, indicate that high frequency sampling reveals a small
 140 number of relatively large changes are observed, which are expected to lead to a “fat-tail” type of
 141 distribution of subsequently computed CH₄ emissions (Delkash et al., 2016).

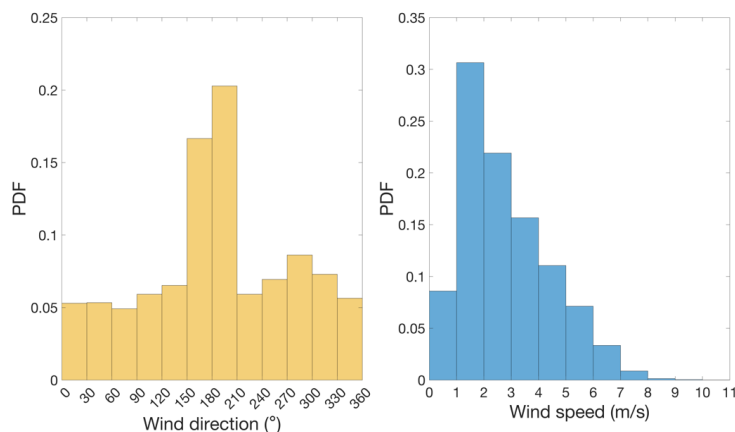
142 The temperature and pressure data were measured by a handheld meteorological instrument
 143 (HWS1000, ZOGLAB, China) with a temporal frequency of 1min. The boundary layer data were
 144 obtained from <https://zenodo.org/records/6498004> (Guo et al., 2022) based on a merging of reanalysis
 145 data with observations (Guo et al., 2024).



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Figure 4. The wind rose of all observed wind speeds from August 10, 2022 to August 25, 2022.



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Figure 5. Probability density function (PDF) for all observed wind direction (yellow) and wind speed (blue) from August 10, 2022 to August 25, 2022.

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2.4 Quantitative Estimation of CH₄ Emissions

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A mass conserving approach was used to estimate the CH₄ emissions in connection with the high frequency observations of CH₄ and meteorological data, hereafter called the Mass Conserving Model of Measured CMM (MCM²). This approach is based on previous dynamic emissions estimates of tropospheric atmospheric column observations of short-lived NO₂ (Li et al., 2023; Qin et al., 2023b), but has never been applied to surface observations in general, or CH₄ in specific. Adopting this approach to solve for CH₄ is done starting with the continuity equation for the conservation of mass (Equation 1), reorganizing the individual terms and converting coordinates from space to time (Equation 2) and finally combining the terms (Equation 3) as follows:

160

$$\frac{\partial CH_4}{\partial t} = E_{CH_4} - \nabla(U \times CH_4) \quad (1)$$

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$$\nabla(U \times CH_4) = CH_4 \times \nabla U + U \times \nabla CH_4 = \alpha \times \left(CH_4 \times \frac{\partial U}{\partial t} + U \times \frac{\partial CH_4}{\partial t} \right) \quad (2)$$



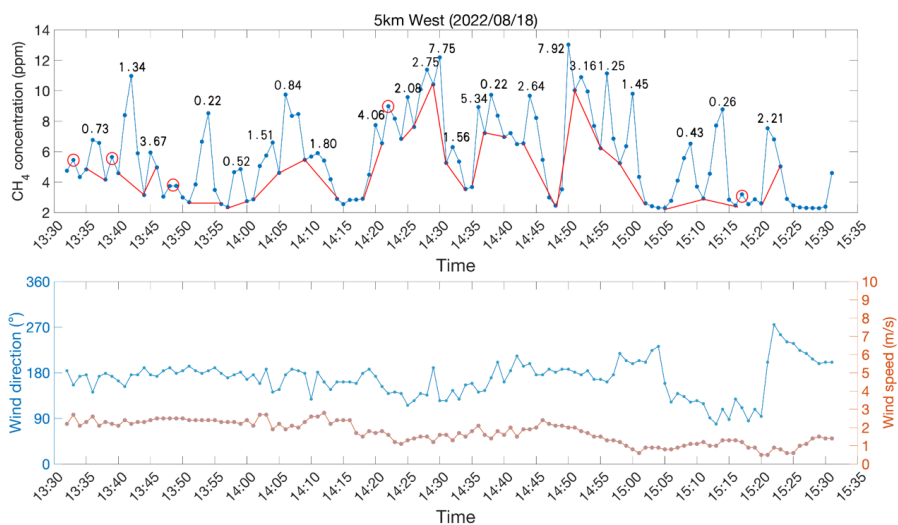
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$$\frac{\partial CH_4}{\partial t} = E_{CH_4} - \alpha \times \left(CH_4 \times \frac{\partial U}{\partial t} + U \times \frac{\partial CH_4}{\partial t} \right) \quad (3)$$

163 where CH_4 is the CH_4 concentration (ppm), t is the time (min), E_{CH_4} is the CH_4 emissions flux (ppm min⁻¹), and U is the wind speed (m s⁻¹). The ∇ is a mathematical operator that takes the gradient on spatially
164 distributed variables. However, when considering motion along one-dimension, the relationship between
165 distance, speed, and time can be used to rewrite the spatial derivatives of $\nabla(CH_4)$ and $\nabla(U)$ as temporal
166 derivatives (Brasseur and Jacob, 2017), where α is a conversion coefficient between distance and wind
167 speed.

168
169 The gradient term used in these equations take into account the local topography of Shanxi, which
170 is known for its significant features and surrounding mountains. These geographical features can impact
171 the transport and dispersion of CH_4 , and their effects are incorporated into the wind field in the continuity
172 equation. Notably, when dealing with a non-divergent wind field, the gradient term simplifies to the term
173 ($U \times \nabla CH_4$) (Sun, 2022). Uncertainty analysis was conducted before calculating the CH_4 emissions to
174 ensure only reliable data was used, since observed variation of CH_4 over time is influenced not only by
175 CH_4 emissions, but also changes in wind speed and pressure. Specifically, $CH_4 \times \frac{\partial U}{\partial t}$ represents the
176 change in CH_4 influenced by pressure, while $U \times \frac{\partial CH_4}{\partial t}$ represents the change in CH_4 influenced by
177 advection. Furthermore, since there is uncertainty in the observations, this work takes a conservative
178 approach, and only considers data when the threshold given by equation (4) is observed to be considered
179 influenced by emissions (a lower threshold can be selected like 25% or 15% et al., but uncertainty will
180 increase).

181
$$u \times \frac{\partial CH_4}{\partial t} / \nabla(U \times CH_4) > 30\% \quad (4)$$

182 The remaining data (approximately 22%, presented as red circle indicators in Figure 6) is not processed
183 in the emissions calculation as the signal is most likely due to a combination of observational uncertainty
184 and white noise (Prinn et al., 1987).



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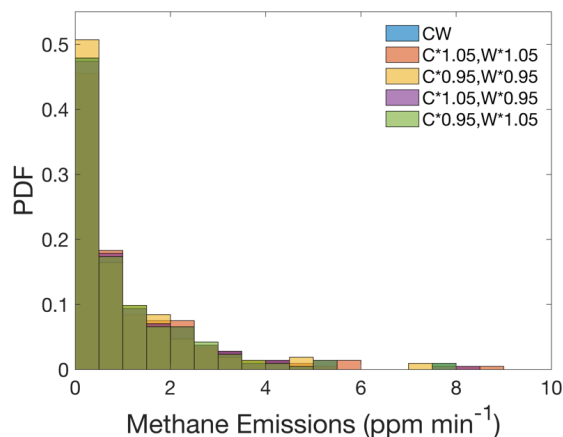
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Figure 6. Time series of CH₄ concentrations (top, blue), background concentrations (top, red), wind direction (bottom, blue) and wind speed (bottom, red) measured 5km west of CM-A on August 18, 2022. MCM² computed CH₄ emissions (top, numbers) (ppm min⁻¹) are computed for all regions where the observations are enhanced compared with the background for at least 3 consecutive observations, and which further pass the noise threshold (Equation 4).

2.5 Uncertainty Analysis

In order to reduce the uncertainty of the CH₄ emission estimation, only data above the threshold given by equation (4) is considered. Prior to this, uncertainty analysis was also conducted on the relevant variables in actual experiments. As shown in Figure 7, a 5% uncertainty was assigned to both the CH₄ concentration and wind speed data, and the CH₄ emission flux was calculated. The uncertainty analysis results indicate that probability distribution of all possible calculated emissions are consistent, and the errors are smaller than 5% in each case, consistent with Equation 3 leading to a dampening of the uncertainty, as also observed in Qin et al. (2023a) study. Therefore, we believe that the results of CH₄ emissions in this study can be trusted.



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Figure 7. The PDFs of uncertain analysis results (C represents concentration, W represents wind speed).

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2.6 Attribution Analysis

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A 2-box mass conserving model (based on equation 5) was used to attribute CH₄ emissions from the more than one suspected source of CH₄ in the 5k west. The changes in CH₄ over time *t* (min) at the observation point $C_{coal\ mine}$ (ppm) is driven by emissions from the upwind coal mine $E_{coal\ mine}$ (ppm min⁻¹) and the concentration gradient computed using the wind *U* (m s⁻¹), and the background concentration $C_{background}$ (ppm) as demonstrated in Figure 8.

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$$\frac{\partial C_{coal\ mine}}{\partial t} = E_{coal\ mine} + U \times C_{background} - U \times C_{coal\ mine} \quad (5)$$

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All observed data points and computed emissions are used when wind direction is capable of transporting the CH₄ from either CM-A or CM-B towards the observation site (Figure 8), while the remaining data is not used. A discretized version of Equation 5 is given in Equation 6 and solved using a first order finite difference approach:

213

$$C_{coal\ mine_{\tau_{i+1}}} - C_{coal\ mine_{\tau_i}} = E_{coal\ mine_{\tau_i}} + U_{\tau_i} \times C_{background_{\tau_i}} - U_{\tau_i} \times C_{coal\ mine_{\tau_i}} \quad (6)$$

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where τ_i and τ_{i+1} are the current and next time step, and the other terms are defined as in equation 5.

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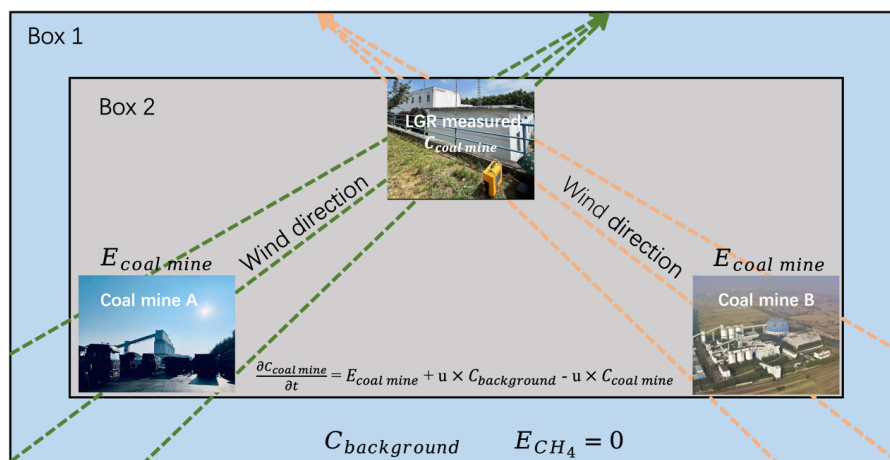
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All possible sets of steady-state concentrations are computed using all possible combinations of emissions and concentrations as boundary and initial conditions and running the equation forward to equilibrium. The computed concentrations are analyzed probabilistically by comparing the modeled CH₄ probability density function (PDF) with the observed CH₄ PDF. Differences between the PDFs are clearly associated with the different wind directions and hence geophysical locations of the sources can be distinguished.



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Figure 8. Schematic diagram of the 2-Box model.

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2.7. Converting Emissions into Policy-Relevant Units

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In order to compare the emissions with some other studies, the units (ppm min^{-1}) were converted into policy-relevant units (kg h^{-1}), although as outlined below, this conversion leads to a larger uncertainty range. According to the attribution analysis in Section 2.6, when the wind direction is located within a 60° arc of coal mine A or coal mine B (Figure 8), the respective CH_4 emissions which successfully passed attribution were assigned to the respective coal mine. Therefore, based on the wind direction, CH_4 emissions of coal mine A were screened from the CH_4 emissions captured at North 1km, and CH_4 emissions of coal mine B were screened from the CH_4 emissions captured at West 5km. adopting the following equation (7) to convert the units from ppm min^{-1} to kg h^{-1} :

232

$$E'_{CH_4} = E_{CH_4} \times \rho_{air} \times H \times A \times 60 \quad (7)$$

233

$$\rho_{air} = \frac{P \times M_{air}}{R \times T} \quad (8)$$

234

where E'_{CH_4} is the CH_4 emissions (kg h^{-1}), E_{CH_4} is the CH_4 emissions (ppm min^{-1}), ρ_{air} is the dry gas density (kg m^{-3}) (based on equation 8), H is the height of the vertical rise that the emissions undergo within their first minute (m), A is the area (m^2) swept over an arc, which ranges linearly from 60° under slow wind conditions to 30° over very fast wind conditions, based on the wind speed when the direction is found to lead to successful attribution, P is the atmosphere pressure (Pa) over the sampling duration, M_{air} is the molecular weight of dry air, which is a fixed constant ($28.97 \times 10^{-3} \text{ kg mol}^{-1}$), R is the universal gas constant ($8.314 \text{ J mol}^{-1} \text{ K}^{-1}$), T is the air temperature (K) over the sampling duration.

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Two different assumptions are made for the vertical extent of the plume rise, since the emissions are computed minute-by-minute which is shorter than the adjustment time throughout the entire boundary layer (Vaughn et al., 2018; Zinchenko et al., 2002). The first is to assume it has mixed within the bottom one fourth of the boundary layer, and the second is that it has mixed based on a steady vertical rise equal to one tenth of the horizontal wind. In this work, results using both assumptions will be presented.

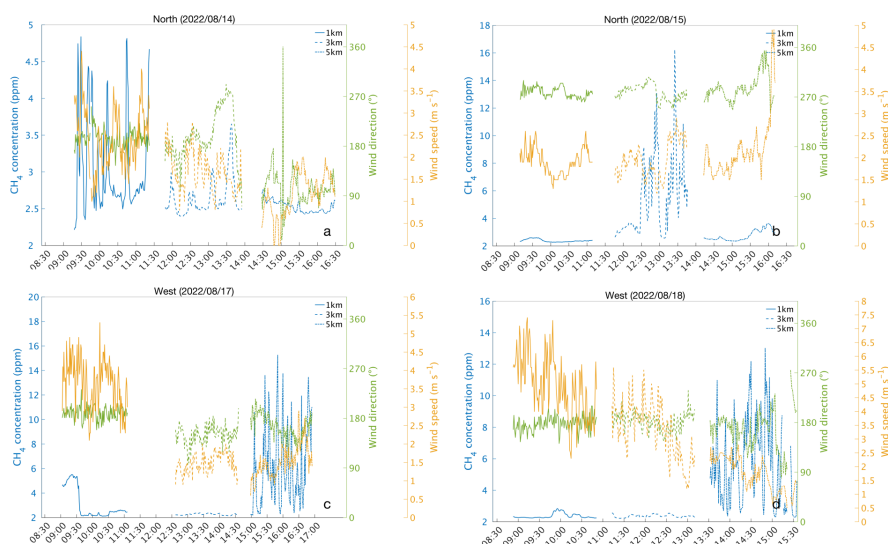
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3. Results and Discussion



247 **3.1 Spatial Distribution Characteristics of CH₄ Concentration Around Coal Mine**

248 Time series of CH₄ concentration, wind speed, and direction at 1 km, 3 km and 5 km north of CM-
 249 A are given in Figure 9a and Figure 9b. The wind direction blew from CM-A towards the observations
 250 (between 150° and 210°) 59.2% of the time, with only one day observed at 1km north (August 15) with
 251 a significant amount of wind from the west (between 240° and 300°) 92.8% of the time. Consistent with
 252 CM-A being the major source at 1km, when the wind blew from the south, the CH₄ concentration
 253 (3.45±0.79 ppm) was both higher and had a larger variation than when the wind blew from the west
 254 (2.40±0.17 ppm) which was similar to background conditions. This is consistent with there being no
 255 known significant sources to the west from this observation location, as shown in Figure 2. Similarly,
 256 under faster than average wind conditions from the direction of CM-A (on August 21 the mean wind was
 257 5.70 m s⁻¹ with 14.9% of observations faster than 7 m s⁻¹), the observed concentrations were slightly
 258 lower, yet similarly variable (3.17±0.82 ppm). All of these findings are consistent with transport
 259 dominating the concentrations at 1 km north, and that high frequency wind and concentration
 260 observations are required in tandem to compute the required spatial gradients in the CH₄, otherwise there
 261 is no basis to objectively separate the effects of the emitting region (CM-A) from the background.



262
 263 **Figure 9. Time series of CH₄ concentration (ppm) (blue), wind speed (m s⁻¹) (yellow) and wind direction (°)**
 264 **(orange lines) measured at 1 km (solid), 3km (dashed) and 5km (dash-dot) located north (top) and west**
 265 **(bottom) of CM-A on four different days.**

266 A similar set of findings were observed at 3 km north, while 5 km north is generally similar to the



267 background. At 3 km north, when the wind was from the south (59.3% of data), the concentration was
268 lower and more variable (3.16 ± 1.48 ppm, with 78.7% of observations below 3.0 ppm) than at 1 km north,
269 consistent with advection from CM-A and a relatively stable atmosphere with a small contribution from
270 diffusion between the plume and the background. When the wind blew from other directions, the
271 distribution of concentrations broadened considerably, with a range from background (2.25 ppm) through
272 extremely polluted (16.2 ppm). One subset of this was observed on August 15 (observed over a total of
273 61 mins of observations, 6.68% of the total observations at 3 km north) when the wind was from the west
274 and slow, where the concentration was (5.44 ± 2.82 ppm), as depicted in Figure 9. The data on this day
275 aligned with the presence of a major highway west of the observation site, which was observed in-person
276 to have heavy traffic consisting of vehicles carrying coal (which could still be outgassing) as well as
277 others powered by compressed natural gas (CNG). At 5 km north the overall concentration (2.40 ± 0.28
278 ppm) was generally lower than at 3 km and had much lower variability, consistent with background CH₄.

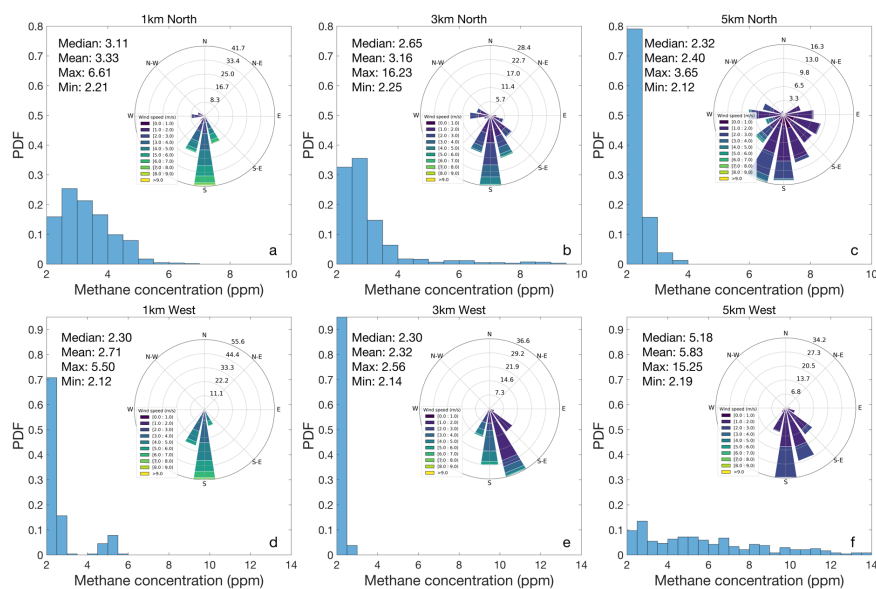
279 Time series of CH₄ measured at 1 km, 3 km and 5 km west of CM-A and corresponding wind
280 direction and speed are given in Figure 9c and Figure 9d. Overall, the main wind direction is from the
281 south 98.4% of the time at 1 km, 74.5% of the time at 3 km, and 70.2% of the time at 5 km, and the wind
282 speed was very high when measuring CH₄ at 1 km west, with an average value of 4.28 ± 1.13 m s⁻¹ and a
283 maximum of 7.4 m s⁻¹. This set of findings is consistent with clean upwind sources. Accordingly at 1 km
284 west, the observed CH₄ concentration was slightly higher than background and had similar variability to
285 1 km and 3 km north (2.71 ± 0.94 ppm and 86.5% of the data below 3 ppm). At 3 km west, CH₄ was
286 observed to be similar to the background (2.32 ± 0.09 ppm). The only exception was found at 1 km west
287 between 9:00 am and 9:30 am on August 17, in which all of the observations were greater than 4 ppm.
288 Since the areas to the west from 1 km west contains mostly farmland, there was no expected strong source
289 of CH₄, as shown in Figure 2. This indicates that during this special short time, the observed slow increase
290 and rapid fall-off in CH₄ concentration must be due an unidentified source, or a change in the boundary
291 layer or vertical mixing structure.

292 Following this, it was anticipated that the 5 km west site would exhibit background types of
293 conditions, however the observed data deviates significantly. Wind speed was low (1.63 ± 0.54 m s⁻¹,
294 maximum 3.0 m s⁻¹), CH₄ concentration was both very high and exhibited substantial temporal
295 variability (5.83 ± 2.99 ppm, 66.7% exceeding 4 ppm, and peak of 15.3 ppm), and 70.2% of the
296 observations were from the south as demonstrated in Figure 10d,e,f. From Figure 1, it can be seen that

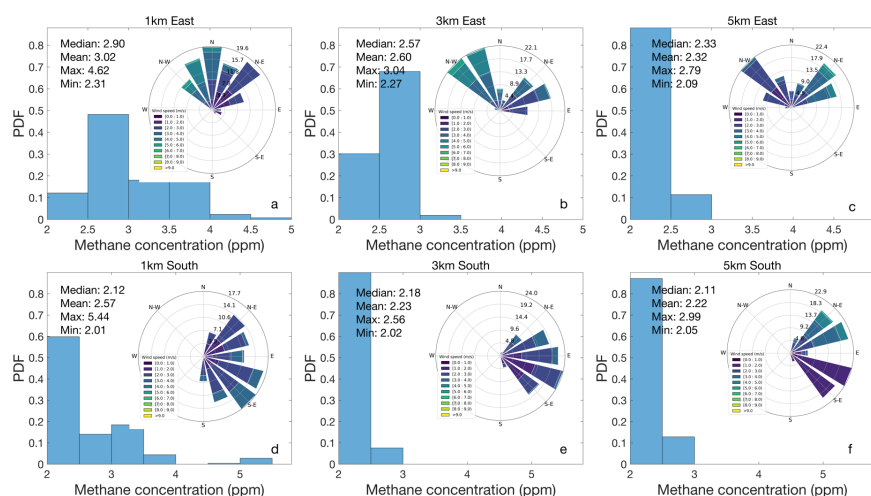


297 there is another coal mine (CM-B) located about 1 km away from the 5 km west measurement point, to
 298 its southwest, although CM-B has an annual production of about 3 million tons (smaller than CM-A) and
 299 not considered to be high gas (like CM-A), and therefore was not previously considered important. The
 300 overlap of high concentrations with low a priori emissions, suggests that formal attribution is essential
 301 to quantitatively confirm whether CM-B is the source responsible for both typical conditions at 5 km
 302 west, as well as the long-range transport event at 1 km west.

303 CH₄ concentrations and wind observations in all directions except to the west, and except for the
 304 small number of special events documents above, exhibit PDFs that show there is a decrease in
 305 concentration the further the distance from CM-A (Figures 10 and 11), indicating that CM-A is consistent
 306 with the major sources in these regions. These decreases are observed in terms of the median, mean,
 307 distribution width, and percentage over 4.0 ppm all decreasing from 1 km north to 3 km north and again
 308 from 3 km north to 5 km north.



309
 310 **Figure 10. Probability density map for CH₄ concentration and wind rose measured at 1 km north (top left),**
 311 **1 km west (bottom left), 3 km north (top center), 3 km west (bottom center), 5 km north (top right), and 5 km**
 312 **west (bottom right) of CM-A.**



313

314 **Figure 11. Probability density map for CH₄ concentration measured at 1 km, 3 km and 5 km east and south**
 315 **of CM-A and corresponding wind rose chart.**

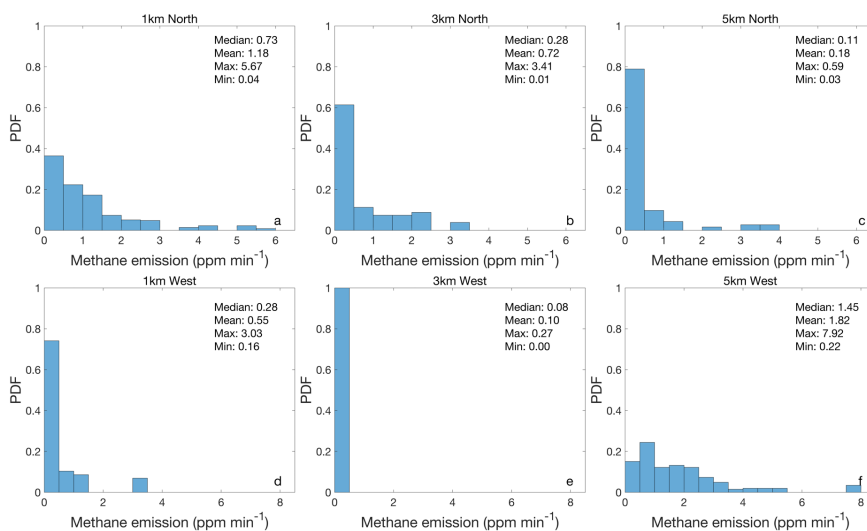
316 The observed CH₄ concentration gradient as one moves westward from CM-A is inconsistent with
 317 the other ordinal directions (Figure 10d, e, f). While there was a small decrease in the mean and
 318 distribution breadth from 1 km west to 3 km west, there was a large increase in the median, mean,
 319 distribution width, percentage over 4.0 ppm from 3 km west to 5 km west. Furthermore, the data at 5
 320 km west was found to be skewed differently than at the other sites, with approximately 70% of the data
 321 greater than 4.0 ppm. The data clearly indicates that the 5 km west site behaves more like a source region
 322 than even the 1 km north site.

323 3.2 Quantification and Emission Characteristics of CMM

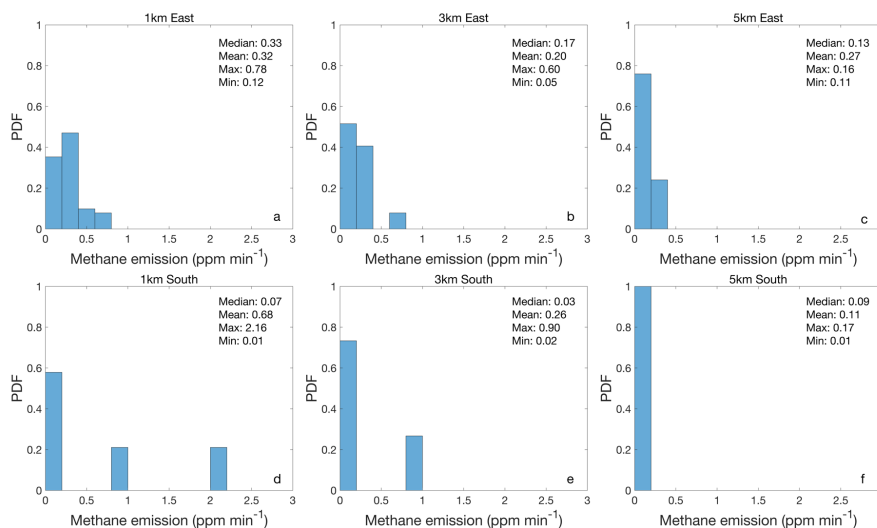
324 The CH₄ emissions have been computed at each of the observation points, with 25.7% of observations
 325 yielding emissions results. The PDFs of the CH₄ emissions (Figures 12 and 13) reveal that the three
 326 stations in the north and the 5km west station all are relatively high and variable, while the remainder are
 327 relatively low and non-variable. Among all the CH₄ emissions results, the highest median, mean,
 328 maximum, and breadth of the distribution are all observed at 5km west. The 3km south location has the
 329 lowest emissions of all points observed (by median), with a respective median, mean, maximum, and
 330 percentage greater than 1.0ppm/min of (0.03 ppm min⁻¹, 0.26 ppm min⁻¹, 0.90 ppm min⁻¹, 0%) (Figure
 331 13), and is subsequently considered representative of background emissions in this work. It is important
 332 to note that there is no area within this region that has 0ppm/min emissions and that the minimum
 333 concentration on average is about 2.23 ppm (Figure 11), both of which are considered very high or



334 polluted compared with most other current studies (Irakulis-Loitxate et al., 2021; Sadavarte et al., 2021).



335
336 **Figure 12. Probability density functions (PDF) of computed CH₄ emissions located at 1 km north (a), 3 km**
337 **north (b), 5 km north (c), 1 km west (d), 3 km west (e), and 5 km west (f) of CM-A, including median, mean,**
338 **maximum, and minimum statistics.**



339
340 **Figure 13. Probability density functions (PDF) of computed CH₄ emissions located at 1 km east (a), 3 km east**
341 **(b), 5 km east (c), 1 km south (d), 3 km south (e), and 5 km south (f) of CM-A, including median, mean,**
342 **maximum, and minimum statistics.**

343 The spatial distribution Characteristics of the CH₄ emissions is similar to that of the CH₄
344 concentration observations (Figure 12). First, there is a decrease as one moves northward along the axis
345 away from CM-A, with the median, mean, maximum, and percentage of emissions greater than 1.0 ppm



346 min^{-1} at 1 km north (0.73 ppm min^{-1} , 1.18 ppm min^{-1} , 5.67 ppm min^{-1} , and 42%) all larger than at 3 km
347 north (0.28 ppm min^{-1} , 0.72 ppm min^{-1} , 3.41 ppm min^{-1} , and 29%). The values at 3 km north are also
348 larger than those at 5 km north, which respectively are 0.11 ppm min^{-1} , 0.18 ppm min^{-1} , and 0.59 ppm
349 min^{-1} , and 0%. The subset of emissions under low wind speed conditions exhibited a larger decline from
350 1 km to 3 km and from 3 km to 5 km. The observations are further consistent with transport from a single
351 dominant source located at CM-A being the primary driving factor, and diffusion from other industrial
352 sources in Changzhi city center being a secondary factor.

353 Consistent with there being few to no sources impacting the 1 km west and 3 km west sites, except
354 for considerably less transport from CM-A the computed PDFs at these sites (Figure 12) demonstrate
355 low emissions and low variability, with the respective median, mean, maximum, and percentage of
356 emissions greater than 1.0 ppm min^{-1} at 1 km west being 0.28 ppm min^{-1} , 0.55 ppm min^{-1} , 3.03 ppm
357 min^{-1} , and 16% and at 3 km west being even lower (0.08 ppm min^{-1} , 0.10 ppm min^{-1} , 0.27 ppm min^{-1} , and
358 0%). However, the CH_4 emissions computed at 5 km west were the highest and most variable of all results
359 computed in this work, with the respective statistics being 1.45 ppm min^{-1} , 1.82 ppm min^{-1} , 7.92 ppm
360 min^{-1} , and 60%. Furthermore, the skewness of the distribution at 5 km west (which has 30% of the CH_4
361 emissions above 2.0 ppm min^{-1}) is much larger than at 1 km north (which only has 15% of emissions
362 above 2.0 ppm min^{-1}). Combining these pieces of information, at first look it seems that the site at 5 km
363 west is not related to the emissions from CM-A, or at best are a mixture of emissions from CM-A and
364 those at another site, herein proposed to be CM-B. The remainder of this study focuses on disentangling
365 and attributing contributions from CM-A and CM-B at 5 km west, with the observations at the remaining
366 sites ruled out in terms of having a contribution from CM-B.

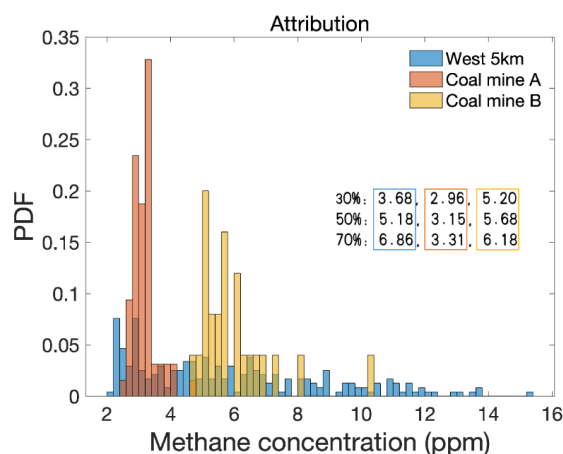
367 **3.3 Attribution of CH_4 Emissions**

368 This work applied the 2-box model at the 5 km west site and quantified the contribution of both
369 CM-A and CM-B emissions to the observed CH_4 concentration distributions as given in Figure 14. First,
370 the results of the 2-box model produce PDFs which overlap with the overall observed CH_4 PDF,
371 indicating that the results are reasonable. Second, space of the emissions computed from the two different
372 two coal mines do not overlap, and cover two independent portions of the observed CH_4 PDF. Specifically,
373 the 30%, 50%, and 70% values of CH_4 concentration observed at 5 km west are 3.68 ppm, 5.18 ppm, and
374 6.86 ppm respectively. The emissions from CM-A yield a CH_4 concentration less than 4 ppm most of
375 the time, with a 30%, 50%, 70%, and maximum concentration of 2.96 ppm, 3.15 ppm, 3.31 ppm, and



376 4.60 ppm, while the emissions from CM-B yield a CH₄ concentration more than 5 ppm most of the time,
 377 with a minimum, 30%, 50%, 70%, and maximum concentration of 4.76 ppm, 5.20 ppm, 5.68 ppm, and
 378 6.18 ppm.

379 Overall, the emissions from CM-B cover well the observed concentration values from the range of
 380 50% to 70%, with a single high value around the 90% value, while the emissions from CM-A cover well
 381 the observed concentration values in the range from 10% to 30%. One weakness is that the length of
 382 observations is not as comprehensive as at the other sites, and therefore it is possible that had more
 383 observations been made, the contributions from CM-B would have filled more of the space between the
 384 70% and 90% levels, and some combination of sources from CM-A and CM-B would have better filled
 385 the space between the 30% and 50% levels. The results indicate to a high degree of certainty that the
 386 emissions from the two respective coal mines are distinct, with CM-A the source of emissions in the
 387 lower range of the concentration distribution and CM-B the source for emissions in the higher
 388 concentration range, covering values in the middle and upper range. Improvements in modeling,
 389 additional observations, considering possible contributions from additional missing sources, and
 390 consideration of longer-range transport could add further improvement and better explore the
 391 intermediate range of observed concentrations.



392
 393 **Figure 14.** The PDFs of CH₄ concentration measured at 5 km west (blue) and simulated using the 2-Box model
 394 under conditions when the source is CM-A (red), and when the sources is CM-B (orange), including
 395 representative 30%, 50%, and 70% bounds are in.

396 **3.4 Policy-Relevant Emissions**

397 In order to compare the values of CH₄ emissions from the Shanxi coal mines computed in section



398 2.7, the units (ppm min^{-1}) are transformed into units of (kg h^{-1}) via a conversion factor based on equations
399 (7) and (8). This conversion increases the overall uncertainty, since since it involves approximations of
400 the area swept, the boundary layer height, and and other uncertainties. In this study, average CH_4
401 emissions from CM-A and CM-B are $13670 \pm 7400 \text{ kg h}^{-1}$ and $5070 \pm 2270 \text{ kg h}^{-1}$, and the CH_4 emissions
402 range is from 200 kg hr^{-1} to 67700 kg h^{-1} and 430 kg hr^{-1} to 15300 kg h^{-1} repectively (Table 2). Both mines
403 display a fat tail distribution, with respective 25th, median, and 75th percentile values of [870, 7500, 38700]
404 kg h^{-1} and [431, 1590, 7000] kg h^{-1} . respectively. These findings are demonstrated to be higher than CH_4
405 emissions from equivalent oil and gas operations in the USA (Chen et al., 2022), with one site being
406 roughly double and the other similar to and slightly lower than day-to-day emissions inverted over 5-
407 years from TROPOMI (Hu et al., 2024) over the same region. This is consistent with the fact that the
408 results herein target very high frequency and spatially confined emissions, while satellites provide day-
409 to-day values over a larger pixel area, as well as associated significant uncertainties involved in
410 converting from ppm to kg. In specific, at CM-A, the minimum (200 kg h^{-1}) and maxmum (67700 kg h
411 $^{-1}$) values of CH_4 emissions results are larger than the respective minimum (8 kg h^{-1}) and maxmum (37300
412 kg h^{-1}) values of CH_4 emissions inverted from TROPOMI, with the statistical values roughly double,
413 while at CM-B, the minimum (430 kg h^{-1}) value of CH_4 emissions results is greater than the minimum
414 (20 kg h^{-1}) from TROPOMI, the maxmum value (15300 kg h^{-1}) of CH_4 emissions is less than the
415 maximum (37300 kg h^{-1}) from TROPOMI, and the statistical values are slightly smaller although similar.
416 In this study, observations were made within 1km of the coal mines on a minute-to-minute basis, while
417 the the TROPOMI observed the xCH_4 over a space scale ($5.5 \times 7 \text{ km}^2$) and on a day-to-day average basis,
418 both of which indicate advantages of higher sampling diversity, especially so at the extreme values
419 observed, when compared with TROPOMI's results. For this reason, it is likely that the sampling time at
420 CM-B was insufficient to fully capture the fat tail of the emissions, since the maxmum is smaller than
421 the maximum estimated CH_4 from TROPOMI (consistent with the limited two days of data at coal mine
422 B), (Table 1). Given that the emissions are generally larger at the higher production coal mine, they are
423 consistent with the concept that over geologically similar environments, higher coal mine production
424 leads to increased CH_4 emissions, although the increase is not linear as most current models assume.
425
426
427
428



429

430

431

432 **Table 2. The CH₄ emissions (kg h⁻¹) of CM-A and CM-B using different observation methods and statistical**

433 **methods**

Coal mines	High frequency ground observation CH ₄ emissions (kg h ⁻¹)								TROPOMI inverted CH ₄ emissions (kg h ⁻¹)			
	1/4 boundary layer				1/10 horizontal wind				Mean±SE	Min	Median	Max
	Mean±SE	Min	Median	Max	Mean±SE	Min	Median	Max				
CM-A	13670±7400	200	7500	67700	19100±9800	70	5790	63800	5500±700	8	2130	24900
CM-B	5070±2270	430	6060	15300	1000±444	30	1600	2850	6200±1000	20	1450	37300

434 **4. Conclusions**

435 This study presents a high frequency ground observation campaign and a new analytical top down
 436 emissions estimation approach to quantify the emissions of CH₄ from a high gas coal mine region with
 437 multiple mines. The base observations are made using a portable greenhouse gas analyzer in connection
 438 with meteorological and other optical measurements. Observations have been made over 15 days at a
 439 frequency of 1 second, at various locations of known distance from an existing high production coal
 440 mine. The high frequency observations are then used in connection with a mass conserving modeling
 441 platform to estimate the CH₄ emission rate. A mass-conserving Two-Box model was used for attribution
 442 analysis in this study. The results show that the spatial characteristics of CH₄ concentration/emissions are
 443 consistent with the distance from a well characterized of single coal mine within 5km distance, and CH₄
 444 emissions demonstrate clear first order effects of both transport and diffusion, with methane emission
 445 rates of 0.73, 0.28 and 0.15 ppm min⁻¹ at 1, 3 and 5 km downwind respectively. At 5 km north the overall
 446 concentration (2.40±0.28 ppm) was generally lower than at 3km and had much lower variability,
 447 consistent with background CH₄, which demonstrate that the CMM emissions mainly affect the
 448 surrounding area with 5km distance. However, the overlap of two coal mines (CM-A and CM-B) have a
 449 far more complex distribution of emissions intensity, ranging as high as 1.82 ppm min⁻¹, which is much
 450 higher than the emissions of single source at any directions. Another, the background concentration of
 451 surface CH₄ in the mining areas is very high compared with other studies, with a value always at and
 452 above 2.23 ppm. Finally, in order to compare these results with results from other parts of the world, the
 453 subset of emissions which successfully underwent attribution were converted into the unit of kg h⁻¹ using
 454 an approximation of the volume swept by the wind and other approximations of the atmosphere. The
 455 resulting values were found to be 13670±7400 kg h⁻¹ and 5070±2270 kg h⁻¹ respectively, which are higher



456 than CH₄ emissions from equivalent oil and gas operations in the USA, and in one case are higher than
457 but in the other case similar to day-to-day emissions inverted from 5-years of TROPOMI over the same
458 region.

459 This work demonstrates that high frequency surface observations of CH₄, in combination with high
460 frequency observations of wind can provide deep insights into emissions by accounting for high
461 frequency changes in space and time at the same time, which tend to be missing from models which used
462 more idealized approaches (such as average plume shapes and sizes, levels of coal production, and
463 interpreting gradients from a small number of fixed images). A significant source of CH₄ emissions from
464 a previously unknown or improperly classified mine may pose a vastly different range of observed
465 concentration as well as computed emissions than expected. The importance of observations at both high
466 frequency and regional spatial coverage are demonstrated, and a set of practical methods that are freely
467 open and can be adopted and modified rapidly are provided. The approach to source attribution used
468 herein can provide insights to policymakers to formulate regional emission control policies and provide
469 a check on or a priori assumption for the new generation of advance satellite-based top-down emissions
470 estimates, while demonstrating that spatial attribution is a critical next-step for satellite approximations
471 and CH₄ control policies.

472 **Data Availability**

473 All underlying data herein are available for access by the editors and reviewers at
474 <https://figshare.com/s/1a393772d7b72ae17e62> and will be made available to the community upon
475 publication.

476 **Author contributions**

477 K.Q., J.B.C. and F.L. designed the research; F.L., C.Y., and Y.S. collected the data; J.B.C. and F.L.
478 analyzed the data; Q.T., W. H. and Q.X. provided the support for data analysis and drawing; Q.H., S.W
479 gave suggestion on running the 2-Box model; F.L. wrote the manuscript with inputs from J.B.C., Q.H.
480 and P.T.; All authors discussed the results and contributed to the final manuscript.

481 **Competing interests**

482 The authors declare that they have no known competing financial interests or personal relationships that
483 could have appeared to influence the work reported in this paper.

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495

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